

1 Application of Landsat-derived vegetation trends over South Africa: Potential for 2 monitoring land degradation and restoration

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9 Abstract

10 Monitoring vegetation change is important because the nature, extent and rate of change in key
11 measures, such as plant biomass, cover and species composition, provides critical insight into broader
12 environmental and land use drivers and leads to the development of appropriate policy. We used
13 Landsat data between 1984 and 2018 to produce a map of Enhanced Vegetation Index (EVI) change
14 over South Africa at 30 m resolution and an interactive web application to make the analysis both
15 globally applicable and locally meaningful. We found an increase in EVI of $0.37 \pm 0.59\% \text{ yr}^{-1}$ (mean \pm
16 standard deviation), confirming global vegetation greening trends observed with lower-resolution
17 satellites. Mesic, productive biomes including the Albany Thicket and Savanna, exhibited the largest
18 greening trends while browning trends were dominant in more arid biomes, such as the Succulent
19 Karoo and Desert. Although overall EVI trends correspond to vegetation index trends derived from the
20 Advanced Very-High-Resolution Radiometer (8 km resolution), the relative scarcity of Landsat data
21 availability during the 1980s is a potential source of error. Using repeat very-high-resolution satellite
22 (ca. 3 m resolution) imagery and ground-based photography as reference, we found good
23 correspondence with EVI trends, revealing patterns of degradation (e.g. woody plant encroachment,
24 desertification), and restoration (e.g. increased rangeland productivity, alien clearing) over selected
25 landscapes. The utility of the EVI trend layer to government and industry for monitoring ecosystem
26 changes will be enhanced by the ability to distinguish climatic from anthropogenic drivers of change.
27 This may be partially achieved through interactive exploration of the EVI trends using the application
28 found here: <http://evitrend.zsv.co.za>

29 **Key words:** *bush encroachment; desertification; ecosystem accounts; net primary productivity; remote*
30 *sensing; repeat photography*

31

32 Introduction

33 Land degradation can be defined as the reduction or loss of ecosystem function caused by both human
34 and non-human processes (Aynekulu et al., 2017). Reduced ecosystem function results in the loss of
35 biodiversity and ecosystem services that sustain livelihoods around the globe, and is estimated to
36 reduce the world's gross domestic product by 10-17% annually (ELD, 2015). Human-induced land
37 degradation results from activities, such as the over-exploitation of natural resources, where
38 vegetation is the primary terrestrial resource (Haberl et al., 2007). This is often associated with a
39 reduction of vegetation cover, loss of natural habitat, and pollution and waste. In some environments
40 an increase in vegetation cover, associated with woody plant thickening (Belay et al., 2013), or the
41 invasion of alien species (Witt et al., 2017), can also result in the loss of ecosystem functionality,
42 although this is not universally the case (Eldridge et al., 2011). Degradation trends can also be a result
43 of climatic variability and trends, particularly in rainfall and temperature, that are unrelated to human
44 activity (Ellis and Swift, 1988). Regardless of the cause, the Sustainable Development Goals
45 acknowledge the global impact of degradation and now include Land Degradation Neutrality (LDN) as
46 an important target with the aim of reversing existing land degradation and avoiding future
47 degradation (Cowie et al., 2018). Achieving LDN requires an account of current ecosystem state, the
48 extent of change relative to a given baseline, and an understanding of the main drivers of degradation.
49 While climate change can exacerbate land degradation (Gonzalez et al., 2012), distinguishing
50 anthropogenic and climatic drivers of vegetation cover change remains difficult.

51 In drylands, which cover 41% of the Earth's terrestrial surface (Maestre et al., 2016), degradation has
52 historically been synonymous with desertification (Bauer, 2016; Sinclair and Fryxell, 1985) and is often
53 associated with the overgrazing of rangelands by pastoralists (Hilker et al., 2014; Mganga et al., 2015).
54 This narrative, however, has long been challenged by tenants of the non-equilibrium theory (Behnke
55 et al., 1994), which holds that, in most dryland environments, vegetation cover is not in equilibrium
56 with human-enforced herbivore pressures (Ellis and Swift, 1988). Rather, climatic variation can
57 override the effect of human-induced degradation, so that vegetation cover can be restored or further
58 degraded during rainfall extremes regardless of herbivore pressures. Despite the important influence
59 that climate has on vegetation cover and composition, recent evidence from southern Africa suggests
60 that the broad-scale recovery of vegetation in some of the region's dryland environments is probably
61 linked more strongly to a release from historically high grazing pressure than to any underlying change
62 in rainfall (Hoffman et al., 2018). Regardless of the drivers of change, distinguishing between
63 trajectories of recovery and degradation is important in order to prioritise areas in need of restoration
64 intervention.

65 Measuring trends in vegetation cover change has been made easier and more affordable at regional
66 scales with the advent of free satellite remote sensing data (Wulder et al., 2012). Various satellite-
67 derived vegetation indices have been used to measure attributes of ecosystem status including
68 grassland and rangeland productivity (e.g. Cawkwell et al., 2016; Svoray et al., 2013) and degradation
69 (e.g. Wessels et al., 2007), forest intactness (e.g. Hansen et al., 2013), land cover change (e.g. Defries
70 and Townshend, 1994) and even ecosystem biodiversity (e.g. Gould, 2000), to name a few. These
71 vegetation indices mostly rely on reflectance in the near-infrared wavelengths because plants have a
72 distinctive spectral signature, characterised by a low reflectance in the visible spectrum, and a high
73 reflectance in the near-infrared (Gates, 2012). A wealth of studies have used low spatial, but high
74 temporal resolution satellite sensors such as the Moderate Resolution Imaging Spectrometer (MODIS)
75 and the Advanced Very-High-Resolution Radiometer (AVHRR) to perform time series analysis of
76 vegetation indices including the Normalised Difference Vegetation Index (NDVI) and the Enhanced
77 Vegetation Index (EVI) as proxies for vegetation cover and primary productivity (Pettorelli et al., 2005;
78 Tucker, 1979; Verbesselt et al., 2010). A wide-spread vegetation greening trend has been observed
79 globally, which has been attributed to atmospheric CO₂ enrichment and increasing rainfall (Hickler et
80 al., 2005; Zhu et al., 2016). Isolating local-scale drivers of change is made difficult by the low spatial
81 resolution of MODIS (250 m) and AVHRR (8 km) imagery. In the context of LDN, useful tools, including
82 Trends.Earth, have been developed to implement systematic accounts of land degradation using
83 MODIS primary productivity trajectories (Gonzalez-Roglich et al., 2019). However, MODIS only extends
84 back to 2000, and human-induced degradation footprints can occur at scales smaller than the 250 m
85 MODIS resolution. Thus, higher resolution imagery might elucidate finer-scale dynamics.

86 Recent advances in cloud computing technology have allowed scientists to utilise the full archive of
87 higher resolution (30 m) Landsat 5, 7 and 8 satellite imagery (extending back to 1984) to perform
88 decadal time series analyses of vegetation cover change (Pasquarella et al., 2016). Remote sensing
89 analysis has been able to move from static, bi-temporal measures of change toward more continuous
90 time series analysis. Applications have included quantifying forest cover change (Hansen et al., 2013)
91 and phenology (Melaas et al., 2013), land cover change (Gong et al., 2019), woody plant encroachment
92 (Venter et al., 2018), riparian zone restoration (Hausner et al., 2018), and surface coal mining (Yang et
93 al., 2018). Such studies provide a distinct advantage over low resolution regional measures of greening
94 and browning trends because they give more nuanced insight into the type and cause of vegetation
95 degradation or restoration.

96 This is particularly important in southern Africa where vegetation greening can signal both restoration
97 and degradation, depending on the local context. For example, the encroachment of woody plants
98 (alien or native) into open areas and an increase in grass cover can both be registered as greening by

99 satellites. However, woody plant encroachment often reduces the grazing capacity of a rangeland,
100 while an increase in grass cover has the opposite effect (Anadón et al., 2014; Macharia and Ekaya,
101 2005). Previous efforts at identifying degradation using satellite remote sensing have focussed on
102 distinguishing climate- versus human-driven browning (Wessels et al., 2007). Such efforts have either
103 relied on expert opinion mapping at administrative unit scales, or have relied on the interpretation of
104 single-temporal NDVI snapshots (Thompson et al., 2009). High-resolution land cover change maps do
105 exist (e.g. GeoTerraImage, 2015) but they rely on a bi-temporal and categorical analysis of change
106 only, that masks gradual changes along a continuum of vegetation cover. Government investment in
107 alien clearing, wetland restoration, land development monitoring, rangeland management, and
108 ecosystem accounts would benefit from monitoring tools that quantify trends in vegetation condition
109 at high-resolution in an interactive environment that also allows for its integration with local
110 knowledge. Furthermore, the Intergovernmental Science-Policy Platform on Biodiversity and
111 Ecosystem Services (IPBES, 2018) has recognized the global importance of improving the detection,
112 monitoring and verification of land degradation trends.

113 To improve upon previous analyses of vegetation cover change over South Africa at low spatial
114 resolutions, we aimed to develop a high-resolution map of vegetation cover change and perform a
115 quantitative and qualitative validation using ancillary satellite and fixed-point repeat photograph data
116 at select locations over the country. We mapped linear trends in annual Landsat EVI time series for
117 South Africa between 1984 and 2018 and hypothesised that the country has undergone a net greening
118 in accordance with previous national-scale studies (Bai & Dent 2007). To make the layer relevant to
119 policy makers and land managers, we have packaged it in an interactive web application that allows
120 users with local knowledge to interpret vegetation dynamics at the landscape-scale. We illustrate this
121 interpretation using very high-resolution satellite imagery and fixed-point repeat photographs for
122 select locations across the country.

123

124 **Methods**

125 *Remote sensing*

126 All remote sensing data collection and analysis was performed using the Google Earth Engine (GEE)
127 JavaScript API, which is a cloud computing platform for earth observation and analysis (Gorelick et al.,
128 2017). We used the near-complete set of USGS Landsat 5, 7 and 8 Surface Reflectance Tier 1 imagery
129 over South Africa at 30 m resolution between 1984 and 2019 (Woodcock et al., 2008). The Landsat
130 data provided by GEE have been pre-processed to orthorectified surface reflectance and have been

131 atmospherically corrected using LEDAPS (Masek et al., 2006). We masked clouds, cloud shadow and
132 snow using the 'pixel_qa' band. Due to slight differences between Landsat 7 and 8 sensors (Holden
133 and Woodcock, 2016), cross-calibration of reflectance values is important when implementing time
134 series analysis (Zhu, 2017). We applied published cross-calibration coefficients to harmonise Landsat
135 8 reflectance values with the other Landsat collections (Roy et al., 2016).

136 The most widely used satellite-derived measure of vegetation growth or vigour is (NDVI devised by
137 Tucker (1979). However, NDVI saturates over high biomass areas and is sensitive to background soil
138 reflectance and atmospheric contamination (Xue and Su, 2017). Despite the disadvantage of being
139 less sensitive to sparse vegetation (Heute et al., 2012), EVI simultaneously corrects for atmospheric
140 and soil effects and does not saturate over high vegetation biomass (Liu and Huete, 1995). Therefore
141 we decided to use EVI even though EVI and NDVI are highly correlated over space and time (Fensholt
142 et al., 2006). EVI is calculated as:

$$143 \quad EVI = G \times \frac{\rho_{nir} - \rho_{red}}{\rho_{nir} + (C_1 \times \rho_{red} - C_2 \times \rho_{blue}) + L}$$

144 where ρ_{nir} , ρ_{red} , and ρ_{blue} represent reflectance in the near-infrared, red and blue wavelengths, and
145 $G=2.5$, $C_1=6$, $C_2=7.5$, and $L=1$ (Huete et al., 2002).

146 *Trend analysis*

147 We derived time-integrated measures of EVI using medoid composites of reflectance values within
148 each year. The medoid is a multi-dimensional analogue of the median which is robust against extreme
149 outliers (e.g. remnant cloudy pixels after cloud masking) and is a better representative of vegetation
150 conditions over the given time period compared to the more commonly used compositing methods
151 such as median or maximum (Flood, 2013). The magnitude and significance of the trend in EVI
152 between 1986 and 2019 was then calculated using the Sen's slope (Sen, 1968) estimator and Mann-
153 Kendall test (Mann, 1945), respectively. The Sen's slope estimator is a non-parametric linear
154 regression that is robust against outliers and skewed data (Wilcox, 2010). We calculated per-pixel
155 percentage change in EVI as:

$$156 \quad \Delta EVI = \frac{m}{c} \times 100$$

157 where m and c are the slope and y-intercept in the regression line defined as $y = mx + c$. Change in EVI
158 is thus expressed relative to the baseline EVI value (y-intercept) for each pixel so that the magnitude
159 of change is comparable across productivity gradients. Although the Sen's slope deals well with
160 outliers, trends with large magnitudes are not necessarily significant because of large variance over
161 time. Thus, the significance of the trend was calculated using the Mann-Kendall S statistic, which is

162 the most common non-parametric method to detect monotonic trends in climatic (e.g. Ahmad et al.,
163 2015) and remote sensing (e.g. de Jong et al., 2011) data. To avoid the pitfalls of setting absolute
164 significance levels (Amrhein et al., 2019), we used the S statistic as a weighting variable to calculate
165 weighted means of EVI trends for spatial aggregations. We did this by scaling the S statistic scores over
166 the country between 0 and 1, and multiplying each pixel's trend value by its scaled S statistic.

167 Mean EVI trend values were calculated for South African biomes and South Africa Vegetation Types
168 (Mucina et al., 2018) grouped into ecological functional groups (Dayaram et al., 2019). South African
169 Vegetation Types are mapped in terms of their historic or pre-European extent, i.e. natural vegetation
170 extent prior to conversion to other land uses. We used a Landsat-derived 2014 land cover map
171 (GeoTerraImage, 1990) to mask out water and non-natural (mine/industrial, plantation, cropland,
172 built up) land cover categories when calculating mean EVI trends. Areas classified as "desert" in the
173 land cover map were not masked because they were found to contain sparse vegetation elements
174 that are of relevance to rangeland and conservation management.

175 *Remote sensing validation*

176 Validating our remote sensing product is made difficult by the fact that there are currently no
177 published maps of decadal-scale ecosystem degradation or vegetation cover trends in South Africa
178 (Driver et al., 2012). Thus, we utilised three reference datasets, namely (1) long-term, low resolution
179 satellite measures of NDVI, (2) very high-resolution satellite red-green-blue (RGB) imagery and (3)
180 fixed-point repeat photography for select locations over the country.

181 NDVI data from AVHRR sensors (in operation since 1981) have been widely used to quantify broad-
182 scale trends in vegetation cover over the globe at a spatial resolution of 8 km (e.g. Cook and Pau, 2013;
183 Fensholt et al., 2009; Gichenje and Godinho, 2018). Unlike the Landsat archive, the AVHRR data have
184 a high temporal frequency and produce annual composites that are less affected by outliers
185 introduced by cloud cover. Landsat data since 1984 have been collected by three different sensors
186 and, without cross-sensor calibration, produce artificial anomalies in reflectance time series (Roy et
187 al., 2016). Although the AVHRR sensors lack reliable calibration devices (Staylor, 1990; Burrell et al.,
188 2018), and have relatively broad spectral bands, which reduce atmospheric correction accuracies
189 (Tanre et al., 1992), they do provide a useful reference dataset for Landsat-derived vegetation indices
190 (Beck et al., 2011). We calculated annual medoid mosaics for the Global Inventory Modeling and
191 Mapping Studies (GIMMS; Tucker et al., 2005) third generation AVHRR NDVI dataset provided by GEE
192 at a spatial resolution of 5 arc minutes between 1984 and 2013 (the AVHRR sensor was discontinued
193 in 2013). The annual Landsat EVI medoids were aggregated to the spatial resolution of GIMMS data

194 using a mean reducer for comparison with GIMMS using a linear regression of all pixel values over
195 South Africa for each year.

196 We performed a further qualitative validation and interpretation of landscape-scale trends using
197 repeat satellite imagery and fixed-point repeat photographs. This validation was neither
198 comprehensive nor quantitative because of the limited availability of metadata associated with
199 validation photos which precluded quantitative estimates of vegetation cover change. They merely
200 provide an analytical tool with which to illustrate what a range of browning and greening trends look
201 like at landscape scales. Very high-resolution (<5 m) RGB satellite imagery from DigitalGlobe was
202 filtered for 2002/4 and 2019 imagery using Google Earth Pro. Based on expert knowledge of known
203 vegetation cover trends over the country, we manually located 14 landscapes expected to cover a
204 variety of EVI trends (Fig. 1A). We manually selected Landsat pixel locations for comparison with
205 visually-interpreted trends visible in DigitalGlobe images. Repeat fixed-point photographs, archived at
206 the Plant Conservation Unit at the University of Cape Town (see <http://rephotosa.adu.org.za>), were
207 filtered for photo pairs in which the repeat photograph was taken after 2010 so as to overlap with the
208 Landsat data (1984-2018). Some historical photographs were taken decades before 1984 and thus
209 comparisons with the Landsat EVI trends assume linear trends in vegetation cover reflected in repeat
210 photographs. For qualitative analysis and interpretation, we selected 8 historical-repeat photo pairs
211 showing a range of vegetation cover trends over different vegetation types. Photo locations (Fig. 1A)
212 reported by photographers using a GPS were visualised in Google Earth Pro using the 3D view mode
213 to digitise the photograph field-of-view over the landscape for comparison with the Landsat EVI trend
214 raster image. One or two pixels were selected over the landscape for EVI trend inspection.

215

216 **Results**

217 *Overall trends*

218 South Africa has undergone a net greening between 1984 and 2018 with a mean EVI trend of $0.67 \pm$
219 $0.77\% \text{ yr}^{-1}$ (mean \pm standard deviation) when quantifying change relative to the baseline EVI value for
220 each pixel (Fig. 1). After weighting trend values by the Mann-Kendall statistic to account for trend
221 significance, the mean value was $0.37 \pm 0.59\% \text{ yr}^{-1}$ which equates to $12.6 \pm 20\%$ increase over the 34
222 years. The mean absolute change in EVI values over the country revealed an increase of 0.02 ± 0.046
223 EVI units yr^{-1} . Large greening trends dominate the central and northern parts of the country while
224 browning trends are evident in parts of the east and over much of the extreme north western part of
225 the country (Fig. 1A). The change in EVI was not evenly distributed across the long-term mean EVI

226 gradient (Fig. 1B) nor the South African biomes (Fig. 1C) or functional vegetation types (Fig. S2).
227 Browning trends were largely restricted to areas with low average EVI (< 0.07), characteristic of the
228 arid and semi-arid areas of the Succulent Karoo and Desert biomes and parts of the Nama-Karoo
229 biome. Greening trends were highest in environments with long-term mean EVI values of ca. 0.2,
230 particularly the Albany Thicket and Indian Ocean Coastal Belt biomes. Greening trends were relatively
231 consistent in more productive, mesic areas, notably the Savanna biome which covers 32% of the
232 country (Fig. 1C).

233 *GIMMS validation*

234 The mean Landsat EVI and GIMMS NDVI values are strongly correlated over South Africa (Fig. 2A).
235 Annual deviations from long-term means for each pixel display positive trends for both Landsat and
236 GIMMS sensors (Fig. 2B). However, Landsat data exhibit large positive biases compared to the GIMMS
237 data during 1985, 1987 and again in 2009 and 2010 (Fig. 2B). This is likely due to the low number of
238 available cloud-free pixels during those years, particularly during the 1980s. This introduces variation
239 in the day-of-year at which the annual medoid value for each pixel was acquired (Fig. S2B). This may
240 also explain the variance in trend values depending on the start and end date of EVI time series (Fig.
241 S3).

242 *Landscape-level validation*

243 We present EVI greening and browning trends over selected landscapes to illustrate different change
244 scenarios outlined in Table 1. Readers can explore trends at this resolution for different localities
245 around the world interactively here: <http://evitrend.zsv.co.za>. Gradual greening trends may be
246 associated with woody plant encroachment into previously grassy areas (Fig. 3A, Table 1), or an
247 increase in the cover of woody plants on previously bare areas, such as eroded gullies in agricultural
248 landscapes (Fig. 3B). Bush encroachment into grassy areas may be associated with a reduction in EVI
249 seasonality over time (e.g. Fig. S9 & S10) due to the perennial phenology of woody vegetation
250 greenness. Greening trends can also be indicative of an increase in herbaceous vegetation cover or
251 vigour while woody plant cover has remained constant (Fig. 3C & D). Here, there is an increase in EVI
252 seasonality due to the senescent phenology of many grass species.

253 Significant browning trends over areas previously covered by woody plants may indicate tree
254 harvesting or clearing (Fig. 4D) and in some cases, the expansion of settlements in response to
255 population growth (Fig. 4A). In areas dominated by herbaceous vegetation, browning may be
256 indicative of overgrazing, which can result in a reduction in grass vigour and basal cover (Fig. 4A). In
257 arid landscapes, where vegetation cover is generally low, browning trends might indicate a loss in

258 vegetation vigour (Fig. 4C). However, the relative change in arid areas should be viewed in terms of
259 the magnitude of absolute changes, which may be very low (e.g. 0.01 EVI units in Fig. 4C). Browning
260 trends are, therefore, not necessarily an indication of management-induced degradation but when
261 viewed in context of rainfall trends, may be a product of long- or short-term drought conditions. The
262 clearing of alien vegetation is an example of where restoration activity might exhibit a browning trend
263 in the vegetation EVI (Fig. 4D). Abrupt browning, typically caused by fire (e.g. Fig. 4D), land cover
264 change, or deforestation (see supplement for illustrations), are smoothed by the long-term EVI trend
265 layer.

266

267 **Discussion**

268 Recent advances in computing power and the availability of large data have expanded the utility of
269 satellite remote sensing in providing repeatable and standardised monitoring of long-term
270 environmental trends (Pettorelli et al., 2014; Zhu, 2017). Despite this, there have been few national
271 level analyses of how remotely-sensed indices of net primary productivity (NPP) have changed over
272 time in South Africa. Using Landsat data, available from the Google Earth Engine platform, our results
273 show an overall net increase in the enhanced vegetation index (EVI) of about $0.37 \pm 0.59 \text{ \% yr}^{-1}$ over
274 South Africa for the period 1984-2018. The general increasing trend corresponds to evidence at the
275 global-scale for a net greening during the past three decades (Zhu et al., 2016). Increasing trends were
276 greatest in the more humid eastern parts of the country where long-term, mean annual rainfall totals
277 are above 200 mm. Negative trends were apparent in the arid and hyper-arid western parts of the
278 country. This adds a different perspective on change, in both its temporal and spatial pattern, from
279 earlier national-level assessments that were conducted over different time spans. For example, using
280 NDVI values from fortnightly images at 8 km spatial resolution derived from the AVHRR sensor and
281 produced by the Global Inventory Modelling and Mapping Studies (GIMMS) group, Bai and Dent
282 (2007) suggested that for the period 1981-2003, net primary productivity had decreased over ca. 30%
283 of South Africa. Their results also showed that the decrease was greatest in the humid and sub-humid
284 eastern areas of South Africa and smallest in the more arid and hyper-arid, western parts of the
285 country where NDVI values had generally increased over their study period. Although these findings
286 cannot be directly compared to our results, the differences in spatio-temporal patterns of change
287 underscore some of the difficulties in using remotely-sensed indices of NPP which can be strongly
288 influenced by the start and end dates of the time series, the type of index used as well as by the spatial
289 and temporal resolution of the data (Wessels et al., 2012).

290 While an analysis of the climatic and anthropogenic drivers of greening and browning trends was not
291 within the scope of this study, we found that using fixed-point repeat photographs and high resolution
292 satellite images as a form of qualitative validation gave insight into the attribution of degradation or
293 restoration signals. This is particularly the case when interpreted relative to the local context of
294 change. For example, much of the increase in EVI recorded in this study for the Albany Thicket, Indian
295 Ocean Coastal Belt and Savanna biomes can probably be attributed to an increase in woody plant
296 cover. Here, assigning the greening trend as degradation or restoration is complex given that woody
297 plant encroachment can be detrimental for livestock agriculture, and yet beneficial for carbon
298 sequestration (Foden et al., 2019). This encroachment or thickening of woody plants has been widely
299 reported in these biomes (O'Connor et al., 2014; Skowno et al., 2017) although reasons for the
300 increase differ. Widespread abandonment of cultivation (Shackleton et al., 2019), a change in burned
301 area (Venter et al., 2018) or fire frequency (Singh et al., 2018), an increase in the concentration of CO₂
302 in the atmosphere (Bond and Midgley, 2000) and an increase in invasive alien plants (Nkambule et al.
303 2017) or even plantation forestry (Turpie et al. 2007) are all potential explanations for the pattern. In
304 contrast, an increase in the cover of herbaceous grasses and not woody plants is probably responsible
305 for the greening of the wide ecotone between the Nama-Karoo and Grassland biome in the central
306 part of the country (Masubelele et al., 2015). For the Fynbos biome, the increase in EVI values within
307 the Cape Fold Mountains as well as the more arid Kamiesberg and Roggeveld high-lying areas are more
308 difficult to explain, especially in light of the severe drought that has occurred across much of the winter
309 rainfall region over the period 2015-2017.

310 There has been a reduction in EVI recorded for the arid and hyper-arid north western parts of the
311 winter rainfall region. Results from several repeat photography studies in the same region report an
312 increase in the cover of perennial plants over landscape units composed of plains and ephemeral
313 streams (Hoffman & Rohde, 2011) although these pre-date the more recent severe drought
314 experienced in the winter rainfall region. Other areas which exhibited a browning trend in South
315 Africa, such as parts of KwaZulu-Natal and the southern section of Kruger National Park, can be
316 explained by factors other than climate. For example, Jewitt et al., (2015), showed that the expansion
317 of agriculture, plantation forestry, settlements, dams and mines in KwaZulu-Natal were largely
318 responsible for the transformation of 1.2% of the province's natural landscapes each year over the
319 period 1994 to 2011. Therefore, the decrease in EVI values over parts of the province has likely
320 occurred as a result of this increase in anthropogenic activities. A different explanation is likely for
321 the browning trend recorded for Kruger National Park which is best understood in terms of the
322 reported impact that megaherbivores have had on vegetation biomass in the park (Pellegrini et al.,
323 2017).

324 To generate or map more generalized explanations for greening and browning trends over South
325 Africa will require a more comprehensive validation approach. Due to the limited availability of
326 ground-truth datasets, we resorted to a nuanced and qualitative validation using repeat fixed-point
327 ground and satellite images. The repeat photograph epochs extend beyond the Landsat time frame
328 (1984-2018) and therefore one has to assume linearity in trends that predate this. The use of GIMMS
329 dataset was an attempt to supplement the qualitative approach and we found good correspondence
330 with the Landsat EVI trends. Apart from merely validating the accuracy of EVI trends, making value
331 judgements about the implications of greening or browning trends by assigning them to degradation
332 or restoration categories requires more information on the drivers of change. For example, there are
333 many methods to distinguish climatic from human-induced drivers of browning or greening using
334 remote sensing data (e.g. Abel et al., 2019; Burrell et al., 2017; Horion et al., 2016; Wessels et al.,
335 2007).

336 Remote sensing with both high- (e.g. Gonzalez et al., 2012) and low-resolution (e.g. Bai et al., 2008)
337 satellite-derived indices of vegetation productivity can be used to monitor changes in the environment
338 over relatively large spatial and temporal scales. However, these indices usually provide little more
339 than a general screening tool and cannot adequately provide information on the type of degradation
340 or improvement in a landscape. More detailed information on the nature of the changes observed
341 and the reasons for the changes at smaller spatial and temporal scales can often only be inferred from
342 extensive ground-truthing, from the observer's knowledge of the region, or from published studies of
343 the area. Even then, it is difficult to include a temporal interpretation of the trajectories observed
344 since very few locations have independent, long-term monitoring data sets to draw on. Data are also
345 very seldom collected at time-scales relevant to those covered by satellite-derived data and
346 mismatches between satellite spatial resolution and field data further complicate validation. In the
347 approach presented here, however, two novel contributions are made.

348 Firstly, when used in combination with other remotely sensed measures, such as those resulting from
349 repeat very-high-resolution satellite and ground photography, a more insightful and detailed
350 interpretation of both the magnitude and direction of change, recorded by Landsat-derived EVI, is
351 possible. We found good correspondence with EVI trends showing patterns of degradation (e.g.
352 woody plant encroachment, desertification) and restoration (e.g. increased rangeland productivity,
353 alien clearing) in South Africa. The inclusion of fine-scale measurement, such as is presented here,
354 enables the observer to check on the broad-scale pattern provided by the satellite-derived vegetation
355 index and, in some cases, to better understand local-level processes responsible for the change. The
356 different spatial scales used in such an analysis also provide a more complete understanding of
357 environmental change than is usually provided by analyses carried out at one spatial scale only.

358 A second contribution of this study relates to the accessibility and usability of satellite-derived
359 vegetation indices. Cloud computing platforms such as Google Earth Engine (Gorelick et al., 2017) are
360 making data collection, analysis and visualisation more accessible to scientists and non-specialists
361 alike. The delivery of remote sensing products in web applications as opposed to static maps is
362 becoming more common (e.g. Donchyts et al., 2016; Huntington et al., 2017). Users are allowed to
363 interact with mapped data at their highest resolution, affording more locally-relevant insights than
364 those gained from a printed map in a publication. The EVI trend data presented here can be explored
365 at full resolution in our web application (<http://evitrend.zsv.co.za>). We intend for users to potentially
366 decipher landscape-scale degradation and restoration trends based on local knowledge with the aid
367 of plotted EVI trends for user-selected pixels over user-defined time periods (Fig. S3). With access to
368 the EVI time series, users might be able to also identify periods of rapid change caused by fire or
369 vegetation clearing events that are lost when summarising the time series using a linear regression
370 line. In this manner, accuracy errors introduced by the low availability of Landsat data during the
371 1980s, and variation in trend values depending on start and end dates for EVI time series can be further
372 interrogated. The biases in the day-of-year (Fig. S1B) during the 80s should inform user interpretation
373 of trends depending on local vegetation phenology. For instance, the large bias toward Jan/Dec
374 medoid EVI values in 1986-7 may introduce false-negative long-term trends in summer rainfall areas
375 where EVI values are expected to be relatively high in Dec/Jan. Conversely, false-positive long-term
376 trends might be more likely in winter rainfall areas.

377 Governments and industry partners within, but not limited to, South Africa would benefit from a cost-
378 effective tool for measuring and monitoring ecosystem management and restoration interventions
379 such as vegetation recovery after alien clearing or mine restoration, or the impact of de-stocking on
380 rangeland vegetation cover. The EVI trend analysis presented here has wide application and relevance
381 to a range of land users and decision making interested in incorporating information on the state and
382 trend of ecosystems into land use management or economic enterprises. For example, livestock or
383 wildlife managers might be able to assess seasonal patterns in vegetation productivity in the context
384 of long-term trends in productivity at the grazing camp or farm-scale. Objective and repeatable
385 measures of ecosystem extent and condition are essential for the successful implementation of
386 ecosystem accounts (United Nations 2014, Driver *et al.* 2015). This EVI trend analysis is a step towards
387 the development of an ecosystem condition index that meets these criteria and with a spatial
388 resolution relevant to ecosystem accounting. By applying time series processing to intra-annual
389 vegetation indices, real-time monitoring of land clearing is already being implemented using a fusion
390 of Landsat and Sentinel satellite data to monitor tropical deforestation (Reiche et al., 2018).
391 Translating such approaches to more arid environments should be possible with relevant ground-truth

392 data. Finally, by supplementing the EVI trend data with time series analyses aimed at disaggregating
393 climatic from anthropogenic drivers (Burrell et al., 2017), it will be possible to understand their relative
394 contribution to the observed trend. More importantly it will highlight where land use interventions
395 are likely to have greater impacts on ecosystems and may contribute to global ecosystem accounting
396 initiatives such as those outlined in the IPBES framework (Diaz et al., 2015).

397

398 **Conclusion**

399 While we observed an overall vegetation greening over South Africa between 1984 and 2018, we
400 found the direction and magnitude of this trend to be spatially variable at regional and landscape
401 scales. We suggest that while aggregating results at regional-scales is interesting, the ability to derive
402 locally relevant insights is only realised at higher resolutions, as illustrated by validation and
403 interpretation of the EVI trend layer with fixed-point repeat photography and very-high-resolution
404 satellite imagery. The delivery of the EVI analysis and results in an interactive web application should
405 allow users to gain local-scale insight at 30 m resolution, interrogate per-pixel EVI time series to
406 identify sources of error, and to potentially integrate knowledge of local ecosystems to distinguish
407 restoration from degradation signals in the EVI time series. The utility of the EVI trend layer to
408 government and industry for monitoring ecosystem changes will be enhanced by the ability to
409 distinguish climatic from anthropogenic drivers of change. This may be soon realised by fusing Landsat
410 and Sentinel satellite data, allowing for high resolution monitoring in near real-time.

411

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415 © SANBI and licensed under Creative Commons licence CC-BY. More information can be found here:
416 <http://rephotosa.adu.org.za/>

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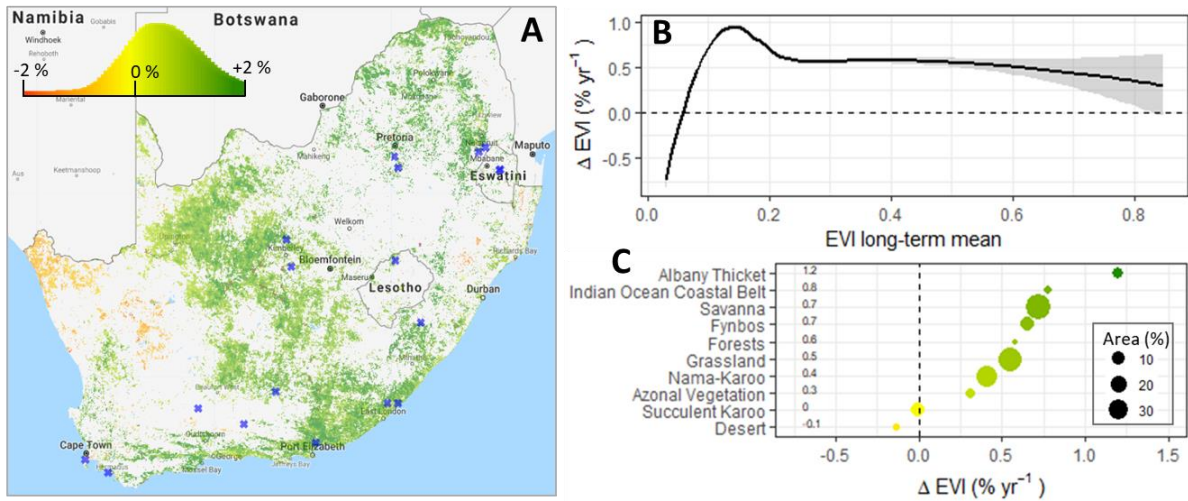
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696 Table 1. Description of greening and browning scenarios illustrated in Fig. 3 and 4. Repeat photographs
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Trend	Biome	Figure reference	Coordinates	Scenario	Photo attribution
Greening	Grassland	3A	-30.526, 29.050	Bush encroachment by <i>Acacia mearnsii</i> and <i>Leucosidea sericea</i>	rePhotoSA image #4
	Grassland	3B	-28.873, 28.242	Increase of woody plant cover in erosion gullies - potential restoration signal	DigitalGlobe
Browning	Nama-Karoo	3C	-32.331, 24.441	Increase in grass cover and vegetation vigor	rePhotoSA image #3
	Savanna	3D	-28.308, 24.779	Woody cover loss on slopes but increase in grass vigor on plains	DigitalGlobe
	Savanna	4A	-26.502, 31.596	Clearing of woody plants for establishing residential buildings	DigitalGlobe
	Nama-Karoo	4B	-33.174, 23.430	Right-hand side of the fence line reflects decline in cover of the dominant shrub, <i>Portulacaria afra</i> (spekboom)	rePhotoSA image #3
	Fynbos	4D	-34.083, 18.402	Manual clearing of alien pine trees from the crest of the slope.	rePhotoSA image #5
	Succulent-Karoo	4C	-28.589, 20.156	Gradual loss of vegetation vigor with potential loss of perennial herbs	rePhotoSA image #3

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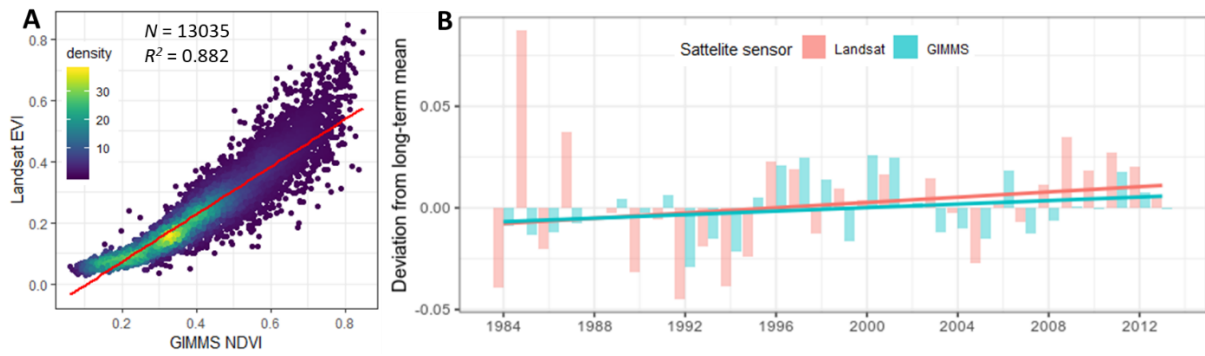
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701 **Figure 1.** Average annual change in EVI between 1984 and 2018 (A). Change is expressed as a
 702 percentage of the baseline (1984) EVI value and is derived from the slope and y-intercept of the
 703 regression line through annual medoid composites of Landsat EVI. Non-significant trends are masked
 704 ($p>0.05$) and points of local-scale validation are shown in blue. The colour scale depicts the distribution
 705 of the data in A prior to significance masking. A loess regression line between the annual average
 706 change and the long-term mean is plotted with a 95% confidence interval ribbon based on a 10 x 10
 707 km aggregation grid (B). Mean annual EVI change values are presented for the major biomes in South
 708 Africa after masking any non-natural land use (C). Data points are sized relative to the area covered
 709 by the respective biome and mean values have been added to the left of the plot space.

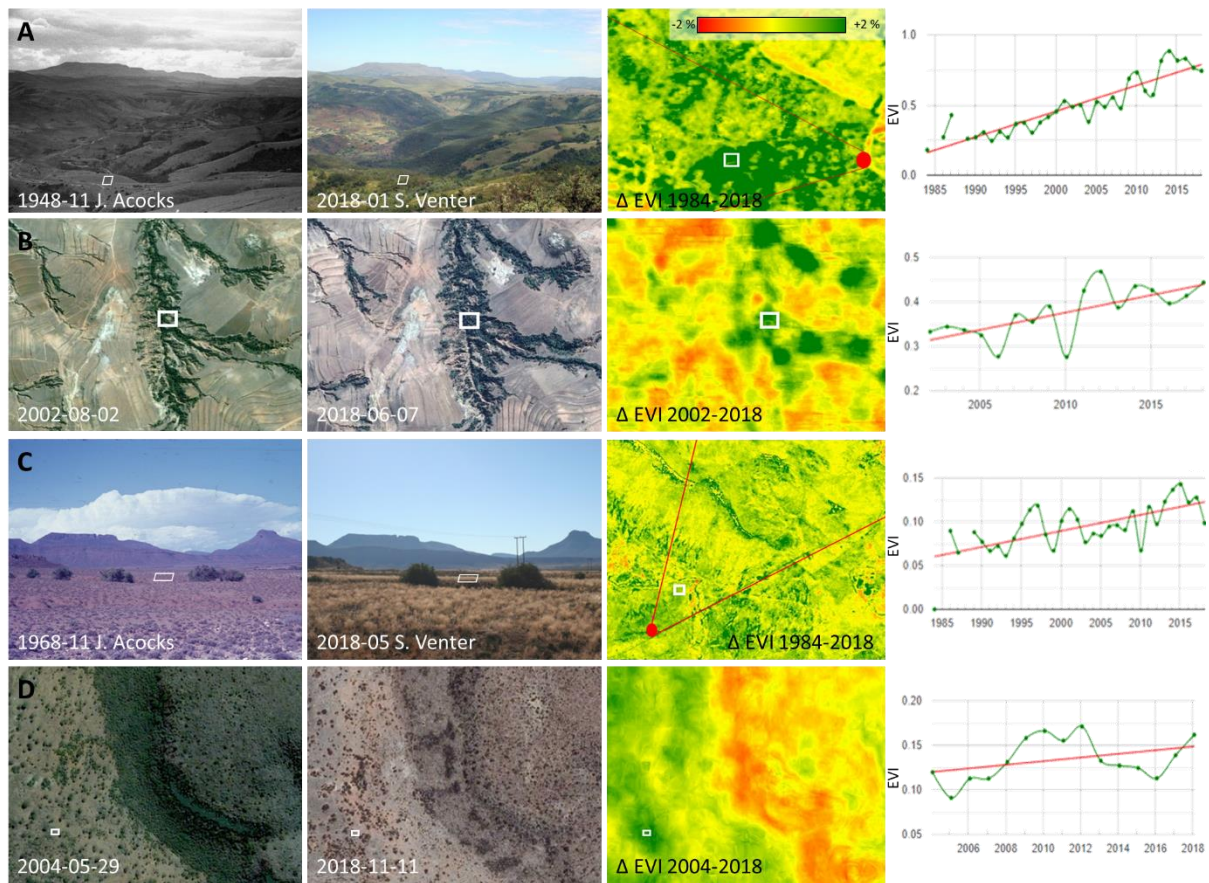
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712 **Figure 2.** Comparison of mean EVI and NDVI values derived from Landsat and GIMMS sensors,
 713 respectively, between 1984 and 2013 for all GIMMS pixels within South Africa (A). The density of data
 714 points is indicated by the colour scale and the sample size (N) and R² value for the linear trend line
 715 (red) are presented. The annual deviation from the long-term mean is plotted in B with a linear trend
 716 line for each sensor.

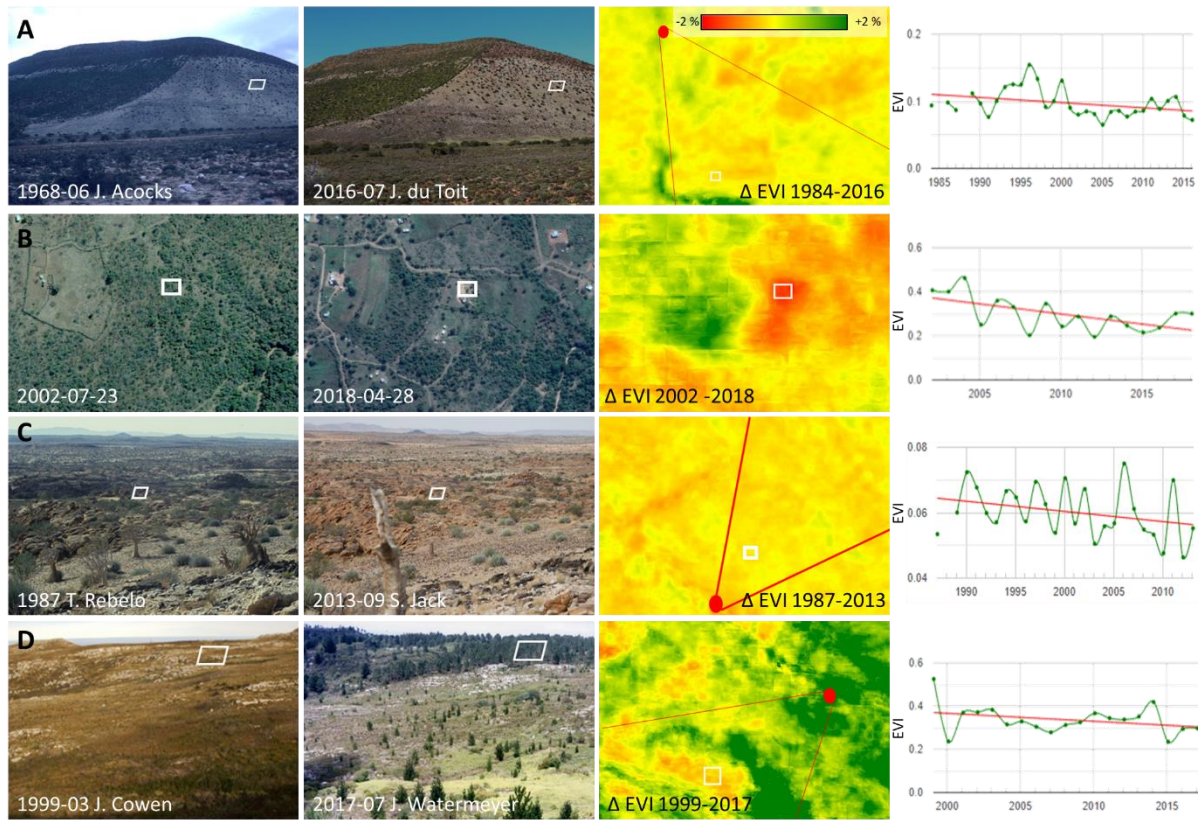
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719 **Figure 3.** Landscape-scale EVI greening trends illustrated and interpreted with repeat ground (A, C)
 720 and very-high resolution satellite (B, D) photographs. The photo location and approximate field of view
 721 for ground photographs are shown in red in A and C. The average Landsat-derived annual change in
 722 EVI are mapped for each pixel in the third column and is expressed as a percentage of the baseline
 723 EVI value. Annual timeseries are plotted for the selected pixels (white squares) in the fourth column and
 724 Sen's slope regression lines are plotted in red. See Table 1 for scenario descriptions.

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727 **Figure 4.** Landscape-scale EVI browning trends illustrated and interpreted with repeat ground (A, C,
 728 D) and very-high resolution satellite (B) photographs. See Fig. 3 caption for further details on figure
 729 layout and Table 1 for scenario descriptions.

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