

Acid mine drainage sludge as an alternative raw material for M-type hexaferrite preparation

著者	Mei Liu, Atsushi Iizuka, Etsuro Shibata
journal or publication title	Journal of Cleaner Production
volume	224
page range	284-291
year	2019-07-01
URL	http://hdl.handle.net/10097/00132128

doi: 10.1016/j.jclepro.2019.03.224

1
2
3
4
5
6 Acid Mine Drainage Sludge as an Alternative
7
8
9
10 Raw Material for M-type Hexaferrite
11
12
13
14
15 Preparation
16
17
18
19

20 Mei Liu^{a,b,*}, Atsushi Iizuka^c, Etsuro Shibata^c
21
22

23
24 ^aGraduate School of Environmental Studies, Tohoku University, Sendai 980-0845,
25
26 Japan
27

28
29 ^bSchool of Material Science and Engineering, Jingdezhen Ceramic Institute, Jingdezhen
30
31 333403, China
32
33

34
35 ^cInstitute of Multidisciplinary Research for Advanced Materials (IMRAM), Tohoku
36
37 University, Sendai 980-8577, Japan
38
39
40

41
42
43
44
45 ***Corresponding author**
46

47 *M. Liu: liu.mei.q5@dc.tohoku.ac.jp; maggieliujci@gmail.com
48
49

50
51 A. Iizuka: atsushi.iizuka.e4@tohoku.ac.jp
52

53
54 E. Shibata: etsuro.shibata.e3@tohoku.ac.jp
55
56
57
58
59
60
61
62
63
64
65

1
2 **Abstract**
3
4
5

6 Two types of dewatered acid mine drainage sludge were sampled from abandoned
7 Japanese mines. The sludge shows great potential to be completely reused as alternative
8 iron and calcium sources to synthesize M-type calcium substituted barium hexaferrite
9 with addition of BaCO₃ by solid-state reaction. The precursors and as-prepared samples
10 were investigated by X-ray diffraction (XRD), field-emission scanning electron
11 microscopy, and thermogravimetry–mass spectroscopy. The effect of sludge addition on
12 phase formation and microstructure evolution during calcination was discussed in
13 details. For sludge content up to 93.9 mass % with addition of 6.1 mass % BaCO₃, M-
14 type hexaferrite can be identified in the XRD patterns of as-prepared samples calcined
15 at 1100, 1200, and 1300 °C. The impurities in the sludge, such as silicon, and
16 aluminum, result in formation of secondary aluminosilicates phases; another impurity
17 sulfur in the sludge may increase M-type hexaferrite formation temperature owing to
18 barium or calcium sulfate formation in the precursor. The impurities also contribute to
19 the difference in the microstructure for lower melting point. Possible applications of the
20 sludge-derived hexaferrite products such as magnetic material or microwave absorber
21 are suggested.
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44

45 **Keywords:** Waste valorization; Sludge; Acid mine drainage; M-type hexaferrite; Solid-
46 state reaction
47
48
49

50
51 **1. Introduction**
52

53 Acid mine drainage is one of the major pollutants of the mining industry. This
54 generation of acidity can be continuous even when a mine has been abandoned for
55 hundreds of years. Lime neutralization and its enhanced processes are the most widely
56
57
58
59
60
61
62
63
64
65

1
2 used methods to treat acid mine drainage (Zinck and Griffith, 2013). As a result, large
3
4 amounts of metal-rich hydroxide sludge are produced during the treatment process
5
6 (Macías et al., 2017). In addition to land occupation, the high disposal cost, and risk of
7
8 toxic metal migration from sludge to land associated with sludge disposal at mine sites
9
10 necessitate novel techniques for sludge recycling (Rakotonimaro et al., 2017).
11
12

13
14 Since it was discovered in the 1950s, M-type barium hexaferrite $\text{BaFe}_{12}\text{O}_{19}$ and its
15
16 substituted or doped materials have been widely used in magnets, microwave devices,
17
18 magnetic recording media, microwave absorbers, gas sensors, adsorbent, and so forth
19
20 (Went, et al., 1952; Sixtus, et al., 1956; Harris, et al., 2009; Sözeri, et al., 2014;
21
22 Karmakar, et al., 2014; Patel, et al., 2012). Annual global production of M-type barium
23
24 hexaferrite exceeds 300,000 tons, which constitutes half of the total production of
25
26 magnetic materials (Pullar, 2012). $\text{BaFe}_{12}\text{O}_{19}$ exists in a hexagonal magnetoplumbite
27
28 structure, which affords it excellent magnetic properties and chemical stability
29
30 (Goldman, 2006). Other sufficiently large divalent metal ions, such as strontium and
31
32 lead, can substitute barium ions without too much distortion of the structure (Kojima,
33
34 1982). Calcium belongs to the same group as barium and strontium, and it is cheaper
35
36 and more abundant. Unfortunately, calcium ions are too small to maintain the
37
38 magnetoplumbite structure of undoped $\text{CaFe}_{12}\text{O}_{19}$. Doping with a small amount of
39
40 lanthanum ions is an effective method to improve the structural stability of $\text{CaFe}_{12}\text{O}_{19}$
41
42 (Ichinose and Kurihara, 1963; Lotgering and Huyberts, 1980). In addition, calcium-
43
44 substituted M-type barium and strontium hexaferrite have been successfully prepared
45
46 (Blanco and Gonzalez, 1991; Iqbal, et al., 2010; Sanghi and Agarwal, 2012; Anbarasu,
47
48 et al., 2013).
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

1
2 As summarized in the authors' review (Liu, et al., 2018), previous studies on waste-
3
4 derived ferrite have mainly focused on using metal-rich wastes to fabricate spinel
5
6 ferrites. Whereas studies of waste-derived hexaferrite, which has a wider range of
7
8 applications and greater market share than spinel ferrite, are relatively rare. Pullar and
9
10 co-workers were the first to use iron-rich sludge produced by a steel/iron-based
11
12 wiredrawing process to synthesize undoped $\text{SrFe}_{12}\text{O}_{19}$ and Co/Mn-doped
13
14 $\text{SrCo}_{0.5}\text{Mn}_{0.5}\text{Fe}_{11}\text{O}_{19}$ and $\text{SrCoMnFe}_{10}\text{O}_{19}$ with addition of SrCO_3 , Co_3O_4 , and
15
16 $\text{MnCl}_2 \cdot 4\text{H}_2\text{O}$ (Pullar, et al., 2013). They subsequently used the same sludge to prepare
17
18 undoped $\text{BaFe}_{12}\text{O}_{19}$ with addition of BaCO_3 (Pullar, et al., 2016). In both studies, the
19
20 sludge was recycled and used only as the iron source in hexaferrite processing. And the
21
22 reactions and mechanism of ferrite derived from wastes have not been sufficiently
23
24 investigated.

25
26
27
28
29
30
31 In this study, acid mine drainage sludge was recycled as both alternative iron and
32
33 calcium sources to synthesize M-type hexaferrite powders by solid-state reaction with
34
35 addition of BaCO_3 . Mixtures with initial compositions of $\text{Ca}_x\text{Ba}_{1-x}\text{Fe}_{12}\text{O}_{19}$ ($x = 0, 0.3,$
36
37 $0.5, 0.7,$ and 1.0) were prepared from different combinations of two types of acid mine
38
39 drainage sludge (A and B), $\alpha\text{-Fe}_2\text{O}_3$, and BaCO_3 . The effect of sludge addition on phase
40
41 formation and microstructure evolution during calcination was investigated, with special
42
43 focus on the chemical reactions and mechanism of hexaferrite preparation from sludge.
44
45 The crystalline M-type hexaferrite phase can be identified by XRD in as-prepared
46
47 samples with addition of up to 93.9% (mass%) acid mine drainage sludge. Value-added
48
49 products derived from the sludge and reduction of sludge accumulation in deposits at
50
51 mine sites can therefore be expected.
52
53
54
55
56
57
58
59
60
61
62
63
64
65

2. Materials and methods

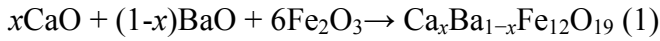
2.1. Materials

Two types of dewatered acid mine drainage sludge (A and B) were sampled from different sludge deposits of acid mine drainage treatment plants in Japan. Sludge A has higher calcium content than sludge B and was generated during a prevalent calcium carbonate and lime two-step neutralization process, whereas sludge B was generated during a less common bacterial oxidation and calcium carbonate neutralization process. The two types of sludge were dried in an oven at 105 °C for 24 h and their main elemental compositions (see Figure 1.) were determined by inductively coupled plasma atomic emission spectroscopy (ICP-AES) (Spectro Arcos, Ametek, Berwyn, PA, USA) after acid digestion by a HCl–HNO₃–HF–H₃BO₃ acid system. Figure S1 and S2 show the XRD patterns, micrographs and TG-MS curves of Sludge A and Sludge B, respectively. Both sludge flocs agglomerated to a few microns in length. Sludge A was characterized as amorphous iron hydroxide with excess neutralizer Mg_{0.064}Ca_{0.936}CO₃, whereas Sludge B mainly consisted of goethite (α -FeOOH) and ferric sulfate. α -Fe₂O₃ and BaCO₃ (99.9% purity, Wako Pure Chemical Industries, Ltd., Osaka, Japan) were used to adjust the initial molar ratio of Ca:Ba:Fe for the Ca_xBa_{1-x}Fe₁₂O₁₉ mixtures used in M-type hexaferrite preparation.

Figure 1. Main elemental compositions of the acid mine drainage sludge determined by ICP-AES (mass %): (a) sludge A and (b) sludge B.

2.2. Preparation of M-type hexaferrite

In previous studies, the impurities of wastes as alternative raw materials in ferrite production were purified by several processes (Liu, et al., 2018). However, these purification processes inevitably increase costs and consequently diminish the practicality of using wastes to prepare ferrite compounds. Considering the economic feasibility, the authors decided to completely reuse the acid mine drainage sludge without a preliminary purification process. The synthesis reaction of $\text{Ca}_x\text{Ba}_{1-x}\text{Fe}_{12}\text{O}_{19}$ hexaferrite is represented by reaction (1).



Mixtures of the sludge and pure chemicals (see Table 1) with Ca:Ba:Fe = $x:(1-x):12$ (molar ratio) were wet ball-milled (Pulverisette 6, Fritsch GmbH, Idar-Oberstein, Germany) for 30 min, followed by drying at 105 °C for 24 h. The dried mixtures were then crushed and homogenized by mortar grinding to obtain the precursors, which were subsequently calcined in an alumina crucible at the target temperature with a holding time of 4 h using a box furnace without atmospheric control (KBF524N1, Koyo Thermo System Co., Ltd., Nara, Japan). Finally, the calcined samples were air-cooled in the furnace and ground into a powder in a mortar for analysis. The sludge-free mixture A_0F composed of $\alpha\text{-Fe}_2\text{O}_3$ and BaCO_3 is used as a reference.

Table 1. Ingredient compositions of the $\text{Ca}_x\text{Ba}_{1-x}\text{Fe}_{12}\text{O}_{19}$ (mass %) mixtures

Mixture	x	Sludge A	Sludge B	$\alpha\text{-Fe}_2\text{O}_3$	BaCO_3
A_0F	0.0	0	0	82.9	17.1
A_1F	0.3	19.1	0	69.4	11.5
A_2F	0.5	31.1	0	61.0	7.9

A ₃ F	0.7	42.3	0	53.0	4.7
A ₄ F	1.0	57.8	0	42.1	0
A ₀ B	0.04	0	88.0	0	12.0
A ₁ B	0.3	12.3	79.2	0	8.5
A ₂ B	0.5	22.0	71.9	0	6.1
A ₃ B	0.7	31.7	64.6	0	3.7
A ₄ B	1.0	46.5	53.5	0	0

2.3. Characterization

2.3.1. Phase identification

The X-ray diffraction (XRD) patterns of the precursors and calcined samples were recorded by a Bruker D2 Phaser diffractometer with Cu-K α radiation at 30 kV and 10 mA using a LYNXEYE detector (Bruker AXS GmbH, Karlsruhe, Germany). The 2 θ scan range was 10°–80° with a step size of 0.02° and a scan speed of 0.5 s per step. Phase identification was performed by matching the XRD patterns of the samples to standard patterns retrieved from the database of the International Centre for Diffraction Data (ICDD, PDF 2010).

2.3.2. Morphological study

The morphologies of the calcined samples were determined by field-emission scanning electron microscopy (FE-SEM, SU6600, Hitachi High-Technologies Corporation, Tokyo, Japan).

2.3.3. Thermal analysis

The thermogravimetric and differential thermogravimetric (TG–DTG) curves of the precursors were recorded from room temperature to 1300 °C at a heating rate of 10 °C/min under a simulated air atmosphere (20% O₂–80% He) using a STA449F1 Jupiter simultaneous thermal analyzer (NETZSCH-Gerätebau GmbH, Wittelsbacherstraße, Germany) by NETZSCH Japan K.K, coupled with a JMS-Q1500GC mass spectrometer (JEOL Ltd., Tokyo, Japan) to analyze the evolved gas.

3. Results and discussion

3.1. Effect of sludge addition on phase formation during calcination

The crystalline phases in the precursors and samples prepared from sludge free A₀F and sludge added mixtures (see Table 1.) are summarized in Table 2. The XRD patterns are shown in Figures S3–S12. The results for the precursors indicate that the iron species in sludge A are quite different from those in sludge B. Crystalline iron oxyhydroxide (goethite, α -FeOOH, PDF 01-076-7156) can be identified only in the precursors containing sludge B. The iron in sludge A exists in an amorphous form, whereas residual crystalline CaCO₃ neutralizer is contained in the precursors not containing sludge B. Calcium or barium carbonate in the ingredients transforms to the sulfate with sludge B addition. High sulfur content in sludge B is believed to be responsible for this phenomenon.

Table 2. Changes in the crystalline phases observed during calcination.

Mixture	Crystalline phases observed					
	Precursor	900 °C/4 h	1000 °C/4 h	1100 °C/4 h	1200 °C/4 h	1300 °C/4 h

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

A ₀ F	α -Fe ₂ O ₃ ; BaCO ₃	M-type hexaferrite; α -Fe ₂ O ₃	M-type hexaferrite; α -Fe ₂ O ₃	M-type hexaferrite; α -Fe ₂ O ₃	M-type hexaferrite; α -Fe ₂ O ₃	M-type hexaferrite; α -Fe ₂ O ₃
A ₁ F	α -Fe ₂ O ₃ ; BaCO ₃	M-type hexaferrite; α -Fe ₂ O ₃ ; hexacelsian	M-type hexaferrite; α -Fe ₂ O ₃ ; hexacelsian	M-type hexaferrite; α -Fe ₂ O ₃	M-type hexaferrite; α -Fe ₂ O ₃	M-type hexaferrite; α -Fe ₂ O ₃ ; hexacelsian
A ₂ F	α -Fe ₂ O ₃ ; BaCO ₃ ; CaCO ₃	α -Fe ₂ O ₃ ; M-type hexaferrite; hexacelsian	α -Fe ₂ O ₃ ; M-type hexaferrite; hexacelsian	α -Fe ₂ O ₃ ; M-type hexaferrite;	α -Fe ₂ O ₃ ; M-type hexaferrite;	α -Fe ₂ O ₃ ; M-type hexaferrite;
A ₃ F	α -Fe ₂ O ₃ ; CaCO ₃	α -Fe ₂ O ₃ ; hexacelsian	α -Fe ₂ O ₃ ; hexacelsian	α -Fe ₂ O ₃ ; M-type hexaferrite; hedenbergite aluminian	α -Fe ₂ O ₃ ; M-type hexaferrite; hedenbergite aluminian; hexacelsian	α -Fe ₂ O ₃ ; M-type hexaferrite; hedenbergite aluminian
A ₄ F	α -Fe ₂ O ₃ ; CaCO ₃	α -Fe ₂ O ₃	α -Fe ₂ O ₃ ; hedenbergite aluminian	α -Fe ₂ O ₃ ; hedenbergite aluminian	α -Fe ₂ O ₃ ; hedenbergite aluminian	α -Fe ₂ O ₃
A ₀ B	α -FeOOH; BaSO ₄ ; BaCO ₃	α -Fe ₂ O ₃ ; BaSO ₄ ; hexacelsian	α -Fe ₂ O ₃ ; BaSO ₄ ; hexacelsian	M-type hexaferrite; α -Fe ₂ O ₃ ; hexacelsian	M-type hexaferrite; α -Fe ₂ O ₃ ; hexacelsian	M-type hexaferrite; α -Fe ₂ O ₃ ; hexacelsian
A ₁ B	α -FeOOH; BaSO ₄ ;	α -Fe ₂ O ₃ ; BaSO ₄ ;	α -Fe ₂ O ₃ ; BaSO ₄ ;	α -Fe ₂ O ₃ ; M-type	α -Fe ₂ O ₃ ; M-type	α -Fe ₂ O ₃ ; M-type

	BaCO ₃	hexacelsian	hexacelsian	hexaferrite; celsian	hexaferrite; celsian	hexaferrite; celsian
A ₂ B	α -FeOOH; BaSO ₄ ; gypsum	α -Fe ₂ O ₃ ; BaSO ₄ ; hexacelsian; celsian	α -Fe ₂ O ₃ ; BaSO ₄ ; hexacelsian; celsian	α -Fe ₂ O ₃ ; M-type hexaferrite; celsian	α -Fe ₂ O ₃ ; M-type hexaferrite; celsian	α -Fe ₂ O ₃ ; M-type hexaferrite; celsian
A ₃ B	α -FeOOH; gypsum; BaSO ₄	α -Fe ₂ O ₃ ; hexacelsian; celsian; BaSO ₄	α -Fe ₂ O ₃ ; hexacelsian; celsian;	α -Fe ₂ O ₃ ; celsian;	α -Fe ₂ O ₃ ; celsian;	α -Fe ₂ O ₃ ; celsian; hexacelsian
A ₄ B	α -FeOOH; gypsum	α -Fe ₂ O ₃ ; anorthite	α -Fe ₂ O ₃ ; anorthite	α -Fe ₂ O ₃ ; anorthite	α -Fe ₂ O ₃ ; anorthite	α -Fe ₂ O ₃ ; anorthite

It is worth noting that calcium-substituted and unsubstituted BaFe₁₂O₁₉ are difficult to distinguish by XRD because of their similar crystal structures (Anbarasu, et al., 2013). Therefore, the phase is labeled “M-type hexaferrite” rather than BaFe₁₂O₁₉ (PDF 00-043-0002). In the sludge-free A₀F series, the characteristic reflections of the (110), (107), (114), and (203) planes of M-type hexaferrite are observed in the XRD patterns of the samples calcined from 900 to 1300 °C. However, weak peaks corresponding to α -Fe₂O₃ (PDF 01-079-0007) are also present in the above temperature range.

Using sludge A as an alternative to α -Fe₂O₃ and BaCO₃ (series A₁F–A₄F), with increasing addition of sludge, there is an increase in the formation temperature and decreases in the intensities of the main peaks of M-type hexaferrite. In addition to the α -Fe₂O₃ phase, impurity phases, such as hexacelsian (BaAl₂Si₂O₈, PDF 01-072-7502) and hedenbergite aluminian (Ca(Fe_{0.821}Al_{0.179})(SiAl_{0.822}Fe_{0.178}O₆), PDF 01-078-1546), are also present in the calcined samples. Because of the structural instability of undoped

1
2 CaFe₁₂O₁₉, it is not surprising that no M-type hexaferrite can be identified in the A₄F
3
4 series without BaCO₃ addition over the whole temperature range.
5
6

7 Using both sludge A and sludge B as alternative materials (series A₀B–A₄B), the
8
9 XRD patterns of the calcined samples are more complex compared with those of series
10
11 A₀F–A₄F. The intensities of the main peaks of M-type hexaferrite for the calcined
12
13 samples are weaker than those for the A₀F–A₄F samples under the same conditions. For
14
15 low BaCO₃ addition (3.7 mass %) in the A₃B series and no BaCO₃ addition in the A₄B
16
17 series, M-type hexaferrite cannot be identified. Another difference is that the peaks
18
19 corresponding to M-type hexaferrite start to appear above 1100 °C, which is
20
21 accompanied by the peaks corresponding to BaSO₄ disappearing. This suggests that M-
22
23 type hexaferrite forms after decomposition of BaSO₄. Some of the hexacelsian impurity
24
25 phase transforms to another barium aluminosilicate crystalline phase (celsian,
26
27 BaAl₂Si₂O₈, PDF 00-038-1450). The results indicate that higher calcium content or
28
29 higher temperature promotes this phase transformation. Hedenbergite aluminian is not
30
31 observed for these series, and anorthite (CaAl₂Si₂O₈, PDF 01-075-1587) is present in
32
33 the A₄B series.
34
35

36 For M-type hexaferrite production, it is clear that the iron source α-Fe₂O₃ can be
37
38 completely replaced by acid mine drainage sludge. The iron in the sludge, regardless of
39
40 whether it is in the form of crystalline iron oxyhydroxide, poorly crystalline ferrihydrite,
41
42 or ferric sulfate, first transforms to the intermediate α-Fe₂O₃ phase and then participates
43
44 in the reactions of hexaferrite production. In contrast, the other raw material BaCO₃ can
45
46 only be partially replaced by the calcium in the sludge to form calcium-substituted M-
47
48 type barium hexaferrite.
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

1
2 The presence of aluminum and silicon impurities in the acid mine drainage sludge
3
4 leads to formation of barium or calcium aluminosilicates in the calcined samples. Sulfur
5
6 (another impurity in the sludge) may increase the temperature when M-type hexaferrite
7
8 starts to form owing to barium or calcium sulfate formation in the precursor.
9
10

11 12 13 14 3.2. Effect of sludge addition on microstructure evolution 15

16 Using acid mine drainage sludge as an alternative raw material in M-type hexaferrite
17
18 preparation also affects the microstructure of the calcined sample. Figure 2. shows a
19
20 comparison of the FE-SEM micrographs of samples with and without sludge addition
21
22 calcined at 900–1300 °C for 4 h. The sludge-free A₀F series exhibit plate-like shaped
23
24 grains of M-type hexaferrite with random orientations. The grains grow larger and
25
26 thicker with increasing calcination temperature. In the case of the A₀B series, bulk α-
27
28 Fe₂O₃ is covered by irregularly shaped particles at 900 and 1000 °C. The sintering
29
30 phenomenon is observed above 1200 °C, and the particles grow as spiral hillocks with
31
32 small steps. The impurities in the sludge are believed to lower the melting point, which
33
34 results in a different crystal growth model from the sludge-free samples (Elwell and
35
36 Neate, 1971; Vinnik, et al., 2017).
37
38
39
40
41
42
43
44
45

46 **Figure 2.** Comparison of the micrographs of A₀F and A₀B calcined at 900–1300 °C for
47
48 4 h.
49
50

51 52 53 54 3.3. Proposed reactions and mechanism 55

56 Figure 3. shows the TG–DTG curves of precursors A₀F, A₀B, A₂B, and A₄B,
57
58 respectively. The $m/z = 18$ (H₂O⁺), 44 (CO₂⁺), and 64 (SO₂⁺) ion current curves of the
59
60
61
62
63
64
65

1
2 above four precursors are shown in Figure 4. Significant differences are observed in the
3
4 decomposition processes of the precursors with different addition of acid mine drainage
5
6
7 sludge.
8
9

10
11 **Figure 3.** TG–DTG curves of the representative precursors: (a) A₀F, (b) A₀B, (c) A₂B,
12
13 and (d) A₄B.
14
15

16
17
18 **Figure 4.** Effect of sludge addition on gas evolution during heating under a simulated
19
20 air atmosphere (20% O₂ and 80% He): (a) $m/z = 18$ (H₂O⁺), (b) $m/z = 44$ (CO₂⁺), and (c)
21
22 $m/z = 64$ (SO₂⁺).
23
24
25

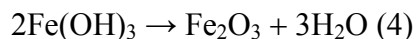
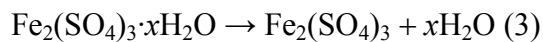
26
27 As shown in Table 1, the A₀F precursor is a mixture of the pure chemicals α -Fe₂O₃
28
29 and BaCO₃, and it exhibits two DTG peaks at 294 and 757 °C (Figure 3(a)). From the
30
31 MS curves, the former peak corresponds to mass loss of water, whereas the latter peak
32
33 corresponds to decomposition of BaCO₃ (eq 2).
34
35



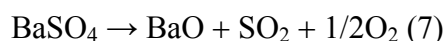
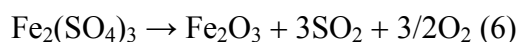
40
41 Another contribution to CO₂ evolution in the range 200–400 °C is decomposition of
42
43 carbonaceous matter. The peak at 1079 °C in the $m/z = 64$ ion current curve of the A₀F
44
45 precursor can be attributed to release of trace SO₂. The possible sources of
46
47 carbonaceous matter and trace sulfate are water added during the wet ball milling
48
49 process and trace impurities in the pure chemicals.
50
51

52
53 With replacement of α -Fe₂O₃ by sludge B (A₀B precursor), three DTG peaks and
54
55 larger mass loss are observed (Figure 3(b)). Both dehydration and dehydroxylation
56
57 reactions are considered to be responsible for the mass loss of water of the A₀B
58
59
60
61

1
2 precursor (Alcolea, et al., 2010). The probable reactions are represented by reaction (3),
3
4 (4) and (5).

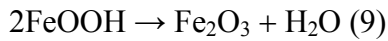
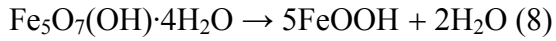


14
15 Compared with the A₀F precursor, there is a marked decrease in mass loss
16
17 corresponding to carbonate decomposition and an increase in mass loss ascribed to
18
19 carbonaceous matter decomposition for the A₀B precursor. The latter is believed to be a
20
21 result of flocculants added in the acid mine drainage treatment process. Combined with
22
23 a clear increase in SO₂⁺ evolution, the former can be attributed to BaCO₃ transforming
24
25 to BaSO₄ during the mixing process with sludge B addition. The peak at 620 °C in the
26
27 DTG curve of the A₀B precursor is believed to be transformation of Fe₂(SO₄)₃ to Fe₂O₃,
28
29 and the peak at 1136 °C corresponds to decomposition of BaSO₄. The desulfation
30
31 reactions to be considered are represented by reaction (6) and (7).
32
33
34
35
36

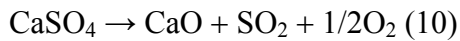


41
42 For the A₂B and A₄B precursors (Figure 3(c) and (d)), the total mass loss is almost
43
44 the same as that of the A₀B precursor. However, the mass loss steps and gas evolution
45
46 are different owing to the different proportions of sludge A, sludge B, and BaCO₃. As
47
48 previously mentioned, sludge A and sludge B are generated by different acid mine
49
50 drainage treatment processes, which results in the two types of sludge having different
51
52 compositions. A possible reason for the discrepancy in water evolution at around 110
53
54 °C between A₄B and the other three precursors might be the abundance of poorly
55
56
57
58
59
60
61
62
63
64
65

1
2 crystalline ferrihydrite-like $\text{Fe}_5\text{O}_7(\text{OH})\cdot 4\text{H}_2\text{O}$ in the A_4B precursor (Pulišová, et al.,
3
4 2014). The DTG peak at 109 °C may correspond to the loss of water during the
5
6 transformation of ferrihydrite to goethite (eq 8), and the peak at 244 °C corresponds to
7
8 dehydroxylation of goethite (eq 9):
9
10



11
12
13
14
15
16
17 The peaks corresponding to carbonate decomposition are completely absent in the
18
19 curves of the A_2B and A_4B precursors. For SO_2 evolution, the peaks at around 1000 °C
20
21 are considered to be because of CaSO_4 decomposition (eq 10):
22
23



24
25
26
27 The variation in SO_2^+ evolution corresponding to CaSO_4 and BaSO_4 decomposition
28
29 among the A_0B , A_2B , and A_4B precursors depends on the amounts of sludge A, sludge
30
31 B, and BaCO_3 .
32
33

34
35 A combined XRD and TG–MS study shows that the chemical composition and
36
37 mineral composition of acid mine drainage sludge have a considerable effect on the
38
39 reaction pathways and final calcination products. The proposed reaction pathways of (a)
40
41 the sludge-free sample (A_0F), (b) only sludge A as an alternative raw material (A_1F –
42
43 A_4F), and (c) both sludge A and sludge B as alternative raw materials (A_0B – A_4B) are
44
45 shown in Figure 5. The $\alpha\text{-Fe}_2\text{O}_3$ intermediate forms by dehydroxylation or desulfation
46
47 reactions in sludge addition pathways (b) and (c). In particular, for addition of sludge B
48
49 with relatively high sulfur content, a barium sulfate or calcium sulfate intermediate
50
51 forms during the wet ball-milling process, which results in an increase in the preparation
52
53 temperature of M-type hexaferrite. Impurities such as Al and Si are common in acid
54
55 mine drainage sludge. There is competition between the CaO – BaO – Fe_2O_3 and CaO –
56
57
58
59
60
61
62
63
64
65

1
2 BaO–Al₂O₃–SiO₂ systems, which would have a negative effect on the efficiency of Ca
3
4 and Ba in M-type hexaferrite preparation.
5
6
7

8
9 **Figure 5.** Proposed reaction pathways of M-type hexaferrite preparation: (a) sludge
10 free, (b) sludge A, and (c) both sludge A and sludge B.
11
12
13
14
15

16 3.4 Possible applications of the acid mine drainage sludge-derived hexaferrite

17
18 The pure M-type hexaferrite has a wide range of applications as we mentioned before.
19
20 In the case of the sludge- derived hexaferrite, the effect of secondary phases need to be
21 considered.
22
23
24

25
26 For applications as a magnetic material, the nonmagnetic secondary phases have
27 negative effects on performance. Separation of magnetic phases from nonmagnetic
28 phases might enhances the performance of the sludge-derived products. But it is
29 evidently a big challenge for practice. And the residues after separation have to be
30 considered.
31
32
33
34
35
36

37
38 For application in microwave absorber, the secondary phases are not always
39 detrimental. For example, celsian also contributes to microwave absorption (Li, et al.,
40 2016). Therefore the acid mine drainage sludge-derived hexaferrite products have
41 potential to use without further purification process after calcination.
42
43
44
45
46
47

48
49 Furthermore, future work will deal with optimization of the amount of sludge addition
50 for hexaferrite synthesis in microwave absorption and other applications.
51
52
53
54
55
56

57 4. Conclusion

1
2 The present study indicates that acid mine drainage sludge can be completely reused
3
4 to produce value-added M-type hexaferrite by a solid-state reaction process. It is
5
6 expected that reduction in sludge accumulation in deposits at mine sites using this
7
8 process. The findings suggest how the impurities in the sludge, such as silicon,
9
10 aluminum, and sulfur, are involved in phase formation and microstructure evolution of
11
12 the calcined samples. A fundamental understanding of the behavior of metal-rich sludge
13
14 in M-type hexaferrite preparation can be achieved. In addition, possible applications of
15
16 M-type hexaferrite prepared from acid mine drainage sludge as a microwave absorber or
17
18 magnetic material are proposed. The magnetic properties and microwave absorption
19
20 properties of the as-prepared products will be investigated in future work.
21
22
23
24
25
26
27
28
29
30

31 **Acknowledgments**

32
33 We gratefully acknowledge financial support from the Management Expenses Grants
34
35 for National Universities Corporations “Dynamic Alliance for Open Innovation
36
37 Bridging Human, Environment and Materials” from the Ministry of Education, Culture,
38
39 Sports, Science and Technology (MEXT) of Japan, and the Hatakeyama Culture
40
41 Foundation. Mei Liu also acknowledges support from the Japanese government
42
43 (MEXT) for a doctoral scholarship. Part of this research was performed as joint research
44
45 with the Metals Environmental Management Department, Japan Oil, Gas and Metals
46
47 National Corporation (JOGMEC), and we would like to express our thanks to them for
48
49 valuable technical discussions and support in sludge sampling. We also thank the
50
51 Fundamental Technology Center, Research Institute of Electrical Communication,
52
53 Tohoku University for the FE-SEM observations performed with a Hitachi SU6600
54
55
56
57
58
59
60
61
62
63
64
65

1
2 scanning electron microscope. We thank Edanz Group (www.edanzediting.com/ac) for
3
4 editing a draft of this manuscript.
5
6
7
8
9

10 **Appendix A. Supplementary data**

11
12
13
14
15

16 **References**

17

- 18 Alcolea, A., Ibarra, I., Caparrós, A., Rodríguez, R. **Study of the MS response by TG–**
19 **MS in an acid mine drainage efflorescence.** *J. Therm. Anal. Calorim.* 101 (2010),
20 pp. 1161–1165. <https://doi.org/10.1007/s10973-009-0556-8>.
21
22
- 23 Anbarasu, V., Md Gazzali, P.M., Karthik, T., Manigandan, A., Sivakumar, K. **Effect of**
24 **divalent cation substitution in the magnetoplumbite structured BaFe₁₂O₁₉**
25 **system.** *J. Mater. Sci. Mater. Electron.* 24 (2013), pp. 916–926.
26 <https://doi.org/10.1007/s10854-012-0850-2>.
27
28
- 29 Blanco, A.M., Gonzalez, F.C. **Magnetic properties and distribution of Ca ions in Ca-**
30 **substituted Ba-ferrite.** *J. Phys. D: Appl. Phys.* 24 (1991), pp. 612–618.
31 <https://doi.org/10.1088/0022-3727/24/4/013>.
32
33
- 34 Elwell, D., Neate, B.W. **Mechanisms of crystal growth from fluxed melts.** *J. Mater.*
35 *Sci.* 6 (1971), pp. 1499–1519. <https://doi.org/10.1007/BF02403090>.
36
37
- 38 Goldman, A. **Modern Ferrite Technology, 2nd ed.** Springer Science and Business
39 Media, Inc., Pittsburgh (2006).
40
- 41 Harris, V.G., Geiler, A., Chen, Y., Yoon, S.D., Wu, M., Yang, A., Chen, Z., He, P.,
42 Parimi, P. V., Zuo, X., Patton, C.E., Abe, M., Acher, O., Vittoria, C. **Recent**
43 **advances in processing and applications of microwave ferrites.** *J. Magn. Magn.*
44 *Mater.* 321 (2009), pp. 2035–2047. <https://doi.org/10.1016/J.JMMM.2009.01.004>.
45
46
- 47 Ichinose, N., Kurihara, K. A new ferrimagnetic compound. *J. Phys. Soc. Japan* 18
48 (1963), pp. 1700–1701. <https://doi.org/10.1143/JPSJ.18.1700>.
49
50
- 51 Iqbal, M.J., Ashiq, M.N., Hussain, I.H. **Physical, electrical and dielectric properties**
52 **of Ca-substituted strontium hexaferrite (SrFe₁₂O₁₉) nanoparticles synthesized**
53 **by co-precipitation method.** *J. Magn. Magn. Mater.* 322 (2010), pp. 1720–1726.
54 <https://doi.org/10.1016/j.jmmm.2009.12.013>.
55
56
- 57 Karmakar, M., Mondal, B., Pal, M., Mukherjee, K. **Acetone and ethanol sensing of**
58 **barium hexaferrite particles: A case study considering the possibilities of non-**
59
60
61
62
63
64
65

1
2 **conventional hexaferrite sensor.** *Sensors Actuators B Chem.* 190 (2014), pp.
3 627–633. <https://doi.org/10.1016/J.SNB.2013.09.035>.

4
5
6 Kojima, H. **Fundamental properties of hexagonal ferrites with magnetoplumbite**
7 **structure, in: Wohlfarth, E.P. (Ed.), Ferromagnetic Materials.** North-Holland
8 Publishing Company, New York (1982), pp. 305.
9

10
11 Li, M., Yin, X., Chen, L., Han, M., Cheng, L., Zhang, L. **Dielectric and**
12 **electromagnetic wave absorption properties of reduced graphene**
13 **oxide/barium aluminosilicate glass–ceramic composites.** *Ceram. Int.* 42 (2016),
14 pp. 7099–7106. <https://doi.org/10.1016/J.CERAMINT.2016.01.098>.
15

16
17 Liu, M., Iizuka, A., Shibata, E. **Recent progress on utilization of metal-rich wastes in**
18 **ferrite processing: a review.** *Waste Biomass Valori.* 9 (2018), pp. 1669-1679.
19 <https://doi.org/10.1007/s12649-017-9909-x>.
20

21
22 Lotgering, F.K., Huyberts, M.A.H. **Composition and magnetic properties of**
23 **hexagonal Ca, La ferrite with magnetoplumbite structure.** *Solid State Commun.*
24 34 (1980), pp. 49–50. [https://doi.org/10.1016/0038-1098\(80\)90627-4](https://doi.org/10.1016/0038-1098(80)90627-4).
25

26
27 Macías, F., Pérez-López, R., Caraballo, M.A., Cánovas, C.R., Nieto, J.M. **Management**
28 **strategies and valorization for waste sludge from active treatment of**
29 **extremely metal-polluted acid mine drainage: A contribution for sustainable**
30 **mining.** *J. Clean. Prod.* 141 (2017), pp. 1057–1066.
31 <https://doi.org/10.1016/j.jclepro.2016.09.181>.
32

33
34 Patel, H. A.; Byun, J.; Yavuz, C. T. **Arsenic removal by magnetic nanocrystalline**
35 **barium hexaferrite.** *J. Nanoparticle Res.* 14 (2012), pp. 881.
36 <https://doi.org/10.1007/s11051-012-0881-x>.
37

38
39
40
41 Pulišová, P., Máša, B., Michalková, E., Večerníková, E., Maříková, M., Bezdička, P.,
42 Murafa, N., Šubrt, J. **Thermal behaviour of natural and synthetic iron**
43 **precipitates from mine drainage.** *J. Therm. Anal. Calorim.* 116 (2014), pp. 625–
44 632. <https://doi.org/10.1007/s10973-014-3759-6>.
45
46

47
48 Pullar, R.C. **Hexagonal ferrites: A review of the synthesis, properties and**
49 **applications of hexaferrite ceramics.** *Prog. Mater. Sci.* 57 (2012), pp. 1191–1334.
50 <https://doi.org/10.1016/J.PMATSCI.2012.04.001>.
51

52
53 Pullar, R.C., Hajjaji, W., Amaral, J.S., Seabra, M.P., Labrincha, J. A. **Magnetic**
54 **properties of ferrite ceramics made from wastes.** *Waste Biomass Valori.* 5
55 (2013), pp. 133–138. <https://doi.org/10.1007/s12649-013-9207-1>.
56
57
58
59
60
61
62
63
64
65

- 1
2 Pullar, R.C., Saeli, M., Novais, R.M., Amaral, J.S., Labrincha, J.A. **Valorisation of**
3 **industrial iron oxide waste to produce magnetic barium hexaferrite.** Chem. Sel.
4 1 (2016), pp. 819–825. <https://doi.org/10.1002/slct.201500042>.
5
6
7 Rakotonimaro, T. V., Neculita, C. M., Bussière, B., Benzaazoua, M., Zagury, G. J.
8 **Recovery and Reuse of Sludge from Active and Passive Treatment of Mine**
9 **Drainage-Impacted Waters: A Review.** Environ. Sci. Pollut. Res. 24 (2017), pp.
10 73–91. <https://doi.org/10.1007/s11356-016-7733-7>.
11
12
13 Sanghi, S., Agarwal, A. **Rietveld refinement, electrical properties and magnetic**
14 **characteristics of Ca–Sr substituted barium hexaferrites.** J. Alloys Compd. 513
15 (2012), pp. 436–444. <https://doi.org/10.1016/j.jallcom.2011.10.071>.
16
17
18 Sixtus, K.J., Kronenberg, K.J., Tenzer, R.K. **Investigations on barium ferrite magnets.**
19 J. Appl. Phys. 27 (1956), pp. 1051–1057. <https://doi.org/10.1063/1.1722540>.
20
21
22 Sözeri, H., Deligöz, H., Kavas, H., Baykal, A. **Magnetic, dielectric and microwave**
23 **properties of M–Ti substituted barium hexaferrites (M=Mn²⁺, Co²⁺, Cu²⁺, Ni²⁺,**
24 **Zn²⁺).** Ceram. Int. 40 (2014), pp. 8645–8657.
25 <https://doi.org/10.1016/J.CERAMINT.2014.01.082>.
26
27
28 Vinnik, D., Tarasova, A., Zhrebtsov, D., Gudkova, S., Galimov, D., Zhivulin, V.,
29 Trofimov, E., Nemrava, S., Perov, N., Isaenko, L., Niewa, R. **Magnetic and**
30 **structural properties of barium hexaferrite BaFe₁₂O₁₉ from various growth**
31 **techniques.** Materials (Basel) 10 (2017), pp. 578.
32 <https://doi.org/10.3390/ma10060578>.
33
34
35 Went, J.J., Rathenau, G.W., Gorter, E.W., van Sterhout, G.W. **Ferroxdure, a class of**
36 **new permanent magnet materials.** Philips Techn. Rev. 13(1952), pp. 194–208.
37
38
39 Zinck, J., Griffith, W. **Review of mine drainage treatment and sludge management**
40 **operations.** MEND Report 3.43.1, Mining Association of Canada, Ottawa (2013).
41 [http://mend-nedem.org/wp-](http://mend-nedem.org/wp-content/uploads/3.43.1_ReviewMineDrainageTreatmentSludge.pdf)
42 [content/uploads/3.43.1_ReviewMineDrainageTreatmentSludge.pdf](http://mend-nedem.org/wp-content/uploads/3.43.1_ReviewMineDrainageTreatmentSludge.pdf).
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

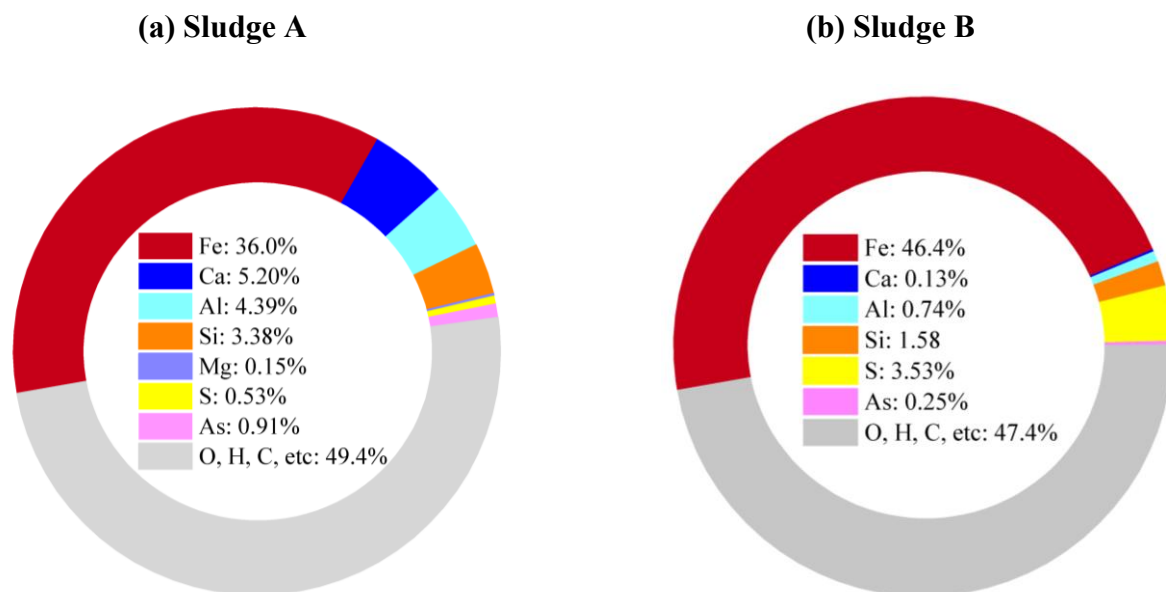
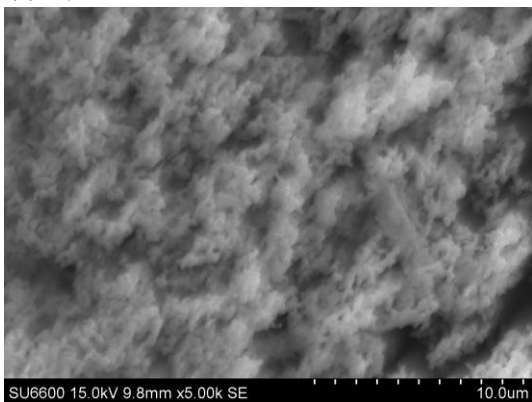
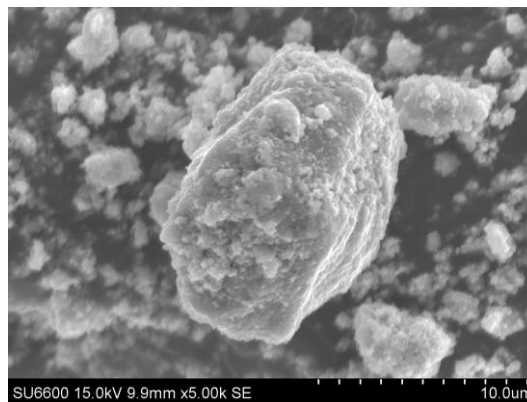


Fig. 1. Main elemental compositions of the acid mine drainage sludge determined by ICP-AES (mass %).

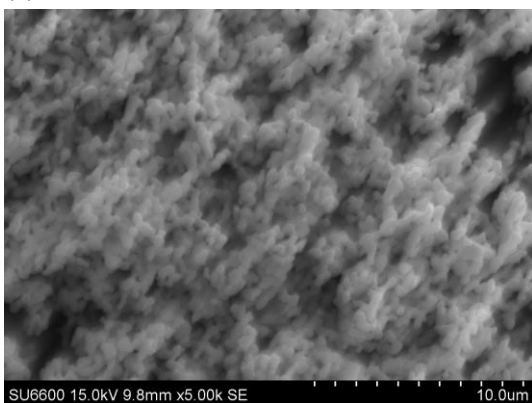
(a) $A_0F-900\text{ }^\circ\text{C}$



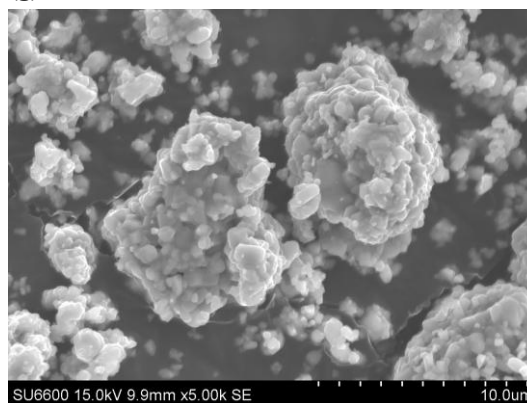
(f) $A_0B-900\text{ }^\circ\text{C}$



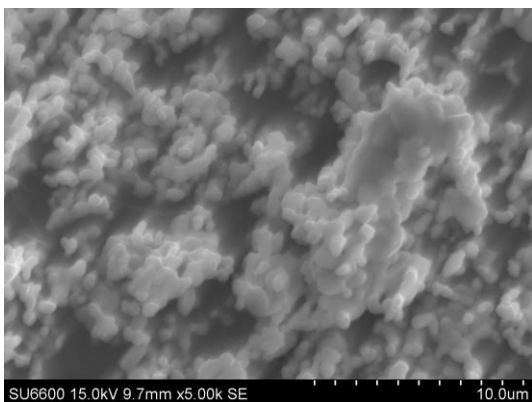
(b) $A_0F-1000\text{ }^\circ\text{C}$



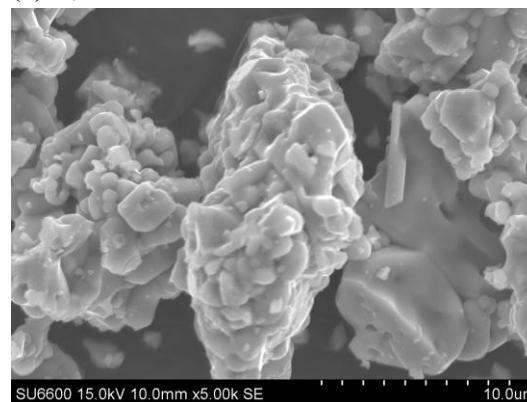
(g) $A_0B-1000\text{ }^\circ\text{C}$



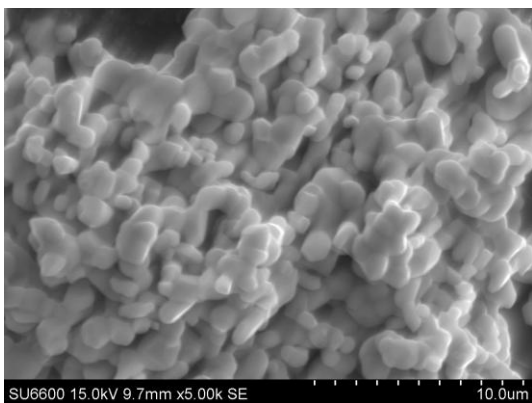
(c) $A_0F-1100\text{ }^\circ\text{C}$



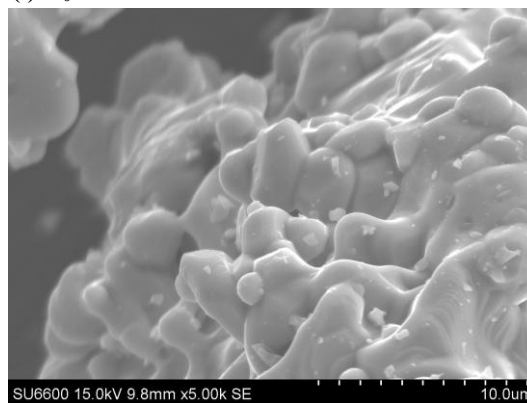
(h) $A_0B-1100\text{ }^\circ\text{C}$



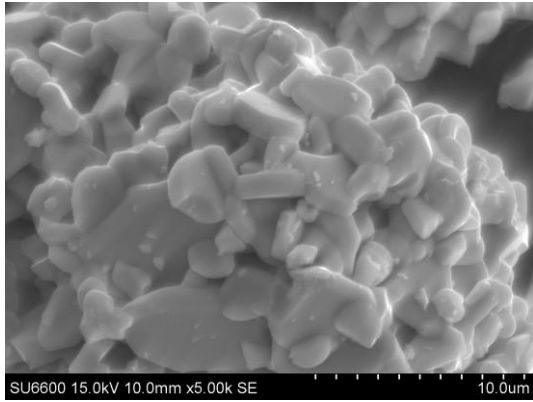
(d) $A_0F-1200\text{ }^\circ\text{C}$



(i) $A_0B-1200\text{ }^\circ\text{C}$



(e) A_0F -1300°C



(j) A_0B -1300°C

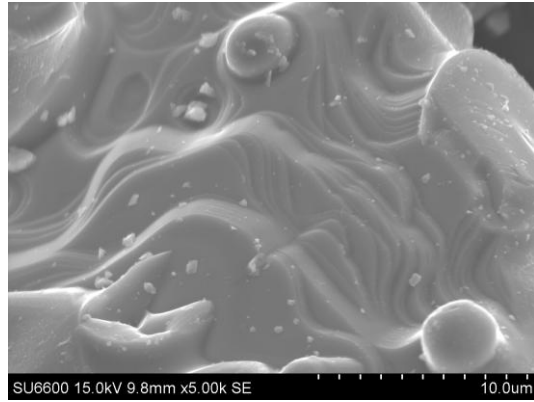


Fig. 2. Comparison of the micrographs of A_0F and A_0B calcined at 900–1300 °C for 4 h.

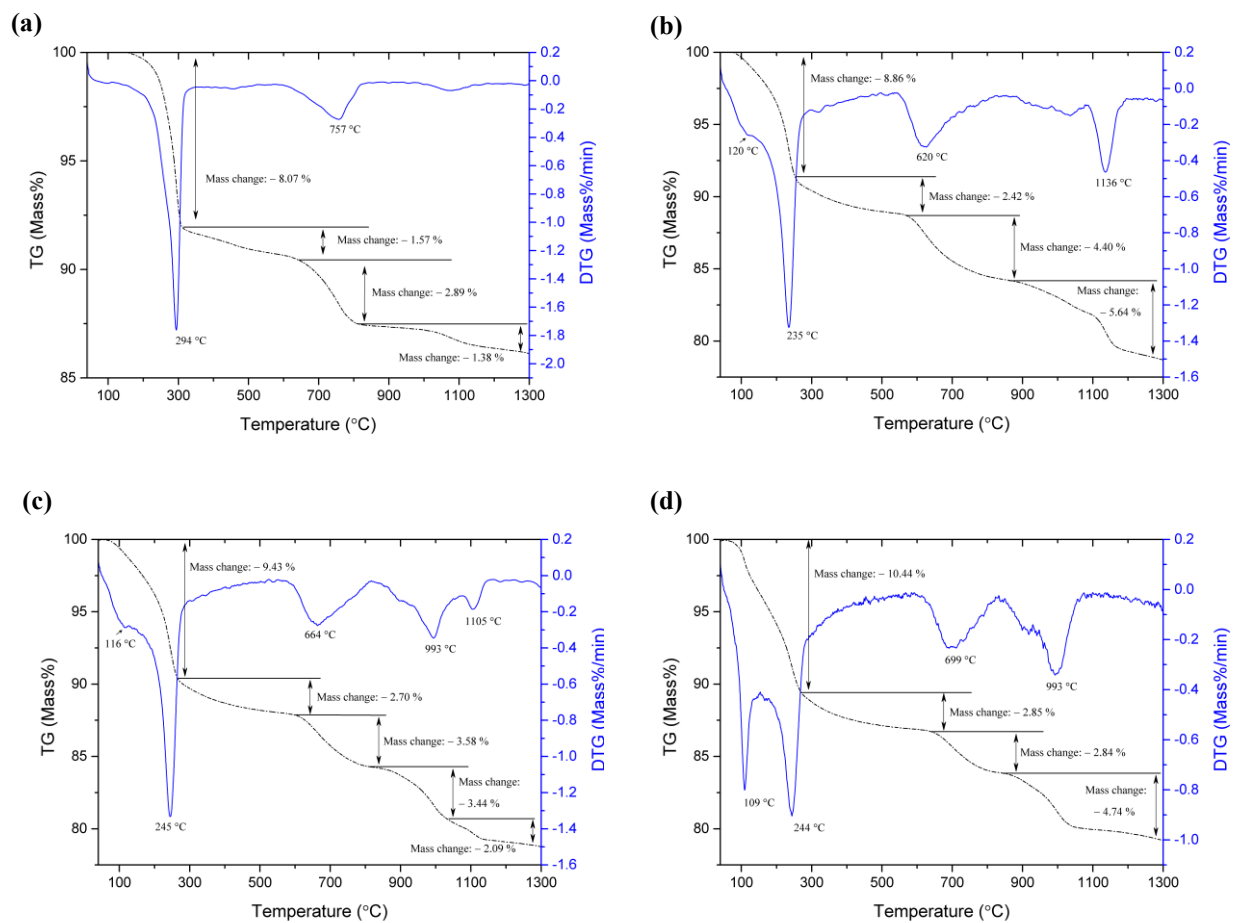


Fig. 3. TG–DTG curves of the representative precursors: (a) A₀F, (b) A₀B, (c) A₂B and (d) A₄B.

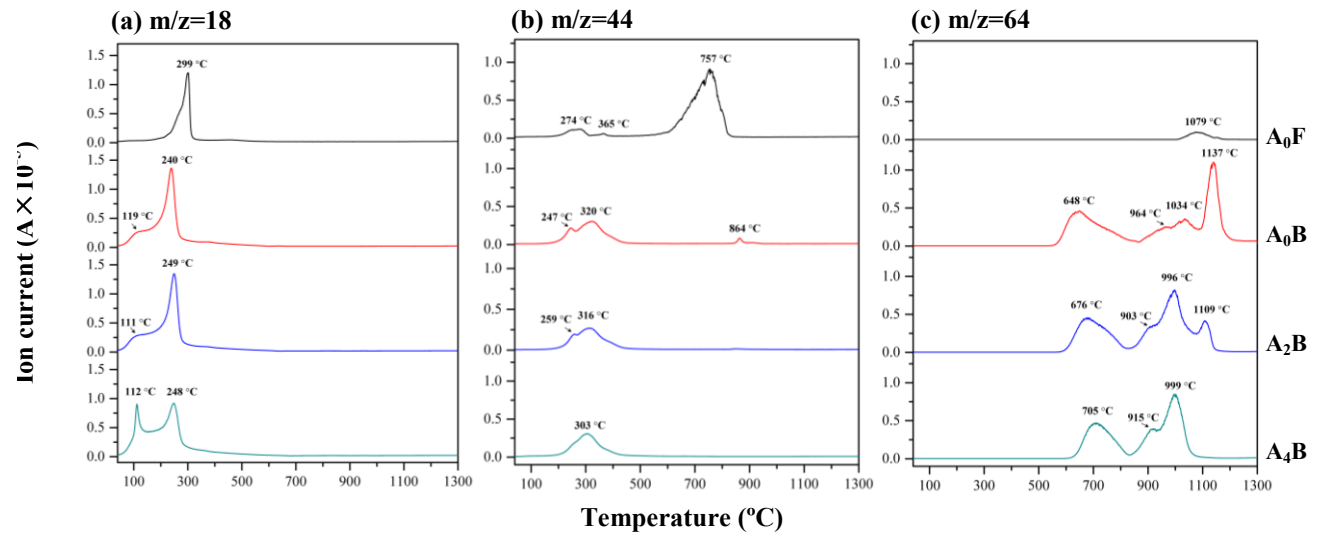
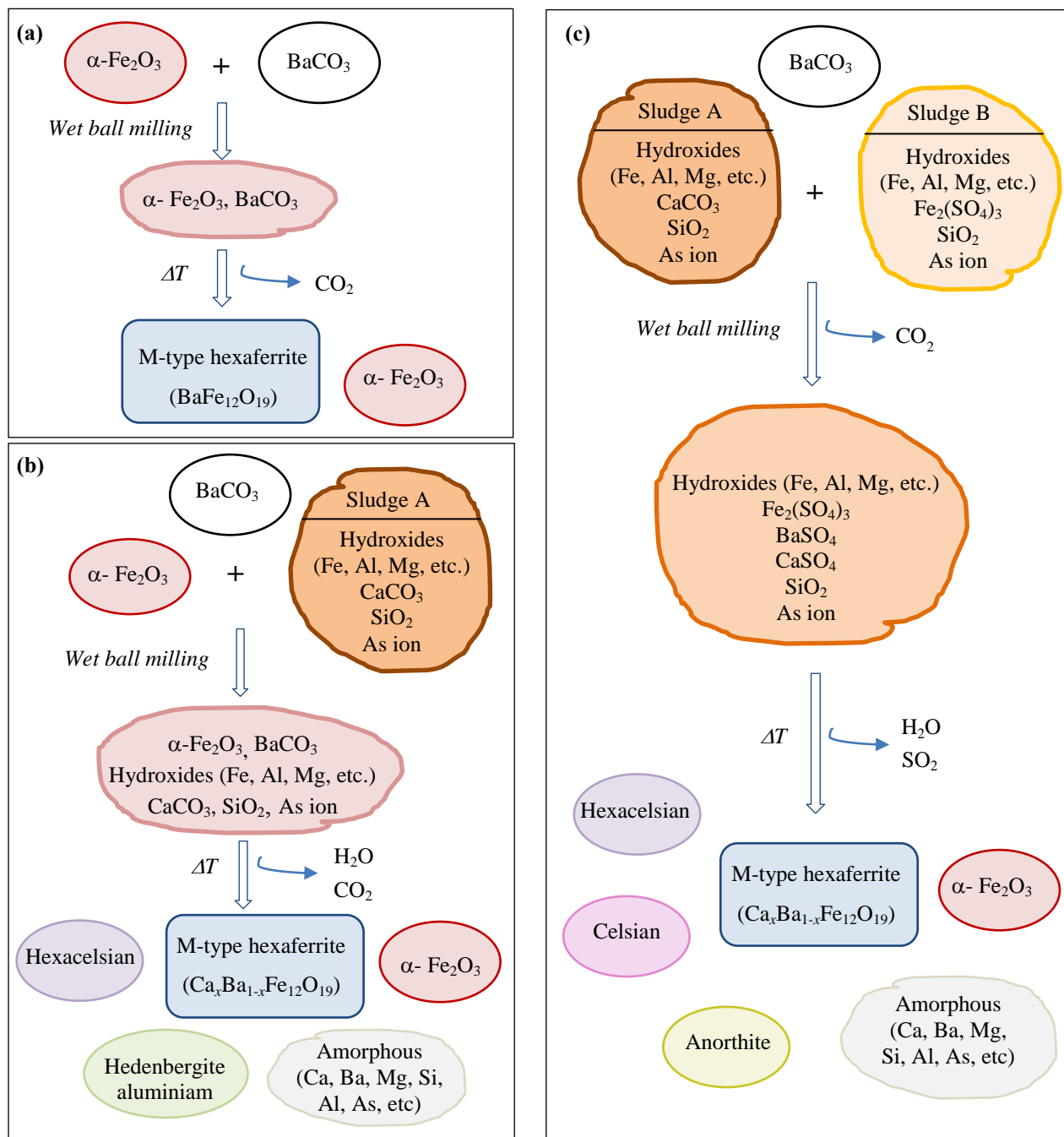
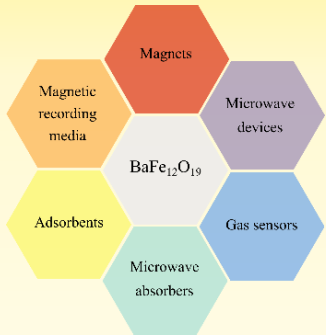


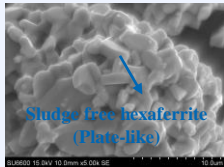
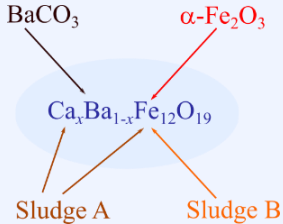
Fig. 4. Effect of sludge addition on gas evolution during heating under a simulated air atmosphere (20% O_2 and 80% He): (a) $m/z = 18$ (H_2O^+), (b) $m/z = 44$ (CO_2^+), and (c) $m/z = 64$ (SO_2^+).



Graphical Abstract



Reuse of AMD sludge in M-type hexaferrite preparation



Supporting Information for

Acid Mine Drainage Sludge as an Alternative Raw Material for M-type
Hexaferrite Preparation

Mei Liu^{a, b, *}, Atsushi Iizuka^c, Etsuro Shibata^c

^a Graduate School of Environmental Studies, Tohoku University, Sendai 980-0845,
Japan

^b School of Material Science and Engineering, Jingdezhen Ceramic Institute,
Jingdezhen 333403, China

^c Institute of Multidisciplinary Research for Advanced Materials (IMRAM), Tohoku
University, Sendai 980-8577, Japan

*To whom correspondence should be addressed.

* M. Liu: liu.mei.q5@dc.tohoku.ac.jp; maggieliujci@gmail.com

Prof. A. Iizuka: atsushi.iizuka.e4@tohoku.ac.jp

Prof. E. Shibata: etsuro.shibata.e3@tohoku.ac.jp

Crystalline iron oxide or hydroxide was not identified from XRD patterns. Iron may exist as amorphous iron hydroxide.

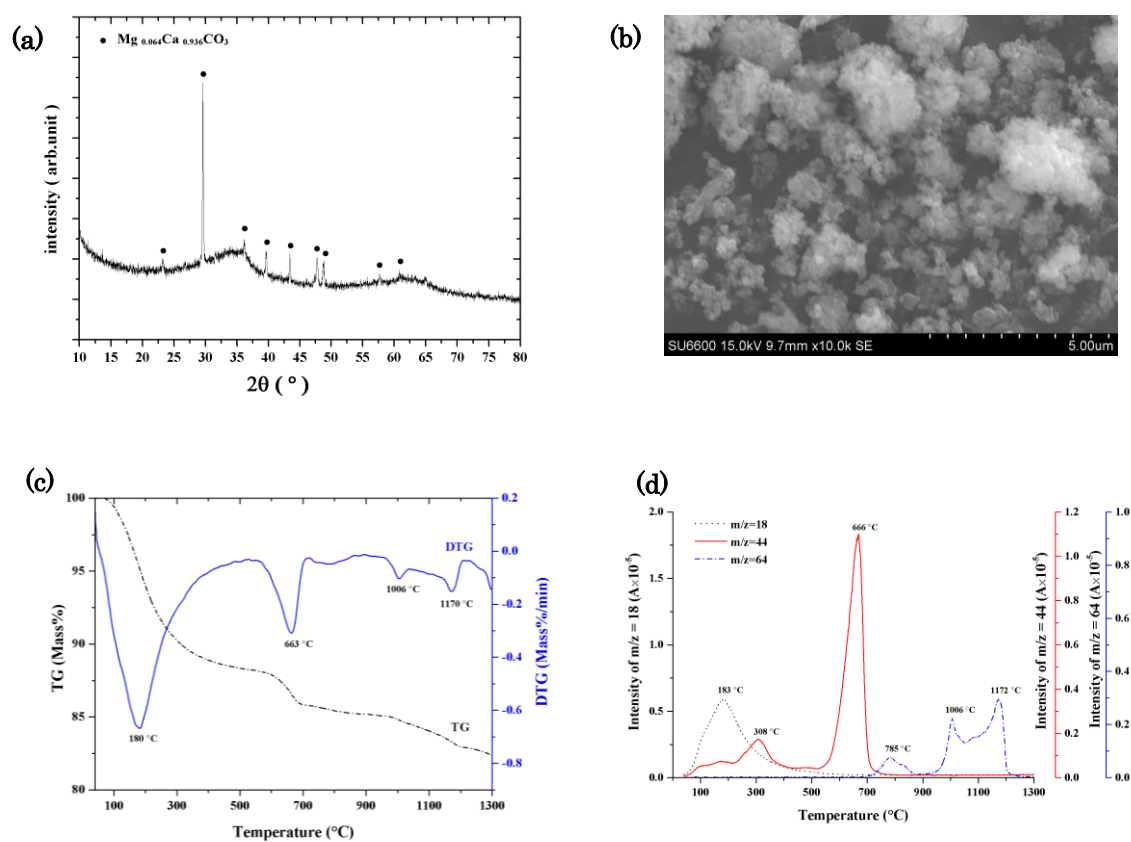


Figure S1. XRD patterns, micrograph, and TG-MS curves of dried Sludge A. (a) XRD; (b) micrograph; (c) TG-DTG curves; (d) MS curves. ^[1]

α -FeOOH was identified by XRD.

The TG-MS suggested that iron also existed as ferric sulfate.

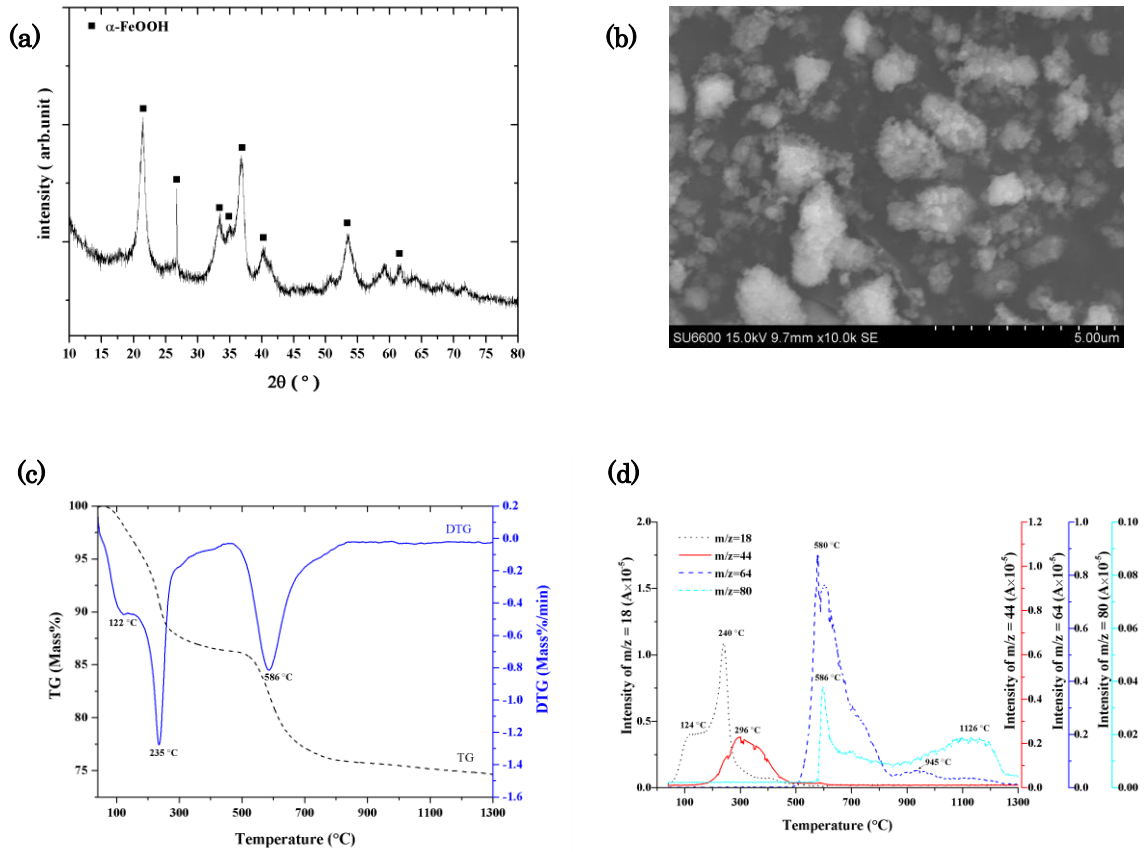


Figure S2. XRD patterns, micrograph, and TG-MS curves of dried Sludge B. (a) XRD; (b) micrograph; (c) TG-DTG curves; (d)MS curves. ^[1]

Ref.

[1] Liu, M., Iizuka, A., Shibata, E. **Effect of Temperature on Phase Transformation and Leaching Behavior of Acid Mine Drainage Sludge.** Mater. Trans. (in press)

<https://doi.org/10.2320/matertrans.M-M2018848>

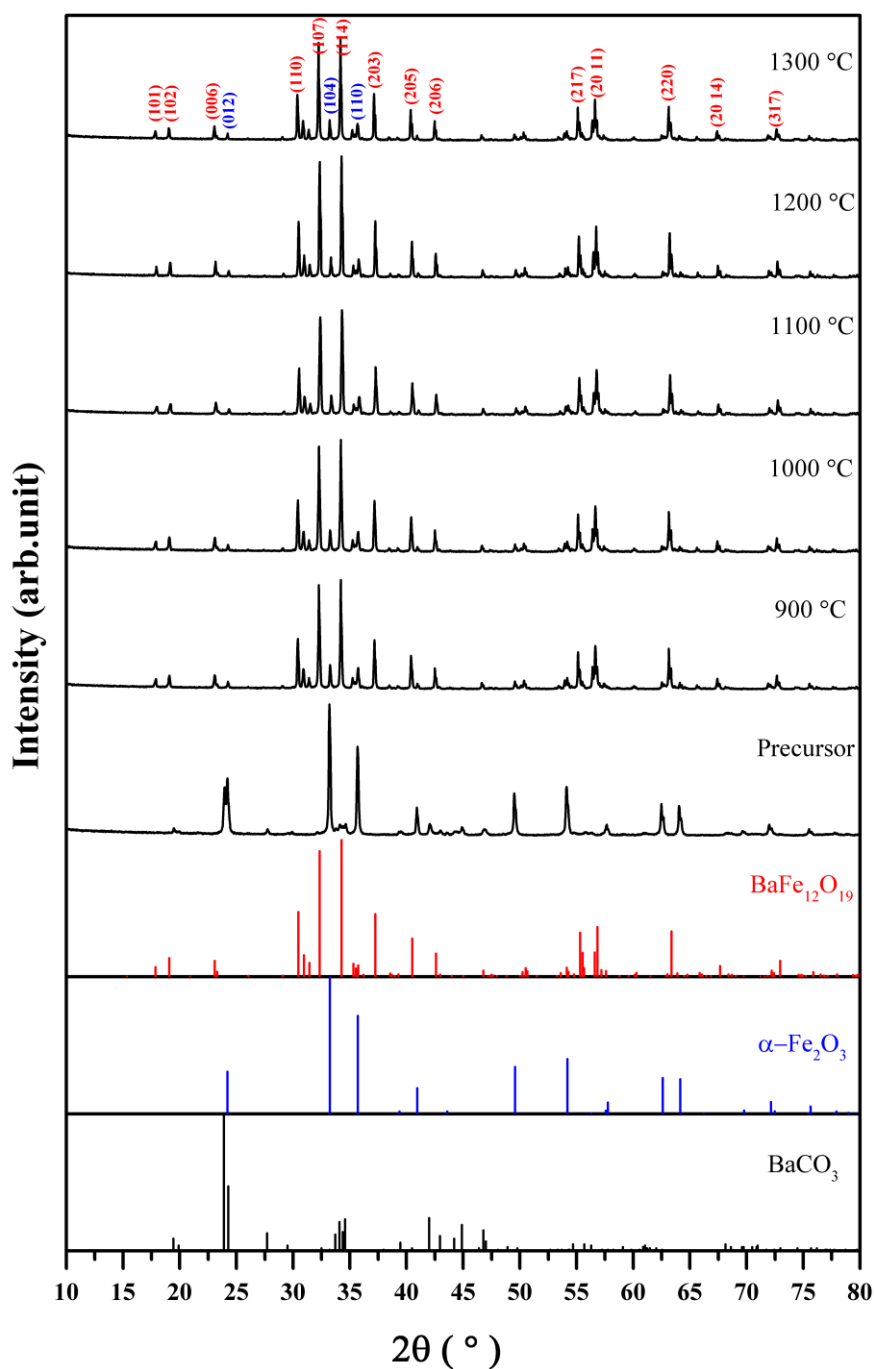


Figure S3. XRD patterns of A₀F before and after calcination at the target temperature for 4 h. The standard patterns retrieved from the database of the ICDD include BaFe₁₂O₁₉ (PDF 00-043-0002), α-Fe₂O₃ (PDF 01-079-0007), and BaCO₃ (PDF 00-044-1487).

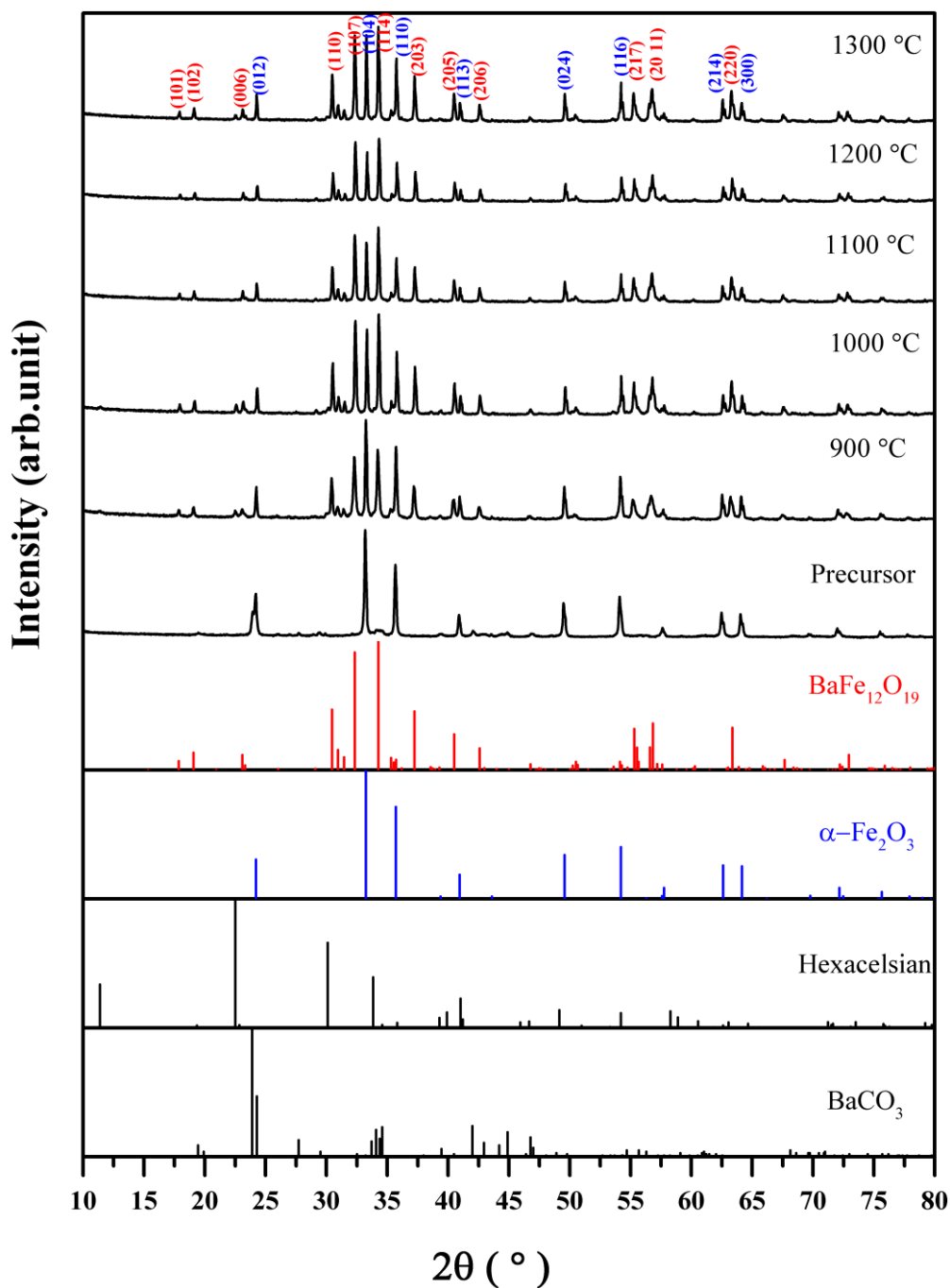


Figure S4. XRD patterns of A₁F before and after calcination at the target temperature for 4 h. The standard patterns retrieved from the database of the ICDD include BaFe₁₂O₁₉ (PDF 00-043-0002), α-Fe₂O₃ (PDF 01-079-0007), hexacelsian (BaAl₂Si₂O₈, PDF 01-072-7502), and BaCO₃ (PDF 00-044-1487).

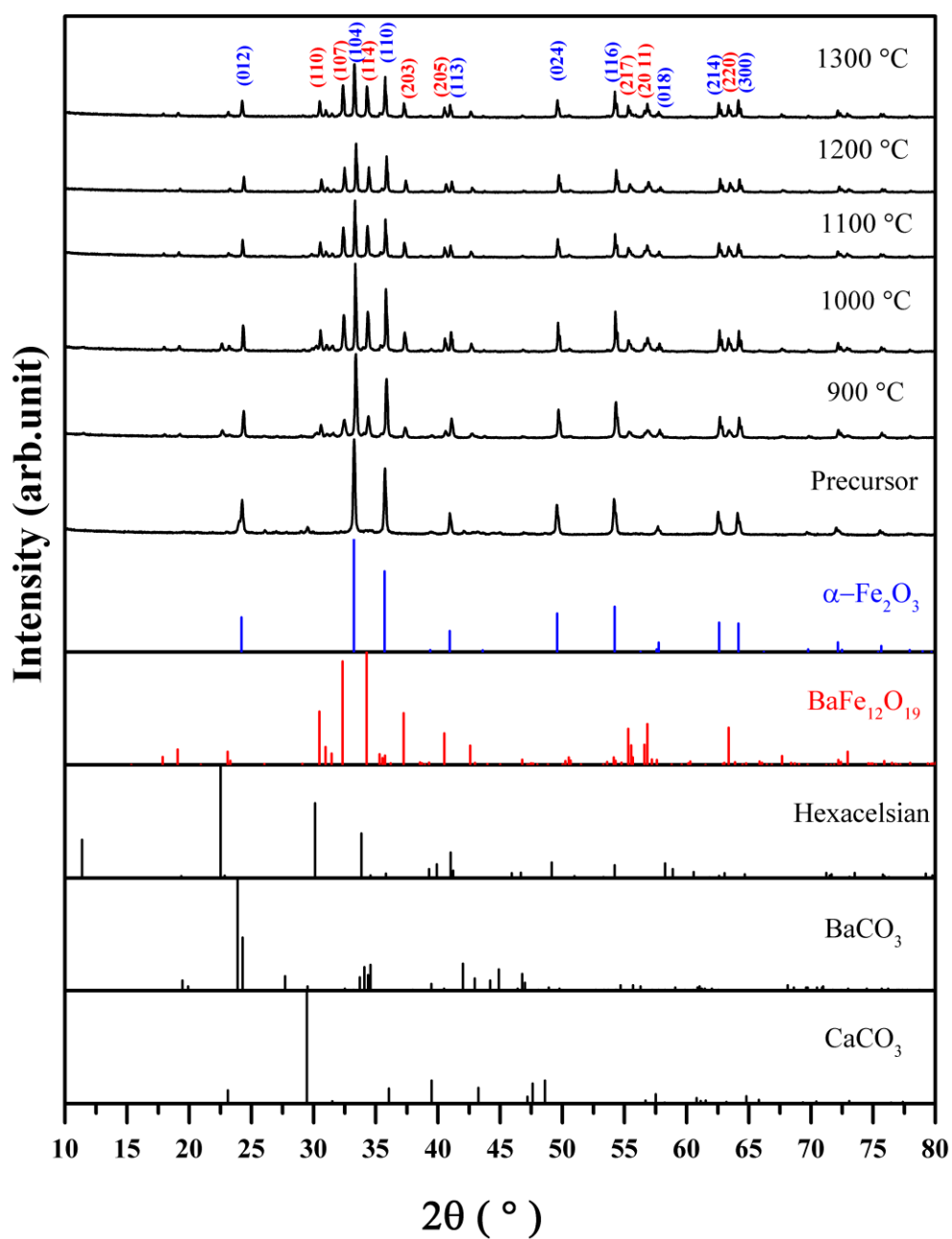


Figure S5. XRD patterns of A₂F before and after calcination at the target temperature for 4 h. The standard patterns retrieved from the database of the ICDD include BaFe₁₂O₁₉ (PDF 00-043-0002), α-Fe₂O₃ (PDF 01-079-0007), hexacelsian (BaAl₂Si₂O₈, PDF 01-072-7502), BaCO₃ (PDF 00-044-1487), and CaCO₃ (PDF 01-085-0849).

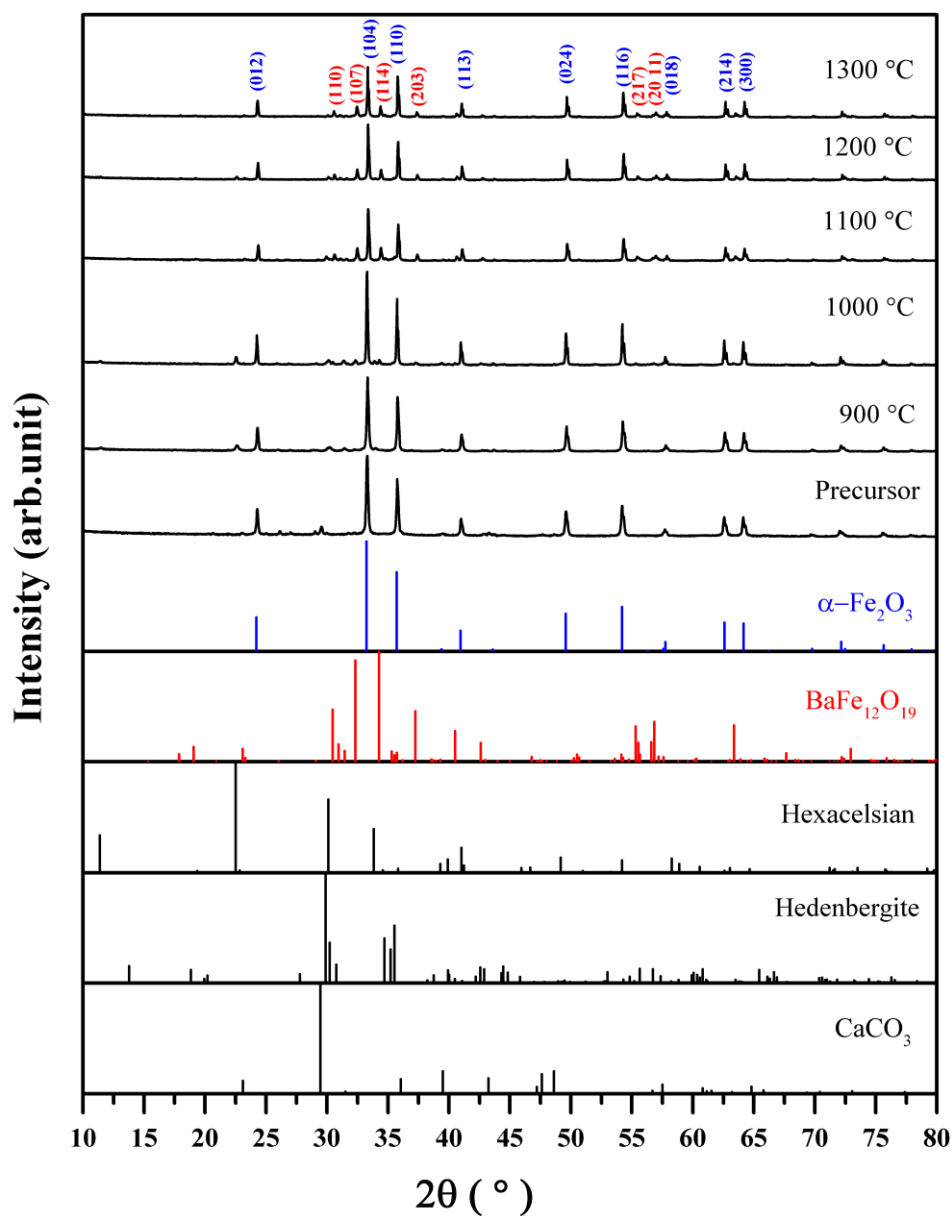


Figure S6. XRD patterns of A_3F before and after calcination at the target temperature for 4 h. The standard patterns retrieved from the database of the ICDD include $BaFe_{12}O_{19}$ (PDF 00-043-0002), $\alpha\text{-Fe}_2\text{O}_3$ (PDF 01-079-0007), hexacelsian ($BaAl_2Si_2O_8$, PDF 01-072-7502), hedenbergite aluminian ($Ca(Fe_{0.821}Al_{0.179})(SiAl_{0.822}Fe_{0.178}O_6)$, PDF 01-078-1546), and $CaCO_3$ (PDF 01-085-0849).

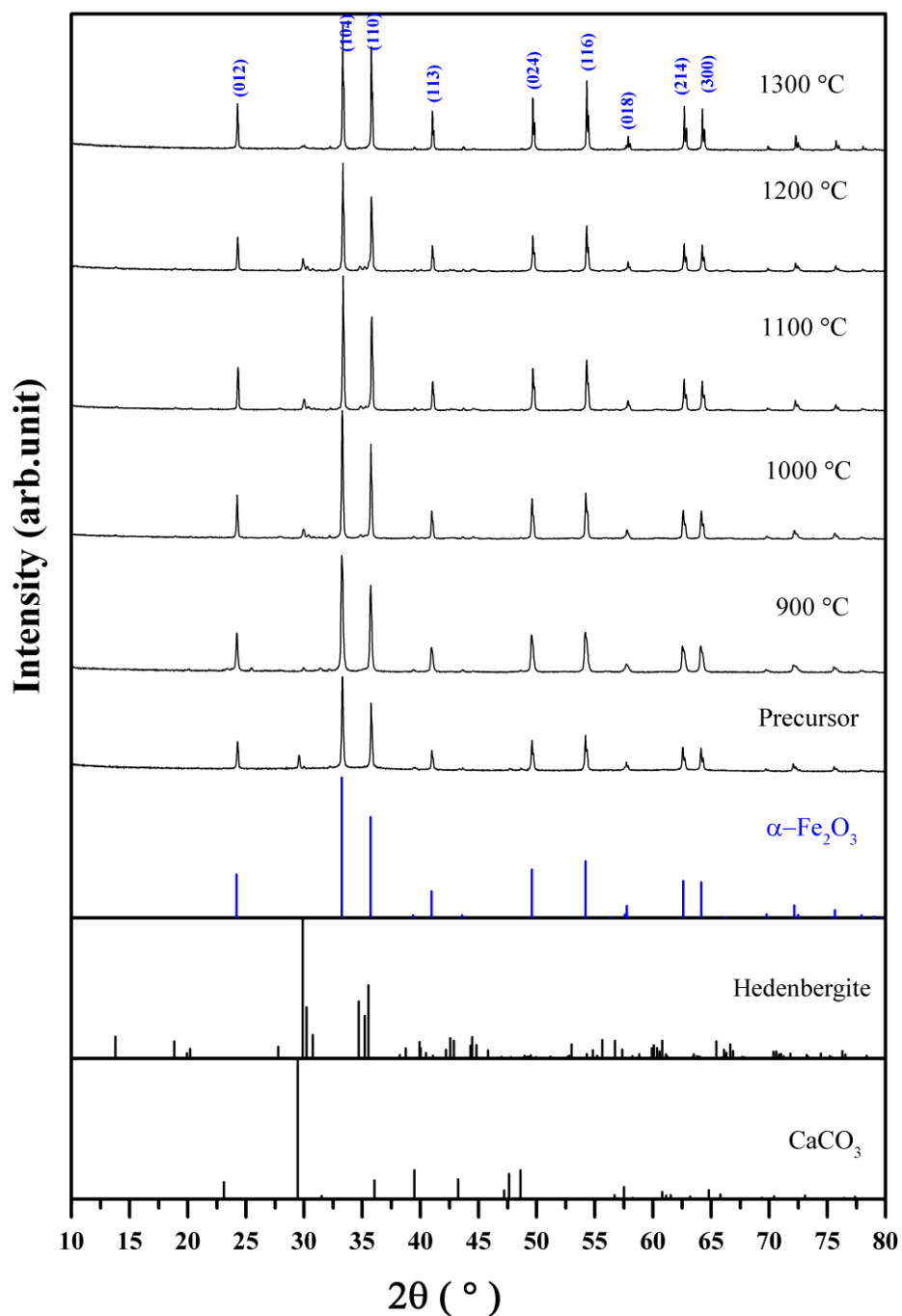


Figure S7. XRD patterns of A₄F before and after calcination at the target temperature for 4 h. The standard patterns retrieved from the database of the ICDD include α-Fe₂O₃ (PDF 01-079-0007), hedenbergite aluminian (Ca(Fe_{0.821}Al_{0.179})(SiAl_{0.822}Fe_{0.178}O₆), PDF 01-078-1546), and CaCO₃ (PDF 01-085-0849).

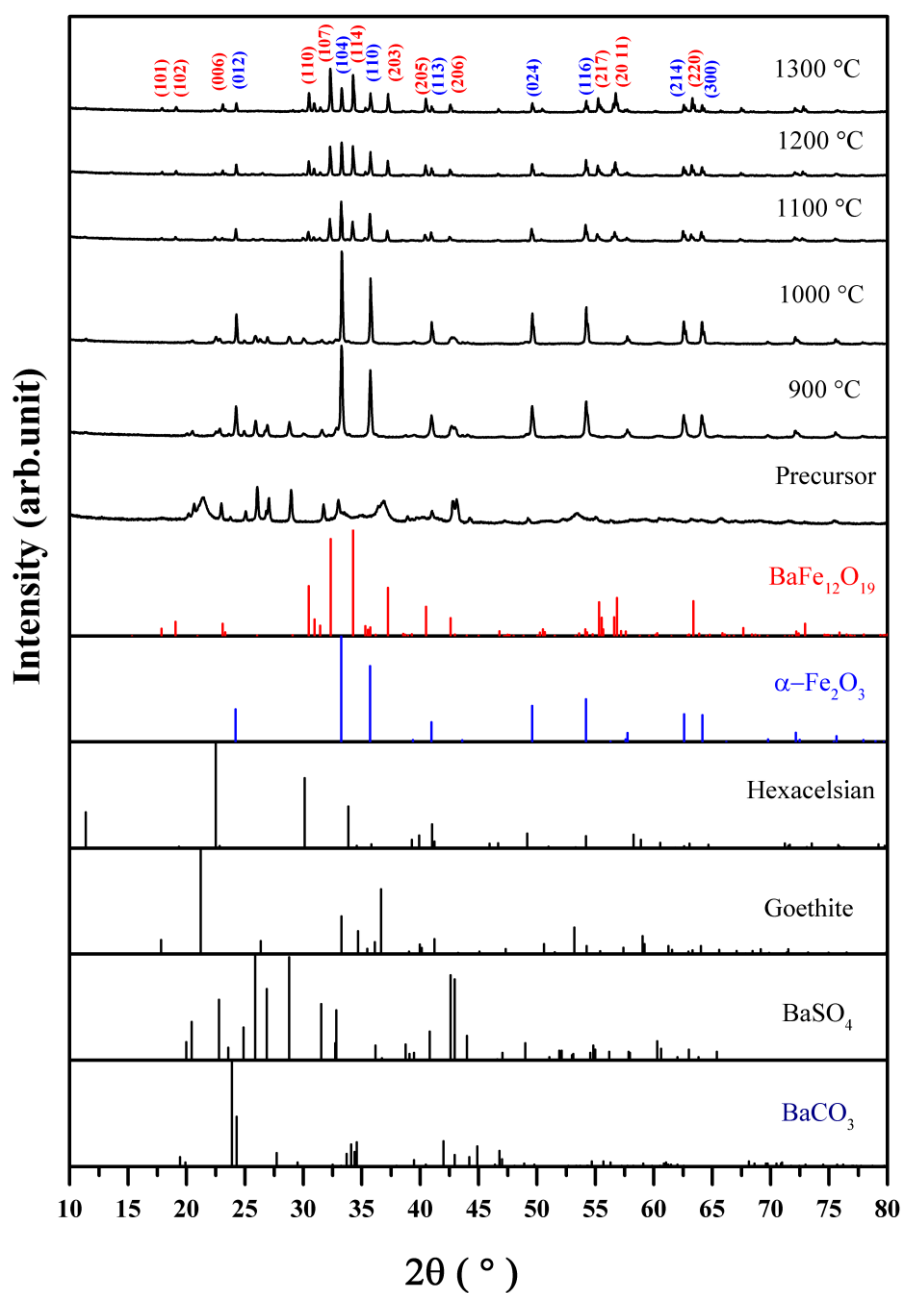


Figure S8. XRD patterns of A_0B before and after calcination at the target temperature for 4 h. The standard patterns retrieved from the database of the ICDD include $BaFe_{12}O_{19}$ (PDF 00-043-0002), $\alpha\text{-Fe}_2\text{O}_3$ (PDF 01-079-0007), hexacelsian ($BaAl_2Si_2O_8$, PDF 01-072-7502), goethite ($\alpha\text{-FeOOH}$, PDF 01-076-7156), $BaSO_4$ (PDF 00-005-0448), and $BaCO_3$ (PDF 00-044-1487).

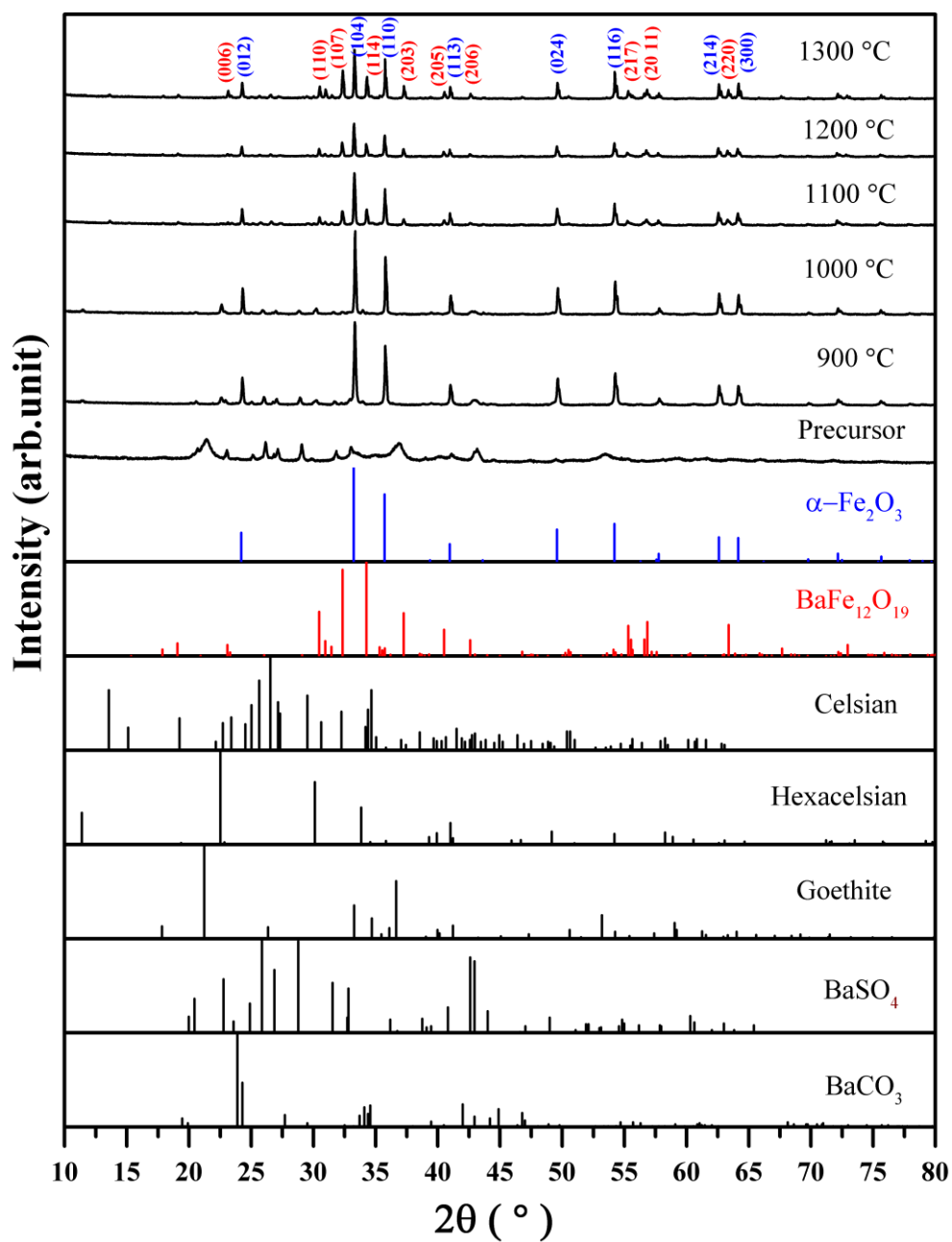


Figure S9. XRD patterns of A_1B before and after calcination at the target temperature for 4 h. The standard patterns retrieved from the database of the ICDD include $BaFe_{12}O_{19}$ (PDF 00-043-0002), $\alpha\text{-Fe}_2\text{O}_3$ (PDF 01-079-0007), celsian ($BaAl_2Si_2O_8$, PDF 00-038-1450), hexacelsian ($BaAl_2Si_2O_8$, PDF 01-072-7502), goethite ($\alpha\text{-FeOOH}$, PDF 01-076-7156), $BaSO_4$ (PDF 00-005-0448), and $BaCO_3$ (PDF 00-044-1487).

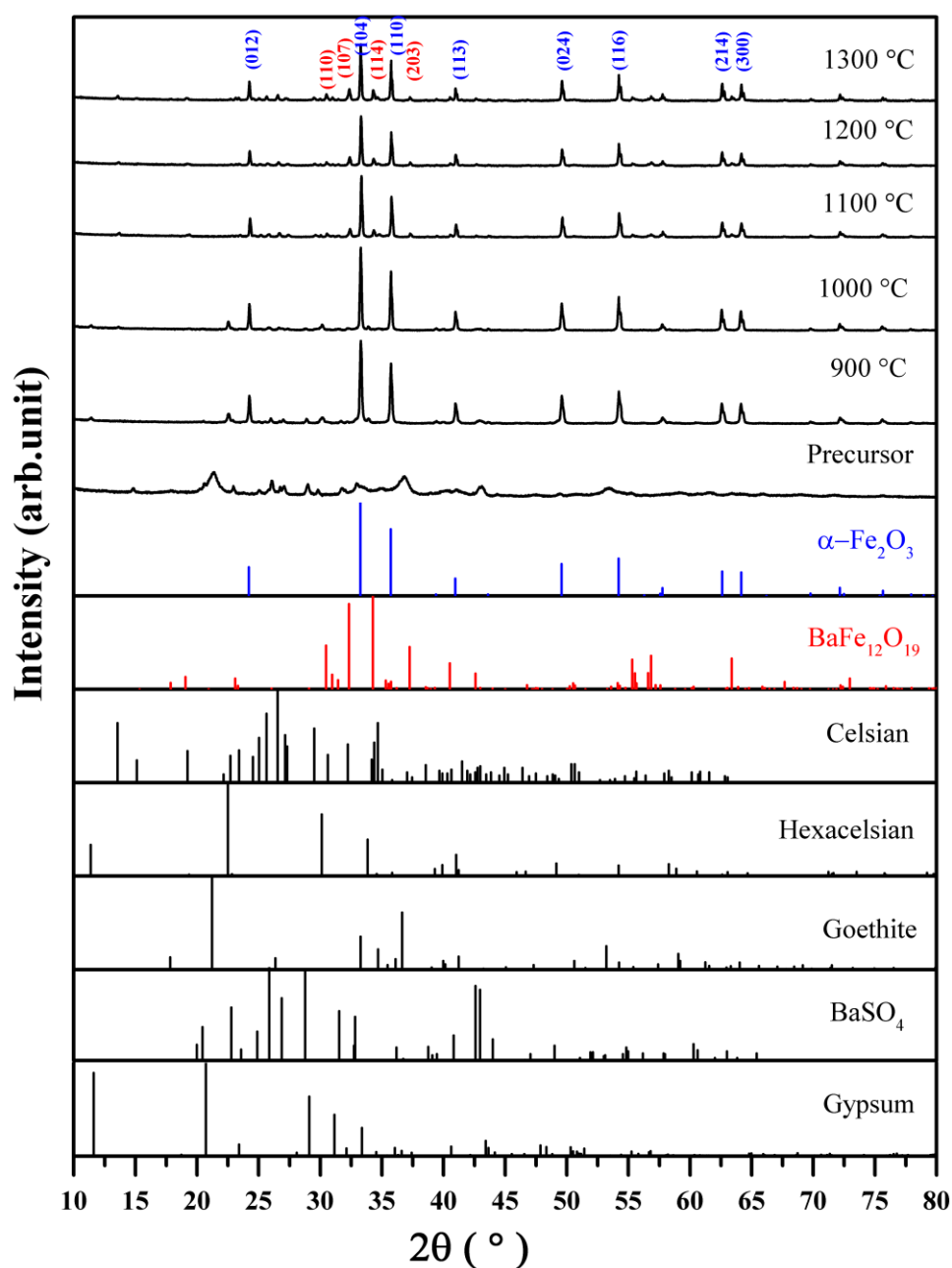


Figure S10. XRD patterns of A₂B before and after calcination at the target temperature for 4 h. The standard patterns retrieved from the database of the ICDD include BaFe₁₂O₁₉ (PDF 00-043-0002), α-Fe₂O₃ (PDF 01-079-0007), celsian (BaAl₂Si₂O₈, PDF 00-038-1450), hexacelsian (BaAl₂Si₂O₈, PDF 01-072-7502), goethite (α-FeOOH, PDF 01-076-7156), BaSO₄ (PDF 00-005-0448), and gypsum (CaSO₄·(H₂O)₂, PDF

01-070-0983).

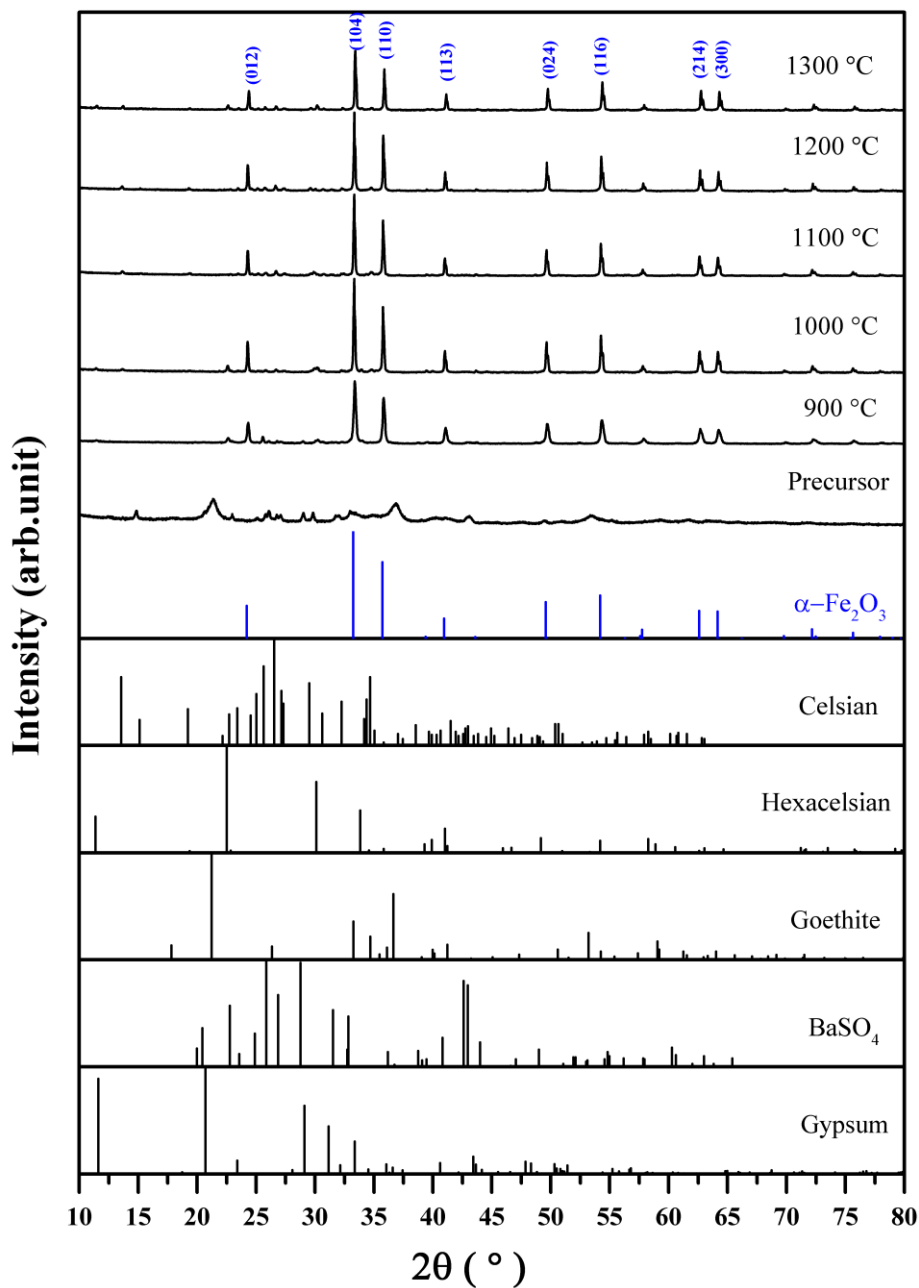


Figure S11. XRD patterns of A₃B before and after calcination at the target temperature for 4 h. The standard patterns retrieved from the database of the ICDD include α-Fe₂O₃ (PDF 01-079-0007), celsian (BaAl₂Si₂O₈, PDF 00-038-1450), hexacelsian (BaAl₂Si₂O₈, PDF 01-072-7502), goethite (α-FeOOH, PDF 01-076-7156), BaSO₄ (PDF

00-005-0448), and gypsum ($\text{CaSO}_4 \cdot (\text{H}_2\text{O})_2$, PDF 01-070-0983).

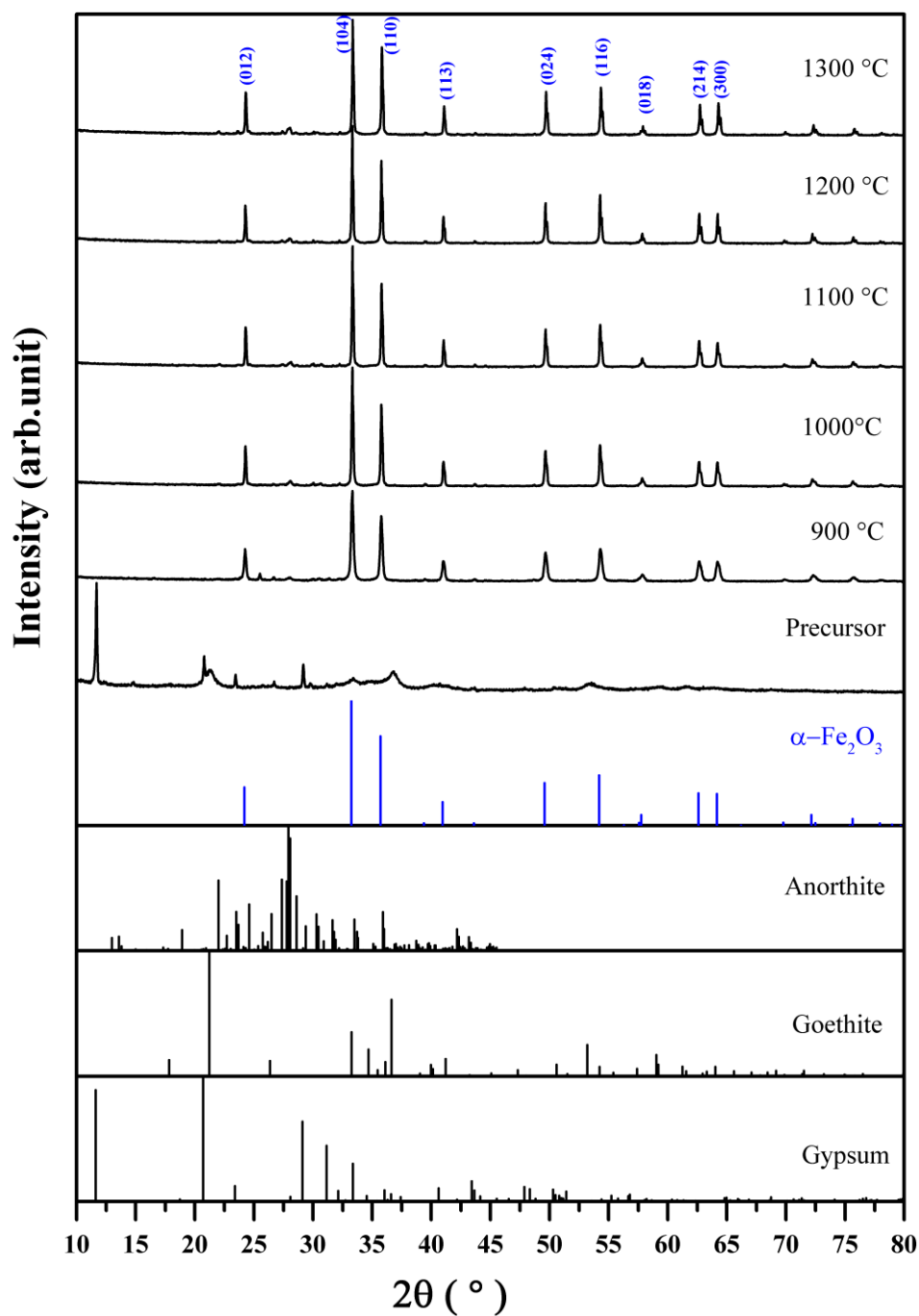


Figure S12. XRD patterns of A₄B before and after calcination at the target temperature for 4 h. The standard patterns retrieved from the database of the ICDD include $\alpha\text{-Fe}_2\text{O}_3$ (PDF 01-079-0007), anorthite ($\text{CaAl}_2\text{Si}_2\text{O}_8$, PDF 01-075-1587), goethite ($\alpha\text{-FeOOH}$,

PDF 01-076-7156), and gypsum ($\text{CaSO}_4 \cdot (\text{H}_2\text{O})_2$, PDF 01-070-0983).