

2
3 **Increasing soil age drives shifts in plant-plant interactions from positive to negative and**
4 **affects primary succession dynamics in a subalpine glacier forefield**
5

6 Mengya Song ^a, Lei Yu ^b, Yonglei Jiang ^b,

7 Helena Korpelainen ^c and Chunyang Li ^{b,*}
8

9 ^a Key Laboratory of Geospatial Technology for the Middle and Lower Yellow River Regions,
10 College of Environment and Planning, Henan University, Kaifeng 475004, Henan, China

11 ^b College of Life and Environmental Sciences, Hangzhou Normal University, Hangzhou 310036,
12 Zhejiang, China

13 ^c Department of Agricultural Sciences, Viikki Plant Science Centre, University of Helsinki, P.O. Box
14 27, FI-00014, Finland

15 * Corresponding author: Chunyang Li

16 E-mail address: licy@hznu.edu.cn, tel: 86-571-28865327, fax: 86-571-28865333.
17

18 **Highlights**

- 19 • Facilitation frequently occurring in subalpine environments promotes survival.
20 • *Populus-Salix* interplay switches from positive to negative during primary succession.
21 • Soil age regulates plant-plant interactions in a subalpine glacier forefield.
22 • Nitrogen availability mediates plant-soil feedbacks between neighboring plants.

23 **Abstract** The stress gradient hypothesis predicts that plant-plant interactions switch between
24 facilitation (positive) and competition (negative) along environmental gradients, with facilitation
25 being more common under high abiotic stress conditions relative to more moderate abiotic stress
26 conditions. Our aim was to reveal, whether the interactions between *Populus purdomii* Rehder and
27 *Salix rehderiana* Schneider switch from positive to negative during the early stages of primary
28 succession in the Gongga Mountain glacier retreat region. We also investigated, whether soil age is
29 a major driving factor for the transformation of interactions between neighboring plants. We
30 analyzed differences between intraspecific interactions and interspecific interactions of *Populus* and
31 *Salix* under 20- and 40-year-old soil conditions, including plant biomass accumulation and
32 allocation, nutrient absorption and utilization, relative competition intensity, non-structural
33 carbohydrates, foliar carbon and nitrogen isotope composition, mesophyll cell ultrastructure, soil
34 microbial biomass and community structure, extracellular enzyme activities, and soil organic carbon
35 (SOC), soil total nitrogen (TN), soil ammonium ($\text{NH}_4^+\text{-N}$), and soil nitrate ($\text{NO}_3^-\text{-N}$) contents. We
36 found that *P. purdomii* and *S. rehderiana* growing under interspecific interactions had greater
37 contents of aboveground dry matter, belowground dry matter and total dry matter compared to
38 intraspecific interactions in 20-year-old soil. Furthermore, in 40-year-old soil conditions, the
39 phospholipid fatty acid (PLFA) analysis showed that *Populus* and *Salix* exposed to interspecific
40 interactions exhibited lower amounts of gram-positive bacteria, fungi (18:1 ω 9c) and actinomycetes,
41 and lower levels of total PLFAs than those growing under intraspecific interactions. The redundancy
42 analysis (RDA) results demonstrated that soil N was the most important parameter contributing to
43 the composition of microbial communities. In addition, the ^{15}N stable isotope labeling method
44 showed that *Populus* and *Salix* growing under interspecific interactions had higher foliage $\delta^{15}\text{N}$

45 derived from NO_3^- ($\delta^{15}\text{N}-\text{NO}_3^-$) than those growing under intraspecific interactions in 20-year-old
46 soil. In summary, our results demonstrated that *Populus-Salix* interactions exhibited positive effects
47 on survival in 20-year-old soil. Conversely, under 40-year-old soil conditions, *Populus-Salix*
48 interactions presented negative effects in relation to nutrients and elimination by neighboring plants.
49 Moreover, soil age is a major driving factor for plant-plant interactions that shift from positive to
50 negative with an increasing soil age in the Gongga Mountain glacier retreat area. In all, our results
51 support the stress gradient hypothesis. Our findings improve understanding of plant-plant
52 interactions and plant-soil feedbacks during the early stages of soil development, and of the
53 construction of vegetation communities.

54

55 **Keywords:** Plant-plant interactions; Positive and negative effects; ^{13}C and ^{15}N stable isotope
56 composition; Microbial community structure; Primary succession; Glacier retreat area.

57 **1. Introduction**

58

59 The stress gradient hypothesis (SGH) predicts that the dynamic balance between facilitation and
60 competition varies along environmental gradients, with facilitation dominating under high
61 environmental stress and competition under low environmental stress (Bertness and Callaway, 1994;
62 Travis et al., 2006; He et al., 2013). Experiments that have provided support for the stress gradient
63 hypothesis have demonstrated that the impact of facilitation increases relative to competition with
64 increasing abiotic stress (Callaway et al., 2002; Dohn et al., 2013; Michalet et al., 2014). Switching
65 between facilitation and competition could be primarily attributed to changes in external abiotic and
66 biotic conditions (e.g., changes in light, water, space, temperature, soil quality and nutrient
67 availability).

68

69 Plant-plant competitive interactions influence plant growth, distribution and abundance, species
70 coexistence, and the composition of plant communities in terrestrial ecosystems (Connell, 1983;
71 Choler et al., 2001; Brooker, 2006). Recent studies have focused on plant-plant facilitative
72 interactions that regulate individual fitness, species composition and diversity, and vegetation
73 community structures, particularly those in alpine ecosystems (Lin et al., 2012; Arroyo et al., 2015;
74 Kéfi et al., 2016; Qi et al., 2018). Some studies have shown that facilitative interactions mainly
75 occur at high elevations and under high biotic and abiotic environmental stress conditions (Choler et
76 al., 2001; Bruno et al., 2003; Kikvidze et al., 2006). Competition and facilitation are major drivers
77 of plant community structure and composition, as well as plant species dynamics (Fowler, 1986;
78 Choler et al., 2001).

79

80 Morphological and physiological traits respond to plant-plant interactions (positive or negative), as
81 well as soil carbon:nitrogen (C:N) and nitrogen:phosphorus (N:P) ratios, with repercussions on soil
82 microbial communities and soil fertility (Huston and DeAngelis, 1994; Callaway et al., 1997;
83 Zechmeister-Boltenstern et al., 2015). Jiang et al. (2018) showed that the leaf N:P ratio increased
84 steadily in a glacier retreat area, and the limiting factor for plant growth shifted from nitrogen to
85 phosphorus. Plants have developed different physiological responses to environmental stresses
86 (such as drought, nutrient deficiency or competition): they perceive and transmit stress signals, as
87 well as regulate morphology and physiological properties (Chen et al., 2010; Zhang et al., 2014;
88 Chen et al., 2015). Furthermore, aboveground and belowground subsystems and their feedbacks are
89 important processes for studying the relationships of neighboring plants (Wardle et al., 2004; Ushio
90 et al., 2016).

91

92 Plant-soil microbes associations are an important link between aboveground and belowground
93 subsystems. For instance, plant residues (e.g., leaf litter, woody debris and dead roots) affect
94 microbial activities and serve as a major source of carbon and nutrients for microbial decomposer
95 communities (Chapin et al., 1994; Kaye et al., 1997; van der Heijden et al., 2003). Meanwhile, the
96 chemical composition of plant residues affects microbial activities and community structures and, in
97 turn, soil microorganisms are primary decomposers of dead plant material and they influence plants'
98 nutritional quality (Aneja et al., 2006; Göransson et al., 2011). Aboveground and belowground
99 interactions drive the structure and functioning of ecosystems, including carbon and nutrient cycling
100 (Kardol and Wardle, 2010).

101

102 During primary succession in a glacier forefield, changes in the biochemical properties of soil (e.g.,
103 increasing nutrient availability and accumulation of organic matter) and in the composition of
104 microbial communities influence plant growth and performance traits (Hodkinson et al., 2003;
105 Walker et al., 2010; Castle et al., 2016). Nutrient resources are limited during the early stages of
106 primary succession (He and Tang, 2008; Göransson et al., 2011). Low nutrient availability
107 frequently limits plant growth and competitive ability, particularly in cold ecosystems, such as
108 alpine or subalpine environments, where low temperature restricts underground decomposition
109 processes of soil organic matter (Chapin and Shaver, 1985; Nadelhoffer et al., 1991; Robinson et al.,
110 1995). The decomposition of plant litter is closely related to soil nutrient (N and P) contents, and
111 soil microbes are crucial for plant litter decomposition (Fujii and Takeda, 2010; Zhao et al., 2013;
112 Zechmeister-Boltenstern et al., 2015).

113

114 Primary succession in the Gongga Mountain glacier forefield has resulted in the establishment of
115 following predominant plant populations: bare land, *Astragalus mahoshanicus* and *Epilobium*
116 *amurense* herb vegetation, *Hippophae hamnoides* and *Salix rehderiana* sapling scrub forest, *Salix*
117 *rehderiana* and *Populus purdomii* deciduous broad-leaved forest, *Betula utilis* and *Rhododendron*
118 *simsii* evergreen broad-leaved forest, and *Abies fabri* and *Picea brachytyla* evergreen coniferous
119 forest (Zhou et al., 2013; Lei et al., 2015). During primary succession, *S. rehderiana* initially
120 appears, then *P. purdomii*, and later, *P. purdomii* replaces *S. rehderiana* as the dominant trees in the
121 community, followed by multiple interactive relationships between *P. purdomii* and *S. rehderiana*.
122 Soil nitrogen, which plays a key role in plant growth, is limited during the early primary succession

123 (Chapin et al., 1994; He and Tang, 2008). Previous studies have shown that litter decomposition by
124 N-fixing species contributes to the accumulation of soil N during succession, with the topsoil having
125 the highest amounts of nutrients (Jia et al., 2005). Song et al. (2017) showed that N-addition could
126 regulate changes in the competitive ability between *P. purdomii* and *S. rehderiana*. However, our
127 understanding of the major effects of soil age on plant-plant interactions and plant-soil feedbacks
128 during soil development and primary succession in a glacial forefield, especially considering
129 climate warming, is limited.

130

131 In the present study, we examined the relationship between *P. purdomii* and *S. rehderiana*
132 individuals (e.g., from positive to negative, from negative to positive, or no change) along with
133 aging soil under intra- and interspecific interactions in the Gongga Mountain glacier retreat area.
134 Moreover, soil microbial community structures and soil nutrient availabilities were measured in
135 order to elucidate the mechanism of the switch between facilitation and competition. We
136 hypothesized the following: 1) plant-plant interactions predominantly show facilitation in severe
137 environments during the early stages of primary succession in subalpine glacier retreat areas,
138 whereas under relatively benign nutrient conditions, the interactions between neighboring plants
139 change into competition; 2) the composition of soil microbes and nutrient abundance are the main
140 driving forces for the shift of *Populus-Salix* interactions from positive to negative with increasing
141 soil age. To validate these hypotheses, we examined a set of morphological and physiological
142 indexes of plants, soil microbial biomass and community composition, extracellular enzyme
143 activities, and nutrient availability.

144

145 **2. Materials and methods**

146

147 *2.1. Study site*

148

149 The experiment was conducted at the Gongga Mountain Alpine Ecosystem Observation and
150 Experiment Station of the Chinese Academy of Sciences (29°34' N, 101°59' E, 3,000 m a.s.l.),
151 located on the south-eastern fringe of the Tibetan Plateau (Fig. 1). The mean annual temperature is
152 4.2°C, mean annual precipitation is approximately 1,949 mm (rainfall mainly occurring from June
153 to September), and mean annual air relative humidity is about 90.2% (Yang et al., 2014). The
154 Hailuogou glacier is one of the most representative monsoonal temperate valley glaciers. Its length
155 is around 2,000 m, width about 50-200 m and elevation between 2,850-3,000 m (He and Tang, 2008;
156 Liu et al., 2010). It is located on the eastern slope of the Gongga Mountain, with a glacier area of
157 approximately 25 km² (Liu et al., 2010). During recent years, the Hailuogou glacier has decreased in
158 size due to climate warming. During the past 120 years, the glacier retreat area has undergone a
159 long-term primary succession from pioneer species to an evergreen coniferous forest following a
160 chronosequence of soil development. Previously, this glacier forefield has been utilized when
161 studying primary succession, soil formation processes, plant-soil feedbacks and plant-plant
162 interactions (Wang et al., 2016; Song et al., 2017; Jiang et al., 2018; Jiang et al., 2019).

163

164 *2.2. Plant materials and experimental design*

165

166 Current season seedlings of *P. purdomii* and *S. rehderiana* were selected from the nursery

167 surrounding the station in September 2014. A total of 72 seedlings were collected (36 of each
168 species). The seedlings had uniform basal stem diameter (about 3.5 mm) and height (about 15 cm).
169 Two seedlings (situated 10 cm apart) were planted into a plastic pot (internal diameter: 30 cm;
170 height: 25 cm) with 20- and 40-year-old soil (two seedlings per pot; two seedlings of *P. purdomii* or
171 *S. rehderiana*, or one *P. purdomii* and one *S. rehderiana*). The 20- and 40-year-old soils used in the
172 experiments were collected from the topsoil (0-15 cm) in the glacier retreat area. The 20- and
173 40-year-old soils had soil organic carbon contents of 11.81 g·kg⁻¹ and 16.59 g·kg⁻¹, total nitrogen of
174 0.34 g·kg⁻¹ and 0.54 g·kg⁻¹, total phosphorus of 0.34 g·kg⁻¹ and 0.29 g·kg⁻¹, and total potassium of
175 6.62 g·kg⁻¹ and 10.73 g·kg⁻¹, respectively. The experimental layout was completely randomized with
176 three factors (species, soil age and planting pattern). Experiments with two species (*P. purdomii* and
177 *S. rehderiana*), two soil ages (20 and 40 years old), and three planting patterns (*P. purdomii* and *P.*
178 *purdomii* intraspecific treatment; *S. rehderiana* and *S. rehderiana* intraspecific treatment; and *P.*
179 *purdomii* and *S. rehderiana* interspecific treatment) were established (Fig. S1). Six replicates per
180 treatment were included in the study. The experiments were performed on an open site in the field
181 under natural rain conditions. Furthermore, we used the ¹⁵N isotope tracer method to study
182 differences in the absorption and utilization of N forms (NH₄⁺ and NO₃⁻) in *P. purdomii* and *S.*
183 *rehderiana* individuals. The ¹⁵N tracer solution, with labelled ¹⁵NH₄NO₃ and NH₄¹⁵NO₃, was
184 injected into the soil at a depth of 5 cm around the plants (30 mg ¹⁵NH₄⁺-N per plant or 30 mg
185 ¹⁵NO₃⁻-N mg per plant). Then, 72 h after the application of the ¹⁵N solution, the plants were
186 harvested and foliage ¹⁵N values were determined (Chen et al., 2014).

187

188 2.3. Morphological and physiological indexes

189

190 Four pots from six replicates were randomly harvested at the end of the experiment on August 25,
191 2015. All individuals were sorted into roots, stems and leaves, and all biomass samples were dried
192 (70°C, 72 h) to a constant weight to measure aboveground, belowground, and total dry matter
193 accumulation. The root to shoot (R/S) ratios were calculated. Dried leaf and root samples were
194 ground to a fine powder and were passed through a mesh (pore diameter: 250 µm). Then, the C, N
195 and P contents of these tissues were measured by the rapid dichromate oxidation technique (Nelson
196 and Sommers, 1982), the semi-micro Kjeldahl method (Fawcett, 1954), and induced plasma
197 emission spectroscopy (Parkinson and Allen, 1975), respectively.

198

199 Non-structural carbohydrates of the root and leaf tissues were estimated as starch, soluble sugar,
200 fructose, and sucrose contents. Approximately 50 mg of each dry powdered plant sample was mixed
201 with 6 mL of 80% (v/v) ethanol, incubated for 30 min in a water bath at 80°C, then centrifuged for 5
202 min at 7,000 g (Guo et al., 2016; Song et al., 2017). The ethanol extract solution was used for the
203 determination of the fructose content according to a modified resorcinol method (Murata et al.,
204 1968), the sucrose content was measured according to the procedure of Green et al. (1994), and the
205 total soluble sugar content was measured according to the anthrone-sulfuric acid method (Yemm
206 and Willis, 1954). The solid residue was hydrolyzed with 2 mL of 9.2 mol·L⁻¹ HClO₄ for 30 min to
207 measure the starch content (Yemm and Willis, 1954).

208

209 *2.4. Foliage C and N isotope composition*

210

211 Stable isotope analyses were performed at the Stable Isotope Laboratory, Chinese Academy of
212 Forestry, using a DELTA V Advantage Isotope Ratio Mass Spectrometer (Thermo Fisher Scientific,
213 Inc., Waltham, MA, USA). The carbon isotope composition ($\delta^{13}\text{C}$) was used to estimate long-term
214 tree water use efficiency (Dong et al., 2015). The $\delta^{13}\text{C}$ value of the samples was expressed relative
215 to the standard Pee Dee Belemnite (Farquhar et al., 1989) as follows: $\delta^{13}\text{C} (\text{‰}) = (R_{\text{sample}}/R_{\text{standard}} - 1)$
216 $\times 1,000$, where R_{sample} is the $^{13}\text{C}/^{12}\text{C}$ ratio of the sample, and R_{standard} is that of the standard substance.
217 The ^{15}N concentration values were converted to $\delta^{15}\text{N}$ using the following equation: $\delta^{15}\text{N} (\text{‰}) =$
218 $(T_{\text{sample}}/T_{\text{standard}} - 1) \times 1,000$, where T_{sample} is the $^{15}\text{N}/^{14}\text{N}$ ratio of the sample, and T_{standard} is that of the
219 standard substance. The overall precision of the $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ estimations was better than 0.1‰, as
220 determined from four replicates in each case.

221

222 *2.5. Leaf ultrastructural assessment*

223

224 Mesophyll cell observations are widely used to investigate changes in leaf organelles under various
225 abiotic stresses (e.g., competitive pressure or nutrient deficiency). Sections (2 mm in length) from
226 middle leaf parts were selected for transmission electron microscopy (TEM) analyses according to
227 the procedures of Zhao et al. (2009) and Song et al. (2017). In brief, the sections were fixed in 3%
228 glutaraldehyde (v/v) in 0.2 M sodium phosphate buffer (pH 7.2) for 6-8 h, post-fixed in 1% osmium
229 tetroxide for 1 h, and finally, again, in 0.2 M sodium phosphate buffer (pH 7.2) for 1-2 h. Ultrathin
230 sections (80 nm thick) were sliced, stained with uranyl acetate and lead citrate, and mounted on
231 copper grids for viewing on H-600IV TEM (Hitachi, Tokyo, Japan).

232

233 *2.6. Soil sampling and biochemical analysis*

234

235 Four soil samples from each treatment were randomly selected at the end of the experiment to
236 analyze soil biochemical properties and microbial community structures. Topsoil samples were
237 collected from depths of 0-15 cm around the seedlings, immediately transported to the laboratory for
238 further analysis and stored at 4 °C for a later assessment of microbial biomass and enzyme activities,
239 and at -20 °C for a subsequent phospholipid fatty acid analyses (performed within one week). The
240 moist soil samples were sieved (pore size: 2 mm) to remove large organic debris prior to analyses.
241 The remaining soil samples were air-dried at room temperature and used for the determination of
242 chemical properties.

243

244 Soil organic C concentrations were measured using the potassium dichromate oxidation-ferrous
245 sulfate titrimetry method (Nelson and Sommers, 1982). Total N concentrations were determined
246 using a LECO EPS-2000 CNS thermal combustion furnace (LECO Corp., St Jose, MI). Total P was
247 melted by sodium carbonate, and determined by molybdenum-blue colorimetry. Total potassium (K)
248 was digested by HF-HClO₄ and detected by flame photometry. Alkali-hydrolyzable N was extracted
249 with 2 M KCl and measured using a microplate reader (Biotek, Winooski, USA). Available P was
250 extracted with 0.5 M NaHCO₃ and determined using the molybdenum blue method. Available K
251 was extracted with 1 M ammonium acetate and quantified by flame photometry (FP640, INASA,
252 China). Soil pH was determined using a 1:2.5 ratio of soil (in grams) to water (in milliliters) using a
253 pH electrode (FE20, Mettler Toledo, Switzerland). Ammonium (NH₄⁺-N) and nitrate (NO₃⁻-N) were
254 extracted with 2 M KCl, and determined using an autoanalyser (SEAL-AA3, Germany). Microbial

255 biomass carbon (MBC) and nitrogen (MBN) were measured using fresh soil and the chloroform
256 fumigation-extraction method (Brookes et al. 1985; Vance et al. 1987) with a Multi-N/C2100
257 analyzer (Analytik Jena, Germany). MBC and MBN were calculated from the differences between
258 extractable C and N concentrations in fumigated and unfumigated samples using conversion factors
259 (k_{EC} and k_{EN}) of 0.45 for both (Vance et al. 1987). The urease activity was measured by colorimetry
260 according to the method of Kandeler and Gerber (1988) and was expressed as $\mu\text{g NH}_4\text{-N}\cdot\text{g}^{-1}\cdot\text{soil}\cdot\text{h}^{-1}$.
261 The nitrate reductase activity was determined using the colorimetric method (Kandeler, 1996) and
262 was expressed as $\mu\text{g NO}_2\text{-N}\cdot\text{g}^{-1}\cdot\text{soil}\cdot\text{h}^{-1}$. The saccharase activity was determined with a
263 spectrophotometer (U-2800, Japan) at a wavelength of 508 nm by the method described by Guan
264 (1986) and was expressed as $\mu\text{g glucose}\cdot\text{g}^{-1}\cdot\text{soil}\cdot\text{h}^{-1}$.

265

266 *2.7. Phospholipid fatty acid extraction and analysis*

267

268 The composition of soil microbial communities was assessed by the PLFA analysis using a modified
269 method of Frostegård et al. (1991), as described by Xu et al. (2015) and Duan et al. (2015). Briefly,
270 6 g of frozen soil was extracted with a chloroform:methanol:citrate acid buffer mixture (1:2:0.8,
271 v/v/v). The lipids were separated into neutral lipids, glycolipids, and phospholipids on silicic acid
272 columns. The phospholipids were subjected to mild alkaline methanolysis (Schindlbacher et al.,
273 2011). The extracted fatty acid methyl esters were analyzed using a Hewlett-Packard 6890 gas
274 chromatograph with an Ultra 2-HP capillary column (cross-linked with 5% phenyl-methyl silicone;
275 25 m, 0.22 mm ID, and 0.33 μm thickness). The peaks were identified using bacterial fatty acids as
276 standards and utilizing the Sherlock peak identification software (MIDI, Inc. Newark, DE, USA).

277 Fatty acids were quantified through a comparison of the sample peak areas with those of internal
278 standards (19:0 nonadecanoic methyl ester) (Smithwick et al., 2005). The areas measured by
279 GC-FID were used to calculate the abundance of PLFA markers, which were expressed as
280 nmol·PLFA g⁻¹ dry soil (Schindlbacher et al., 2011). We used the terminal-branched saturated PLFA
281 peaks i14:0, a15:0, i15:0, i15:0 G, i16:0, i17:0, and a17:0 as markers for gram-positive (G+) bacteria
282 (Zelles, 1997). The mono-unsaturated and cyclopropyl-saturated peaks 16:1 2OH, 17:1ω8c,
283 18:1ω5c, cy17:0, and cy19:0ω8c were used as markers for gram-negative (G-) bacteria (Federle,
284 1986; Frostegård et al., 2011). The 18:1ω9c marker was used as an indicator of fungi, whereas
285 16:1ω5c and 16:1ω11c were used as representatives of arbuscular mycorrhizal fungi (Bossio et al.,
286 1998; Swallow et al., 2009; Huang et al., 2013; Xu et al., 2015). The PLFA peaks 10 Me17:0 and 10
287 Me18:0 were selected as indicators of actinomycetes (Waldrop et al., 2004).

288

289 2.8. Data analyses

290

291 The experimental layout was completely randomized with three factors (species, planting pattern
292 and soil age). Experiments with *P. purdomii* and *S. rehderiana* were established to test intraspecific
293 interactions (two of the same species per mesocosm) or interspecific interactions (two different
294 species per mesocosm) in soil of two ages (20 and 40 years old). We calculated the relative
295 competition intensity (RCI) of both *P. purdomii* and *S. rehderiana* when subjected to different
296 planting patterns and soil ages, according to the formula described by Grace (1995) and Guo et al.
297 (2017) as follows: $RCI = (B_m - B_n)/B_n$, where B_m represents the dry biomass matter of one seedling
298 from interspecific planting and B_n represents the average dry biomass matter of corresponding

299 plants from intraspecific planting. If the RCI value is positive, the interspecific interaction has a
300 facilitative effect on the species, while if the RCI value is negative, the interspecific interaction has
301 a competitive effect on the species. Statistical analyses were conducted with the SPSS 16.0 for
302 Windows statistical software package (SPSS, Chicago, IL, USA). Individual differences among
303 treatments were compared by Tukey's tests following one-way ANOVAs. In addition, Generalized
304 Linear Model was performed using the R statistical software (version 3.3.2) to evaluate the effects
305 of species, soil age, planting pattern, and their interactions on each variable. All statistical tests were
306 considered significant at $P < 0.05$. Redundancy analysis (RDA) was further undertaken to visualize
307 the relationship between soil microbial community composition and soil properties (soil organic C,
308 biochemical parameters, and enzyme activities), performed using Canoco 5.0 (Microcomputer
309 Power, Ithaca, NY, USA).

310

311

312

313

314

315

316

317

318

319

320

321 **3. Results**

322

323 *3.1. Effects of soil age and plant-plant interactions on morphological and physiological*
324 *characteristics*

325

326 In 20-year-old soil conditions, *S. rehderiana* individuals exhibited significantly higher aboveground
327 dry matter (ADM), belowground dry matter (BDM) and total dry matter (TDM) weights when
328 grown under interspecific interactions compared to intraspecific conditions (Fig. 2). However, in
329 40-year-old soil conditions, *P. purdomii* individuals showed significantly lower ADM, BDM and
330 TDM under interspecific growth conditions when compared to intraspecific treatments (Fig. 2). On
331 the other hand, *S. rehderiana* individuals from interspecific interactions showed lower ADM, BDM
332 and TDM in 40-year-old soil compared to the values detected in 20-year-old soil. Furthermore, no
333 significant differences were observed in the R/S ratio of *P. purdomii* and *S. rehderiana* between
334 interspecific and intraspecific conditions in 20-year-old soil, whereas a significantly lower R/S ratio
335 was detected in 40-year-old soil (Fig. 2d). In 20-year-old soil conditions, the relative competition
336 intensity (RCI) value of *P. purdomii* and *S. rehderiana* was positive, which indicated that there was
337 a cooperative relationship between the two species. However, in 40-year-old soil conditions, the
338 RCI value of *P. purdomii* was negative, which indicated that *P. purdomii* was inhibited by
339 competitive stress caused by *S. rehderiana* (Fig. 3). The growth traits of seedling (ADM, BDM and
340 TDM) were significantly affected by species, soil age, species \times soil age, species \times planting pattern,
341 and soil age \times planting pattern (Table 4).

342

343 In 20-year-old soil conditions, *P. purdomii* individuals subjected to interspecific interactions showed
344 significantly higher leaf and root starch contents than those exposed to intraspecific interactions,
345 whereas *S. rehderiana* showed no significant differences in starch contents (Fig. 4a). *P. purdomii*
346 individuals subjected to interspecific treatments had significantly higher leaf and root starch
347 contents than those under intraspecific planting conditions, whereas *S. rehderiana* plants presented
348 significantly lower leaf starch contents under interspecific planting than those subjected to
349 intraspecific interactions when grown in 40-year-old soil (Fig. 4a). Under 20-year-old soil and
350 interspecific growth conditions, *P. purdomii* individuals exhibited significantly higher root total
351 sugar contents compared to those under intraspecific planting, whereas when grown in 40-year-old
352 soil conditions, *P. purdomii* individuals growing under interspecific competition showed
353 significantly lower root total sugar contents than those under intraspecific interactions (Fig. 4b). *P.*
354 *purdomii* plants from intraspecific interactions exhibited significantly lower fructose and sucrose
355 contents when grown in 40-year-old soil compared to those grown in 20-year-old soil (Fig. 4c and
356 d). The interaction of species \times soil age, species \times planting pattern and species \times soil age \times planting
357 pattern significantly affected leaf starch, while root starch was significantly affected by species, soil
358 age, planting pattern, species \times soil age and species \times planting pattern (Table 4).

359

360 *3.2. Effects of soil age and plant-plant interactions on nutrient absorption and allocation, and on*
361 *foliage C and N isotopic composition*

362

363 *P. purdomii* and *S. rehderiana* exhibited higher N and P contents but lower C:N ratios in leaves than
364 in roots (Fig. 5). In 20-year-old soil conditions, *S. rehderiana* experiencing interspecific interactions

365 showed significantly higher leaf N contents than plants grown under intraspecific planting, whereas
366 no significant differences in leaf N contents were observed in *P. purdomii*. The root N content of *P.*
367 *purdomii* under interspecific planting was significantly higher in 40-year-old soil compared to that
368 in 20-year-old soil. *P. purdomii* and *S. rehderiana* individuals exposed to interspecific interactions in
369 20-year-old soil showed higher root C:N ratios than those subjected to intraspecific conditions (Fig.
370 5c). *S. rehderiana* plants subjected to interspecific planting had higher leaf N:P ratios than those
371 under intraspecific interactions in 20-year-old soil (Fig. 5d). Using a combination of 40-year-old
372 soil and interspecific conditions, *S. rehderiana* plants exhibited lower root C:N ratios than those
373 subjected to intraspecific treatment (Fig. 5c). The interaction of species \times planting pattern, soil age
374 \times planting pattern, and species \times soil age \times planting pattern significantly affected leaf N contents
375 and leaf C:N ratio. Root N contents and root C:N ratio were significantly affected by soil age,
376 planting pattern, species \times soil age, and soil age \times planting pattern (Table 4).

377

378 Under 20- and 40-year-old soil conditions, *P. purdomii* and *S. rehderiana* subjected to interspecific
379 treatments had lower levels of $\delta^{15}\text{N}$ derived from NH_4^+ ($\delta^{15}\text{N-NH}_4^+$) than those growing under
380 intraspecific conditions (Fig. 6a). *P. purdomii* and *S. rehderiana* exposed to interspecific treatments
381 exhibited higher levels of $\delta^{15}\text{N}$ derived from NO_3^- ($\delta^{15}\text{N-NO}_3^-$) than plants grown under
382 intraspecific conditions in 20-year-old soil (Fig. 6b). Under 40-year-old soil and interspecific
383 conditions, the $\delta^{15}\text{N-NO}_3^-$ level of *S. rehderiana* significantly decreased compared to plants
384 subjected to intraspecific interactions (Fig. 6b). In all treatments, *P. purdomii* individuals presented
385 significantly higher $\delta^{13}\text{C}$ than *S. rehderiana*, whereas *S. rehderiana* individuals exposed to intra-
386 and interspecific treatments exhibited no significant differences in $\delta^{13}\text{C}$. When using 40-year-old

387 soil, *P. purdomii* plants that experienced interspecific planting had significantly lower $\delta^{13}\text{C}$ than
388 those exposed to intraspecific interactions (Fig. 6c).

389

390 3.3. Changes in leaf ultrastructure due to soil age and plant-plant interactions

391

392 When using 20-year-old soil, *P. purdomii* plants exposed to interspecific treatments showed
393 smoother and thicker cell membranes and cell walls than those grown under intraspecific
394 interactions (Fig. 8a and c). In addition, *S. rehderiana* plants that underwent interspecific treatments
395 were characterized by fewer plastoglobuli and smaller starch granules than those subjected to
396 intraspecific interactions (Fig. 8b and d). Under 40-year-old soil conditions, *P. purdomii* and *S.*
397 *rehderiana* exposed to interspecific treatments presented negative plant-plant interactions, as
398 indicated by fewer mitochondria, larger plastoglobuli and developing plasmolysis compared to
399 intraspecific treatments (Fig. 8e-h). However, in 20-year-old soil conditions, *P. purdomii* and *S.*
400 *rehderiana* grown under interspecific conditions were clearly influenced by positive plant-plant
401 interactions, as indicated by typical chloroplast structures and well-arranged thylakoid membranes,
402 and thick and continuous cytomembranes and cytodermis (Fig. 8a and b). In 40-year-old soil
403 conditions, *P. purdomii* plants subjected to intraspecific treatments showed higher numbers of
404 mitochondria with a normal structure and clear cristae, whereas *P. purdomii* plants exhibited more
405 severe damage under interspecific planting patterns, as indicated by the disappearance of
406 mitochondria, swollen chloroplasts and numerous plastoglobuli (Fig. 8e).

407

408 3.4. Response of soil biochemical properties and enzyme activities

409

410 Soil age significantly affected soil organic carbon (SOC), total nitrogen (TN), total potassium (TK),
411 alkali-hydrolyzable nitrogen (AN), available potassium (AK), pH levels, and soil enzyme activities,
412 whereas different planting patterns showed no differences in SOC, TN, total phosphorus (TP), and
413 pH values (Table 1). *P. purdomii* and *S. rehderiana* individuals exposed to interspecific treatments
414 showed higher soil AN contents than those grown under intraspecific interactions in 20-year-old soil.
415 However, in 40-year-old soil conditions, interspecific *Populus-Salix* plantings showed lower soil TN
416 contents compared to intraspecific conditions. Intra- and interspecific treatments using 40-year-old
417 soil presented higher SOC and TN contents than those using 20-year-old soil. Thus, SOC and TN
418 contents increased with aging soil. Small changes in soil enzyme activities were observed under
419 different planting patterns in the same-age soil. In addition, *Populus-Salix* interactions resulted in
420 significantly higher soil nitrate reductase activities compared to those of *Salix-Salix* interactions
421 under 20-year-old soil conditions (Table 1). On the other hand, when using 40-year-old soil,
422 *Populus-Salix* interactions exhibited significantly lower soil nitrate reductase activities. The
423 activities of soil urease, invertase, and nitrate reductase were significantly higher in 40-year-old soil
424 than in 20-year-old soil.

425

426 3.5. Soil microbial biomass and community structure characteristics

427

428 Under 20-year-old soil conditions, no significant differences were observed in MBC, MBN, or in
429 the MBC:MBN ratio among the three planting patterns (interspecific *Populus-Salix*, intraspecific
430 *Populus-Populus*, and intraspecific *Salix-Salix* planting). In 40-year-old soil conditions, interspecific

431 *Populus-Salix* planting resulted in lower MBC and MBN contents when compared to intraspecific
432 *Salix-Salix* planting (Table 1). Intraspecific *Salix-Salix* treatments showed higher MBC and MBN
433 contents in 40-year-old soil than in 20-year-old soil. Also, intraspecific *Populus-Populus* treatments
434 presented higher MBC and MBN contents in 40-year-old soil than in 20-year-old soil. MBC and
435 MBN contents were higher in 40-year-old soil compared to those in 20-year-old soil, and increasing
436 soil microbial biomass affected plant-soil interactions as well (Table 1). The increase in soil
437 microbial biomass is associated with a series of physicochemical processes between plant and soil.
438 Under 20-year-old soil conditions, the *Populus-Salix* planting pattern showed no differences in soil
439 microbes, whereas there were differences when 40-year-old soil was used.

440

441 No significant differences were observed in the total amounts of PLFAs among planting patterns in
442 20-year-old soil, whereas in 40-year-old soil, significant differences were observed (Table 2).
443 Furthermore, under 40-year-old soil conditions, interspecific *Populus-Salix* planting exhibited lower
444 densities of gram-positive bacteria, fungi (18:1 ω 9c), arbuscular mycorrhizal fungi (AMF), and
445 actinomycetes, and lower total PLFA levels when compared to intraspecific planting. Under
446 20-year-old soil conditions, interspecific planting (PS) showed higher amounts of AMF and fungi
447 (18:1 ω 9c) and higher ratios of gram+/gram- bacteria PLFAs and fungal/bacterial PLFAs when
448 compared to intraspecific planting. However, in 40-year-old soil conditions, interspecific planting
449 (PS) presented significantly lower amounts of AMF and fungi (18:1 ω 9c) and a lower ratio of
450 fungal/bacterial PLFAs compared to intraspecific treatments (PP and SS). In 20-year-old soil under
451 different planting patterns conditions, no significant changes were observed in the community
452 structure of soil microbes, whereas under 40-year-old soil conditions, the effects of interspecific

453 *Populus-Salix* planting were significantly different from those detected in intraspecific treatments.

454 The community structures of soil microbes differed between 40-year-old soil and 20-year-old soil.

455

456 3.6. Coupling relationship between soil microbes and abiotic factors

457

458 The Pearson correlation analysis on soil microbial communities and abiotic factors indicated that

459 gram-positive bacteria (G+) and gram-negative bacteria (G-) were significantly correlated with soil

460 TN, alkali-hydrolyzable N, $\text{NH}_4^+\text{-N}$, and nitrate reductase activity ($P < 0.05$, Table 3). The total

461 amounts of PLFAs showed significant positive correlations with TN, $\text{NH}_4^+\text{-N}$, $\text{NO}_3^-\text{-N}$, and nitrate

462 reductase activity. Soil actinomycetes showed significant positive correlations with SOC, TN,

463 $\text{NH}_4^+\text{-N}$, $\text{NO}_3^-\text{-N}$, and enzyme activities. In addition, AM fungi exhibited significant positive

464 correlations with TN, $\text{NH}_4^+\text{-N}$, and $\text{NO}_3^-\text{-N}$. Moreover, a significant positive correlation was

465 observed between fungi and alkali-hydrolyzable N contents. Therefore, the composition of soil

466 microbial communities was largely affected by soil nutrients (particularly N).

467

468 Redundancy analysis (RDA) revealed the effects of environmental factors on the composition of

469 soil microbial communities. The model explained 67.7% of PLFAs; RDA1 and RDA2 explained

470 53.1% and 14.6%, respectively, of the changes in the PLFA data (Fig. 7). RDA1 was positively

471 correlated with TN, AN, $\text{NH}_4^+\text{-N}$, $\text{NO}_3^-\text{-N}$, and nitrate reductase activities. RDA2 was positively

472 correlated with SOC, pH, AP, urease, and saccharase. The analysis included SOC, TN, AN, TP, AP,

473 $\text{NH}_4^+\text{-N}$, $\text{NO}_3^-\text{-N}$, pH, soil urease, saccharase, and nitrate reductase activities as environmental

474 factors. RDA showed that soil AN (which explained 36.8% of the variance, $P = 0.002$) was the most

475 important parameter contributing to the composition of soil microbial communities. Thereafter, the

476 most important ones were soil $\text{NH}_4^+\text{-N}$ (which explained 33.2% of the variance, $P = 0.002$), soil TN
477 (which explained 32.1% of the variance, $P = 0.002$), soil saccharase (which explained 29.5% of the
478 variance, $P = 0.002$), SOC (which explained 28.3% of the variance, $P = 0.006$), soil NRA (which
479 explained 28.2% of the variance, $P = 0.004$), soil urease (which explained 25.1% of the variance, P
480 = 0.01), and soil $\text{NO}_3^-\text{-N}$ (which explained 24.1% of the variance, $P = 0.006$).

481

482

483

484

485

486

487

488

489

490

491

492

493

494

495

496

497

498 **4. Discussion**

499

500 Plant-plant interactions may induce negative and positive effects on subalpine forest communities.
501 In fact, such negative and positive interactions occur widely in nature, particularly at high altitudes,
502 thereby indicating that plant species are not independently distributed. The stress gradient
503 hypothesis predicts that plant-plant interactions vary along external environmental gradients and
504 that negative interactions are more frequent in relatively productive environments, whereas positive
505 interactions are more common in severe environments (Bertness and Callaway, 1994; He et al.,
506 2013). Indeed, many previous studies have shown that plant-plant interactions play a positive role in
507 adaptation to severe environments in terms of survival, whereas neighboring plants undergo
508 competition in low-stress environments to avoid elimination by other species (Callaway and Walker,
509 1997; Fynn et al., 2005; Craine and Dybzinski, 2013). Contrary to facilitative interactions,
510 competitive interactions between neighboring species induce negative interference (Bertness and
511 Callaway, 1994; Brooker and Callaghan, 1998).

512

513 The present study demonstrated that positive and negative interactions affect the growth traits
514 (biomass allocation, nutrient absorption, mesophyll cells) of *P. purdomii* and *S. rehderiana*
515 individuals in 20- and 40-year-old soil. Furthermore, our data suggested that when soil nutrient
516 resources are limited (e.g., 20-year-old soil), *Populus-Salix* interactions result in higher above- and
517 belowground dry matter accumulation, foliage starch contents, root C/N ratios and $\delta^{15}\text{N-NO}_3^-$
518 values compared with intraspecific conditions. Thus, there are positive interactions between *P.*
519 *purdomii* and *S. rehderiana* individuals when grown in 20-year-old soil. Previous research has

520 showed that neighboring plants could improve nutrient availability and micro-climate, which would
521 then ameliorate plant performance under high abiotic stress (Maestre et al., 2005). Conversely, when
522 soil nutrient resources are moderate (e.g., 40-year-old soil), *P. purdomii* and *S. rehderiana* subjected
523 to interspecific interactions show a lower rate of dry matter accumulation and lower root/shoot
524 ratios, thus indicating that root growth is inhibited by interspecific competition and there is a
525 negative interaction. In addition, mesophyll cells of *P. purdomii* plants showed tissue damage under
526 interspecific conditions when grown in 40-year-old soil, as indicated by the disappearance of
527 mitochondria, numerous plastoglobuli and developing plasmolysis. As also previously shown, the
528 effect of neighbors is facilitation under low abiotic stress and competition under high stress
529 (Maestre et al., 2005). For instance, Kikvidze et al. (2006) found that plant-plant interactions
530 switched from competition to facilitation when the stress intensity increased (water limitation) in
531 subalpine plant communities. Moreover, Olsen et al. (2016) showed a temperature-driven switch in
532 plant-plant interactions from facilitation to competition in seminatural grasslands. Our study, which
533 suggests that changes in the level of soil nutrients mediate *Populus-Salix* interactions from positive
534 to negative in the subalpine glacier retreat areas, provides further support for the results of Brooker
535 and Callaghan (1998) indicating that the intensity of positive interactions increases and that of
536 negative interactions decreases with increasing stress and disturbance.

537

538 Plant-soil feedbacks influence plant performance (leaf and root traits) and competitive ability
539 (Kardol et al., 2007; Teste et al., 2017), which could regulate nutrient cycling in ecosystems (Bohlen
540 et al., 2001; Mariotte et al., 2018). Plant growth could change the biochemical properties of soil in a
541 way that affects plant-plant interactions. Our results showed that soil alkali-hydrolyzable N and

542 NO₃⁻-N contents decrease under interspecific treatment (PS) compared to intraspecific treatments
543 (SS and PP) in 40-year-old soil, whereas in 20-year-old soil, alkali-hydrolyzable N and NO₃⁻-N of
544 PS increase compared to those in SS and PP treatments (Table 1). Soil N gradually accumulates
545 through primary succession, which directly affects the composition of microbial communities and
546 indirectly regulates plant-plant interactions in a glacier forefield. Consequently, plant-soil feedbacks
547 lead to the replacement of early-successional species by late-successional species, as
548 early-successional species suffer from negative plant-soil feedbacks (Kulmatiski et al., 2008).
549 Moreover, our results demonstrated that SOC, NH₄⁺-N and NO₃⁻-N contents were significantly
550 higher in 40-year-old soil than in 20-year-old soil, thereby indicating that soil C and N contents
551 significantly accumulate in the studied glacier forefield over time. In addition, significantly higher
552 soil enzyme activities were observed in older soil, whereas no differences in enzyme activities in
553 relation to intra- and interspecific interactions were observed, thus indicating that the effects of
554 plant-plant interactions on enzyme activities were relatively small (Table 1). Biochemical properties
555 and extracellular enzyme activities are considered as important indicators of soil fertility
556 (Vepsäläinen et al., 2001; Mairura et al., 2007).

557

558 Soil microbes play an important role in the soil development process of primary succession,
559 particularly in glacier retreat regions. Microbes affect plant growth, and feedbacks occur because
560 plants alter microbial communities through root exudates (Westover et al., 1997; Stephan et al.,
561 2000). Wardle et al. (2004) have shown that the decomposer subsystem breaks down dead plant
562 material that indirectly regulates plant growth, and plant litter, in turn, provides energy (C) and
563 nutrients that are required by soil microbes. Moreover, Kardol et al. (2007) have reported that

564 early-successional plants generally change the composition of soil microbial communities and
565 increase the likelihood for the establishment of mid-successional species. Our findings suggest that
566 under 20-year-old soil conditions, there are no significant differences in the soil microbial biomass
567 and community composition between intra- and interspecific treatments of *P. purdomii* and *S.*
568 *rehderiana*. However, under 40-year-old soil conditions, interspecific planting of *Populus-Salix*
569 shows significantly lower amounts of gram-negative bacteria, gram-positive bacteria, AMF, fungi,
570 actinomycetes, and total PLFAs in soil when compared to intraspecific conditions (Table 2). These
571 findings suggest that the composition of soil microbial communities under interspecific conditions
572 substantially differs from that under intraspecific treatments and that belowground competitive
573 interactions decrease soil microbial biomass and change microbial community structures.

574

575 The aboveground and belowground linkages have a widely influence on ecosystem processes. The
576 aboveground biota can influence belowground subsystems. As a feedback, belowground organisms
577 could influence plant growth and the structure and function of the aboveground community (Wardle
578 et al., 2004). Plant-soil feedbacks are mechanisms of plant-plant interactions that are dependent on
579 competition for resources and changes in soil properties. Bever et al. (1997) have demonstrated that
580 soil microbes could alter the coexistence of plant species through indirect feedbacks. Furthermore,
581 the composition of soil microbial communities has been shown to have a strong effect on plant-plant
582 interactions, community dynamics and coexistence (Van der Putten and Peters, 1997; Van Der
583 Heijden et al., 2006). Our findings indicate that microbial community compositions (e.g., the
584 amount of PLFAs) have significant correlations with soil TN, AN, $\text{NH}_4^+\text{-N}$, $\text{NO}_3^-\text{-N}$ and nitrate
585 reductase activity (Table 3, $P < 0.05$). These results suggest that nitrogen availability in soil is

586 essential for microbial growth, and it directly affects the growth of plants and the interactions of
587 neighboring plants.

588

589 Soil microbial effects are incorporated into the dynamic changes of plant communities through
590 niche modification and plant-soil feedbacks (Kardol et al., 2007; Bever et al., 2010). Plant-plant
591 interactions could be directly driven by sharing of resources or competition, and the interactions of
592 plant and soil microbes could mediate the N transformation process of soil (Bohlen et al., 2001).
593 Göransson et al. (2011) have shown that the availability of resources and nutrient limitation for
594 microbial growth vary along a chronosequence in a glacier forefield, thereby suggesting that
595 microbial growth is mainly limited by C and plant growth is mainly limited by N, whereas soil
596 microbes are important competitors for N in young soil. Our results show that soil N availability
597 (NH_4^+ and NO_3^-) are the most important parameters that contribute to the composition of soil
598 microbial communities (Fig. 7). Furthermore, Reynolds et al. (2003) have demonstrated that
599 microbe-mediated partitioning of soil resources could contribute to the coexistence of plant species
600 through associations with different microbial symbionts. Hence, soil microbes alter the availability
601 of different forms of N, and indirectly affect plant-plant interactions by mediating soil resource
602 partitioning.

603 **5. Conclusions**

604

605 The present study shows that there are soil age-driven changes in *Populus-Salix* interactions, from
606 positive to negative during the primary succession in the Gongga Mountain glacier retreat area. Our
607 study reveals that *Populus-Salix* interactions exhibit facilitation for survival under 20-year-old soil,
608 whereas under 40-year-old soil conditions, *Populus-Salix* interactions involve competition to avoid
609 elimination. Moreover, soil microbes and nutrients (particularly N) are major factors for the
610 transformation of *Populus-Salix* interactions from positive to negative. Our results support the stress
611 gradient hypothesis, which predicts that positive and negative interactions vary inversely along
612 abiotic stress gradients, with positive interaction being more common under high abiotic stress
613 when compared to more benign abiotic conditions. These findings improve our understanding of
614 plant-plant interactions and plant-soil feedbacks in subalpine forest ecosystems. However, further
615 research is still needed on energy flows and nutrient cycles in plant-soil-microbe systems, on the
616 effects of neighboring plants and microbial feedbacks, and on the dynamic models of primary
617 succession in a glacier retreat area.

618

619

620 **Acknowledgements** This work was supported by the Talent Program of the Hangzhou Normal
621 University (2016QDL020). We thank LetPub (www.letpub.com) for its linguistic assistance during
622 the preparation of this manuscript.

623 **References**

- 624 Aneja, M.K., Sharma, S., Fleischmann, F., Stich, S., Heller, W., Bahnweg, G., Munch, J.C., Schloter,
625 M., 2006. Microbial colonization of beech and spruce litter-influence of decomposition site and
626 plant litter species on the diversity of microbial community. *Microb. Ecol.* 52, 127-135.
- 627 Arroyo, A.I., Pueyo, Y., Saiz, H., Alados, C.L., 2015. Plant-plant interactions as a mechanism
628 structuring plant diversity in a Mediterranean semi-arid ecosystem. *Ecol. Evol.* 5, 5305-5317.
- 629 Bertness, M.D., Callaway, R., 1994. Positive interactions in communities. *Trends Ecol. Evol.* 9,
630 191-193.
- 631 Bever, J.D., Dickie, I.A., Facelli, E., Facelli, J.M., Klironomos, J., Moora, M., Rilling, M.C., Stock,
632 W.D., Tibbett, M., Zobel, M., 2010. Rooting theories of plant community ecology in microbial
633 interactions. *Trends Ecol. Evol.* 25, 468-478.
- 634 Bever, J.D., Westover, K.M., Antonovics, J., 1997. Incorporating the soil community into plant
635 population dynamics: the utility of the feedback approach. *J. Ecol.* 85, 561-573.
- 636 Bohlen, P.J., Groffman, P.M., Driscoll, C.T., Fahey, T.J., Siccama, T.G., 2001. Plant-soil-microbial
637 interactions in a northern hardwood forest. *Ecology* 82, 965-978.
- 638 Bossio, D.A., Scow, K.M., Gunapala, N., Graham, K.J., 1998. Determinants of soil microbial
639 communities: effects of agricultural management, season, and soil type on phospholipid fatty
640 acid profiles. *Microb. Ecol.* 36, 1-12.
- 641 Brookes, P.C., Landman, A., Pruden, G., Jenkinson, D.S., 1985. Chloroform fumigation and the
642 release of soil nitrogen: a rapid direct extraction method to measure microbial biomass nitrogen
643 in soil. *Soil Biol. Biochem.* 17, 837-842.
- 644 Brooker, R.W., Callaghan, T.V., 1998. The balance between positive and negative plant interactions

645 and its relationship to environmental gradients: a model. *Oikos* 81, 196-207.

646 Brooker, R.W., 2006. Plant-plant interactions and environmental change. *New Phytol.* 171, 271-284.

647 Bruno, J.F., Stachowicz, J.J., Bertness, M.D., 2003. Inclusion of facilitation into ecological theory.

648 *Trends Ecol. Evol.* 18, 119-125.

649 Callaway, R.M., 1997. Positive interactions in plant communities and the individualistic-continuum

650 concept. *Oecologia* 112, 143-149.

651 Callaway, R.M., Walker, L.R., 1997. Competition and facilitation: a synthetic approach to

652 interactions in plant communities. *Ecology* 78, 1958-1965.

653 Callaway, R.M., Brooker, R.W., Choler, P., Kikvidze, Z., Lortie, C.J., Michalet, R., Paolini, L.,

654 Pugnaire, F.I., Cook, B.J., Aschehoug, E.T., Armas, C., Newingham, B., 2002. Positive

655 interactions among alpine plants increases with stress: a global experiment. *Nature* 417,

656 844-848.

657 Castle, S.C., Lekberg, Y., Affleck, D., Cleveland, C.C., 2016. Soil abiotic and biotic controls on

658 plant performance during primary succession in a glacial landscape. *J. Ecol.* 104, 1555-1565.

659 Chapin, F.S., Shaver, G.R., 1985. Individualistic growth response of tundra plant species to

660 environmental manipulations in the field. *Ecology* 66, 564-576.

661 Chapin, F.S., Walker, L.R., Fastie, C.L., Sharman, L.C., 1994. Mechanisms of primary succession

662 following deglaciation at Glacier Bay, Alaska. *Ecol. Monogr.* 64, 149-175.

663 Chen, J., Duan, B.L., Wang, M.L., Korpelainen, H., Li, C.Y., 2014. Intra- and inter-sexual

664 competition of *Populus cathayana* under different watering regimes. *Funct. Ecol.* 28, 124-136.

665 Chen, J., Dong, T.F., Duan, B.L., Korpelainen, H., Niinemets, Ü., Li, C.Y., 2015. Sexual

666 competition and N supply interactively affect the dimorphism and competitiveness of opposite

667 sexes in *Populus cathayana*. Plant Cell Environ. 38, 1285-1298.

668 Choler, P., Michalet, R., Callaway, R.M., 2001. Facilitation and competition on gradients in alpine
669 plant communities. Ecology 82, 3295-3308.

670 Connell, J.H., 1983. On the prevalence and relative importance of interspecific competition:
671 evidence from field experiments. Amer. Nat. 122, 661-696.

672 Craine, J.M., Dybzinski, R., 2013. Mechanisms of plant competition for nutrients, water and light.
673 Funct. Ecol. 27, 833-840.

674 Dohn, J., Dembélé, F., Karembé, M., Moustakas, A., Amévor, K. A., Hanan, N. P., 2013. Tree
675 effects on grass growth in savannas: competition, facilitation and the stress-gradient hypothesis.
676 J. Ecol. 101, 202-209.

677 Dong, T.F., Li, J.Y., Zhang, Y.X., Korpelainen, H., Niinemets, Ü., Li, C.Y., 2015. Partial shading of
678 lateral branches affects growth, and foliage nitrogen- and water-use efficiencies in the conifer
679 *Cunninghamia lanceolata* growing in a warm monsoon climate. Tree physiol. 35, 632-643.

680 Duan, B.L., Zhang, Y.B., Xu, G., Chen, J., Paquette, A., Peng, S., 2015. Long-term responses of
681 plant growth, soil microbial communities and soil enzyme activities to elevated CO₂ and
682 neighbouring plants. Agri. For. Meteorol. 213, 91-101.

683 Farquhar, G.D., Ehleringer, J.R., Hubick, K.T., 1989. Carbon isotope discrimination and
684 photosynthesis. Annu. Rev. Plant Biol. 40, 503-537.

685 Fawcett, J.K., 1954. The semi-micro Kjeldahl method for the determination of nitrogen. J. Med. Lab.
686 Technol. 12, 1-22.

687 Federle, T.W., 1986. Microbial distribution in the soil-new techniques. In: Megusar, F., Gantar, M.
688 (Eds.), Perspectives in Microbial Ecology. Slovene Society for Microbiology, Ljubljana,

689 Slovenia, pp. 493-498.

690 Fowler, N., 1986. The role of competition in plant communities in arid and semiarid regions. *Annu.*
691 *Rev. Ecol. Evol. S.* 17, 89-110.

692 Frostegård, Å., Tunlidb, A., Bååth, E., 1991. Microbial biomass measured as total lipid phosphate in
693 soils of different organic content. *J. Microbiol. Meth.* 14, 151-163.

694 Frostegård, Å., Tunlidb, A., Bååth, E., 2011. Use and misuse of PLFA measurements in soils. *Soil*
695 *Biol. Biochem.* 43, 1621-1625.

696 Fujii, S., Takeda, H., 2010. Dominant effects of litter substrate quality on the difference between
697 leaf and root decomposition process above- and belowground. *Soil Biol. Biochem.* 42,
698 2224-2230.

699 Fynn, R.W.S., Morris, C.D., Kirkman, K.P., 2005. Plant strategies and trait trade-offs influence
700 trends in competitive ability along gradients of soil fertility and disturbance. *J. Ecol.* 93,
701 384-394.

702 Göransson, H., Venterink, H.O., Bååth, E., 2011. Soil bacterial growth and nutrient limitation along
703 a chronosequence from a glacier forefield. *Soil Biol. Biochem.* 43, 1333-1340.

704 Grace, J.B., 1995. On the measurement of plant competition intensity. *Ecology* 76, 305-308.

705 Green, T.H., Mitchell, R.J., Gjerstad, D.H., 1994. Effects of nitrogen on the response of loblolly
706 pine to drought. *New Phytol.* 128, 145-152.

707 Guan, S.Y., 1986. Soil enzymes and their research methodology. *Agriculture, Beijing*, pp. 274-338.

708 Guo, Q.X., Li, J.Y., Zhang, Y.B., Zhang, J.X., Lu, D.L., Korpelainen, H., Li, C.Y., 2016.
709 Species-specific competition and N fertilization regulate non-structural carbohydrate contents
710 in two *Larix* species. *For. Ecol. Manage.* 364, 60-69.

711 Guo, Q.X., Zhang, Y.X., Wang, D.L., Zhang, Y.B., Korpelainen, H., Li, C.Y., 2017. Influence of soil
712 qualities on intra- and interspecific competition dynamics of *Larix kaempferi*, and *L. olgensis*.
713 Environ. Exp. Bot. 135, 96-105.

714 He, Q., Bertness, M.D., Altieri, A.H., 2013. Global shifts towards positive species interactions with
715 increasing environmental stress. Ecol. Lett. 16, 695-706.

716 He, L., Tang, Y., 2008. Soil development along primary succession sequences on moraines of
717 Hailuoguo Glacier, Gongga Mountain, Sichuan, China. Catena 72, 259-269.

718 Hodgkinson, I.D., Coulson, S.J., Webb, N.R., 2003. Community assembly along proglacial
719 chronosequences in the high Arctic: vegetation and soil development in north-west Svalbard. J.
720 Ecol. 91, 651-663.

721 He, Q., Bertness, M.D., Altieri, A.H., 2013. Global shifts towards positive species interactions with
722 increasing environmental stress. Ecol. Lett. 16, 695-706.

723 Huang, Z.Q., Wan, X.H., He, Z.M., Yu, Z.P., Wang, M.H., Hu, Z.H., Yang, Y.S., 2013. Soil
724 microbial biomass, community composition and soil nitrogen cycling in relation to tree species
725 in subtropical China. Soil Biol. Biochem. 62, 68-75.

726 Huston, M.A., DeAngelis, D.L., 1994. Competition and coexistence: the effects of resource
727 transport and supply rates. Amer. Nat. 144, 954-977.

728 Jia, G.M., Cao, J., Wang, C., Wang, G., 2005. Microbial biomass and nutrients in soil at the different
729 stages of secondary forest succession in Ziwoulin, northwest China. For. Ecol. Manage. 217,
730 117-125.

731 Jiang, Y.L., Song, M.Y., Zhang, S., Cai, Z.Q., Lei, Y.B., 2018. Unravelling community assemblages
732 through multi-element stoichiometry in plant leaves and roots across primary successional

733 stages in a glacier retreat area. *Plant Soil* 428, 291-305.

734 Jiang, Y.L., Lei, Y.B., Qin, W., Korpelainen, H., Li, C.Y., 2019. Revealing microbial processes and
735 nutrient limitation in soil through ecoenzymatic stoichiometry and glomalin-related soil
736 proteins in a retreating glacier forefield. *Geoderma* 338, 313-324.

737 Kandeler, E., 1996. In: Schinner, F., Ohlinger, R., Kandeler, E., Margesin, R. (Eds.), *Methods in Soil*
738 *Biology*. Springer-Verlag, Heidelberg, New York, pp. 176-179.

739 Kandeler, E., Gerber, H., 1988. Short-term assay of soil urease activity using colorimetric
740 determination of ammonium. *Biol. Fert. Soils* 6, 68-72.

741 Kardol, P., Cornips, N.J., van Kempen, M.M., Bakx-Schotman, J.M., van der Putten, W.H., 2007.
742 Microbe-mediated plant-soil feedback causes historical contingency effects in plant community
743 assembly. *Ecol. Monogr.* 77, 147-162.

744 Kardol, P., Wardle, D.A., 2010. How understanding aboveground-belowground linkages can assist
745 restoration ecology. *Trends Ecol. Evol.* 25, 670-679.

746 Kaye, J.P., Hart, S.C., 1997. Competition for nitrogen between plants and soil microorganisms.
747 *Trends Ecol. Evol.* 12, 139-143.

748 Kéfi, S., Holmgren, M., Scheffer, M., 2016. When can positive interactions cause alternative stable
749 states in ecosystems? *Funct. Ecol.* 30, 88-97.

750 Kikvidze, Z., Khetsuriani, L., Kikodze, D., Callaway, R.M., 2006. Seasonal shifts in competition
751 and facilitation in subalpine plant communities of the central Caucasus. *J. Veg. Sci.* 17, 77-82.

752 Kulmatiski, A., Beard, K., Stevens, J., Cobbold, S., 2008. Plant-soil feedbacks: a meta-analytical
753 review. *Ecol. Lett.* 11, 980-992.

754 Lei, Y.B., Zhou, J., Xiao, H.F., Duan, B.L., Wu, Y.H., Korpelainen, H., Li, C.Y., 2015. Soil

755 nematode assemblages as bioindicators of primary succession along a 120-year-old
756 chronosequence on the Hailuogou Glacier forefield, SW China. *Soil Biol. Biochem.* 88,
757 362-371.

758 Lin, Y., Berger, U., Grimm, V., Ji, Q.R., 2012. Differences between symmetric and asymmetric
759 facilitation matter: exploring the interplay between modes of positive and negative plant
760 interactions. *J. Ecol.* 100, 1482-1491.

761 Liu, Q., Liu, S.Y., Zhang, Y., Wang, X., Zhang, Y.S., Guo, W.Q., Xu, J.L., 2010. Recent shrinkage
762 and hydrological response of Hailuogou glacier, a monsoon temperate glacier on the east slope
763 of Mount Gongga, China. *J. Glaciol.* 56, 215-224.

764 Mairura, F.S., Mugendi, D.N., Mwanje, J.I., Ramisch, J.J., Mbugua, P.K., Chianu, J.N., 2007.
765 Integrating scientific and farmers' evaluation of soil quality indicators in Central Kenya.
766 *Geoderma* 139, 134-143.

767 Maestre, F.T., Valladares, F., Reynolds, J.F., 2005. Is the change of plant-plant interactions with
768 abiotic stress predictable? A meta-analysis of field results in arid environments. *J. Ecol.* 93,
769 748-757.

770 Mariotte, P., Mehrabi, Z., Bezemer, T. M., De Deyn, G. B., Kulmatiski, A., Drigo, B., Veen, G.F.,
771 van der Heijden M.G.A., Kardol, P., 2018. Plant-soil feedback: bridging natural and agricultural
772 sciences. *Trends Ecol. Evol.* 33, 129-142.

773 Michalet, R., Le Bagousse-Pinguet, Y., Maalouf, J. P., Lortie, C. J., 2014. Two alternatives to the
774 stress-gradient hypothesis at the edge of life: the collapse of facilitation and the switch from
775 facilitation to competition. *J. Veg. Sci.* 25, 609-613.

776 Murata, T., Akazawa, T., Fukuchi, S., 1968. Enzymic mechanism of starch breakdown in

777 germinating rice seeds I. An analytical study. *Plant Physiol.* 43, 1899-1905.

778 Nadelhoffer, K.J., Giblin, A.E., Shaver, G.R., Laundre, J.A., 1991. Effects of temperature and
779 substrate quality on element mineralization in six arctic soils. *Ecology* 72, 242-253.

780 Nelson, D.W., Sommers, L.E., 1982. Total carbon, organic carbon, and organic matter. *Methods of*
781 *soil analysis. Part 2: Chemical and microbiological properties.* The American Society of
782 *Agronomy, Madison, pp. 539-579.*

783 Olsen, S.L., Töpper, J.P., Skarpaas, O., Vandvik, V., Klanderud, K., 2004. From facilitation to
784 competition: temperature-driven shift in dominant plant interactions affects population
785 dynamics in seminatural grasslands. *Global Change Biol.* 22, 1915-1926.

786 Parkinson, J.A., Allen, S.E., 1975. A wet oxidation procedure suitable for the determination of
787 nitrogen and mineral nutrients in biological material. *Commun. Soil Sci. Plan.* 6, 1-11.

788 Qi, M., Sun, T., Xue, S.F., Yang, W., Shao, D.D., Martínez-López, J., 2018. Competitive ability,
789 stress tolerance and plant interactions along stress gradients. *Ecology* 99, 848-857.

790 Reynolds, H.L., Packer, A., Bever, J.D., Clay, K., 2003. Grassroots ecology: plant-microbe-soil
791 interactions as drivers of plant community structure and dynamics. *Ecology* 84, 2281-2291.

792 Robinson, C.H., Wookey, P.A., Parsons, A.N., Potter, J.A., Callaghan, T.V., Lee, J.A., Press, M.C.,
793 Welker, J.M., 1995. Responses of plant litter decomposition and nitrogen mineralization to
794 simulated environmental change in a high arctic polar semi-desert and a subarctic dwarf shrub
795 heath. *Oikos* 74, 503-512.

796 Schindlbacher, A., Rodler, A., Kuffner, M., Kitzler, B., Sessitsch, A., Zechmeister-Boltenstern, S.,
797 2011. Experimental warming effects on the microbial community of a temperate mountain
798 forest soil. *Soil Biol. Biochem.* 43, 1417-1425.

- 799 Smithwick, E.A., Turner, M.G., Metzger, K.L., Balser, T.C., 2005. Variation in NH_4^+ mineralization
800 and microbial communities with stand age in lodgepole pine (*Pinus contorta*) forests,
801 Yellowstone National Park (USA). *Soil Biol. Biochem.* 37, 1546-1559.
- 802 Song, M.Y., Yu, L., Jiang, Y.L., Lei, Y.B., Korpelainen, H., Niinemets, Ü., Li, C.Y., 2017.
803 Nitrogen-controlled intra- and interspecific competition between *Populus purdomii* and *Salix*
804 *rehderiana* drive primary succession in the Gongga Mountain glacier retreat area. *Tree Physiol.*
805 37, 799-814.
- 806 Stephan, A., Meyer, A.H., Schmid, B., 2000. Plant diversity affects culturable soil bacteria in
807 experimental grassland communities. *J. Ecol.* 88, 988-998.
- 808 Swallow, M., Quideau, S.A., Mackenzie, M.D., Kishchuk, B.E., 2009. Microbial community
809 structure and function: the effect of silvicultural burning and topographic variability in northern
810 alberta. *Soil Biol. Biochem.* 41, 770-777.
- 811 Teste, F.P., Kardol, P., Turner, B.L., Wardle, D.A., Zemunik, D., Renton, M., Laliberté, E., 2006.
812 Plant-soil feedback and the maintenance of diversity in Mediterranean-climate shrublands.
813 *Science* 355, 173-176.
- 814 Travis, J.M.J., Brooker, R.W., Clark, E.J., Dytham, C., 2006. The distribution of positive and
815 negative species interactions across environmental gradients on a dual-lattice model. *J. Theor.*
816 *Biol.* 241, 896-902.
- 817 Ushio, M., Aiba, S.I., Takeuchi, Y., Iida, Y., Matsuoka, S., Repin, R., Kitayama, K., 2016. Plant-soil
818 feedbacks and the dominance of conifers in a tropical montane forest in Borneo. *Ecol. Monogr.*
819 87, 105-129.
- 820 Van der Putten, W.H., Peters, B.A.M., 1997. How soil-borne pathogens may affect plant competition.

821 Ecology 78, 1785-1795.

822 Van der Heijden, M.G.A., Bardgett, R.D., Van Straalen, N.M., 2008. The unseen majority: soil
823 microbes as drivers of plant diversity and productivity in terrestrial ecosystems. *Ecol. Lett.* 11,
824 296-310.

825 Van der Heijden, M.G.A., Bakker, R., Verwaal, J., Scheublin, T.R., Rutten, M., Van Logtestijn, R.,
826 Staehelin, C., 2006. Symbiotic bacteria as a determinant of plant community structure and plant
827 productivity in dune grassland. *FEMS Microbiol. Ecol.* 56, 178-187.

828 Vance, E.D., Brookes, P.C., Jenkinson, D.S., 1987. An extraction method for measuring soil
829 microbial biomass C. *Soil Biol. Biochem.* 19, 703-707.

830 Waldrop, M.P., Zak, D.R., Sinsabaugh, R.L., 2004. Microbial community response to nitrogen
831 deposition in northern forest ecosystems. *Soil Biol. Biochem.* 36, 1443-1451.

832 Vepsäläinen, M., Kukkonen, S., Vestberg, M., Sirviö, H., Niemi, R.M., 2001. Application of soil
833 enzyme activity test kit in a field experiment. *Soil Biol. Biochem.* 33, 1665-1672.

834 Walker, L.R., Wardle, D.A., Bardgett, R.D., Clarkson, B.D., 2010. The use of chronosequences in
835 studies of ecological succession and soil development. *J. Ecol.* 98, 725-736.

836 Wang, J.P., Wu, Y.H., Zhou, J., Bing, H.J., Sun, H.Y., 2016. Carbon demand drives microbial
837 mineralization of organic phosphorus during the early stage of soil development. *Biol. Fert.*
838 *Soils* 52, 825-839.

839 Wardle, D.A., Bardgett, R.D., Klironomos, J.N., Setälä, H., Van Der Putten, W.H., Wall, D.H., 2004.
840 Ecological linkages between aboveground and belowground biota. *Science* 304, 1629-1633.

841 Westover, K.M., Kennedy, A.C., Kelley, S.E., 1997. Patterns of rhizosphere microbial community
842 structure associated with co-occurring plant species. *J. Ecol.* 85, 863-873.

843 Xu, G., Chen, J., Berninger, F., Pumpanen, J., Bai, J.W., Yu, L., Duan, B.L., 2015. Labile,
844 recalcitrant, microbial carbon and nitrogen and the microbial community composition at two
845 *Abies faxoniana* forest elevations under elevated temperatures. *Soil Biol. Biochem.* 91, 1-13.

846 Yang, Y., Wang, G.X., Shen, H.H., Yang, Y., Cui, H.J., Liu, Q., 2014. Dynamics of carbon and
847 nitrogen accumulation and C:N stoichiometry in a deciduous broadleaf forest of deglaciated
848 terrain in the eastern Tibetan Plateau. *For. Ecol. Manage.* 312, 10-18.

849 Yemm, E.W., Willis, A.J., 1954. The estimation of carbohydrates in plant extracts byanthrone.
850 *Biochemistry* 57, 508-514.

851 Zechmeister-Boltenstern, S., Keiblinger, K.M., Mooshammer, M., Peñuelas, J., Richter, A., Sardans,
852 J., Wanek, W., 2015. The application of ecological stoichiometry to plant-microbial-soil organic
853 matter transformations. *Ecol. Monogr.* 85, 133-155.

854 Zelles, L., 1997. Phospholipid fatty acid profiles in selected members of soil microbial communities.
855 *Chemosphere* 35, 275-294.

856 Zhao, H.M., Huang, G, Ma, J., Li, Y., Tang, L.S., 2014. Decomposition of aboveground and root
857 litter for three desert herbs: mass loss and dynamics of mineral nutrients. *Biol. Fert. Soils* 50,
858 745-753.

859 Zhao, H.X., Li, Y., Duan, B.L., Korpelainen, H., Li, C.Y., 2009. Sex-related adaptive responses of
860 *Populus cathayana* to photoperiod transitions. *Plant Cell Environ.* 32, 1401-1411.

861 Zhou, J., Wu, Y.H., Prietzel, J., Bing, H.J., Yu, D., Sun, S.Q., Luo, J., Sun, H.Y., 2013. Changes of
862 soil phosphorus speciation along a 120-year soil chronosequence in the Hailuogou Glacier
863 retreat area (Gongga Mountain, SW China). *Geoderma* 195, 251-259.

864 **Table 1.** Changes of soil chemical properties, extracellular enzyme activities, soil microbial biomass C, and
 865 microbial biomass N under different soil age and plant-plant interaction conditions.

Soil properties	20-year-old soil			40-year-old soil		
	PP	SS	PS	PP	SS	PS
SOC (g·kg ⁻¹)	6.52 ± 0.24 b	7.57 ± 0.31 b	6.98 ± 0.11 b	16.75 ± 0.41 a	17.47 ± 0.89 a	16.38 ± 0.19 a
TN (g·kg ⁻¹)	0.49 ± 0.03 b	0.56 ± 0.02 b	0.56 ± 0.04 b	0.74 ± 0.03 a	0.81 ± 0.06 a	0.73 ± 0.02 a
TP (g·kg ⁻¹)	1.30 ± 0.05 a	1.23 ± 0.02 a	1.26 ± 0.05a	1.18 ± 0.10 a	1.54 ± 0.19 a	1.17 ± 0.07 a
TK (g·kg ⁻¹)	48.77 ± 0.32 ab	48.77 ± 0.17 ab	50.02 ± 0.26 a	48.12 ± 0.80 abc	45.97 ± 0.23 c	46.97 ± 0.78 bc
AN (mg·kg ⁻¹)	93.30 ± 2.77 c	95.15 ± 3.00 c	102.68 ± 5.58 c	143.98 ± 6.16 ab	154.21 ± 3.73 a	129.77 ± 3.91 b
AP (mg·kg ⁻¹)	60.79 ± 0.80 ab	54.76 ± 1.27 bc	61.17 ± 1.83 ab	53.46 ± 1.49 c	62.54 ± 0.59 a	60.25 ± 2.48 ab
AK (mg·kg ⁻¹)	187.04 ± 1.85 ab	183.47 ± 3.74 b	207.64 ± 5.89 a	181.17 ± 4.98 b	165.38 ± 1.94 b	183.63 ± 7.06 b
NH ₄ ⁺ -N (mg·kg ⁻¹)	15.47 ± 0.93 b	15.76 ± 0.42 b	15.63 ± 1.37 b	21.04 ± 1.07 a	22.51 ± 1.39 a	19.63 ± 0.79 a
NO ₃ ⁻ -N (mg·kg ⁻¹)	33.68 ± 2.22 b	31.44 ± 1.35 b	33.53 ± 2.94 b	42.38 ± 1.47 a	44.16 ± 1.85 a	40.58 ± 3.19 a
MBC (mg·kg ⁻¹)	256.26 ± 4.67 c	274.11 ± 3.25 c	246.69 ± 5.44 cd	345.40 ± 5.98 ab	347.05 ± 4.96 a	323.76 ± 3.51 b
MBN (mg·kg ⁻¹)	9.24 ± 0.25 c	9.51 ± 0.14 c	8.90 ± 0.19 c	15.49 ± 0.38 ab	16.55 ± 0.35 a	14.53 ± 0.13 b
MBC:MBN ratio	27.76 ± 0.96 a	28.83 ± 0.49 a	27.72 ± 0.14 a	22.33 ± 0.73 b	20.98 ± 0.22 b	22.29 ± 0.34 b
Urease activity (µg NH ₄ -N·g ⁻¹ ·soil·h ⁻¹)	15.19 ± 0.39 b	13.81 ± 0.28 b	13.17 ± 0.47 b	19.23 ± 0.47 a	19.94 ± 0.41 a	19.52 ± 0.81 a
Nitrate reductase activity (µg NO ₂ -N·g ⁻¹ ·soil·h ⁻¹)	5.39 ± 0.25 bc	4.06 ± 0.51 b	4.83 ± 0.33 bc	6.73 ± 0.57 ab	7.68 ± 0.17 a	6.47 ± 0.57 ab
Saccharase activity (µg glucose·g ⁻¹ ·soil·h ⁻¹)	132.79 ± 3.21 b	139.46 ± 3.38 b	126.98 ± 1.86 b	184.81 ± 2.92 a	180.71 ± 4.09 a	175.89 ± 4.06 a
pH value	6.65 ± 0.05 a	6.80 ± 0.10 a	6.68 ± 0.05 a	6.03 ± 0.15 b	6.52 ± 0.19 ab	6.00 ± 0.11 b

866 SOC, soil organic carbon; TN, soil total nitrogen; TP, soil total phosphorus; TK, soil total potassium; AN, soil
 867 alkali-hydrolyzable nitrogen; AP, soil available phosphorus; AK, soil available potassium; NH₄⁺-N, soil
 868 ammonium nitrogen; NO₃⁻-N, soil nitrate nitrogen; MBC, soil microbial biomass C; MBN, soil microbial biomass
 869 N. PP, soil sample from *P. purdomii* and *P. purdomii* intraspecific interaction; SS, soil sample from *S. rehderiana*
 870 and *S. rehderiana* intraspecific interaction; PS, soil sample from *P. purdomii* and *S. rehderiana* interspecific
 871 interaction. Values followed by same letters in the same row are not significantly different at the $P < 0.05$ level
 872 according to Tukey's test. Values are expressed as means ± SE ($n = 4$).

873 **Table 2.** Changes of soil microbial PLFA parameters (nmol g⁻¹ soil) under difference soil age and plant-plant
 874 interactions conditions.

PLFA markers	20-year-old soil			40-year-old soil		
	PP	SS	PS	PP	SS	PS
Gram-positive						
i14:0	0.036±0.003 a	0.034±0.003 a	0.022±0.005 a	0.036±0.008 a	0.049±0.007 a	0.036±0.004 a
i15:0	0.497±0.039 a	0.503±0.033 a	0.533±0.035 a	0.534±0.121 a	0.627±0.056 a	0.405±0.010 a
i15:1 G	0.083±0.009 a	0.082±0.007 ab	0.028±0.016 b	0.090±0.015 a	0.093±0.013 a	0.071±0.005 ab
i16:0	0.147±0.004 b	0.171±0.006 ab	0.164±0.003 b	0.204±0.012 a	0.205±0.007 a	0.151±0.007 b
i17:0	0.156±0.006 ab	0.143±0.010 ab	0.137±0.008 ab	0.174±0.015 ab	0.205±0.029 a	0.115±0.009 b
a15:0	0.329±0.024 a	0.337±0.015 a	0.361±0.038 a	0.406±0.077 a	0.466±0.033 a	0.303±0.013 a
a17:0	0.145±0.003 ab	0.167±0.006 ab	0.093±0.017 b	0.190±0.013 ab	0.206±0.023 a	0.126±0.005 ab
Σ Gram-positive	1.392±0.076 ab	1.437±0.054 ab	1.339±0.008 ab	1.634±0.244 ab	1.851±0.169 a	1.207±0.033 b
Gram-negative						
17:1 ω8c	0.060±0.003 a	0.085±0.011 a	nd	0.063±0.005 a	0.072±0.018 a	0.048±0.004 a
18:1 ω5c	0.126±0.011 a	nd	nd	0.164±0.022 a	0.139±0.008 a	0.126±0.012 a
cy17:0	0.333±0.005 bc	0.400±0.005 ab	0.360±0.016 bc	0.377±0.017 abc	0.468±0.038 a	0.285±0.017 c
cy19:0 ω8c	0.362±0.004 a	0.363±0.018 a	0.369±0.009 a	0.359±0.032 a	0.439±0.016 a	0.185±0.018 b
16:1 2OH	nd	0.041±0.005 b	nd	0.028±0.005 bc	0.063±0.007 a	0.015±0.001 cd
Σ Gram-negative	0.882±0.018 bc	0.889±0.054 bc	0.729±0.036 c	0.991±0.145 ab	1.181±0.151 a	0.659±0.050 c
AM fungi						
16:1 ω5c	0.245±0.003 c	0.264±0.003 bc	0.341±0.043 abc	0.364±0.034 ab	0.400±0.024 a	0.233±0.008 c
16:1 ω11c	0.115±0.005 b	0.125±0.005 ab	0.132±0.008 ab	0.163±0.022 ab	0.176±0.017 a	0.106±0.005 b
Σ AMF	0.360±0.003 c	0.389±0.005 bc	0.473±0.042 abc	0.527±0.036 ab	0.576±0.032 a	0.340±0.012 c
Fungi						
18:1 ω9c	0.565±0.024 bc	0.533±0.031 bc	0.571±0.012 bc	0.669±0.053 ab	0.768±0.065 a	0.390±0.021 c
Actinomycetes						
10 Me17:0	0.021±0.002 a	0.021±0.001 a	0.023±0.002 a	0.039±0.009 a	0.039±0.008 a	0.021±0.005 a
10 Me18:0	0.136±0.004 b	0.148±0.012 b	0.108±0.011 b	0.262±0.023 a	0.228±0.024 a	0.122±0.002 b
Σ Actinomycetes	0.157±0.002 b	0.169±0.012 b	0.131±0.009 b	0.301±0.029 a	0.267±0.032 a	0.144±0.005 b
Total PLFAs	3.356±0.162 bc	3.418±0.211 bc	3.244±0.122 bc	4.122±0.765 ab	4.644±0.639 a	2.739±0.109 c
Gram+/Gram- ratio	1.576±0.098 a	1.617±0.094 a	1.842±0.127 a	1.632±0.180 a	1.563±0.051 a	1.839±0.155 a
Fungi/Bacteria ratio	0.249±0.028 abc	0.229±0.012 bc	0.276±0.005 a	0.258±0.019 ab	0.253±0.009 abc	0.209±0.015 c

875 PP, soil sample from *P. purdomii* and *P. purdomii* intraspecific interaction; SS, soil sample from *S. rehderiana* and
 876 *S. rehderiana* intraspecific interaction; PS, soil sample from *P. purdomii* and *S. rehderiana* interspecific interaction.
 877 Values followed by different letters in the same row are significantly different at the $P < 0.05$ level according to
 878 Tukey's test. nd, not detected. Values are expressed as means ± SE ($n = 4$).

Table 3. Pearson correlation coefficients between soil biochemical properties and microbial lipid biomarkers.

	SOC	TN	AN	pH	NH ₄ ⁺ -N	NO ₃ ⁻ -N	Urease	Saccharase	Nitrate reductase
G+	0.392	0.549*	0.559*	0.058	0.525*	0.415	0.346	0.407	0.507*
G-	0.368	0.481*	0.529*	0.078	0.508*	0.402	0.382	0.399	0.521*
G+/G- ratio	-0.016	0.028	-0.059	-0.082	-0.103	-0.058	-0.175	-0.077	-0.172
Bacteria	0.391	0.533*	0.559*	0.068	0.529*	0.418	0.368	0.413	0.524*
Fungi	0.275	0.473*	0.471*	0.096	0.438	0.377	0.232	0.276	0.433
F/B ratio	-0.211	-0.104	-0.070	0.158	-0.081	-0.023	-0.249	-0.246	-0.069
AM Fungi	0.426	0.609**	0.644**	-0.066	0.597*	0.528*	0.333	0.416	0.515**
Actinomycetes	0.650**	0.690**	0.746***	-0.346	0.669**	0.598**	0.582*	0.707**	0.615**
Total PLFAs	0.419	0.568*	0.601**	0.017	0.562*	0.475*	0.378	0.438	0.541*

880 G+, gram-positive bacterial PLFAs; G-, gram-negative bacterial PLFAs; G+/G-, the ratio of gram-positive to
881 gram-negative bacterial PLFAs; F/B, the ratio of fungal (18:1 ω9c) to bacterial PLFAs; AM Fungi, arbuscular
882 mycorrhizal fungi; SOC, soil organic carbon; TN, soil total nitrogen; AN, alkali-hydrolyzable nitrogen. Significant
883 correlations are shown in bold. * $P < 0.05$; ** $P < 0.01$; *** $P < 0.001$.

884 **Table 4.** Generalized Linear Model (GLM) results of species, soil age, plant pattern, and their interactions effects
885 on morphological traits, nutrient allocation and non-structural carbohydrates. F_S , species effect; F_Y , soil age effect;
886 F_P , planting pattern effect; $F_{S \times Y}$, species \times soil age effect; $F_{S \times P}$, species \times planting pattern effect; $F_{Y \times P}$, soil age \times
887 planting pattern effect; $F_{S \times Y \times P}$, species \times soil age \times planting pattern effect. Significant effects ($P < 0.05$) are shown
888 in a bold font.

Variable	$P > F_S$	$P > F_Y$	$P > F_P$	$P > F_{S \times Y}$	$P > F_{S \times P}$	$P > F_{Y \times P}$	$P > F_{S \times Y \times P}$
Aboveground	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	0.203
Belowground	<0.001	<0.001	0.055	0.002	<0.001	<0.001	0.444
Total biomass	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	0.099
Root:shoot ratio	<0.001	0.033	0.025	0.126	0.868	0.002	0.868
Leaf N	0.019	<0.001	0.641	0.051	<0.001	0.009	<0.001
Leaf P	0.322	0.098	0.953	0.269	0.150	0.860	0.861
Leaf C:N	0.002	<0.001	0.099	0.118	0.007	0.015	0.013
Leaf N:P	0.472	<0.001	0.811	0.023	<0.001	0.052	0.012
Root N	0.927	<0.001	<0.001	<0.001	0.116	<0.001	0.415
Root P	0.015	0.015	0.002	0.166	0.351	0.352	0.814
Root C:N	0.307	<0.001	<0.001	<0.001	<0.001	<0.001	0.298
Root N:P	<0.001	0.883	0.401	<0.001	0.011	0.089	0.304
Leaf starch	0.333	0.866	0.004	<0.001	<0.001	0.275	<0.001
Leaf total sugar	<0.001	<0.001	0.595	<0.001	0.149	0.814	0.072
Leaf sucrose	<0.001	<0.001	<0.001	0.912	0.407	0.231	0.013
Leaf fructose	<0.001	<0.001	0.013	0.618	0.266	0.192	0.197
Root starch	<0.001	0.002	0.039	0.001	<0.001	0.772	0.295
Root total sugar	<0.001	<0.001	0.269	<0.001	0.538	<0.001	0.031
Root sucrose	<0.001	0.002	0.408	0.406	0.051	0.363	0.442
Root fructose	<0.001	<0.001	0.134	0.515	0.006	0.468	0.093

889

890 **Figure legends**

891 **Figure 1.** The experiment was conducted on the Gongga Mountain glacier forefield, which is located on the
892 south-eastern fringe of the Tibetan Plateau (modified from Zhou et al. 2013). A, experimental station; S3,
893 20-year-old soil; S4, 40-year-old soil. S1-S7 belong to a 120-year-old soil chronosequence.

894

895 **Figure 2.** Growth traits of *P. purdomii* and *S. rehderiana* as affected by different soil age and plant-plant
896 interaction conditions. Different letters above bars denote statistically significant differences between treatments at
897 the $P < 0.05$ level according to Tukey's test. Generalized Linear Model was conducted to evaluate the effects of
898 different factors and their interactions (Table 4). Treatment codes are as follows: P/PP, *P. purdomii* individuals
899 from intraspecific interaction; P/PS, *P. purdomii* individuals from interspecific interaction; S/SS, *S. rehderiana*
900 individuals from intraspecific interaction; S/PS, *S. rehderiana* individuals from interspecific interaction. Values
901 are expressed as means \pm SE ($n = 4$).

902

903 **Figure 3.** Relative competition intensity of *P. purdomii* and *S. rehderiana* under different soil age and plant-plant
904 interaction conditions. Different letters above bars denote statistically significant differences between treatments at
905 the $P < 0.05$ level according to Tukey's test.

906

907 **Figure 4.** Leaf and root starch (a), total sugar (b), sucrose (c) and fructose (d) contents in *P. purdomii* and *S.*
908 *rehderiana* plants under different soil age and plant-plant interaction conditions. Treatment codes are the same as
909 those in Figure 2. Different letters above bars denote statistically significant differences between treatments at the
910 $P < 0.05$ level according to Tukey's test. P values for interactive effects according to Generalized Linear Model
911 was in Table 4. Values are expressed as means \pm SE ($n = 4$).

912

913 **Figure 5.** Leaf and root nitrogen (a), phosphorus (b), the ratio of carbon to nitrogen (c) and the ratio of nitrogen to
914 phosphorus (d) contents in *P. purdomii* and *S. rehderiana* plants under different soil age and plant-plant interaction
915 conditions. Treatment codes are the same as those in Figure 2. Different letters above bars denote statistically
916 significant differences between treatments at the $P < 0.05$ level according to Tukey's test. P values for interactive
917 effects according to Generalized Linear Model was in Table 4. Values are expressed as means \pm SE ($n = 4$).

918

919 **Figure 6.** Foliage $\delta^{15}\text{N}$ derived from NH_4^+ ($\delta^{15}\text{N-NH}_4^+$), $\delta^{15}\text{N}$ derived from NO_3^- ($\delta^{15}\text{N-NO}_3^-$) and carbon isotope

920 composition ($\delta^{13}\text{C}$) of *P. purdomii* and *S. rehderiana* under different soil age and plant-plant interaction conditions.
921 Treatment codes are the same as those in Figure 2. Different letters above bars denote statistically significant
922 differences between treatments at the $P < 0.05$ level according to Tukey's test. Values are expressed as means \pm
923 SE ($n = 4$).

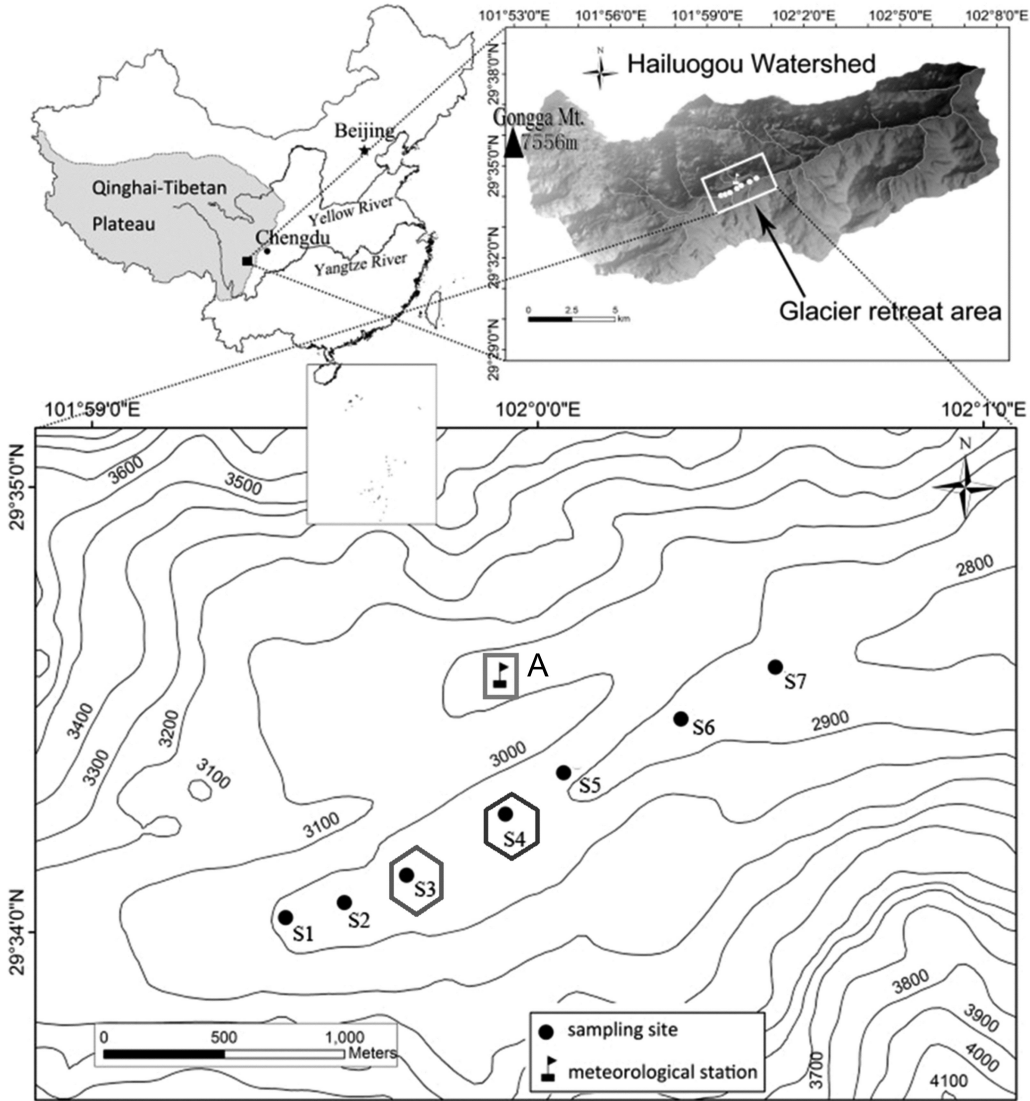
924

925 **Figure 7.** Redundancy analysis (RDA) between soil microbial PLFAs and environmental parameters. White
926 square, soil sample from interspecific interaction of *P. purdomii* and *S. rehderiana* in 20-year-old soil; white star,
927 soil sample from intraspecific interaction of *P. purdomii* and *P. purdomii* in 20-year-old soil; white inverse triangle,
928 soil sample from intraspecific interaction of *S. rehderiana* and *S. rehderiana* in 20-year-old soil; black square, soil
929 sample from interspecific interaction of *P. purdomii* and *S. rehderiana* in 40-year-old soil; black star, soil sample
930 from intraspecific interaction of *P. purdomii* and *P. purdomii* in 40-year-old soil; black inverse triangle, soil sample
931 from intraspecific interaction of *S. rehderiana* and *S. rehderiana* in 40-year-old soil; G+, gram-positive bacteria
932 PLFAs; G-, gram-negative bacteria PLFAs; G+/G-, the ratio of gram-positive bacteria to gram-negative bacteria
933 PLFAs; F/B, the ratio of fungi (18:1 $\omega 9\text{c}$) to bacteria PLFAs; AM Fungi, arbuscular mycorrhizal fungi; MBC, soil
934 microbial biomass C; MBN, soil microbial biomass N; SOC, soil organic carbon; TN, soil total nitrogen; TP, soil
935 total phosphorus; AN, soil alkali-hydrolyzable nitrogen; AP, soil available phosphorus; $\text{NH}_4^+\text{-N}$, soil ammonium
936 nitrogen; $\text{NO}_3^-\text{-N}$, soil nitrate nitrogen.

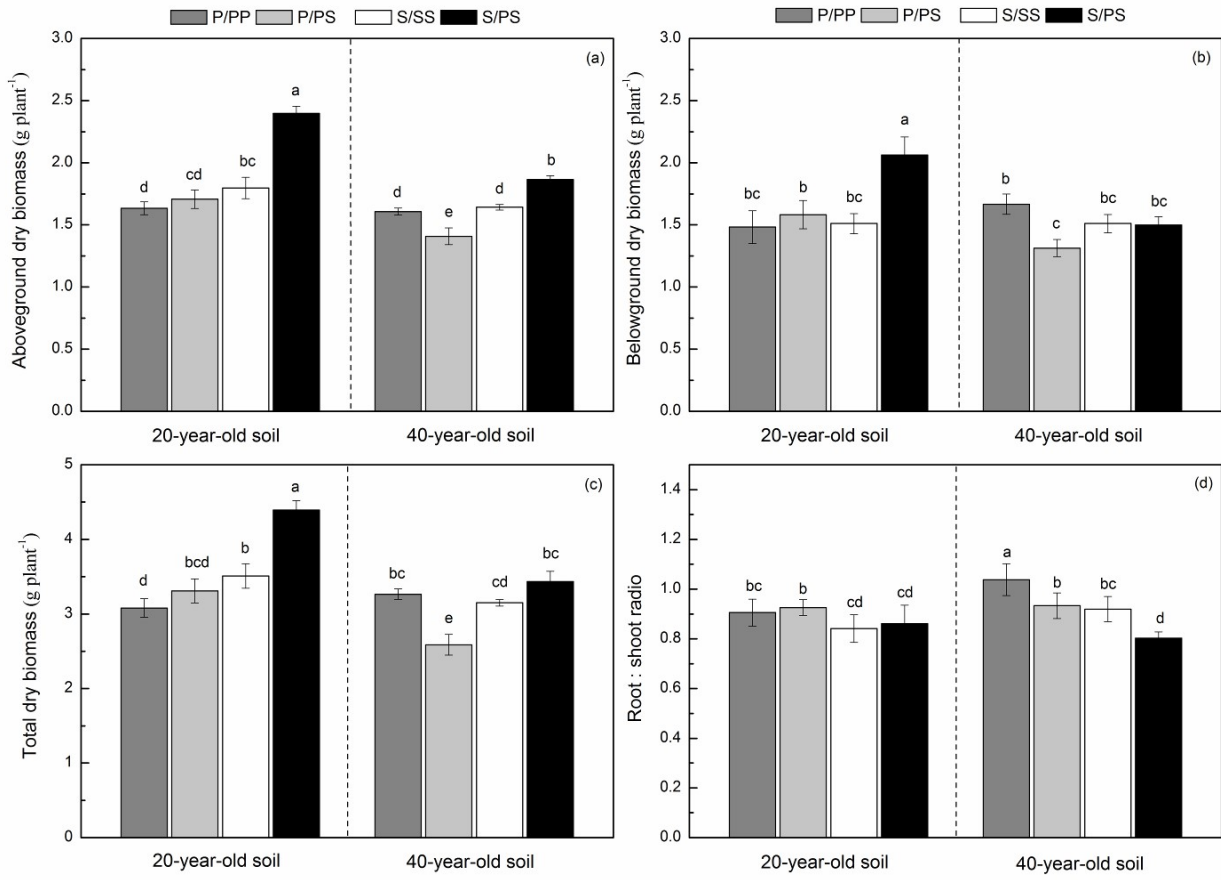
937

938 **Figure 8.** Transmission electron microscopy observations of mesophyll cells in *P. purdomii* and *S. rehderiana*
939 under different soil age and planting pattern conditions. (a) *P. purdomii* from interspecific interaction in
940 20-year-old soil; (b) *S. rehderiana* from interspecific interaction in 20-year-old soil; (c) *P. purdomii* from
941 intraspecific interaction in 20-year-old soil; (d) *S. rehderiana* from intraspecific interaction in 20-year-old soil; (e)
942 *P. purdomii* from interspecific interaction in 40-year-old soil; (f) *S. rehderiana* from interspecific interaction in
943 40-year-old soil age; (g) *P. purdomii* from intraspecific interaction in 40-year-old soil; (h) *S. rehderiana* from
944 intraspecific interaction in 40-year-old soil. C, chloroplast; S, starch granule; P, plastoglobulus; G, granum; CW,
945 cell wall; V, vacuole; M, mitochondrion; N, nucleus. The bars correspond to 1 μm in all panels.

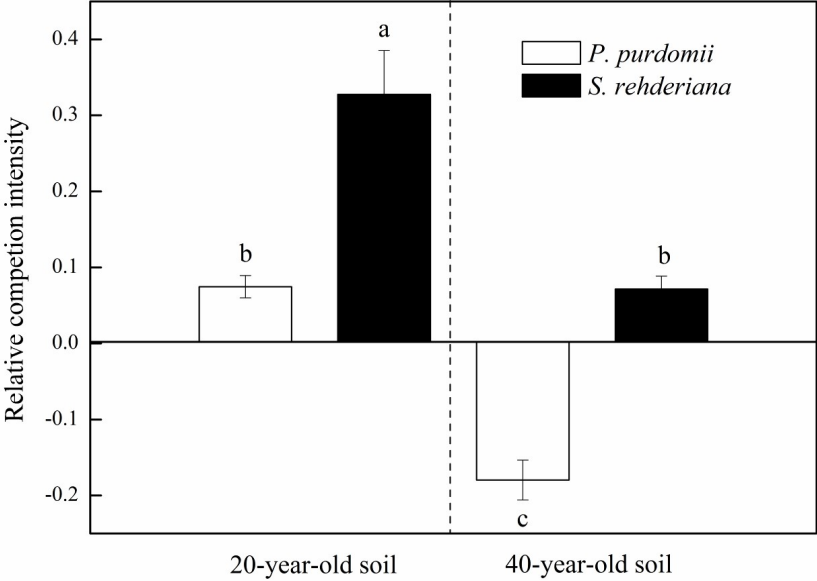
946 **Figure 1.**



947

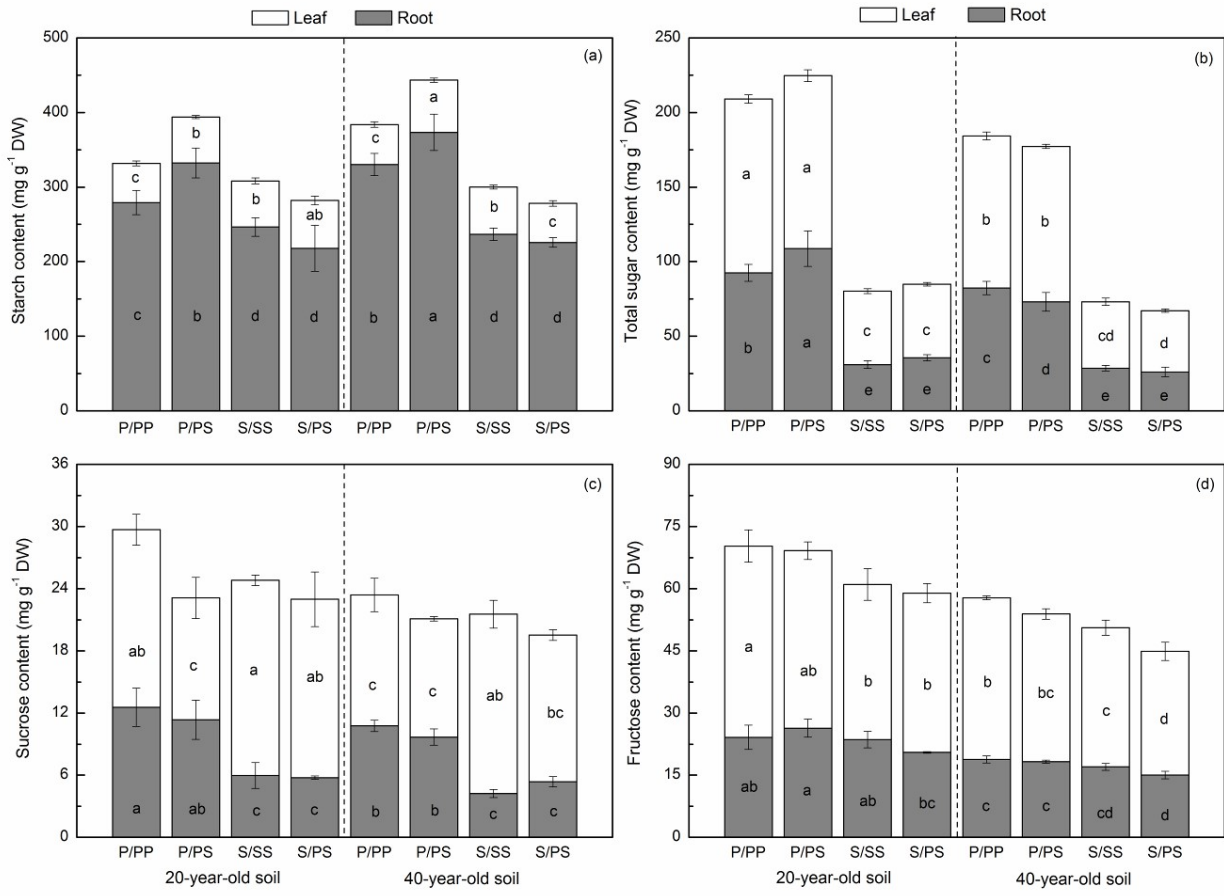


951 **Figure 3.**



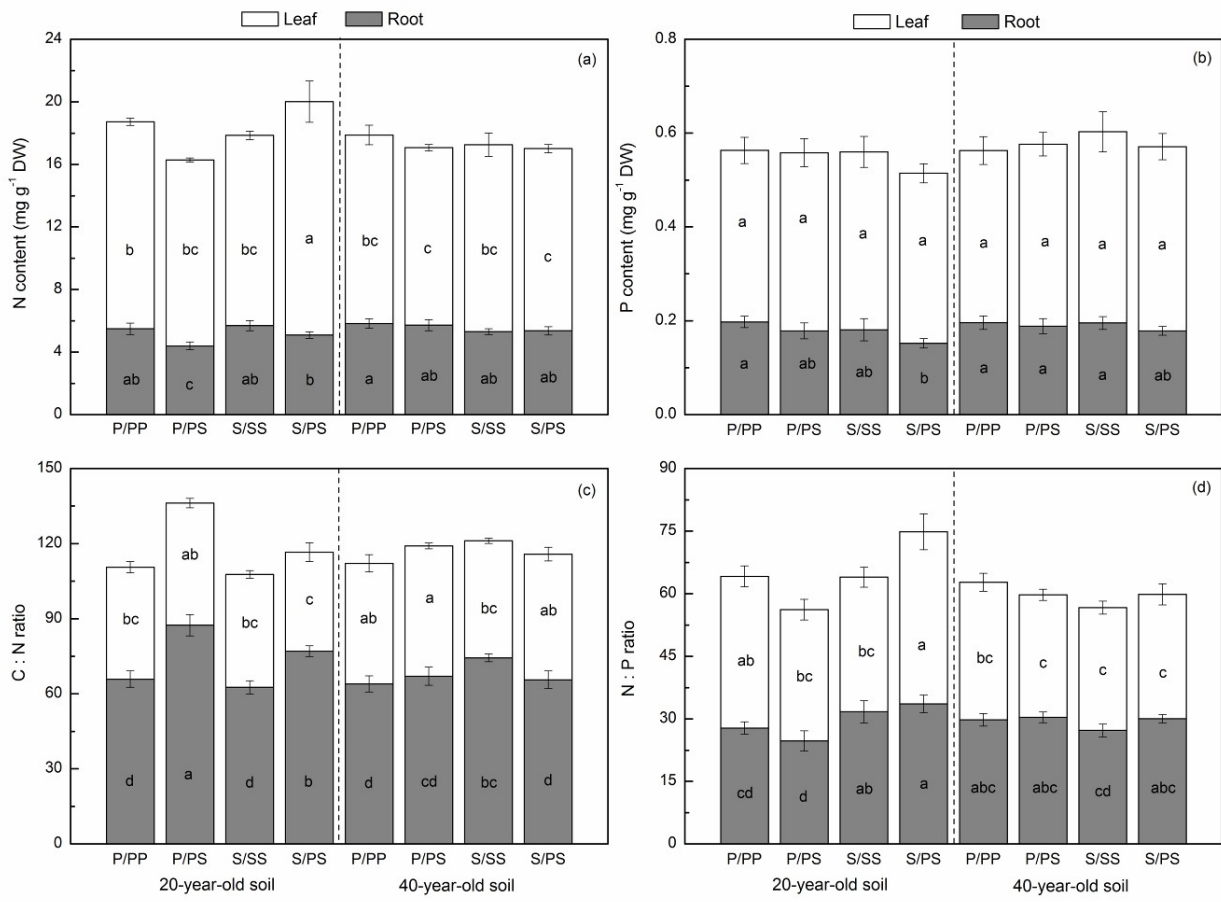
952

953



955

956

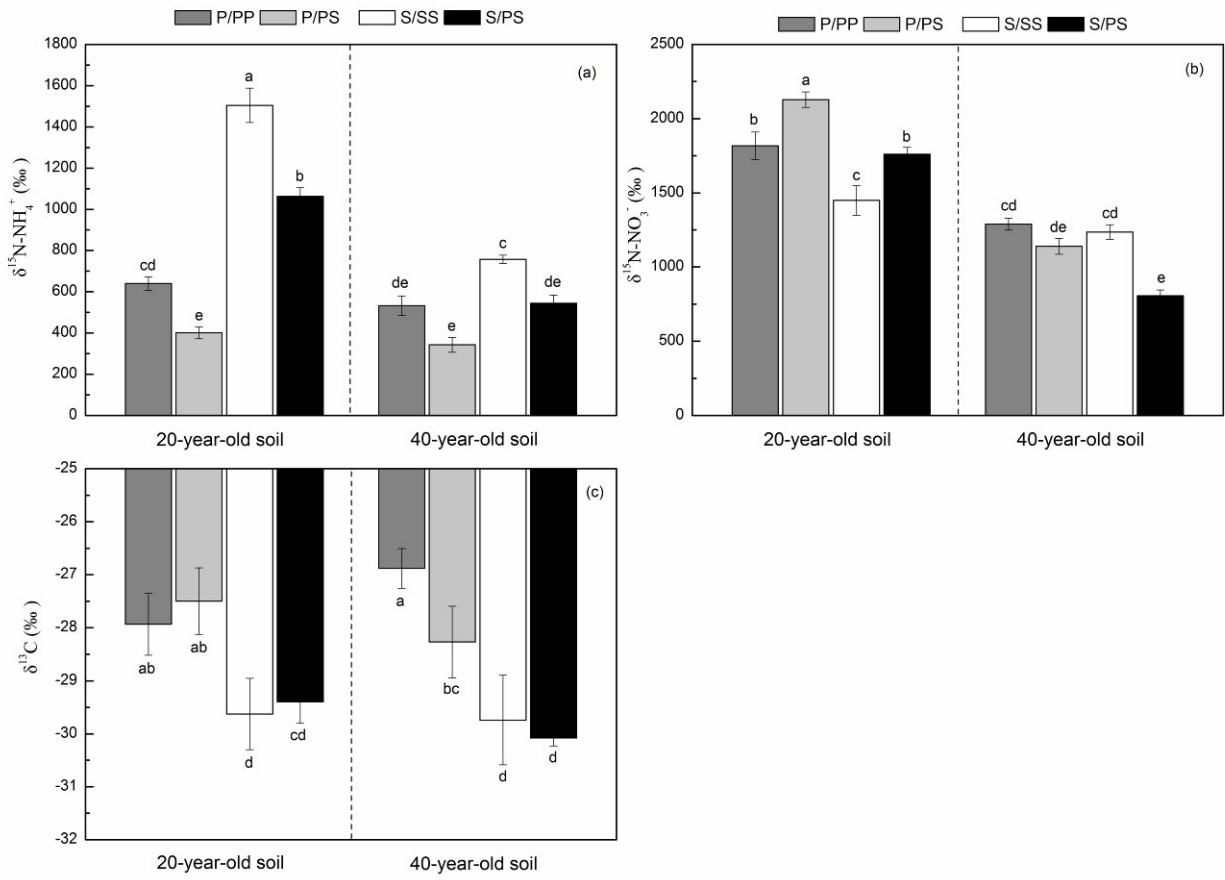


958

959

960

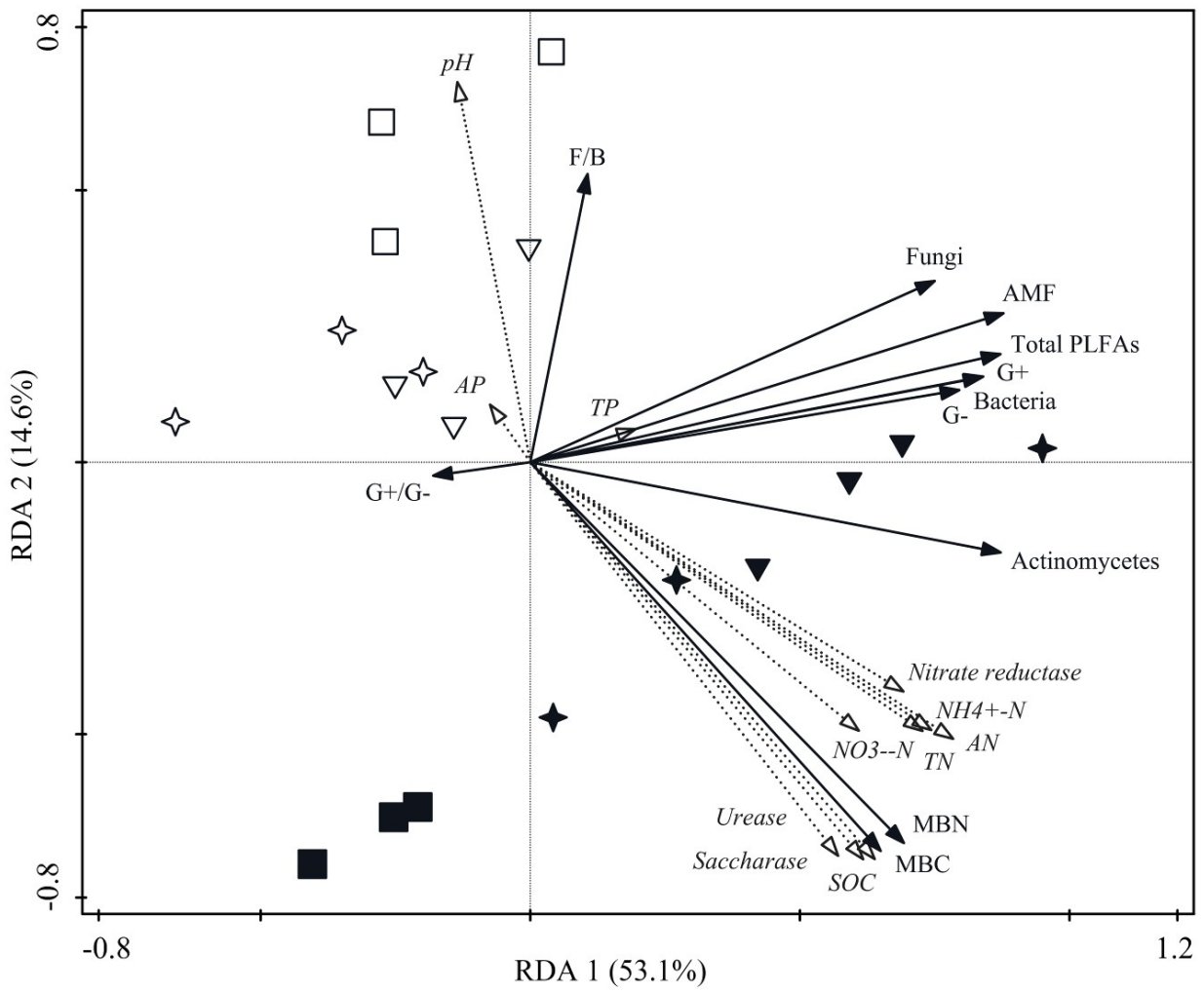
961 **Figure 6.**



962

963

964 **Figure 7.**



965

966

Figure 8.

