

THE UNIVERSITY of EDINBURGH

Edinburgh Research Explorer

Phylogeography of the iconic Australian pink cockatoo, Lophochroa leadbeateri

Citation for published version:

Ewart, KM, Johnson, RN, Joseph, L, Ogden, R, Frankham, GJ & Lo, N 2021, 'Phylogeography of the iconic Australian pink cockatoo, Lophochroa leadbeateri', Biological journal of the linnean society. https://doi.org/10.1093/biolinnean/blaa225

Digital Object Identifier (DOI):

10.1093/biolinnean/blaa225

Link:

Link to publication record in Edinburgh Research Explorer

Document Version: Peer reviewed version

Published In: Biological journal of the linnean society

General rights

Copyright for the publications made accessible via the Edinburgh Research Explorer is retained by the author(s) and / or other copyright owners and it is a condition of accessing these publications that users recognise and abide by the legal requirements associated with these rights.

Take down policy The University of Edinburgh has made every reasonable effort to ensure that Edinburgh Research Explorer content complies with UK legislation. If you believe that the public display of this file breaches copyright please contact openaccess@ed.ac.uk providing details, and we will remove access to the work immediately and investigate your claim.



Phylogeography of the iconic Australian pink cockatoo, Lophochroa leadbeateri

Running title: Phylogeography of the pink cockatoo

Kyle M. Ewart^{1,2}, Rebecca N. Johnson^{1,2}, Leo Joseph³, Rob Ogden⁴, Greta J. Frankham^{2,5}, Nathan Lo¹

¹ The University of Sydney, School of Life and Environmental Sciences, NSW, Australia

² Australian Centre for Wildlife Genomics, Australian Museum Research Institute, NSW, Australia

³ Australian National Wildlife Collection, National Research Collections Australia, CSIRO, Canberra, Australia

⁴ Royal (Dick) School of Veterinary Studies and the Roslin Institute, University of Edinburgh, Edinburgh, UK

⁵ Centre for Forensic Science, University of Technology Sydney, PO Box 123, Broadway, NSW, 2007 Australia

1 Abstract

The pink cockatoo (Lophochroa leadbeateri; or Major Mitchell's cockatoo) is one of 2 Australia's most iconic bird species. Two subspecies based on morphology are separated by a 3 biogeographical divide, the Evrean Barrier. Testing the genetic basis for this subspecies 4 5 delineation, clarifying barriers to gene flow and identifying any cryptic genetic diversity will 6 likely have important implications for conservation and management. Here, we used genomewide SNPs and mitochondrial DNA data to conduct the first range-wide genetic assessment 7 8 of the species. The aims were to investigate the pink cockatoo's phylogeography, characterise 9 conservation units and reassess subspecies boundaries. We found consistent but weak genetic 10 structure between the two subspecies based on nuclear SNPs. However, phylogenetic analysis 11 of nuclear SNPs and mitochondrial DNA sequence data did not recover reciprocally 12 monophyletic groups, indicating that the subspecies are not evolutionarily distinct. Consequently, we have proposed that the two currently recognized subspecies be treated as 13 14 separate management units rather than evolutionarily significant units. Because poaching is suspected to be a threat to this species, we assessed the utility of our data for wildlife forensic 15 applications. We demonstrated that a subspecies identification test could be designed using as 16 17 few as twenty SNPs.

18

19 Key words

20 conservation genetics - *Lophochroa leadbeateri* - phylogeography - population genomics 21 wildlife forensics - wildlife trade

22

24 Introduction

The pink cockatoo (also known as the Major Mitchell's cockatoo), *Lophochroa leadbeateri*(Vigors, 1831), is an iconic bird species endemic to Australia. It is considered by many to be
the most beautiful and spectacular of the cockatoos (Cacatuidae; Rowley & Chapman, 1991;
Schodde, 1994), having pink-white plumage and an impressive bright red, yellow and white
crest. The pink cockatoo is a hardy species that occurs in low densities throughout Australia's
harsh arid and semi-arid regions.

31

Within the pink cockatoo's wide yet patchy distribution, four core breeding regions are 32 33 apparent (Blakers et al., 1984; Fig. 1a). Although previous authors have recognised a variable number (0-4) of subspecies (ssp; e.g. Mathews, 1912 – 3 ssp; Peters, 1937 – 4 ssp; Condon, 34 1975 - 0 ssp; Hall, 1974 - 3 ssp; Schodde, 1997 - 2 ssp), two subspecies, L. l. leadbeateri and 35 L. l. mollis (cf Forshaw & Cooper, 1981) have been generally accepted since the publication 36 37 of Schodde's 1994 study (Fig. 1a) on the basis of body size and colour and pattern of the crest. These subspecies are separated by the Eyrean Barrier (Fig. 1a): a well-documented 38 biogeographic barrier in southern Australia for a range of bird species (Ford, 1974; Schodde, 39 40 1982; Kearns et al., 2009; Dolman & Joseph, 2012). Lophochroa leadbeateri leadbeateri is 41 east of the Eyrean Barrier and has a more prominent yellow band in its crest and is larger in body size, while L. l. mollis is west and north of the Eyrean Barrier (Schodde 1994, 1997; 42 43 Forshaw & Cooper, 2002).

44

Despite its wide distribution, the pink cockatoo is of conservation concern. In the eastern part
of its distribution it is listed as Vulnerable (New South Wales and Queensland - Biodiversity
Conservation Act, 2016; Nature Conservation Wildlife Regulation, 2006), or Threatened
(Victoria - Flora and Fauna Guarantee Act, 1988; Walker *et al.*, 1999) (see Fig. 1a for State

49 localities). The species' abundance and range in north-western Victoria and western New 50 South Wales have been greatly reduced through the removal of habitat, in particular the loss of hollow-bearing trees (Garnett et al., 2011). Like other cockatoo species, the pink cockatoo 51 52 is unable to excavate its own hollows for nesting and so requires naturally occurring tree hollows (Mackowski, 1984; Cameron, 2007). Further, increased agriculture and clearing of 53 54 feeding habitat have impacted the species, particularly in the southwest of its range in the Western Australian wheat belt region (Rowley & Chapman, 1991). Another threat to this 55 species is poaching (Forshaw & Cooper, 1981; Higgins, 1999), which Rowley & Chapman 56 (1991) found to impact the most critical stage of the species' life cycle: recruitment of young. 57 Poaching is directly linked to demand for the species in the illegal pet trade. Together, these 58 59 factors indicate a need for improved understanding of phylogeographic patterns within the 60 species to aid in the conservation management of the species.

61

62 Genomic tools allow researchers to investigate how genetic diversity is distributed among 63 populations. They may help to identify and manage at-risk populations. Characterising discrete units of genetic variation, termed conservation units (Ryder, 1986), and clarifying 64 65 barriers to gene flow within the pink cockatoo will facilitate conservation strategies that maximize the evolutionary potential of the species (Frankham et al., 2010). The putative 66 67 subspecies barrier, the Evrean Barrier, comprises the Flinders Ranges and Lake Evre Basin 68 (Schodde, 1982). It is thought to have limited dispersal during the Plio-Pleistocene due to the presence of vast lakes associated with the Lake Eyre Basin, and then in the Pleistocene due to 69 70 extreme aridity (Ford & Parker, 1973; Ford, 1974; Schodde, 1982; Joseph et al., 2006). 71 However, the timing and strength with which the Eyrean Barrier has separated populations 72 within species is known to vary between avian taxa (Schodde, 1982; Dolman & Joseph, 2012; McElroy et al., 2018). Whether the morphological differences between pink cockatoo 73

subspecies at this barrier reflect underlying genetic divergence and potential conservation
units is unknown. Schodde (1994) suggested that there is currently no dispersal between
subspecies over this barrier, and that the two may even warrant recognition at species rank.
Further, it is unknown whether cryptic genetic structure exists across other well-characterised
southern Australian arid-zone biogeographic barriers within the pink cockatoo distribution,
such as the Nullarbor and Murravian Barriers (see Schodde & Mason, 1999). The impact of
these biogeographic barriers varies considerably between species (Neaves *et al.*, 2012).

82 Clarifying the species' evolutionary history and intraspecific taxonomy have been problematic due to a combination of poor sampling, relatively weak morphological 83 84 divergence across the species (e.g. see Forshaw, 2011), and the need to disentangle patterns 85 of geographical, sexual and age-related variation. Genomic analyses have the potential to 86 help to characterize conservation units, investigate connectivity among core breeding 87 populations, and resolve lingering taxonomic uncertainties about subspecies boundaries 88 (Baumsteiger et al., 2017; Marie et al., 2019; Tonzo et al., 2019; Ewart et al., 2020). 89 Furthermore, genetic data could facilitate the development of wildlife forensic tools, such as 90 geographical provenance and progeny testing, to increase the capacity for detection and 91 prosecute trafficking crimes involving this species (Walker et al., 1999; Huffman & Wallace, 92 2011). The pink cockatoo is listed under CITES Appendix II, and trade in the species is 93 strictly regulated under Australian legislation.

94

Here we perform the first comprehensive phylogeographic assessment of the pink cockatoo
to address the topics we have raised above. This builds on two earlier genetic studies
involving this species based on allozymes (Adams *et al.*, 1984), and a multilocus nuclear and
mitochondrial DNA (mtDNA) data set (White *et al.*, 2011); both used only a few individuals

99 to address the species' systematic position with respect to other cockatoos. Pink cockatoo 100 specimens from across the species' range have been collected over many decades and are 101 stored in museums throughout Australia and elsewhere. Owing to developments in museum 102 genomics, genome-wide data of use in population-level studies can be generated from old museum specimens (Rowe et al., 2011; Ewart et al., 2019). We generated thousands of 103 104 genome-wide single nucleotide polymorphism markers (SNPs) and sequence data at three mtDNA markers from pink cockatoo frozen tissue and toe pad samples across their entire 105 106 distribution. We performed comprehensive population genomic analyses to investigate 107 potential barriers to gene flow for the purposes of clarifying taxonomy and informing conservation management. These data can be interpreted in light of the biogeography and 108 109 palaeoenvironmental history of Australia's arid and semi-arid zones, and compared to the 110 steadily increasing body of phylogeographic analyses of species having broadly similar distributions across southern Australia (Neaves et al., 2009; Dolman & Joseph, 2012, 2015; 111 112 Engelhard et al., 2015; Ansari et al., 2019).

113

114 Methods

115 Sample acquisition and DNA extractions

116 We acquired pink cockatoo frozen tissue (frozen liver/muscle) (n=45) and toe pad (n=51)

samples from across their distribution (Fig. 1a & Table S1). Samples were obtained from: the

118 Australian National Wildlife Collection, Canberra (ANWC); the Australian Museum, Sydney

119 (AM); Museum Victoria, Melbourne (MV); and the Western Australian Museum, Perth

120 (WAM). Collection dates for these samples ranged from 1883 to 2011 (Table S1).

Thinly sliced toe pads (~2 mm thick) were sampled from traditional museum specimens, and
DNA was extracted following Ewart *et al.* (2019). These DNA extractions were performed in
a clean room facility dedicated to historical museum samples likely to have degraded DNA.
Genomic DNA was extracted from frozen tissue samples following the manufacturer's
protocols for the 'Bioline Isolate II Genomic DNA kit' Bioline (Australia). DNA
concentration was measured using a Qubit 2.0 Fluorometer (Thermo Fisher Scientific).

128

129 *SNP genotyping*

130 SNP data was generated using DArTseq[™], a reduced representation sequencing method 131 (methods described in Kilian et al., 2012; Cruz et al., 2013). This was performed by Diversity Arrays Technology (DArT) in Canberra, Australia. DArTseq has previously been 132 used to generate SNP data for a range of phylogeographic, phylogenetic, and population 133 genetic studies on vertebrate species (Melville et al., 2017). Briefly, different combinations 134 135 of restriction enzymes were tested, and the PstI-SphI enzymes were selected for digestion of cockatoo DNA. DNA was then processed as per Kilian et al. (2012), using two different 136 137 adaptors that correspond with the restriction site overhangs, both containing an Illumina flow 138 cell attachment sequence, and one (the PstI-compatible adapter) also containing a sequencing 139 primer sequence and varying length barcode region. The library was subject to PCR (using REDTaq DNA Polymerase, Sigma-Aldrich) as follows: initial denaturation at 94°C for 1 140 141 min, then 30 cycles of 94°C for 20 sec, 58°C for 30 sec and 72°C for 45 sec, and a final extension step at 72°C for 7 min. The library was then normalized and sequenced by first 142 143 performing a c-Bot (Illumina) bridge PCR, followed by single end sequencing for 77 cycles on an Illumina Hiseq2500. 144

145

146 The resultant short-read sequences were processed using the DArT Pty Ltd analytical 147 pipelines. First, poor quality sequences were removed (using a Phred score ≥ 10), and 148 sequences were demultiplexed (using a barcode Phred score \geq 30). Second, sequences were 149 trimmed to 69 bp and clustered with a Hamming distance threshold of 3. Low-quality regions from singleton tags were corrected where possible. Third, SNPs were called using the 150 151 proprietary DArTsoft14 SNP calling pipeline. Real alleles were discriminated from paralogous sequences by assessing a range of parameters including sequence depth, allele 152 153 count and call rate.

154

155 SNP filtering

We applied numerous SNP filtering criteria depending on the analysis (following Ewart et 156 157 al., 2019). First, we removed the duplicate/triplicate samples with the highest amount of 158 missing data. Second, we removed potentially erroneous SNPs, and SNPs with a high level of 159 missing data, based on reproducibility (100%) and call rate (>80%), using the R package dartR version 1.0.5 (Gruber et al., 2018). Third, to meet the population genetic assumptions 160 161 of some analyses, we removed linked SNPs, outlier SNPs that potentially represented loci 162 under selection, and SNPs out of Hardy-Weinberg equilibrium (HWE). To remove linked 163 SNPs to meet the assumption of linkage disequilibrium for some of the analyses, we retained 164 only one SNP per DArTseq locus using the R package dartR. To identify and remove outlier 165 SNPs that are potentially under directional or balancing selection to meet the assumption of neutrality for some analyses, we used LOSITAN (Beaumont & Nichols, 1996; Antao et al., 166 167 2008). For this analysis, samples were divided into subspecies, then performed 100,000 simulations, applying the 'infinite alleles' mutation model, a 0.95 confidence interval and a 168 169 0.1 false discovery rate. To identify departure from HWE, we used ARLEQUIN version 3.5 170 (Excoffier & Lischer, 2010), implementing 1,000,000 Markov Chain steps and a burnin of

171 100,000. We removed loci with a p-value <0.01 that potentially deviate from HWE. For this
172 analysis, we considered all samples as one population, which is likely a conservative
173 approach, as we would expect some false positives due to the Wahlund effect.
174

To investigate whether remnant poor-quality SNPs were skewing results, additional filters
were applied to represent a 'stringently filtered' data set, and analyses were repeated. Here,
we filtered SNPs for average locus coverage (>20 X) using the R package dartR, and minor
allele frequency (MAF) (>0.05) using the R package poppr version 2.6.1 (Kamvar *et al.*,
2014, 2015).

180

Additionally, to ensure that the inclusion of toe pad samples from old museum specimens did
not skew results, SNPs were re-called using only the more contemporary tissue samples
(using the SNP calling methods outlined in the previous section). SNPs were subsequently refiltered. Additional details on SNP filtering methods and variants are provided in the
Supplementary Material (Appendix I).

186

187 SNP quality control

To quantify genotyping error, we included 18 replicate and 4 triplicate samples among the 96 pink cockatoo samples analysed (indicated in Table S1). We used various replicate/triplicate types to investigate the factors that may influence error, including: frozen tissue replicates (from the same and different DArTseq plates), toe pad replicates (from the same and different DArTseq plates), frozen tissue / toe pad replicates (i.e. a frozen tissue and toe pad from the same individual), and tissue DNA replicates (from the same and different DArTseq plates).

We calculated SNP error rates (i.e. the number of SNP mismatches between replicate pairs
over the total number of SNPs that were not missing in both replicates) using R functions
from Mastretta-Yanes *et al.* (2015). Error rates were calculated pre- and post- SNP-filtering.

199 Generation of mitochondrial DNA sequence data

200 To generate mitochondrial reference genomes, we performed low-coverage whole genomic 201 sequencing for four pink cockatoo samples (indicated in Table S1), following the NEBNext 202 DNA library preparation protocol, with a pre-treatment of 500 bp shearing using Covaris M220. The libraries were then sequenced on an Illumina MiSeq using paired-end 251 bp 203 204 sequencing. Library preparation and sequencing were performed at the Monash University 205 Malaysia Genomics Facility (Selangor, Malaysia). The resultant paired sequence reads were 206 trimmed using the BBDuk plugin in Geneious version 10.2.4 (Kearse et al., 2012), then 207 assembled using Geneious and NOVOPlasty (Dierckxsens et al., 2017). We then designed 208 primers for the ND4 and ND5 genes and d-loop (for ND2, we used primers from Sorenson, 209 2003), and amplified and sequenced 15 samples from across the pink cockatoo range 210 (indicated in Table S1). Thus, the mtDNA analyses were carried out using 19 samples (4 211 using low-coverage whole genomic sequencing, and 15 using Sanger sequencing). The dloop marker was subsequently excluded as it was unable to be reliably sequenced (possibly 212 213 due to the presence of control region duplications, which are often found in parrot species; 214 Schirtzinger et al., 2012; Eberhard & Wright, 2016). Additional details on mitochondrial 215 genome assemblies, primers, PCR conditions and sequencing can be found in the 216 Supplementary Material (Appendix II). 217

218

220 Identifying population structure

221 We used five methods to investigate population structure present in the SNP genotype data. Details of the different SNP filtering strategies and samples used in the different analyses are 222 223 provided in the Supplementary Material (Appendix I; Table S1). First, genetic variation was summarized and visualized using a principal coordinates analysis (PCoA). This was 224 225 performed using the R packages dartR and ade4 version 1.7 (Chessel et al., 2004). Second, 226 STRUCTURE version 2.3 (Pritchard et al., 2000) was used to investigate genetic structure and admixture. For this analysis, we modelled up to five ancestral populations (K=1-5), 227 228 implementing 10 replicates for each K, assuming admixture and correlated allele frequencies (Porras-Hurtado et al., 2013). We ran the analysis for 2 million iterations with a burn-in of 1 229 million. This analysis was parallelized and automated using StrAuto version 1.0 (Chhatre & 230 231 Emerson, 2017). We considered six different estimators to determine the optimal value of K, 232 generated using StructureSelector (Li & Liu, 2018). Replicate runs were merged and bar plots 233 were generated using CLUMPAK (Kopelman et al., 2015), implemented through 234 StructureSelector. We took a hierarchical approach, whereby the population clusters 235 identified using the full dataset were separated, re-filtered, then run independently. Third, to 236 investigate whether patterns of genetic differentiation derived from continuous (i.e. isolation 237 by distance; IBD) or discrete (e.g. biogeographic barriers) phylogeographic processes, we performed a conStruct analysis (Bradburd et al., 2018), implementing the spatial model. A 238 239 conStruct (i.e. 'continuous structure') analysis is similar to the STRUCTURE analysis, but 240 controls for geographic distance between samples. Based on initial optimization, we ran two 241 independent conStruct analyses, with the 'adapt delta' parameter (the target average proposal acceptance probability) set at 0.85, implementing two chains with 100,000 MCMC iterations 242 243 for each run. We checked for consistency between chains and independent runs, and visually checked for convergence using the trace plots generated by conStruct. To determine an 244

appropriate level of parameterization, we ran five replicates of a cross-validation analysis comparing the spatial and non-spatial models for K = 1-5 for each replicate. We used a random 90% subsample as the training partition, and ran the analysis for 10,000 MCMC iterations.

249

Fourth, to measure genetic divergence between subspecies, we calculated pairwise F_{ST} values 250 251 (Weir & Cockerham, 1984) using the R package hierfstat version 0.4.22 (Goudet & Jombart, 2015). F_{ST} values were considered significant if their associated confidence intervals (based 252 253 on 0.025 and 0.975 quantiles, implementing 1000 bootstraps) did not encompass 0. To 254 investigate differentiation within and between subspecies we performed an AMOVA using 255 the R package poppr, and checked for significance using 10,000 permutations implemented 256 in the R package ade4. To investigate whether any genetic structure patterns were driven by closely related individuals (e.g. cousins), we performed an inter-individual kinship analysis 257 258 using the R package SNPRelate version 1.14 (Zheng et al., 2012). 259 260 We performed a haplotype network analysis to investigate population structure within the 261 mtDNA sequence dataset. We performed this analysis using PopART (Leigh & Bryant, 2015), based on concatenated ND2, ND4 and ND5 sequences (a total of 2037 bp) and 19 262 263 samples, implementing the statistical parsimony TCS method (Clement et al., 2000). 264 Additionally, we calculated net nucleotide divergence (Da) between the two subspecies based 265 on the mtDNA sequence dataset using the R package strataG version 2.4.905 (Archer et al., 2017). 266

267

268

270 *Gene flow patterns*

271 To investigate the influence of geographic distance in our genetic structure results, we investigated the correlation between genetic and geographic distance (i.e. IBD). As there are 272 273 no discrete sampling sites (reflecting the pink cockatoo's continuous distribution; Fig. 1a), we analysed inter-individual distances. Individual-based genetic distances were based on 274 275 PCA-based Euclidean distance, following Shirk et al. (2017), calculated using 45 principal 276 components (35 when using only fresh tissue samples), and performed using the R package adegenet version 2.1.0 (Jombart, 2008). We then performed a Mantel test using these 277 278 Euclidean genetic distances and geographic distance (in kilometres) using the R packages 279 adegenet and dartR.

280

281 Due to the ongoing debate surrounding the use of Mantel tests to infer IBD patterns (e.g. 282 Diniz-Filho et al., 2013), especially when considering inter-individual distances, we analysed 283 interpopulation gene flow along a transect following methods in Ogden & Thorpe (2002). Indirect gene flow inferences were based on pairwise F_{ST} measurements (calculated as above, 284 but scaled by pairwise geographic distance) between five 'sample clusters' (three individuals 285 per cluster) across Australia, focusing on the putative subspecies barrier (Fig. 3b; Table S1). 286 Willing *et al.* (2012) demonstrated that F_{ST} values can be estimated with relatively small 287 288 sample sizes when using thousands of SNPs. To complement this analysis of gene flow 289 across the Eyrean Barrier, we ran a conStruct analysis using the same 15 samples in the 290 transect above. We used the same settings as the previous conStruct analysis, except the 291 'adapt delta' parameter was set to 0.7.

292

293

295 *Genetic diversity*

296 To measure the genetic diversity within each subspecies, we calculated allelic richness,

297 heterozygosity and private allele counts for each SNP marker. Allelic richness was calculated

using the R package PopGenReport version 3.0.4 (Adamack & Gruber, 2014), implementing

299 rarefaction to account for differences in sample size. Observed and expected heterozygosity

300 were calculated using GenAlEx (Peakall & Smouse, 2006, 2012). A count of private alleles

301 per population was calculated using the R package poppr. Mitochondrial DNA diversity was

302 measured in terms of nucleotide diversity, proportion of polymorphic sites, and number of

- haplotypes using Geneious and the R package pegas version 0.1 (Paradis, 2010).
- 304

298

305 *Population growth*

To investigate factors that may have caused discordant mtDNA and nuclear DNA clustering
patterns (see the '*Results*' section) and to test for population growth, we computed Tajima's

308 D (Tajima, 1989), Fu's Fs (Fu, 1997) and Ramos-Onsin's R2 (Ramos-Onsins & Rozas, 2002)

309 statistics using DnaSP 6.12.03 (Rozas *et al.*, 2017), based on mtDNA sequence data (2037 bp

of concatenated *ND2*, *ND4* and *ND5* sequences). The significance of the statistics was

311 inferred using coalescent simulations with 1000 replicates. Additionally, a mismatch

312 distribution plot was generated using the R package pegas.

313

314 *Phylogenetic methods*

315 We performed phylogenetic analyses to investigate whether genetic units identified in the

- 316 population genetic analyses were evolutionarily distinct within a phylogenetic framework.
- 317 Phylogenetic analyses based on SNPs were performed using SNAPP (Bryant *et al.*, 2012),
- 318 implemented in BEAST version 2.4 (Bouckaert et al., 2014), to compare 'species'
- 319 hypotheses (ESU hypotheses in this case). We used SNAPP to compare the relative support

320 for two models: one enforcing monophyly of each of the two subspecies (which corresponds 321 to two genetic units in population genetic analyses; see Results section), and one without 322 enforcing monophyly. As SNAPP is computationally intensive, we included four individuals 323 per subspecies and 1000 randomly selected SNPs with no missing data from the putatively neutral SNP data set (see Supplementary Material, Appendix I, for more details) to improve 324 325 computational tractability. We ran SNAPP for 4 million Markov chain Monte Carlo (MCMC) steps, sampling every 1000 steps after a burn-in of 400,000 steps. We used allele frequencies 326 327 for the forward and backward mutation rates, and the default settings for priors. Model 328 support was subsequently estimated using the AICM (Akaike information criterion through MCMC) method in Tracer version 1.6 (Rambaut et al., 2014). AICM was chosen over the 329 330 preferred stepping-stone and path sampling analyses to improve computational tractability. 331 As AICM has been shown to suffer from poor repeatability (Baele et al., 2012), we ran three replicate SNAPP analyses for each model (i.e. three enforcing monophyly of subspecies, and 332 333 three not enforcing monophyly) and subsequently estimated AICM for each of the six runs. 334

To complement the SNAPP analysis, we performed a maximum likelihood phylogenetic analysis using RAxML (Stamatakis, 2014) based on concatenated SNP data (see Supplementary Material, Appendix I). We implemented the GTR substitution model with gamma-distributed rates among sites and the Lewis-type ascertainment bias correction to account for the exclusion of invariant sites, and performed 1000 bootstrap replicates to estimate node support. Trees were rooted using the midpoint method and visualised using Figtree 1.4.2 (Rambaut, 2009).

342

We performed a Bayesian phylogenetic analysis of mtDNA data (2,037 bp of concatenated *ND2*, *ND4* and *ND5*) using MrBayes 3.2 (Ronquist *et al.*, 2012). This analysis was performed

using four independent Markov chains, each run for 100 million steps with a 25% burn-in,
and sampled every 100 steps with convergence diagnostics calculated every 100 steps. We
implemented the HKY substitution model with gamma-distributed rates among sites.
Convergence diagnostics were assessed using Tracer (ESS values <200 were considered
inadequate). This analysis was performed with and without an outgroup (*Cacatua pastinator*;
GenBank accession: JF414240). Trees were rooted using either the midpoint method or an
outgroup, and visualized using Figtree.

352

353 Testing SNPs for wildlife forensic applications

354 We filtered a subset of SNPs based on their utility in a geographic provenance assignment test by investigating SNP contributions in a discriminant analysis of principal components 355 356 (DAPC). DAPC minimizes variation within groups, and maximizes variation between 357 groups. First, we performed DAPC on the entire SNP dataset with no missing data (see extra 358 filtering details in Supplementary data, Appendix I) using the R package adegenet. We considered two populations (K=2), corresponding to separation of the two subspecies, then 359 360 repeated the analysis considering three populations (K=3) to investigate whether more fine-361 scale geographic assignment was possible. Second, SNPs were ranked based on their 362 contribution to the clustering analysis. Third, we iterated through decreasing numbers of SNPs (increments of five SNPs) to investigate the minimum number of SNPs required to 363 364 separate the two subspecies clusters. Finally, we tested the utility of a refined set of SNPs for geographic/subspecies assignment by assigning six randomly selected individuals (three 365 366 individuals per subspecies) in separate tests using GeneClass2 (Piry et al., 2004). For this analysis, we implemented the frequency-based assignment method (Paetkau et al., 1995) and 367 368 a 0.05 assignment threshold. The individual being tested was removed from the 'reference'

369 data before each analysis. Likelihood ratios were calculated from the assignment likelihood370 results, considering different prosecution and defence hypotheses.

371

372 **Results**

373 *SNP genotyping*

374 Seventy-eight samples were successfully genotyped using DArTseq (Table S1). DNA

extracts from one frozen tissue sample (out of 45) and 20 toe pad samples (out of 51) were

376 unsuitable for successful DArTseq library preparation. The oldest sample successfully

377 genotyped was collected in 1912; all samples collected before this date failed. The

378 DArTsoft14 pipeline called 20,324 SNPs from the successfully genotyped 78 samples (with

379 36.32% missing data). This SNP data set was reduced to 4,135 SNPs (with 12.26% missing

data) after filtering for quality and missing data, 2,131 SNPs (with 11.78% missing data) after

381 filtering for neutrality and linkage, and 1,279 SNPs (with 10.35% missing data) after

applying more stringent filtering (see Supplementary Material, Appendix I for data filtering

details, and Table S1 to view which individuals were used in each analyses). When using

only the more contemporary tissue samples for SNP calling, the DArTsoft14 pipeline called

385 16,472 SNPs (with 16.79% missing data), which was reduced to 6,466 SNPs (with 3.07%

missing data) after filtering for quality and missing data, and to 4,891 SNPs (with 1.95%

387 missing data) after filtering for neutrality and linkage.

388

389 *SNP quality control*

390 Of the 18 replicate and 4 triplicate samples examined, some failed. We found two additional 391 replicate samples based on their genetic signature (i.e. they had different sample numbers and 392 were held in different Museums but they were parts from the same individual in two

collections). This was subsequently confirmed with the relevant Museums. Overall, a total of
13 replicates and 4 triplicates were used to quantify genotyping error (Table S2).

395

396 Filtering reduced the allele error rate in all samples except one (ANWC B38557; this sample also had a very high proportion of missing data) (Table S1). After filtering, SNP error rates 397 398 for frozen tissue and DNA replicates/triplicates were all <3%. The SNP error rate and/or shared missing data (missing in both replicates) was particularly high in eight 'toe pad/toe 399 400 pad' and 'tissue/toe pad' replicates (ranging from 12.10-23.08% and 0.63-97.17% 401 respectively after filtering). Although several problematic samples were removed from many of the population genetic analyses (see Supplementary Material, Appendix I), error in toe pad 402 403 samples was variable, ranging from 2.87-23.08% in 'toe pad/toe pad' replicates after 404 filtering, hence toe pad samples with relatively high error rates are likely present in some analyses. 405

406

407 *Genetic structure*

408 The PCoA revealed three distinct clusters: one L. l. mollis cluster and two L. l. leadbeateri 409 clusters (Fig. 1b). Within L. l. leadbeateri, five individuals from central Queensland formed a 410 cluster that was distant from the other samples. Kinship between these individuals was relatively high (0.045-0.144; Table S3) compared to the average kinship of the entire dataset 411 412 (0.008; excluding self-kinship values), which may distort the level of genetic structure in this region. When removing four out of the five central Queensland samples in a PCoA, the 413 414 remaining sample clusters with the other L. l. leadbeateri individuals (this result is consistent when different central Queensland individuals are used; Figure S1). The only other 415 416 Queensland individual in the data set, from southern Queensland (see Fig. 1a), clustered with 417 the other L. l. leadbeateri samples. There were five other outlier samples. The four outliers

near the origin of the PCoA plot (Fig. 1b) are likely explained by their high level of missing
data (>70%) (missing data are replaced by the mean allele frequency in the PCoA analysis).
The origin of the outlier from the Northern Territory (MV Z50083) is unclear. It may have
been either a migrant, an escaped aviary bird from the *L. l. leadbeateri* range, or the result of
a processing error (e.g. mislabelling, DNA contamination etc.).

423

438

439

424 Genetic variability in the STRUCTURE analysis was best explained using *K*=2-5, depending on 425 the estimator considered (Fig. S2). We present the major modes generated by CLUMPAK for 426 K=2 and K=3 (Figs 1c, d). The STRUCTURE analysis revealed a clear genetic break between the two subspecies, with the exception of one outlier sample from the Northern Territory 427 428 (identified in the PCoA; Fig. 1). Individuals from central Queensland were distinct when 429 using K=3 (Fig. 1d) and in the analysis based on L. l. leadbeateri samples only (Fig. S3a). 430 Similar to the case for the PCoA, this result is likely driven by the relatively high relatedness 431 between these central Queensland individuals. In the STRUCTURE analysis based on L. l. 432 *mollis* samples only, subtle population differentiation, although not robustly supported, coincided with samples from the south-western wheatbelt region (Fig. S3b-c). 433 434 Genetic variability in the conStruct analysis was best explained using K=2-3 (Fig. S4). Some 435 436 isolation by distance is evident as the spatial model is preferred over the non-spatial model. 437 In the conStruct analysis using K=2, there is clear population differentiation between the two

subspecies (excepting the Northern Territory outlier sample identified above; this sample was

removed from subsequent analyses; Fig. S5a-b), corroborating the STRUCTURE analysis (Fig.

440 1c-d). There was slight variability in the admixture plots between different chains and

441 independent analyses, however the main patterns were consistent (we present one chain from

442 each independent analysis; Fig. S5a-b). Inadequate convergence and consistency between

443 chains/analyses when using K=3 indicated that the results were unreliable at this level of 444 parameterization.

445

446	Relatively low but significant genetic differentiation was evident between the two subspecies
447	($F_{ST} = 0.039$; confidence interval: 0.035, 0.042). In the AMOVA based on the full dataset
448	(i.e. 56 individuals and 2131 SNPs), the proportion of genetic variation within individuals
449	was 69.8%. This is significantly lower than expected based on random permutations (p $<$
450	0.001). The proportion of genetic variation within and between subspecies (25.8% and 4.4% ,
451	respectively) were, however, both greater than expected ($p < 0.001$) (Table S4 & Fig. S6).
452	These patterns are indicative of population structure, and not a single panmictic population.
453	In the PCoA, STRUCTURE, F_{ST} and AMOVA analyses, use of different SNP datasets (i.e.
454	SNPs based on only tissues, and SNPs that underwent more stringent filtering) exhibited very
455	similar results (Figs S6, S7, S8 & Tables S4, S5).
456	
457	Ten haplotypes were observed from the 19 mtDNA samples that were sequenced (i.e. 2,037
458	bp of concatenated ND2, ND4 and ND5 genes; Table S7). The haplotype network analysis
459	based on mtDNA exhibited a star-like pattern (Fig. 2a). A central haplotype predominated,

460 while other haplotypes were connected by the common haplotype. The common central

461 haplotype comprises individuals from both subspecies from across the species range. The

462 mtDNA *Da* between subspecies was 0.004%. Overall, mtDNA structure did not reflect

463 patterns found in SNP clustering analyses.

464

465 *Gene flow patterns*

466 The inter-individual Mantel tests revealed significant IBD when analysing the full dataset and 467 when analysing only more contemporary frozen tissue samples (all p < 0.001) (Fig. S9).

However, inter-individual genetic distances were found to be relatively invariable (note the
near-horizontal relationship between genetic and physical distance in Fig. S9a). Relatively
low genetic distances across Australia indicate that differentiation among geographic
locations is weak. Further, in some cases, spatial patterns inferred from Mantel tests are
problematic (Legendre & Fortin, 2010; Legendre *et al.*, 2015). We did not consider mtDNA
in this analysis, as mtDNA is known to produce unreliable IBD results (Teske *et al.*, 2018).

There was a reduction in gene flow between the 'sample clusters' spanning the putative 475 476 subspecies along the transect (Fig. 3). Although the level of differentiation was relatively low, all pairwise F_{ST} estimates along the transect were significant except for one (between 477 478 'cluster 1' and '2'; see Fig. 3b). The conStruct analysis based on these 15 transect samples 479 corroborated the other population structure analyses. Clear genetic differentiation was evident between the two subspecies (Fig. S5c-d). Although there was slight variability between the 480 481 independent analysis and separate chains, the main population structure patterns were 482 consistent.

483

484 *Genetic diversity*

Lophochroa leadbeateri mollis had the highest genetic diversity for all metrics, although not
considerably higher than *L. l. leadbeateri* (Table 1). Genetic diversity measurements varied
when using different SNP datasets, but were qualitatively consistent (Table S6). As expected,
when applying more stringent filtering (including a MAF filter), the number of private alleles
and allelic richness decreased. Without subspecies divisions, mtDNA nucleotide diversity
was 0.0012 (Table S7); *ND2* was considerably more diverse than *ND4* and *ND5*.

491

493 *Population growth*

494 Analyses of 'randomness', 'neutrality', Tajima's D (-1.851), Fu's Fs (-4.865) and Ramos-495 Onsin's R2 (0.052), were all significant (p < 0.05 in each case). The unimodal mismatch 496 distribution (with a high value at zero mismatches) of the mtDNA data also indicates the 497 occurrence of an expansion event (Figure S10; Rogers & Harpending, 1992). These results 498 are consistent with a scenario of rapid growth in population size.

499

500 *Phylogenetics*

501 The SNAPP model for which monophyly was not enforced received the highest support

502 (Table S8). AICM was relatively consistent between replicates, ranging from 16838.7 to

503 16846.6 for model one (monophyly not enforced), and from 16874.8 to 16879.1 for model 2

504 (monophyly enforced). The two subspecies each exhibited monophyly in the RAxML

analysis (excepting the one aforementioned outlier sample from Northern Territory), although

506 bootstrap support was relatively low (i.e. 73%; Figure S11). These results indicate that the

507 existence of two ESUs corresponding each of the two subspecies are not unambiguously

supported.

509

510 Similar to the haplotype network analysis, phylogenetic analysis of mtDNA did not

511 correspond to the SNP population structure results and did not exhibit any discernible

512 geographic patterns (Fig. 2b, S12).

513

514 *Wildlife forensics*

The initial DAPC used for SNP selection clearly separated the two subspecies (Fig. 4a), in
line with the other genetic structure analyses. We retained 35 principal components for this

analysis. The minimum number of SNPs required to separate the subspecies via DAPC was twenty (Fig. 4b). We considered adequate separation when all samples were correctly sorted into their corresponding subspecies clusters. We retained five principal components when performing the DAPC using twenty SNPs. When considering three populations (K=3), the central Queensland individuals formed a separate cluster having no overlap but only when ≥ 75 SNPs were utilized (Fig. S13). It should be noted, however, that this clustering is likely driven by the high relatedness between these central Queensland samples.

524

525 The GeneClass2 analyses correctly assigned all six individuals with high support. When

signing an individual to the correct subspecies (e.g. claiming a *L. l. leadbeateri* individual

527 was *L. l. leadbeateri*), all likelihood ratios were >28.71, and averaged 1.81×10^7 (Table S9).

528 The likelihood ratios were higher when assigning *L*. *l*. *mollis* than when assigning *L*. *l*.

leadbeateri, averaging 3.54×10^8 and 8.23×10^6 respectively. When assigning an individual to

the incorrect subspecies (e.g. claiming an *L. l. leadbeateri* individual was an *L. l. mollis*

individual), all likelihood ratios were $<3.48 \times 10^{-2}$, and averaged 5.81×10^{-3} .

532

533 **Discussion**

We have performed the first comprehensive phylogeographic study of one of Australia's most charismatic but relatively understudied parrots, the pink cockatoo. Our extensive data set revealed two major genetic clusters corresponding to the currently recognized subspecies, and an additional, divergent cluster comprising closely related Central Queensland members of *L. l. leadbeateri* (importantly, this cluster disappeared when only one representative was used). We use these results to reassess the species' conservation priorities and taxonomy,

540 which are currently based on morphology.

542 *Population structure*

Lophochroa leadbeateri is a widespread species that does not have defined geographically 543 disjunct population isolates. Our SNP data show consistent but relatively weak levels of 544 545 genetic structure between the two currently recognized subspecies at the Eyrean Barrier. It is important to determine whether this result is derived from historical biogeography (i.e. the 546 547 Eyrean Barrier) or sampling gaps (i.e. IBD) as has been highlighted by several authors (Latch 548 et al., 2014; Bradburd et al., 2018; Chambers & Hillis, 2020). We found that genetic structure 549 between the two subspecies based on SNPs was apparent even when accounting for 550 geographic distance (Figs. 3 & S5). Contrastingly, distinct subspecies clusters were not 551 apparent in the mtDNA analyses. This is possibly due to incomplete lineage sorting and/or higher female dispersal, and is consistent with the weak and/or recent phylogeographic 552 553 structure across the continent inferred by the SNP analyses. Large effective population sizes 554 retaining ancestral variation even after long periods of isolation and/or recent divergence times could potentially preclude signals of population divergence in mtDNA (Hartl et al., 555 1997; Maddison, 1997). 556

557

558 The significant population expansion result, further evidenced by the star-like haplotype 559 network (Fig. 2a), may have proliferated the frequency of a common haplotype and explain 560 the absence of distinct geographically disjunct haplotype clusters. The common haplotype 561 (see Fig. 2a) comprised individuals from across the species' range, including an individual 562 from central Queensland (B28102) and individuals from south-west Western Australia 563 (A35378, Z23813 and B53847), indicating that the species has the capacity to disperse over long-distances. However, the weak differentiation detected by SNPs indicates that the Eyrean 564 565 barrier may have limited dispersal, similar to other vertebrate species found in this region 566 (Neaves et al., 2012; McElroy et al., 2018).

567

Overall, these data suggest that the Eyrean barrier has been either a somewhat effective,
although relatively recent biogeographic barrier to gene flow in this species, or a more longterm but porous barrier. The subtle morphological divergence between subspecies reported
by Schodde (1994) is consistent with a relatively recent divergence time. Morphological
differences can accumulate rapidly in bird taxa, often before mtDNA genetic divergence
(Zink & Barrowclough, 2008; Safran *et al.*, 2016).

574

The weak substructure evident within each of the two subspecies is consistent with relatively 575 regular gene flow between members of the four core breeding populations (Fig. 1a). In L. l. 576 577 leadbeateri, the genetic differentiation we identified between individuals from central 578 Queensland individuals and all other populations is likely an artefact of analysing related 579 individuals. Although the relatively high relatedness between these individuals may be due to 580 actual genetic structure in this region (i.e. higher levels of inbreeding in a genetically isolated 581 population), it is more likely that individuals from a family unit were sampled. All five central Queensland individuals were collected in the same region, four of which were 582 583 collected three days apart (while the other was collected ~3 years later), and the kinship 584 analysis suggests these individuals could be second- and/or third-order relatives (Table S3). 585 In L. l. mollis, there is limited genetic differentiation between the population in the south-586 westernmost 'wheatbelt' area and other populations (Fig. S3b). This population inhabits 587 mulga shrubland, and was previously considered a separate subspecies (Peters, 1937). 588 However, the genetic structure in this region is subtle and inconsistent; notably, some of the associated samples do have high levels of missing data. Analysing additional geographically 589 intermediate samples may help clarify the presence of potential cryptic genetic diversity 590

within the two subspecies, and hence elucidate management strategies to conserve theirgenetic variation.

593

594 The pink cockatoo's shallow phylogeographic structure across its range corresponds to that seen in some other Australian arid zone bird species (Joseph & Wilke, 2006; Dolman & 595 596 Joseph, 2015). Engelhard et al. (2015), for example, found mtDNA genetic structure, albeit weak, in another cockatoo in the same subfamily (Cacatuinae), the galah (Eolophus 597 598 roseicapilla). However, there are numerous examples of similarly distributed bird species 599 that do exhibit more marked genetic differentiation across much the same range, such as the copper-backed and chestnut quail-thrush (Cinclosoma clarum and C. castanotum, 600 601 respectively), the white-eared honeyeater (Nesoptilotis leucotis), the splendid fairy-wren 602 (Malurus splendens), and the Australian ringneck (Barnardius zonarius) (Joseph & Wilke, 603 2006; Kearns et al., 2009; Dolman & Joseph, 2015, 2016). We recently found evolutionarily 604 distinct isolates within arid zone populations of another inland cockatoo species, the red-605 tailed black-cockatoo Calyptorhynchus banksii. In that case the southwestern 'wheatbelt' population was found to be genetically and taxonomically distinct (Ewart et al., 2020). 606 607 Varying responses to biogeographical barriers among the pink cockatoo and these other arid 608 bird taxa are likely due to differences in habitat specificity and vagility (Toon et al., 2007). 609

610 *Conservation implications*

Robust delineation of conservation units is vital for effective conservation prioritization.
Conservation units can be apportioned as either management units (i.e. a demographically
independent unit of genetic variation; Moritz, 1994; Palsbøll *et al.*, 2007) or evolutionarily
significant units (i.e. independently evolving units of genetic variation; Ryder, 1986; Moritz,

615 1994). Based on the genetic structure results presented above, the two subspecies should be

616 considered separate management units. Given the lack of support for two evolutionarily 617 distinct clades (i.e. they do not exhibit reciprocal monophyly) in the phylogenetic analysis 618 based on nuclear SNPs, the low F_{ST} values, and the lack of mtDNA support, these 619 conservation units do not appear to constitute separate ESUs.

620

621 Assessing population fragmentation within each of the two subspecies is critical, as small isolated populations often suffer from genetic erosion (Frankham et al., 2017). The additional 622 623 substructure we identified in central Queensland could indicate that this population is at risk 624 of genetic isolation, although it is likely that the genetic differentiation detected in this region 625 is likely driven by high relatedness among the samples examined (see above). Denser 626 sampling of unrelated individuals, and geographically wider sampling to fill gaps in this 627 study should be implemented to clarify the genetic structure in this region and determine 628 whether or not it should be regarded as separate management unit.

629

630 Taxonomic reassessment

631 Incorrect delineation of subspecies can misguide subsequent studies and conservation 632 strategies (Zink, 2004; Braby et al., 2012; Huang & Knowles, 2016). Typically, different 633 subspecies exhibit at least some mtDNA phylogenetic resolution (e.g. Kearns et al., 2015, 2016). Net divergence, Da, at the mtDNA ND2 gene between the two nominal pink cockatoo 634 635 subspecies was only 0.009%. In several other avian species that exhibit ND2 differentiation 636 at the Eyrean Barrier, the value is much higher. Examples include the white-eared honeyeater 637 (2.23%; Dolman & Joseph, 2015), the mulga parrot subspecies (1.92%; McElroy et al., 2018), and the Australian ringneck (1.72%; Joseph & Wilke, 2006). Accordingly, the 638 639 minimal mtDNA differentiation may be taken to suggest that the species is monotypic (i.e., 640 no subspecies). Conversely, a lack of mtDNA-based subspecies divergence does not

641 necessarily justify/dictate taxonomic modifications (Ball & Avise, 1992; Funk & Omland,

642 2003; Omland *et al.*, 2006). Traits other than genetics and morphology, including

643 vocalizations, ecological characteristics, and frequency of subspecies hybrids, can be taken

644 into account (Remsen, 2005; also see Ford & Parker, 1973). Therefore, although they may

not be evolutionary distinct genetically (i.e. they may not represent separate ESUs), we

646 advocate continued recognition of two subspecies within the pink cockatoo.

647

648 Wildlife forensics implications

649 The generation of SNP data and the population genetic inferences presented in this study 650 could facilitate the development of wildlife forensic techniques for the pink cockatoo (Ogden, 2011). Typically, a species or subspecies identification test is based on analysis of mtDNA 651 652 due its high mutation rate, lack of recombination, availability of homologous reference data, 653 and the ease with which it is amplified and sequenced (Linacre & Tobe, 2011; Johnson et al., 654 2014). However, the lack of reciprocal monophyly of subspecies/populations in our analyses of mtDNA loci means they may not be suitable for performing a subspecies identification or 655 656 geographic provenance tests. Any forensic testing of pink cockatoo subspecies should 657 therefore rely on nuclear DNA markers. We have provided proof of concept that reliable 658 population identification testing can be performed in this species using as few as 20 SNPs (all 659 likelihood ratios were >28 when the prosecution hypothesis was correct). Including more 660 SNPs and samples would intuitively yield greater assignment power and confidence. 661 Furthermore, different SNPs could be selected that are more informative to identify individuals in certain subregions. Being able to identify source populations will help direct 662 enforcement and compliance resources to areas most vulnerable to illegal collection and 663 664 allow repatriation of seized animals to their subspecies/population of origin (Alacs & 665 Georges, 2008). This study serves as an example on how to construct subspecies/population

666 identification or geographic provenance tests for species with relatively shallow phylogenetic667 structure.

668

Additionally, the SNPs generated in this paper could be utilized in the development of
parentage testing. A parentage test of offspring along with their putative mother and/or father
could determine in an investigation whether a pink cockatoo is wild or captive-bred. SNPs
with a high MAF are particularly useful for parentage analysis (Andrews *et al.*, 2018); even
after filtering for missing data, locus quality, outliers, HWE, and linkage, the dataset
contained 176 SNPs with a MAF >0.4 (data not shown).

675

676 Benefits and caveats of this genome-wide SNP data set

This study serves as another example of genome-wide SNP data being able to resolve 677 678 populations where mtDNA and/or relatively few nuclear markers lacked resolution or were 679 misleading (Leslie & Morin, 2016; Rodríguez-Ezpeleta et al., 2016; Younger et al., 2017). Utilizing many genetic markers alleviates issues faced when basing important conservation 680 681 decisions and/or taxonomy on a small number of markers. Further, we have harnessed 682 advancements in museum genomics to successfully genotype numerous old museum 683 specimens (toe pads) collected over decades with reasonably high success (61% of toe pads 684 were successfully genotyped), which provided critical representation of the species' 685 distribution (Fig. 1a).

686

687 Utilizing SNPs generated from old museum specimens, however, presents a number of

688 issues. Even after filtering, some toe pad genotypes had high SNP error rates and

689 considerable missing data (Table S2). Although some troublesome genotypes were removed

690 from most analyses (Appendix I), given the variability of the error rates of toe pad samples,

691 some samples with relatively high error rates may not have been excluded during the filtering 692 stages. The error and missing data in these old museum specimens are expected to be biased 693 towards low diversity SNPs, and random errors are expected to homogenize genetic structure 694 (Ewart et al., 2019). Further, even though the pink cockatoo is a very long-lived species 695 (Brouwer et al., 2000), it is possible that the genetic structure may have changed over the 696 sampling period (i.e. genotyped individuals were sampled between 1912 and 2011). To ensure these biases did not lead to false population genetic inferences, we demonstrated that 697 698 comparable results were obtained when analysing a SNP dataset called exclusively from 699 more contemporary tissue samples. These results corroborate the value of using toe pad 700 samples genotyped with this platform to support spatial conclusions, but may present 701 problems for temporal inferences (Ewart et al., 2019).

702

703 Conclusion

This is the first species-wide genetic study on the pink cockatoo. The extensive dataset
provides a basis for effective conservation management for this species. Well-informed
management strategies based on genetics can now be implemented with the aim of
maximising the species' genetic diversity and its potential to adapt to changing environments
(Frankham, 2005; Huffman & Wallace, 2011). Further, preliminary analyses with these data
indicate they could be successfully used in the development of a wildlife forensic toolbox to
detect and prosecute trafficking crimes associated with this species.

711

712 The phylogeographic analyses we have performed represent a robust approach for

713 investigating species that are widespread, yet have shallow phylogeographic structure. SNPs

revealed subtle patterns of genetic differentiation that were not detected through analysis of

715 mtDNA and morphology.

716

717 Acknowledgements

We are grateful to the University of Sydney Industry and Community Engagement Seed 718 719 Fund, Birdlife Australia (Stuart Leslie Bird Research Award), Australian Museum 720 Foundation, Australian Government Research Training Program (RTP) Scholarship and 721 University of Sydney Merit Award, and Michael Sverns, Glenn Sharp, and Stephen Lavelle from the Department of Environment, Land, Water and Planning (State Government of 722 Victoria) for funding. We thank the following people/organizations for their help in obtaining 723 724 samples: Robert Palmer (Australian National Wildlife Collection), Scott Ginn, Leah Tsang 725 and Richard Major (Australian Museum), Ron Johnstone and Rebecca Bray (Western 726 Australian Museum), Claire Keely (Museum Victoria), and John Martin (Royal Botanic 727 Gardens, Sydney). We also thank Frankie Sitam and Jeffrine Rovie-Ryan for providing access and training at the National Wildlife Forensic Laboratory (Kuala Lumpur, Malaysia), 728 729 Linda Neaves for her assistance with the mtDNA PCRs and sequencing, Kate FitzGerald for 730 her assistance with the graphics, and Simon Ho and Mark Eldridge for their expert advice. 731 732 Collection of pink cockatoos for research purposes of this project was approved by the Australian Museum Animal Care and Ethics Committee (approval number 16-02). All earlier 733 734 samples had been collected under all appropriate ethics and scientific collecting permits. Sulphur-crested cockatoo samples were obtained as part of another project, under the ethics 735 736 permit AEC 151020-0. Transfer of pink cockatoo DNA from Australia to Malaysia was 737 authorized under CITES permit number PWS2018-AU-000019. 738

739 Conflict of interest: none

741 Data accessibility

742 SNP data and mtDNA sequence data will be made available on Dryad Digital Repository and743 GenBank respectively upon acceptance.

744

745 **References**

746

747 Adamack AT, Gruber B. 2014. PopGenReport: simplifying basic population genetic analyses
748 in R. *Methods in Ecology and Evolution* 5: 384-387.

Adams M, Baverstock PR, Saunders DR, Schodde R, Smith GT. 1984. Biochemical
systematics of the Australian cockatoos (Psittaciformes: Cacatuinae). *Australian Journal of Zoology* 32: 363-377.

Alacs E, Georges A. 2008. Wildlife across our borders: a review of the illegal trade in
Australia. *Australian Journal of Forensic Sciences* 40: 147-160.

Andrews KR, Adams JR, Cassirer EF, Plowright RK, Gardner C, Dwire M, Hohenlohe PA,
Waits LP. 2018. A bioinformatic pipeline for identifying informative SNP panels for
parentage assignment from RAD seq data. *Molecular Ecology Resources* 18: 12631281.

Ansari MH, Cooper SJB, Schwarz MP, Ebrahimi M, Dolman G, Saint KM, Donnellan SC,
Bull CM, Gardner MG. 2019. Plio-Pleistocene diversification and biogeographic
barriers in southern Australia reflected in the phylogeography of a widespread and
common lizard species. *Molecular Phylogenetics and Evolution* 133: 107-119.

762 Antao T, Lopes A, Lopes RJ., Beja-Pereira A, Luikart G. 2008. LOSITAN: A workbench to

763 detect molecular adaptation based on a F_{ST} -outlier method. *BMC Bioinformatics* 9: 764 323.

765 Archer FI, Adams PE, Schneiders BB. 2017. stratag: an r package for manipulating,

summarizing and analysing population genetic data. *Molecular Ecology Resources* 17:
5-11.

Baele G, Lemey P, Bedford T, Rambaut A, Suchard MA, Alekseyenko AV. 2012. Improving
the accuracy of demographic and molecular clock model comparison while
accommodating phylogenetic uncertainty. *Molecular Biology and Evolution* 29: 21572167.

- Ball Jr RM, Avise JC. 1992. Mitochondrial DNA phylogeographic differentiation among
 avian populations and the evolutionary significance of subspecies. *The Auk* 109: 626636.
- Baumsteiger J, Moyle PB, Aguilar A, O'Rourke SM, Miller MR. 2017. Genomics clarifies
 taxonomic boundaries in a difficult species complex. *PloS One* 12: e0189417.
- 777 Beaumont MA, Nichols RA. 1996. Evaluating loci for use in the genetic analysis of
- population structure. *Proceedings of the Royal Society of London B* 263: 1619-1626.
- Biodiversity Conservation Act, 2016. NSW State Government, Office of Environment and
 Heritage. Available at:
- 781 https://www.environment.nsw.gov.au/threatenedspeciesapp/profile.aspx?id=10116
- Blakers M, Davies SJJF, Reilly PN. 1984. *The atlas of Australian birds*. Melbourne:
 Melbourne University Press.
- Bouckaert R, Heled J, Kühnert D, Vaughan T, Wu CH, Xie D, Suchard MA, Rambaut A,
 Drummond AJ. 2014. BEAST 2: a software platform for Bayesian evolutionary
 analysis. *PLoS Computational Biology* 10: e1003537.
- Braby MF, Eastwood R, Murray N. 2012. The subspecies concept in butterflies: has its
 application in taxonomy and conservation biology outlived its usefulness?. *Biological Journal of the Linnean Society* 106: 699-716.
- Bradburd GS, Coop GM, Ralph PL. 2018. Inferring continuous and discrete population
 genetic structure across space. *Genetics* 210: 33-52.
- Brouwer K, Jones ML, King CE, Schifter H. 2000. Longevity records for Psittaciformes in
 captivity. *International Zoo Yearbook* 37: 299-316.
- Bryant D, Bouckaert R, Felsenstein J, Rosenberg NA, RoyChoudhury A. 2012. Inferring
 species trees directly from biallelic genetic markers: bypassing gene trees in a full
 coalescent analysis. *Molecular Biology and Evolution* 29: 1917-1932.
- 797 Cameron M. 2007. Cockatoos. Melbourne: CSIRO Publishing.
- Chambers EA, Hillis DM. 2020. The multispecies coalescent over-splits species in the case of
 geographically widespread taxa. *Systematic Biology* 69: 184-193.
- 800 Chessel D, Dufour AB, Thioulouse J. 2004. The ade4 package-I-One-table methods. *R News*801 4: 5-10.
- 802 Chhatre VE, Emerson KJ. 2017. StrAuto: Automation and parallelization of STRUCTURE
 803 analysis. *BMC Bioinformatics* 18: 192.
- Clement M, Posada DCKA, Crandall KA. 2000. TCS: a computer program to estimate gene
 genealogies. *Molecular Ecology* 9: 1657-1659.

- 806 Condon HT. 1975. *Checklist of the Birds of Australia: Non-passerines (Vol. 1)*. Melbourne:
 807 Royal Australasian Ornithologists Union.
- 808 Cruz VM, Kilian A, Dierig DA. 2013. Development of DArT marker platforms and genetic
 809 diversity assessment of the U.S. collection of the new oilseed crop lesquerella and
 810 related species. *PLoS ONE* 8: e64062.
- B11 Dierckxsens N, Mardulyn P, Smits G. 2017. NOVOPlasty: de novo assembly of organelle
 genomes from whole genome data. *Nucleic Acids Research* 45: e18.
- 813 Diniz-Filho JAF, Soares TN, Lima JS, Dobrovolski R, Landeiro VL, Telles MPDC, Rangel
- 814 TF, Bini LM. 2013. Mantel test in population genetics. *Genetics and Molecular*815 *Biology* 36: 475-485.
- B16 Dolman G, Joseph L. 2012. A species assemblage approach to comparative phylogeography of
 B17 birds in southern Australia. *Ecology and Evolution* 2: 354-369.
- 818 Dolman G, Joseph L. 2015. Evolutionary history of birds across southern Australia: structure,
- history and taxonomic implications of mitochondrial DNA diversity in an ecologically
 diverse suite of species. *Emu* 115: 35-48.
- Dolman G, Joseph L. 2016. Multi-locus sequence data reveal Pleistocene speciation in semiarid southern Australian birds (Cinclosoma spp.) was associated with increased genetic
 drift. *BMC Evolutionary Biology* 16: 226.
- Eberhard JR, Wright TF. 2016. Rearrangement and evolution of mitochondrial genomes in
 parrots. *Molecular Phylogenetics and Evolution* 94 (Part A): 34-46.
- Engelhard D, Joseph L, Toon A, Pedler L, Wilke T. 2015. Rise (and demise?) of subspecies in
 the Galah (*Eolophus roseicapilla*), a widespread and abundant Australian cockatoo. *Emu-Austral Ornithology* 115: 289-301.
- Ewart KM, Johnson RN, Ogden R, Joseph L, Frankham GJ, Lo N. 2019. Museum specimens
 provide reliable SNP data for population genomic analysis of a widely distributed but
 threatened cockatoo species. *Molecular Ecology Resources* 00: 1-15.
- 832 Ewart KM, Johnson RN, Ogden R, Joseph L, Ho SYW, Frankham G, Eldridge MDB, Lo N.
- 833 2020. Genome-wide SNP analyses identify novel evolutionarily significant units in an
 834 iconic Australian bird species, the red-tailed black-cockatoo (*Calyptorhynchus*835 *banksii*). *Heredity* 1-16.
- Excoffier L, Lischer HEL. 2010. Arlequin suite version 3.5: A new series of programs to
 perform population genetics analyses under Linux and Windows. *Molecular Ecology Resources* 10: 564-567.

- 839 Fenster CB, Ballou JD, Dudash MR, Eldridge MD, Frankham R, Lacy RC, Ralls K, Sunnucks
- P. 2018. Focus: ecology and evolution: conservation and genetics. *The Yale Journal of Biology and Medicine* 91: 491.
- 842 Flora and Fauna Guarantee Act, 1988. Victorian State Government, Department of
- 843 Environment, Land, Water & Planning. Available at:
- 844 https://www.environment.vic.gov.au/__data/assets/pdf_file/0024/115827/FFG-
- 845 Threatened-List.doc.pdf
- Ford J. 1974. Speciation in Australian birds adapted to arid habitats. *Emu* 74: 161-168.
- Ford J. 1987. Minor isolates and minor geographical barriers in avian speciation in continental
 Australia. *Emu* 87: 90-102.
- Ford J, Parker SA. 1973. A second species of wedgebill? *Emu* 73: 113-118.
- 850 Forshaw JM. 2011. *Parrots of the world*. Melbourne: CSIRO Publishing.
- Forshaw JM, Cooper WT. 1981. *Australian Parrots. Second edition*. Melbourne: Lansdowne
 Press.
- Forshaw JM, Cooper WT. 2002. *Australian Parrots. Third edition*. Melbourne: Lansdowne
 Press.
- 855 Frankham R. 2005. Genetics and extinction. *Biological Conservation* 126: 131-140.
- Frankham R, Ballou JD, Briscoe DA. 2010. *Introduction to conservation genetics. Second edition.* Cambridge: Cambridge University Press.
- 858 Frankham R, Ballou JD, Ralls K, Eldridge MDB, Dudash MR, Fenster CB, Lacy RC,
- 859 Sunnucks P. 2017. *Genetic management of fragmented animal and plant populations*.
 860 Oxford: Oxford University Press.
- Fu Y-X. 1997. Statistical tests of neutrality of mutations against population growth,
 hitchhiking and background selection. *Genetics* 147: 915-925.
- Funk DJ, Omland KE. 2003. Species-level paraphyly and polyphyly: frequency, causes, and
 consequences, with insights from animal mitochondrial DNA. *Annual Review of Ecology, Evolution, and Systematics* 34: 397-423.
- 866 Garnett ST, Szabo JK, Dutson G. 2011. *The action plan for Australian birds 2010*. Melbourne:
 867 CSIRO Publishing.
- Goudet J, Jombart T. 2015. hierfstat: Estimation and Tests of Hierarchical F-Statistics. R
 package version 0.04-22.
- Gruber B, Unmack PJ, Berry OF, Georges A. 2018. dartr: An r package to facilitate analysis
 of SNP data generated from reduced representation genome sequencing. *Molecular*
- 872 *Ecology Resources* 18: 691-699.

- Hall BP. 1974. *Birds of the Harold Hall Australian Expeditions: 1962-1970*. Trustees of the
 British Museum of Natural History, London.
- Hartl DL, Clark AG, Clark AG. 1997. *Principles of population genetics, Vol. 116*. Sunderland:
 Sinauer Associates.
- Higgins PJ. 1999. *Handbook of Australian, New Zealand & Antarctic Birds. Vol. 4, Parrots to Dollarbird*. Oxford: Oxford University Press.
- 879 Huang JP, Knowles LL. 2016. The species versus subspecies conundrum: Quantitative
- delimitation from integrating multiple data types within a single Bayesian approach in
 hercules beetles. *Systematic Biology* 65: 685-699.
- Huffman JE, Wallace JR. 2011. *Wildlife forensics: methods and applications*. Hoboken: John
 Wiley & Sons.
- Johnson RN, Wilson-Wilde L, Linacre A. 2014. Current and future directions of DNA in
 wildlife forensic science. *Forensic Science International: Genetics* 10: 1-11.
- Jombart T. 2008. adegenet: a R package for the multivariate analysis of genetic markers. *Bioinformatics* 24: 1403-1405.
- Joseph L, Wilke T. 2006. Molecular resolution of population history, systematics and
 historical biogeography of the Australian ringneck parrots *Barnardius*: are we there
 yet?. *Emu-Austral Ornithology* 106: 49-62.
- Joseph L, Wilke T, Ten Have J, Terry Chesser R. 2006. Implications of mitochondrial DNA
 polyphyly in two ecologically undifferentiated but morphologically distinct migratory
 birds, the masked and white-browed woodswallows Artamus spp. of inland Australia. *Journal of Avian Biology* 37: 625-636.
- Kamvar ZN, Brooks JC, Grünwald NJ. 2015. Novel R tools for analysis of genome-wide
 population genetic data with emphasis on clonality. *Frontiers in Genetics* 6: 208.
- Kamvar ZN, Tabima JF, Grünwald NJ. 2014. poppr: an R pack- age for genetic analysis of
 populations with clonal, partially clonal, and/or sexual reproduction. *PeerJ* 2: e281.
- Kearns A, Joseph L, Double M, Edwards S. 2009. Inferring the phylogeography and
 evolutionary history of the Splendid Fairy-wren (*Malurus splendens*) from
- 901 mitochondrial DNA and spectrophotometry. *Journal of Avian Biology* 40: 7-17.
- 902 Kearns AM, Joseph L, White LC, Austin JJ, Baker C, Driskell AC, Malloy JF, Omland KE.
- 903 2016. Norfolk Island Robins are a distinct endangered species: ancient DNA unlocks
- 904 surprising relationships and phenotypic discordance within the Australo-Pacific
- 905 Robins. *Conservation Genetics* 17: 321-335.

906	Kearns AM, White LC, Austin JJ, Omland KE. 2015. Distinctiveness of Pacific Robin
907	subspecies in Vanuatu revealed from disparate patterns of sexual dichromatism,
908	plumage colouration, morphometrics and ancient DNA. Emu-Austral Ornithology 115:
909	89-98.
910	Kearse M, Moir R, Wilson A, Stones-Havas S, Cheung M, Sturrock S, Buxton S, Cooper A,
911	Markowitz S, Duran C, Thierer T, Ashton B, Mentjies P, Drummond A. 2012.
912	Geneious Basic: an integrated and extendable desktop software platform for the
913	organization and analysis of sequence data. Bioinformatics 28: 1647-1649.
914	Kilian A, Wenzl P, Huttner E, Carling J, Xia L, Blois H, Aschenbrenner-Kilian M. 2012.
915	Diversity arrays technology: a generic genome profiling technology on open platforms.
916	In Data production and analysis in population genomics. Totowa: Humana Press, 67-
917	89.
918	Kopelman NM, Mayzel J, Jakobsson M, Rosenberg NA, Mayrose I. 2015. Clumpak: a
919	program for identifying clustering modes and packaging population structure
920	inferences across K. Molecular Ecology Resources 15: 1179-1191.
921	Latch EK, Reding DM, Heffelfinger JR, Alcalá-Galván CH, Rhodes Jr OE. 2014. Range-wide
922	analysis of genetic structure in a widespread, highly mobile species (Odocoileus
923	hemionus) reveals the importance of historical biogeography. Molecular Ecology 23:
924	3171-3190.
925	Legendre P, Fortin MJ. 2010. Comparison of the Mantel test and alternative approaches for
926	detecting complex multivariate relationships in the spatial analysis of genetic data.
927	Molecular Ecology Resources 10: 831-844.
928	Legendre P, Fortin MJ, Borcard D. 2015. Should the Mantel test be used in spatial analysis?.
929	Methods in Ecology and Evolution 6: 1239-1247.
930	Leigh JW, Bryant D. 2015. popart: full-feature software for haplotype network construction.
931	Methods in Ecology and Evolution 6: 1110-1116.
932	Leslie MS, Morin PA. 2016. Using genome-wide SNPs to detect structure in high-diversity
933	and low-divergence populations of severely impacted eastern tropical Pacific spinner
934	(Stenella longirostris) and pantropical spotted dolphins (S. attenuata). Frontiers in
935	Marine Science 3: 253.
936	Li YL, Liu JX. 2018. StructureSelector: A web based software to select and visualize the
937	optimal number of clusters using multiple methods. Molecular Ecology Resources 18:
938	176–177.

- 939 Linacre A, Tobe SS. 2011. An overview to the investigative approach to species testing in
 940 wildlife forensic science. *Investigative Genetics* 2: 2.
- 941 Mackowski CM. 1984. The ontogeny of hollows in blackbutt (*Eucalyptus pilularis*) and its
- 942 relevance to the management of forests for possums, gliders and timber. In: Smith AP,
- 943 Hume ID, eds. *Possums and Gliders*. Sydney: Australian Mammal Society and Surrey
- 944 Beatty and Sons, 553-67.
- 945 Maddison WP. 1997. Gene trees in species trees. *Systematic Biology* 46: 523-536.
- 946 Marie AD, Stockwell BL, Rico C. 2019. DNA analysis of juvenile scalloped hammerhead
- 947 sharks *Sphyrna lewini* (Griffith & Smith 1834) reveals multiple breeding populations
 948 and signs of adaptive divergence in the South Pacific. *Frontiers in Marine Science* 6:
 949 718.
- Mastrantonio V, Porretta D, Urbanelli S, Crasta G, Nascetti G. 2016. Dynamics of mtDNA
 introgression during species range expansion: insights from an experimental
 longitudinal study. *Scientific Reports* 6: 30355.
- 953 Mastretta-Yanes A, Arrigo N, Alvarez N, Jorgensen TH, Piñero D, Emerson BC. 2015.
- 954 Restriction site-associated DNA sequencing, genotyping error estimation and de novo
 955 assembly optimization for population genetic inference. *Molecular Ecology Resources*956 15: 28–41.
- 957 Mathews GM. 1912. *The Austral Avian Record: A Scientific Journal Devoted Primarily to the*958 *Study of the Australian Avifauna (Volume 1)*. London: Witherby & Company.
- McElroy K, Beattie K, Symonds MR, Joseph L. 2018. Mitogenomic and nuclear diversity in
 the Mulga Parrot of the Australian arid zone: cryptic subspecies and tests for selection. *Emu-Austral Ornithology* 118: 22-35.
- 962 Melville J, Haines ML, Boysen K, Hodkinson L, Kilian A, Smith Date KL, ... Parris KM.
- 963 2017. Identifying hybridization and admixture using SNPs: application of the DArTseq
 964 platform in phylogeographic research on vertebrates. *Royal Society Open Science* 4:
 965 161061.
- 966 Menkhorst P, Rogers D, Clarke R, Davies J, Marsack P, Franklin K. 2017. *The Australian bird*967 *guide*. Melbourne: CSIRO Publishing, 260.
- 968 Moritz C. 1994. Defining "evolutionarily significant units" for conservation. Trends in
 969 *Ecology and Evolution* 9: 373–375.
- 970 Nature Conservation (Wildlife) Regulation, 2006. Queensland State Government. Available
 971 at: https://www.legislation.qld.gov.au/view/pdf/inforce/2010-05-21/sl-2006-0206.

972 Neaves LE, Zenger KR, Prince RIT, Eldridge MDB, Cooper DW. 2009. Landscape

- 973 discontinuities influence gene flow and genetic structure in a large, vagile Australian
 974 mammal, *Macropus fuliginosus*. *Molecular Ecology* 18: 3363–78.
- 975 Neaves LE, Zenger KR, Prince RI, Eldridge MD. 2012. Impact of Pleistocene aridity
- 976 oscillations on the population history of a widespread, vagile Australian mammal,
 977 *Macropus fuliginosus. Journal of Biogeography* 39: 1545-1563.
- 978 Ogden R. 2011. Unlocking the potential of genomic technologies for wildlife forensics.
 979 *Molecular Ecology Resources* 11: 109-116.
- Ogden R, Thorpe RS. 2002. Molecular evidence for ecological speciation in tropical habitats.
 Proceedings of the National Academy of Sciences 99: 13612-13615.
- 982 Omland KE, Baker JM, Peters JL. 2006. Genetic signatures of intermediate divergence:
 983 population history of Old and New World Holarctic ravens (*Corvus corax*). *Molecular*984 *Ecology* 15: 795-808.
- Paetkau D, Calvert W, Stirling I, Strobeck C. 1995. Microsatellite analysis of population
 structure in Canadian polar bears. *Molecular Ecology* 4: 347-354.
- Palsbøll PJ, Berube M, Allendorf FW. 2007. Identification of management units using
 population genetic data. *Trends in Ecology and Evolution* 22: 11-16.
- Paradis E. 2010. pegas: an R package for population genetics with an integrated-modular
 approach. *Bioinformatics* 26: 419-420.
- 991 Peakall R, Smouse PE. 2006. GENALEX 6: Genetic analysis in Excel. Population genetic
 992 software for teaching and research. *Molecular Ecology Notes* 6: 288-295.
- Peakall R, Smouse PE. 2012. GenAlEx 6.5: Genetic analysis in Excel. Population genetic
 software for teaching and research-an update. *Bioinformatics* 28: 2537-2539.
- 995 Peters JL. 1937. *Check-list of birds of the world, volume III*. Cambridge: Harvard University
 996 Press, 176.
- 997 Piry S, Alapetite A, Cornuet JM, Paetkau D, Baudouin L, Estoup A. 2004. GENECLASS2: a
 998 software for genetic assignment and first-generation migrant detection. *Journal of*999 *Heredity* 95: 536-539.
- Porras-Hurtado L, Ruiz Y, Santos C, Phillips C, Carracedo Á, Lareu M. 2013. An overview of
 STRUCTURE: applications, parameter settings, and supporting software. *Frontiers in Genetics* 4: 98.
- Pritchard JK, Stephens M, Donnelly P. 2000. Inference of population structure using
 multilocus genotype data. *Genetics* 155: 945-959.

- 1005 Ralls K, Ballou JD, Dudash MR, Eldridge MD, Fenster CB, Lacy RC, Sunnucks P, Frankham
- 1006 R. 2018. Call for a paradigm shift in the genetic management of fragmented
 1007 populations. *Conservation Letters* 11: e12412.
- 1008 Rambaut A. 2009. FigTree: tree figure drawing tool version 1.4.2. Available at:
 1009 <u>http://tree.bio.ed.ac.uk/software/figtree</u>.
- 1010 Rambaut A, Suchard MA, Xie D, Drummond AJ. 2014. Tracer v1.6. Available at:
 1011 http://beast.bio.ed.ac.uk/Tracer.
- 1012 Ramos-Onsins SE, Rozas J. 2002. Statistical properties of new neutrality tests against
 1013 population growth. *Molecular Biology and Evolution* 19: 2092-2100.
- 1014 Remsen Jr JV. 2005. Pattern, process, and rigor meet classification. *The Auk* 122: 403-413.
- 1015 Rodríguez-Ezpeleta N, Bradbury IR, Mendibil I, Álvarez P, Cotano U, Irigoien X. 2016.

1016 Population structure of Atlantic mackerel inferred from RAD-seq-derived SNP

- 1017 markers: effects of sequence clustering parameters and hierarchical SNP selection.
- 1018 *Molecular Ecology Resources* 16: 991-1001.
- Rogers AR, Harpending H. 1992. Population growth makes waves in the distribution of
 pairwise genetic-differences. *Molecular Biology and Evolution* 9: 552-569.
- 1021 Rowe KC, Singhal S, Macmanes MD, Ayroles JF, Morelli TL, Rubidge EM, Bi KE, Moritz
- 1022 CC. 2011. Museum genomics: low-cost and high-accuracy genetic data from historical
 1023 specimens. *Molecular Ecology Resources* 11: 1082-1092.
- 1024 Rowley I, Chapman G. 1991. The breeding biology, food, social-organization, demography
- 1025and conservation of the Major Mitchell or pink cockatoo, Cacatua-Leadbeateri, on the1026margin of the Western Australian wheat-belt. Australian Journal of Zoology 39: 211-1027261.
- Rozas J, Ferrer-Mata A, Sanchez DelBarrio JC, Guirao-Rico S, Librado P, Ramos-Onsins SE,
 Sánchez-Gracia A. 2017. DnaSP v6: DNA sequence polymorphism analysis of large
 datasets. *Molecular Biology and Evolution* 34: 3299-3302.
- 1031 Ryder OA. 1986. Species conservation and systematics: The dilemma of subspecies. *Trends in* 1032 *Ecology and Evolution* 1: 9-10.
- 1033 Safran RJ, Scordato ESC, Wilkins MR, Hubbard JK, Jenkins BR, Albrecht T, ... Nosil P.
- 10342016. Genome-wide differentiation in closely related populations: the roles of selection
- and geographic isolation. *Molecular Ecology* 25: 3865-3883.

- Schirtzinger EE, Tavares ES, Gonzales LA, Eberhard JR, Miyaki CY, Sanchez JJ, ... Wright
 TF. 2012. Multiple independent origins of mitochondrial control region duplications in
 the order Psittaciformes. *Molecular Phylogenetics and Evolution* 64: 342-356.
- Schodde R. 1982. Origin, adaptation and evolution of birds in arid Australia. In: Barker WR,
 Greenslade PJM, eds. *Evolution of flora and fauna of arid Australia*. Adelaide:
- 1041 Peacock Publications, 191-224.
- 1042 Schodde R. 1994. The bird fauna of western New South Wales: geography and status. In:
- Lunney D, Hand D, Reed P, Butcher D, eds. *Future of the fauna of western New South Wales*. Mosman: Royal Zoological Society of NSW, 107-121.
- 1045 Schodde R. 1997. Cacatuidae. In: Houston WWK, Wells A, eds. *Zoological Catalogue of*1046 *Australia. Vol. 37.2: Aves (Columbidae to Coraciidae).* Melbourne: CSIRO
 1047 Publishing, 64-108.
- 1048 Schodde R, Mason IJ. 1999. Directory of Australian birds: passerines: Passerines.
- 1049 Melbourne: CSIRO publishing.
- Shirk AJ, Landguth EL, Cushman SA. 2017. A comparison of individual-based genetic
 distance metrics for landscape genetics. *Molecular Ecology Resources* 17: 1308-1317.
- 1052 Sorenson MD. 2003. Avian mtDNA Primers. Available at: http://people.bu.edu/msoren/
 1053 primers.html.
- Stamatakis A. 2014. RAxML version 8: a tool for phylogenetic analysis and post-analysis of
 large phylogenies. *Bioinformatics* 30: 1312-1313.
- Tajima F. 1989. Statistical method for testing the neutral mutation hypothesis by DNA
 polymorphism. *Genetics* 123: 585-595.
- Teske PR, Golla TR, Sandoval-Castillo J, Emami-Khoyi A, van der Lingen CD, von der
 Heyden S, ... Beheregaray LB. 2018. Mitochondrial DNA is unsuitable to test for
 isolation by distance. *Scientific Reports* 8: 8448.
- Toews DPL, Brelsford A. 2012. The biogeography of mitochondrial and nuclear discordance
 in animals. *Molecular Ecology* 21: 3907-3930.
- Tonzo V, Papadopoulou A, Ortego J. 2019. Genomic data reveal deep genetic structure but no
 support for current taxonomic designation in a grasshopper species complex.
 Molecular Ecology 28: 3869-3886.
- Toon A, Mather PB, Baker AM, Durrant KL, Hughes JM. 2007. Pleistocene refugia in an arid
 landscape: analysis of a widely distributed Australian passerine. *Molecular Ecology*16: 2525-2541.

1069 Walker IS, Sluiter IRK, Hawker P. 1999. Flora and fauna guarantee action statement, Major

- 1070 Mitchell's Cockatoo *Cacatua leadbeateri*. Department of Sustainability and
- 1071 Environment. Available at:
- 1072 https://www.environment.vic.gov.au/__data/assets/pdf_file/0025/32875/Major_Mitche
- 1073 lls_Cockato_Cacatua_leadbeateri.pdf
- 1074 Weir BS, Cockerham CC. 1984. Estimating F-statistics for the analysis of population
 1075 structure. *Evolution* 38: 1358-1370.
- 1076 White NE, Phillips MJ, Gilbert MTP, Alfaro-Núñez A, Willerslev E, Mawson PR, Spencer
 1077 PB, Bunce M. 2011. The evolutionary history of cockatoos (Aves: Psittaciformes:
 1078 Cacatuidae). *Molecular Phylogenetics and Evolution* 59: 615-622.

1079 Willing EM, Dreyer C, van Oosterhout C. 2012. Estimates of genetic differentiation measured 1080 by F_{ST} do not necessarily require large sample sizes when using many SNP markers.

1081 *PLOS ONE* 7: e42649.

1082 Younger JL, Clucas GV, Kao D, Rogers AD, Gharbi K, Hart T, Miller KJ. 2017. The

- 1083 challenges of detecting subtle population structure and its importance for the
 1084 conservation of emperor penguins. *Molecular Ecology* 26: 3883-3897.
- Zheng X, Levine D, Shen J, Gogarten SM, Laurie C, Weir BS. 2012. A high-performance
 computing toolset for relatedness and principal component analysis of SNP data.

1087 *Bioinformatics* 28: 3326-3328.

- Zink RM. 2004. The role of subspecies in obscuring avian biological diversity and misleading
 conservation policy. *Proceedings of the Royal Society of London. Series B: Biological Sciences* 271: 561-564.
- 1091 Zink RM, Barrowclough GF. 2008. Mitochondrial DNA under siege in avian phylogeography.
 1092 *Molecular Ecology* 17: 2107-2121.
- 1093

1094 Figure and table legends

1095

1096 Figure 1. (a) The distribution of Lophochroa leabeateri leadbeateri (blue) and L. l. mollis 1097 (orange) in Australia, adapted from Schodde (1994) and Menkhorst et al. (2017), and 1098 localities of the frozen tissue samples (stars) and toe pads samples (circles) genotyped in this study. The thick grey line represents the Eyrean Barrier, the darker shading represents core 1099 1100 breeding zones, and the lighter shading and blurred fringes represent areas of potentially sparser distribution and/or non-breeding based on records from the Atlas of Living Australia 1101 database (https://www.ala.org.au; accessed 4 November 2020). (b) A PCoA plot for 57 pink 1102 cockatoo individuals using 4,135 SNPs. (c) and (d) STRUCTURE plots for 57 pink cockatoo 1103 individuals based on 2,131 SNPs when K=2 and K=3, respectively. The bottom-left photo is 1104 1105 of L. l. leadbeateri, Mt. Hope, NSW. Photo: Corey Callaghan.

- 1106
- **Figure 2.** (a) TCS-based haplotype network analysis based on 19 pink cockatoo individuals
- using 2,037 bp of concatenated ND2, ND4 and ND5 genes. (b) Phylogeny of the pink
- 1109 cockatoo based on mtDNA data (see Table S1 for samples details), isolated from Fig. S12a
- 1110 (outgroup removed for clarity). Bayesian posterior probabilities are given above relevant
- branches. The 'CQ' and 'SW' labels next to the haplotypes (a) and taxon names (b) represent
- samples from central Queensland and south-west Western Australia respectively (see Fig. S3
- for additional details). NB: the common haplotype in the haplotype network (a) contains
- haplotypes from both *L. l. leadbeateri* populations and the south-western Western Australia *L. l. mollis* population, but not the more north-easterly *L. l. mollis* population.
- 1115

Figure 3. (a) Genetic divergence of populations along a transect based on inter-population pairwise $F_{ST}/(1 - F_{ST})$ calculated using 2,131 SNPs, divided by pairwise geographic distance, and plotted against the midpair distance of adjacent localities. (b) For this analysis, 15 pink cockatoo individuals were divided into five 'sample clusters' (3 individuals per cluster) along a transect. The vertical dotted red line in (a) indicates the pairwise comparison across the putative subspecies barrier.

1123

Table 1. Genetic diversity measurements based on 2,131 SNPs in 56 pink cockatoo

individuals, and 2,037 bp of concatenated *ND2*, *ND4* and *ND5* genes in 19 pink cockatoo
individuals. Genetic diversity was measured within subspecies. Note, the haplotype common

1127 to both subspecies (see Fig. 2a) was counted twice in the 'number of haplotypes.'

1128

1129 Figure 4. DAPC analyses showing separation between *Lophochroa leadbeateri mollis*

- 1130 (orange) and L. l. leadbeateri (blue). The analyses were based on 49 pink cockatoo
- 1131 individuals using (a) 1,307 SNPs, and (b) 20 informative SNPs.