

**SUPPLEMENTAL MATERIAL FOR**

**Carbon and nitrogen recycling during cyanoHABs in dreissenid invaded and non-invaded US  
midwestern lakes and reservoirs**

Trinity L. Hamilton<sup>a,b,\*</sup>, Jessica R. Corman<sup>c</sup>, Jeff R. Havig<sup>d</sup>

Author 1: Hamilton T. L. (trinityh@umn.edu)

Author 2: Corman, J. R. (jcorman3@unl.edu)

Author 3: Havig J. R. (jhavig@umn.edu)

<sup>a</sup> Department of Plant and Microbial Biology, University of Minnesota, St. Paul, MN 55108, USA

<sup>b</sup> BioTechnology Institute, University of Minnesota, St. Paul, MN, USA 55108

<sup>c</sup> School of Natural Resources, University of Nebraska, Lincoln, NE 68583, USA

<sup>d</sup> Department of Earth Sciences, University of Minnesota, Minneapolis, MN 55455, USA

\*correspondence:

Trinity L Hamilton. 218 Cargill Building, Plant and Microbial Biology, University of Minnesota, St. Paul, USA, 55108. Phone: +16126256372, Email: trinityh@umn.edu.

**Running title:** Carbon and nitrogen cycling during cyanoHABS in midwestern lakes

**Keywords:** cyanobacteria, blooms, carbon stable isotopes, carbon, nitrogen, biogeochemical cycling, eutrophication, invasive mussel species, cyanoHABs

## **DISCUSSION**

### **TOTAL CARBON**

The overall C concentration at GLSM in the surface sediments is 3.4 mol C/kg for organic carbon and 1.8 mol C/kg for carbonate carbon, and at Buckeye Lake 3.5 mol C/kg for organic carbon and 1.7 mol C/kg for carbonate carbon. These give totals of 5.3 mol C/kg buried in the surface sediments at GLSM and 5.1 mol C/kg at Buckeye Lake, or about half that estimated for eutrophic lakes in Minnesota (Dean, 1999). These data suggest shallow eutrophic lakes or reservoirs impacted by blooms are efficient at removing organic C from the water column, and thus potentially efficient at sequestering C.

The overall surface sediment C concentration at Brookville Lake and Sandusky Bay were the lowest values measured in this study, at 4.13 mol C/kg and 4.37 mol C/kg, respectively. However, at Brookville Lake most of the C was present as carbonate (3.62 mol C/kg) while at the Sandusky Bay site most of the C was present as organic carbon (2.68 mol C/kg). The Sandusky Bay site was from a protected mud flat (with no invasive mussel shells or shell fragments present in the sediment sample) where anoxia near the water-sediment interface coupled to the predominantly silt and clay sediments may inhibit colonization by invasive mussels locally, while the greater bay and Lake Erie area (including the rocky substrates of the nearby jetty a few meters away) are more densely populated by mussels. The Brookville Lake site had sediments that ranged in size from sand to gravel, and included abundant invasive mussel shells and shell fragments, indicating an active invasive mussel colony present or very close to the sampling location. Regardless of this local heterogeneity of mussel population density, both sites exhibited the lowest C concentration values for the surface sediments, suggesting that cyanoHABs coupled to the presence of invasive mussel species may in fact reduce carbon deposition and thus C sequestration potential.

### **CONCEPTUAL MODEL**

Our data suggest the presence or absence of invasive mussel species is not correlative to the composition of cyanobacteria in cyanoHABs. However, we acknowledge that our single time point samples do not capture temporal or seasonal variability in bloom community structure. Regardless, we examined sites with and without invasive mussel species that experience cyanoHABs annually to begin to

constrain how the presence of invasive mussels may affect carbon sequestration in eutrophic systems. We constructed a conceptual model for the movement and sequestration of carbon in the different types of lakes and reservoirs sampled in this study (Fig. 9 in the main text).

Lakes and reservoirs that have experienced eutrophication that drives cyanoHABs such as Buckeye Lake and GLSM have potential for efficient sequestration of both  $C_{\text{org}}$  and  $C_{\text{carb}}$ . However, the  $C_{\text{org}}$  sediment concentrations at those two lakes (4.14 to 4.17 %) indicate an approximate three-fold reduction in  $C_{\text{org}}$  burial compared to an average of surface sediments at 20 lakes across Minnesota (Dean, 1999). Furthermore, the  $C_{\text{carb}}$  sediment concentrations (1.99 % at Buckeye Lake and 2.17 % at GLSM) indicate less carbonate burial compared to an average of 3.13 % C as  $\text{CaCO}_3$  (Dean, 1999). From summing the averages of Buckeye Lake and GLSM, the carbon burial in the surface sediments of cyanoHAB-impacted lakes and reservoirs is 6.2 % (5.19 mol C/ kg dry sediment), or about 54 % reduction in C burial compared to that of Minnesota lakes (average of 11.21 mol C/kg dry sediment; Dean, 1999), suggesting the Ohio lakes are less effective at burying C. Loss of  $C_{\text{org}}$  as  $\text{CH}_4$  due to breakdown of  $C_{\text{org}}$  by methanogens may also play an important role in impacting the carbon sequestration potential of Ohio lakes, which warrants further study.

Lakes and reservoirs with cyanoHABs which have also been colonized by invasive mussel species (such as Brookville Lake and Sandusky Bay, Lake Erie) appeared to exhibit a reduced capacity for carbon sequestration compared to the cyanoHAB-only sites (Fig. 9 in the main text). Reduced burial of organic carbon is likely linked directly to consumption and respiration/metabolic breakdown of organic matter from the water column by mussels. Burial of inorganic carbon may be influenced by several factors. First, reduced cyanobacterial biomass may be related to decreased rates of photosynthetically driven calcite production due to decreased nucleation sites (fewer cells in the water column). Second, increased uptake of calcium by mussels for shell formation may lead to less calcium available for calcite precipitation. And, finally, increased growth of mussels may lead to increased production of  $\text{CO}_2$  in the water column (due to mussel metabolism) generating carbonic acid which will react with calcite crystals in the water column to dissolve it. Local heterogeneity of mussel population density can lead to a range of values for  $C_{\text{org}}$  and  $C_{\text{carb}}$ , with higher  $C_{\text{org}}$  and lower  $C_{\text{carb}}$  with lower local mussel population density. But total C values were similar for both the lower local mussel density site (Sandusky Bay, 5.24 %) and

higher local mussel density site (Brookville Lake, 4.96 %), giving an average of 5.1 % C, or an average of 4.25 mol C/kg dry sediment. This represents a nearly 20 % reduction in C burial compared to cyanoHAB-only sites, or nearly a 62 % reduction in C burial compared to eutrophic lakes in Minnesota (Dean, 1999).

One caveat for Brookville Lake's carbon burial potential involves mussel shells as a means for  $C_{\text{carb}}$  sequestration. For this study, while we did not take into account carbon sequestered as carbonate precipitated as shells by mussels, invasive mussel species shells collected from Lake Erie ( $n = 27$ ) showed that a mature Zebra mussel shell (two produced per individual mussel) had an average long axis of 23.4 mm ( $\pm 3.35$  mm) and weighed an average of 0.378 g ( $\pm 0.149$  g), containing a total of 3.78 mmol C as  $\text{CaCO}_3$ . Using mussel shells to compensate for the surface sediments of the Brookville Lake site (total of 4.13 mol C/kg dry sediment) to match the average C burial of surface sediments in Minnesota lakes (11.21 mol C/kg dry sediment), each kilogram of dry surface sediment deposited at Brookville Lake site would also require the production of 1875 mussel shells. Assuming a loose packing volume of 1.28  $\text{cm}^3$  per mussel shell (based on the shells measured for this study), the volume of 2856 mussel shells would be  $\sim 2,400 \text{ cm}^3$ . Assuming a sediment density of  $\sim 1 \text{ g/cm}^3$  for Brookville Lake surface sediments, this would mean for every 1000  $\text{cm}^3$  of sediment, you would need nearly a 2.5 fold equivalent volume of mussel shell deposition. In other words, a 10 cm x 10 cm surface would have 10 cm deep sediment layer, and an additional 24 cm of mussel shells. This level of mussel shell accumulation could be further complicated by low concentrations of calcium in fresh water systems as well as remineralization of organic carbon producing  $\text{CO}_2$ , generating carbonic acid ( $\text{CO}_2 + \text{H}_2\text{O} \Rightarrow \text{H}_2\text{CO}_3$ ), and potentially driving carbonate dissolution ( $\text{H}_2\text{CO}_3 + \text{CaCO}_3 \Rightarrow \text{Ca}^{2+} + 2 \text{HCO}_3^-$ ) in anoxic sediments.

## REFERENCES

Dean W. E. and Gorham E. (1998) Magnitude and significance of carbon burial in lakes, reservoirs, and peatlands. *Geology* 26: 535–538.

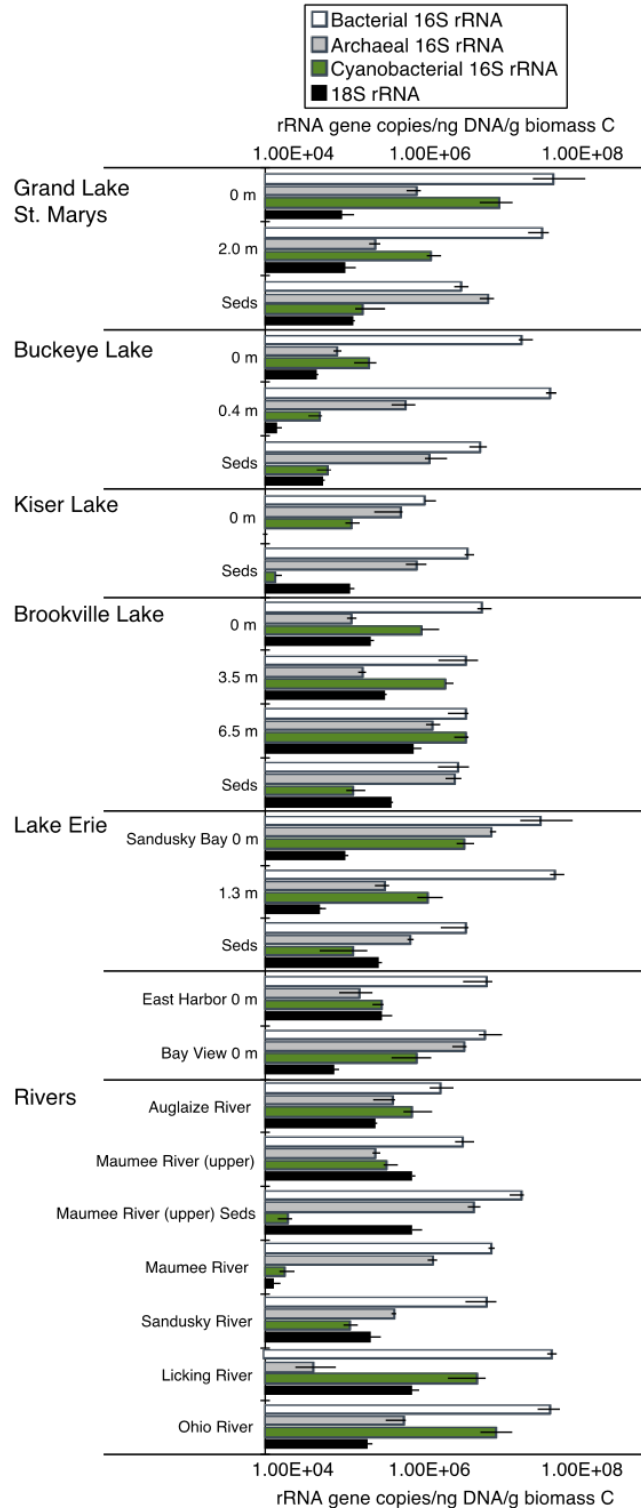
**Table S1.** Physical characteristics of lakes sampled in this study.

| Lake/Bay                | Surface area<br>(km <sup>2</sup> ) | Max depth<br>(m) | mean depth<br>(m) |
|-------------------------|------------------------------------|------------------|-------------------|
| Grand Lake St. Marys    | 58.3                               | 3.0              | 1.8               |
| Buckeye Lake            | 12.9                               | 4.3              | 1.8               |
| Kiser Lake              | 1.6                                | 3.7              | 1.9               |
| Brookville Lake         | 21.3                               | 36.0             | 4.9               |
| Sandusky Bay, Lake Erie | 77.0                               | 3.4              | 1.8               |

**Table S2.** Accession numbers and file names for the 16S and 18S rRNA amplicon libraries included in the present study.

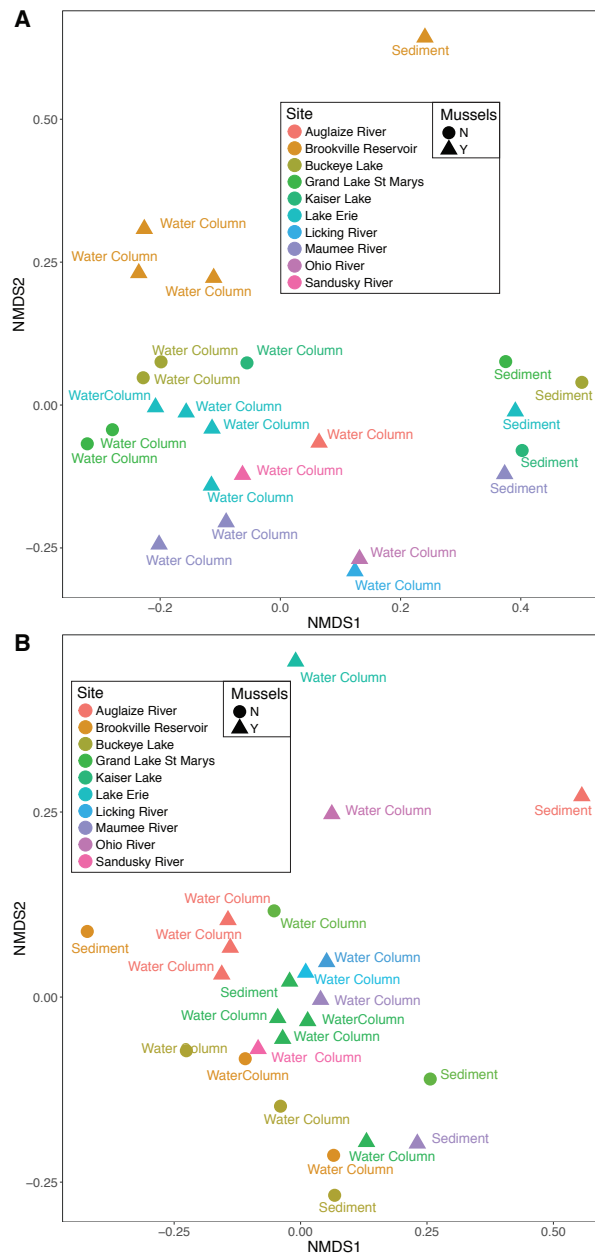
|   | BioSample Accession |              |
|---|---------------------|--------------|
|   | 16S rRNA            | 18S rRNA     |
| Kaiser Lake - 0m                        | SAMN07125357        | SAMN07125357 |
| Kaiser Lake sediments                   | SAMN07125358        | SAMN07125358 |
| GLSM - 0m                               | SAMN07125359        | SAMN07125359 |
| GLSM - 2m                               | SAMN07125360        | SAMN07125361 |
| GLSM sediments                          | SAMN07125361        | SAMN07125360 |
| Maumee River (upper) - 0m               | SAMN07125362        | SAMN07125362 |
| Maumee River sediments (upper)          | SAMN07125363        | SAMN07125363 |
| Auglaize River - 0m                     | SAMN07125364        | SAMN07125364 |
| Maumee River (lower)                    | SAMN07125365        | SAMN07125365 |
| Lake Erie - Bay View Boat Launch        | SAMN07125366        | SAMN07125366 |
| Lake Erie - Sandusky Bay - 0m           | SAMN07125367        | SAMN07125369 |
| Lake Erie - Sandusky Bay - 1.3m         | SAMN07125368        | SAMN07125367 |
| Lake Erie - Sandusky Bay sediments      | SAMN07125369        | SAMN07125368 |
| Lake Erie - East Harbor State Park - 0m | SAMN07125370        | SAMN07125370 |
| Sandusky River - 0m                     | SAMN07125371        | SAMN07125371 |
| Buckeye Lake - 0m                       | SAMN07125372        | SAMN07125372 |
| Buckeye Lake - 0.4m                     | SAMN07125373        | SAMN07125373 |
| Buckeye Lake Sediments                  | SAMN07125374        | SAMN07125374 |
| Brookville Reservoir - 0m               | SAMN07125380        | SAMN07125379 |
| Brookville Reservoir - 3.5m             | SAMN07125381        | SAMN07125380 |
| Brookville Reservoir - 6.5m             | SAMN07125382        | SAMN07125381 |
| Brookville Reservoir sediments          | SAMN07125383        | SAMN07125382 |
| Ohio River - 0m                         | SAMN07125384        | SAMN07125383 |
| Licking River - 0m                      | SAMN07125385        | SAMN07125384 |

**Figure S1.** qPCR of SSU rRNA genes from reservoirs, rivers, and lakes in Ohio, Kentucky and Indiana. Results are presented as the mean of triplicate qPCR assays and error bars represent the standard deviation of replicates. Copy number is normalized to per gram biomass C (Table 1). Sample collection depths are indicated when more than just the surface was collected. Sed, sediments collected at the water sediment interface.

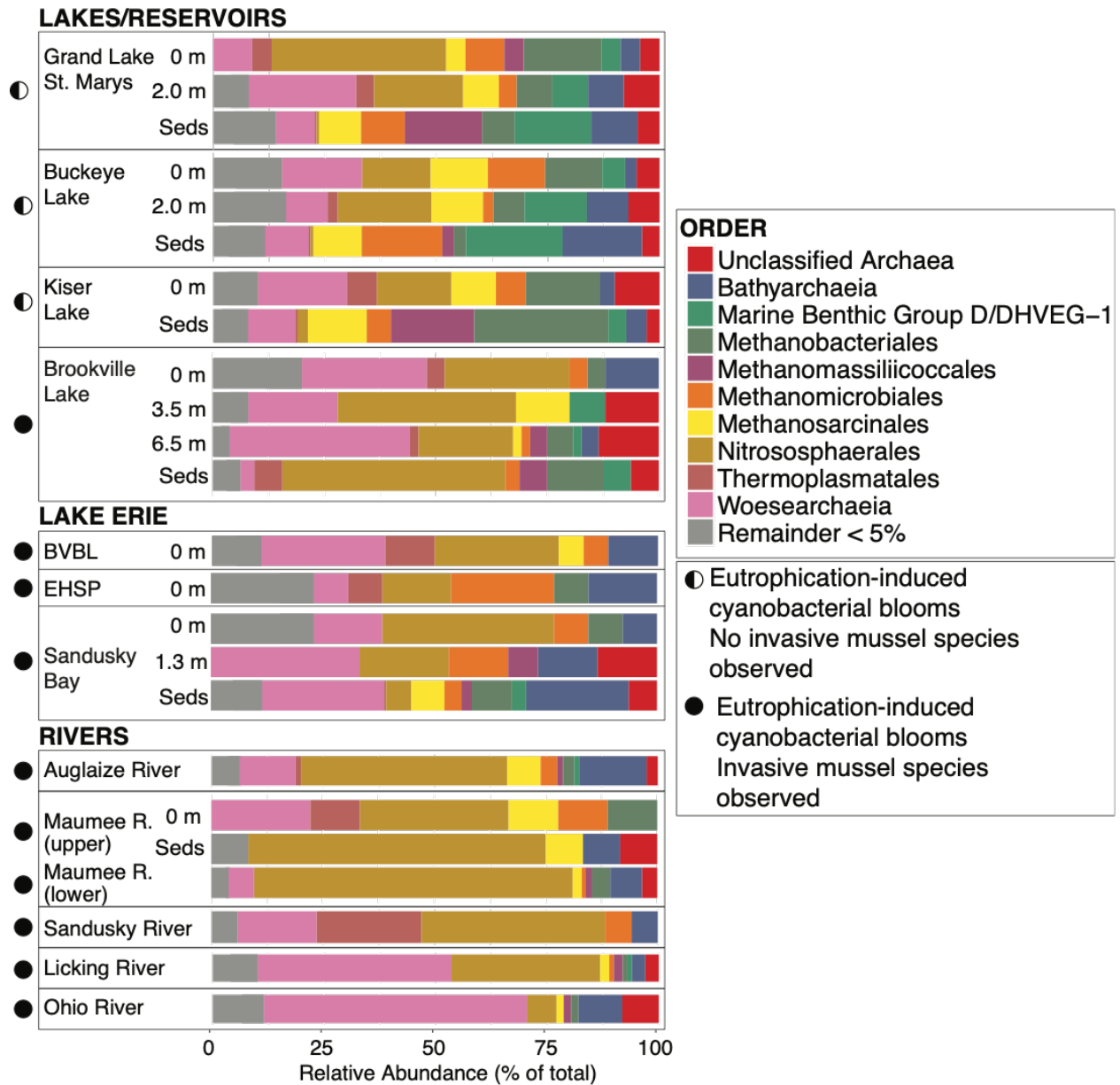




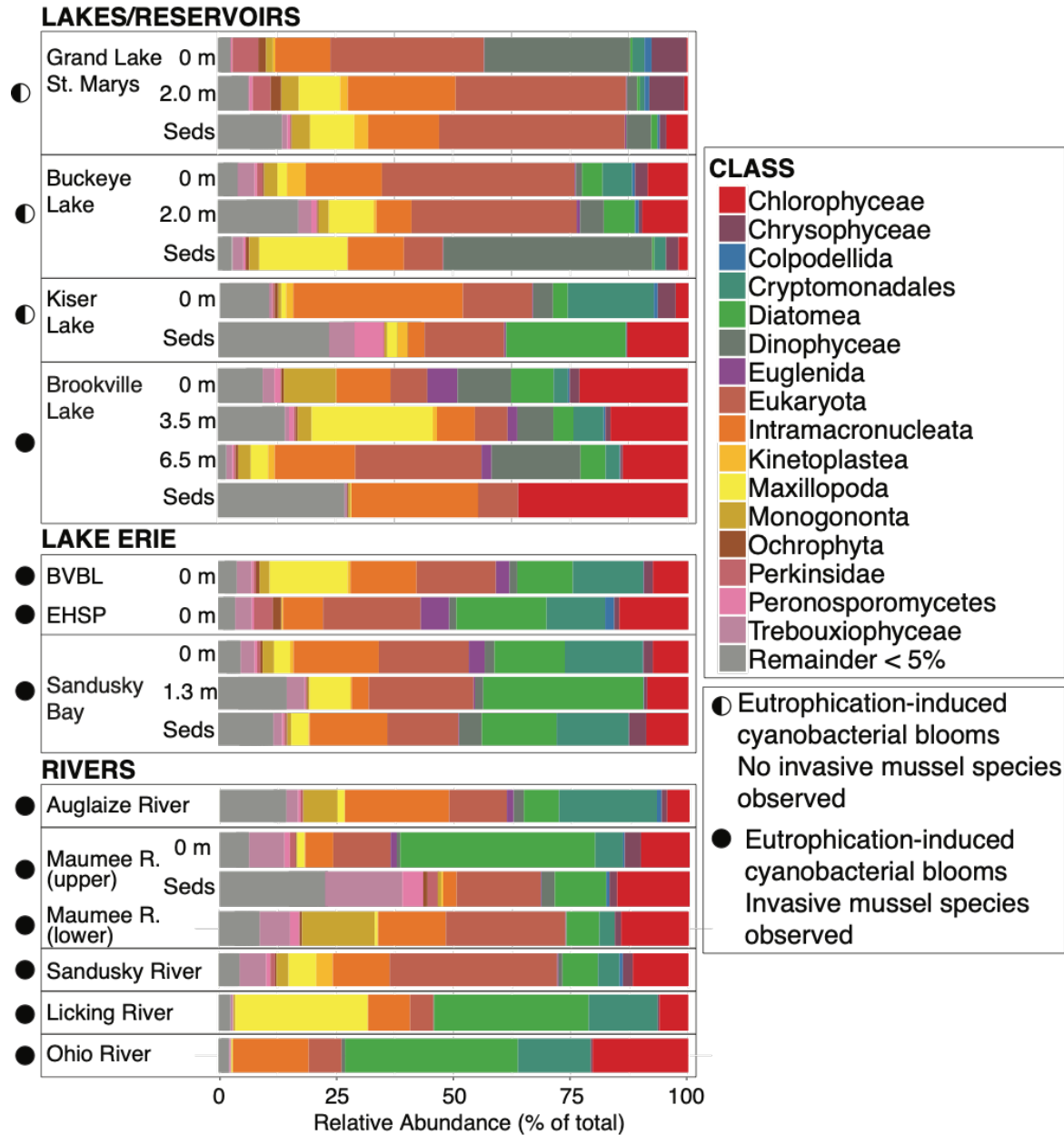
**Figure S2.** Nonmetric multidimensional scaling (NMDS) plot showing dissimilarities of 16S (A) and 18S (B) rRNA gene sequences among the samples with Bray–Curtis distances.



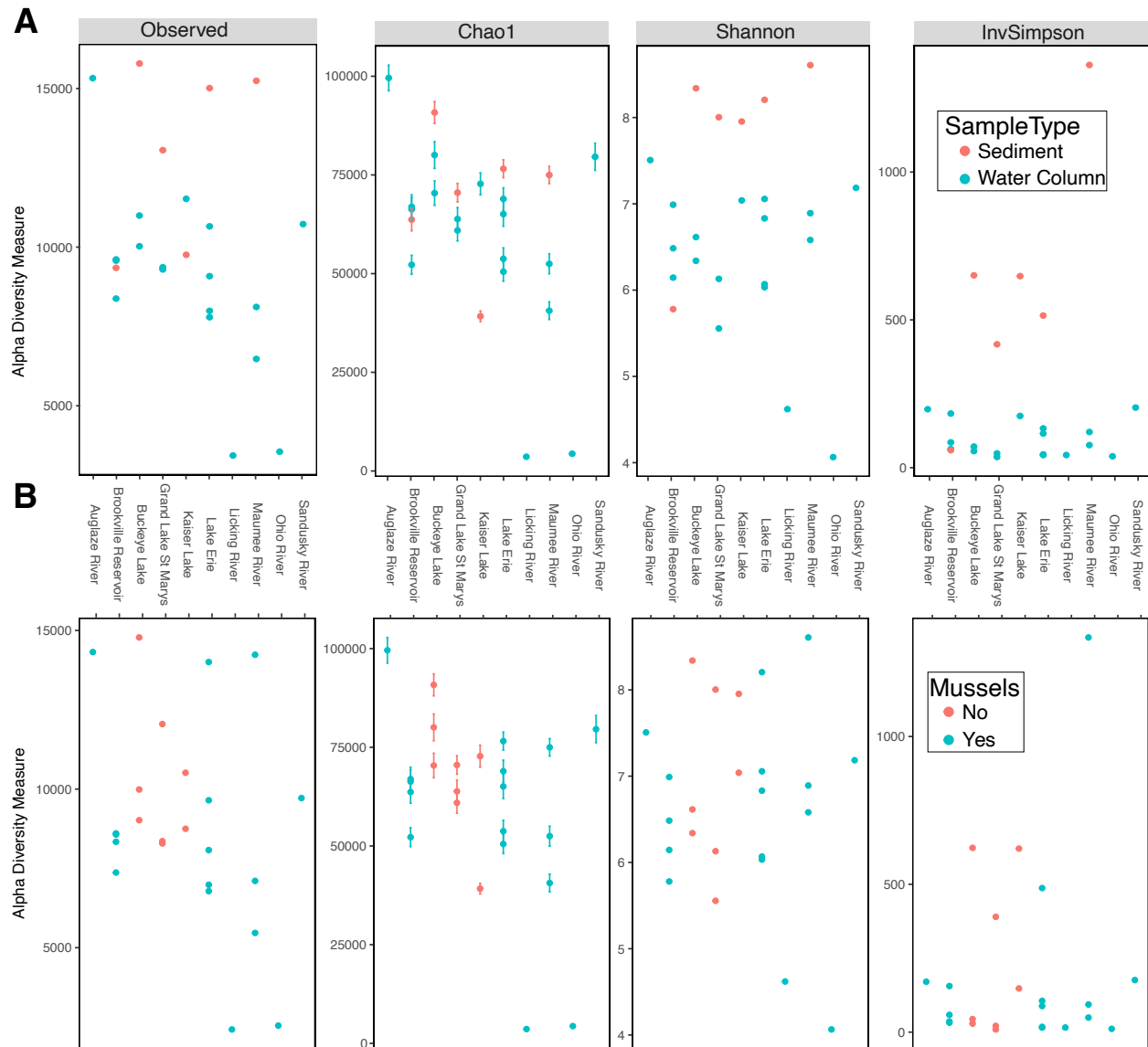
**Figure S3.** Composition of 16S rRNA gene sequences affiliated with Archaea recovered from sediment and planktonic biomass in reservoirs, rivers, and lakes in Ohio, Kentucky and Indiana. Representative OTUs for each library were binned at the Order level. Sample collection depths are indicated when more than just the surface was collected. Seds, sediments collected at the water sediment interface; BVBL, Bay View Boat Launch; EHSP, East Harbor State Park.



**Figure S4.** Composition of 18S rRNA gene sequences recovered from sediment and planktonic biomass in reservoirs, rivers, and lakes in Ohio, Kentucky and Indiana. Representative OTUs for each library were binned at the Class level. Sample collection depths are indicated when more than just the surface was collected. Seds, sediments collected at the water sediment interface; BVBL, Bay View Boat Launch; EHSP, East Harbor State Park.



**Figure S5.** Alpha diversity metrics of 16S rRNA gene sequences recovered from reservoirs, rivers, and lakes in Ohio, Kentucky and Indiana. Diversity metrics are plotted by by sample type (water column or sediment) in (A) and by the presence (+) or absence (-) of mussels in (B).



**Figure S6.** Alpha diversity metrics of 18S rRNA gene sequences recovered from reservoirs, rivers, and lakes in Ohio, Kentucky and Indiana. Diversity metrics are plotted by by sample type (water column or sediment) in (A) and by the presence (+) or absence (-) of mussels in (B).

