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Low burrow occupancy and breeding success of burrowing petrels at Gough Island: a consequence of mouse predation

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Summary

The predatory behaviour of introduced house mice *Mus musculus* at Gough Island is known to impact on albatross and petrels, resulting in the Tristan Albatross *Diomedea dabbenena* and Atlantic Petrel *Pterodroma incerta* being listed as “Critically Endangered” and “Endangered”, respectively. Although predation has been documented for two burrowing petrels and one albatross species, the impact of house mice on other burrowing petrels on Gough Island is unknown. We report burrow occupancy and breeding success of Atlantic Petrels, Soft-plumaged Petrels *Pterodroma mollis*, Broad-billed Prions *Pachyptila vittata*, Grey Petrels *Procellaria cinerea* and Great Shearwaters *Puffinus gravis*. With the exception of the Great Shearwater, breeding parameters of burrowing petrels at Gough Island were very poor, with low burrow occupancy (range 4–42%) and low breeding success (0–44%) for four species, and high rates of chick mortality in Atlantic Petrel burrows. Breeding success decreased with mass, suggesting that smaller species are hardest hit, and winter-breeding species had lower breeding success than summer breeders. The results indicate that introduced house mice are having a detrimental impact on a wider range of species than previously recorded and are likely to be causing population declines among most burrowing petrels on Gough Island. The very low values of burrow occupancy recorded for Soft-plumaged Petrels and Broad-billed Prions and greatly reduced abundance of burrowing petrels in comparison to earlier decades indicate that Gough Island’s formerly abundant petrel populations are greatly threatened by the impact of predatory house mice which can only be halted by the eradication of this species from the island.

Introduction

Invasive species impact upon one third of the world’s threatened bird species and invasive mammalian predators are a major driver in the ongoing deterioration in the status of the world’s birds (Butchart 2008, Hilton and Cuthbert 2010). Among the invaders, rodents are near-ubiquitous in their ability to colonise islands and the negative impacts of rats *Rattus* spp. are well known (Moors and Atkinson 1984, Atkinson 1985, Jones *et al.* 2008). Unlike rats, the impact of the house mouse *Mus musculus* has until recently been thought to be relatively small, but the discovery that the house mouse is a predator of albatross and petrel chicks at Gough Island has demonstrated that mice can be an important threat to a wide range of seabirds (Cuthbert and Hilton 2004, Wanless *et al.* 2007, Angel *et al.* 2009). Gough Island in the South Atlantic Ocean is renowned for being one of the world’s most important seabird breeding islands (Swales 1965, Collar and Stuart 1985), supporting at least 20 species of seabird including tens of millions of burrowing petrels and albatrosses (Ryan *et al.* 2001, Cuthbert 2004, Cuthbert and Sommer 2004a). More recently, Gough Island has become as well known for the predatory behaviour of

mice that were accidentally introduced to the island by sealers in the early 19th century (Rowe-Rowe and Crafford 1992) and which have been shown to kill burrowing petrel chicks and even chicks of the Tristan Albatross *Diomedea dabbenena* (Cuthbert and Hilton 2004, Wanless et al. 2007). The impact of this predation is severe and is driving population declines of both Tristan Albatrosses and Atlantic Petrels *Pterodroma incerta* (Cuthbert 2004, Wanless et al. 2009), resulting in these species being listed as “Critically Endangered” and “Endangered”, respectively (BirdLife International 2004). The endemic Gough Bunting *Rowettia goughensis* is also “Critically Endangered” due to predation and competition with house mice (Ryan and Cuthbert 2008).

To date, direct evidence for mouse predation on Gough’s seabird species has been recorded on chicks of the Tristan Albatross, Atlantic Petrel, and Great Shearwater *Puffinus gravis* (Wanless et al. 2007). Although predation events have been recorded throughout the year, the impact of this predation appears to be most severe for winter-breeding species such as Tristan Albatross and Atlantic Petrel (Cuthbert and Hilton 2004). In contrast, summer-breeding species such as Great Shearwater, Southern Giant Petrel *Macronectes giganteus*, Atlantic Yellow-nosed Albatross *Thalassarche chlororhynchos* and Sooty Albatross *Phoebastria fusca* have high values of chick survival and overall breeding success (Cuthbert 2005, Cuthbert and Hilton 2004, Cuthbert et al. submitted.). Due to these seasonal differences in predation rates, it is likely that other winter-breeding species on Gough Island also are severely impacted by mouse predation (Angel and Cooper 2006). It is also likely that smaller petrels on Gough Island, such as storm petrels and diving petrels, also suffer high predation rates (Angel and Cooper 2006), and anecdotal observations point to large-scale decreases in the numbers of several smaller species (Ryan 2010). In this study we report rates of burrow occupancy and breeding success for a wider range of burrowing petrels on Gough Island. In light of our results we discuss the likely impact of house mice on their populations and on populations of other burrowing petrel species on the island.

Methods

Study site and species

Gough Island (40°21’S, 9°53’W) is part of the UK Overseas Territory of Tristan da Cunha. It is a volcanic island of c.65 km² in area with steep mountainous terrain. Four main vegetation types are found on Gough Island (Wace 1961, Cooper and Ryan 1994): coastal tussock (mainly comprising *Spartina arundinacea* and *Parodiocloa flabellata*), fernbush (dominated by the deciduous fern *Histiopteris incisa*, the island tree *Phylica arborea* and bog-ferns *Blechnum palmiforme*), upland wet heath (comprising a diverse assemblage of species found in all other vegetation types), feldmark, and peat bogs. There are no native terrestrial mammals and the only introduced vertebrate is the house mouse that arrived with sealers prior to the 1880s (Rowe-Rowe and Crafford 1992). At least 12 species of burrowing petrels breed on the island and are distributed within all habitats with the exception of peat bogs, areas of bare rock and the highest exposed peaks.

Fieldwork methods

Teams were based on Gough Island in three seasons (2008/09, 2009/10 and 2010/11), with fieldwork running for 13 months from September to September of each season. Studies of burrow occupancy and breeding success were carried out for five species: Atlantic Petrel, Soft-plumaged Petrel *Pterodroma mollis*, Broad-billed Prion *Pachyptila vittata*, Grey Petrel *Procellaria cinerea* and Great Shearwater. Attempts were also made to monitor breeding success of Great-winged Petrels *Pterodroma macroptera*, Common Diving Petrels *Pelecanoides urinatrix*, Grey-backed Storm Petrels *Garrodia nereis*, White-faced Storm Petrels *Pelagodroma marina* and White-bellied/Black-bellied Storm Petrels (*Fregatta grallaria/tropica*). Where applicable, data are also presented from fieldwork undertaken in 2000/01 (Cuthbert 2004, 2005, unpubl. data).

Breeding petrels were monitored through digging study-hatches in burrows (2000/2001 and 2008/09) and/or through the use of an infra-red burrow-scope (manufactured by Sextant Technologies, New Zealand) (2009/10 and 2010/11). The identity of petrel burrows was inferred from the size and shape of the burrow entrance, vegetation type and altitude (Swales 1965, Cuthbert 2004, Cuthbert and Sommer 2004b). For all species, burrows were checked in the pre-breeding period and early in the incubation period in order to identify burrows containing prospecting pairs and/or incubating birds. After an incubating bird was detected, the burrow was left undisturbed throughout the incubation period to reduce the risk of desertions. Burrows were checked 7–10 days after the estimated hatching date to quantify hatching success. The timing of breeding was based on dates reported in Ryan (2007) and the known breeding phenology for Great Shearwaters and Atlantic Petrels (Cuthbert 2004, 2005) or from the relationship between petrel egg size and incubation period (from Warham 1990). After hatching, some burrows were checked every 5–14 days until fledging, while others were checked only once, close to the end of the chick-rearing period but prior to the estimated fledging date. Large chicks which had grown contour feathers and lost most of their down were assumed to survive to fledge. A further sample of Broad-billed Prion chicks was observed at a cave site on the island, as large numbers of prions breed in caves on Gough Island (Cuthbert *et al.* 2011). This cave was visited at monthly intervals from 20 January 2009 when small, newly hatched chicks were present and observations were made on the presence of dead or injured chicks.

For all species, we report values of burrow occupancy (the proportion of burrows occupied by breeding pairs) and calculate maximum breeding success based on the proportion of incubating eggs producing a fledged chick. Mean estimates are presented \pm one standard error. The total sample of burrows monitored included a proportion of burrows where the contents could not be determined, as some burrows were either, too long and twisted, or too complex, to determine the contents with complete confidence. As a consequence, estimates of burrow occupancy were only calculated for burrows where it was certain that they were empty or contained an incubating bird or pair. Due to the similar size of the species it was difficult to separate potential burrows of Soft-plumaged Petrels from those of Broad-billed Prions, although it is likely that both species use the same burrows. As a consequence we present two estimates of burrow occupancy: (a) the number containing incubating birds divided by the total number of burrows searched and (b) the number containing incubating birds divided by the number of burrows searched corrected for the relative ratio of Broad-billed Prions to Soft-plumaged Petrels (a ratio of 2.17 : 1; Cuthbert 2004). This second estimate provides a more conservative estimate of burrow occupancy. Because house mice are known to predate even newly hatched chicks being brooded by adult birds (Wanless *et al.* 2007), estimates of hatching success are only reported for Great Shearwaters where the timing of breeding is known and highly synchronous (Cuthbert 2005) and where chicks could be checked in the days immediately after hatching. Due to the timing of breeding of Atlantic Petrels (which lay eggs from mid-June to mid-July, hatch chicks from mid-August to mid-September and fledge during December and January; Cuthbert 2004) and the timing of field visits to the island (from September to September) “burrow occupancy” is reported for small chicks in September and for large, nearly fledged chicks in December, and chick survival is estimated for this 3-month period.

Results

Atlantic Petrel

The contents of 206 burrows were determined with a burrow-scope in September 2010, with additional data from 50 and 63 burrows with study hatches during 2000 and 2001, respectively (Table 1). Burrow occupancy with small chicks was consistent for these three seasons, ranging from 23 to 28% and an overall mean of $26 \pm 1.4\%$. Chick survival from September to December was variable (Table 1), with survival averaging $49 \pm 10\%$ for this 3-month period in the three monitored seasons. Totals of 9, 10 and 6 dead chicks were found within burrows in the 2000, 2001

Table 1. Breeding data for Atlantic Petrels on Gough Island indicating season, number of burrows with known contents, numbers of small chicks in September, number of large chicks in December and corresponding values of burrow occupancy (%) and chick survival (%) for September-December.

Season	Burrows	Sep chicks	Dec chicks	Chick occupancy	Survival
2000	50	14	5	28%	36%
2001	63	17	7	27%	41%
2010	206	48	33	23%	69%

and 2010 seasons, respectively, giving minimum chick mortality estimates in these three seasons (from September to December) of 64%, 59% and 13%, respectively and mean chick mortality of $45 \pm 15\%$. Regular inspection (every 7–10 days) of these Atlantic Petrel burrows indicated that chicks were alive and in apparently good condition and health prior to being found dead in the burrow: suggesting a rapid cause of mortality likely to have been mouse predation.

Soft-plumaged Petrel

A total of 333 Soft-plumaged Petrel sized burrows were checked with a burrow-scope during November and December 2009, with occupancy determined for 281 burrows (84.4%). Burrow occupancy for this period was estimated to be 11–18% (Table 2). Egg-laying for this species occurred from 23 November to 2 December. Of 22 burrows with incubating birds, only 3 produced fledged chicks, giving a crude breeding success of 14%. Limited data from 2000/2001 indicated a breeding success of 44% (RJC unpubl. data); the pooled estimate of breeding success for both years is $26 \pm 15\%$.

Broad-billed Prion

A total of 579 burrows were checked with a burrow-scope in 2009/2010 and 2010/2011, with the occupancy status determined for 461 burrows (79.6%; Table 2). Burrow occupancy was very low in both seasons with just 4–12% of burrows estimated to contain incubating birds. While the sample sizes of incubating birds are very small, values of breeding success were poor in both monitored seasons (Table 2), with an overall pooled estimate of breeding success of $6 \pm 5\%$. Data from the cave site indicated large numbers of birds breeding at this site, with 60 chicks found on 20 January 2009. Further checks of this site revealed the presence of eggshells and dead chicks that had incisor marks on eggs and chick carcasses consistent with house mouse predation. These records include a large, nearly fledged chick which was found with feeding signs indicating it had been partly consumed by mice.

Table 2. Breeding data for Soft-plumaged Petrels, Broad-billed Prions and Grey Petrels on Gough Island indicating season, number of burrows with known contents (values in parentheses are total numbers of burrows checked), numbers of eggs monitored and chicks fledged, and corresponding values of burrow occupancy (%) and breeding success (%).

Species	Season	Burrows	Occupied	Eggs	Fledge	Occupancy	Breed
Soft-plumaged Petrel	2000-01	-	-	16	7	-	44%
	2009-10	281 (333)	32	22	3	11-18%	14%
Broad-billed Prion	2009-10	256 (360)	11	11	1	4-13%	9%
	2010-11	205 (219)	8	7	0	4-11%	0%
Grey Petrel	2009	-	-	26	8	-	31%
	2010	50 (141)	21	17	6	42%	35%

Grey Petrel

A total of 141 burrows were checked during 2010, with the contents of 50 burrows determined with certainty with a burrow-scope (35.5%). The remaining 91 burrows were too long, wet or twisted for the burrow-scope. Burrow occupancy was 42% (Table 2). Estimates of breeding success from the 2009 and 2010 seasons were similar (Table 2) with an overall pooled average of $33 \pm 2\%$.

Great Shearwater

Burrow occupancy and breeding success of Great Shearwaters were monitored in four years, with 67–180 burrows monitored each year (Table 3). Burrow occupancy was consistently high in all four seasons: $68.7 \pm 2.7\%$ (range 65–77%). Values of hatching success, fledging success and overall breeding success averaged $56.9 \pm 8.5\%$, $81.7 \pm 22.1\%$ and $43.2 \pm 9.4\%$, respectively. Fledging success was high in two seasons (93–96%) and lower (56%) in one season (Table 3).

Discussion

Globally, seabirds are among the most threatened groups of birds with 97 of 346 species (28%) listed as globally threatened, including 17 species classified as “Critically Endangered” (Croxall *et al.* 2012). On their breeding grounds, the gravest threat to seabirds is the impact of introduced mammalian predators (Baker *et al.* 2002, Croxall *et al.* 2012), with invasive rodents being among the most serious invasive species (Jones *et al.* 2008). Our results from Gough Island present a worrying situation for almost all of the burrowing petrels studied here, with low values of burrow occupancy and breeding success recorded for Atlantic Petrel, Broad-billed Prion, Soft-plumaged Petrel and Grey Petrel. The ranges of breeding success recorded for Atlantic Petrel (20–23%; Cuthbert 2004, Wanless *et al.* 2012), Soft-plumaged Petrels (14–44%), Broad-billed Prions (0–9%) and Grey Petrels (31–35%) from Gough Island are well below the breeding success of most petrels in areas free from introduced predators (Table 4). Only Great Shearwaters had levels of burrow occupancy and breeding parameters that can be considered normal for *Puffinus* species (Cuthbert and Davis 2002; Table 4).

We consider the predatory behaviour of house mice at Gough Island to be the most logical explanation for these low values of reproductive success. Mouse predation of chicks has been confirmed by direct observation for Atlantic Petrels, Great Shearwaters and Tristan Albatrosses on Gough Island (Wanless *et al.* 2007) and patterns of injuries indicate predation also occurs on nestlings of Gough Buntings (Ryan and Cuthbert 2008). Further records now suggest that mice also predate Atlantic Yellow-nosed and Sooty Albatross chicks (RSPB unpubl. data). Further direct evidence for predation was recorded for Broad-billed Prions in the current study, where both eggs and chicks of birds breeding in caves (where mice are abundant; Cuthbert *et al.* 2011)

Table 3. Breeding data for Great Shearwaters on Gough Island indicating season, number of burrows with known contents, numbers of eggs monitored, number of hatched chicks and number of chicks fledged, and corresponding values of burrow occupancy (%), hatching success (%), fledging success (%) and breeding success (%).

Season	Burrows	Eggs	Hatch	Fledge	Occupancy	Hatching	Fledging	Breeding
2000-01	67	44	24 (30) ¹	(28) ¹	65.7%	55.6%	93.0%	52.0%
2008-09	80	54	32	18	67.5%	59.3%	56.3%	33.3%
2009-10	80	52	24	23	65.0%	46.2%	95.8%	44.2%
2010-11	180	138 (126) ²	84	-	76.7%	66.7%	-	-

¹An additional 6 hatched chicks were added to the study: fledging and breeding success include these.

²A total of 138 burrows were found with eggs and 126 were monitored for hatching success.

Table 4. Measured values of breeding success for species on Gough Island (in bold type) and values for the same species or congeneric species at sites with no, or low rates, of predation rates.

Species	Breeding success %	Reference
<i>Puffinus gravis</i>	43%	This study
<i>Puffinus huttoni</i>	47%	Cuthbert and Davis (2002)
<i>Puffinus puffinus</i>	60%	Thomson and Furness (1991)
<i>Puffinus tenuirostris</i>	48%	Serventy and Curry (1984)
<i>Puffinus pacificus</i>	43%	Floyd and Swanson (1983)
<i>Pterodroma mollis</i>	14-44%	This study
<i>Pterodroma incerta</i>	20-23%	Cuthbert (2004), Wanless <i>et al.</i> (2012)
<i>Pterodroma macroptera</i>	60%	Cooper and Fourie (1991)
<i>Pterodroma pycrofti</i>	59%	Pierce (2002)
<i>Pterodroma hypoleuca</i>	82%	Seto and Conant (1996)
<i>Pterodroma axillaris</i>	71-78%	Brooke (2004)
<i>Pterodroma leucoptera</i>	50% ¹	DEC-NSW (2006)
<i>Pachyptila vittata</i>	0-9%	This study
<i>Pachyptila vittata</i>	60-72% ²	van Rensburg and Bester (1988)
<i>Pachyptila belcheri</i>	52%	Nevoux and Barbraud (2005)
<i>Pachyptila belcheri</i>	63% ³	Catry <i>et al.</i> (2007)
<i>Procellaria cinerea</i>	31-35%	This study
<i>Procellaria cinerea</i>	40-60%	Inchausti <i>et al.</i> (2003)
<i>Procellaria cinerea</i>	75-80%	Schulz <i>et al.</i> (2005)

¹Following predator control and habitat management.

²In predator free enclosures.

³In lowland areas with minimal predation rates.

were found with injuries consistent with mouse predation (Brown *et al.* 1998; Cuthbert and Hilton 2004). While breeding success of many seabirds can be highly variable, often in response to at-sea conditions (Warham 1990), we consider it very unlikely that environmental conditions would be responsible for consistently low breeding success and low burrow occupancy across four different species and over several years of monitoring. A further cause of low breeding success could be due to high rates of burrow competition, with larger species (such as Great Shearwaters) displacing smaller species on Gough, as has been observed for other seabirds (e.g. Ramos *et al.* 1997, Was *et al.* 2000). While such competition may occur on Gough the low levels of burrow occupancy seems likely to reduce the scale and impact of this upon breeding success. It is exceptional for house mice to be predators on the scale observed at Gough Island (Cuthbert and Hilton 2004, Wanless *et al.* 2007), and our study indicates that this source of mortality is impacting a wider range of species than previously considered.

Given the low values of breeding success observed in four species of petrels at Gough Island, are their populations likely to be stable or declining? Petrels are able to withstand several years of poor breeding success as they are long-lived, and even with consistently low breeding success, population declines are likely to be slow (e.g. Brooke *et al.* 2010). Population modelling undertaken for the Henderson Petrel *Pterodroma atrata* suggests that with values of annual adult survival set at 93%, breeding success needs to average more than 25% in order for the population to be close to stable (Brooke *et al.* 2010). Values of breeding success for both *Pterodroma* species on Gough are close to this level, suggesting that despite poor breeding success populations could be close to stable provided adult survival is sufficiently high. Published information for Grey Petrels is limited, but values of breeding success from Kerguelen of 40–60% (Inchausti *et al.* 2003) and Macquarie Island 75–80% (Schulz *et al.* 2005) indicate that in comparison breeding success for this species on Gough Island is low (Table 4). As well as suffering

low breeding success on Gough Island, Grey Petrels are also often killed by commercial fishing operations (Barbraud *et al.* 2009) and in combination these two sources of mortality could be leading to population declines. Population modelling predicts a similar scenario for the Tristan Albatross, which like the Grey Petrel is impacted at sea and on land (Wanless *et al.* 2009). No demographic data are available for Broad-billed Prions, but with a breeding success of just 0–9% the species is almost certainly declining.

At least 12 species of burrowing petrel breed at Gough Island (Table 5) and our results demonstrate that breeding success was low in four out of five species studied. Angel and Cooper (2006) hypothesised that winter-breeding species and smaller species (diving petrels and storm petrels) are likely to be most impacted by mouse predation on Gough Island, and other studies have reported that invasive rats have a greater impact on smaller seabird species (Baker *et al.* 2002, Jones *et al.* 2008). These hypotheses are supported by our data which indicate a significant relationship between body size and breeding success as well as lower values of breeding success for the two winter breeding species (Figure 1). If this relationship holds for all species of burrowing petrel on Gough Island then we predict that the impact of house mice on the remaining species is likely to be severe, as they are either winter breeders (Great-winged Petrel) or of small body size (< 350 g; Table 5 and Figure 1). Given these results we predict that wide-scale population declines are occurring among most burrowing petrels on Gough Island. The very low rates of burrow occupancy found for Broad-billed Prions (4–12%) and Soft-plumaged Petrels (11–18%) suggest that large-scale population declines may already have taken place for these two species, although some caution should be attached to these occupancy figures as it is difficult to separate burrows of the two species and large numbers of prions breed in caves on the island. If burrow occupancy values found in this study are applied to the population estimates in Cuthbert (2004) then populations of both Broad-billed Prions and Soft-plumaged Petrels could be far lower than previously estimated.

Earlier reports and observations of burrowing petrels at Gough Island lend some further credence to the fact that populations of many species have declined. The first detailed ornithological survey of Gough Island in 1955/56 (Swales 1965) described Little Shearwaters *Puffinus assimilis* as “extremely abundant”, recorded that Kerguelen Petrels *Lugensa brevirostris* were so abundant that “so many thousands... were attracted to the lights... that it became necessary to turn them off

Table 5. Species of burrowing petrel on Gough Island indicating breeding season, breeding success, body mass and likely population trend based on the results of this study (underlined) and inferred population trend for other species (italics). Data for body mass are from Brooke (2004), Cuthbert (2004, 2005), Hockey *et al.* (2005) and authors’ unpubl. data.

Species	Breeding season	Breeding success (%)	Body mass (g)	Trend
Great Shearwater <i>Puffinus gravis</i>	Summer	43%	737–1130	Stable
Soft-plumaged Petrel <i>Pterodroma mollis</i>	Summer	26%	202–342	<u>Decrease</u>
Little Shearwater <i>Puffinus assimilis</i>	Summer	Unknown	210–295	<i>Decrease</i>
Broad-billed Prion <i>Pachyptila vittata</i>	Summer	6%	155–235	<u>Decrease</u>
Grey Petrel <i>Procellaria cinerea</i>	Winter	33%	950–1220	<u>Decrease</u>
Great-winged Petrel <i>Pterodroma macroptera</i>	Winter	Unknown	460–745	<i>Decrease</i>
Atlantic Petrel <i>Pterodroma incerta</i>	Winter	20%	420–720	<u>Decrease</u>
Kerguelen Petrel <i>Pterodroma brevirostris</i>	Summer	Unknown	220–450	<i>Decrease</i>
Grey-backed Storm Petrel <i>Garrodia nereis</i>	Summer	Unknown	26–40	<i>Decrease</i>
White-faced Storm Petrel <i>Pelagodroma marina</i>	Summer	Unknown	45–60	<i>Decrease</i>
<i>Fregatta</i> storm petrels (<i>F. grallaria/tropica</i> *)	Summer	Unknown	45–65	<i>Decrease</i>
Common Diving Petrel <i>Pelecanoides urinatrix</i>	Summer	Unknown	100–160	<i>Decrease</i>

*recent evidence suggests that both *F. grallaria* and *F. tropica* breed on Gough Island.

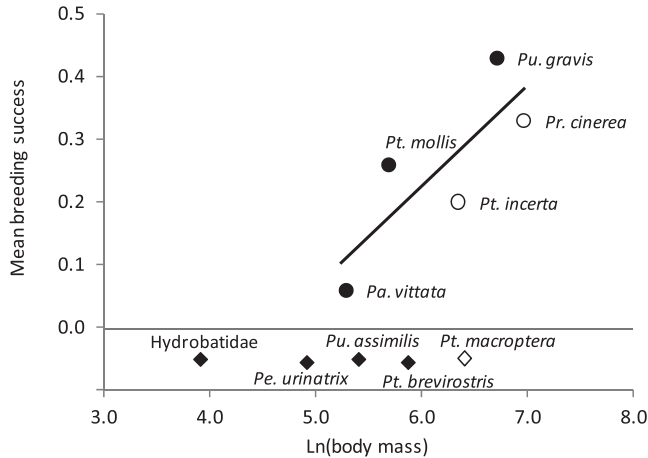


Figure 1. Relationship between Ln body mass (g) and mean breeding success for monitored populations of burrowing petrels reported in this paper indicated by filled circles for summer breeding species and open circles for winter breeding species (the fitted correlation line for the 5 species is statistically significant: $R^2 = 0.806$, $P < 0.05$) and Ln body mass of unstudied species on Gough Island below with summer and winter breeding indicated by filled and unfilled diamond symbols. Species codes are: Atlantic petrel (*Pt. incerta*), soft-plumaged petrel (*Pt. mollis*), great-winged petrel (*Pt. macroptera*), Kerguelen petrel (*Pt. brevirostris*), grey petrel (*Pr. cinerea*), great shearwater (*Pu. gravis*), little shearwater (*Pu. assimilis*), broad-billed prion (*Pa. vittata*), common diving petrel (*Pe. urinatrix*) and storm petrels (Hydrobatidae).

again at once for the safety of the observers, the birds and the lights themselves”, reported “large numbers” of Great-winged Petrels and described Common Diving Petrels as “abundant”. Similarly, Ryan (2010) estimates on a single misty night in October 1984 that there were “tens of thousands of Broad-billed Prions, thousands of Soft-plumaged and Kerguelen Petrels, and hundreds of Atlantic Petrels, Little Shearwaters, Common Diving-petrels and White-faced and Grey-backed storm-petrels”. By contrast numbers of birds seen spot-lighting at the same location over many nights in September–October 2009, recorded no Kerguelen Petrels, one Little Shearwater and two Grey-backed Storm Petrels. Ryan (2010) also reports that Little Shearwaters used to be commonly heard calling in tussock areas in the southeast of the island during the mid 1980s but are now no longer recorded. While there are obviously limitations to these conclusions based on a single night’s observations from the mid- 1980s, nonetheless it is striking that no similar observations of numbers on this scale have been recorded during the last decade, when year-round teams of fieldworkers have been present for five years since 2000/01. Despite specific searches, they have failed to locate any breeding burrows of Great-winged Petrels, Little Shearwaters, Common Diving Petrel or storm petrels, and have only found a few Kerguelen Petrel burrows. Some caution should be applied to comparing the surveys from 1955/56 with results from the last decade, as the earlier surveys were undertaken on the east coast at The Glen, whereas most recent fieldwork has been undertaken in upland areas of the island and at Transvaal Bay. Nonetheless, despite some evidence from birds attracted to the lights of fishing boats anchored offshore from The Glen that this site may hold greater numbers of Little Shearwaters and Kerguelen Petrels than other areas (PGR and ES pers obs.) it is clear from overnight visits to the site that there are considerably fewer birds present than described by Swales (1965). While these comparisons from The Glen are qualitative, the observations of Ryan (2010) from 1984 and 2009 were undertaken at the same site and provide stronger evidence for large-scale declines.

House mice have been present on Gough Island since the early 19th century (Rowe-Rowe and Crafford 1992) and given that numbers of burrowing petrels appeared to remain abundant on Gough Island until at least the 1960s, the question arises as to why large scale impacts of mice and apparent reductions in petrel numbers have only been observed in the last few decades? The negative impacts of rats on seabird populations are usually acute and relatively rapid (Atkinson 1985), although in some circumstances gradual population declines can take place over hundreds of years (Brooke *et al.* 2010). Comparable knowledge on the impacts of House mice on seabirds is lacking, because large-scale predatory behaviour by the species appears to be the exception rather than the rule. Given the impacts that mice are now having on Gough Island and the presence of very large numbers of birds until the mid 20th century we consider it likely that, for currently unknown reasons, the predatory behaviour of mice has increased in recent decades to the point whereby the impact of predation is now observed across most species on the island.

In conclusion, this study indicates that the predicted impacts of house mice on winter-breeding petrels (Angel and Cooper 2006) are indeed severe, and moreover that several summer-breeding petrels appear to be impacted, especially the smaller species. If the levels of breeding success observed in this study are typical, populations of most burrowing petrels on Gough Island are likely to be declining. Very low levels of burrow occupancy by two formerly abundant species and comparison of numbers to earlier records strongly suggest that large-scale declines may have already taken place for many species and that Gough Island's burrowing petrel community could be at the start of a catastrophic collapse. The only feasible and practical solution to this situation is the eradication of house mice from the island. This conservation action previously has been recommended to protect the "Critically Endangered" Tristan Albatross and Gough Bunting, and the "Endangered" Atlantic Petrel (Cuthbert and Hilton 2004, Wanless *et al.* 2009). Our study indicates that a far larger range of species would benefit from removing mice from the island. The eradication of invasive rodents has become one of the most important conservation actions available for protecting native species on islands (Howald *et al.* 2007) and while eradicating house mice from Gough Island would be challenging and expensive, a review has indicated that it should be feasible (Parkes 2008) and planning for this operation is underway (Torr *et al.* 2011). If an eradication operation is successful, populations of burrowing petrels should recover relatively rapidly as large numbers of adult breeding birds remain and improved breeding success should occur immediately after the removal of mice. A successful eradication would remove the key conservation threat to more than ten seabird species and restore Gough Island's status as one of the world's most important seabird breeding islands.

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