- 1 Title: Change in terrestrial human footprint drives continued loss of intact ecosystems
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32 Summary

Our ability to map numarity's influence across Earth has evolved, thanks to powerful
computing, a network of earth observing satellites, and new bottom-up census and crowd-
sourced data. Here, we provide the latest temporally inter-comparable maps of the terrestrial
Human Footprint, and assessment of change in human pressure at global, biome, and
ecoregional scales. In 2013, 42% of terrestrial Earth could be considered relatively free of
anthropogenic disturbance, and 25% could be classed as 'wilderness' (the least degraded end
of the human footprint spectrum). Between 2000 and 2013, 1.9 million km^2 - an area the size
of Mexico - of land relatively free of human disturbance became highly modified. The
majority of this occurred within tropical and subtropical grasslands, savannah, and shrubland
ecosystems, but the rainforests of Southeast Asia also underwent rapid modification. Our
results show that humanity's footprint is eroding Earth's last intact ecosystems, and greater
efforts are urgently needed to retain them.

45

Key words: Human pressure, cumulative pressure mapping, ecosystem degradation, human
modification, human footprint, wilderness, wild lands, biodiversity, conservation, land use
change.

49 Introduction

Humans have influenced the terrestrial biosphere for millennia, converting much of Earth's surface to anthropogenic land uses¹. Nevertheless, there are still some ecosystems that remain free from significant human pressure, thereby providing crucial habitats for imperilled species^{2,3} and maintaining the ecosystem processes that underpin planetary life-support systems^{4,5}. As a consequence, calls for the global identification, monitoring, and retention of the remaining lands that are relatively free of direct anthropogenic disturbance are increasing^{6–8}.

Over the past two decades, cumulative pressure maps that combine remotely-sensed 57 data with survey data are being increasingly used to assess the full range of human pressures 58 on land spatially⁹. These advances have facilitated the mapping of Earth's remaining marine 59 and terrestrial wilderness^{8,10,11}, improved measures and estimates of species extinction risk¹², 60 underpinned broader assessments of human impacts on ecosystems¹³ and biodiversity^{14–16}, 61 and enabled the identification of protected areas and world heritage sites in danger^{14,17,18}. The 62 results of these mapping efforts are influencing global policy discussions^{6,19}, and informing 63 on-the-ground decisions about where to undertake biodiversity conservation action²⁰⁻²². 64

Here, we provide the latest global maps of cumulative human pressure 23,24 for the 65 years 2000, 2005, 2010, and 2013, and use them to assess how changes in human pressure are 66 altering Earth's terrestrial ecosystems. We used a human footprint threshold of <4 (on 0-5067 scale) to identify where land is considered ecologically intact (below the threshold) or highly 68 69 modified and thus ecologically degraded (equal to or above the threshold). Areas below this threshold are ecosystems that may be subject to some level of human pressure (for example 70 low-density transitory human populations or pasture lands grazed at a low intensity), but still 71 contain the majority of their natural habitat and ecological processes^{14,25}. This threshold has 72

been found to be robust from a species conservation perspective because once surpassed,
species extinction risk increases dramatically¹², and several ecosystem processes are
altered^{12,16,26}.

76 We assess transitions from intact to highly modified land at global, biome, and ecoregional scales²⁷ and ascertain which nations contain Earth's remaining intact systems, 77 and had the greatest amounts of habitat loss. Previous global assessments of human pressure 78 have attempted to identify at risk ecosystems by determining a 'safe limit' of biodiversity loss 79 for ecosystem functionality 28,29 , assessing protection levels³⁰, and analysing habitat 80 conversion using land cover^{31,32}. But all of these ignore a broad range of threats that occur 81 beyond land use such as accessibility via roads, railways and navigable waterways, human 82 population density, and light pollution. These pressures have environmental impacts well 83 beyond the local development footprint^{33,34,36}. As such, our results provide the latest spatially 84 explicit understanding of the state of human pressure on the natural environment, and how it 85 is changing over time. We show that the human footprint methodology can be continually 86 updated and, when more recent data becomes available, allow for assessment of habitat loss 87 at scales relevant to planning activities. 88

89 **Results**

90 State of terrestrial Earth

91	As of 2013, 55.8 million km ² (41.6%) of Earth's surface was intact (which includes
92	wilderness, human footprint of <4), and 33.5 million km^2 (25.0%) was wilderness (human
93	footprint of <1). The remaining (human footprint of \geq 4) 78.4 million km ² (58.4%) was
94	under moderate or intense human pressure (and therefore highly modified), which was
95	widespread, encompassing over half the area of 11 (or 78.6%) of Earth's 14 biomes
96	(Figure 1). Temperate broadleaf and mixed forests were the most altered biome, with 11.6
97	million km ² (91.0%) being highly modified, followed by tropical and subtropical dry
98	broadleaf forests with 2.72 million km^2 (90.5%), and Mediterranean forests, woodlands
99	and scrubs with 2.88 million km^2 (89.7%). Wilderness areas have all but disappeared in
100	many biomes, for example, only $82,000 \text{ km}^2$ (0.81%) remained in temperate grasslands,
101	savannahs, and shrublands, 29,000 km^2 (0.96%) in tropical and subtropical dry broadleaf
102	forests, and just 12,000 km^2 (1.69%) in tropical and subtropical coniferous forests.
103	Earth's 14 biomes consist of 795 ecoregions, which represent distinct biotic
104	assemblages and abiotic features (such as landforms) at a finer scale than biomes ²⁷ . We
105	found the entire extent of 46 (5.76%) ecoregions were highly modified. These 46
106	ecoregions span 10 biomes, with the majority located in tropical and subtropical moist
107	broadleaf forests ($n = 17, 37.0\%$), tropical and subtropical dry broadleaf forests ($n=6$,
108	13.0%), and temperate broadleaf and mixed forests (n=6, 13.0%). One-quarter of all
109	ecoregions (n=187) have lost all wilderness.

110 The majority of land in tundra, boreal and taiga forests, and deserts and xeric
111 shrubland biomes remains intact. At the ecoregion level, just 52 (6.53%) still have >90% of
112 their land intact, and a mere 21 (2.64%) are >90% wilderness. These ecoregions with >90%

- wilderness are found in just four biomes, tundra (n = 12), boreal forests/taiga (n = 5), tropical
- and subtropical moist broadleaf forests (Rio Negro campinarana and Juruá-Purus moist
- 115 forests), and tropical and subtropical grasslands, savannahs and shrublands (Northwestern
- 116 Hawaii scrub).



Figure 1. The global human footprint map for the year 2013. The surrounding pie charts represent the proportion of each terrestrial biome that was completely free of direct anthropogenic disturbance ('wilderness', dark green, human footprint value of <1), relatively free of direct anthropogenic disturbance ('intact', light green, human footprint value of <4 and \geq 1), or highly impacted by anthropogenic disturbance ('highly modified', red, human footprint value of \geq 4) in the year 2013. Circles sizes represent relative biome area.

104 *Contemporary changes in human pressure*





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Figure 2. Density plot depicting change in the global terrestrial human footprint between the years 2000 and 2013 (n = 134, 154, 306). The x-axis represents the human footprint

value of a pixel in the year 2000, and the y-axis represents the human footprint value of

that pixel in the year 2013. The number of pixels that made that particular transition are

represented by the colour within the plot. Red represents a high number of pixels and blue

represents low. Legend is log-scaled. Between 2000 and 2013, 25,348,514 km² (18.9%) of 120

- 121 pixels deteriorated (human pressure increased), while 7,995,464 km² (5.96%) improved
- 122 (human pressure decreased).



Figure 3. The global human footprint map for the year 2000 (a). Areas completely free of direct anthropogenic disturbance ('wilderness', dark green, human footprint value of <1), relatively free of direct anthropogenic disturbance ('intact', light green, human footprint value of <4 and \geq 1), or highly impacted by anthropogenic disturbance ('highly modified', red, human footprint value of \geq 4). The change between 2000 and 2013 within each 2000 state can be seen for intact land (b) and highly modified land (c), which leads to the 2013 state (d).

Intact lands were lost in all biomes during the assessment period, with the highest loss 123 occurring in tropical and subtropical grassland, savannah and shrublands (655,000 km² was lost 124 representing 11.3% of all intact lands within the biome, an area approximately the size of 125 France; Figure 4). The tropical and subtropical moist broadleaf forests and mangrove biomes 126 also lost substantial areas of intact land (559,000 km², 6.90% and 9,000 km², 14.7% 127 respectively). While the largest absolute loss of intact lands occurred in savannah and woodland 128 ecoregions, the largest proportional losses occurred in tropical forest ecoregion types. For 129 example, intact areas were completely lost in seven forested ecoregions including the Louisiade 130 Archipelago rainforests (Papua New Guinea), and Sumatran freshwater swamp forests 131 (Indonesia; see Supplemental 1). 132

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Figure 4. The proportion of biome that transitioned between wilderness (human footprint value of <1), intact (human footprint value between <4 and \geq 1), and highly modified (human footprint value of \geq 4) states between 2000 and 2013, represented by circles. If part of the biome transitions to a worsened condition it moves upwards into the red area, if part of the biome does not transition it remains on the diagonal, and if part of the biome improves, it moves downwards into the blue area. For exact values see Supplemental 1.

142 The largest losses of wilderness between 2000 and 2013 occurred in biomes that contained the largest areas of wilderness in 2000. For example, deserts and xeric shrublands 143 lost 426,000 km² (5.08%) of their remaining wilderness. This was concentrated in desert, 144 woodland and savannah ecoregion types (see Supplemental 1). Wilderness in the tundra and 145 boreal/taiga forests suffered the most extreme transitions, with 22,000 km² and 15,000 km² 146 respectively changing from wilderness to highly modified land (human footprint <1 to >4) 147 (Figure 4). The ecoregions of the Russian tundra and taiga lost the most wilderness, for 148 example, the Yamal-Gydan tundra lost 8,000 km², and the East Siberian taiga lost 5,000 km². 149

150 National responsibility

151 In 2013, only 26 nations (out of 221) had the majority (>50%) of their land intact.

152 Excluding island territories, the two countries with the highest proportion of intact land

included Guyana (88.8% of country; 187,000 km²) and Suriname (88.5%; 125,000 km²).

154 The African continent contained 11 ecoregions that lost the largest areas of intact land.

155 Between 2000 and 2013, more intact land was lost in the Democratic Republic of the

156 Congo (DRC) than any other country $(316,000 \text{ km}^2; 13.6\% \text{ of the country}; 37.3\% \text{ of its})$

intact lands). This was followed by Indonesia and Brazil which lost 122,000 km² (6.98% of

the country or 20.2% of its intact lands) and 87,000 km^2 (29% of the country or 1.88% of

159 its intact lands) respectively.

Russia, Canada, Brazil, and Australia are responsible for the largest areas of Earth's
remaining intact areas (which includes wilderness, human footprint score of <4).

162 Combined, these four countries harbour more than 60% of Earth's wilderness (human

- 163 footprint score of <1, Figure 5). Brazil also lost the most wilderness (human footprint
- increasing above 1) of any country (109,880 km², 3.87% of its wilderness area). The
- 165 largest areas of wilderness lost to high levels of human modification (human footprint

increasing above 4) were in Russia (23,000 km²), Canada (10,000 km²), and Brazil (6,000





- 169 Figure 5 Proportion of each country's terrestrial land that was completely free of direct
- anthropogenic disturbance ('wilderness', dark green, human footprint value of <1),
- 171 relatively free of direct anthropogenic disturbance ('intact', light green, human footprint
- value between <4 and ≥ 1), or highly impacted by anthropogenic disturbance ('highly
- 173 modified', red, human footprint value of \geq 4) in the year 2013.

174 Discussion

The terrestrial human footprint presented here is one of the most comprehensive and up-to-175 date measures of cumulative human pressure across Earth, and will be continuously 176 177 improved as more data on the eight included pressures (built environments, population density, night-time lights, crop lands, pasture lands, accessibility via roads, railways, and 178 navigable waterways) become available. While this latest update is already seven years out 179 of date, advances in data generation and modelling³⁵ will facilitate more rapid updates of 180 the human footprint in the near future. Our analyses show that between 2000 and 2013 181 182 substantial areas of intact land, including wilderness areas, have been lost. This loss has profound implications for the biodiversity that require intact land for their continued 183 survival³, and for people who rely on the services that intact ecosystems provide^{8,37}. The 184 185 transition from intact ecosystems to highly modified land is the greatest predictor of why species face increasing extinction risk¹² as this transition is where habitat is considered 186 functionally unavailable for many terrestrial vertebrates^{38,39}. This transition also negatively 187 impacts wildlife population viability, because intact ecosystems are proven strongholds for 188 genetic diversity⁴⁰. Climate change mitigation efforts are also undermined by these losses 189 because intact lands make crucial contributions to the residual terrestrial carbon sink^{37,41}. 190 For example, a recent study found that carbon impacts of intact forest loss are 626% worse 191 than originally estimated⁴¹. 192

We also demonstrate that patterns of degradation due to increasing human pressure are now changing within biomes. Past studies note that dry forested biomes have suffered the highest rates of habitat loss^{25,31} but our results now show that recent increases in human pressure predominantly occurred in tropical savannah and grassland ecosystems, which lost 11.3% of their intact area between 2000 and 2013. This finding is consistent with previous evidence that savannahs are the current development frontier in many regions

worldwide^{42,43}. Proactive conservation planning is urgently needed to prevent the last 199 intact savannahs, such as Australia's northern savannahs⁴⁴ and Colombia's Llanos in the 200 Orinoquia region⁴³, suffering the same losses that occurred in places such as Brazil's 201 Cerrado⁴⁵. Conservation planning needs to utilise tools that take into account past and 202 future risk, so that preventative conservation action can be implemented in places where 203 development is most likely to occur^{46–48}. Our analysis helps inform where proactive 204 conservation planning activity must occur, and demonstrates the potential of human 205 pressure mapping for informing global conservation action. 206

207 Nearly three decades ago, the world came together to ratify the Rio Conventions, including the Convention on Biological Diversity (CBD), the UN Convention to Combat 208 Desertification (UNCCD), and the UN Framework Convention on Climate Change 209 (UNFCCC). Despite the fact that almost all nations are signatories on these three 210 international environmental agreements, intact habitats continue to be lost at a rapid rate⁴⁹, 211 including within the borders of many signatory nations, such as the DRC, Indonesia, and 212 Brazil. One possible explanation for this trend is the challenge of collectively identifying 213 intact landscapes and then using this information to take coordinated action across the 214 globe to protect them. Given the growing body of scientific evidence demonstrating the 215 exceptional value of intact ecosystems (including wilderness areas) for conserving 216 biodiversity⁵⁰, mitigating climate change⁴¹, and providing essential ecosystem services³⁷, 217 the importance of data on intactness should be elevated when undertaking efforts to 218 219 develop international and national targets and shaping actions under these Conventions. 220 For example, at the end of 2020 nations that are party to the Convention on Biological Diversity will sign off on the Post-2020 Global Biodiversity Framework that will set global 221 targets on nature for the coming decades. Negotiations around the Post-2020 Framework 222

present an opportunity for countries to include targets specifically for the protection and
 complete retention of intact ecosystems⁸.

Halting the loss of intact ecosystems cannot be achieved alongside current trajectories 225 of development, population growth, and resource consumption⁵¹. Retention of Earth's 226 remaining intact lands can only be achieved through a combination of strategic policy mixes 227 that better regulate deleterious activities across all sectors, levels of governance and 228 229 jurisdictions, and on the ground site-based action such as well-resourced protected areas in conjunction with other effective area-based conservation measures (OECMS) such as 230 payment schemes for safeguarding ecosystem services^{51–54}. While many pathways on how 231 intact retention can be achieved are being developed^{8,51,54,55}, the challenge is ensuring action 232 occurs at the scale and speed necessary to ensure all intact ecosystems are secured. 233

The highest losses of intact lands occurred in African nations, where the highest 234 biodiversity impact from future socio-economic development is also predicted to occur⁵⁶. 235 Parts of Africa also have the largest gap between food consumption and production in the 236 world, we can therefore only infer that increasing agricultural production is a key driver of 237 savannah and grassland loss^{57,58}. Other regions experiencing extreme levels of intact 238 ecosystem loss are the rainforests of Indonesia (which covers 1.3% of Earth but contains 239 10% of the world's plants, 12% of mammals, 16% of reptile-amphibians, and 17% of 240 birds⁵⁹) and Papua New Guinea (which covers less than 1% of Earth but contains 5% of its 241 biodiversity⁶⁰). This extreme habitat loss is likely due to the spike in habitat conversion to 242 grow cash crops such as oil palm^{61,62}, driven by international demand⁶³. Thus, research 243 244 must be oriented to understanding these drivers, and subsequently to find mechanisms that facilitate socio-economic development without further degrading intact ecosystems^{51,64}. 245

Conclusion 247

248	We have presented the latest comprehensive assessment of humanity's footprint on
249	terrestrial Earth using the best available data. We find human pressure is extending ever
250	further into the last ecologically intact, and wilderness areas. With important policy
251	discussions on the Convention on Biological Diversity's Post-2020 Global Biodiversity
252	Framework well underway ⁶⁵ , this is a timely opportunity for nations to take stock and to
253	set explicit targets for retaining Earth's remaining intact lands. Proactively protecting
254	Earth's intact ecosystems is humanity's best mechanism for protecting against climate
255	change, ensuring large-scale ecological and evolutionary processes persist, and
256	safeguarding biological diversity into the future.
257	Experimental procedures

Overview 258

We updated the Human Footprint²³ terrestrial cumulative human pressure maps for the years 259 2000, 2005, 2010 and 2013 and used it to define the state of Earth's biomes, ecoregions and 260 261 countries, and their transitions between states between 2000 and 2013. All analyses, and 262 creation of the human footprint maps, were conducted in the Mollweide equal area projection at 1 km² resolution. 263

Updating the human footprint 264

To recreate the human footprint maps we followed broadly the methods developed by 265 Sanderson and colleagues²⁴ and Venter and colleagues²³. Significant areas missing in the 266 original Human Footprint²³ (which carried over into subsequent releases), including 267 Azerbaijan, areas along the western former-USSR border, and along the Orange River in 268 South Africa, among others, have been included in this update. We used data on human 269 270 pressures across the periods 2000 to 2013 to map: 1) the extent of built human environments, 271 2) population density, 3) electric infrastructure, 4) crop lands, 5) pasture lands, 6) roadways, 7) railways, and 8) navigable waterways. To facilitate comparison across pressures we placed 272 each human pressure within a 0-10 scale, weighted within that range according to estimates 273 of their relative levels of human pressure following Sanderson and colleagues²⁴. The resulting 274 standardized pressures were then summed together to create the standardized human footprint 275 maps for all non-Antarctic land areas. Pressures are not intended to be mutually exclusive, 276 and many will co-occur in the same location. Three pressures only had data from a single 277 time period or have poorly annotated temporal information, and these are treated as static in 278 279 the human footprint maps.

We used free and open-source GRASS GIS 7.2.2⁶⁶ to create a series of scripts that 280 integrate the spatial data on human pressures, yielding 134,064,303 pixels for Earth's 281 282 terrestrial surface (excluding Antarctica). For any grid cell, the human footprint can range between 0–50. We carried out a validation of the human footprint map using visual 283 interpretation of high resolution imagery across 3114×1 km² sample plots randomly located 284 across the Earth's non-Antarctic land areas. We found strong agreement between the human 285 footprint measure of pressure and pressures scored by visual interpretation of high resolution 286 287 imagery, with a root mean squared error (RMSE) of 0.116 and a Kappa statistic of 0.806 (P \leq 0.01). For further details on the validation exercise see Supplemental 2. The following 288 289 sections (and Table S1) describe in detail the source data for each pressure, the processing 290 steps applied, and the rationale behind the pressure weighting. The code and underlying data for generating these maps is available online at 291

<u>https://github.com/scabecks/humanfootprint_2000-2013</u>, and can be used to easily regenerate
 them with updated or alternate datasets, as well as to apply the same methodology at national
 or regional scales.

295 Built environments

296 Built environments, in the context of the human footprint, are anthropogenic areas that represent urban settings, including buildings, paved land and urban parks. These 297 environments do not provide viable habitats for many species of conservation concern, nor do 298 they provide high levels of ecosystem services $^{67-70}$. As such, built environments were 299 assigned a pressure score of 10. 300 To map built environments, we used the Defence Meteorological Satellite Program 301 302 Operational Line Scanner (DMSP-OLS) composite images which gives the annual average brightness of 30 arc second (~1 km at the equator) pixels in units of digital numbers 303 (DN)^{71,72}. This data was collected from six different satellite missions over the period 1992 to 304

2013. We extracted data for the years 2000, 2005, 2010, and 2013, and all datasets were then inter-calibrated to facilitate comparison⁷¹. Using the DMSP-OLS datasets, we considered pixels to be 'built' if they exhibited a calibrated DN greater than 20. This threshold is based on a global analysis of the implications of a range of thresholds for mapped extent of cities⁷³, and visual validation against Landsat imagery for 10 cities spread globally.

The DMSP-OLS has limitations for the purpose of mapping human settlements, including hyper sensitivity of the sensors causing detection of over-glow adjacent to built environments⁷³ and bright lights associated with gas flaring from oil production facilities⁷⁴. However, no other data exist to map built environments in a consistent way globally over our time horizon. While more recent satellite platforms launches – such as VIIRS – offer higher spatial resolution and greater light sensitivity⁷⁵ than DMSP-OLS, they aren't presently comparable or integrated across the temporal range we required.

317 **Population density**

The intensity of human pressure on the environment is often associated with proximity tohuman populations, such as human disturbance, hunting and the persecution of non-desired

- species⁷⁶. Even low-density human populations with limited technology and development can 320 have significant impacts on biodiversity^{77,78}. 321
- We incorporated human population density using the Gridded Population of the 322 323 World dataset developed by the Centre for International Earth Science Information Network (CIESEN)⁷⁹. The dataset provides a 1 km² gridded summary of population census data for the 324 years 2000, 2005, 2010, and 2013. We used linearly interpolated densities for year 2013 from 325 data for years 2010 and 2015. For all locations with more than 1000 people km^{-2} , we 326 assigned a pressure score of 10. For more sparsely populated areas with densities lower than 327 1000 people km⁻², we logarithmically scaled the pressure score using, 328 329 Pressure score = $3.333 \times \log (\text{population density} + 1) (1)$ 330 Human population density is scored in this way under the assumption that the pressures 331
- people induce on their local natural systems increase logarithmically with increasing 332

population density, and saturate at a level of 1000 people km⁻². 333

Night-time lights 334

340

The high sensitivity of the DMSP-OLS⁷² dataset provides a means for mapping the sparser 335 electric infrastructure typical of more rural and suburban areas. In 2009, 79% of the lights 336 337 registered in the DMSP-OLS dataset had a Digital Number less than 20, and are therefore not included in our 'built environments' layers. However, these lower DN values are often 338 important human infrastructures, such as rural housing or working landscapes, with 339 associated pressures on natural environments.

To include these pressures, we used the inter-calibrated DMSP-OLS layers^{71,72,80} used 341 for the built environments mapping. The 2013 calibration parameters were conveyed through 342 personal communications from the creators of the dataset, and are not yet published. The 343 equations for inter-calibrating across years are second order quadratics trained using data 344

345 from Sicily, which was chosen as it had negligible infrastructure change over this period and where DN average roughly 14⁷². For our purposes, DN values of six or less where excluded 346 from consideration prior to calibration of data, as the shape of the quadratic function leads to 347 348 severe distortion of very low DN values. The inter-calibrated DN data from 2000 were then rescaled using an equal quantile approach into a 0-10 scale. To scale the data, we divided the 349 calibrated night light data into 10 equal sample bins (each bin with a DN greater than 1 350 contains the same number of pixels) based on the DN values and then assigned them scores 351 of 1 through 10, starting with the lowest DN bin. DN values of 0 were assigned a score of 0. 352 353 The thresholds used to bin the 2000 data where then used to convert the 2005, 2010, and 2013 data into a comparable 0-10 scale. 354

355 Crop and pasture lands

Crop lands vary in their structure from intensely managed monocultures receiving high inputs of pesticides and fertilizers, to mosaic agricultures such as slash and burn methods that can support intermediate levels of natural values^{82,84}. For the purposes of the human footprint, we focused only on intensive agriculture because of its greater direct pressure on the environment, as well as to circumvent the shortcomings of using remotely sensed data to map mosaic agriculture globally, namely the tendency to confound agriculture mosaics with natural woodland and savannah ecosystems⁸⁶.

363 Spatial data on remotely sensed agriculture extent were extracted from the MERIS 364 CCI Landcover annual dataset ⁸¹. Although intensive agriculture often results in whole-scale 365 ecosystem conversion, we gave it a pressure score of 7, which is lower than built 366 environments because of their less impervious cover.

Pasture lands cover 22% of the Earth's land base or almost twice that of agricultural
 crop⁸³, making them the most extensive direct human pressure on the environment. Land

grazed by domesticated herbivores is often degraded through a combination of fencing,
intensive browsing, soil compaction, invasive grasses and other species, and altered fire
regimes⁸⁸. We mapped grazing lands for the year 2000 using a spatial dataset that combines
agricultural census data with satellite derived land cover to map pasture extent⁸³. We assigned
pasture a pressure score of 4, which was then scaled from 0–4 using the percent pasture for
each 1 km² pixel.

375 Roads and railways

As one of humanity's most prolific linear infrastructures, roads are an important direct driver of habitat conversion⁸⁹. Beyond simply reducing the extent of suitable habitat, roads can act as population sinks for many species through traffic induced mortality⁹⁰. Roads also fragment otherwise contiguous blocks of habitat, and create edge effects such as reduced humidity⁹¹ and increased fire frequency that reach well beyond the roads immediate footprint⁹². Finally, roads provide conduits for humans to access nature, bringing hunters and nature users into otherwise wilderness locations⁹³.

Data from OpenStreetMaps (OSM) on roads and railways was extracted from the 383 global OSM planet database⁸⁵. We include all categories of tagged highway in the OSM 384 planet database. OSM is a volunteer driven, open-source global mapping project that has 385 grown enormously in spatial completeness since its inception in 2004⁹⁴. The volume and 386 coverage of global transportation networks in the OSM database has far surpassed previously 387 available roads data (e.g., gRoads⁹⁵) which was used in earlier iterations of the Human 388 Footprint²³; however, the OSM dataset still does not provide full coverage outside of urban 389 areas in some global regions, notably in central Africa, at the time of data extraction. 390 Therefore, to benefit both from the larger OSM database while maintaining road coverages in 391 regions that are currently poorly mapped in OSM, we merged the OSM data with gRoads 392

data. The merged dataset performed best globally when we validated the three data layers(gRoads only, OSM only, and the union of gRoads/OSM).

We mapped the direct and indirect influence of roads by assigning a pressure score of 395 396 8 for 0.5 km out for either side of roads, and access pressures were awarded a score of 4 at 0.5 km and decaying exponentially out to 15 km either side of the road. While railways are an 397 important component of our global transport system, their pressure on the environment 398 399 differs in nature from that of our road networks. By modifying a linear swath of habitat, railways exert direct pressure where they are constructed, similar to roads. However, as 400 passengers seldom disembark from trains in places other than rail stations, railways do not 401 402 provide a means of accessing the natural environments along their borders. The direct pressure of railways where assigned a pressure score of 8 for a distance of 0.5 km on either 403 side of the railway. We exclude railways tagged as abandoned or disused. 404

Importantly, neither gRoads nor OSM datasets provide true and comprehensive
temporal information (gRoads not at all); as such both datasats were used in their most up-todate version in all time periods considered.

408 Navigable waterways

Like roads, coastlines and navigable rivers act as conduits for people to access nature. While all coastlines are theoretically navigable, for the purposes of the human footprint we only considered coasts⁹⁶ as navigable for 80 km either direction of signs of a human settlement, which were mapped as a night lights signal with a DN⁷² greater than 6 within 4 km of the coast. We chose 80 km as an approximation of the distance a vessel can travel and return during daylight hours. As new settlements can arise to make new sections of coast navigable, coastal layers were generated for the years 2000, 2005, 2010, and 2013.

416	Large lakes can act essentially as inland seas, with their coasts frequently plied by				
417	trade and harvest vessels. Based on their size and visually identified shipping traffic and				
418	shore side settlements, we treated the great lakes of North America, Lake Nicaragua, Lake				
419	Titicaca in South America, Lakes Onega and Peipus in Russia, Lakes Balkash and Issyk Kul				
420	in Kazakhstan, and Lakes Victoria, Tanganyika and Malawi in Africa as we did navigable				
421	marine coasts.				
422	Rivers were considered as navigable if their depth was greater than 2 m and there				
423	were signs of night-time lights (DN>=6) within 4 km of their banks, or if contiguous with a				
424	navigable coast or large inland lake, and then for a distance of 80 km or until stream depth is				
425	likely to prevent boat traffic. To map rivers and their depth we used the hydrosheds				
426	(hydrological data and maps based on shuttle elevation derivatives at multiple				
427	scales) ⁸⁷ dataset on stream discharge, and the following formulae ^{97,98} :				
	stream width = $8.1 \times (discharge[m^3/s])0.58$ (2)				
428	and				
	velocity = $4.0 \times (discharge[m^3/s])0.6/(width[m]) (3)$				
429	and				
	cross-sectional area = discharge/velocity (4)				
430	and				
431	depth = $1.5 \times \text{area/width}(5)$				
432	Assuming second order parabola as channel shape.				
433	Navigable rivers layers were created for the years 2000, 2005, 2010, and 2013, and				

434 combined with the navigable coasts and inland seas layers for the same years to create the

final navigable waterways layers. The access pressure from navigable water bodies were

awarded a score of 4 adjacent to the water body, decaying exponentially out to 15 km.

437 Defining low-pressure areas and wilderness

We defined intact areas with low human pressure as a human footprint value of <4, and the areas of high human pressure, or 'damaged' areas, as \geq 4. This value of \geq 4 equates to a human pressure score equal to pasture lands, representing a reasonable approximation of when anthropogenic land conversion has occurred to an extent that the land can be considered human-dominated and no longer 'natural'. This threshold, which is considered significant at the landscape level²⁵, is also the point where species are far more likely to be threatened by habitat loss¹².

Within the intact state, we defined areas that are pressure-free, or 'wilderness', as a human footprint value of <1 following previous global wilderness assessments¹⁰. We defined wilderness because it increasingly holds special importance in global policy dialogue, as they contain the highest densities of Earth's biomass, remaining intact mega-faunal assemblages, provide life-supporting ecosystem services, act as controls against which to measure planetary health, provide the last strongholds for many of the world's languages and have spiritual and cultural value for many of the world's people of many religions^{7,8,99,100}.

452 Units of analysis

Biomes and ecoregions are ecologically distinct geographical units that reflect the
distributions of a broad range of fauna and flora across the entire planet ²⁷. These entities are
now critical for policy and decision makers, being considered core units of reporting in global
treaties, and as such can direct legislation, management and conservation efforts towards
crisis locations and ecosystems^{6,30,31,101,102}. We use biomes and ecoregions described by
Olson and colleagues in 2001²⁷ to define terrestrial biomes and ecoregions, excluding Lakes

and Rock and Ice. We excluded ecoregions that either fell within the Lakes, Rock and Ice
biomes or were not covered by the human footprint. World borders were described by
Sandvik 2009¹⁰³, both datasets are freely downloadable.

462 Assessing human footprint change

We calculated transitions in levels of human pressure by first assessing human footprint 463 scores for the year 2000, then identified pixels that had changed to a different intensity 464 through to the year 2013. We assume that once a pixel has moved from a score of 0 (a 465 wilderness state), it cannot return to this condition as by definition, once transformed an area 466 is no longer wilderness^{8,104}. Therefore, any pixel that was <1 in 2013, but greater than 1 in 467 any other year was given a value of 1 so that it is considered intact land rather than 468 wilderness. All other comparisons directly report changes between 2000 and 2013, including 469 470 positive changes when a pixel has a lower human footprint value in the year 2013 than it did in the year 2000. We assess both total area and proportional losses, as smaller losses in 471 smaller units may potentially more significant to those unique assemblages as large ones ¹⁰⁵. 472 In addition to calculating the overall state of biomes and ecoregions for the years 2000 and 473 2013, we calculated the state for each time period in the human footprint dataset (2000, 2005, 474 2010 and 2013). All spatial analyses were carried out using ArcMap 10.5¹⁰⁶. We report on 475 values rounded to the nearest ten throughout for readability. For all values see Supplemental 476 477 1.

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482 Author Contributions

483	LW conceived the	idea and B.W.	and J.W.	designed the rese	arch. B.W	carried out the
405	J. W CONCEIVED the	\mathbf{D}		ucoigned the rese		. carried out the

- 484 analysis and led the writing of the manuscript. O.V and S.A created the updated human
- 485 footprint maps with the support of J.E, S.G, A.H, P.J, R.P, S.R.B, C.S, and A.V. All authors
- 486 contributed to and edited the manuscript.

487 Supplementary material

- 488 Supplemental 1 Excel sheets detailing the area in each state, and the area that transitioned
- 489 between each state at the global, biome, ecoregional, and national scales.
- 490 Supplemental 2 Technical validation for the human footprint
- 491 Table S1 Summary of the data and methodology used to create the human footprint maps
- 492 for the years 2000, 2005, 2010 and 2013

493 **Competing interests**

494 None declared.

495 Data availability

- 496 The updated human footprint maps, and all the code for generating them are freely
- 497 downloadable from <u>https://github.com/scabecks/humanfootprint_2000-2013</u>. All other
- 498 geographic layers used to carry out this analysis are available online from the reference
- 499 sources.

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502 **References**

- Ellis, E.C., and Ramankutty, N. (2008). Putting people in the map: anthropogenic
 biomes of the world. Front. Ecol. Environ. *6*, 439–447.
- Morales-Hidalgo, D., Oswalt, S.N., and Somanathan, E. (2015). Status and trends in global primary forest, protected areas, and areas designated for conservation of biodiversity from the Global Forest Resources Assessment 2015. For. Ecol. Manage. 352, 68–77.
- Hazlitt, S.L., Martin, T.G., Sampson, L., and Arcese, P. (2010). The effects of
 including marine ecological values in terrestrial reserve planning for a forest-nesting
 seabird. Biol. Conserv. *143*, 1299–1303.
- 512 4. Bonan, G.B. (2008). Forests and climate change: forcings, feedbacks, and the climate
 513 benefits of forests. Science. *320*, 1444–1449.
- 5. Sheil, D., and Murdiyarso, D. (2009). How forests attract rain: an examination of a new hypothesis. Bioscience *59*, 341–347.
- Dinerstein, E., Vynne, C., Sala, E., Joshi, A.R., Fernando, S., Lovejoy, T.E., Mayorga,
 J., Olson, D., Asner, G.P., and Baillie, J.E.M. (2019). A Global Deal For Nature:
 Guiding principles, milestones, and targets. Sci. Adv. 5, eaaw2869.
- 519 7. Lovejoy, T.E. (2016). Conservation biology: the importance of wilderness. Curr. Biol. 26, R1235–R1237.
- 8. Watson, J.E.M., Venter, O., Lee, J., Jones, K.R., Robinson, J.G., Possingham, H.P.,
 and Allan, J.R. (2018). Protect the last of the wild. Nature 563, 27–30.
- 523 9. Watson, J.E.M., and Venter, O. (2019). Mapping the continuum of humanity's footprint on land. OneEarth *1*, 175–180.
- Allan, J.R., Venter, O., and Watson, J.E.M. (2017). Temporally inter-comparable maps
 of terrestrial wilderness and the Last of the Wild. Sci. data *4*, 170187.
- Jones, K.R., Klein, C.J., Halpern, B.S., Venter, O., Grantham, H., Kuempel, C.D.,
 Shumway, N., Friedlander, A.M., Possingham, H.P., and Watson, J.E.M. (2018). The
 location and protection status of Earth's diminishing marine wilderness. Curr. Biol. 28,
 2506–2512.
- 531 12. Di Marco, M., Venter, O., Possingham, H.P., and Watson, J.E.M. (2018). Changes in human footprint drive changes in species extinction risk. Nat. Commun. 9, 4621.
- Beyer, H.L., Venter, O., Grantham, H.S., and Watson, J.E.M. (2019). Substantial
 losses in ecoregion intactness highlight urgency of globally coordinated action.
 Conserv. Lett. *e12592*.
- Jones, K.R., Venter, O., Fuller, R.A., Allan, J.R., Maxwell, S.L., Negret, P.J., and
 Watson, J.E.M. (2018). One-third of global protected land is under intense human
 pressure. Science. *360*, 788–791.
- Allan, J.R., Watson, J.E.M., Di, M.M., O'Bryan, C.J., Possingham, H.P., Atkinson,
 S.C., and Venter, O. (2019). Hotspots of human impact on threatened terrestrial
 vertebrates. PLoS Biol. *17*, e3000158.
- 542 16. Tucker, M.A., Böhning-Gaese, K., Fagan, W.F., Fryxell, J.M., Van Moorter, B.,

Alberts, S.C., Ali, A.H., Allen, A.M., Attias, N., and Avgar, T. (2018). Moving in the 543 Anthropocene: Global reductions in terrestrial mammalian movements. Science. 359, 544 466-469. 545 17. Allan, J.R., Venter, O., Maxwell, S., Bertzky, B., Jones, K., Shi, Y., and Watson, 546 J.E.M. (2017). Recent increases in human pressure and forest loss threaten many 547 Natural World Heritage Sites. Biol. Conserv. 206, 47-55. 548 18. Geldmann, J., Joppa, L.N., and Burgess, N.D. (2014). Mapping change in human 549 550 pressure globally on land and within protected areas. Conserv. Biol. 28, 1604-1616. 19. Watson, J.E.M., and Venter, O. (2017). A global plan for nature conservation. Nature 551 550, 48-49. 552 20. Tulloch, V.J.D., Tulloch, A.I.T., Visconti, P., Halpern, B.S., Watson, J.E.M., Evans, 553 554 M.C., Auerbach, N.A., Barnes, M., Beger, M., and Chadès, I. (2015). Why do we map threats? Linking threat mapping with actions to make better conservation decisions. 555 Front. Ecol. Environ. 13, 91–99. 556 21. Vörösmarty, C.J., McIntyre, P.B., Gessner, M.O., Dudgeon, D., Prusevich, A., Green, 557 P., Glidden, S., Bunn, S.E., Sullivan, C.A., and Liermann, C.R. (2010). Global threats 558 to human water security and river biodiversity. Nature 467, 555-561. 559 22. Allan, J.R., Grossmann, F., Craig, R., Nelson, A., Maina, J., Flower, K., Bampton, J., 560 ste Deffontaines, J.-B., Miguel, C., and Araquechande, B. (2017). Patterns of forest 561 loss in one of Africa's last remaining wilderness areas: Niassa National Reserve 562 (Northern Mozambique). Parks 23, 39-50. 563 Venter, O., Sanderson, E.W., Magrach, A., Allan, J.R., Beher, J., Jones, K.R., 23. 564 565 Possingham, H.P., Laurance, W.F., Wood, P., and Fekete, B.M. (2016). Global terrestrial Human Footprint maps for 1993 and 2009. Sci. data 3, sdata201667. 566 Sanderson, E.W., Jaiteh, M., Levy, M.A., Redford, K.H., Wannebo, A. V, and 24. 567 Woolmer, G. (2002). The human footprint and the last of the wild: the human footprint 568 is a global map of human influence on the land surface, which suggests that human 569 beings are stewards of nature, whether we like it or not. AIBS Bull. 52, 891–904. 570 Watson, J.E.M., Jones, K.R., Fuller, R.A., Marco, M. Di, Segan, D.B., Butchart, 25. 571 572 S.H.M., Allan, J.R., McDonald-Madden, E., and Venter, O. (2016). Persistent disparities between recent rates of habitat conversion and protection and implications 573 for future global conservation targets. Conserv. Lett. 9, 413–421. 574 26. Crooks, K.R., Burdett, C.L., Theobald, D.M., King, S.R.B., Di Marco, M., Rondinini, 575 576 C., and Boitani, L. (2017). Quantification of habitat fragmentation reveals extinction 577 risk in terrestrial mammals. Proc. Natl. Acad. Sci. 114, 7635-7640. 27. Olson, D.M.D., Dinerstein, E., Wikramanayake, E.D., Burgess, N.D., Powell, G.V.N., 578 579 Underwood, E.C., D'amico, J.A., Itoua, I., Strand, H.E., and Morrison, J.C. (2001). Terrestrial Ecoregions of the World: A New Map of Life on Earth: A new global map 580 of terrestrial ecoregions provides an innovative tool for conserving biodiversity. 581 582 Bioscience 51, 933–938. Newbold, T., Hudson, L.N., Arnell, A.P., Contu, S., De Palma, A., Ferrier, S., Hill, 583 28. S.L.L., Hoskins, A.J., Lysenko, I., and Phillips, H.R.P. (2016). Has land use pushed 584 terrestrial biodiversity beyond the planetary boundary? A global assessment. Science. 585

- 586 *353*, 288–291.
- Steffen, W., Richardson, K., Rockström, J., Cornell, S.E., Fetzer, I., Bennett, E.M.,
 Biggs, R., Carpenter, S.R., De Vries, W., and De Wit, C.A. (2015). Planetary
 boundaries: Guiding human development on a changing planet. Science. 347,
 1259855.
- 30. Dinerstein, E., Olson, D., Joshi, A., Vynne, C., Burgess, N.D., Wikramanayake, E.,
 Hahn, N., Palminteri, S., Hedao, P., and Noss, R. (2017). An ecoregion-based
 approach to protecting half the terrestrial realm. Bioscience 67, 534–545.
- Hoekstra, J.M., Boucher, T.M., Ricketts, T.H., and Roberts, C. (2005). Confronting a
 biome crisis: global disparities of habitat loss and protection. Ecol. Lett. *8*, 23–29.
- Hannah, L., Carr, J.L., and Lankerani, A. (1995). Human disturbance and natural
 habitat: a biome level analysis of a global data set. Biodivers. Conserv. 4, 128–155.
- Tulloch, A.I.T., Gordon, A., Runge, C.A., and Rhodes, J.R. (2019). Integrating
 spatially realistic infrastructure impacts into conservation planning to inform strategic
 environmental assessment. Conserv. Lett., e12648.
- Gaston, K.J., Visser, M.E., and Hölker, F. (2015). The biological impacts of artificial light at night: the research challenge.
- Batcheller, A.L., Duke, C.S., and Porter, J.H. (2013). Big data and the future of
 ecology. Front. Ecol. Environ. 11, 156–162.
- 606 36. Carr, D.L. (2004). Proximate population factors and deforestation in tropical agricultural frontiers. Popul. Environ. 25, 585–612.
- Watson, J.E.M., Evans, T., Venter, O., Williams, B., Tulloch, A., Stewart, C.,
 Thompson, I., Ray, J.C., Murray, K., Salazar, A., *et al.* (2018). The exceptional value of intact forest ecosystems. Nat. Ecol. Evol. *2*, 599–610.
- 611 38. Fleischner, T.L. (1994). Ecological costs of livestock grazing in western North
 612 America. Conserv. Biol. 8, 629–644.
- Senior, R.A., Börger,
 L., Bennett, D.J., Choimes, A., and Collen, B. (2015). Global effects of land use on
 local terrestrial biodiversity. Nature *520*, 45.
- 40. Miraldo, A., Li, S., Borregaard, M.K., Flórez-Rodríguez, A., Gopalakrishnan, S.,
 Rizvanovic, M., Wang, Z., Rahbek, C., Marske, K.A., and Nogués-Bravo, D. (2016).
 An Anthropocene map of genetic diversity. Science. *353*, 1532–1535.
- 41. Maxwell, S.L., Evans, T., Watson, J.E.M., Morel, A., Grantham, H., Duncan, A.,
 Harris, N., Potapov, P., Runting, R.K., and Venter, O. (2019). Degradation and
 forgone removals increase the carbon impact of intact forest loss by 626%. Sci. Adv. 5,
 eaax2546.
- 42. Vargas, L.E.P., Laurance, W.F., Clements, G.R., and Edwards, W. (2015). The
 impacts of oil palm agriculture on Colombia's biodiversity: what we know and still
 need to know. Trop. Conserv. Sci. *8*, 828–845.
- 43. Williams, B.A., Grantham, H.S., Watson, J.E.M., Alvarez, S.J., Simmonds, J.S.,

627 628 629		Rogéliz, C.A., Da Silva, M.A., Forero-Medina, G., Etter, A., Nogales, J., <i>et al.</i> (2020). Minimising the loss of biodiversity and ecosystem services in an intact landscape under risk of rapid agricultural development. Environ. Res. Lett. <i>15</i> , 14001.
630 631 632	44.	Australian Government (2015). Our north, our future: White paper on developing northern Australia Available at: https://www.industry.gov.au/data-and-publications/our-north-our-future-white-paper-on-developing-northern-australia.
633 634 635	45.	Strassburg, B.B.N., Brooks, T., Feltran-Barbieri, R., Iribarrem, A., Crouzeilles, R., Loyola, R., Latawiec, A.E., Oliveira Filho, F.J.B., Scaramuzza, C.A. de M., and Scarano, F.R. (2017). Moment of truth for the Cerrado hotspot. Nat. Ecol. Evol. <i>1</i> , 99.
636 637 638 639	46.	Monteiro, L.M., Brum, F.T., Pressey, R.L., Morellato, L.P.C., Soares-Filho, B., Lima-Ribeiro, M.S., and Loyola, R. (2018). Evaluating the impact of future actions in minimizing vegetation loss from land conversion in the Brazilian Cerrado under climate change. Biodivers. Conserv. <i>29</i> , 1–22.
640 641 642	47.	Pressey, R.L., and Taffs, K.H. (2001). Scheduling conservation action in production landscapes: priority areas in western New South Wales defined by irreplaceability and vulnerability to vegetation loss. Biol. Conserv. <i>100</i> , 355–376.
643 644 645	48.	Wilson, K., Pressey, R.L., Newton, A., Burgman, M., Possingham, H., and Weston, C. (2005). Measuring and incorporating vulnerability into conservation planning. Environ. Manage. <i>35</i> , 527–543.
646 647 648	49.	Lenton, T.M., Rockström, J., Gaffney, O., Rahmstorf, S., Richardson, K., Steffen, W., and Schellnhuber, H.J. (2019). Climate tipping points—too risky to bet against. Nature <i>575</i> , 592–595.
649 650 651	50.	Di Marco, M., Ferrier, S., Harwood, T.D., Hoskins, A.J., and Watson, J.E.M. (2019). Wilderness areas halve the extinction risk of terrestrial biodiversity. Nature <i>573</i> , 582–585.
652 653 654 655	51.	IPBES (2019). Global assessment report on biodiversity and ecosystem services of the Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services (Bonn, Germany) Available at: https://www.ipbes.net/global-assessment-report- biodiversity-ecosystem-services.
656 657	52.	Visseren-Hamakers, I.J. (2015). Integrative environmental governance: enhancing governance in the era of synergies. Curr. Opin. Environ. Sustain. 14, 136–143.
658 659	53.	Wunder, S. (2007). The efficiency of payments for environmental services in tropical conservation. Conserv. Biol. <i>21</i> , 48–58.
660 661	54.	Maron, M., Simmonds, J.S., and Watson, J.E.M. (2018). Bold nature retention targets are essential for the global environment agenda. Nat. Ecol. Evol. <i>2</i> , 1194.
662 663	55.	Watson, J.E.M., Keith, D.A., Strassburg, B.B.N., Venter, O., Williams, B., and Nicholson, E. (2020). Set a global target for ecosystems. Nature <i>578</i> , 360–362.
664 665 666 667	56.	Di Marco, M., Harwood, T.D., Hoskins, A.J., Ware, C., Hill, S.L.L., and Ferrier, S. (2019). Projecting impacts of global climate and land-use scenarios on plant biodiversity using compositional-turnover modelling. Glob. Chang. Biol. <i>25</i> , 2763–2778.
668	57.	Van Ittersum, M.K., Van Bussel, L.G.J., Wolf, J., Grassini, P., Van Wart, J., Guilpart,

669 670		N., Claessens, L., de Groot, H., Wiebe, K., and Mason-D'Croz, D. (2016). Can sub-Saharan Africa feed itself? Proc. Natl. Acad. Sci. 113, 14964–14969.
671 672 673 674 675 676	58.	Shiferaw, B., Negassa, A., Koo, J., Wood, J., Sonder, K., Braun, J.A., and Payne, T. (2011). Future of wheat production in Sub-Saharan Africa: analyses of the expanding gap between supply and demand and economic profitability of domestic production. In Increasing Agricultural Productivity & Enhancing Food Security in Africa: New Challenges and Opportunities (Africa Hall, UNECA, Addis Ababa, Ethiopia: International Food Policy Research Institute (IFPRI)).
677 678	59.	Margono, B.A., Potapov, P. V, Turubanova, S., Stolle, F., and Hansen, M.C. (2014). Primary forest cover loss in Indonesia over 2000–2012. Nat. Clim. Chang. <i>4</i> , 730.
679 680 681	60.	Australian government About Papua New Guinea. Available at: https://web.archive.org/web/20110518125558/http://www.ausaid.gov.au/country/png/p ng_intro.cfm.
682 683 684	61.	Nelson, P.N., Gabriel, J., Filer, C., Banabas, M., Sayer, J.A., Curry, G.N., Koczberski, G., and Venter, O. (2014). Oil palm and deforestation in Papua New Guinea. Conserv. Lett. 7, 188–195.
685 686 687	62.	Austin, K.G., Mosnier, A., Pirker, J., McCallum, I., Fritz, S., and Kasibhatla, P.S. (2017). Shifting patterns of oil palm driven deforestation in Indonesia and implications for zero-deforestation commitments. Land use policy <i>69</i> , 41–48.
688 689 690	63.	Lenzen, M., Moran, D., Kanemoto, K., Foran, B., Lobefaro, L., and Geschke, A. (2012). International trade drives biodiversity threats in developing nations. Nature <i>486</i> , 109.
691 692 693 694	64.	Costanza, R., Daly, L., Fioramonti, L., Giovannini, E., Kubiszewski, I., Mortensen, L.F., Pickett, K.E., Ragnarsdottir, K.V., De Vogli, R., and Wilkinson, R. (2016). Modelling and measuring sustainable wellbeing in connection with the UN Sustainable Development Goals. Ecol. Econ. <i>130</i> , 350–355.
695 696 697	65.	Secretariat of the Convention on Biological Diversity (2020). Zero Draft of the Post-2020 Global Biodiversity Framework Available at: https://www.cbd.int/article/2020-01-10-19-02-38.
698 699	66.	OS Geo Project (2017). GRASS GIS 7.2.2. Available at: https://grass.osgeo.org/news/68/15/GRASS-GIS-7-2-2-released/.
700 701	67.	Tratalos, J., Fuller, R.A., Warren, P.H., Davies, R.G., and Gaston, K.J. (2007). Urban form, biodiversity potential and ecosystem services. Landsc. Urban Plan. <i>83</i> , 308–317.
702 703 704 705	68.	Aronson, M.F.J., La Sorte, F.A., Nilon, C.H., Katti, M., Goddard, M.A., Lepczyk, C.A., Warren, P.S., Williams, N.S.G., Cilliers, S., and Clarkson, B. (2014). A global analysis of the impacts of urbanization on bird and plant diversity reveals key anthropogenic drivers. Proc. R. Soc. B Biol. Sci. <i>281</i> , 20133330.
706 707 708	69.	Butchart, S.H.M., Walpole, M., Collen, B., Van Strien, A., Scharlemann, J.P.W., Almond, R.E.A., Baillie, J.E.M., Bomhard, B., Brown, C., and Bruno, J. (2010). Global biodiversity: indicators of recent declines. Science. <i>328</i> , 1164–1168.
709 710 711	70.	Chamberlain, D.E., Cannon, A.R., Toms, M.P., Leech, D.I., Hatchwell, B.J., and Gaston, K.J. (2009). Avian productivity in urban landscapes: a review and meta- analysis. Ibis (Lond. 1859). <i>151</i> , 1–18.

712 71. Elvidge, C.D., Hsu, F.-C., Baugh, K.E., and Ghosh, T. (2014). National trends in satellite-observed lighting. Glob. urban Monit. Assess. through earth Obs. 23, 97–118. 713 72. Elvidge, C.D., Imhoff, M.L., Baugh, K.E., Hobson, V.R., Nelson, I., Safran, J., Dietz, 714 J.B., and Tuttle, B.T. (2001). Night-time lights of the world: 1994–1995. ISPRS J. 715 Photogramm. Remote Sens. 56, 81–99. 716 Small, C., Elvidge, C.D., Balk, D., and Montgomery, M. (2011). Spatial scaling of 717 73. stable night lights. Remote Sens. Environ. 115, 269-280. 718 74. Elvidge, C.D., Ziskin, D., Baugh, K.E., Tuttle, B.T., Ghosh, T., Pack, D.W., Erwin, 719 E.H., and Zhizhin, M. (2009). A fifteen year record of global natural gas flaring 720 derived from satellite data. Energies 2, 595–622. 721 75. Elvidge, C.D., Baugh, K.E., Zhizhin, M., and Hsu, F.-C. (2013). Why VIIRS data are 722 723 superior to DMSP for mapping nighttime lights. Proc. Asia-Pacific Adv. Netw. 35. Brashares, J.S., Arcese, P., and Sam, M.K. (2001). Human demography and reserve 724 76. size predict wildlife extinction in West Africa. Proc. R. Soc. London. Ser. B Biol. Sci. 725 268, 2473-2478. 726 727 77. Miller, G.H., Fogel, M.L., Magee, J.W., Gagan, M.K., Clarke, S.J., and Johnson, B.J. (2005). Ecosystem collapse in Pleistocene Australia and a human role in megafaunal 728 extinction. Science. 309, 287-290. 729 78. Burney, D.A., and Flannery, T.F. (2005). Fifty millennia of catastrophic extinctions 730 731 after human contact. Trends Ecol. Evol. 20, 395-401. 79. CIESIN, and SEDAC (2017). Gridded Population of the World Version 4. Cent. Int. 732 Earth Sci. Inf. Netw., 1–21. Available at: 733 734 http://sedac.ciesin.columbia.edu/data/set/gpw-v4-population-density [Accessed 735 September 11, 2017]. NOAA (2013). Version 4 DMSP-OLS Nighttime Lights Time Series. Available at: 80. 736 https://ngdc.noaa.gov/eog/dmsp/downloadV4composites.html#AVSLCFC [Accessed 737 738 March 17, 2020]. ESA (2017). 300 m annual global land cover time series from 1992 to 2015. Available 739 81. at: http://maps.elie.ucl.ac.be/CCI/viewer/ [Accessed July 13, 2017]. 740 82. Fischer, J., Brosi, B., Daily, G.C., Ehrlich, P.R., Goldman, R., Goldstein, J., 741 742 Lindenmayer, D.B., Manning, A.D., Mooney, H.A., and Pejchar, L. (2008). Should agricultural policies encourage land sparing or wildlife-friendly farming? Front. Ecol. 743 Environ. 6, 380-385. 744 745 83. Ramankutty, N., Evan, A.T., Monfreda, C., and Foley, J.A. (2008). Farming the planet: 1. Geographic distribution of global agricultural lands in the year 2000. Global 746 Biogeochem. Cycles 22, GB1003. 747 Luck, G.W., and Daily, G.C. (2003). Tropical countryside bird assemblages: richness, 84. 748 749 composition, and foraging differ by landscape context. Ecol. Appl. 13, 235-247. 85. OpenStreetMap Contributors (2017). Planet OSM. Available at: https://planet.osm.org 750 751 [Accessed May 29, 2017]. Herold, M., Mayaux, P., Woodcock, C.E., Baccini, A., and Schmullius, C. (2008). 752 86.

753 754		Some challenges in global land cover mapping: An assessment of agreement and accuracy in existing 1 km datasets. Remote Sens. Environ. <i>112</i> , 2538–2556.
755 756	87.	Lehner, B., Verdin, K., and Jarvis, A. (2008). New global hydrography derived from spaceborne elevation data. Eos, Trans. Am. Geophys. Union <i>89</i> , 93–94.
757 758 759	88.	Kauffman, J.B., and Krueger, W.C. (1984). Livestock impacts on riparian ecosystems and streamside management implications a review. Rangel. Ecol. Manag. Range Manag. Arch. <i>37</i> , 430–438.
760 761	89.	Trombulak, S.C., and Frissell, C.A. (2000). Review of ecological effects of roads on terrestrial and aquatic communities. Conserv. Biol. 14, 18–30.
762 763	90.	Woodroffe, R., and Ginsberg, J.R. (1998). Edge effects and the extinction of populations inside protected areas. Science. 280, 2126–2128.
764 765	91.	Laurance, W.F., Goosem, M., and Laurance, S.G.W. (2009). Impacts of roads and linear clearings on tropical forests. Trends Ecol. Evol. 24, 659–669.
766 767	92.	Adeney, J.M., Christensen Jr, N.L., and Pimm, S.L. (2009). Reserves protect against deforestation fires in the Amazon. PLoS One <i>4</i> , e5014.
768 769	93.	Forman, R.T.T., and Alexander, L.E. (1998). Roads and their major ecological effects. Annu. Rev. Ecol. Syst. 29, 207–231.
770 771	94.	OpenStreetMap, and OpenStreetMap Contributors (2020). OpenStreetMap. Available at: https://www.openstreetmap.org/about [Accessed March 18, 2020].
772 773 774 775	95.	Center for International Earth Science Information Network (2010). Global Roads Open Access Data Set (gROADS), v1 (1980 – 2010). NASA Socioecon. Data Appl. Cent. Available at: https://sedac.ciesin.columbia.edu/data/set/groads-global-roads- open-access-v1.
776 777 778	96.	National Imagery and Mapping Agency (1997). National Imagery and Mapping Agency. Vector Map Level 0. Available at: https://earth-info.nga.mil/publications/vmap0.html.
779 780 781	97.	Bjerklie, D.M., Dingman, S.L., Vorosmarty, C.J., Bolster, C.H., and Congalton, R.G. (2003). Evaluating the potential for measuring river discharge from space. J. Hydrol. <i>278</i> , 17–38.
782 783 784	98.	Bjerklie, D.M., Moller, D., Smith, L.C., and Dingman, S.L. (2005). Estimating discharge in rivers using remotely sensed hydraulic information. J. Hydrol. <i>309</i> , 191–209.
785 786 787	99.	Mittermeier, R.A., Mittermeier, C.G., Brooks, T.M., Pilgrim, J.D., Konstant, W.R., Da Fonseca, G.A.B., and Kormos, C. (2003). Wilderness and biodiversity conservation. Proc. Natl. Acad. Sci. <i>100</i> , 10309–10313.
788 789 790	100.	Watson, J.E.M., Shanahan, D.F., Di Marco, M., Allan, J., Laurance, W.F., Sanderson, E.W., Mackey, B., and Venter, O. (2016). Catastrophic declines in wilderness areas undermine global environment targets. Curr. Biol. <i>26</i> , 2929–2934.
791 792	101.	Forero-Medina, G., and Joppa, L. (2010). Representation of global and national conservation priorities by Colombia's protected area network. PLoS One 5, e13210.
793	102.	Rodríguez, J.P., Rodríguez-Clark, K.M., Baillie, J.E.M., Ash, N., Benson, J., Boucher,

794 795		T., Brown, C., Burgess, N.D., Collen, B.E.N., and Jennings, M. (2011). Establishing IUCN red list criteria for threatened ecosystems. Conserv. Biol. <i>25</i> , 21–29.
796	103.	Sandvik, B. (2009). World Borders Dataset. Available at: thematicmapping.org.
797 798 799	104.	Crouzeilles, R., Curran, M., Ferreira, M.S., Lindenmayer, D.B., Grelle, C.E. V, and Benayas, J.M.R. (2016). A global meta-analysis on the ecological drivers of forest restoration success. Nat. Commun. 7, 1–8.
800 801	105.	He, F., and Hubbell, S.P. (2011). Species–area relationships always overestimate extinction rates from habitat loss. Nature 473, 368–371.
802	106.	ESRI (2017). ArcGIS Release 10.5.1. Redlands, CA.
803		