

1 **The importance of regional differences in vulnerability to climate change for demersal**  
2 **fisheries**

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18

19 **Abstract**

20 Regional differences in climate vulnerability are particularly important in many countries with  
21 socio-ecological gradients or geographical and environmental spatial segregation. Many studies  
22 are regularly performed at the national level, but regional assessments can provide more  
23 detailed information and important insights into intra-national vulnerabilities. They require  
24 detailed information of many socio-ecological components that are often neglected at the  
25 regional scale but are meaningful and operational at national and international levels. In this  
26 work, we developed a climate vulnerability assessment to investigate the vulnerability of  
27 demersal fisheries based on 19 indicators covering exposure, fisheries sensitivity, species  
28 sensitivity and adaptive capacity for nine coastal regions of Spain, contrasting Mediterranean to  
29 Atlantic areas. Exposure was consistently larger in the Mediterranean than Atlantic regions,  
30 while adaptive capacity showed the opposite trend. While fisheries and species sensitivity did  
31 not display a clear Atlantic-Mediterranean pattern, they were critical for capturing regional  
32 differences that have an impact on fisheries vulnerability. Fisheries-related indicators such as  
33 employment, species mean price, catch stability, landings of species with low commercial value  
34 or fleet power may be able to capture significant socio-economic differences at the regional  
35 level. Our results highlight the generally higher vulnerability of Mediterranean demersal  
36 fisheries, mainly due to the lower adaptive capacity and higher exposure of Mediterranean  
37 regions, while providing key regional elements for guiding national and international actions for  
38 adaptation. This study demonstrates that the spatial scale considered in the development of  
39 climate vulnerability assessments must recognise the spatial heterogeneity in the socio-  
40 ecological system within its unit of analysis in order to be a relevant tool for management and  
41 policy makers.

42

## 43 **Introduction**

44 The ocean plays a crucial role in the food security and livelihoods of millions of people  
45 worldwide. From 1961 human fish consumption has grown at a rate almost twice that of the  
46 world population, and higher than all other animal protein foods, fuelling a sector that currently  
47 creates almost 40 million direct employment positions worldwide (Barange et al., 2018; FAO,  
48 2020). Nevertheless, numerous studies suggest that fisheries production is being compromised  
49 by climate change impacts on the marine environment (Hollowed et al., 2013; Moore et al.,  
50 2018; Barange et al., 2018). Among the most studied effects are the shifts in species distribution  
51 and marine productivity, with potential consequences for fisheries landings, the economy and  
52 food security. Evidence of alarming rates in these shifts has emerged over the last two decades,  
53 and they are expected to continue to increase in the future (e.g., Hollowed et al., 2013; IPCC,  
54 2014a; Gattuso et al., 2015; Barange et al., 2018; Pinsky et al 2018; Free et al., 2019). However,  
55 these effects are not experienced homogeneously at a global scale, and the impacts of climate  
56 change depend on the environmental and socio-economic characteristics of each region (Alison  
57 et al., 2009; IPCC, 2014a; Payne et al. 2020).

58 Climate vulnerability assessment (CVA) is an analytical framework developed to  
59 understand, quantify and synthesize climate change impacts on socio-ecological systems (IPCC,  
60 2001). In it, “vulnerability” is given as a function of the “sensitivity” of a system to climate  
61 change, the degree of “exposure” to climate hazards, and “adaptive capacity”, as the  
62 community's ability to prevent or compensate potential impacts of climate damage (IPCC, 2001).  
63 The CVA approach has recently been revised, and its focus has shifted from “vulnerability” to  
64 “risk”, partly intending to use terms that embrace uncertainties in simulations of future climate  
65 impacts (IPCC, 2014a). Despite these changes in nomenclature, the core of the CRA and CVA  
66 analysis framework remains essentially unchanged.

67 To date, several studies have used CVAs and CRAs to investigate the impacts of climate  
68 change on the fisheries sector to prioritise where to allocate adaptations funds and provide  
69 policy-makers with sufficient knowledge to undertake practical decisions. These studies  
70 consistently emphasise the importance of investing in adaptation mechanisms in light of climate  
71 change threats (e.g., Alison et al. 2009; Cinner et al., 2012; Colburn et al., 2016; Pinnegar et al.  
72 2019; Payne et al., 2020). This approach has recently been applied at a sub-national level,  
73 achieving more detailed information and leading to more accurate analyses of regional  
74 vulnerability (Pinnegar et al., 2019; Barnes et al., 2020; Payne et al., 2020). It follows that  
75 regional differences are particularly important in many countries with socio-ecological gradients  
76 or geographical (and environmental) spatial segregation, for which differences should be  
77 expected in the combined effects of: (i) the impacts of climate change (Exposure); (ii) the  
78 relevance of fisheries to their economies and diets (Sensitivity); and (iii) the limited capacity for  
79 social adaptation (Adaptive capacity) (Cinner et al., 2012; Pinnegar et al. 2019; Payne et al.  
80 2020). However, quantifying these spatial differences poses the challenge of searching for a  
81 more diverse set of socio-ecological indicators that capture these important differences at the  
82 regional scale, while being meaningful and operational at the national level.

83 Spain is a paradigmatic example regarding all these characteristics, with two large  
84 contrasting areas associated to the Atlantic Ocean and the Mediterranean Sea, and clear  
85 gradients and spatial heterogeneity in the ecological, fisheries and socio-ecological contexts  
86 (e.g., Hidalgo et al., 2017; Punzón et al., 2020). Indeed, the characteristics of the biological  
87 communities and the fisheries that depend on them are also markedly different (FAO 2018,  
88 2020). Among the common characteristics, demersal fishing plays an important role in both the  
89 Atlantic and Mediterranean areas with over 1/3 of the fleet and more than half the total Spanish  
90 fleet's power and gross tonnage (MAPA, 2019). To highlight the importance of regional  
91 differences in vulnerability to climate change, we use the demersal fishery to apply a  
92 vulnerability assessment framework and identify the Spanish coastal regions that are more

93 threatened by climate change according to the following criteria: where the continental shelf is  
94 most exposed to the climate change impacts, where the fishing industry contributes the most  
95 to livelihoods and economic growth, and/or where social resources and infrastructure may limit  
96 adaptive capacity.

## 97 **Material and methods**

98 Spain is among the 25 largest seafood producers globally and hosts the largest fishing industry  
99 in the EU, representing 21% of the total of active companies and 25% of its turnover (STECEF,  
100 2019a). In 2017, Spain produced around 1 million tons of seafood, 18% of the total EU fleet  
101 landings, with an estimated revenue of 2,000 million euros (STECEF, 2019b). The Spanish  
102 demersal fleet accounted in 2017 for more than 70% of the total revenue with an estimated  
103 value of 1,500 million euros (STECEF, 2019b). With approximately 8,000 of coastline of km and  
104 340 registered fishing ports (EUMOFA, 2020), Spanish coastal regions have distinct  
105 characteristics in terms of the relative contribution of different fleets, their dependence on  
106 fishing and their socio-economic realities. There are also marked differences between the  
107 Atlantic and Mediterranean regions, each having its own fisheries governance structures, with  
108 the International Council for the Exploration of the Sea (ICES) active in the Atlantic and the  
109 General Fisheries Commission for the Mediterranean (GFCM) in the Mediterranean, developing  
110 the fisheries assessment of fisheries resources (e.g., GFCM, 2019; ICES, 2019, for demersal  
111 species).

112 In this context, taking into account: (i) the importance of demersal fishing in Spain; (ii) the  
113 heterogeneity of its coastal regions; (iii) and the importance of the fishing sector for livelihood  
114 and food security, we investigated the regional vulnerability of demersal fisheries based on 19  
115 indicators covering exposure (E), fisheries sensitivity (FS), species sensitivity (SS) and adaptive  
116 capacity (AC) for nine coastal regions of Spain: Galicia, Asturias, Cantabria, Basque Country,

117 Catalonia, Valencia, Balearic Islands, Murcia and Andalusia. This spatial coverage allowed us to  
118 assess differences among these administrative areas and, more generally, between the Atlantic  
119 and Mediterranean domains to measure the impact of climate change on fisheries, contrasting  
120 these two large areas. Note that Andalusia was considered a Mediterranean region, although  
121 part of its coast (the Gulf of Cádiz) is within the Atlantic domain. We made this assumption in  
122 order to maintain this region in our analyses, as most indicators could not be partitioned for the  
123 Atlantic and Mediterranean regions. The Macaronesian region (i.e., Canary Islands) has not been  
124 included in the study due to the minor role played by demersal fishing on its insular shelf and  
125 also, in order to maintain the comparability between the Atlantic and Mediterranean areas.

126 Exposure (E) is defined as the nature and degree to which a system is subjected to significant  
127 climatic variations such as temperature anomalies, extreme weather events, or other climate  
128 change effects (Füssel & Klein, 2006; IPCC, 2007). The analysis of physical variables is widely  
129 considered in vulnerability analysis studies to determine the degree of exposure of the fisheries  
130 sector, affecting its operational capacity, economic benefits and food security (Alison et al.,  
131 2009; Pinnegar et al., 2019). In this study, we focused on the sea surface temperature trends  
132 and the continental shelf area of each region, which was included as a proxy of the probability  
133 of exposure to other climate impacts (IPCC, 2014b; Barange et al., 2018; Maxwell et al., 2019)  
134 (Table 1). The continental shelf contains an important part of the fishing resources and the  
135 effects of climate change on the species that inhabit this area directly threaten the fishing  
136 activity and economic productivity. With this in mind, we assume that the regions with the  
137 largest shelf areas will tend to have a greater number of impacts associated with climate change,  
138 as well as the occasional and/or extreme impact of some of them. Fisheries sensitivity (FS) is  
139 defined as the component of total sensitivity that refers to the socio-economic sub-system:  
140 employment, food security, among others (Colburn et al., 2016; Pinnegar et al., 2019; Tables 1).  
141 We used four indicators to assess the fisheries sensitivity in each of the nine coastal regions:  
142 Fleet power, Fleet age, Fish consumption, Employment (Table 1). Among the characteristics of

143 the fleet, fleet power is highly correlated with tonnage and indicates larger vessels which are  
144 less flexible regarding fishing areas and fishing gear.

145 Adaptative capacity (AC) is the degree of adjustment that ecological, social or economic systems  
146 can achieve to balance the actual or projected climate and its impacts, including actions that  
147 moderate, prevent damage or exploit beneficial opportunities (Noble et al., 2014; UNFCCC,  
148 2018). It includes elements such as the level of social capital, human capital, and the adequacy  
149 and effectiveness of governance structures, and it presupposes an indissoluble relationship  
150 between climate change actions and the central imperatives of reducing poverty, increasing  
151 food security and ending hunger (Haddad, 2005; Yohe et al., 2006; Tol & Yohe 2007; Vincent,  
152 2007). In this study, in addition to these elements that are widely used in climate vulnerability  
153 analysis studies we considered factors of adaptive capacity related to: the mean number of  
154 landing ports per fishing vessel, or the potential of small scale and recreational fishing (Tables  
155 1). The possibility of landing outside the base port offers the fishing fleet options for avoiding  
156 extreme weather events and seeking a higher economic return on sea voyages. Artisanal and  
157 recreational fishing emerge in this scenario as an option for migrating boats and fishing crews  
158 to these small-scale fisheries and the tourist market, an opportunity for adaptation since climate  
159 change affects fishing yields and viability. Note than some of these factors could serve as either  
160 sensitivity or adaptive capacity indicators. To assess the species sensitivity (SS) to climate  
161 change, we collected information on relevant biological traits of the most landed species in both  
162 the Atlantic and Mediterranean areas between 2016 and 2018. These traits included  
163 temperature preferences (mean and range), spawning duration (months), and depth range, in  
164 addition to landing stability and market data (Alison et al., 2009; Hare et al., 2016). These  
165 indicators were calculated at the species level and subsequently weighted by the mean CPUE of  
166 each species in each region during the period 2016-2018, to represent the sensitivity of the  
167 biological communities (Tables 1).

168

169 ***Construction of the vulnerability index***

170 A consistent application of weighting criteria has been suggested as an important step to avoid  
171 an arbitrary classification of the indicators (Johnson et al., 2016). Because each indicator has its  
172 own units, we rescaled the indicators between 0 and 1 to make them unitless and comparable.  
173 However, by doing so, we artificially equalled the range of variability of all the indicators. To  
174 compensate for this, we use the Coefficient of Variation of each indicator (CV) as a weighting  
175 factor, and thus we ended up with a set of unitless and comparable indicators ranked according  
176 to their original variability. Following this approach, we computed the exposure (E), fisheries  
177 sensitivity (FS), species sensitivity (SS) and adaptive capacity (AC) indices which varied between  
178 0 and 1. The final vulnerability was calculated by adding exposure (E), fisheries sensitivity (FS),  
179 species sensitivity (SS), and subtracting adaptive capacity (-AC).

180 
$$V = E + FS + SS - AC$$

181 The final vulnerability scores were higher for the most vulnerable regions (i.e., high exposure,  
182 high sensitivity and low adaptive capacity), and lower for the least vulnerable ones. These final  
183 index scores were normalised between 0 to 1, with 0 being the lowest vulnerability and 1 the  
184 highest vulnerability.



**Table 1.** Summary of variables and data sources used to calculate exposure, fisheries and species sensitivity and adaptive capacity of the coastal regions to vulnerabilities associated with climate change impacts on marine demersal fisheries.

Component	Indicator	Variable	Interpretation	Data sources
Exposure	Sea surface temperature	Mean projected sea surface temperature increase (°C at 0.5 m depth) by 2050	Projected trends in the regionalisation of the IPCC climate scenarios indicate future exposure of the region to water temperature changes.	COPERNICUS, 2020
	Continental shelf area	Surface of the continental shelf (km <sup>2</sup> ) between 0 and 200 m	A larger extension of continental shelves (0-200 m) makes fishing regions more exposed to the impacts of climate change as extreme weather events.	EMODNET, 2019
Fisheries sensitivity	Employment	Persons working in marine fisheries as % of the total economically active population	Employment in the sector is a direct indicator of its size. In this case, the number of registrations in the sea regime quantifies the fisheries sector capture component. The larger the size, the greater the sensitivity to change.	INE, 2018
	Fleet power	Mean average power of demersal fleet (kW) (2016-2018)	Average vessel power is a measure of fleet capacity; the more powerful fleets are considered the most sensitive and possibly the least adaptive.	STECF, 2019
	Fleet age	Average vessel age (2016-2018)	Older and therefore less efficient fleets are considered to be more sensitive to the effects of climate change.	STECF, 2018
	Fish consumption	Per capita consumption of fishery products (kg/year).	High consumption of seafood is directly related to dependence on this source of protein and, consequently, to sensitivity to climate change.	MAPA, 2019
Species sensitivity	Price	Price of the main commercial species	The species with the highest market value typically are subject to the greatest fishing pressure, which is why it is estimated that they are more sensitive to climate change.	STECF, 2017

	CPUE stability	Coefficient of variation of the CPUE for the main commercial species by region (2015-2018)	Species with the most variable CPUE are generally those most dependent on environmental conditions and are therefore considered the most sensitive to climate change.	IEO, unpublished data.
	Temperature sensitivity	Preferred temperature range of a species (i.e., range between the 10th and 90th percentile, T10 and T90) divided by its preferred mean temperature.	The species most sensitive to temperature changes will have to withstand narrower temperature ranges and have an affinity for lower temperatures.	AQUAMAPS, 2020
	Depth range	Bathymetric range (quantile-10 (D10) of the depth distribution and the quantile-90 (D90))	Inhabiting narrower depth ranges increases species' sensitivity to climate change by limiting their ability to migrate at higher depths.	IEO, unpublished data
	Spawning period	Spawning season of each species (months/year)	Species with shorter spawning periods will be more sensitive to the seasonal temperature changes expected due to climate change.	FishBase; SeaLife, 2020
Adaptive capacity	Low value species landings	(CPUE of species with low commercial value/total CPUE) * 100 (2016-2018)	The percentage of landings of species with low commercial interest indicates less dependence on resources of high commercial value and is considered an indicator of greater adaptability.	IEO, unpublished data
	Landings outside the base port	Mean number of ports where each vessel landed (2016-2018)	Landings outside the main port can be considered as indicators of the fleet's ability to adapt, for example to adverse weather conditions or market opportunities.	IEO, unpublished data.
	Per capita GDP	Per capita gross domestic product	Higher GDP per capita allows for less economic dependence on any single activity, including fisheries, and therefore indicates greater ability to adapt to change.	INE, 2019
	Education	% of the population that complete non-university studies	The percentage of education is an indicator of the population's empowerment, allowing greater adaptation to changing conditions.	MEFP, 2019

Small scale fisheries	Number of boats using small-scale fishing gear	A larger small-scale fleet represents more opportunities for workers to move within the fishing sector from large-scale to artisanal fishing in order to guarantee their profits and livelihoods.	MAPA, 2019
Recreational fisheries	Number of recreational fishers	Recreational fishing emerges as an option for migrating boats and fishing crews to the tourist market, an opportunity for adaptation since climate change affects fishing yields and viability.	Gordoa et al., 2019
Associations	Number of Local Fisheries Action Groups (GALPs) added to the number of Associative Entities of Autonomous Scope (AAS) of the fishing sector	The associative capacity of the sector is related to a greater capacity to adapt to climate change.	REGP, 2019
Gear diversity	Shannon diversity index of main fishing gear (2016-2018)	A greater diversity of fishing gear within the fleet is related to a greater diversity of target species. Diversification increases the capacity to adapt.	STECF, 2019

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COPERNICUS - European Union Earth Observation Programme; EMODNET – European Marine Observation and Data Network; INE - National Statistical Institute of Spain; STECF - Scientific, Technical and Economic Committee for Fisheries; MAPA - Ministry of Agriculture, Fisheries and Food of Spain; IEO - Spanish Institute of Oceanography; MEFP - Spanish Ministry of Education and Vocational Training; REGP - Spanish Network of Fisheries Groups.

## 192 **Results**

### 193 ***Exposure (E)***

194 The increase sea surface temperature marked the difference in exposure between the Atlantic  
195 and Mediterranean (Wilcoxon  $\rho = 0.019$ ,  $p < 0.05$ ). The Mediterranean regions presented higher  
196 warming trends (between  $0.023^{\circ}\text{C}/\text{year}$  and  $0.028^{\circ}\text{C}/\text{year}$ ) than the Atlantic regions (between  
197  $0.014^{\circ}\text{C}/\text{year}$  to  $0.04^{\circ}\text{C}/\text{year}$ ). Note that the highest sea surface temperature increases in the  
198 Mediterranean are twice the highest increases in the Atlantic (Fig. 1a). The continental shelf  
199 indicator was characterised by high inter-regional variability (Fig. 1b). However, no significant  
200 difference was found between the Atlantic and Mediterranean areas (Wilcoxon  $\rho = 0.540$ ,  $p >$   
201  $0.05$ ).

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### 203 ***Fisheries Sensitivity (FS)***

204 In terms of fisheries sensitivity, the indicator with the greatest inter-regional variability was the  
205 fisheries employment. The result in the Atlantic region of Galicia (2.1%) was by far the largest,  
206 while this indicator ranged between 0.2 and 0.5% in the other regions. The difference between  
207 Atlantic and Mediterranean areas was significant (Wilcoxon  $\rho = 0.049$ ,  $p < 0.05$ ). The per capita  
208 consumption of fishery products was between 21 and 30 kg/year throughout Spain, with higher  
209 consumption in the Atlantic than in the Mediterranean area (Wilcoxon  $\rho = 0.019$ ,  $p < 0.05$ ). The  
210 demersal fleet was generally older in Mediterranean regions (Mean  $\approx 38$  years) than the Atlantic  
211 ones (Mean  $\approx 27$  years), while this difference was not significant (Wilcoxon  $\rho = 0.109$ ,  $p > 0.05$ ;  
212 Figure 1e). Fleet power, used as a proxy for the demersal fleet catch capacity, showed high inter-  
213 regional variability, but no distinct pattern between Atlantic and Mediterranean areas was  
214 identified (Wilcoxon  $\rho = 0.902$ ,  $p > 0.05$ ; Figure 1f).

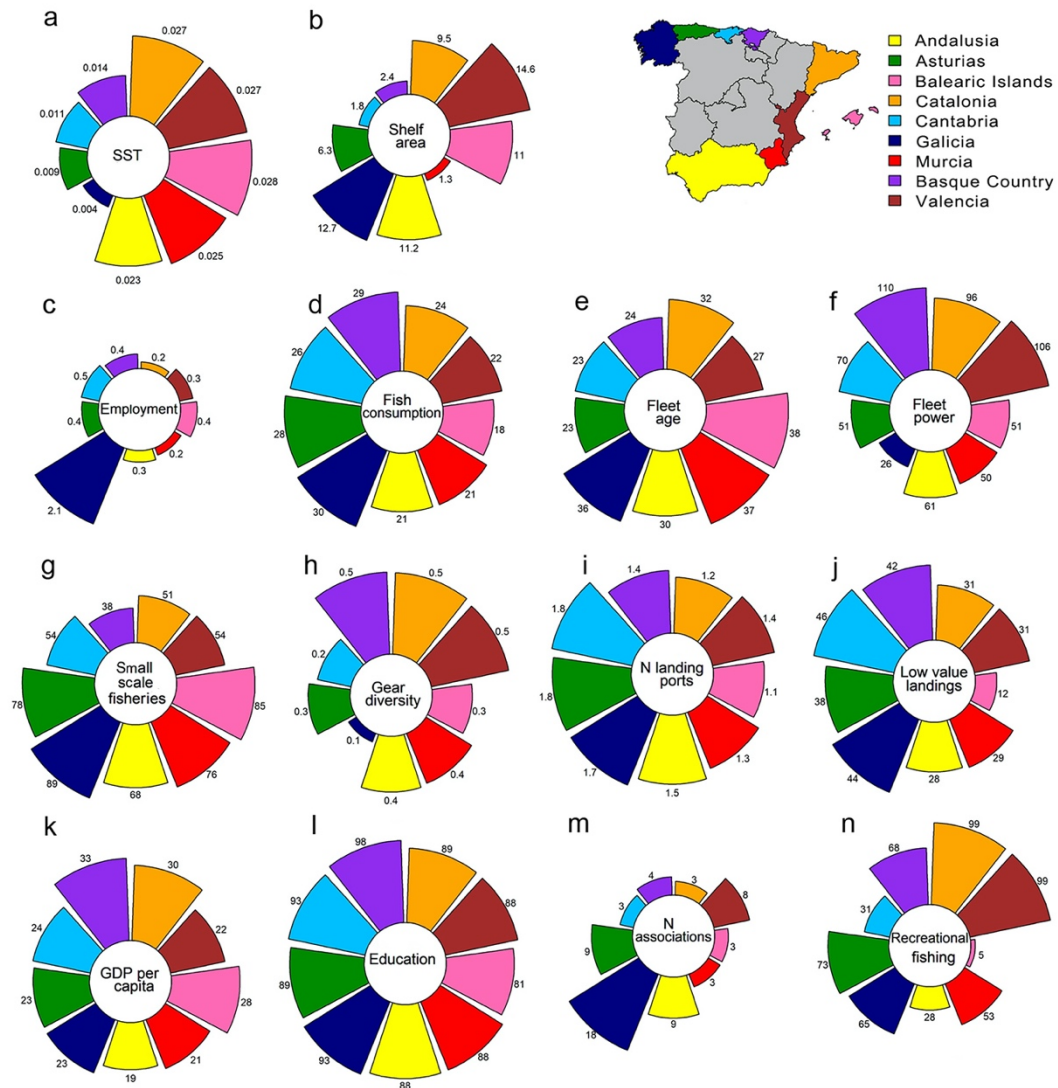
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### 216 ***Adaptive capacity (AC)***

217 Our results showed that three out of the five largest artisanal fleets belong to Mediterranean  
218 regions and two to Atlantic ones (Figure 1g). Although there was no clear pattern between these  
219 two major areas (Wilcoxon  $\rho = 0.903$ ,  $p > 0.05$ ), the inter-regional scale differences were high.  
220 Gear diversity was generally higher in the Mediterranean regions, although no significant  
221 difference was revealed between the two areas (Wilcoxon  $\rho = 0.208$ ,  $p > 0.05$ ; Figure 1h).  
222 Regarding the mean number of ports for landing, there were more for the Atlantic than for the  
223 Mediterranean, showing a clear difference (Wilcoxon  $\rho = 0.048$ ,  $p < 0.05$ ; Figure 1i). The inter-  
224 regional variability was also pronounced for this indicator. In terms of low commercial value  
225 species landings, we also found significant differences between the Atlantic and Mediterranean  
226 areas (Wilcoxon  $\rho = 0.019$ ,  $p < 0.05$ ), with the Mediterranean regions showing a lower  
227 percentage of catches of low commercial value than the Atlantic (Figure 1j).

228

229 The Gross domestic product (GDP) ranged between 19.000 and 33.000 € throughout Spain, with  
230 the highest GDP belonging to the Atlantic Basque Country region, while Andalusia and Murcia,  
231 both located in the Mediterranean were the regions with the lowest GDP (Figure 1k). No  
232 significant difference was found between the Atlantic and Mediterranean areas (Wilcoxon  $\rho =$   
233  $0.389$ ,  $p > 0.05$ ). The rate of non-university education was higher for the Atlantic than for the  
234 Mediterranean (Wilcoxon  $\rho = 0.023$ ,  $p < 0.05$ ; Figure 1l). In terms of civil participation in fisheries  
235 management organisations, the Atlantic region of Galicia was noteworthy for having the largest  
236 number of associations, including Local Support Groups (GALP's) and Associative Entities of  
237 Autonomous Scope for the fishing sector (Figure 1m). Apart from this region, no pattern  
238 emerged between the Mediterranean and Atlantic areas (Wilcoxon  $\rho = 0.368$ ,  $p > 0.05$ ).  
239 Regarding the number of recreational fishers, the largest numbers belonged to Mediterranean  
240 regions but still, inter-regional variability was high and no generalised pattern between Atlantic  
241 and Mediterranean areas could be recognised (Wilcoxon  $\rho = 0.902$ ,  $p > 0.05$ ; Figure 1n).



242  
 243 **Figure 1.** The figure shows the indicators of *Exposure*: (a) Continental shelf area, (b) Sea surface temperature trend; *Fisheries*  
 244 *Sensitivity*: (c) Fisheries employment, (d) Fish consumption per capita, (e) Fleet age, (f) Demersal fleet power; and *Adaptive capacity*  
 245 (g) Small scale fisheries, (h) Gear diversity, (i) Landings outside the base port, (j) Low-value species landings, (k) Per Capita GDP, (l)  
 246 Non-university education, (m) Autonomous Community Associations, (n) Recreational fisheries. The map shows the colours  
 247 designated for each coastal region that follow the same pattern in the presentation of the indicator results.

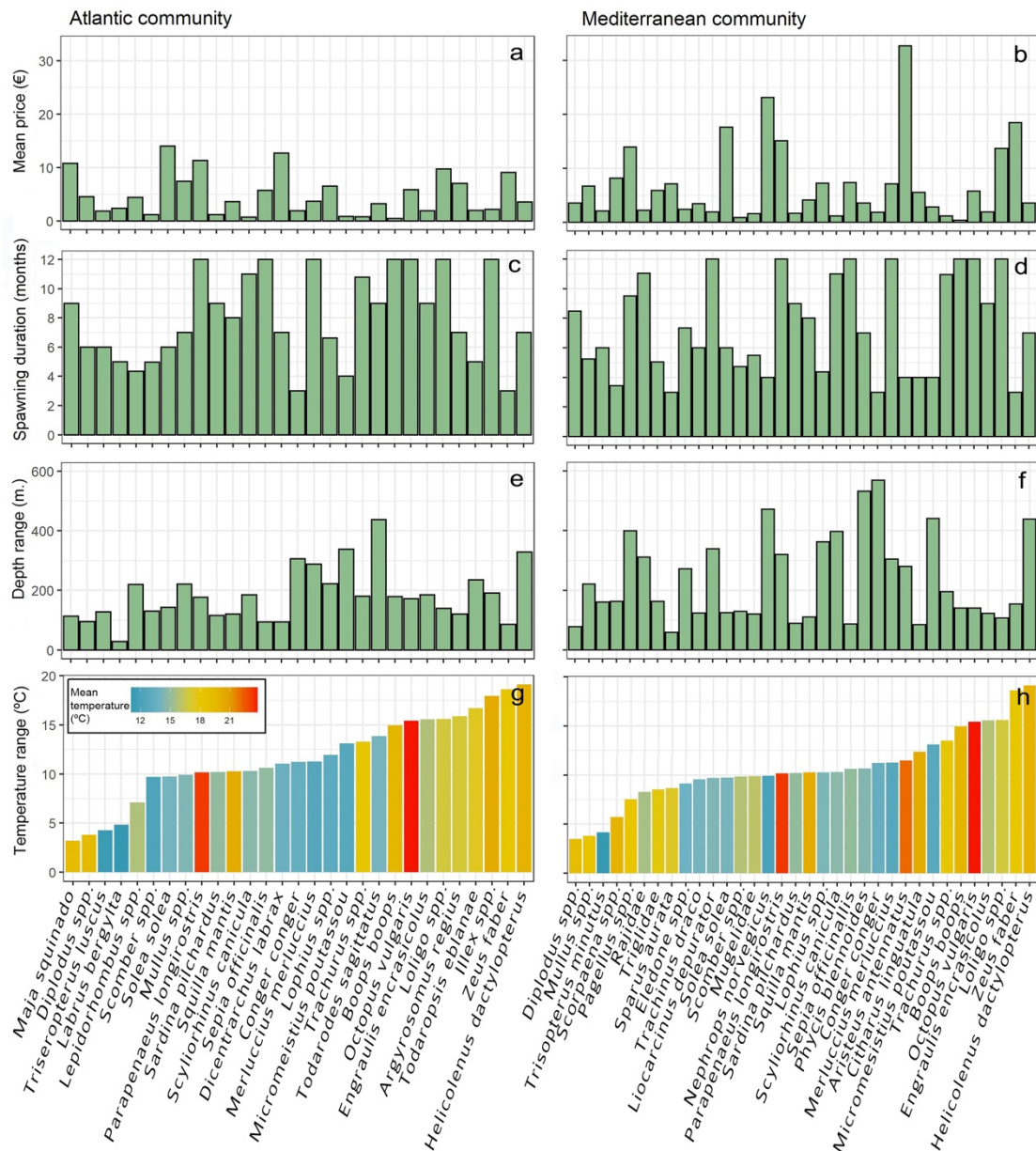
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### 249 **Species Sensitivity (SS)**

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251 In terms of the economic value of commercial species, our results showed major price variations  
 252 with the Mediterranean showing a higher price for most of the species investigated ( $5.48 \pm 5.71$   
 253 € in the Atlantic and  $6.49 \pm 6.71$  € in the Mediterranean; Figure 2a, 2b). Despite major price  
 254 variations not significant differences between the Atlantic and Mediterranean areas was found

255 (Wilcoxon  $\rho = 0.389$ ,  $p > 0.05$ ). In terms of life history traits, the spawning period showed similar  
256 patterns between the Atlantic and Mediterranean biological communities (Wilcoxon  $\rho = 0.474$ ,  
257  $p > 0.05$ ; Figure 2c, 2d). Regarding the species depth ranges, we found differences between the  
258 species when contrasting Atlantic and Mediterranean populations, such as for Conger conger,  
259 *Scyliorhinus canicula* and *Lophius* spp., with generally wider ranges in the Mediterranean  
260 populations ( $181.78 \pm 90.08$  m. in the Atlantic and  $235.59 \pm 145.38$  m. in the Mediterranean;  
261 Figure 2e, 2f). No significant difference was found between the two major areas (Wilcoxon  $\rho =$   
262  $0.777$ ,  $p > 0.05$ ). In terms of temperature, the Mediterranean and Atlantic areas showed a similar  
263 temperature range ( $11.71 \pm 4.27$  °C in the Atlantic and  $10.67 \pm 3.65$  °C in the Mediterranean;  
264 Figure 2g, 2h). Species that require low mean temperatures displayed intermediate temperature  
265 ranges (except for *Parapenaeus longirostris*), and species requiring higher mean temperatures  
266 displayed both high and low temperature ranges. This interaction translated into temperature  
267 sensitivity showed a similar pattern between Atlantic and Mediterranean areas (Wilcoxon  $\rho =$   
268  $0.74$ ,  $p > 0.05$ ), being slightly higher in the Mediterranean community ( $1.7 \pm 1.23$  in the Atlantic  
269 and  $1.81 \pm 1.16$  in the Mediterranean).



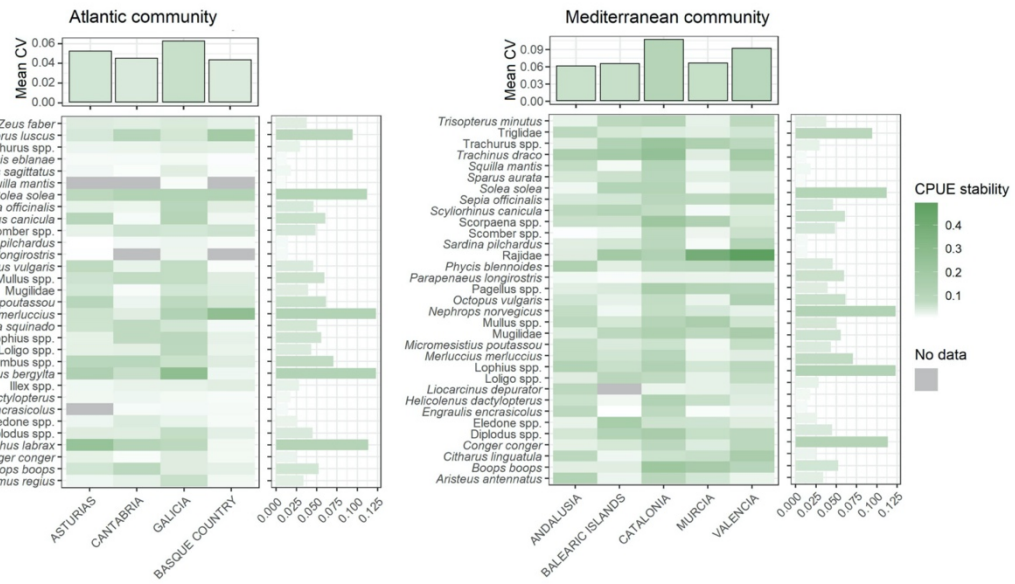
**Figure 2.** Species sensitivity indicators for the Atlantic and Mediterranean main commercial species targeted by demersal fisheries. Mean price of the main commercial species (a, b); spawning period (c, d); depth range (e, f) and temperature range (g, h).

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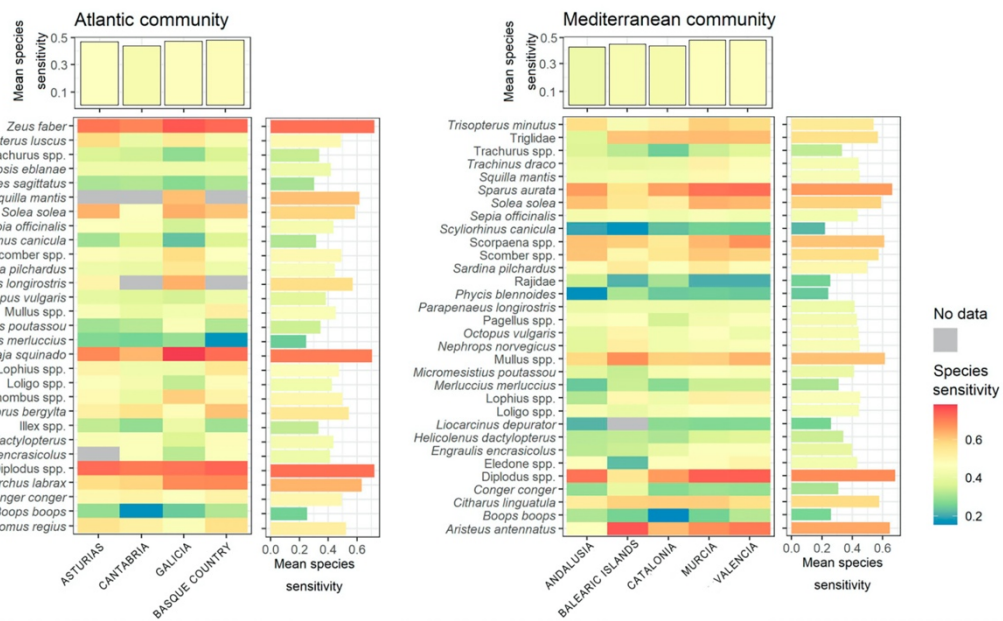
275 The stability of landings showed high inter-regional variability. In general terms, stability was  
276 lower in the Atlantic area than in the Mediterranean (Wilcoxon  $\rho = 0.037$ ,  $p < 0.05$ ;  $0.051 \pm 0.047$   
277 and  $0.078 \pm 0.064$  respectively; Figure 3). The mean species sensitivity was similar in the Atlantic  
278 and Mediterranean areas (Wilcoxon  $\rho = 0.493$ ,  $p < 0.05$ ;  $0.46 \pm 0.24$  and  $0.44 \pm 0.22$  respectively;  
279 Figure 4). Some species were particularly sensitive in the Atlantic (e.g., *Zeus faber* and *Maja*  
280 *squinado*) and in the Mediterranean (e.g., *Aristeus antennatus* and *Sparus aurata*), but



281 generally, the species present in both areas had similar sensitivity values; e.g., *Diplodus spp.* And  
 282 *Solea solea* were among the most sensitive while *Boops boops* and *Scyliorhinus canicula* were  
 283 among the least sensitive.



284  
 285 **Figure 3.** Indicator of stability of landings for Atlantic (left) and Mediterranean (right) species based on the mean CPUE stability  
 286 (1/CV). Differences between species and regions are shown for both large areas.



287  
 288 **Figure 4.** General species sensitivity for Atlantic (left) and Mediterranean (right) based on the combination of the species sensitivity  
 289 indicators. Differences between species and coastal regions are shown for both large regions.

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291 ***Integrative indicators and overall vulnerability index***

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293 The integrative exposure index showed that the regions with the highest degree of exposure

294 were: Valencia, Catalonia, Balearic Islands and Andalusia (Figure 5a), marking a clear distinction

295 between the Atlantic and Mediterranean areas (Wilcoxon  $\rho = 0.025$ ,  $p < 0.05$ ). The regions with

296 the greatest fisheries sensitivity were Galicia and the Basque Country (Figure 5b). The result in

297 Galicia was driven mainly by the large number of jobs in the fishing sector, which was the

298 indicator ranking highest regarding inter-regional variability and thus had the largest

299 contribution to the fisheries sensitivity index (Figure 5b). The Basque Country had the largest

300 demersal fleet power among all the regions studied, which justifies its position as the second

301 most sensitive region in terms of its fishing fleet. As these two regions are located in the Atlantic,

302 the mean fisheries sensitivity of the Atlantic area was generally higher than that of the

303 Mediterranean area (Figure 5b). However, no significant difference was found between Atlantic

304 and Mediterranean areas (Wilcoxon  $\rho = 0.867$ ,  $p < 0.05$ ).

305

306 In terms of species sensitivity Murcia, Valencia and Andalusia were the most sensitive regions,

307 pointing to a more sensitive pattern for the Mediterranean area (Wilcoxon  $\rho = 0.493$ ,  $p < 0.05$ ;

308 Figure 5c) mostly due to the differences in the mean price of the main commercial species. Our

309 analyses have shown that the regions with the least adaptive capacities were the Balearic Islands

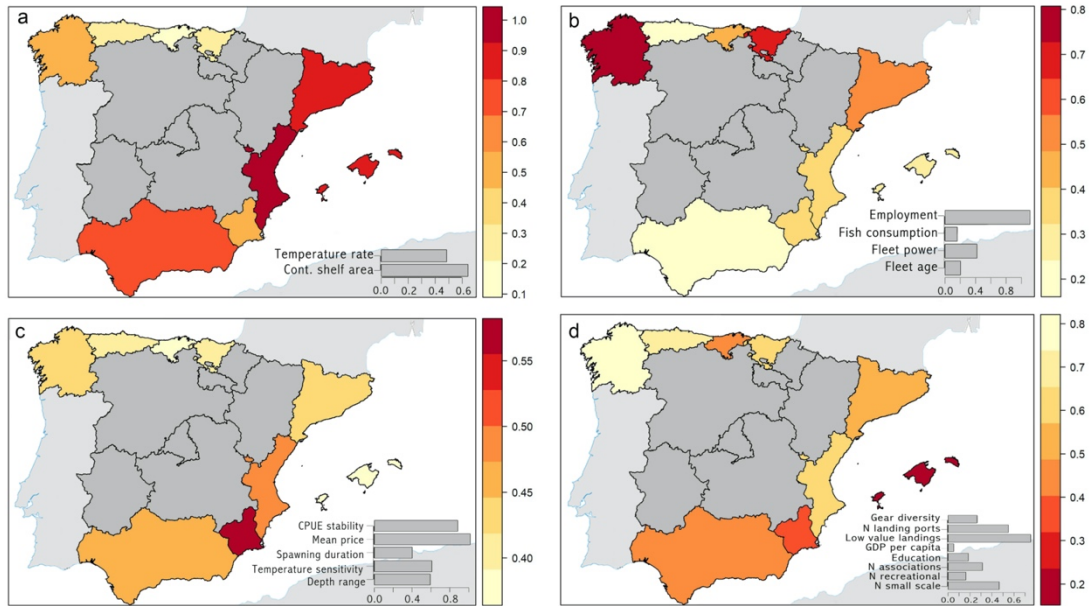
310 and Murcia (Figure 5d). The results showed a distinct pattern between the Atlantic and

311 Mediterranean regions (Wilcoxon  $\rho = 0.025$ ,  $p < 0.05$ ), with a lower adaptive capacity in the

312 Mediterranean region mainly due to the lower number of landing ports, and the lower landings

313 of species of low commercial value (Figure 5d), both indicators being relatively higher than in

314 the Atlantic regions.



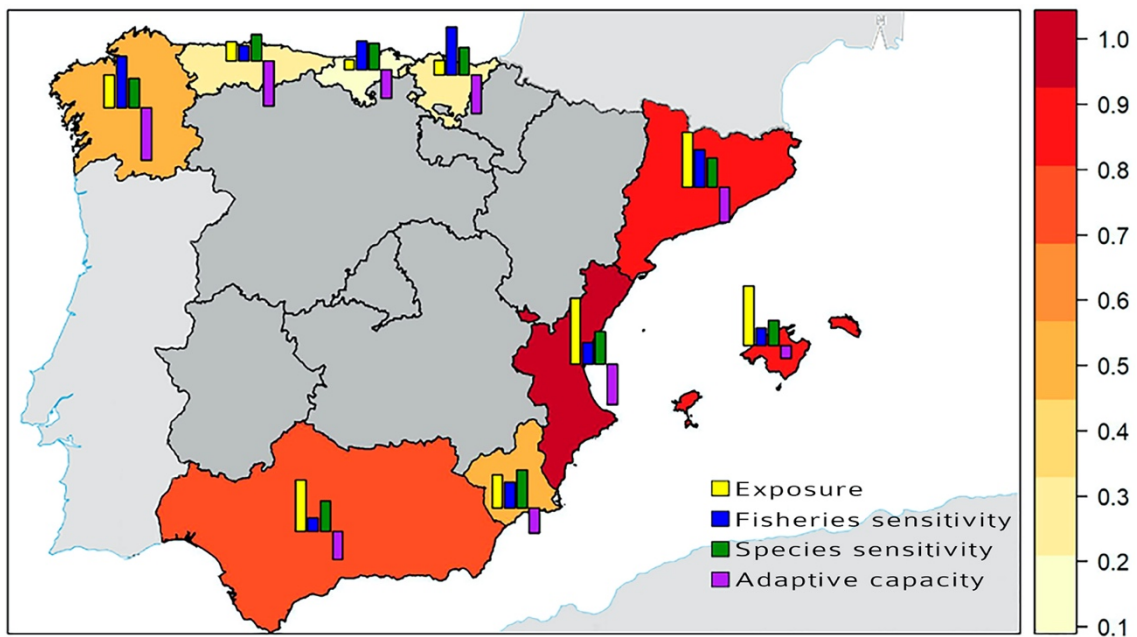
315  
 316 **Figure 5.** Thematic maps illustrating Exposure (a), Fisheries Sensitivity (b), Species Sensitivity (c) and reverse Adaptive Capacity (d)  
 317 of the Spanish coastal regions (see scale bar; darker colours indicate higher scores associated with higher sensitivity, higher  
 318 exposure, and lower Adaptive capacity). The grey bars indicate the weight of each indicator in constructing the indexes considering  
 319 their respective Coefficients of Variation.

320

321 In the Mediterranean area Valencia, Balearic Islands, Catalonia and Andalusia emerged as the  
 322 most vulnerable regions, while the less vulnerable regions were in the Atlantic, i.e., Cantabria,  
 323 Asturias and Basque Country (Figure 6). These results highlight the generally higher vulnerability  
 324 of Mediterranean demersal fisheries (Wilcoxon  $\rho = 0.026$ ,  $p < 0.05$ ), partly due to the lower  
 325 adaptive capacity of Mediterranean regions and their higher exposure. Galicia was an exception  
 326 in the Atlantic, with intermediate vulnerability values associated with relative higher exposure  
 327 (a broad continental shelf) and high fisheries sensitivity (Figure 6).

328

329



330

331 **Figure 6.** Map illustrating climate vulnerability for each Spanish coastal region based on combined scores for all four attributes (E,  
 332 FS, SS, AC). The vulnerability index is rated between 0 and 1 (see scale bar; darker colours indicate higher scores associated with  
 333 higher vulnerability).

334

### 335 **Discussion**

336 Our study conducted a CVA at the regional level based on 19 indicators of exposure, adaptive  
 337 capacity, and demersal fisheries and species sensitivity. The work sought to balance the  
 338 vulnerability of species' biological traits and each region's specific socio-economic  
 339 characteristics, highlighting each coastal region's capacity to be resilient against the observed  
 340 and expected diversity of climate change impacts. Despite the increasing number of CVAs (and  
 341 CRAs), there is a general lack of assessments specifically considering socio-economic and  
 342 ecological uniqueness at regional scales, particularly in the European regions. This research gap  
 343 could be related to the results of previous CVAs performed at the national level, which indicated  
 344 a generally low vulnerability of the European continent to the risks of climate change, mostly  
 345 due to its greater adaptive capacity, when compared to less economically developed regions  
 346 (Allison et al., 2009; Ding et al., 2017). However, recent examples show the need to downscale  
 347 CVAs (e.g., Pinnegar et al, 2019), mainly when designing effective management approaches

348 (Holsman et al., 2019, 2020). In this regard, the European regions present unique challenges in  
349 terms of climate risks on a regional scale (Payne et al., 2020), requiring further detailed analyses  
350 at a finer resolution. Within Europe, Spain emerges as one of the countries with the most  
351 vulnerable fishing sector, despite recent studies at a sub-national level that point to  
352 intermediate vulnerability values for the Spanish coastal regions (Payne et al., 2020). Beyond  
353 clear Atlantic-Mediterranean differences, our study also revealed and characterised previously  
354 unknown spatial differences in demersal fisheries vulnerability, of concern to national and  
355 international bodies in charge of the development of climate change adaptation plans.

356

357 Our results point to a generally higher exposure of the Mediterranean coastal regions to climate  
358 change risk, than is the case for the Atlantic regions. Indeed, sea surface warming has been more  
359 rapid in the Mediterranean Sea than in the Atlantic in the last decades (e.g., Belkin, 2009) and  
360 these contrasting warming rates are predicted to persist (Adloff et al., 2015; Aznar et al., 2016).  
361 Temperature predictions are also the most widely accepted output of climate forcing models  
362 and have good short to medium-term predictive power (e.g., Aznar et al., 2016). However, other  
363 climate change related factors such as acidification, changes in rainfall and, frequency and  
364 intensity of extreme meteorological events such as hurricanes or cyclones are not so ubiquitous  
365 in the CVA literature, even though they can have clear effects on fisheries vulnerability (Colburn  
366 et al., 2006; Metcalf et al., 2015; Wabnitz et al., 2018; Pinnegar et al., 2019). This is a limitation  
367 for most CVA, which fail to incorporate additional climate factors as these often lack regionalised  
368 projections under climate change scenarios. Thus, in addition to the predicted temperature  
369 trends, we considered the continental shelf area of each coastal region to account for other  
370 climate change-related factors. Indeed, the relatively broader shelf areas in most Mediterranean  
371 regions leave their demersal biological communities and fisheries more exposed and thus more  
372 vulnerable to the diversity of impacts expected to increase in the Mediterranean, such as the  
373 increase of extreme events and heatwaves, changes in vertical mixing and productivity regimes,

374 or changes in the regional circulation and population connectivity (Hidalgo et al., 2018;  
375 Darmaraki et al., 2019; Ser-Giacomi et al., 2020, MedECC 2020). In contrast, CVAs published on  
376 European fisheries do not capture this difference, since the exposure indicators considered were  
377 different and mainly focused on the diversity of the portfolio of species captured as an indicator  
378 of fishery resilience (Payne et al., 2020). This metric is highly relevant over large geographic  
379 scales (i.e., broad Europe) as it captures the differences between regions with low diversity of  
380 fish in the catches, such as the northern European countries, and the high diversity in southern  
381 Europe. However, within the national scale, the number of species captured in different regions  
382 would generally be more similar, as is the case in Spain (Punzón et al., 2020).

383

384 Fisheries sensitivity proved to be mostly related to the number of jobs in the extractive sector,  
385 with Galicia being the most sensitive region in this regard at the national level (STECF, 2019a).  
386 The important fishing tradition in Galicia has led to the development of maritime industry that  
387 includes activities such as processing and manufacture of canned fish, crustaceans and molluscs.  
388 The relevance and socio-economic impact of these activities, have made the fishing sector one  
389 of Galicia's main economic drivers. Indeed, employment dependency, either directly as the  
390 number of jobs or indirectly as the percentage of the active population within the fisheries  
391 sector, is the most widely used attribute for exploring fisheries sensitivity (Allison et al., 2009;  
392 Morzaria-Luna et al., 2014; Ding et al., 2017; Wabnitz et al., 2018; Pinnegar et al., 2019).  
393 However, it is also important to notice the high spatial heterogeneity in other factors beyond  
394 employment that contribute to fisheries sensitivity. As an alternative to the number of fishing  
395 boats or licenses (as in Colburn et al., 2006; Morzaria- Luna et al., 2014), which might correlate  
396 with the employment dependency, we considered two attributes related to fleet capacity, i.e.,  
397 the mean age of the fishing vessels and their mean power. Fleet age can be considered a proxy  
398 for efficiency in fishing and fuel consumption. The demersal fishing fleet in Spain is generally  
399 above 20 years mean age, with several Mediterranean regions approaching 40 years mean age,

400 double the life expectancy for this kind of fishing vessel (Knittweis et al., 2016). An old fishing  
401 fleet will suffer from physical deterioration and normal obsolescence, decreasing its capacity to  
402 cope with change because it is technologically outdated. While the fishing fleet in Europe has  
403 been generally aging and reducing its size in the last decades, fishing power is still increasing in  
404 most countries (Rousseau et al., 2019). Fishing power is highly correlated with tonnage and  
405 depicts how much a vessel can fish. While the variability in fleet age is moderate, there is high  
406 variability in fishing power, with few coastal regions having a larger share of more powerful  
407 industrial vessels, both in the Atlantic and Mediterranean (i.e., Basque Country, Catalonia and  
408 Valencia). On the other end, Galicia shows the lowest fishing power, as a large percentage of its  
409 fishing fleet is composed of small-scale fishing vessels (MAPA, 2019). Another important factor  
410 considered in our study was the consumption of fish protein as a measure of nutritional  
411 dependence. The Atlantic regions presented higher results than Mediterranean ones.

412

413 Regarding species sensitivity, the Atlantic-Mediterranean pattern was not as evident. Species  
414 sensitivity was the least variable index, pointing to the crucial role of socio-economic differences  
415 in explaining fisheries vulnerability patterns among regions. While about 15% of species in both  
416 Atlantic and Mediterranean regions showed high sensitivity ( $>0.6$ ), their contribution to the  
417 overall community sensitivity was low, as they were not among the most abundant species  
418 captured by demersal fisheries. Species sensitivity is closely associated with biological and  
419 ecological species traits pertaining to their distribution and life-history (Musick, 1999), which  
420 was relatively similar between species captured in the Atlantic and Mediterranean. Indeed, the  
421 species captured in higher abundances ranked as having low or moderate sensitivities, which  
422 reduced species sensitivity index variability among regions. In contrast, to the most used  
423 temperature statistics (i.e., maximum preferred temperature ranges [TP90]) (as in Sunday et al.,  
424 2015; Hare et al., 2016; Pinnegar et al., 2019), we combined mean and temperature range  
425 attributes in an easily computed and broadly applicable indicator of temperature sensitivity. This

426 allowed us to identify species-specific sensitivities within the Atlantic and Mediterranean  
427 regions, though not differences between regions. It is important to highlight the difference in  
428 species-specific responses to climate change impacts on the marine environment. This  
429 reinforces the need for regional studies linking species and climate change all over different  
430 ontogenetic stages in order to produce better and more efficient management plans (Catalán et  
431 al., 2019; Holsman et al. 2019, Twinaime et al., 2020). In addition to the most widely used  
432 distribution and phenology parameters such as the depth range and spawning time, which are  
433 important as resilience indicators (Sunday et al., 2015; Hare et al., 2016), we also estimated the  
434 stability of the stocks through the CV of their landings. Together with landing stability, the price  
435 analysis emerged as a tool to highlight the pressure suffered by each fish stock, considering that  
436 the species with the highest commercial value are therefore subject to higher exploitation rates  
437 and will be more vulnerable to climate change risks (Pinnegar et al., 2019; Hiddink et al., 2019).  
438 The Mediterranean area has shown itself to be slightly more sensitive, and this could be related  
439 to the market characteristics of some of its regions (i.e., the Balearic Islands), which operate in  
440 a market that is more local and therefore more restricted compared to the Atlantic region,  
441 influencing its ability to compete in terms of prices and product varieties.

442

443 Another important factor that marked the difference in the vulnerability index between Atlantic  
444 and Mediterranean areas was adaptive capacity, driven partially by lower levels of landings of  
445 species of low commercial value in the Mediterranean regions. This could be explained due to  
446 the combination of a lower availability of these species and the fact that they are mostly  
447 consumed as a proximity fisheries product, as opposed to what occurs in the Atlantic, where  
448 there is a higher availability of these species, a higher contribution of catches from other  
449 European fishing areas and a higher contribution and supply to the national market (STECF,  
450 2019a). A larger catch of low commercial value species can be related to a greater capacity to  
451 adapt, since the fishery is less dependent on a small group of high commercial value species that



452 are under greater fishing pressure. The capture of species of lower commercial value emerges  
453 as an alternative when the species of higher value may suffer from the impacts of climate change  
454 such as change in distribution and changes in biological patterns. Other indicator that provides  
455 flexibility is the possibility of landing part of the catch in different ports, which allows the fleet  
456 to adapt in the face of difficulties and impacts associated with climate change, such as adverse  
457 weather (e.g., extreme events), but also providing the possibility of taking advantage of better  
458 market opportunities. The combination of these two indicators also explains the lower adaptive  
459 capacity in Mediterranean area and particularly the Balearic Islands region, indicating a great  
460 dependency of a local and more constrained market (UE, 2011). The transition of fishing crews  
461 from large-scale fishing to artisanal fishing and the recreational fishing sector has also been  
462 considered an important indicator of adaptation capacity. In this context, the Atlantic region of  
463 Galicia shows an adaptative advantage over the other regions with a much larger artisanal  
464 fishing fleet that provides more opportunities for the workers to move around within the fishing  
465 sector in order to guarantee their profits and livelihood (Gordoa et al., 2019). Note than some  
466 indicators could benefit from scaling up by making their values relative to size of other indicators  
467 improving the perception of the size to the opportunity (e.g. ratio between recreational and  
468 small-scale fishery). However, with the methodological approach taken, this would increase  
469 correlation with other indicators affecting the interpretation of our results.

470

471 The overall vulnerability index showed a clear pattern between the Atlantic and Mediterranean  
472 areas and regions within these two areas. The results for each region are unique in the  
473 combination of their dependence on the fishing sector, socio-economic development, and  
474 exposure to climate risks regionally. Nevertheless, some regions demand more specific studies  
475 due to their biogeographic complexity. For example, Andalusia presents a more pronounced  
476 geographical heterogeneity than other regions with part of its territory on the Mediterranean  
477 and part on the Atlantic, calling for more specific vulnerability studies for this region which

478 consider its uniqueness. Using demersal fishing on the Spanish coast as an example, the present  
479 study highlights the importance of regional scale analyses to achieve more refined diagnoses in  
480 climate vulnerability assessments. These studies may be instrumental in supporting decision-  
481 making at both national and international levels, as the design of efficient adaptation  
482 management strategies requires cross-scale risks (including exposure, sensitivities and  
483 differences in adaptive capacities) (Holsman et al., 2020). Future analyses should be conducted  
484 to explore the complexity of natural and socio-economic systems, their interactions and their  
485 trade-offs. This shows that customised fisheries adaptation planning is urgently needed at the  
486 regional level, given that the expected large-scale policies may limit flexibility and compromise  
487 their effectiveness (Holsman et al., 2019, 2020). Management results can be more effective  
488 when they adopt dynamic measures that locally consider social and environmental variability  
489 (Levin et al., 2013). Besides the possibility of not achieving the expected results, the adoption of  
490 large-scale adaptation planning can lead to a lack of confidence in fisheries management even  
491 within well-managed systems (Levin et al., 2013; Mumby et al., 2017). While the spatial scale for  
492 CVA should minimise the complexity and variability (spatial heterogeneity) in the socio-  
493 ecological system within its analysis unit, it must be relevant and operational for management  
494 purposes. With this in mind, our study calls for a more detailed consideration of the intra-  
495 national vulnerabilities in other countries to reveal additional and important socio-ecological  
496 sources of fisheries vulnerability to climate change.

497

#### 498 *Acknowledgements*

499 This study was supported by the VADAPES project funded with the support of the  
500 *Biodiversity Foundation* of the Spanish Ministry for the Ecological Transition and the  
501 Demographic Challenge. GA, LLL and MH acknowledge also funding from PANDORA  
502 project (H2020, grant agreement 773713). All authors additionally acknowledge funding  
503 from the COCOCHA project (PID2019-110282RA-I00, Spanish Ministry of Science and

504 Innovation). The authors also acknowledge the development and maintenance of online  
505 repositories and databases such as COPERNICUS, EMODNET, INE, STECF,  
506 AQUAMAPS, FISHBASE or SEALIFEBASE, and their fundamental role in advancing  
507 knowledge in marine and fisheries science. Finally, we thank all the dedicated people,  
508 who have worked on the research vessel surveys of the Spanish Institute of Oceanography  
509 (IEO) included in the study.

510

## 511 **References**

512

513 Adloff, F., Somot, S., Sevault, F., Jordà, G., Aznar, R., Déqué, M., ... & Alvarez-Fanjul, E. 2015.  
514 Mediterranean Sea response to climate change in an ensemble of twenty first century  
515 scenarios. *Climate Dynamics*, 45:2775-2802.

516

517 Allison, E. H., Perry, A. L., Badjeck, M. C., Neil Adger, W., Brown, K., Conway, D. & Dulvy, N. K. 2009.  
518 Vulnerability of national economies to the impacts of climate change on fisheries. *Fish and*  
519 *fisheries*, 10: 173-196.

520

521 Aznar, R., Padorno, M. E., Pérez, B., Gómez Lahoz, M., García Sotillos, M., Álvarez Fanjul, E., & Gomis,  
522 D. 2016. Vulnerability of Spanish ports to climate change Vol. 1: Trends in physical oceanic and  
523 atmospheric variables over the last decades and projections for the 21st century.  
524 <http://hdl.handle.net/20.500.11765/8809>

525

526 Barange, M., Bahri, T., Beveridge, M. C. M., Cochrane, K. L., Funge-Smith, S., and Poulain, F. (eds) 2018.  
527 Impacts of climate change on fisheries and aquaculture: synthesis of current knowledge, adaptation  
528 and mitigation options. *FAO Fisheries and Aquaculture. Technical Paper*, 627. Rome, FAO.

529

530 Barnes, M. L., Wang, P., Cinner, J. E., Graham, N. A., Guerrero, A. M., Jasny, L., ... & Zamborain-Mason, J.  
531 2020. Social determinants of adaptive and transformative responses to climate change. *Nature*  
532 *Climate Change*, 10: 1-6.

533

534 Belkin, I. M. 2009. Rapid warming of large marine ecosystems. *Progress in Oceanography*, 81: 207-  
535 213.

536

537 Catalán, I. A., Auch, D., Kamermans, P., Morales-Nin, B., Angelopoulos, N. V., Reglero, P., ... & Peck, M.  
538 A. 2019. Critically examining the knowledge base required to mechanistically project climate  
539 impacts: A case study of Europe's fish and shellfish. *Fish and Fisheries*, 20: 501-517.

540

541 Cinner, J. E., Mc Clanahan, T. R., Graham, N. A., Daw, T. M., Maina, J., Stead, S. M., ... & Bodin, Ö. 2012.  
542 Vulnerability of coastal communities to key impacts of climate change on coral reef fisheries. *Global*  
543 *Environmental Change*, 22: 12-20.

544

545 Colburn, L. L., Jepson, M., Weng, C., Seara, T., Weiss, J., & Hare, J. A. 2016. Indicators of climate change  
546 and social vulnerability in fishing dependent communities along the Eastern and Gulf Coasts of the  
547 United States. *Marine Policy*, 74: 323-333.

548

549 Darmaraki, S., Somot, S., Sevault, F., Nabat, P., Narvaez, W. D. C., Cavicchia, L. & Sein, D. V. (2019).  
550 Future evolution of marine heatwaves in the Mediterranean Sea. *Climate Dynamics*, 53: 1371-1392.

551

- 552 Ding, Q., Chen, X., Hilborn, R., & Chen, Y. 2017. Vulnerability to impacts of climate change on marine  
553 fisheries and food security. *Marine Policy*, 83: 55-61.  
554
- 555 FAO. 2018. *The State of Mediterranean and Black Sea Fisheries*. General Fisheries Commission for  
556 the Mediterranean. Rome. 172 pp.
- 557 FAO. 2020. *The State of World Fisheries and Aquaculture 2020. Sustainability in action*. Rome.  
558 <https://doi.org/10.4060/ca9229en>
- 559 Free, C. M., Thorson, J. T., Pinsky, M. L., Oken, K. L., Wiedenmann, J., & Jensen, O. P. 2019. Impacts of  
560 historical warming on marine fisheries production. *Science*, 363: 979-983.
- 561 Füssel, H. M., & Klein, R. J. 2006. Climate change vulnerability assessments: an evolution of conceptual  
562 thinking. *Climatic change*, 75: 301-329.
- 563 Gattuso, J.-P., Magnan, A., Bille, R., Cheung, W. W. L., Howes, E. L., Joos, F., Allemand, D. et al. 2015.  
564 Contrasting futures for ocean and society from different anthropogenic CO2 emissions scenarios.  
565 *Science*, 349: 1-10.
- 566 GFCM. 2019. *Working Group on Stock Assessment of Demersal Species*. Final report. Scientific  
567 Advisory Committee on Fisheries. FAO, Rome, Italy, 13-18 November 2019.  
568 <http://www.fao.org/gfcm/technical-meetings/detail/en/c/1274921/>  
569
- 570 Gordo, A., Dedeu, A. L., & Boada, J. 2019. Recreational fishing in Spain: First national estimates of  
571 fisher population size, fishing activity and fisher social profile. *Fisheries Research*, 211: 1-12.  
572
- 573 Haddad, B. M. 2005. Ranking the adaptive capacity of nations to climate change when socio-political  
574 goals are explicit. *Global Environmental Change*, 15: 165-176.  
575
- 576 Hare JA, Morrison WE, Nelson MW, Stachura MM, Teeters EJ, Griffis RB, et al. (2016) A Vulnerability  
577 Assessment of Fish and Invertebrates to Climate Change on the Northeast U.S. Continental Shelf. *PLoS*  
578 *one* 11: e0146756. <https://doi.org/10.1371/journal.pone.0146756>
- 579 Hidalgo, M., Kaplan, D. M., Kerr, L. A., Watson, J. R., Paris, C. B., & Browman, H. I. 2017. Advancing the  
580 link between ocean connectivity, ecological function and management challenges. *ICES Journal of*  
581 *Marine Science*, 74: 1702-1707.
- 582 Hidalgo, M., Mihneva, V., Vasconcellos, M., & Bernal, M. 2018. Climate change impacts, vulnerabilities  
583 and adaptations: Mediterranean Sea and the Black Sea marine fisheries. *Impacts of climate change*  
584 *on fisheries and aquaculture*, 139 pp.  
585
- 586 Hiddink, J. G., Jennings, S., Sciberras, M., Bolam, S. G., Cambiè, G., McConnaughey, R. A., ... & Parma, A.  
587 M. 2019. Assessing bottom trawling impacts based on the longevity of benthic invertebrates. *Journal*  
588 *of Applied Ecology*, 56: 1075-1084.
- 589 Hollowed, A. B., Barange, M., Beamish, R. J., Brander, K., Cochrane, K., Drinkwater, K., Foreman, M. G.  
590 G. et al. 2013. Projected impacts of climate change on marine fish and fisheries. *ICES Journal of Marine*  
591 *Science*, 70: 1023-1037.
- 592 Holsman, K. K., Hazen, E. L., Haynie, A., Gourguet, S., Hollowed, A., Bograd, S. J. & Aydin, K. 2019.  
593 *Towards climate resiliency in fisheries management*. *ICES Journal of Marine Science*, 76: 1368-1378.
- 594 Holsman, K. K., Haynie, A. C., Hollowed, A. B., Reum, J. C. P., Aydin, K., Hermann, A. J. & Punt, A. E. 2020.  
595 *Ecosystem-based fisheries management forestalls climate-driven collapse*. *Nature*  
596 *communications*, 11: 1-10.

- 597 IPCC. 2001. *Climate Change 2001: Impacts, Adaptation, and Vulnerability. Contribution of Working*  
598 *Group II to the Third Assessment Report of the IPCC.* Cambridge University Press, Cambridge, UK.  
599 1032 pp.
- 600 IPCC. 2007. *Climate Change 2007: Impacts, Adaptation and Vulnerability. Contribution of Working*  
601 *Group II to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change,* M.L.  
602 *Parry, O.F. Canziani, J.P. Palutikof, P.J. van der Linden and C.E. Hanson, Eds.,* Cambridge University  
603 *Press, Cambridge, UK, 976pp.*
- 604 IPCC. 2014a. *Climate Change 2014: Impacts, Adaptation, and Vulnerability. Part A: Global and*  
605 *Sectoral Aspects. Contribution of Working Group II to the Fifth Assessment Report of the*  
606 *Intergovernmental Panel on Climate Change.* Ed. by C. B. Field, V. R. Barros, D. J. Dokken, K. J. Mach,  
607 *M. D. Mastrandrea, T. E. Bilir, M. Chatterjee. et al.* Cambridge University Press, Cambridge, UK and  
608 *New York, New York. 1132 pp.*
- 609 IPCC, 2014b. *Climate Change 2014: Impacts, Adaptation, and Vulnerability. Part B: Regional Aspects.*  
610 *Contribution of Working Group II to the Fifth Assessment Report of the Intergovernmental Panel on*  
611 *Climate Change [Barros, V.R., C.B. Field, D.J. Dokken, M.D. Mastrandrea, K.J. Mach, T.E. Bilir, M.*  
612 *Chatterjee, K.L. Ebi, Y.O. Estrada, R.C. Genova, B. Girma, E.S. Kissel, A.N. Levy, S. MacCracken, P.R.*  
613 *Mastrandrea, and L.L. White (eds.)].* Cambridge University Press, Cambridge, United Kingdom and  
614 *New York, NY, USA, 688 pp.*
- 615 Johnson, J. E., Welch, D. J., Maynard, J. A., Bell, J. D., Pecl, G., Robins, J., & Saunders, T. 2016. *Assessing*  
616 *and reducing vulnerability to climate change: Moving from theory to practical decision-*  
617 *support.* *Marine Policy*, 74: 220-229.  
618
- 619 Knittweis, L., Carvalho, N., & Casey, J. 2016. *Assessment of balance indicators for key fleet segments*  
620 *and review of national reports on member states efforts to achieve balance between fleet capacity*  
621 *and fishing opportunities (STECF-16-18).*  
622 <https://www.um.edu.mt/library/oar/handle/123456789/26251>
- 623 Levin, S., Xepapadeas, T., Crépin, A.-S., Norberg, J., de Zeeuw, A., Folke, C., Hughes, T. et al. 2013. *Social-*  
624 *ecological systems as complex adaptive systems: modelling and policy implications.* *Environment*  
625 *and Development Economics*, 18: 111-132.
- 626 MAPA. 2019. *The Spanish fleet, situation on 31 December 2019.* Secretariat-General for Fisheries,  
627 *Directorate-General for Fisheries Management and Aquaculture, Sub-Directorate-General for*  
628 *Competitiveness and Social Affairs, Madrid, Spain. 14 pp.*  
629
- 630 Maxwell, S. L., Butt, N., Maron, M., McAlpine, C. A., Chapman, S., Ullmann, A. & Watson, J. E. 2019.  
631 *Conservation implications of ecological responses to extreme weather and climate events.* *Diversity*  
632 *and Distributions*, 25: 613-625.
- 633 MedECC. 2020. *Climate and Environmental Change in the Mediterranean Basin – Current Situation*  
634 *and Risks for the Future. First Mediterranean Assessment Report [Cramer, W., Guiot, J., Marini, K.*  
635 *(eds.)] Union for the Mediterranean, Plan Bleu, UNEP/MAP, Marseille, France, 600pp, in press*
- 636 Metcalf, S. J., van Putten, E. I., Frusher, S., Marshall, N. A., Tull, M., Caputi, N. & Pecl, G. T. 2015.  
637 *Measuring the vulnerability of marine social-ecological systems: a prerequisite for the identification*  
638 *of climate change adaptations.* *Ecology and Society*, 20: 35.  
639
- 640 Moore, J. K., Fu, W., Primeau, F., Britten, G. L., Lindsay, K., Long, M. & Randerson, J. T. 2018. *Sustained*  
641 *climate warming drives declining marine biological productivity.* *Science*, 359: 1139-1143.  
642
- 643 Morzaria-Luna, H. N., Turk-Boyer, P., & Moreno-Baez, M. 2014. *Social indicators of vulnerability for*  
644 *fishing communities in the Northern Gulf of California, Mexico: implications for climate*  
645 *change.* *Marine Policy*, 45: 182-193.  
646

- 647 Mumby, P. J., Sanchirico, J. N., Broad, K., Beck, M. W., Tyedmers, P., Morikawa, M. & Kleypas, J. A. 2017.  
648 Avoiding a crisis of motivation for ocean management under global environmental change. *Global*  
649 *change biology*, 23: 4483-4496.
- 650  
651 Musick, J. A. 1999. Criteria to define extinction risk in marine fishes: the American Fisheries Society  
652 initiative. *Fisheries*, 24: 6-14.
- 653  
654 Noble, I.R., S. Huq, Y.A. Anokhin, J. Carmin, D. Goudou, F.P. Lansigan, B. Osman-Elasha, and A.  
655 Villamizar, 2014. Adaptation needs and options. In: *Climate Change 2014: Impacts, Adaptation, and*  
656 *Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II*  
657 *to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change* [Field, C.B., V.R. Barros,  
658 D.J. Dokken, K.J. Mach, M.D. Mastrandrea, T.E. Bilir, M. Chatterjee, K.L. Ebi, Y.O. Estrada, R.C. Genova,  
659 B. Girma, E.S. Kissel, A.N. Levy, S. MacCracken, P.R. Mastrandrea, and L.L. White (eds.)]. Cambridge  
660 University Press, Cambridge, United Kingdom and New York, NY, USA, pp. 833-868.
- 661  
662 Payne, M. R., Kudahl, M., Engelhard, G. H., Peck, M. A., & Pinnegar, J. K. 2020. Climate risk to European  
663 fisheries and coastal communities. *BioRxiv*. <https://doi.org/10.1101/2020.08.03.234401>
- 664  
665 Sunday, J. M., Pecl, G. T., Frusher, S., Hobday, A. J., Hill, N., Holbrook, N. J. & Watson, R. A. 2015. Species  
666 traits and climate velocity explain geographic range shifts in an ocean-warming hotspot. *Ecology*  
667 *letters*, 18: 944-953.
- 668  
669 Pinnegar, J. K., Engelhard, G. H., Norris, N. J., Theophille, D., & Sebastien, R. D. 2019. Assessing  
670 vulnerability and adaptive capacity of the fisheries sector in Dominica: long-term climate change and  
671 catastrophic hurricanes. *ICES Journal of Marine Science*, 76: 1353-1367.
- 672  
673 Pinsky, M. L., Reygondeau, G., Caddell, R., Palacios-Abrantes, J., Spijkers, J., & Cheung, W. W. 2018.  
674 Preparing ocean governance for species on the move. *Science*, 360: 1189-1191.
- 675  
676 Punzón, A., Rueda, L., Rodríguez-Basalo, A., Hidalgo, M., Oliver, P., Castro, J. & Massutí, E. 2020. History  
677 of the Spanish demersal fishery in the Atlantic and Mediterranean Seas. *ICES Journal of Marine*  
*Science*, 77: 553-566.
- 678  
679 Rousseau, Y., Watson, R. A., Blanchard, J. L., & Fulton, E. A. (2019). Evolution of global marine fishing  
680 fleets and the response of fished resources. *Proceedings of the National Academy of Sciences*, 116:  
681 12238-12243.
- 682  
683 Ser-Giacomi, E., Sánchez, G. J., Soto-Navarro, J., Thomsen, S., Mignot, J., Sevault, F., & Rossi, V. (2020).  
684 Impact of climate change on surface stirring and transport in the Mediterranean Sea. *Geophysical*  
*Research Letters*, <https://doi.org/10.1029/2020GL089941>
- 685  
686 STECF. 2019a. Scientific, Technical and Economic Committee for Fisheries. The EU Fish Processing  
687 Sector. Economic Report (STECF-19-15). Publications Office of the European Union, Luxembourg,  
2019. 172-180 pp.
- 688  
689 STECF. 2019b. Scientific, Technical and Economic Committee for Fisheries (STECF): The 2019 Annual  
690 Economic Report on the EU Fishing Fleet (STECF 19-06), Carvalho, N., Keatinge, M. and Guillen Garcia,  
691 J. editor(s), EUR 28359 EN, Publications Office of the European Union, Luxembourg, 2019, ISBN 978-  
92- 76-09517-0, doi:10.2760/911768, JRC117567.
- 692  
693 Tol, R. S., & Yohe, G. W. 2007. The weakest link hypothesis for adaptive capacity: an empirical  
test. *Global Environmental Change*, 17: 218-227.
- 694  
695 Turner, B.L., Kasperson, R.E., Matsone, P.A. et al. 2003. A framework for vulnerability analysis in  
696 sustainability science. *Proceedings of the National Academy of Sciences of the United States Of*  
*America* 100: 8074-8079.

- 697 Twiname, S., Audzijonyte, A., Blanchard, J. L., Champion, C., de la Chesnais, T., Fitzgibbon, Q. P., ... &  
698 Oellermann, M. 2020. A cross-scale framework to support a mechanistic understanding and  
699 modelling of marine climate-driven species redistribution, from individuals to  
700 communities. *Ecography*. 15 pp.  
701
- 702 UE 2011 Farnet Guide 3: Adding value to local fisheries and aquaculture products European  
703 Commission, Maritime Affairs and Fisheries. Belgium. 58pp.
- 704 UNFCCC. 2018. United Nations Framework Convention on Climate Change. Climate change annual  
705 report, Luxembourg. 62 pp.
- 706 Vincent, K. 2007. Uncertainty in adaptive capacity and the importance of scale. *Global Environmental*  
707 *Change*, 17: 12-24.
- 708 Wabnitz, C. C., Lam, V. W., Reygondeau, G., Teh, L. C., Al-Abdulrazzak, D., Khalfallah, M., ... & Cheung,  
709 W. W. 2018. Climate change impacts on marine biodiversity, fisheries and society in the Arabian  
710 Gulf. *PloS one*, 13: e0194537.
- 711 Yohe, G. W., Malone, E., Brenkert, A., Schlesinger, M., Meij, H., & Xing, X. 2006. Global distributions of  
712 vulnerability to climate change. *Integrated Assessment*, 6: 35-44.  
713