

12-2016

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### Recommended Citation

Soti, P. G., & Jayachandran, K. (2016). Effect of exotic invasive old world climbing fern (*lygodium microphyllum*) on soil properties. *Journal of soil science and plant nutrition*, 16(4), 930-940.  
<http://dx.doi.org/10.4067/S0718-95162016005000066>

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# Effect of exotic invasive old world climbing fern (*Lygodium microphyllum*) on soil properties

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## Abstract

Old World climbing fern (*Lygodium microphyllum*) has become one of the most serious ecological threats to the integrity of the greater Everglades ecosystem of south Florida. In this study, we analyzed the effects of Old World climbing fern on surface soil characteristics at invaded sites in Florida. We compared soil characteristics of six invaded and adjacent uninvaded plots at three different locations. Our results show that the fern can grow and thrive in a wide range of soil types and the impact on the soil was site specific with effects being more prominent in sites with low nutrient status. Additionally, there were significant differences in the soil nutrient status and microbial population in the invaded and uninvaded sites. Sites with Old World climbing fern had significantly higher nutrient concentrations that correlated with higher soil organic matter. Overall our results indicate that this exotic pest plant can potentially alter its below ground environment to its own benefit by enhancing the soil nutrient status by adding soil organic matter.

**Keywords:** Invasive plants, nutrient cycle, pH, soil organic matter, soil nutrients

## 1. Introduction

Invasion by exotic invasive plants has a substantial impact on the structure, function, and composition of the native communities (Evans *et al.* 2001; Ehrenfeld, 2003; Rice and Emery, 2003; Vila *et al.* 2011; Lanta *et al.* 2015). Existing literature provides evidence that invasive plant species can modify physical, chemical, and biological properties of the soil including inputs and cycling of nutrients (Ehrenfeld, 2003; Hawkes *et al.* 2005; Sperry *et al.* 2006), soil pH (Kourtev *et al.* 2003), soil organic matter and aggregation (Sag-

gar *et al.* 1999). Invasive plants also modify the biotic composition of the soil by affecting the soil food web (Duda *et al.* 2003), total microbial communities (Kourtev *et al.* 2003), and fungal communities (Hawkes *et al.* 2006). Some invasive plants are also reported to exude allelochemicals which could inhibit soil borne pathogens, defend against disease, and repel insects (Yuan *et al.* 2012; and references there in). However, the documented impacts of invasive species

on soil characteristics are diverse. Most of the studies have reported increased soil nutrient stock in invaded sites compared to non-invaded sites creating a positive feedback benefiting invasive species (Duda *et al.* 2003; Vanderhoeven *et al.* 2005; Liao *et al.* 2008; Perkins, 2011). A meta-analysis of litter decomposition rates by Liao *et al.* (2008) showed that the litter decomposition rate of the invasive plants was on average, 117% faster than the co-occurring native species. However, Ehrenfeld (2010) demonstrated slower rates of litter decay in exotic species compared with the native plant species. This species-specific variation was attributed to be the most important factor in determining the decomposition rates (Hoorens *et al.* 2003), as each plant species has a unique biochemical composition, including the nitrogen concentration and carbon-to-nitrogen ratio. Gordon (1998) in her meta-analysis has shown that out of 31 species considered most invasive, 12–20 (39–64%) potentially alter the ecosystem properties of geomorphology, hydrology, biogeochemistry, and disturbance in Florida.

*Lygodium microphyllum* (Old World Climbing Fern) is a highly invasive species distributed throughout the freshwater and moist habitats of south Florida. It is common in cypress swamp, pine flatlands, wet prairies, sawgrass marshes, mangrove communities, and the Everglades tree islands (Pemberton and Ferriter, 1998). According to an estimate by (Ferriter and Pernas, 2006), *L. microphyllum* covers 183,080 acres in the entire South/Central Florida region. Managing *L. microphyllum* has been a significant challenge for land resource managers and researchers due to its extensive rapid invasion in natural areas of south Florida. Once established, *L. microphyllum* dominates both understory and overstory native wetland habitats, and has the ability to grow in varying hydrological (Gandiaga *et al.* 2009), nutrient (Volin *et al.* 2010), soil pH (Soti *et al.* 2014), and light gradients (Volin *et al.* 2004). Soti *et al.* (2014) have shown that *L. micro-*

*phyllum* is highly dependent on mycorrhizal fungi for growth and phosphorus uptake which could highly enhance its invasiveness.

Above-ground changes caused by *L. microphyllum* in south Florida natural areas are obvious and have remained the focus of land managers and researchers, but below ground changes caused by the plant-soil feedback have not gained much research interest so far. However, for the management of this exotic invasive and restoration of invaded sites, it is essential to understand the direct and indirect impacts of *L. microphyllum* on soil processes and how these changes can influence the successful management of invaded areas. If *L. microphyllum* can successfully modify soil processes such as nutrient cycling, litter decomposition, and soil microbial communities, simply removing it may not be an effective management strategy. Furthermore, it is necessary to assess the effects of soil modification by *L. microphyllum* on the invasibility of plant communities, whether this modification facilitates other invasive species and if it has a negative impact on the native species. The aim of this paper is to obtain baseline information on how *L. microphyllum* alters the physical, chemical, and microbial characteristics of the invaded sites, which would help to better understand and interpret effects of various management techniques used to control this pest plant species. We compared an invaded and a nearby non-invaded site to examine the impacts of *L. microphyllum* invasion on the topsoil chemistry and microbial populations at three sites with different soil characteristics.

## 2. Methods

### 2.1. Sampling sites

Two sites in south Florida were selected for the comparison of soil characteristics: Tree Tops Park (Bro-

ward county) and Jonathan Dickinson Park (Martin county), and one site in central Florida, the Trustcorp/Tiedtke property (Lake County). These sites were paired to include one plot with *L. microphyllum* and another with native plants. To minimize the probability of pre-existing difference in the soil characteristics, the sites were selected based on the following criteria: 1) a well-established monospecific population; 2) sites that had not undergone any management

activities for at least the last five years; 3) homogeneous soil type under both the native and invasive plants (texture and parent material); and 4) adjacent plots with native plants. 5) no known difference in the disturbance status at the plot with invasive plants and corresponding invasive plot. These criteria are similar to the ones used by Vanderhoeven et al. (2005). Site location and dominant vegetation in the sampling sites are given in Table 1.

**Table 1.** Site locations, vegetation, and site type.

Site	Coordinates	Dominant species in the invaded sites	Site Type
<b>Tree Tops Park</b>	26°4'0.04" N, 80°16'5.88" W	<i>Chrysobalanus icaco</i> , <i>Osmunda regalis</i> var. <i>spectabilis</i> , <i>Annona glabra</i>	County Park, disturbed habitat, poorly drained
<b>Jonathan Dickinson</b>	27°0'37.33" N, 80°7'20.28" W	<i>Pinus elliotii</i> , <i>Myrica cerifera</i> , <i>Ilex cassine</i> , <i>Serenoa repens</i>	State Park, well drained, undisturbed habitat
<b>Central Florida</b>	28°23'4.03" N, 81°44'41.30" W	<i>Pinus elliotii</i> , <i>Quercus geminata</i> , <i>Quercus nigra</i> , <i>Serenoa repens</i>	Private property, sand mine spoil

## 2.2. Soil sampling and analysis

At each sampling site, six 1 m x 1 m plots were selected randomly and soil from the 10-15 cm deep zone was collected from each of the four corners and the center of each plot with a soil corer ( $\emptyset$  18 mm) and mixed homogeneously into one bulk sample for each plot. The soil samples from south Florida were transported to the laboratory in a cooler and the samples from central Florida were cooled to 4 °C and shipped overnight. A portion of the soil samples from all sites were stored in a 4 °C refrigerator until analysis for biological measure-

ments. A small portion of each soil sample was air dried and passed through a 2 mm sieve for analysis of physicochemical properties. Those subsamples were then ground to fine powder with a mortar and pestle, and stored at room temperature in air-tight containers for further analysis of nutrients and trace elements. The soil pH was measured with a pH meter (soil solution ratio 1:2 in DI water), texture was measured by the hydrometer method, and total organic matter was measured based on the standard loss-on-ignition method (500 °C, 5 hours; Storer, 1984). Total C and N were measured with a Truspec CN analyzer. Total Ca, Fe, Al, Mg, K, Mn,

and P were measured with an ICP–MS at USDA, ARS Laboratory, Homestead, Florida after following the acid digestion Method 3050 (USEPA, 1996). One gram of each finely ground soil sample was transferred to a large glass tube and mixed with 10 ml of 30% HNO<sub>3</sub>. The tubes were covered with a vapor recovery system and heated to 95 ± 5° C and refluxed for 10 minutes without boiling under the hood in a heating block maintained with a Partlow Mic 6000 Profile Process Controller. After cooling to 40 °C, 5 ml of concentrated HNO<sub>3</sub> was added and the sample was heated again until no brown fumes were given off. After cooling to 40 °C, 2 ml of DI water and 3 ml of 30% H<sub>2</sub>O<sub>2</sub> was added to each tube and heated until the effervescence subsided. The samples were cooled and diluted to 50 ml with DI water, centrifuged at 2000 rpm for 10 minutes and filtered with Whatman No. 41 filter paper.

The total colony-forming-units (CFU) of bacteria and fungi were determined by the standard dilution spread plate method as described by Seely and VanDemark (1981). The dry equivalent of one-gram soil was mixed in 9 ml sterile water (autoclaved) and was diluted serially. Samples were vigorously mixed during dilution to assist in dislodging the bacteria from the soil particles. A serial dilution of 10<sup>-2</sup>, 10<sup>-3</sup>, 10<sup>-4</sup>, and 10<sup>-5</sup> was made for fungi and 10<sup>-4</sup>, 10<sup>-5</sup>, 10<sup>-6</sup>, and 10<sup>-7</sup> for bacteria. A total of 100 µl of diluted soil suspension was spread on three plates per soil sample for both bacteria and fungi at each dilution level. Nutrient agar containing cycloheximide solution (to prevent fu-

gal growth) was used for bacteria and Rose Bengal Agar (RBA) with streptomycin sulphate (to prevent bacteria growth) was used for the estimation of fungal colonization. Sterilized water was spread on the agar plates that were used as controls.

### 2.3. Data analysis

Differences in soil characteristics between the invaded and uninvaded plots were compared by means of paired t-tests. Additionally, a two-way ANOVA was done with site and vegetation type (invaded and uninvaded) as the fixed main effects for selected soil parameters. Pearson's correlation analysis was done with all sites pooled to determine relationships between the measured soil variables. Differences are reported as significant for tests with *P*-values ≤ 0.05. All the parameters were analyzed with SAS Version 9.2 software.

## 3. Results

Our results show that there was a clear contrast between the rhizosphere soils from the three sites (Table 2, 3 & 4). All the study sites were acidic (soil pH 4.95-6.36), but the sites in central Florida and at Tree Tops Park had lower soil pH compared to the sites at Jonathan Dickinson Park. Likewise, the soil texture differed among the sites; Jonathan Dickinson sites had sandy soils, while the soils in central Florida were clayey, and the soil at Tree Tops Park was sandy loam. Soil organic matter was highest at Tree Tops Park and was lowest at Jonathan Dickinson Park.

**Table 2.** Comparison of the topsoil chemical characteristics (means with standard deviations in parentheses) at the three sites with and without *L. microphyllum*.

Site		Al (mg/g)	C (%)	Ca (mg/g)	N (%)	P (mg/g)	Zn (µg/g)	OM (%)	pH (H <sub>2</sub> O)
Central FL	Native	2.62	2.18	0.68	0.04	0.67	21.46	5.18	4.95
		(0.64)	(0.50)	(0.34)	(0.11)	(0.10)	(9.94)	(0.80)	(0.45)
	Invasive	5.07	4.03	0.41	0.195	1.03	15.93	8.65	5.78
		(0.68)	(0.84)	(0.13)	(0.27)	(0.26)	(8.25)	(1.09)	(0.12)
	P Level <sup>a</sup>	**	**	*	ns	**	ns	**	**
Jonathan Dickinson	Native	0.43	3.13	1.19	0.26	1.02	7.24	1.08	6.36
		(0.07)	(1.10)	(0.21)	(0.08)	(0.06)	(1.26)	(0.41)	(0.09)
	Invasive	0.93	7.02	3.35	0.44	1.15	8.77	4.32	6.57
		(0.30)	(1.88)	(1.77)	(0.12)	(0.09)	(1.86)	(0.90)	(0.1)
	P Level <sup>a</sup>	*	**	*	**	*	ns	**	*
Tree Tops	Native	1.62	16.55	9.11	1.31	1.11	17.48	36.75	5.54
		(0.14)	(3.02)	(1.15)	(0.40)	(0.04)	(5.17)	(0.10)	(0.04)
	Invasive	1.88	22.43	17.21	1.27	1.22	23.21	44.42	5.60
		(0.25)	(4.15)	(6.30)	(0.17)	(0.16)	(4.53)	(2.71)	(0.06)
	P Level <sup>a</sup>	*	*	*	ns	ns	*	**	ns

<sup>a</sup> Significance for paired t-test, ns: not significant; Probability levels: \*: P<0.05; \*\* P<0.01; \*\*\*P<0.0001. OM: soil organic matter.

Significant differences were found in the soil chemistry among invaded and non-invaded soils in different between the invaded plots and non-invaded plots in all three sampling sites (Table 2). Most notably, there were significant differences in the total soil Al, C, Ca, and OM% at all three sites (Table 2). At all the sites, soils invaded by *L. microphyllum* had higher concentrations of Al, Ca, C and organic matter. Differences in the concentrations of N, P, Zn, and pH were site-specific. Soils under native vegetation were gener-

ally more acidic; this was statistically significant at the central Florida and Jonathan Dickinson sites but was not statistically significant at the Tree Tops Park site. There was no significant difference in the Cu, Fe, K, Mg, and Mn concentrations under native vegetation and *L. microphyllum* at all the three sites (data not shown). The most significant effects on soil characteristics were seen at sites with the lowest nutrient concentrations.

Two-way ANOVA results of sampling sites, plant type (native/invasive) and their interaction (Table 3), shows that there was a significant difference in the sites soil chemistry. Plant type had a significant impact on soil Al, C, Ca, K, Mg, P, OM, and pH while there was no significant impact on Cu, Fe, N, and Zn. Similarly, the interaction of site and plant had a significant impact only on soil Al, Ca, Mg, OM, and pH.

Though it was site specific, *L. microphyllum* invasion caused a shift of the bacteria and fungi CFUs (Table 4). Largest CFUs of bacteria were observed under native plants in the Tree Tops Park site ( $282.5 \times 10^6$ ) while the lowest CFUs were in the Jonathan Dickinson Park site under *L. microphyllum* ( $39 \times 10^6$ ). The bacteria population was significantly higher under the native species compared to *L. microphyllum* in two sites; the difference was not statistically significant at the central Florida site. On the other hand, the fungal population was significantly higher under *L. microphyllum* compared to the native plants at all three sites. The Central FL site under

*L. microphyllum* has the highest fungal CFUs ( $88.83 \times 10^3$ ), while it was lowest under native plants at Jonathan Dickinson Park site ( $33.5 \times 10^3$ ).

Pearson's correlation analysis among the soil variables indicated that there was a strong correlation between soil nutrients and soil organic matter (Table 5). There a strong negative correlation between soil OM and Al concentration ( $r = -0.030, p < 0.05$ ) and strong positive correlation with C, Ca, and N ( $r = 0.928, 0.874, 0.875; p < 0.0001$ ). Soil pH had negative correlation with Al concentration ( $r = -0.385, p < 0.05$ ), Zn ( $r = -0.716$ ), and positive correlation with total P ( $r = -0.716, p < 0.0001$ ). Total bacteria CFUs was positively correlated with C ( $r = 0.463, p < 0.01$ ), Ca ( $r = 0.339, p < 0.05$ ), N ( $r = 0.533, p < 0.01$ ) OM ( $r = 0.651, p < 0.0001$ ) and Zn ( $r = 0.469, p < 0.01$ ), and negatively correlated with soil pH ( $r = -0.632, p < 0.0001$ ). Similarly, total fungi CFUs was positively correlated with soil Al ( $r = 0.776, p < 0.0001$ ) and Zn ( $r = 0.524, p < 0.01$ ), and negatively correlated with pH ( $r = -0.432, p < 0.01$ ).

**Table 3.** Results of two-way analyses of variance (ANOVA) with degree of freedom (DF), F-value and probability levels for the effects of site, plant type and the interaction of the two on the soil characteristics.

Source	DF	Al	C	Ca	Cu	Fe	K	Mg	Mn	N	P
Site	2	195.3 ***	179.89 ***	75.91 ***	46.87 ***	19.5 ***	10.09 ***	28.59 ***	3.47 *	99.65 ***	16.78 ***
Plant	1	65.24 ***	25.33 ***	13.52 **	0.14 ns	2.55 ns	14.74 ***	12.16 **	2.54 ns	1.9 ns	19.04 ***
Site x Plant	2	27.05 ***	2.29 ns	7.53 **	0.52 ns	0.55 ns	0.83 ns	4.19 *	3.31 ns	0.94 ns	2.8 ns

  

Source	DF	OM	Zn	pH
Site	2	2132.23 ***	14.73 ***	102.88 ***
Plant	1	85.22 ***	0.08 ns	30.13 ***
Site X Plant	2	7.68 **	2.66 ns	12.29 ***

Probability levels: \*:  $P < 0.05$ ; \*\*:  $P < 0.01$ ; \*\*\* $P < 0.0001$ . ns: not significant; OM: soil organic matter.

**Table 4.** Comparison of the topsoil microbial populations (means with standard deviations in parentheses) at the three sites with and without *L. microphyllum*.

		TBC	TFC
Central FL	Native	152.5	61.16
		(16.42)	(9.82)
	Invasive	138.33	88.83
		(10.78)	(3.31)
	P Level <sup>a</sup>	ns	**
Jonathan Dickinson	Native	57.33	33.5
		(7.25)	(4.84)
	Invasive	39	46.66
		(4.28)	(7.76)
	P Level <sup>a</sup>	**	*
Tree Tops	Native	282.5	51.66
		(11.07)	(5.68)
	Invasive	143.66	73.83
		(13)	(12.27)
	P Level <sup>a</sup>	***	**

<sup>a</sup> Significance for paired t-test, ns: not significant; Probability levels: \*: P<0.05; \*\* P<0.01; \*\*\*P<0.0001. TBC: total bacterial count (count x 10<sup>6</sup> g<sup>-1</sup> dw); TFC: total fungal count (count x 10<sup>3</sup>g<sup>-1</sup> dw).



**Table 5.** Pearson's correlation coefficients between the selected soil parameters with all sites pooled.

	Al	C	Ca	N	OM	P	Zn	pH	TBC	TFC
Al	-									
C	-0.171 ns	-								
Ca	-0.173 ns	<b>0.892</b> ***	-							
N	-0.241 ns	0.894 ***	<b>0.792</b> ***	-						
OM	-0.030 *	<b>0.928</b> ***	<b>0.874</b> ***	<b>0.875</b> ***	-					
P	-0.099 ns	0.542 **	0.571 **	0.524 **	0.433 **	-				
Zn	0.255 ns	0.389 *	0.447 **	0.325 ns	0.470 **	-0.151 ns	-			
pH	-0.385 *	-0.127 ns	-0.142 ns	-0.069 ns	-0.321 ns	0.457 **	-0.716 ***	-		
TBC	0.254 ns	0.463 **	0.339 *	0.533 **	0.651 ***	-0.057 ns	0.469 **	-0.632 ***	-	
TFC	<b>0.776</b> ***	0.190 ns	0.135 ns	0.029 ns	0.270 ns	-0.035 ns	0.524 **	-0.432 **	0.255 ns	-

Probability levels: \*; P<0.05; \*\* P<0.01; \*\*\*P<0.0001. Coefficients higher than 0.75 are in bold.

OM: soil organic matter; TBC: total bacterial count; TFC: total fungal count.

#### 4. Discussion

In this study we compared soil from three different sites with contrasting soil characteristics and land use history; thus, the differences observed in soil measured soil parameters among these sites are in large part explained by this. The site effect was highly significant for all the soil parameters analyzed, indicating that *L. microphyllum* can adapt to and thrive in sites with a significant variation in nutrients as well as other soil characteristics. Additionally, our results indicate *L. microphyllum* has a significant impact on the soil characteristics, and the effect is highly dependent on existing soil conditions. Impacts on the total nutrient pool and microbial community structure are considered the most prominent effects of invasive species in the ecosystem. Overall, our results show higher

total soil nutrient pool in the rhizosphere of *L. microphyllum* compared to the rhizosphere of adjacent native species. These results follow the general trend reported by various researchers, where the nutrient pools in the invasive species rhizosphere are significantly increased compared to the coexisting natives (Duda *et al.* 2003; Vanderhoven, 2005; Dasonville *et al.* 2008; Liao *et al.* 2008; Perkins, 2011). This effect was most evident at sites with lowest nutrient concentration. As reported by Ehrenfeld (2003) and Liao *et al.* (2008), this may be the direct effect of higher quality (lower C:N and C: P ratios) litter added to the rhizosphere (Soti *et al.* unpublished data) which may result in faster turnover rates and increased nutrient availability.

Our results also show a difference in the rhizosphere microbial community which contributes directly to plant fitness, nutrient acquisition and stress tolerance. There was an increase on the fungi populations under *L. microphyllum* compared to the native species, while a decline in the bacteria population. This could be the result of the difference in the carbon input from plant to soil through root exudation (Merino *et al.* 2015; Violante and Caporale, 2015). Further analysis on the type of microbes and the role of allelochemicals, which are reported to be an important determinant for invasive success of exotic plants (Bais *et al.* 2003), in regulating the soil microbial community structure is necessary, but our results provide some evidence that *L. microphyllum* could regulate the structure of soil microbial communities in its rhizosphere.

Additionally, Since the Tree Tops site was a seasonally flooded site, this site had a high percentage of soil organic matter, which is common in the Florida Everglades. Soil organic matter was significantly higher in the invaded sites compared to the non-invaded sites. Soil organic matter was strongly correlated to the available soil nutrients which indicates that the difference in the organic matter inputs to the soil under the natives and *L. microphyllum* could influence the difference in the nutrient availability.

Although any pre-existing differences in the plots with and without *L. microphyllum* cannot be disregarded with complete certainty, we believe that the differences in the soil characteristics between the invaded and uninvaded plots could be the result of difference in plant species in them.

## 5. Conclusion

*L. microphyllum* invasion is still expanding and pre- and post-invasion comparison could provide a better insight on the alternative hypothesis that persisting differences in site leading to exotic species invasion.

Successful management of habitats invaded by exotic plant species requires a prior knowledge of whether the invaders have significantly altered the ecosystem (Walker and Smith, 1997) because soil properties such as texture, pH, and organic matter content influence herbicide efficiency and therefore control success. Along with the added organic matter, nutrients, and changes in the pH of the soil, other specific ecosystem process as outlined by Gordon (1998) could also be influenced by various exotic invasive species that create positive feedback for themselves and future invaders. Additionally, mechanisms such as production of allelochemicals, and changes in the microbial communities merit future research to achieve successful control of *L. microphyllum* and other exotic invasive species in the Everglades.

## Acknowledgement

This research was supported by the Dissertation Evidence Acquisition Fellowship, to Pushpa Soti from the Graduate School, Florida International University. Helpful comments on the manuscript were made by, Suzanne Koptur, and Michael Sukop. We also thank Stewart T. Reed, USDA, ARS, Miami, Florida for his help in analyzing the samples

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