

# Effect of canopy structure on the performance of tree mapping methods in urban parks

Tanhuanpää, T., Yu, X., Luoma, V., Saarinen, N., Raisio, J., Hyyppä, J., Kumpula, T., Holopainen, M.

doi:<https://doi.org/10.1016/j.ufug.2019.126441>

## Abstract

Urban forests consist of patches of recreational areas, parks, and single trees on roadsides and other forested urban areas. Large number of tree species and heterogeneous growing conditions result in diverse canopy structure. Large variation can be found both at the level of single tree crowns and canopy characteristics of larger areas. As urban forests are typically managed with small-scale, even tree-level operations, there is a need for detailed forest information. In this study, the effect of varying canopy conditions was tested on nine tree mapping methods. All methods utilized canopy height models (CHM) derived from dense airborne laser scanning data. The mapping methods utilized modified watershed segmentation differed in the intensity of CHM filtering. The performance of mapping methods was compared in different canopy conditions. The results showed considerable variation between the methods when tested in varying canopy conditions. Especially, presence of large broadleaved trees affected the accuracy of detecting individual trees.

Key words: LiDAR, Terrestrial laser scanning, Urban forest, Trees outside forests, Airborne laser scanning

## Introduction

There is a growing interest in more detailed information on urban trees. Typically, tree-level data are needed for maintaining the urban tree reserve (allocating tree maintenance etc.) but during recent years, the interest has grown to apply tree-level information for defining the ecosystem services (ES) provided by the trees. Mapping of ES such as storm water catchment or air pollutant removal would likely benefit from detailed tree-level data. Quantifying and mapping the ES provided by urban trees helps in justifying the costs needed for tree management as well as assessing the ES-related benefits in regional level. As manual generation and maintenance of tree-level data is time consuming and thus costly, methods utilizing various high-resolution remote sensing (RS) material have been implemented for detecting and monitoring urban trees.

During the last decade, airborne laser scanning (ALS) datasets have become a common tool for urban mapping and planning (e.g., elevation models and building delineation in urban areas). Therefore, they have a high potential in operative mapping of urban vegetation (Lee et al., 2016; Saarinen et al., 2014; Moskal et al., 2011). The main advantage of ALS data is the ability to capture the vertical structure of forest stands with high detail and accuracy (e.g., White et al., 2015; Suárez et al., 2005). Potentially high spatial resolution, i.e., point density, enables detection of individual tree crowns (e.g., Tanhuanpää et al., 2016; White et al., 2013; Hyyppä and Inkinen, 1999). On the other hand, the spectral resolution has typically been limited to a single wavelength, which makes species interpretation challenging (Kim et al., 2011; Vauhkonen et al., 2009). Hence, ALS datasets have often been accompanied by spectral information from another RS data source (e.g., Alonzo et al., 2014; Koukoulas and Blackburn, 2005).

Delineating individual trees from ALS data has been widely studied, especially in the field of forestry (Mongus and Zalik, 2015; Duncanson et al., 2014; Hyyppä et al., 2012). Individual tree detection (ITD) aims at mapping all trees from a given area and estimating the trees' key attributes. Depending on the attribute of interest

42 and the materials used, tree attributes can be modeled with combined use of ALS and field data (e.g.,  
43 Tanhuanpää et al., 2014; Vauhkonen et al., 2010; Persson et al., 2002) or measured directly from ALS data  
44 (Tanhuanpää et al., 2015; Vauhkonen, 2010; Popescu and Zhao, 2008). A central hindrance in ALS-based ITD  
45 becoming an operational state-of-the-art for producing tree maps for large areas has been the sensitivity of  
46 the approach to the spatial structure of the canopy. Depending on the density and spatial distribution of the  
47 canopy (i.e., forest structure), ITD methods often suffer from either omission or commission errors (Eysn et  
48 al., 2015; Vauhkonen et al., 2011). Omission errors typically occur in multi-layered canopy conditions, where  
49 crowns of larger and older trees cover smaller trees. On the other hand, wide and forked crowns of old  
50 broadleaved trees are likely to cause commission errors.

51 In the last decade, the interest towards utilization of point clouds from digital aerial photogrammetry (DAP)  
52 in providing information about forest resources has increased (White et al. 2016). This is partly due to the  
53 long tradition of utilizing aerial images in forest inventories but mostly because of the cost of DAP has been  
54 estimated to be one-third to one-half of the cost of ALS data (White et al. 2013). Although the capability of  
55 DAP point clouds in penetrating the canopy is not comparable with ALS (Vastaranta et al. 2013) comparable  
56 results between ALS and DAP in identifying individual trees has been obtained (Rahlf et al. 2015, St-Onge et  
57 al. 2015). Tanhuanpää et al. (2016) concluded that although plot-level mean height can be similar to ALS-  
58 based, identifying individual trees and estimating their height with DAP remains challenging. Additionally,  
59 DAP point clouds tend to require an accurate digital terrain model (DTM) that is usually produced with ALS  
60 data due to its capability of providing observations also below the canopy.

61 Several ALS-based tree delineation methods have been published during the last two decades (see review,  
62 Lindberg and Holmgren, 2017). Basically, the methods can be divided into two core approaches where the  
63 ALS point clouds are analyzed either directly (e.g., Lu et al., 2014; Wang et al., 2008) or by first simplifying  
64 them into canopy height models (CHM) that describe the outer envelope of the canopy (e.g., Yu et al., 2011).  
65 The direct analysis of point clouds enables the use of all the data in the delineation process, whereas in the  
66 CHM approach most of the data are typically discarded. The main benefit from simplifying the data into a  
67 CHM is that further processing of the data is computationally less intensive.

68 Previous studies have shown that the performance of different ITD methods varies between different canopy  
69 conditions (e.g., Eysn et al., 2015; Vauhkonen et al., 2011). It has also been reported that, due to their simpler  
70 crown structure, coniferous trees can be detected more accurately than deciduous trees (Jing et al., 2012;  
71 Koch et al., 2006). However, majority of the ALS-based ITD studies have been conducted in forest  
72 environment. Urban forest and trees are often interwoven with buildings, poles, wires, and other built  
73 objects common in urban environment. Because of the special qualities, urban areas are challenging  
74 environments for classification of trees and other objects (e.g., Rottensteiner et al., 2014). In terms of forest  
75 characteristics, urban forests are typically dominated by deciduous tree species and characterized by  
76 heterogeneous and fragmented canopy. Both CHM and point cloud-based ITD methods have been studied in  
77 urban surroundings. In Holopainen et al. (2013), dense ALS data were utilized in CHM-based method for  
78 detecting single trees in urban park area. The automatic method found 65.5% of the trees with DBH of 5 cm  
79 or above. Zhang et al. (2015), detected urban park trees directly from ALS point clouds, thus finding 90.6 %  
80 of the trees with DBH over 10.16 cm (4 inches). To our best knowledge urban comparisons on ITD methods'  
81 performance have not been made. Hence, there is little information available on the effects of canopy  
82 characteristics on ITD or the superiority of ITD methodologies in urban surroundings.

83 Watershed segmentation (WS) and its variations are commonly used in CHM-based approaches for  
84 delineating individual tree crowns (Duncanson et al., 2014; Koch et al., 2006, Hyyppä and Inkinen, 1999). In  
85 WS, the canopy envelope is typically inverted and treated as the floor of a water body. Local minima (i.e., the  
86 sinks of the CHM) are then used as seed points to delineate the basins, i.e., the individual tree crowns. As a  
87 tree crown can result in several local minima in a CHM, a central part of the WS process is to reduce the

88 number of possible seed points. The reduction can be done by using a threshold distance between  
89 neighboring seed points (e.g., Yu et al., 2011). The assumption is that the proportion of tree height and crown  
90 width is rather constant within the study area. The other option is to smooth the CHM, using either varying  
91 or fixed intensity smoothing kernel (e.g., Tanhuanpää et al., 2016; Hyyppä and Inkinen, 1999). For both  
92 approaches, finding the optimal number of seed points requires a priori knowledge about the tree crowns'  
93 height-width ratio. This is problematic in urban surroundings where high spatial variation of species and  
94 growth environments complicates choosing the correct number of seed points.

95 Earlier studies have pointed out the significance of deciduous trees for the accuracy of ITD (e.g., Koch et al.,  
96 2006). The large trees affect the crown delineation in two levels. Firstly, the crown shape affects how the  
97 tree itself appears in the data, and secondly the crown size and shape affect the detectability of the  
98 neighboring trees. It is typical for old trees to have several distinguishable tops within the crown. This type  
99 of crowns often cause commission errors in crown delineation, as each top act as a seed point in  
100 segmentation process. Here, threshold values are used to control the number of resulting tree segments. The  
101 value is used in determining the minimum distance between detectable trees (Yu et al., 2011). However, the  
102 growth pattern of urban open-grown trees has been reported to differ from that of closed canopy conditions,  
103 as bigger share of biomass is allocated in tree crown (e.g., Riikonen et al., 2017; Tanhuanpää et al., 2017). In  
104 varying light and other growth conditions the relation between tree height and crown width varies  
105 considerably, which complicates choosing the correct threshold value. If the value is defined too low, the  
106 single crown is split into several crown segments within the delineation process. On the other hand, too high  
107 threshold value causes losing the smaller tree crowns. The same logic also affects the smoothing of CHMs.  
108 Choosing the level of smoothing according to the large deciduous trees is likely to result in losing the smaller  
109 trees (Tanhuanpää et al., 2014).

110 Because of the heterogeneous nature of urban forest canopy, a widely applicable urban ITD method should  
111 be able to adapt to a wide range of variation in crown shapes and sizes. In this study, performance of nine  
112 ITD approaches was tested in heterogeneous urban park areas. The aim of the study was threefold.

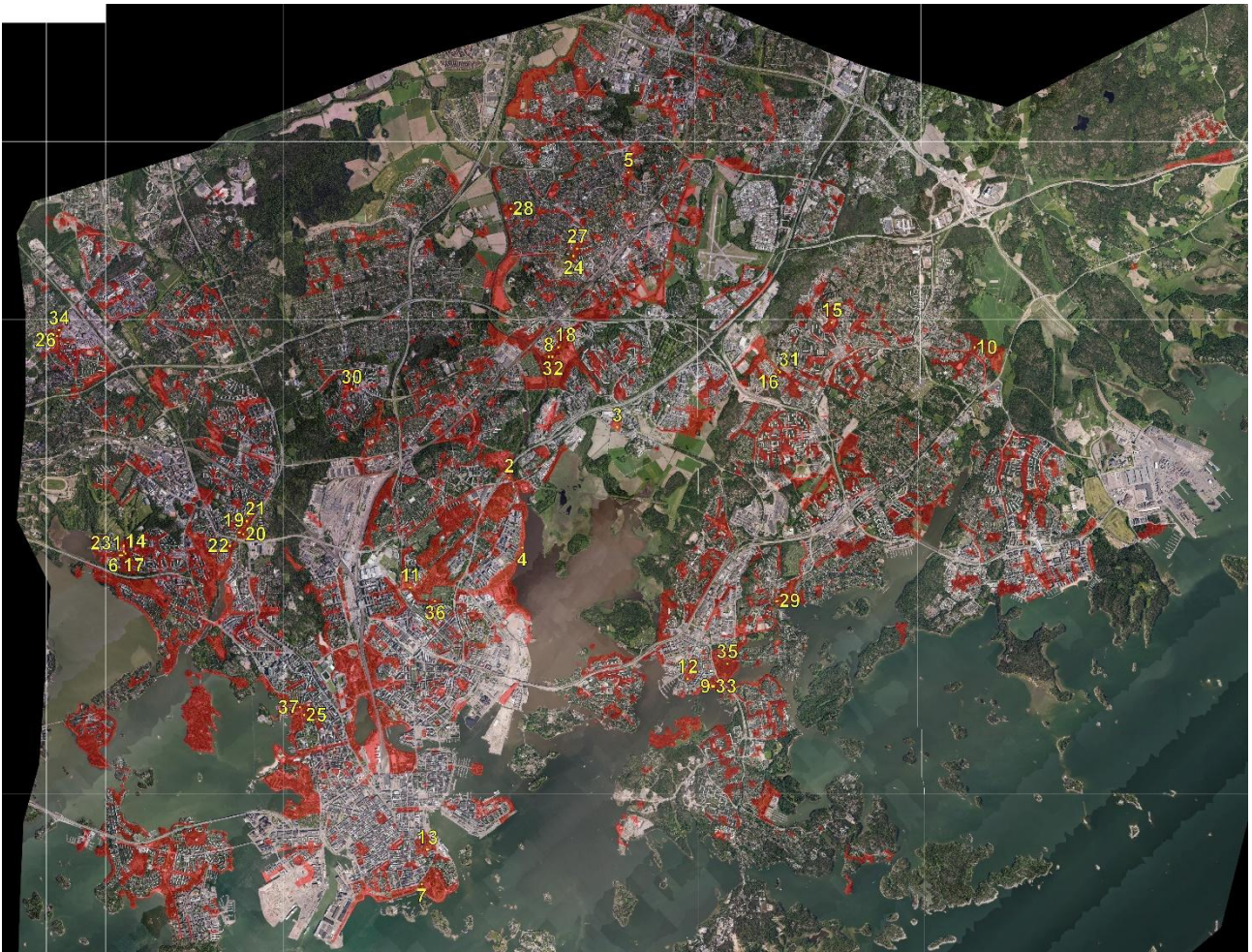
113 Firstly, we wanted to investigate the effect of heterogeneous urban canopy conditions on the performance  
114 of ITD methods. Assumption was that park areas with lower canopy cover would offer more favorable  
115 conditions for ITD than denser areas. Secondly, as large trees have been found to complicate especially urban  
116 ITD, we explored their effect on plot-level accuracy of the ITD methods. Thirdly, expecting that none of the  
117 methods would outperform others in all canopy conditions, we wanted to define the best methods for  
118 different conditions and thus clarify the requirements of a universal ITD methodology oriented to  
119 heterogeneous urban park areas.

120 Material and methods

### 121 *Study area*

122 The study was conducted in the city of Helsinki, located in southern Finland (60°10'10.27"N, 24°56'7.62"E,  
123 Figure 1). The study area consisted of 2000 hectares of parks and recreational urban forests. The area was  
124 separated from other urban areas according to a property classification maintained by the City of Helsinki.  
125 The canopy structure of the selected areas varied from isolated individual tree crowns to multi-layer  
126 recreational areas amidst the city. The species composition ranged from pure deciduous parks to conifer  
127 dominated forest-like conditions.

128



129

130 Figure 1: Aerial image of the study area (in red) and the sample plot locations (yellow numbering).  
 131

132 *ALS data*

133 The City of Helsinki supplied The ALS data. The data were collected in eight separate flights between May  
 134 20<sup>th</sup> and July 7<sup>th</sup>, 2015 using Leica ALS-70 HP and ALS-70 CM scanners. Scanning altitude was 500 m above  
 135 ground level and side overlap between neighboring flight lines was 50%. Pulse density at nadir was 20 p/m<sup>2</sup>.

136 *CHM generation*

137 ALS data were simplified into a CHM for the needs of park stratification and tree delineation. All calculations  
 138 were done in Terra Scan software (Terrasolid, Helsinki, Finland). The CHM was created by first triangulating  
 139 a digital elevation model (DEM) (see, Axelsson, 2000) and generating a digital surface model (DSM),  
 140 representing the outer envelope of the canopy, from the highest echoes. The cell sizes were 1.0 m x 1.0 m  
 141 and 0.5 m x 0.5 m for the DEM and DSM, respectively. Smaller cell size was used for DSM for capturing the  
 142 fine-scale variation in tree crowns. The CHM was generated by subtracting the DEM values from the values  
 143 of DSM. The spatial resolution of the final CHM was 0.5 m. ALS-based CHM is visualized in Figure 2.





144

145 Figure 2: Plot number 37 visualized in an aerial image (left) and an ALS-based canopy height model (right) over an  
 146 aerial image.

147

148 *Stratification of park areas*

149 For capturing the variation within the park areas, the whole study area was first divided in three strata in  
 150 terms of canopy cover and mean height. As open grown trees have typically wider crowns than forest trees,  
 151 we used 32 m by 32 m grid for the stratification. This resulted in four times larger plots compared to  
 152 operational forest attribute mapping carried out by the Finnish Forest Centre (Suomen metsäkeskus). The  
 153 grid was placed over the park areas and each grid cell fully inside park area was characterized by ALS-derived  
 154 mean height ( $H_{\text{mean}}$ ) and canopy cover-% (CC). Both figures were achieved by applying simplified WS (i.e., all  
 155 peaks in the CHM were considered seed points) with cut height of 2 m (i.e., CHM values under 2 m were  
 156 ignored) to the unfiltered ALS-based CHMs. CC was calculated as the proportional area of canopy polygons  
 157 inside each 32 m by 32 m grid cell. Similarly,  $H_{\text{mean}}$  was calculated for each grid cell as a mean of height maxima  
 158 of the canopy polygons. Hence,  $H_{\text{mean}}$  represents the height of vegetation peaks in CHM, not the mean height  
 159 of ALS points. The stratification was used for allocating field plots over the study area proportionally to the  
 160 total area of each strata (Table 1).

161 Table 1: Definition and key characteristics of the field plots within the park strata.  $n_{\text{plot}}$  stands for the number of plots,  
 162  $n_{\text{tree}}$  for the total number of trees,  $CC_{\text{mean}}$  for the ALS-derived mean canopy cover,  $H_{\text{mean}}$  for the ALS-derived mean  
 163 height,  $DBH_{\text{mean}}$  for the mean,  $DBH_{\text{min}}$  for the minimum, and  $DBH_{\text{max}}$  for the maximum field-measured DBH within each  
 164 stratum.

Stratum	Definition	$n_{\text{plot}}$	$n_{\text{tree}}$	$CC_{\text{mean}}$	$H_{\text{mean}}$	$DBH_{\text{mean}}$	$DBH_{\text{min}}$	$DBH_{\text{max}}$
1	CC < 50 %	12	82	26 %	8.0 m	20.6	6.0	108.2
2	CC > 50 % and $H_{\text{mean}} < 15$ m	15	453	74 %	10.5 m	19.6	6.0	65.1
3	CC > 50 % and $H_{\text{mean}} > 15$ m	10	277	83 %	18.0 m	25.6	6.2	116.2
<i>Total</i>	All plots	37	812	63 %	12.2 m	21.7	6.0	116.2

165

166 In terms of individual trees' size and shape, the variation between the three strata was substantial. Stratum  
 167 1 (Figure 3) consisted largely of open or semi-open grown trees, whereas most trees in strata 2 and 3 (Figures  
 168 4 and 5) were growing under significant light competition. Although  $H_{\text{mean}}$  in stratum 1 was the lowest within  
 169 the study, it included some of the thickest trees.



170

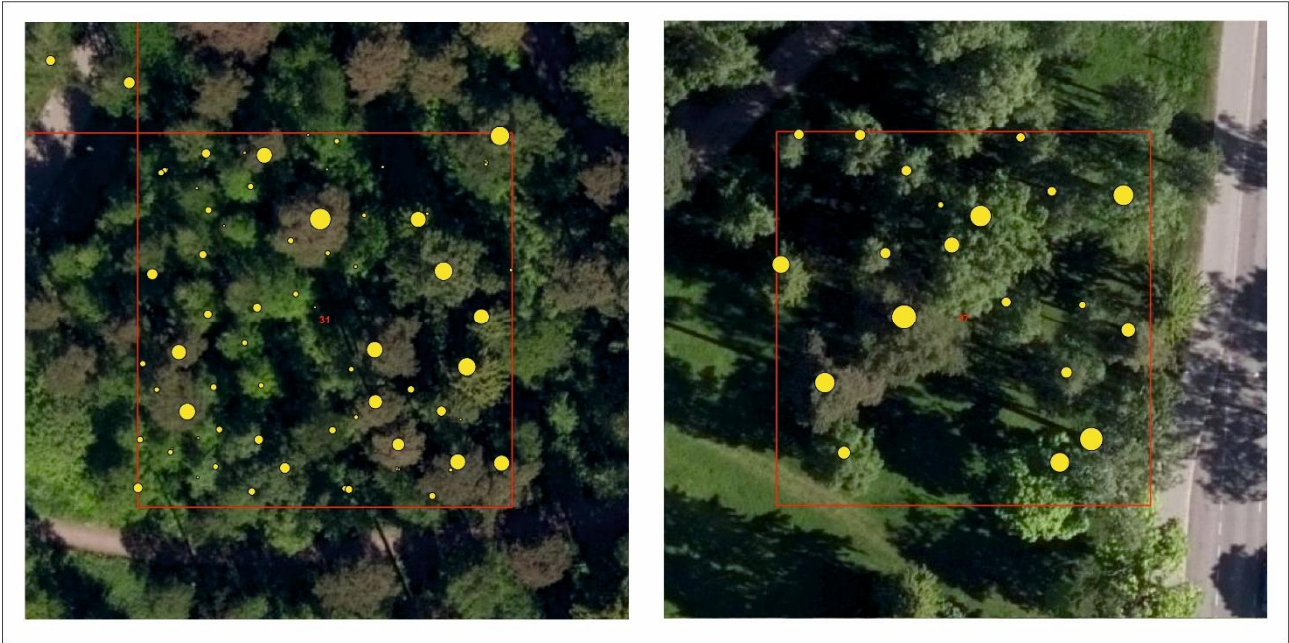
171 Figure 3: Two 32 m x 32 m sample plots describing the canopy conditions in stratum 1. Yellow circles represent the  
 172 position and relative size (DBH) of the field reference.  
 173



174

175 Figure 4: Two 32 m x 32 m sample plots describing the canopy conditions in stratum 2. Yellow circles represent the  
 176 position and relative size (DBH) of the field reference.  
 177





178

179 Figure 5: Two 32 m x 32 m sample plots describing the canopy conditions in stratum 3. Yellow circles represent the  
 180 position and relative size (DBH) of the field reference.  
 181

181

182 *Field measurements*

183 In total, 37 grid cells were field measured as sample plots. All plots were scanned in June-July 2016 using  
 184 Leica HDS 6100 (Leica Geosystems AG, Heerbrugg, Switzerland) terrestrial laser scanner (TLS). The field crew  
 185 scanned each plot from five locations and co-registered the scans using spherical reference targets. The  
 186 resulting point clouds were georeferenced by positioning the scan locations with Trimble R8s GNSS receiver  
 187 (Trimble Navigation Ltd., Sunnyvale, CA, USA). Following the work of Saarinen et al. (2014), the georeferenced  
 188 point clouds were used for creating tree maps for each field plot. The initial tree maps were first generated  
 189 from TLS data by excerpting a 20 cm slice from the point clouds at the height of 1.3 m. In case of multi-stem  
 190 trees, all stems that had branched below 1.3 m (i.e., DBH could be measured) were considered individual  
 191 trees. The TLS-derived tree maps were field-checked for error of omission and commission in April 2017. In  
 192 total, there were 812 trees on the field plots. Diameter at breast height (DBH) was determined for each tree  
 193 by measuring them from two directions with a caliper. All trees with DBH of 6 cm or above were recorded.  
 194 Based on total basal area, the most common tree species in stratum 1 were Salix sp. group (28%), Acer  
 195 platanoides (25%), and Betula pendula (17%). The most common tree species for stratum 2 were Betula  
 196 pendula (21%), Salix sp. group (19%), and Ulmus glabra (12%) and for stratum 3 Betula pendula (25%),  
 197 Quercus robur (23%), and Acer platanoides (11%).

198

199 *Tree delineation methods*

200 Altogether, nine different crown delineation methods were tested for detecting the trees from ALS data  
 201 (Table 2). All the applied methods were based on WS of smoothed CHMs.

202 Table 2: Description of the tree delineation methods tested in the study. WS stands for watershed segmentation.

Method	Filter description	Relative weight of center cell	Kernel size: pixels (effective)	Delineation method
G <sub>0.2</sub>	Gaussian, $\sigma = 0.2$	84.4%	3 x 3 (1.5 m x 1.5 m)	simple WS
G <sub>0.4</sub>	Gaussian, $\sigma = 0.4$	24.9%	5 x 5 (2.5 m x 2.5 m)	simple WS

$G_{0.5}$	Gaussian, $\sigma = 0.5$	15.9%	7 x 7 (3.5 m x 3.5 m)	simple WS
$G_{0.7}$	Gaussian, $\sigma = 0.7$	8.1%	9 x 9 (4.5 m x 4.5 m)	simple WS
$G_{0.9}$	Gaussian, $\sigma = 0.9$	4.9%	11 x 11 (5.5 m x 5.5 m)	simple WS
$G_{1.0}$	Gaussian, $\sigma = 1.0$	4.0%	13 x 13 (6.5 m x 6.5 m)	simple WS
$F_1$	Fixed kernel	25.0%	3 x 3 (1.5 m x 1.5 m)	simple WS
$F_2$	Fixed kernel	42.9%	3 x 3 (1.5 m x 1.5 m)	simple WS
$G_{adapt}$	Gaussian, $\sigma = 0.2 - 2.0$	4.0% - 96%	3 x 3 - 13 x 13 (1.5 m x 1.5 m - 6.5 m x 6.5 m)	marker controlled WS

203

204 The first eight of the methods presented in Table 2 (i.e.,  $G_{0.2} - F_2$ ) utilized WS without any pre-assumptions  
 205 on crown size. Here, the segmentation process was changed only by using different filter when smoothing  
 206 the CHM prior to applying WS. A Gaussian filter was tested using six different values of parameter  $\sigma$ ,  
 207 determining the size and weights of the filter ( $G_{0.2} - G_{1.0}$ ). The size of Gaussian filters ranged from 1.5 to 6.5  
 208 meters and the relative weight of the center cell varied from 84.5% to 4%, respectively. Methods  $F_1$  and  $F_2$   
 209 were adapted from earlier studies (Hyypä et al. 2001 and Kaartinen et al. 2012, respectively). They can be  
 210 considered as modifications of  $G_{0.2}$  filter giving the center cell a lower relative weight with the same spatial  
 211 extent.

212  $G_{adapt}$  was the most sophisticated method tested in the study. It used an automatic algorithm of local maxima  
 213 finding with a moving window of varying size, followed by crown delineation. The details of the algorithm  
 214 are: firstly, the CHM was smoothed with a Gaussian filter with varying window size and value of  $\sigma$  was  
 215 adjusted according to the CHM values, so that the higher the CHM value was the higher the value of  $\sigma$ . All in  
 216 all, the widow size varied between 1.5 m x 1.5 m and 6.5 m x 6.5 m; Secondly, local maxima were searched  
 217 for and considered as tree tops. Relying on the presumed relationship between crown width and tree height,  
 218 the search radius was determined with Equation 1, which is adapted from an earlier study (Pitkänen et al,  
 219 2004)

$$220 \quad C_w = 1.2 + 0.16h, \quad (1)$$

221 where  $C_w$  is crown width and  $h$  tree height in meters. Finally, single tree crowns were delineated using marker  
 222 controlled WS with the tree tops found in the previous step as control markers. Afterwards, tree locations  
 223 were obtained from the CHM pixels having the highest value within each tree segment.

## 224 *Modelling of DBH*

225 DBH was modeled using non-parametric Random Forest (RF) methodology (see, e.g., Yu et al., 2011). The RF  
 226 method classifies target observations by number of randomly selected predictors. Best predictors are chosen  
 227 by repeating the classification procedure several hundred times. In this study, sample trees for the training  
 228 data were selected among the field measured trees. In this step most of the small trees (DBH < 20 cm)  
 229 standing next to larger ones were discarded because their crowns were not represented properly by the  
 230 crown segments. Depending on the tested ITD method, the number of trees in the training set varied  
 231 between 356 and 442. The variation between methods was caused from matching the predicted tree crowns  
 232 with field measurements. Prediction models for DBH were created using 26 tree-level ALS-derived height and  
 233 density metrics as predictors. The DBH estimates were obtained by repeating the classification procedure  
 234 300 times.

## 235 *Accuracy assessment*

236 Stem count, minimum ( $DBH_{min}$ ), maximum ( $DBH_{max}$ ), and mean DBH ( $DBH_{mean}$ ) were used as criteria in  
 237 defining the plot-level accuracy for each method. For DBH, minimum and maximum were included to show



238 whether there are differences in the detection of small and large trees, whereas mean characterized whole  
 239 plots. For each criterion, best-case accuracies were determined for each stratum. All assessments were made  
 240 first at plot-level and combined into stratum-level results. Estimates were compared using root mean square  
 241 error (RMSE) (Equation 2), bias (Equation 3) and their relative values (Equations 4 and 5).

$$242 \quad RMSE = \sqrt{\frac{\sum_{i=1}^n (y_i - \hat{y}_i)^2}{n}}, \quad (2)$$

$$243 \quad bias = \frac{\sum_{i=1}^n (y_i - \hat{y}_i)}{n}, \quad (3)$$

$$244 \quad RMSE_{rel} = \frac{RMSE}{\bar{y}}, \quad (4)$$

$$245 \quad bias_{rel} = \frac{bias}{\bar{y}}, \quad (5)$$

246 where  $n$  is the number of trees,  $y$  the observed value,  $\hat{y}$  the modeled value, and  $\bar{y}$  the mean of observations.

247 In addition, DBH estimates were transformed into DBH distributions with 3 cm bin size. DBH distributions  
 248 were compared to the field-measured distributions using Reynold's Error index (Reynolds et al., 1988)  
 249 (Equation 6) and relative error index (Packalén and Maltamo, 2008) (Equation 7). As both the DBH and the  
 250 tree count are present in the approach, the distribution-based methods illustrate the overall correctness of  
 251 plot-level observations.

$$252 \quad EI = \sum_{i=1}^k w_i |f_i - \hat{f}_i|, \quad (6)$$

$$253 \quad EI_{rel} = \sum_{i=1}^k 0.5 \left| \frac{f_i}{N} - \frac{\hat{f}_i}{\hat{N}} \right|, \quad (7)$$

254 where  $k$  is the number of DBH classes,  $w_i$  is the weight of class  $i$ ,  $f_i$  is the observed tree count in class  $i$ ,  $\hat{f}_i$  is  
 255 the predicted tree count in class  $i$ ,  $N$  is the observed, and  $\hat{N}$  the predicted tree count in the plot. Here, weight  
 256 parameter  $w_i$  was set to 1, giving equal weight for all DBH classes.  $EI$  shows the absolute difference between  
 257 the observed and predicted distribution, whereas  $EI_{rel}$  is its modification and can be interpreted as a  
 258 proportion of incorrect observations in terms of the shape of the reference distribution. In this study,  $EI$  was  
 259 used in comparing the within stratum results, whereas  $EI_{rel}$  was utilized when comparing the performance of  
 260 each method in different strata.

## 261 Results

### 262 Tree count

263 The performance of all ITD methods showed considerable variation between the three strata. Depending on  
 264 the stratum and the ITD method used, the  $RMSE_{rel}$  of tree count varied between 0.36 – 1.79 and  $bias_{rel}$   
 265 between -1.23 – 0.63 (Table 3).

266 In terms of  $RMSE_{rel}$  stratum 2 had the highest accuracies. For  $bias_{rel}$ , stratum 3 showed the best results.  
 267 Methods  $G_{0.5}$  and  $F_1$  resulted in the best accuracies in all three strata in nearly all criteria. An exception to  
 268 this were  $RMSE$  and  $RMSE_{rel}$  in stratum 2, were  $G_{adapt}$  resulted in the best results.

269 Table 3. RMSE and bias for the number of detected trees in all three strata. Column *Total* shows the overall accuracy  
 270 without stratification. The accuracies of the best methods within each stratum are highlighted.

271

$\Sigma \phi$	Stratum 1	Stratum 2	Stratum 3	Total
---------------	-----------	-----------	-----------	-------

	RMSE, trees/plot	RMSE <sub>rel</sub>	bias, trees/plot	bias <sub>rel</sub>	RMSE, trees/plot	RMSE <sub>rel</sub>	bias, trees/plot	bias <sub>rel</sub>	RMSE, trees/plot	RMSE <sub>rel</sub>	bias, trees/plot	bias <sub>rel</sub>	RMSE, trees/plot	RMSE <sub>rel</sub>	bias, trees/plot	bias <sub>rel</sub>
G <sub>0.2</sub>	12.3	1.79	-8.4	-1.23	20.9	0.69	-14.1	-0.47	24.1	0.78	-15.3	-0.50	19.3	0.81	-11.9	-0.50
G <sub>0.4</sub>	5.3	0.78	-3.2	-0.46	12.9	0.43	7.1	0.23	19.3	0.63	7.4	0.24	14.8	0.62	4.8	0.20
G <sub>0.5</sub>	3.7	0.54	-1.2	-0.17	15.2	0.50	10.9	0.36	21.0	0.68	11.8	0.38	16.7	0.70	8.3	0.35
G <sub>0.7</sub>	4.7	0.69	1.6	0.23	18.5	0.61	15.3	0.51	23.8	0.77	15.7	0.51	19.1	0.80	12.0	0.50
G <sub>0.9</sub>	5.3	0.77	1.7	0.24	19.2	0.64	15.7	0.52	24.0	0.78	16.2	0.53	19.8	0.83	12.4	0.52
G <sub>1.0</sub>	6.2	0.90	3.0	0.44	22.9	0.76	18.9	0.63	27.1	0.88	19.4	0.63	22.2	0.93	14.9	0.63
F <sub>1</sub>	6.3	0.92	-3.9	-0.57	11.1	0.37	1.9	0.06	18.6	0.60	1.0	0.03	13.7	0.57	0.8	0.03
F <sub>2</sub>	5.8	0.84	-3.7	-0.54	13.0	0.43	5.3	0.18	19.7	0.64	4.9	0.16	14.8	0.62	3.3	0.14
G <sub>adapt</sub>	4.7	0.69	-2.4	-0.35	11.0	0.36	3.2	0.11	21.2	0.69	12.3	0.40	15.0	0.63	4.8	0.20

272

273 When comparing the accuracy of different methods, method F<sub>1</sub> resulted in the best results overall. F<sub>1</sub> had  
274 the lowest RMSE and bias within stratum 3 (18.6 and 1.0, respectively) and within the unstratified data  
275 (column *Total*, 13.7 and 0.8, respectively). Overall, the methods performing the best resulted in low bias in  
276 all strata. However, the RMSE remained rather high for all methods and strata.

#### 277 *DBH estimates*

278 As the tree candidates were not matched to the field reference, the accuracy of DBH estimates was  
279 investigated at plot-level. The accuracies of estimated minimum, maximum, and mean DBHs were  
280 investigated and the plot-level results are presented in Table 4. The accuracy of minimum and maximum DBH  
281 illustrates the methods' ability to detect the smallest and the largest trees, whereas mean DBH demonstrates  
282 the methods' ability to capture the overall structure of the plot.

283 Table 4. Accuracy of plot-level estimates of minimum, maximum, and mean DBH. Column Total shows the  
 284 accuracy for the entire data set. The bolded figures highlight the most accurate methods within each  
 285 stratum for each DBH measure (min, max, and mean).

286

Method	DBH measure	Stratum 1				Stratum 2				Stratum 3				Total			
		RMSE <sub>L</sub> , cm	RMSE <sub>rel</sub>	bias, cm	bias <sub>rel</sub>	RMSE <sub>L</sub> , cm	RMSE <sub>rel</sub>	bias, cm	bias <sub>rel</sub>	RMSE <sub>L</sub> , cm	RMSE <sub>rel</sub>	bias	bias <sub>rel</sub>	RMSE <sub>L</sub> , cm	RMSE <sub>rel</sub>	bias, cm	bias <sub>rel</sub>
G <sub>0.2</sub>	min	29.3	4.88	11.9	1.98	2.2	<b>0.37</b>	-1.5	-0.25	6.3	1.01	-2.6	-0.41	17.0	2.84	2.5	0.42
	max	16.8	0.16	1.2	0.01	19.8	0.30	-8.9	-0.14	16.9	0.15	<b>0.3</b>	<b>0.00</b>	17.9	0.15	-3.2	-0.03
	mean	23.3	1.13	8.7	0.42	3.9	0.20	-2.8	-0.14	11.0	0.43	-8.4	-0.33	14.8	0.70	-0.8	-0.04
G <sub>0.4</sub>	min	27.2	4.54	10.4	1.73	4.2	0.71	-3.6	-0.59	8.0	1.29	-5.5	-0.89	16.3	2.72	0.3	0.05
	max	17.0	0.16	2.7	0.02	6.6	0.10	<b>1.5</b>	<b>0.02</b>	21.3	0.18	3.8	0.03	15.0	0.13	<b>2.1</b>	<b>0.02</b>
	mean	22.4	1.09	8.2	0.40	4.4	0.23	-3.6	-0.18	11.9	0.47	-9.8	-0.38	14.7	0.69	-1.7	-0.08
G <sub>0.5</sub>	min	27.6	4.61	9.7	1.61	6.4	1.07	-5.1	-0.85	9.3	1.50	-5.3	-0.86	17.0	2.83	-0.5	-0.09
	max	17.7	0.16	1.2	0.01	5.6	0.09	3.2	0.05	17.8	0.15	8.2	0.07	13.9	0.12	3.8	0.03
	mean	23.4	1.13	7.5	0.37	5.5	0.28	-4.4	-0.22	10.0	0.39	-7.5	-0.29	14.9	0.70	-1.6	-0.08
G <sub>0.7</sub>	min	27.4	4.56	8.7	1.44	7.9	1.32	-6.4	-1.06	9.9	1.60	-8.5	-1.37	17.2	2.87	-2.2	-0.36
	max	17.6	0.16	7.4	0.07	<b>4.3</b>	<b>0.07</b>	2.3	0.04	19.3	0.17	12.9	0.11	14.1	0.12	6.6	0.06
	mean	23.8	1.15	8.6	0.42	6.9	0.35	-6.0	-0.30	12.4	0.48	-9.3	-0.36	15.8	0.74	-2.4	-0.11
G <sub>0.9</sub>	min	26.8	4.47	4.9	0.82	11.5	1.91	-10.7	-1.78	20.8	3.36	-17.2	-2.78	19.9	3.32	-7.2	-1.21
	max	10.2	<b>0.09</b>	0.0	<b>0.00</b>	8.2	0.13	2.9	0.05	17.6	0.15	6.4	0.06	<b>11.7</b>	<b>0.10</b>	2.8	0.02
	mean	18.1	<b>0.88</b>	3.6	0.18	9.3	0.48	-8.6	-0.44	17.8	0.70	-14.5	-0.57	15.2	0.72	-6.5	-0.30
G <sub>1.0</sub>	min	<b>22.5</b>	<b>3.75</b>	4.5	<b>0.75</b>	11.7	1.96	-10.6	-1.77	12.3	1.99	-10.1	-1.64	16.2	2.69	-5.7	-0.94
	max	19.0	0.18	6.9	0.06	9.5	0.15	4.9	0.08	16.4	0.14	10.8	0.09	14.8	0.13	6.9	0.06
	mean	20.2	0.98	6.1	0.30	8.3	0.42	-7.4	-0.38	15.2	0.59	-11.4	-0.45	15.1	0.71	-4.4	-0.21
F <sub>1</sub>	min	28.7	4.78	10.4	1.73	3.0	0.51	-2.7	-0.46	8.2	1.32	-5.4	-0.87	17.0	2.84	0.7	0.11
	max	14.0	0.13	3.7	0.03	7.4	0.11	5.1	0.08	17.0	0.15	8.5	0.07	12.5	0.11	5.4	0.05
	mean	23.8	1.15	8.9	0.43	3.3	0.17	-2.6	-0.13	8.5	0.33	-5.9	-0.23	14.6	0.69	-0.1	<b>0.00</b>
F <sub>2</sub>	min	27.6	4.59	10.4	1.73	3.8	0.63	-3.3	-0.55	7.8	1.26	-4.2	-0.68	16.4	2.73	0.8	0.14
	max	16.9	0.16	0.4	0.00	7.9	0.12	3.3	0.05	<b>15.9</b>	<b>0.14</b>	6.6	0.06	13.5	0.12	2.8	0.02
	mean	23.7	1.15	8.5	0.41	3.8	0.19	-3.0	-0.15	11.2	0.44	-8.4	-0.33	15.2	0.71	-1.0	-0.05
G <sub>adapt</sub>	min	26.8	4.47	8.9	1.48	4.3	0.72	-4.0	-0.67	7.9	1.28	-4.5	-0.73	<b>16.0</b>	<b>2.67</b>	0.1	<b>0.02</b>
	max	12.9	0.12	3.9	0.04	9.5	0.15	8.2	0.13	20.9	0.18	15.3	0.13	14.1	0.12	8.6	0.07
	mean	23.6	1.14	9.1	0.44	2.2	0.11	-0.7	-0.04	7.5	0.30	-3.0	-0.12	14.0	0.66	1.8	0.08

287

288 In stratum 1, all tested methods underestimated the minimum DBH and method G<sub>1.0</sub> resulted in the most  
 289 accurate results. The relative bias remained high for all methods, varying from 0.75 – 1.98. The relative RMSE  
 290 ranged between 3.75 and 4.88. For maximum and mean diameters, the results were more accurate, method  
 291 G<sub>0.9</sub> giving the most accurate results. The relative bias of maximum and mean diameters ranged from 0.00 –  
 292 0.07 and 0.18 – 0.44, respectively. The best results in stratum 1 were achieved by using rather heavy filtering  
 293 of CHM, i.e., using methods G<sub>0.9</sub> and G<sub>1.0</sub>.

294 As the minimum diameter was systematically underestimated in stratum 1, the situation was the exact  
 295 opposite in stratum 2. All methods overestimate the minimum DBH, method G<sub>0.2</sub> giving the most accurate  
 296 results. Overall, the relative bias of minimum DBH in stratum 2 ranged from -0.25 to -1.78. Relative RMSE  
 297 ranged between 0.37 and 1.96. Except for G<sub>0.2</sub>, maximum DBH was underestimated by all methods. Method  
 298 G<sub>0.4</sub> delivered the lowest bias<sub>rel</sub> (0.02) but method G<sub>0.7</sub> resulted in the smallest RMSE<sub>rel</sub> (0.07). Relative bias



299 and RMSE for maximum DBH ranged from -0.14 to 0.13 and from 0.07 to 0.30, respectively. For mean DBH,  
300 method  $G_{\text{adapt}}$  resulted in the most accurate results ( $\text{bias}_{\text{rel}} = -0.04$  and  $\text{RMSE}_{\text{rel}} = 0.11$ ). Overall, all methods  
301 overestimated the mean DBH. The relative bias and RMSE for mean DBH ranged from -0.44 to -0.04 and from  
302 0.11 to 0.48, respectively.

303 Also, in stratum 3, all methods overestimated the minimum DBH.  $G_{0.2}$  was the most accurate method for  
304 minimum DBH. The relative bias ranged from -2.78 to -0.41 and relative RMSE from 1.01 to 3.36. For  
305 maximum DBH method  $G_{0.2}$  resulted in the lowest bias ( $\text{bias}_{\text{rel}} = 0.00$ ) and  $F_2$  the lowest RMSE ( $\text{RMSE}_{\text{rel}} = 0.14$ ).  
306 Overall, all methods underestimated the maximum DBH also in stratum 3. Relative bias and RMSE ranged  
307 from 0.00 to 0.13 and 0.14 to 0.18, respectively. The mean DBH was best captured using method  $G_{\text{adapt}}$  ( $\text{bias}_{\text{rel}}$   
308  $= -0.12$  and  $\text{RMSE}_{\text{rel}} = 0.30$ ). All methods overestimated the mean DBH. The relative bias and RMSE for mean  
309 DBH ranged from -0.57 to -0.12 and from 0.30 to 0.70, respectively.

310 When the study area was not stratified,  $G_{\text{adapt}}$  was the most accurate method for determining minimum DBH  
311 ( $\text{bias}_{\text{rel}} = 0.02$  and  $\text{RMSE}_{\text{rel}} = 2.67$ ). The relative bias and RMSE ranged from -1.21 to 0.42 and from 2.67 to  
312 3.32, respectively. For maximum DBH, all methods except  $G_{0.2}$  resulted in slight overestimates.  $G_{0.4}$  ( $\text{bias}_{\text{rel}} =$   
313  $0.02$ ) and  $G_{0.9}$  ( $\text{RMSE}_{\text{rel}} = 0.10$ ) resulted in the most accurate estimates, whereas the relative bias and RMSE  
314 overall ranged from -0.03 to 0.07 and from 0.10 to 0.15, respectively. The most accurate methods for  
315 estimating mean DBH were  $F_1$  ( $\text{bias}_{\text{rel}} = 0.00$ ) and  $G_{\text{adapt}}$  ( $\text{RMSE}_{\text{rel}} = 0.66$ ). The relative bias and RMSE of mean  
316 DBH estimates ranged from -0.30 to 0.08 and from 0.66 to 0.74, respectively.

317 When comparing the accuracy of the DBH estimates from the three strata, stratum 1 clearly stands out. With  
318 few exceptions, in stratum 1 the  $\text{RMSE}_{\text{rel}}$  is at its highest for all methods. Also, it is only in stratum 1 that the  
319 methods utilizing heavy filtering (i.e.,  $G_{0.9}$  and  $G_{1.0}$ ) outperform the others. In strata 2 and 3, heavy filtering  
320 typically resulted in lowest accuracies in all three DBH estimates. For these two strata,  $G_{0.2}$  was found the  
321 most accurate for minimum DBH, which reflects the methods sensitivity to small scale variation in CHM. For  
322 the same two strata,  $G_{\text{adapt}}$  was found the best method for acquiring the mean DBH. For maximum DBH, the  
323 differences between methods were relatively small and none of the methods clearly stood out. Except for  
324  $G_{0.2}$ , all methods underestimated the maximum DBH in all three strata.

### 325 *DBH distributions*

326 Error indices describing the correctness of the ITD-derived DBH distributions are presented in Table 5. As the  
327 tree count varies between the strata,  $EI$  is used to compare the performance of the methods within each  
328 stratum, whereas  $EI_{\text{rel}}$  is used for assessing the performance of each method over all strata.

329

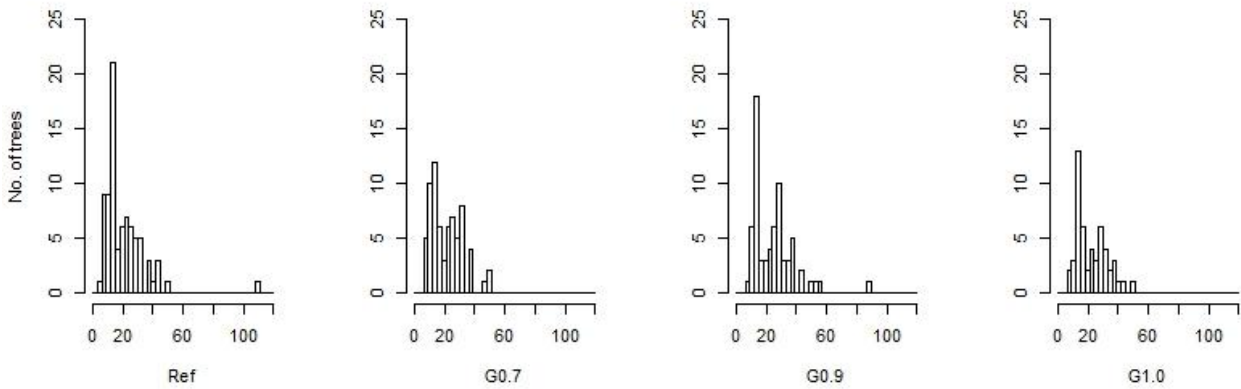
330 Table 5. Error index ( $EI$ ) and Relative error index ( $EI_{rel}$ ) of the ITD-derived DBH distributions in all three strata. Column  
 331 *Total* shows the overall accuracy without stratification. Best  $EI$  within each stratum are bolded.

Method		Stratum			<i>Total</i>
		1	2	3	
$G_{0.2}$	$EI$	160	506	428	1094
	$EI_{rel}$	0.590	0.386	0.573	0.501
$G_{0.4}$	$EI$	116	339	302	757
	$EI_{rel}$	0.594	0.409	0.616	0.522
$G_{0.5}$	$EI$	98	350	276	724
	$EI_{rel}$	0.607	0.499	0.602	0.561
$G_{0.7}$	$EI$	<b>75</b>	343	274	<b>692</b>
	$EI_{rel}$	0.644	0.502	0.655	0.588
$G_{0.9}$	$EI$	82	356	265	703
	$EI_{rel}$	0.715	0.543	0.673	0.633
$G_{1.0}$	$EI$	82	373	265	720
	$EI_{rel}$	0.725	0.611	0.681	0.666
$F_1$	$EI$	115	378	343	836
	$EI_{rel}$	0.621	0.440	0.613	0.543
$F_2$	$EI$	116	<b>338</b>	328	782
	$EI_{rel}$	0.563	0.402	0.618	0.509
$G_{adapt}$	$EI$	89	374	<b>257</b>	720
	$EI_{rel}$	0.475	0.439	0.548	0.478

332  
 333 When examining  $EI$  within stratum 1, method  $G_{0.7}$  results in the lowest number of errors in classification (75  
 334 misclassifications) followed by  $G_{0.9}$  and  $G_{1.0}$  (82 misclassifications). Large filter size seemed to increase the  
 335 accuracy in terms of correctness of DBH distributions. In stratum 2, method  $F_2$  had the lowest number of  
 336 classification errors. However, the differences between  $G_{0.4}$ ,  $G_{0.7}$ , and  $F_2$  are small (339, 343, and 338,  
 337 respectively). Unlike in stratum 1, there is no clear pattern between filter size and classification accuracy. In  
 338 stratum 3, method  $G_{adapt}$  results in the best classification accuracy (257). Still, methods'  $G_{0.9}$  and  $G_{1.0}$   
 339 classification accuracies (265 for both) are close to that of  $G_{adapt}$ . Hence, it seems that, also in stratum 3,  
 340 methods with larger filter size result in more accurate estimates for plot-level DBH distributions. Looking at  
 341 column *Total*, the effect of different methods is clear. The intensity of CHM filtering increases gradually for  
 342 methods  $G_{0.2}$ - $G_{1.0}$ , which decreases the number of detected tree candidates. Method  $G_{0.2}$  strongly  
 343 overestimates, whereas  $G_{0.4}$ - $G_{1.0}$  underestimate the tree count. The best results were achieved with method  
 344  $G_{0.7}$ .

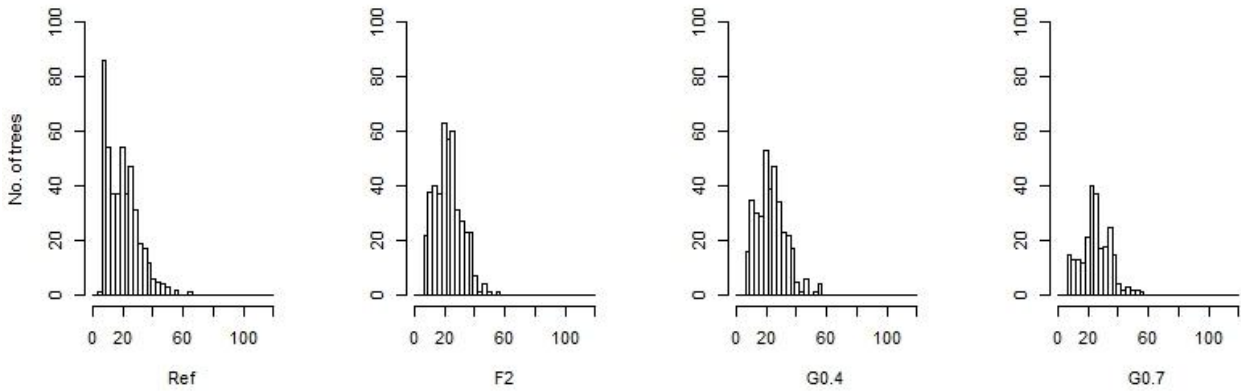
345 Differences between the three best methods within each stratum were visualized through DBH distributions.  
 346 In stratum 1 (Figure 6), the methods with the best  $EI$  values resulted in two-peaked DBH distributions, which  
 347 resembled that of the field reference. However, none of the three methods were able to detect the highest  
 348 DBH values. Also,  $G_{adapt}$  overestimated the tree count both in the smallest and the largest DBH classes. In  
 349 stratum 2 (Figure 7), all three methods underestimated the number of trees with DBH <10 cm, whereas  
 350 method  $F_2$  overestimated the number of trees with DBH between 20-30 cm. The results are similar in stratum  
 351 3 (Figure 8), where the number of trees with DBH <20 cm is underestimated. Considering all three strata, the  
 352 amount of detected small trees decreased as canopy cover and plot mean height increased.

353



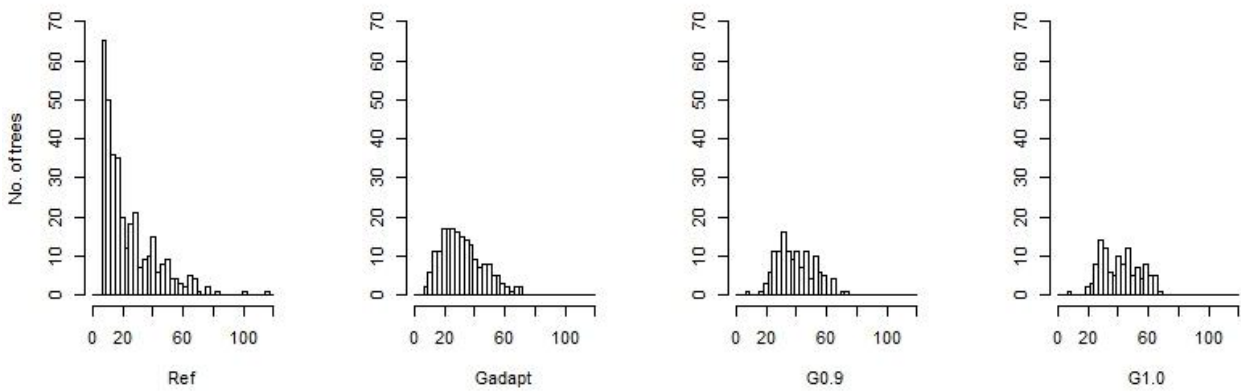
354

355 Figure 6. DBH distributions of the field reference and three best performing methods ( $G_{0.7}$ ,  $G_{0.9}$ , and  $G_{adapt}$ ) in stratum 1.



356

357 Figure 7. DBH distributions of the field reference and three best performing methods ( $F_2$ ,  $G_{0.4}$ , and  $G_{0.5}$ ) in stratum 2.



358

359 Figure 8. DBH distributions of the field reference and three best performing methods ( $G_{adapt}$ ,  $G_{0.9}$ , and  $G_{1.0}$ ) in stratum 3.

360

361 *Overall performance of the best methods*



362 The best performing methods were selected on grounds of EI as it summarized the predicted information of  
 363 number of trees and DBHs. Table 6 summarizes the accuracy for the best methods in terms of EI, DBH, and  
 364 number of trees.

365 Table 6. A summary of the best-performing ITD methods' accuracies for each stratum.  
 366

Stratum	Best method (in terms of EI)	Number of trees		DBH <sub>min</sub>		DBH <sub>max</sub>		DBH <sub>mean</sub>		EI	EI <sub>rel</sub>
		RMSE <sub>rel</sub>	bias <sub>rel</sub>	RMSE <sub>rel</sub>	bias <sub>rel</sub>	RMSE <sub>rel</sub>	bias <sub>rel</sub>	RMSE <sub>rel</sub>	bias <sub>rel</sub>		
1	G <sub>07</sub>	0.69	0.23	4.56	1.44	0.16	0.07	1.15	0.42	75	0.644
2	F <sub>2</sub>	0.43	0.18	0.63	-0.55	0.12	0.05	0.19	-0.15	338	0.402
3	G <sub>adapt</sub>	0.69	0.40	1.28	-0.73	0.18	0.13	0.30	-0.12	257	0.548
all	G <sub>07</sub>	0.80	0.50	2.87	-0.36	0.12	0.06	0.74	-0.11	692	0.588

367

### 368 Discussion

369 In this study, we investigated the effect of heterogeneous canopy conditions on the performance of nine  
 370 CHM-based ITD methods in urban parks. To do so, the park areas were stratified into three groups according  
 371 to ALS-derived canopy mean height and canopy cover. In addition, the effect of large trees on plot-level  
 372 accuracy of ITD methods was explored. Accuracy was evaluated through the correctness of plot-level stem  
 373 count and three DBH-related criteria (DBH<sub>min</sub>, DBH<sub>max</sub>, and DBH<sub>mean</sub>). All criteria were examined separately to  
 374 define the best-case methods for each stratum and criterion. Finally, special requirements of urban ITD  
 375 methodology were brought up by defining the best ITD methods for each stratum, by using Reynold's Error  
 376 index. As expected, the canopy conditions had a considerable effect on the performance of ITD methods.  
 377 Contrary to our expectations, areas with lowest canopy cover turned out as the most demanding for ITD  
 378 approach. This was largely due to heterogeneity of the stand structure, but also large open-grown trees.  
 379 Finally, we demonstrated that selecting stratum-specific ITD methods improves the overall results of urban  
 380 ITD.

381 Although the estimation of DBH was not in the core of this study, the issue is addressed here because DBH  
 382 was a central parameter when assessing the accuracy of ITD methods. The light conditions varied from nearly  
 383 open areas to closed canopy, which is known to affect the growth pattern of trees (e.g., Mäkinen, 2003;  
 384 Niemistö, 1995). Variation in growth pattern weakens the relation between DBH and height and thus lowers  
 385 the accuracy of DBH estimates (Tanhuanpää et al., 2017). Also, as the species of the detected trees were not  
 386 identified, the species information could not be used in DBH modelling. As the growth patterns differ  
 387 between different tree species, including species information to the process would likely enhance the overall  
 388 accuracies. Still, as the DBH modelling procedure was the same for all methods, we presume that the method-  
 389 related variation of DBH estimates remained fairly constant in all ITD methods tested, and that the main  
 390 differences between methods resulted from differences in crown delineation. Considering the estimates for  
 391 DBH<sub>min</sub> and DBH<sub>max</sub>, the number of trees is rather small, which decreases the robustness of these two figures.  
 392 Since only one tree per plot was selected as the smallest or the biggest, the figures rely on 10-15 trees (i.e.,  
 393 one tree per plot) in stratum-level and 37 trees when no stratification was done. As all trees in each plot are  
 394 used to achieve DBH<sub>mean</sub>, this figure is more robust for describing the methods' performance.

395 As expected, the performance of ITD methods varied substantially between the three park strata. Table 6  
 396 shows how the relative RMSE of stem count and DBH-related accuracy indicators increase considerably while  
 397 shifting from stratum 2 to stratum 3, towards denser and higher canopy. As similar results have been  
 398 reported before (e.g., Koch et al., 2006), similar results were also expected to be reported in urban

399 surroundings. What was surprising in the results was that the poorest overall performance of ITD methods  
400 was found in stratum 1, representing the areas with sparse canopy. Table 6 shows how the performance of  
401 the best ITD methods is at the lowest in stratum 1 considering most of the studied criteria. Similarly, looking  
402 at the best-case results in stem count and DBH criteria (Tables 3 and 4), the results in sparse canopy areas  
403 are mainly weaker than in the other two strata. In the open park areas of stratum 1, variation in canopy  
404 structure is wide. The stratum contains the smallest and some of the largest trees in the study area. Also, the  
405 spatial distribution of the trees varied substantially within the stratum as clustering of the trees was not taken  
406 into consideration. Hence, similar composition in terms of canopy cover and mean height may result from  
407 highly different spatial patterns. Evenly distributed crowns are an ideal subject for delineation of individual  
408 trees as every tree stands out from the CHM. Then again, clustered crowns result in challenging crown objects  
409 for crown delineation, because tree crowns are interlocking, and tall trees cover the smaller ones.

410 As shown in earlier studies in managed rural forests, deciduous trees differ from conifers in crown shape  
411 (e.g., Kwack et al., 2007; Koch et al. 2006) and thus pose special requirements for ITD methods. The same  
412 also holds for urban forests where manmade growing conditions result in altered growth patterns  
413 (Tanhuanpää et al., 2014, 2017). As expected, similar findings were also made in this study. The observed  
414 effect of large deciduous trees was twofold. Because of their wide and fragmented crowns, old deciduous  
415 trees tended to be oversegmented with all methods. In addition to visual observations during the process,  
416 the phenomenon was also supported indirectly by the results. Stratum 2, which had the most accurate ITD  
417 results with respect to all criteria (Tables 3-5), had lower  $DBH_{mean}$  and  $DBH_{max}$  than strata 1 and 3 (Table 1).  
418 This means that the relative number of large trees in stratum 2 was lower and that the biggest trees were  
419 notably smaller than in the other two strata. Still, as the crowns were not investigated in detail, only indirect  
420 interpretations can be made. In addition to commission errors, large deciduous crowns also resulted in  
421 omission errors. Especially smaller trees located under or even close to large crowns were easily left  
422 undetected. We found this was especially problematic in case of open-grown crowns. To be correctly  
423 delineated, the wide multi-top crowns required either heavy filtering of CHM or, depending on the method  
424 used, large threshold value for seed point selection. In both cases small tree tops were ignored. Brandtberg  
425 et al. (2001) reported similar findings on delineating unevenly sized trees simultaneously.

426 The effect of large crowns can be seen in the distributions describing the trees in strata 2 and 3 (Figures 4  
427 and 5). Looking at the reference distributions, the largest DBH classes in stratum 2 (Figure 7) consisted of  
428 trees with  $DBH < 10$  cm. However, with all the ITD methods tested, the trees in the largest classes had DBH  
429 of 20 cm and above. We presume that this was caused mainly by two factors. Firstly, small trees remained  
430 undetected under the taller ones which lead to omission errors in small diameter classes. On the other hand,  
431 fragmentation of large tree crowns caused detection of non-existing trees, which lead to commission errors  
432 in larger diameter classes. However, as dense dominant canopy layer is known to suppress shorter trees in  
433 CHM-based ITD regardless of whether there are single large trees or not (Wang et al. 2016), the finding can  
434 be only partly supported by the results. The effects of high canopy cover and large tree crowns are very  
435 similar and thus challenging to separate. In both cases, smaller DBH classes suffer from omission errors. The  
436 results lead to a conclusion that many of the large open-grown trees are likely misdelineated, resulting in  
437 several smaller crown segments, but the effect of higher canopy cover cannot be fully excluded from the  
438 analysis.

439 The effect of large trees was most evident in stratum 1, which turned up to be the most challenging  
440 environment for ITD (see, Tables 3-5). Then again, stratum 2 that consisted of areas with the lowest mean  
441 DBH and the canopy was formed mainly from young trees with rather consistent crowns, seems to represent  
442 the most favorable conditions for ITD in the study area.

443 Comparing the results from stratified and non-stratified data shows that there are no methods that would  
444 be the best in all strata. Despite the differences in smoothing of CHM and applying WS, simple WS and marker

445 controlled WS resulted in similar results. Applying the methods with the lowest EI (i.e.,  $G_{0.7}$ ,  $F_2$  and  $G_{adapt}$ )  
446 leads to total of 670 misclassifications, whereas the best performing method ( $G_{0.7}$ ) for the unstratified dataset  
447 leads to 692 misclassifications.

448 All the best three methods had similar trend of performance between the three strata (Table 5). In terms of  
449  $EI_{rel}$ , stratum 2 had the most accurate results. For strata 1 and 3, the  $EI_{rel}$  values were typically rather similar  
450 between the methods. For the best three, results for stratum 1 were slightly better than for stratum 3. This  
451 means that the shape of the DBH distributions were more accurate in stratum 1. Overall, all the best methods  
452 underestimated the plot-level tree count (Table 6). From the three strata, the biggest underestimations were  
453 found in stratum 3 where majority of the small undergrowth was missed due to dense cover of the  
454 suppressing canopy (Figure 8).

455 In general, ITD approach performed the best in stratum 2. When comparing the results from the best-  
456 performing methods, the accuracy in stratum 2 was the highest in nearly all the criteria presented in Table 6.  
457 The only exception to this was  $DBH_{mean}$ , which had slightly higher relative bias than in stratum 3. We suggest  
458 that the reason for having the best results in stratum 2 lies in the low number of old open-grown trees. This  
459 is supported by the mean characteristics of the strata. On average, the trees in stratum 2 are growing much  
460 denser than in the other two strata. Also, the mean diameter is the lowest and mean height close to that of  
461 stratum 1. Hence the canopy in stratum 2 consists of younger trees with narrow and hence distinct crowns.

462 Stratification of park areas served two purposes. Firstly, it could be means for reaching more accurate results,  
463 as sites can be mapped with methods adjusted for those specific conditions. Secondly, considering  
464 operational applications, it allowed assessing the accuracy of the methods tested separately for each  
465 stratum. Knowing that the accuracy of ITD varies over the mapped area, the stratum-level accuracy could be  
466 used as a quality indicator for the created tree maps. However, as only ALS-derived metrics ( $CC$  and  $H_{mean}$ )  
467 were used for the stratification, the actual structure of the resulting strata varied considerably. The within-  
468 strata variation seemed to affect the accuracy of ITD. In terms of tree size, the park areas in stratum 1 and 3  
469 are more diverse than in stratum 2. High variation in the size of trees lowered the accuracy of ITD regardless  
470 of the method used. In the light of this study, urban ITD methods would benefit from better and broader  
471 description of canopy. Further research effort should be put on meaningful stratification of park areas for  
472 enhancing the performance of ITD in park areas.

473 Overall, applying a single method for all three strata leads to lower overall accuracy with all three indicators  
474 tested (i.e., number of trees, DBH, and Error index). This underlines that applying ITD for heterogeneous  
475 urban forests requires specialized methods. Compared to the ITD methods developed for forest conditions,  
476 urban forests require algorithms that are more agile to cope with the high variation in tree size and shape.

477

## 478 Conclusions

479 The intense management of park areas creates growing conditions that are rather unnatural. This holds  
480 especially for light conditions. Altered growing conditions affect the formation of individual tree crowns and  
481 thus the characteristics of the whole canopy. In this study, we pursued the issue by stratifying the park areas  
482 by ALS-derived features describing the canopy structure. The results show that urban, highly heterogeneous,  
483 and deciduous-dominated forests are a challenging environment for CHM-based detection of individual  
484 trees. In the light of this study, the key elements limiting the accuracy of urban ITD are the presence of large  
485 open-grown deciduous trees and, on the other hand, the heterogeneity of the canopy. Large deciduous trees  
486 result in both omission and commission errors and are thus a single key component in forming a successful  
487 ITD application. In addition, it seems that canopy cover type-specific methods would increase the accuracy.  
488 However, the specialized methods need to tackle a large variation of crown characteristics even within small



489 park areas. To achieve the best possible results, stratifying the park areas into more homogeneous areas is a  
490 viable option. At least for the relatively open park areas, basing the stratification solely on canopy cover and  
491 mean height is not enough for defining the most suitable crown delineation method. As for adaptive  
492 delineation methods, this means that controlling the segmentation algorithm with solely CHM height does  
493 not provide enough information for achieving good ITD accuracy.

494

#### 495 Acknowledgement

496 The project was done in co-operation with the City of Helsinki and funded by the City of Helsinki Innovation  
497 fund and the Academy of Finland (The Centre of Excellence in Laser Scanning Research, project number  
498 2721955). Additional funding was also received from the Academy of Finland's Strategic Research Council  
499 (IBC-Carbon, project number 312559).

500

#### 501 References

502 Alonzo, M., Bookhagen, B., & Roberts, D. A. (2014). Urban tree species mapping using hyperspectral and lidar  
503 data fusion. *Remote Sensing of Environment*, 148, 70-83.

504 Axelsson, P. (2000). DEM generation from laser scanner data using adaptive TIN models. *International*  
505 *archives of photogrammetry and remote sensing*, 33(4), 110-117.

506 Brandtberg, T., Warner, T. A., Landenberger, R. E., & McGraw, J. B. (2003). Detection and analysis of individual  
507 leaf-off tree crowns in small footprint, high sampling density lidar data from the eastern deciduous forest in  
508 North America. *Remote sensing of Environment*, 85(3), 290-303.

509 Duncanson, L. I., Cook, B. D., Hurtt, G. C., & Dubayah, R. O. (2014). An efficient, multi-layered crown  
510 delineation algorithm for mapping individual tree structure across multiple ecosystems. *Remote Sensing of*  
511 *Environment*, 154, 378–386.

512 Eysn, L., Hollaus, M., Lindberg, E., Berger, F., Monnet, J. M., Dalponte, M., Kobal, M., Pellegrini, M., Lingua,  
513 E., Mongus, D., & Pfeifer, N. (2015). A benchmark of lidar-based single tree detection methods using  
514 heterogeneous forest data from the alpine space. *Forests*, 6(5), 1721–1747.

515 Holopainen, M., Kankare, V., Vastaranta, M., Liang, X., Lin, Y., Vaaja, M., Yu, X., Hyypä, J., Hyypä, H.,  
516 Kaartinen, H., Kukko, A., Tanhuanpää, T., & Alho, P. (2013). Tree mapping using airborne, terrestrial and  
517 mobile laser scanning—A case study in a heterogeneous urban forest. *Urban forestry & Urban greening*, 12(4),  
518 546–553.

519 Hyypä, J. (1999). Detecting and estimating attributes for single trees using laser scanner. *Photogramm J*  
520 *Finland*, 16, 27-42.

521 Hyypä, J., Kelle, O., Lehikoinen, M., & Inkinen, M. (2001). A segmentation-based method to retrieve stem  
522 volume estimates from 3-D tree height models produced by laser scanners. *IEEE Transactions on Geoscience*  
523 *and remote Sensing*, 39(5), 969-975.

524 Hyypä, J., Yu, X., Hyypä, H., Vastaranta, M., Holopainen, M., Kukko, A., Kaartinen, H., Jaakkola, A., Vaaja,  
525 M., Koskinen, J., Alho, P. (2012). Advances in forest inventory using airborne laser scanning. *Remote Sensing*,  
526 4(5), 1190–1207.

- 527 Jing, L., Hu, B., Li, J., & Noland, T. (2012). Automated delineation of individual tree crowns from LiDAR data  
528 by multi-scale analysis and segmentation. *Photogrammetric Engineering & Remote Sensing*, 78(12), 1275–  
529 1284.
- 530 Kaartinen, H., Hyyppä, J., Yu, X., Vastaranta, M., Hyyppä, H., Kukko, A., Holopainen, M., Heipke, C.,  
531 Hirschmugl, M., Morsdorf, F., Næsset, E., Pitkänen, J., Popescu, S., Solberg, S., Wolf, B. M., & Wu, J-C. (2012).  
532 An international comparison of individual tree detection and extraction using airborne laser scanning.  
533 *Remote Sensing*, 4(4), 950-974.
- 534 Kim, S., Hinckley, T., & Briggs, D. (2011). Classifying individual tree genera using stepwise cluster analysis  
535 based on height and intensity metrics derived from airborne laser scanner data. *Remote Sensing of*  
536 *Environment*, 115(12), 3329–3342.
- 537 Koch, B., Heyder, U., & Weinacker, H. (2006). Detection of individual tree crowns in airborne lidar data.  
538 *Photogrammetric Engineering & Remote Sensing*, 72(4), 357–363.
- 539 Koukoulas, S., & Blackburn, G. A. (2005). Mapping individual tree location, height and species in broadleaved  
540 deciduous forest using airborne LIDAR and multi-spectral remotely sensed data. *International Journal of*  
541 *Remote Sensing*, 26(3), 431–455.
- 542 Kwak, D. A., Lee, W. K., Lee, J. H., Biging, G. S., & Gong, P. (2007). Detection of individual trees and estimation  
543 of tree height using LiDAR data. *Journal of Forest Research*, 12(6), 425-434.
- 544 Lee, J. H., Ko, Y., & McPherson, E. G. (2016). The feasibility of remotely sensed data to estimate urban tree  
545 dimensions and biomass. *Urban forestry & urban greening*, 16, 208–220.
- 546 Lindberg, E., & Holmgren, J. (2017). Individual Tree Crown Methods for 3D Data from Remote Sensing. *Current*  
547 *Forestry Reports*, 3(1), 19–31.
- 548 Lu, X., Guo, Q., Li, W., & Flanagan, J. (2014). A bottom-up approach to segment individual deciduous trees  
549 using leaf-off lidar point cloud data. *ISPRS Journal of Photogrammetry and Remote sensing*, 94, 1-12.
- 550 Mäkinen, H. (2002). Effect of stand density on the branch development of silver birch (*Betula pendula* Roth)  
551 in central Finland. *Trees*, 16(4-5), 346-353.
- 552 Mongus, D., & Žalik, B. (2015). An efficient approach to 3D single tree-crown delineation in LiDAR data. *ISPRS*  
553 *Journal of Photogrammetry and Remote Sensing*, 108, 219-233.
- 554 Moskal, L. M., Styers, D. M., & Halabisky, M. (2011). Monitoring urban tree cover using object-based image  
555 analysis and public domain remotely sensed data. *Remote Sensing*, 3(10), 2243–2262.
- 556 Niemistö, P. (1995). Influence of initial spacing and row-to-row distance on the crown and branch properties  
557 and taper of silver birch (*Betula pendula*). *Scandinavian Journal of Forest Research*, 10(1-4), 235–244.
- 558 Packalén, P., & Maltamo, M. (2008). Estimation of species-specific diameter distributions using airborne laser  
559 scanning and aerial photographs. *Canadian Journal of Forest Research*, 38(7), 1750-1760.
- 560 Persson, A., Holmgren, J., & Soderman, U. (2002). Detecting and measuring individual trees using an airborne  
561 laser scanner. *Photogrammetric Engineering and Remote Sensing*, 68(9), 925-932.
- 562 Pitkänen J., Maltamo M., Hyyppä J., Yu X., 2004. Adaptive methods for individual tree detection on airborne  
563 laser based canopy height model. *International Archives of Photogrammetry, Remote Sensing and Spatial*  
564 *Information Sciences*, Vol. XXXVI, 187-191.

565 Popescu, S., C., Zhao, K., 2008. A voxel-based lidar method for estimating crown base height for deciduous  
566 and pine trees. *Remote Sensing of Environment* 112 (3), 767–781.

567 Rahlf, J., Breidenbach, J., Solberg, S., Astrup, R. (2015) Forest parameter prediction using an image-based  
568 point cloud: A comparison of semi-itc with ABA. *Forests*, 6, 4059–4071.

569 Reynolds, M. R., Burk, T. E., & Huang, W. C. (1988). Goodness-of-fit tests and model selection procedures for  
570 diameter distribution models. *Forest Science*, 34(2), 373–399.

571 Riikonen, A., Pumpanen, J., Mäki, M., & Nikinmaa, E. (2017). High carbon losses from established growing  
572 sites delay the carbon sequestration benefits of street tree plantings—A case study in Helsinki, Finland. *Urban  
573 Forestry & Urban Greening*, 26, 85-94.

574 Rottensteiner, F., Sohn, G., Gerke, M., Wegner, J. D., Breitkopf, U., & Jung, J. (2014). Results of the ISPRS  
575 benchmark on urban object detection and 3D building reconstruction. *ISPRS Journal of Photogrammetry and  
576 Remote Sensing*, 93, 256–271.

577 Saarinen, N., Vastaranta, M., Kankare, V., Tanhuanpää, T., Holopainen, M., Hyypä, J., & Hyypä, H. (2014).  
578 Urban-tree-attribute update using multisource single-tree inventory. *Forests*, 5(5), 1032-1052.

579 St-Onge, B., Audet, F.-A., and Bégin, J. (2015). Characterizing the height structure and composition of a boreal  
580 forest using an individual tree crown approach applied to photogrammetric point clouds. *Forests*, 6, 3899–  
581 3922.

582 Suárez, J. C., Ontiveros, C., Smith, S., & Snape, S. (2005). Use of airborne LiDAR and aerial photography in the  
583 estimation of individual tree heights in forestry. *Computers & Geosciences*, 31(2), 253–262.

584 Tanhuanpää, T., Kankare, V., Setälä, H., Yli-Pelkonen, V., Vastaranta, M., Niemi, M. T., Raisio, J., & Holopainen,  
585 M. (2017). Assessing above-ground biomass of open-grown urban trees: A comparison between existing  
586 models and a volume-based approach. *Urban forestry & urban greening*, 21, 239–246.

587 Tanhuanpää, T., Saarinen, N., Kankare, V., Nurminen, K., Vastaranta, M., Honkavaara, E., Karjalainen, M., Yu,  
588 X., Holopainen, M., & Hyypä, J. (2016). Evaluating the performance of high-altitude aerial image-based  
589 digital surface models in detecting individual tree crowns in mature boreal forests. *Forests*, 7(7), 143.

590 Tanhuanpää, T., Kankare, V., Vastaranta, M., Saarinen, N., Holopainen, M., & Raisio, J. (2015). Deriving canopy  
591 metrics of urban trees from airborne laser scanning data. In *Joint Urban Remote Sensing Event (JURSE), 2015*  
592 (pp. 1-4). IEEE.

593 Tanhuanpää, T., Vastaranta, M., Kankare, V., Holopainen, M., Hyypä, J., Hyypä, H., Alho, P., & Raisio, J.  
594 (2014). Mapping of urban roadside trees—A case study in the tree register update process in Helsinki City.  
595 *Urban forestry & urban greening*, 13(3), 562–570.

596 Vastaranta, M., Wulder, M.A., White, J.C., Pekkarinen, A., Tuominen, S., Ginzler, C., Kankare, V., et al. 2013.  
597 Airborne laser scanning and digital stereo imagery measures of forest structure: comparative results and  
598 implications to forest mapping and inventory update. *Canadian Journal of Remote Sensing*, 39, 382–395.

599 Vauhkonen, J., Ene, L., Gupta, S., Heinzl, J., Holmgren, J., Pitkänen, J., Solberg, S., Wang, Y., Weinacker, H.,  
600 Hauglin, K. M., Lien, V., Packalén, P., Gobakken, T., Koch, B., Næsset, E., Tokola, T., & Maltamo, M. (2011).  
601 Comparative testing of single-tree detection algorithms under different types of forest. *Forestry*, 85(1), 27-  
602 40.



- 603 Vauhkonen, J., Korpela, I., Maltamo, M., & Tokola, T. (2010). Imputation of single-tree attributes using  
604 airborne laser scanning-based height, intensity, and alpha shape metrics. *Remote Sensing of Environment*,  
605 114(6), 1263–1276.
- 606 Vauhkonen, J., Tokola, T., Packalén, P., & Maltamo, M. (2009). Identification of Scandinavian commercial  
607 species of individual trees from airborne laser scanning data using alpha shape metrics. *Forest Science*, 55(1),  
608 37–47.
- 609 Wang, Y., Hyyppä, J., Liang, X., Kaartinen, H., Yu, X., Lindberg, E., Holmgren, J., Qin, Y., Mallet, C., Ferraz, A.,  
610 Torabzadeh, H., Morsdorf, F., Zhu, L., Liu, J., & Alho, P. (2016). International benchmarking of the  
611 individual tree detection methods for modeling 3-D canopy structure for silviculture and forest ecology using  
612 airborne laser scanning. *IEEE Transactions on Geoscience and Remote Sensing*, 54(9), 5011–5027.
- 613 Wang, Y., Weinacker, H., & Koch, B. (2008). A lidar point cloud based procedure for vertical canopy structure  
614 analysis and 3D single tree modelling in forest. *Sensors*, 8(6), 3938-3951.
- 615 White, J.C., Coops, N.C., Wulder, M.A., Vastaranta, M., Hilker, T., Tompalski, P. (2016). Remote sensing  
616 technologies for enhancing forest inventories: A review. *Canadian Journal of Forest Research* 42, 619-641.
- 617 White, J., Stepper, C., Tompalski, P., Coops, N., & Wulder, M. (2015). Comparing ALS and image-based point  
618 cloud metrics and modelled forest inventory attributes in a complex coastal forest environment. *Forests*,  
619 6(10), 3704-3732.
- 620 White, J., Wulder, M., Vastaranta, M., Coops, N., Pitt, D., & Woods, M. (2013). The utility of image-based  
621 point clouds for forest inventory: A comparison with airborne laser scanning. *Forests*, 4(3), 518–536.
- 622 Yu, X., Hyyppä, J., Vastaranta, M., Holopainen, M., & Viitala, R. (2011). Predicting individual tree attributes  
623 from airborne laser point clouds based on the random forests technique. *ISPRS Journal of Photogrammetry  
624 and remote sensing*, 66(1), 28–37.
- 625 Zhang, C., Zhou, Y., & Qiu, F. (2015). Individual tree segmentation from LiDAR point clouds for urban forest  
626 inventory. *Remote Sensing*, 7(6), 7892–7913.