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1	Spatiotemporal distribution of canopy litter and nutrient resorption in a
2	chronosequence of different development stages of <i>Cunninghamia lanceolata</i> in
3	southeast China
4	
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23 Abstract

Canopy litter is an important component of coarse woody debris (CWD), which 24 25 affects nutrient and carbon cycling in forest ecosystems. For marcescent plant species (characterized by dead branches and leaves remaining in the canopy for several years 26 27 before abscission), nutrient resorption from senescing leaves is an important nutrient conservation strategy. However, investigating the ecological function of canopy litter 28 is challenging due to its limited accessibility and also the heterogeneous canopy 29 microclimate in terms of light transmission, temperature and moisture. We studied the 30 spatiotemporal distribution of canopy litter mass and seasonal dynamics of leaf 31 nutrients and nutrient resorption during senescence in the canopy along a 32 chronosequence of Chinese fir [Cunninghamia lanceolata (Lamb.) Hook] plantations 33 34 in southeast China. The dry mass weight of dead branches and dead leaves in the canopy significantly increased with stand stage (14.6, 14.2, and 17.4 t ha⁻¹ for young, 35 middle-aged, and mature stands respectively), accounting for high proportions of total 36 aboveground litter of 85.7%, 79.1% and 80.0%, respectively, along with annual 37 litterfall production (2.44, 3.75, and 4.34 t ha⁻¹, respectively). The canopy height 38 distribution of dead branches and leaves also increased with stand age, ranging from 39 0-4 m in young stands, 3–8 m in middle-aged stands, to 4–10 m in mature stands. The 40 seasonal pattern of canopy litter mass was the inverse of litterfall production: canopy 41 litter mass peaked, whilst litterfall production was lowest in winter. Mean N, P, K, and 42 Mg nutrient resorption efficiencies across stands at each stage were 53.8-58.9%, 43 64.0-68.9%, 85.0-90.2%, and 46.5-56.6%, respectively, while Ca was not 44

retranslocated from senescing leaves. In summary, Chinese fir plantations retain large amounts of dead branches and leaves in the canopy from which at least ~50% of the nutrients N, P, K and Mg are recycled, representing an important nutrient conservation strategy that has evolved to adapt to nutrient-limited habitats. Canopy litter therefore plays an important role in these forest plantation ecosystems and should be protected instead of being removed from the canopy to the forest floor.

51 Key words: canopy litter, nutrient concentration, nutrient resorption, nutrient
52 recycling, *Cunninghamia lanceolata* (Lamb.) Hook

53

54 **1. Introduction**

Nutrient resorption and litterfall kinetics may play an important role in modifying the 55 return of nutrients to the soil and litterfall chemistry, thereby strongly influencing soil 56 nutrient turnover and fertility. Nutrient resorption is the process by which nutrients are 57 retranslocated from senescing organs to living or storage organs (Aerts, 1996). It is 58 59 also a key nutrient conservation strategy in perennial plants and is frequently studied to understand the internal cycling of nutrients in plants (Brant and Chen, 2015; Zhou 60 61 et al., 2016a). For example, mineral nutrients are absorbed in the tree crown before 62 leaf shedding, which enhances the residence time of nutrients within plant tissues and improves internal nutrient recycling and use efficiency (Cleveland et al., 2013; 63 Maillard et al., 2015; Yan et al., 2015). On average, perennial plants reabsorb more 64 than half of leaf N and P pools during leaf senescence (Van Heerwaarden et al., 65 2003a). For example, Liu et al. (2014) investigated resorption of 13 nutrient elements 66

in karst vegetation and reported that four nutrients (N, P, K and Mg) were resorbed in
varying degrees from senescing leaves, with the extent of resorption increasing with
decreasing concentrations of these nutrients in green leaves. Therefore, insights into
nutrient resorption patterns are important for understanding nutrient cycling in forest
ecosystems.

72

Nutrient resorption is usually quantified as resorption efficiency (RE), which is 73 defined as the ratio between the amount of nutrients reabsorbed prior to leaf shedding 74 75 and leaf nutrient content before the onset of leaf senescence (Huang et al., 2007). For planted forests worldwide, mean nitrogen and phosphorus RE values are 59.0% and 76 60.2%, respectively, with significantly lower nitrogen REs in tropical and subtropical 77 78 forests (Jiang et al., 2019). Differences in RE values reported are associated with leaf life span (evergreen vs. deciduous), leaf nutrient content, and soil nutrient availability 79 (Yuan and Chen, 2015; Tsujii et al., 2017). For instance, nutrient RE is higher in 80 81 deciduous species and graminoids than in evergreen species and forbs (Aert, 1996; Han et al. 2013). However, the relationship between nutrient RE and soil nutrient 82 availability and leaf nutrient status is not clear-cut. Many studies have suggested that 83 plants would resorb more nutrients in nutrient-poor environments (Tully et al. 2013; 84 Yuan and Chen, 2015), but this has been disputed (Tang et al. 2013; See et al. 2015). 85 In addition, whether and how nutrient resorption is regulated by plant nutrient status 86 87 across different functional types and forest ecosystem remains unresolved (Vergutz et al. 2012; Zhou et al. 2016a). 88

The canopy litter (dead branches and leaves that remain attached to the trunk) is not 90 only a component of the aboveground coarse woody detritus (CWD) pool, but also of 91 the nutrient and carbon pool (Fonte and Schowalter, 2004; Yoshida and Hijii, 2006). 92 In marcescent species, such as the conifers Cryptomeria japonica and Pseudotsuga 93 menziesii, dead branches and leaves remain attached to the trunk within the canopy 94 long after death (Ishii and Kadotani, 2006; Yoshida and Hijii, 2006). Some 95 broad-leaved trees, such as Populus euphratica, Fagus grandiflora and many 96 Metrosideros forests also retain brown leaves within the canopy for some time 97 (Escudero and Arco, 1987; Shi et al., 2017). This results in accumulation of large 98 amounts of dead branches and leaves in the canopy. For example, canopy litter 99 100 (foliage, reproductive tissue and fine woody debris) accounted for 66-86% of the mass, 63–90% of the N, and 49–92% of the P returned annually in aboveground litter, 101 dominating recycling of N and P in three Rocky Mountain forests (Laiho and Prescott, 102 1999). The canopy litter plays an important role in nutrient cycling through 103 withdrawing nutrients from senescing tissues and providing litterfall to the forest floor 104 (Prescott, 2002). However, the spatial and temporal distributions of canopy litter have 105 not been reported in previous studies as they are difficult to estimate accurately (Ishii 106 and Wilson, 2001; Ishii and Kadotani, 2006). Furthermore, the nutrient concentrations 107 of leaves have been shown in agricultural crop plants to change substantially between 108 109 senescence and death, significantly influencing the nutrient RE (Osaki and Shinano, 2001). Most previous studies of nutrient RE in trees have focused on comparing 110

nutrient concentrations between green leaves and either senescent leaves or litterfall,
but the changes in nutrient concentrations from leaf senescence to death are not well
known across different stand development stages.

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Chinese fir (Cunninghamia lanceolata (Lamb.) Hook) is an endemic coniferous 115 species in China with high yield of timber and excellent wood quality. It has a 116 planting history of more than 1000 years and is distributed across 17 provinces, 117 ranging from 21°31' to 34°03' N, and 101°30' to 121°53' E (Yu, 2000). The planted 118 area and stand volumes of Chinese fir account for 23.2% and 29.2% of the total 119 plantation area and stand volume in China, respectively, the highest of any tree 120 species (State Forestry Administration, 2014). Chinese fir is characterized by a 121 122 canopy containing large amounts of dead leaves and branches which remain attached to the trunk for more than 4 years (Sheng and Fan, 2002) (Fig. S1). Previous studies 123 reported that the mass of dead leaves and branches increased with stand stage from 124 0.70-1.26 t ha⁻¹ in young stands to 6.51-9.43 t ha⁻¹ in middle-aged stands (Sheng and 125 Fan, 2002). In 10-year-old Chinese fir, the mass of dead branches and leaves in the 126 canopy accounted for 95% of the total aboveground litter, and remained at ~50% even 127 after substantial increase in litterfall production in trees >15-years-old (Liao et al., 128 1996). Several studies have demonstrated that the decay rates of dead branches and 129 leaves in the canopy are much slower than litter decomposition on the forest floor, 130 leading to less nutrient return than in forest soil (Zhang and Sheng, 2001; Gao et al., 131 2015). However, canopy litter mass and its spatiotemporal distribution in Chinese fir 132

has not been quantified, due to the difficulty of sampling.

134

In a previous study conducted at Xinkou National Forest Farm, but at different 135 locations and different-aged stands compared to this research and only comparing 136 green and litterfall leaves, we showed that fresh Chinese fir leaves absorb 50% N and 137 70% P before litterfall to the forest floor (Zhou et al., 2016a). Similar values have 138 been reported for Larix spp. (a dominant coniferous species in China), with resorption 139 from senescing leaves of 72.6%, 69.7% and 44.7% of N, P and K, respectively, but 140 141 not Mg and Ca (Yan et al., 2015). Lower N and P resorption efficiencies of 38% and 46%, respectively, reported for Chinese fir (Liu, 2014), perhaps arise from neglecting 142 the leaf mass loss and leaf shrinkage during senescence (Van Heerwaarden et al., 143 144 2003a). Litter retention in the canopy is not only closely linked with litterfall production and nutrient input into soil, but also with internal nutrient recycling. Long 145 retention times of litter in the canopy and translocation of nutrients to other tissues 146 147 before senescence should potentially increase nutrient reuse efficiency (Escudero et al., 1992; Niinemets and Tamm, 2005). This may be an important adaptation strategy, 148 which enables plants to use nutrients directly from senescent organs instead of 149 absorbing them from soil nutrient pools, helping plants to survive in nutrient-deficient 150 environments (Kobe et al., 2005; Lü et al., 2016). 151

152

We hypothesized that: (1) the spatiotemporal distribution of canopy litter variessignificantly with chronosequence-development stages in Chinese fir plantations; (2)

nutrient concentrations of needles decreased from growth to senescence and varied 155 seasonally; (3) foliar nutrient resorption varied among stand stages and seasons, and 156 nutrients required in large quantities (such as N, P and K) would be resorbed more 157 effectively than nutrients required in small amounts (such as Ca and Mg). The specific 158 objectives of this study were to investigate the spatiotemporal dynamics of canopy 159 litter mass and determine the nutrient (N, P, K, Ca, and Mg) RE and seasonal patterns 160 of dead canopy leaves among young, middle, and mature aged Chinese fir plantations. 161 Our results contribute towards a better understanding of plant adaptation strategies to 162 nutrient availability, and forest nutrient cycling processes, including internal nutrient 163 recycling and external fluxes. 164

165

166 2. Materials and methods

167 **2.1 Study site**

This study was conducted at Xinkou National Forest Farm (26°07'-27°13' N, 168 117°27'-118°14' E), Sanming City, Fujian province, China (Fig. S1). The elevation 169 was ~205-500 m above sea level. The region is characterized by a subtropical 170 monsoon climate with mean annual temperature of 17-19.4 °C, and mean annual 171 precipitation of 1600–1800 mm. The mean annual evaporation is 1585 mm, mean 172 annual relative humidity 81.0%, and the frost-free period is 240–300 days (Zhou et al. 173 2015). The soil was classified as Silty Oxisol (according to US Taxonomy), developed 174 175 on sandstone and shale parent material.

The study area was originally covered by broad-leaved forests that have now been 177 replaced by fast-growing Chinese fir plantations with a short rotation of typically 178 20-30 years (Wu, 1984). Three first-generation monoculture Chinese fir plantation 179 stands of ages 8, 16 and 25 years (established in 2008, 2000 and 1991), representing 180 young, middle, and mature ages, respectively, were selected for the study within a 2 181 km distance. The young and middle-aged stands face northeast and the mature stand is 182 located on a southeast facing slope. All stands were established from seedlings planted 183 in hand-dug holes after clear-cutting and burning of the existing forest. The young 184 185 stage was unthinned, whilst the middle-aged and mature stands were thinned following the standard practice for Chinese fir plantations by removing alternate rows 186 and cutting the crowns of thinned trees on site. The middle-aged stand was thinned in 187 188 2008, and the mature stand in 2000 and 2008. The dominant understory comprised shrubs and grasses, such as Woodwardia japonica, Maesa japonica, Ilex pubescens, 189 Selaginella doederleinii and Miscanthus floridulus. In each stand, three permanent 190 replicate plots (20 m \times 20 m) were established. Soil properties across different 191 developmental stages were: pH 4.24–4.58, soil organic matter 35.3–42.4 g kg⁻¹, bulk 192 density 1.25–1.33 g cm⁻³ and C:N ratio 10.9–13.1. 193

194

195 **2.2 Canopy litter biomass production**

Canopy litter was collected in January, April, July, and October 2015. After measuring
tree diameter at breast height (DBH), tree height, crown diameter and the height
above ground of the lowest dead branch for all the trees in the plot (Table 1), 16

sample trees were tagged in each plot, of which a separate four trees were sampled on 199 each date (Fig. S2). Above the lowest dead branch, all canopy litter was collected 200 201 from each sample tree at 1-m height intervals for the young stage (from a ladder placed against the trunk), and 2-m intervals for the middle and mature stages (using 202 203 branch clippers). If the canopy litter was too high to be accessed by these methods, it was collected by climbing the tree (Fig. S3). Plastic sheets were placed on the forest 204 floor around each sample tree to collect any canopy litter detached during the canopy 205 collection. The fresh weights of all samples of dead branches and dead leaves from 206 each height interval were measured in the field. Sub-samples were taken to determine 207 moisture content in the laboratory to allow the fresh weights to be converted to dry 208 weight. Since there were no obvious branch gaps in the canopy, it was assumed that 209 210 the distribution of canopy litter by height represents the original branch location and that there has not been substantial downwards movement of canopy litter. 211

Table 1. Plot characteristics in Chinese fir plantation stands at different development stages (young, middle and mature). Tree DBH, height, height above

- 213 ground of lowest dead branch, and crown diameter are expressed as mean ± SE of all the trees in each plot.
- 214

Stage	Plot number	Year of planting	DBH (cm)	Tree height (m)	Height above ground of	Crown diameter	Canopy density	Stand density (stem ha ⁻¹)	Slope (°) and aspect
					lowest dead branch (m)	(m)			
Young	1	2008	8.78±0.22	7.69±0.13	0.09±0.01	2.44±0.06	0.9	1789	20° NE
	2	2008	11.1±0.22	9.25±0.11	0.13±0.01	3.08±0.06	0.8	2811	18° NE
	3	2008	11.5±0.25	9.50±0.14	0.14±0.01	3.29±0.07	0.8	2547	17° NE
Middle	1	2000	13.2±0.60	10.1±0.46	2.93±0.13	5.72±0.26	0.8	1775	29° NE
	2	2000	14.2±0.47	11.8±0.39	3.47±0.12	6.23±0.22	0.7	1880	28° NE
	3	2000	15.5±0.53	11.9±0.41	3.44±0.12	6.74±0.23	0.7	1802	26° NE
Mature	1	1991	18.9±0.72	15.3±0.59	5.72±0.22	8.58±0.33	0.6	1475	21° SE
	2	1991	20.9±0.62	17.4±0.52	9.95±0.30	9.52±0.28	0.5	1650	20° SE
	3	1991	21.5±0.62	17.4±0.45	9.09±0.27	8.72±0.26	0.5	1580	22° SE

216 **2.3 Litterfall production**

Litterfall was collected in litter traps $(1 \text{ m} \times 1 \text{ m})$ installed at 1 m above the ground at 217 five randomly selected locations in each plot in October 2014. Chinese fir litter was 218 collected from each trap in January, April, July, and October 2015, with litter from 219 other tree species or understory vegetation discarded. The fresh weight of the Chinese 220 fir litter (unsorted leaves and twigs) was measured and sub-samples oven-dried at 221 80 °C for 48 h to a constant weight. Litterfall production (t ha⁻¹) was calculated as a 222 dry weight taking account of the litter trap area and the moisture content of the fresh 223 litterfall. 224

225

226 **2.4 Needle sampling and nutrient analysis**

227 Chinese fir needles were also sampled from the same sample trees from which canopy litter was collected and at the same times in January, April, July, and October 2015. 228 Chinese fir leaves form a new node annually, therefore, the apex needles in the first 229 230 node are <1-year old, while needles further along the shoot are perennial. At least 10 shoots were collected from each sample tree from different aspects and heights in the 231 canopy. Five categories of needles [<1-year-old, 1-year-old, perennial, senesced 232 (brown), and dead (yellow)] were collected separately from each shoot (Fig. S4). 233 Senesced needles are brown and not completely dry; dead needles are completely dry 234 and yellow, without signs of deterioration (Wright and Westoby, 2003). All the 235 236 needles in the same category were mixed to form a composite sample. All the samples were first oven-dried at 105 °C for 15 min to deactivate enzymes and then at 80 °C for 237

at least 48 h to constant weight (Zou et al., 2015). Dried samples were ground 238 uniformly using a grinder, sieved to <1.0 mm and then stored in brown glass bottles 239 before nutrient concentration determination. Total N was determined using 240 combustion in an ELEMENTAT Vario EL III analyzer (Germany) at Fujian 241 Agriculture and Forestry University (Zhou et al., 2015). High purity glutamic acid 242 containing 9.52% N was used as a standard to calibrate the instrument. It was also 243 analyzed every 10 samples as a quality control check, with instrument performance 244 deemed acceptable if the measured value was $\pm 10\%$ that of the standard. For each 245 246 plant sample, a 0.1 g subsample was digested using HNO₃-H₂SO₄ solution (volume ratio 5:1). Total P in the digests was measured using the molybdenum-antimony 247 colorimetric method with a 723A spectrophotometer (AOXI Instruments, Shanghai, 248 249 China) (He et al., 2003). The absorbance of the solution was determined in a 1-cm cell at 700 nm, using ultrapure water as the blank. Total K, Ca, and Mg were determined 250 using atomic absorption spectrophotometry (AAS) (Puxi Instruments, Beijing, China) 251 252 (Institute of Soil Academia Sinica, 1978).

253

254 **2.5 Resorption efficiency**

Nutrient resorption efficiency (RE), the proportional withdrawal of nutrients during
leaf senescence, was calculated using Equation 1 (Han et al., 2013; Zhou et al., 2016a):

257 (1)
$$RE = (1 - \frac{Nutrient_{yel}}{Nutrient_{gre}} \times MLCF) \times 100$$

where *Nutrient_{yel}* and *Nutrient_{gre}* are nutrient concentrations (N, P, K, Ca, and Mg) of

259 yellow and <1-year-old needles, respectively. Considering that leaf mass decreases

260	during senescence, a mass loss correction factor (MLCF) was used to compensate for
261	the underestimation of RE. Here the MLCF for needles was 0.65, determined in a
262	previous study of Chinese fir trees conducted nearby at different locations within
263	Xinkou National Forest Farm (Zhou et al., 2016a).
264	
265	2.6 Statistical analysis
266	The data are presented as means \pm SE. A three-way analysis of variance (ANOVA)
267	was used to test the effects of canopy strata (height sampled above the ground), stand
268	age, season, and their interactions on the dry mass of canopy litter. Fixed factors were
269	canopy strata, stand stage and season. A one-way ANOVA was used to compare the
270	dry mass differences among seasons in the same vertical strata or among vertical
271	strata in the same season. Means were compared using LSD post-hoc tests ($P < 0.05$).
272	All statistical analyses were performed using SPSS 17.0 software package for
273	Windows (SPSS Inc., Chicago, IL, USA).
274	
275	3. Results

3.1 Spatiotemporal distribution of canopy litter and annual litterfall production 276

Canopy strata, stand age, and season all had significant effects (P < 0.001) on canopy 277 litter mass (Table 2). The interactions between stand age and canopy height, and 278 canopy strata and season were also significant (P < 0.001 and P < 0.01, respectively). 279 There were no significant interaction effects between all three factors. The mean 280 masses of the dead leaves and dead branches in the canopy of Chinese fir trees were 281

282	14.6, 14.2, and 17.4 t ha ⁻¹ for the young, middle-aged and mature stages, respectively
283	(Fig. 1). The height of dead branches and leaves in the tree canopy increased with
284	stand age: 0-4 m in the young stand, 3-8 m in the middle-aged stand, and 4-10 m in
285	the mature stand (Fig. 2). In the young stand, most canopy litter mass was at $0-3$ m
286	height above the ground, accounting for 78.6% of total canopy litter mass. For the
287	middle-aged stand, the majority of the canopy litter mass was at 4-8 m height,
288	accounting for 75.1% of total canopy litter production. In the mature stand, most
289	canopy litter was at 6-10 m height, accounting for 77.7% of the total canopy litter
290	mass.

292Table 2. Three-way analysis of vertical strata (df = 6), season (df = 3), stand stage (df = 2)293and their interaction on canopy litter mass of attached dead branches and leaves of Chinese294fir plantations at different development stages. P values in bold indicate significant factors295and their interactions.

Values	df	F	Р
Stand stage	2	227.643	<0.001
Season	3	12.490	<0.001
Vertical strata	6	1353.813	<0.001
Stand stage × Season	6	0.750	0.610
Stand stage × Vertical strata	1	1584.753	<0.001
Season × Vertical strata	18	2.211	0.003
Stand stage \times Season \times Vertical strata	3	0.643	0.588



Fig. 1. Vertical and seasonal distribution of dry mass of attached dead branches and leaves in the canopy among young, middle and mature stage *C. lanceolata* plantations in southern China (t ha⁻¹). Horizontal bars are means \pm SE (n=3). Different uppercase letters indicate significant differences between seasons for the same vertical strata (*P*<0.05). Different lowercase letters indicate significant differences between vertical strata in the same season (*P*<0.05).

The total mass of canopy litter was significantly higher during the winter (January): 318 15.5, 15.3, and 18.2 for the young, middle, and mature stage stands, respectively (Fig. 319 1). The lowest total mass of canopy litter was 13.6 and 13.5 t ha⁻¹ during autumn 320 (October) for the young and middle stage stands, respectively, and 16.5 t ha⁻¹ during 321 the summer (July) for the mature stand. The seasonal dynamics of litterfall production 322 were in the order autumn > spring > summer > winter for all stand stages (Fig. 2). The 323 total annual litterfall production was 2.44, 3.75, and 4.34 t ha⁻¹ for young, middle-aged, 324 and mature stands, respectively. The canopy litter mass comprised 85.7%, 79.1%, and 325 80.0% of the total mass, including the annual litterfall production, in the three 326 different stages of Chinese fir plantations. 327



Fig. 2. Seasonal dynamics of litterfall production (t ha⁻¹) of Chinese fir plantations at different developmental stages. Vertical bars are means \pm SE (n = 3). Different uppercase letters indicate significant differences between seasons for the same stage (*P*<0.05). Different lowercase letters indicate significant differences between stages for the same season (*P*<0.05).

341 **3.2** Variation of leaf nutrient concentrations during senescence and shedding

Nutrient concentration (N, P, K, Ca and Mg) varied significantly with stand stage, leaf 342 level, and season (Table 3, P < 0.01). There were also significant interaction effects 343 for all nutrients studied between stand stage and needle categories, and needle 344 categories and seasons (P < 0.01). N, P, K and Mg concentrations showed a 345 significant downward trend with leaf senescence and shedding, while Ca 346 concentrations displayed the opposite trend (Fig. 3). Concentrations of N, P, K, and 347 Mg were highest in the <1-year-old needles at 12.2–16.9 g kg⁻¹, 0.93–1.13 mg g⁻¹, 348 6.55–6.83 g kg⁻¹, and 1.83–2.54 g kg⁻¹, respectively. Concentrations of N, P, K, and 349 Mg were lowest in the litterfall needles at 6.96–9.96 g kg⁻¹, 0.44–0.53 g kg⁻¹, 350 0.74–1.24 g kg⁻¹, and 1.30–1.89 g kg⁻¹, respectively. In contrast, the highest Ca 351 concentrations were in yellow (dead) needles (5.62–8.32 g kg⁻¹), and the lowest were 352 in the <1-year-old needles (2.24–3.68 mg g^{-1}). All the needles from different 353 categories and stand stages showed similar seasonal patterns. N and P concentrations 354 were lowest in January, increased from April to July and decreased again. The 355 seasonal variation in Ca concentrations showed the opposite trend, being highest in 356 January and lower in July. K and Mg concentrations were highest in October and 357 lowest in April. 358

359



Fig. 3. Nutrient concentrations (N, P, K Ca and Mg) in the different categories of needles sampled in the canopy (<1-year-old needles, 1-year-old needles, perennial needles, senesced needles, yellow (dead) needles) and in litterfall needles at various seasons among young, middle and mature stage *C. lanceolata* plantations in southern China. Note different y-axis scales between figures. Different uppercase letters indicate significant differences between stages for the same needle category in each season (P<0.05). Different lowercase letters indicate significant differences between needle categories for the same stage in each season (P<0.05).

389 3.3 Nutrient resorption efficiency of needles during senescence in the canopy

During leaf senescence, N, P, K, and Mg were strongly resorbed, with the resorption 390 efficiency (RE) varying significantly among stand stage and seasons (apart from for 391 season for Mg). There were also significant interactions between the two variables 392 (except Mg, Table 4). Overall, across stand stages, the REs of different nutrients were 393 in the order K > P > N > Mg, at 85.0–90.2%, 64.0–68.9%, 53.8–58.9%, and 394 46.5-56.6%, respectively (Fig. 4). In contrast, Ca concentrations in needles were 395 enriched during leaf senescence in all stand stages. N and Mg resorption efficiencies 396 were highest in the young stand (58.4% and 52.8%, respectively), followed by the 397 mature (46.7% and 50.1%, respectively), and the middle-aged stand (45.1% and 398 46.5%, respectively). P resorption efficiency was in the order, young stand (70.0%) >399 middle-aged stand (67.8%) > mature stand (66.2%). K had the largest resorption 400 efficiency at 86.5-88.5% in all three stands. 401

402

It is noteworthy that nutrient RE varied significantly between seasons. N resorption efficiency was highest in April for young and middle-aged stands, and in January for the mature stand. P resorption efficiency was highest in April for young and mature stands, and in July for the middle-aged stand. K resorption efficiency was highest in April and lowest in October for each of the three stands. Mg resorption efficiency showed a similar trend in all stand ages of April > October > January > July. 409 Table 3. Three-way analysis of stand stage (df = 2), needle categories (df = 5), season (df = 3) and their interaction on nutrient concentration (N, P, K, Ca, Mg)

410 of di	fferent needle categories of the second s	Chinese fir plantations at	t different development stages. I	P values in bold indicate significant factors	s and their interactions.
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	N concentr	ation	P concentra	ation K concentration		ation	Ca concentration		Mg concentration	
Values	F	Р	F	Р	F	Р	F	Р	F	Р
Stand stage	6143.548	< 0.01	1.106	0.334	1946.781	< 0.01	200.388	< 0.01	519.069	< 0.01
Needle categories	3440.240	< 0.01	1375.376	< 0.01	48541.201	< 0.01	4350.938	< 0.01	460.773	< 0.01
Season	5720.779	< 0.01	21.521	< 0.01	692.002	< 0.01	954.066	< 0.01	852.737	< 0.01
Stand stage ×Needle categories	294.374	< 0.01	6.185	< 0.01	248.380	< 0.01	323.598	< 0.01	10.893	< 0.01
Stand stage × Season	656.548	< 0.01	2.134	0.053	17.700	< 0.01	64.847	< 0.01	114.785	< 0.01
Needle categories × Season	7.287	< 0.01	3.524	< 0.01	3.399	< 0.01	5.503	< 0.01	2.499	< 0.01
Stand stage \times Needle categories \times Season	5.622	< 0.01	1.062	0.392	2.993	< 0.01	4.717	< 0.01	1.928	< 0.01

411

Table 4. Two-way analysis of stand stage (df = 2), season (df = 3) and their interaction on nutrient resorption efficiency (N, P, K, Ca, Mg) of Chinese fir
 plantations at different development stages. *P* values in bold indicate significant factors and their interactions.

Values	N resorption efficiency		P resorption efficiency		K resorption efficiency		Ca resorption efficiency		Mg resorption efficiency	
	F	Р	F	Р	F	Р	F	Р	F	Р
Stand stage	8912.653	< 0.001	36.701	< 0.001	343.190	< 0.001	2902.120	< 0.001	9.770	0.001
Season	133.390	< 0.001	4.505	0.012	549.263	< 0.001	420.285	< 0.001	2.195	0.115
Stand stage \times Season	48.938	< 0.001	3.024	0.024	21.726	< 0.001	70.357	< 0.001	1.305	0.293



litter of young, middle and mature stage C. lanceolata plantations in southern China. 436 Columns are means \pm SE (n=3). Different uppercase letters indicate significant differences 437

438 between stages in the same season (P < 0.05). Different lowercase letters indicate significant

439 differences between seasons for the same stage (P < 0.05).

441 **4. Discussion**

442 **4.1 Spatiotemporal distribution of canopy litter**

443 Our results verified the first hypothesis that the attached dead leaves and branches in the canopy of Chinese fir plantations at different development stages varied 444 significantly among vertical strata and seasons (Table 2). Canopy litter of Chinese fir 445 in the study stands was 14.2–17.4 t ha⁻¹, accounting for 79.1%–85.7% of both canopy 446 biomass and annual litterfall production. The canopy litter mass of the mature stand 447 was much lower than reported for other tree species in previous studies, e.g., 448 33-year-old *Cryptomeria japonica* stand (34 t ha⁻¹), but that of the young stand was 449 much higher than in a 16-year-old *C. japonica* stand (3.6 t ha⁻¹) (Miyaura and Hozumi, 450 1989; Yoshida and Hijii, 2006). Compared to our results, much lower canopy litter 451 452 masses have been reported in other studies of Chinese fir. For example, a 13-year-old Chinese fir plantation in Jiangxi Province, China, had a canopy litter mass of 4.12 t 453 ha⁻¹ (Zhang and Sheng, 2001), whilst Zheng (1997) reported canopy litter masses of 454 ~ 8.0 t ha⁻¹, 3.5 t ha⁻¹ and 5.5 t ha⁻¹ in 10-, 15-, and 19-year-old Chinese fir stands, 455 respectively, and it increased as soil quality improved. Liao et al. (1996) found that 456 Chinese fir retained large amounts of dead branches and leaves for up to 10 years, 457 with a small amount of litterfall production (canopy litter: 7.52 t ha⁻¹; litterfall 458 production: 0.475 t ha⁻¹). Litterfall production was higher in >15-year-old stands, 459 whilst the canopy litter biomass was only 1.69 and 2.62 t ha⁻¹ for 15- and 19-year-old 460 stands, respectively (Liao et al., 1996). The large variation in estimates of canopy 461 litter mass reported in the literature may be attributed to differing stand stages, site 462

climate, and sample size and methods. In other studies, a smaller number of trees were
sampled and lower forest productivity (and thus litterfall production) is expected due
to cooler and drier conditions compared to our site.

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The canopy dead branches and leaves mainly occupied heights of 0-4 m, 3-8 m, and 467 4-10 m in the canopy of young, middle-aged, and mature stands, respectively, and the 468 canopy litter occurred predominantly in the lower parts of the canopy in the young 469 stand. Several possible factors may be responsible for the large mass of canopy litter 470 471 and vertical distribution pattern of dead branches and leaves. First, the lower foliage is shaded by upper branches and leaves of adjacent trees after canopy closure 472 (>5-year-old stands) (Zhang and Sheng, 2001). Most branches and leaves die from 473 474 lack of sunshine instead of physiological causes. The dead branches and leaves remain attached to the trunk for several years with a slow decomposition rate, if not detached 475 by wind or snow. Second, the canopy litter may continue to accumulate living leaves 476 because Chinese fir leaf longevity is ~4-5 years (Ma et al., 2007) and the fast-growth 477 and stemwood formation periods of Chinese fir are 5-12 years and 13-19-years, 478 respectively (Tian et al., 2002). Third, Zhang and Sheng (2001) reported that dead 479 branches and leaves in the lower canopy may decay more rapidly due to greater 480 moisture and abundant microbial communities than in the upper canopy, which could 481 lead to an increase in the vertical height of occurrence of canopy litter as 482 developmental stage increases for Chinese fir. 483

Our results showed that the seasonal mass dynamics of canopy litter were inverse to 485 those of litterfall production (Figs. 2 and 3). The death of branches and leaves 486 occurred all year, but the canopy litter mass of Chinese fir peaked in winter 487 (15.3–18.2 t ha⁻¹), while litterfall production was lowest in winter. Furthermore, the 488 mass of dead branches and leaves in the canopy was much larger relative to that of 489 litterfall production, accounting for 79.1%-85.7% of total litter mass including 490 litterfall production, which indicated that the death rate (litter input to canopy) was 491 much higher than the litterfall rate (litter output from the canopy to the soil). 492 Assuming that the litterfall rate equals the canopy litter death rate, the turnover rate of 493 the canopy litter was 0.17, 0.26, and 0.25 t year⁻¹ for the young, middle-aged, and 494 mature stands of Chinese fir plantations studied, respectively. The actual turnover 495 496 rates of canopy litter are likely to be much lower. Yoshida and Hijii (2006) reported the average turnover rate of canopy litter was 0.17 year⁻¹ (0.21 t year⁻¹ for canopy leaf 497 litter and 0.09 t year⁻¹ for canopy branch litter) for *C. japonica* forests. Assuming the 498 rates of litterfall input to soil and the output to the deeper soil layers are 499 approximately equal, litter accumulating on the soil surface would be replaced after 500 about 0.63, 0.60, and 0.80 years in the young, middle-aged, and mature stands, 501 respectively (Fig. 5). 502

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Fig. 5. Estimates of litter flow in young, middle and mature *C. lanceolata* plantations based
on results from this study and other studies. The biomass of living branches and leaves and the
mass of dead branches and leaves in the canopy is expressed as t dry wt ha⁻¹, and the amount of
transfer (litterfall) is expressed as t dry wt ha⁻¹ a⁻¹. ^a indicates data from Ma et al. (2002); ^b
indicates data from Zhou et al. (2012).

4.2 Variation and seasonality in nutrient concentrations of needles in the canopy 524 We partly confirmed the second hypothesis that nutrient (N, P, K, Mg) concentrations 525 of needles decreased from growth to senescence and varied seasonally. In contrast, Ca 526 concentrations increased, in agreement with Viers et al. (2013), which may be 527 explained by the relative immobility of Ca involved in the structure and physiological 528 processes of cell membranes (Fircks et al., 2001). We found that nutrient 529 concentration was significantly affected not only by the needle type but also by season. 530 The highest leaf N and P concentrations in different categories of Chinese fir needles 531 532 were measured in the summer (July) samples. A similar tendency has been observed in other conifers, such as Pinus sylvestris (Regina and Tarazona, 2001). In the 533 growing season (summer), intense mitotic activity companied by cellular growth 534 535 requires more nutrients for protein synthesis, in particular N and P (Arneth et al., 1996). The highest Ca and K concentrations in Chinese fir needles occurred in 536 January and October, consistent with Juniperus phoenicea species (Kutbay et al. 537 2005). Due to the relative immobility of Ca, as noted already, it usually accumulates 538 in plant tissues during the growing season, with peak concentrations occurring at the 539 end of the growing season (Hepler, 2005). 540

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In our study, fresh foliar (<1-year-old) nutrient concentrations (mean values of the three development stages are N: 14.0 g kg⁻¹; P: 0.98 g kg⁻¹; K: 6.55 g kg⁻¹; Ca: 2.86 g kg⁻¹; Mg: 2.08 g kg⁻¹) were lower than those in broadleaved species (such as *S. superb*, *Castanopsis sclerophylla schott, Eucalyptus grandis*), implying that nutrient cycling is 2010). The fresh foliar N and P concentrations in our study were lower than the critical nutrient concentrations in a global study (N: 18.6 g kg⁻¹ and P: 1.21 g kg⁻¹) (Han et al. 2005), which further indicated that the Chinese fir plantations studied were both N- and P-limited. Lower nutrient concentrations and higher nutrient use efficiency in Chinese fir plantations might be important adaptation mechanisms for nutrient conservation in nutrient deficient soil conditions (Diehl et al., 2003).

more conserved in coniferous species (Huang et al., 2007; Li et al., 2009; Wang et al.,

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4.3 Nutrient resorption efficiency of needles in the canopy

We partly verified the third hypothesis that nutrient RE varied among stand stages and 555 seasons (Table 4). Many previous studies have focused on N and P resorption, 556 557 overlooking other nutrients that are also important for forest productivity. In our study, leaf N and P nutrient RE (50.1% and 68.0%, respectively) in the canopy of Chinese fir 558 during senescence were of a similar magnitude to the mean values reported for 559 planted forests of 59.0% and 60.2% (Jiang et al., 2019), but slightly lower than in our 560 previous study in the same area in 10-, 22-, and 34-year-old Chinese fir (57% and 561 75%, respectively). 562

563

The average RE of K and Mg was 85.0–90.2% and 46.5–56.6% respectively, and Ca was accumulated in the senescing leaves. The higher RE values for N, P and K may be explained by their large requirement as well as their mobility, in alignment with our third hypothesis. Ca is stable and not easily resorbed, as demonstrated by many

studies (e.g., Viers et al., 2013). In our previous study, we found nutrient REs were 568 most affected by leaf nutrient status and stoichiometric control rather than soil 569 nutrient status (Zhou et al., 2016a). Leaf shedding patterns may also affect nutrient 570 resorption from senescing leaves. Milla et al. (2005) reported that species that shed 571 leaves gradually throughout the year have significantly lower K resorption efficiency 572 than species that shed leaves over a short time period, such as Mediterranean 573 perennials. Thus, many factors influence the nutrient resorption process from 574 senescing leaves to other plant tissues. Which processes are dominant will likely be 575 influenced by the relative status of leaf and soil nutrient conditions in the ecosystem. 576

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In our previous study, we estimated the mass of senescent branches and senescent 578 579 leaves and determined their nutrient concentrations (Zhou et al., 2016b). Based on these, in our study, the nutrient contents of senescent branches and leaves in the 580 canopy for N, P, K, and Mg were 102-125, 5.00-6.39, 19.2-26.9, and 18.2-25.8 kg 581 ha⁻¹ a⁻¹, respectively, much higher than our previous estimates of total annual nutrient 582 return from Chinese fir plantations (young, mature and over-mature stage) to the soil 583 layer (N: 39.3–62.0, P: 1.30–1.63, K: 16.0–22.4 kg ha⁻¹ a⁻¹) (Zhou et al., 2015). The 584 decline in plantation productivity in successive plantings would be greatly eased if 585 these nutrients could be absorbed directly from senescent organs before shedding 586 (Farooq et al., 2019). From the N, P, K, and Mg REs estimated in the present study for 587 the different development stages, nutrient resorption is 68.8-77.0, 4.54-6.30, 588 14.9–20.2, and 10.4–13.3 kg ha⁻¹ a⁻¹, respectively. We previously reported that the 589

nutrient turnover rate (storage/return) for Chinese fir plantations (7.44, 10.4, and 12.0 years for young, mature and overmature stage, respectively) increases as stand development stage increases (Zhou et al., 2016b). Long nutrient turnover time, relatively small litterfall production and low litterfall quality (higher C:N ratio and low P content) could slow the rate of nutrient cycling within Chinese fir plantations and soil fertility recovery, and thus reduce forest productivity (Zhou et al., 2015, 2016 a,b).

597

There are two major reasons why nutrient turnover rate is slow for Chinese fir. Firstly, 598 litterfall production of Chinese fir is much lower than in many broad-leaved evergreen 599 species, such as *Castanopsis kawakamii* (11.0 t ha⁻¹ a⁻¹), *Altingia gracilipes* (6.4 t ha⁻¹ 600 a^{-1}), and *Tsoongiodendron odorum* (6.3 t $ha^{-1}a^{-1}$). Secondly, the decomposition rate of 601 Chinese fir litter is slow, due to the poor litter quality which has a high C:N ratio (62.5) 602 and lignin content (Wang et al. 2007, 2009), restricting nutrient recovery. Several 603 604 studies have reported that the decomposition rate of Chinese fir litterfall on the forest floor is 3–5 times that in the canopy and suggested that knocking the canopy litter off 605 onto the forest soil would be an effective measure to maintain plantation productivity 606 (Zhang et al., 2001; Gao et al., 2015). However, in our study, the dead attached leaves 607 and branches were characterized by high nutrient resorption capacity during needle 608 senescence and death, which could greatly improve nutrient utilization by intra-tree 609 cycling and decrease the dependence on soil nutrients. This is an important nutrient 610 conservation strategy in response to the changes in nutrient availability. 611

613 4.4 Implications for nutrient cycling in Chinese fir

614 This study examined the spatiotemporal dry mass of canopy litter and leaf nutrient resorption prior to death in Chinese fir plantations at different development stages. 615 Large masses of dead leaves and branches were measured in the canopy of 14.6, 14.2, 616 and 17.4 t ha⁻¹, comprising 85.7%, 79.1%, and 80.1% of the total canopy mass of 617 young, middle-aged, and mature stands, respectively, including litterfall production. 618 Apart from for Ca, we found that the extensive canopy litter contained the highest 619 620 concentration of nutrients that were retranslocated to other plant tissues before needles senesced. The N, P, K, and Mg resorption efficiencies across development stages were 621 high, at 53.8-58.9%, 64.0-68.9%, 85.0-90.2%, and 46.5-56.6%, respectively. 622 623 Canopy pruning is an important silvicultural practice for many tree plantations. However, for Chinese fir, the lower litterfall production and slower litter 624 decomposition rate results in less nutrient input over a longer time, which delays soil 625 626 fertility recovery. Our results showed that Chinese fir is characterized by large amounts of dead attached leaves and branches remaining in the canopy for many years, 627 and by high nutrient resorption efficiencies, which decrease the dependence on soil 628 nutrients, especially in nutrient-deficient soil. The characteristics of Chinese fir 629 canopy litter could be considered to have evolved as a nutrient conservation strategy 630 to maximize within tree nutrient recycling. Therefore, maintaining dead leaves and 631 branches in the canopy might be an important influence on nutrient cycling in 632 subtropical plantation forests, and should inform pruning and management decisions. 633

635 **5.** Conclusions

Large amounts of dead branches and leaves accumulated in the canopy of Chinese fir 636 plantations at young, middle-aged and mature development stages. The canopy litter 637 occupied different canopy strata, and the vertical height increased with stand age. The 638 seasonal dynamics of canopy litter mass were almost inverse to those of litterfall 639 production. Macronutrient (N, P, K Ca, and Mg) concentrations varied greatly with 640 leaf growth, senescence, and abscission among the three development stages. N, P, K, 641 and Mg were resorbed during leaf senescence with resorption efficiency of 642 53.8-58.9%, 64.0-68.9%, 85.0-90.2%, and 46.5-56.6%, respectively. The nutrient 643 contribution of resorption of these four nutrients is estimated at 68.8–77.0, 4.54–6.30, 644 14.9–20.2, and 10.4–13.3 kg ha⁻¹ a⁻¹, respectively. This implies the retention of dead 645 branches and leaves in the canopy may have evolved in Chinese fir as a strategy to 646 conserve nutrients in response to nutrient-deficient soil. 647

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656 Declaration of competing interests

The authors declare that they have no known competing financial interests or personalrelationships that could have appeared to influence the work reported in this paper.

659

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671 **References**

Aerts, R., 1996. Nutrient resorption from senescing leaves of perennials: are there general patterns?
J. Ecol. 84, 597–608.

Arneth, A., Kelliher, F.M., Bauer, G., Hollinger, D.Y., Byers, J.N., Hunt, J.E., McSeveny, T.M.,

2675 Ziegler, W., Vygodskay, N.N., Milukova, I., Sogachov, A., Varlagin, A., Schulze, E.-D., 1996.

- 676 Environmental regulation of xylem sap flow and total conductance of *Larix gmelinii* trees in
- eastern Siberia. Tree Physiol. 16, 247–255.

- Brant, A.N., Chen, H.Y.H., 2015. Patterns and mechanisms of nutrient resorption in plants. Crit
 Rev Plant Sci. 34, 471–486.
- 680 Cleveland, C.C., Houlton, B.Z., Smith, W.K., Marklein, A.R., Reed, S.C., Parton, W., Del Grosso,
- 681 S., Running, S.W., 2013. Patterns of new versus recycled primary production in the terrestrial
- biosphere. PNAS. 110, 12733–12737.
- Diehl, P., Mazzarino, M.J., Funes, F., Fontenla, F., Gobbi, M., Ferrari, J., 2003. Nutrient
- 684 conservation strategies in native Andean-Patagonian forests. J Veg Sci. 14, 63–70.
- Escudero, A., del Arco. J.M., 1987. Ecological significance of the phenology of leaf abscission.
- 686 Oikos. 49(1), 11–14.
- 687 Escudero, A., del Arco, J.M., Sanz, I.C., Ayala, J., 1992. Effects of leaf longevity and
- retranslocation efficiency on the retention time of nutrients in the leaf biomass of different
- 689 woody species. Oecologia. 90 (1), 80–87.
- 690 Farooq, T.H., Yan, W., Rashid, M.H.U., Tigabu, M., Gilant, M.M, Zou, X.H., Wu, P.F., 2019.
- 691 Chinese fir (*Cunninghamia lanceolata*) a green gold of China with continues decline in its
- productivity over the successive rotations: a review. App Ecol Env Res. 17(5), 11055-11067.
- 693 Fircks, Y.O., Ericsson, T., Sennerby-Forsse, L., 2001. Seasonal variation of macronutrients in
- 694 leaves, stems and roots of *Salix dasyclados* Wimm. grown at two nutrient levels. Biomass
- 695 Bioenerg. 21, 321–34.
- 696 Fonte, S.J., Schowalter, T.D., 2004. Chapter 21 Decomposition in Forest Canopies. In: Lowman,
- 697 M. and Rinker, H.B. (Eds.), Forest Canopies (2nd ed.), Elsevier Academic Press, Oxford,

- 698 pp .413–422.
- 699 Gao, S.L., He, Z.M., Huang, Z.Q., Lin, S.Z., Liu, Z.M., 2015. Decomposition, carbon and nitrogen

stable isotope and chemical composition of dead leaves clinging in a Chinese fir

- 701 (*Cunninghamia lanceolata*) plantation. Chin J Ecol. 34(9), 2457–2463. (in Chinese)
- Gerrish, G., 1990. Relating carbon allocation patterns to tree senescence in *Metrosideros* forests.
 Ecology, 71: 1176–1184.
- Han, W.X., Fang, J.Y., Guo, D.L., Zhang, Y., 2005. Leaf nitrogen and phosphorus stoichiometry

across 753 terrestrial plant species in China. New Phytol. 168, 377–385.

Han, W.X., Tang, L.Y., Chen, Y.H., Fang, J.Y., 2013. Relationship between the relative limitation

and resorption efficiency of nitrogen vs phosphorus in woody plants. PLoS ONE. 8:e83366

- 708 He, Y., Liao, H., Yan, X.L., 2003. Localized supply of phosphorus induces root morphological and
- architectural changes of rice in split and stratified soil cultures. Plant Soil. 248(1–2),
 247–256.
- Hepler, P.K., 2005. Calcium: a central regulator of plant growth and development. Plant Cell. 17,
 2142–2155.
- Huang, J.J., Wang, X. H., Yan, E.R., 2007. Leaf nutrient concentration, nutrient resorption and
- 714 litter decomposition in an evergreen broad-leaved forest in eastern China. For Ecol Manag.
 715 239, 150–158.
- 716 Institute of Soil Academia Sinica, 1978. Analysis of Soil Physics and Chemistry. Sciences and
- 717 Technology of Shanghai Press, Shanghai, pp. 274–276. (in Chinese)

- 718 Ishii, H., Wilson, M.E., 2001. Crown structure of old-growth Douglas fir in the western Cascade
- Range, Washington. Can J Forest Res. 39, 1250–1261.
- 720 Ishii, H., Kadotani, T., 2006. Biomass and dynamics of attached dead branches in the canopy of
- 450-year-old Douglas-fir trees. Can J Forest Res. 36(2), 378–389.
- Jiang, D., Geng, Q., Li, Q., Luo, Y., Vogel, J., Shi, Z., Ruan, H., Xu, X., 2019. Nitrogen and
- phosphorus resorption in planted forests worldwide. Forests 10, 201.
- 724 Kobe, R.K., Lepczyk, C.A., Iyer, M., 2005. Resorption efficiency decreases with increasing green
- leaf nutrients in a global data set. Ecology. 86, 2780–2792.
- Kutbay, H.G., OK, T., Bilgin, A., Yalçin, E., 2005. Seasonal nutrient levels and foliar resorption in
 Juniperus phoenicea. Belg Journ Bot 138(1), 67–75.
- 728 Laiho, R., Prescott, C.E. 1999. The contribution of coarse woody debris to carbon, nitrogen and
- phosphorus cycles in three Rocky Mountain coniferous forests. Can J For Res. 29(10),
 1592–1603.
- 731 Li, Y.X., Cha, Z.Z., Luo, W., Lin, Z.M., Bei, M.R., 2009. Dynamics and transfer of nutrients in
- the seedling leaves of three Eucalyptus varieties. Scientia Silvae Sinicae.45 (1), 152–157. (in
- 733 Chinese)
- 734 Liao, Z.H., Fan, S.H., Yu, X.T., 1996. Study on the biomass among different-aged Chinese fir
- plantations at different site conditions. IV- Biomass of dead branches and leaves in the canopy.
- For Res. 9 (special issue), 96–99. (in Chinese)
- 737 Liu, C.C., Liu Y.G., Guo, K., Wang S.J., Yang, Y., 2014. Concentrations and resorption patterns of

- 738 13 nutrients in different plant functional types in the karst region of south-western China. Ann
 739 Bot. 5, 873–885.
- Liu, Z.M., 2014. Research on the nutrient dynamics and internal absorption of *Cunninghamia lanceolata* foliar [D]. Thesis. (in Chinese)
- 742 Lü, X.T., Reed, S.C, Yu, Q., Han, X.G., 2016. Nutrient resorption helps drive intra-specific
- coupling of foliar nitrogen and phosphorus under nutrient-enriched conditions. Plant Soil.
 2016, 398(1-2), 111–120.
- 745 Ma, X.Q., Heal, K.V., Liu, A.Q., Jarvis, P.G., 2007. Nutrient cycling and distribution in

different-aged plantations of Chinese fir in southern China. For Ecol Manag. 243, 61–74.

747 Ma, X.Q., Liu, C.J., Hannu, I., Westman, C.J., Liu, A.Q., 2002. Biomass, litterfall and the nutrient

fluxes in Chinese fir stands of different age in subtropical China. J Forestry Res. 13(3),

749 165–170.

- 750 Maillard, A., Diquélou, S., Billard, V., Laîné, P., Garnica, M., Prudent, M., Garcia-Mina, J-M.,
- 751 Yvin, J. C., Ourry, A., 2015. Leaf mineral nutrient remobilization during leaf senescence and
- modulation by nutrient deficiency. Front Plant Sci. 6, 317.doi: 10.3389/fpls.2015.00317
- 753 Marschner, H., 2011. Marschner's Mineral Nutrition of Higher Plants (3rd ed.). Elsevier Academic
- 754 Press, Oxford, pp. 50–66.
- 755 Milla, R., Castro-Díez, P., Maestro-Martínez, M., Montserrat-Martí, G., 2005. Does the
- 756 gradualness of leaf shedding govern nutrient resorption from senescing leaves in
- 757 Mediterranean woody plants? Plant Soil. 278(1-2), 303–313.

- 758 Miyaura, T., Hozumi, K., 1989. Measurement of litterfall in a sugi (Cryptomeria japonica)
- plantation by the cloth-trap method. J Jpn For Soc. 71, 69–73.
- 760 Nadkarni, N.M., Matelson, T.J., 1991.Litter dynamics within the canopy of a neotropical cloud
- forest, Monteverde, Costa Rica. Ecology. 72, 2071–2082.
- 762 Niinemets, Ü., Tamm, Ü., 2005. Species differences in timing of leaf fall and foliage chemistry
- modify nutrient resorption efficiency in deciduous temperate forest stands. Tree Physiol. 25
 (8), 1001–1014.
- Osaki, M., Shinano, T., 2001. Plant growth based on interrelation between carbon and nitrogen
 translocation from leaves. Photosynthetica. 39, 197–203.
- Prescott, C.E., 2002. The influence of the forest canopy on nutrient cycling. Tree Physiol. 22,
 1193–1200.
- 769 Regina, I.S., Tarazona, T., 2001. Nutrient cycling in a natural beech forest and adjacent planted
 770 pine in northern Spain. Forestry 74, 11–28.
- 771 See, C.R., Yanai, R.D., Fisk, M.C., Vadeboncoeur, M.A., Quintero, B.A., Fahey, T.J., 2015. Soil
- nitrogen affects phosphorus recycling: foliar resorption and plant-soil feedbacks in a northern
- 773 hardwood forest. Ecol. 96, 2488–2498.
- Seibold, S., Hagge, J., Müller, J., Grppe, A., Brandl, R., Bässler, C., Thorn, S., 2018. Experiments
- with dead wood reveal the importance of dead branches in the canopy for saproxylic beetle
- conservation. For Ecol Manag. 409, 564–570.
- 777 Sheng, W.T., Fan, S.H., 2002. Impact of growth and development characters of Chinese fir and its

plantation on the long-term site productivity. For Rese, 15 (6), 629–636. (in Chinese)

779	State Forestry Administration of the People's Republic of China, 2014. Forest resources in China:
780	The 8th National Forest Inventory http://211.167.243.162:8085/8/book/jiankuang/index.html
781	Tang, L.Y., Han, W.X., Chen, Y.H., Fang, J.Y., 2013. Resorption proficiency and efficiency of leaf
782	nutrients in woody plants in eastern China. J Plant Ecol, 6, 408–417.
783	Tian, D., Xiang, W., Yan, W., Kang, W., 2002. Effect of successive rotation on productivity and
784	biomass of Chinese fir plantation at fast growing stage. Scientia Silvae Sinicae. 38(4), 14–18.
785	(in Chinese)
786	Tsujii, Y., Onoda, Y., Kitayama, K 2017. Phosphorus and nitrogen resorption from different
787	chemical fractions in senescing leaves of tropical tree species on Mount Kinabalu, Borneo.
788	Oecologia.185(2), 1–10.
789	Tully, K.E., Wood, T.E., Schwantes, A.M., Lawrence, D., 2013. Soil nutrient availability and
790	reproductive effort drive patterns in nutrient resorption in Pentaclethra macroloba. Ecol. 94,
791	930–940.
792	Van Heerwaarden, L. M., Toet, S., Aerts, R., 2003a. Nitrogen and phosphorus resorption
793	efficiency and proficiency in six sub-arctic bog species after 4 years of nitrogen fertilization. J
794	Ecol. 91, 1060-1070.
795	Van Heerwaarden, L.M., Toet, S., Aerts, R., 2003b. Current measures of nutrient resorption
796	efficiency lead to a substantial underestimation of real resorption efficiency: facts and

797 solutions. Oikos. 101(3), 664-669.

798	Vergutz, L., Manzoni, S. Porporato. A., Novais, R.F., Jackson, R.B., 2012. Global resorption
799	efficiencies and concentrations of carbon and nutrients in leaves of terrestrial plants.
800	Ecological Monographs, 82, 205–220.
801	Viers, J., Prokushkin, A. S., Pokrovsky, O. S., Auda, Y., Kirdyanov, A. V., Beaulieu, E. Zouiten,
802	C., Oliva, P., Dupré, B., 2013. Seasonal and spatial variability of elemental concentrations in
803	boreal forest larch foliage of Central Siberia on continuous permafrost. Biogeochemistry. 113,
804	435–49.
805	Wang, L.J., Zhang, Y.Q., Ding, Z.L., Xu, X.N., 2010. Foliar nutrient dynamics and nutrient use
806	efficiency of four dominant tree species in a subtropical evergreen broad-leaved forest in
807	Xiaokeng, southern Anhui. J NE For Uni. 38(7), 10–12 (in Chinese)
808	Wang, Q.K., Wang, S.L., Fan, B., Yu, X.J., 2007. Litter production, leaf litter decomposition and
809	nutrient return in Cunninghamia lanceolata plantations in south China: effect of planting
810	conifers with broadleaved species. Plant Soil. 297(1-2), 201-211.
811	Wang, Q.K., Wang, S.L., Huang, Y., 2009. Leaf litter decomposition in the pure and mixed
812	plantations of Cunninghamia lanceolata and Michelia macclurei in subtropical China. Bio
813	Fert Soils. 45(4), 371–377.
814	Wright, I.J., Westoby, M., 2003. Nutrient concentration, resorption and lifespan: leaf traits of
815	Australian sclerophyll species. Funct Ecol. 17, 10–19.
816	Wu, Z.L (Editor), 1984. Chinese Fir. China Forestry Publishing House, Beijing. (in Chinese)
817	Yan, T., Lü, X.T., Yang, K., Zhu, J.J., 2015. Leaf nutrient dynamics and nutrient resorption: a

comparison between larch plantations and adjacent secondary forests in Northeast China. J

- 819 Plant Ecol. 9, 165–173.
- Yoshida, T., Hijii, N., 2006. Spatiotemporal distribution of aboveground litter in a *Cryptomeria japonica* plantation. J For Res.11(6), 419–426.
- 822 Yu, X.T., 2000. Research progress of Chinese fir in the 1990s. Journal of Fujian Forestry College.
- 823 20(2), 179–188. (in Chinese)
- Yuan, Z.Y., Chen, H. Y.H. 2015. Negative effects of fertilization on plant nutrient resorption. Ecol.
 96, 373–380.
- 826 Zhang, J.C., Sheng, W.T., 2001. The study on decay of dead branches and leaves on living trees
- taken from crown into litter environment in a Chinese fir plantation, compared with decay in
 canopy. Scientia Silvae Sinicae. 37(6), 2-10. (in Chinese)
- 829 Zheng, L.X., 1997. Research on the effect of litter on the nutrient characteristics in Chinese fir
- 830 plantation. For Res. 10(6), 607–611. (in Chinese)
- 831 Zhou. L.L., Cai, L.P., Ma, X.Q., Wu, P.F., Hou, X.L., Chen, N.L., Liu, C.H., Jiang, Z.K., 2012.
- Eco-hydrological function in different developing stages of Chinese fir. Journal of Soil and
- 833 Water Conservation. 5, 249–253. (in Chinese)
- Zhou, L.L., Shalom, A.D.D., Wu, P.F, Li, S.B., Jia, Y.Y., Ma, X.Q., 2015. Litterfall production
- and nutrient return in different-aged Chinese fir (*Cunninghamia lanceolata*) plantations in
- 836 South China. J Forestry Res. 26(1), 79–89.
- 837 Zhou, L.L, Addo-Danso, S.D., Wu, P.F, Li, S.B., Zou, X.H., Zhang, Y., Ma, X.Q., 2016a. Leaf

838	resorption efficiency in relation to foliar and soil nutrient concentrations and stoichiometry of
839	Cunninghamia lanceolata, with stand development in southern China. Journal Soils
840	Sediments. 16(5), 1448–1459.
841	Zhou, L.L., Shalom, A.D.D., Wu, P.F., He, Z.M., Liu, C.H., Ma, X.Q. 2016b. Biomass production,
842	nutrient cycling and distribution in age-sequence Chinese fir (Cunninghamia lanceolata)
843	plantations in subtropical China. J Forestry Res. 27(2), 357–368.
844	Zou, B., Li, Z.A., Ding, Y.Z., Tan, W.N., 2006. Litterfall of common plantations in south
845	subtropical China. Acta Ecol Sin. 26(3), 715–721. (in Chinese)

- Zou, X.H., Wu, P.F., Chen, N.L., Wang, P., Ma, X.Q., 2015. Chinese fir root response to spatial
- and temporal heterogeneity of phosphorus availability in the soil. Can J of Forest Res. 45(4),
- 848 402–410.

Supplementary Material

Spatiotemporal distribution of canopy litter and nutrient resorption in a chronosequence of different development stages of *Cunninghamia lanceolata* in southeast China

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Fig. S1. Location of Xinkou National Forest Farm and of the different development-stage Chinese fir plantations studied.



Fig. S2. Dead branches and leaves in the canopy of middle-age Chinese fir trees at the study site.



Fig. S3. Collection of dead branches and leaves from the tree canopy using clippers (left) and climbing (right).



Fig. S4. Five categories of needles [<1-year-old, 1-year-old, perennial, senesced (brown), and yellow] were collected from each shoot.

