

1 **Coupling ultrasound to the electro-oxidation of methyl paraben**  
2 **synthetic wastewater: effect of frequency and supporting**  
3 **electrolyte**

4  
5 D. Dionisio<sup>1,2</sup>, A.J. Motheo<sup>1,\*</sup>, C. Sáez<sup>2</sup>, M.A. Rodrigo<sup>2</sup>

6  
7 <sup>(1)</sup> São Carlos Institute of Chemistry, University of São Paulo, P.O. Box 780, CEP 13560-970,  
8 São Carlos, SP, Brazil

9 <sup>(2)</sup> Department of Chemical Engineering, Faculty of Chemical Sciences & Technologies,  
10 Universidad de Castilla - La Mancha, Campus Universitario s/n, 13071 Ciudad Real, Spain

11  
12  
13  
14 \*Corresponding author

15 Phone: +55 16 33739932

16 E-mail: [artur@iqsc.usp.br](mailto:artur@iqsc.usp.br)

## 17 **Abstract**

18 In this paper, single electrolysis with diamond electrode is studied and compared to its coupling  
19 with sonolysis for the degradation of methyl paraben (MeP). Complete mineralization of MeP  
20 was achieved for both processes, in chloride and sulfate media. Results showed that, although  
21 the oxidation of pollutant is faster in the presence of chloride, sulfate medium favored the  
22 mineralization. Ultrasound irradiation enhanced the removal of organic matter due to the  
23 activation of oxidant species in both supporting electrolytes. Moreover, the formation of  $\text{Cl}_2$  in  
24 chloride medium, improves the ultrasound cavitation effect, promoting faster depletion of TOC  
25 (total organic carbon) in the first hour of treatment. Regarding the formation of more toxic  
26 products, all possible organochlorinated intermediates were removed and chlorates and  
27 perchlorates only start to be produced after total removal of TOC. Additionally, the coupling  
28 to ultrasound accelerates the destruction of intermediates and delays the formation of  $\text{ClO}_3^-$   
29 and  $\text{ClO}_4^-$ . Low and high ultrasound frequencies were evaluated and were found to produce  
30 different effects of cavitation, which affect the electrolysis in different ways. The final result  
31 will be a balance between those effects and, thus, an optimal frequency may be observed for  
32 distinctive systems.

33

## 34 **Keywords**

35 Methyl paraben, diamond anodes, electrolysis, sonolysis

36

## 37 **Highlights**

38 - Complete mineralization of methyl paraben is achieved in sulfate and chloride media;

39 - Sulfate medium favored the removal of TOC;

40 - Enhancement of electrolysis by the coupling to sonolytic process;

41 - Cavitation effects of low and high intensities are balanced in an optimal frequency;

42 - Low frequency ultrasound promotes the activation of persulfate.

## 43 **Introduction**

44 Methyl paraben (MeP) is one of a series of preservatives, which are esters of *p*-hydroxybenzoic  
45 acid and include among others, methyl, ethyl, propyl, butyl and benzyl parabens. This  
46 compound is widely used in pharmaceutical drugs, cosmetics and personal care products,  
47 singly or in combination to other preservatives, in order to achieve the intended antimicrobial  
48 effect [1, 2]. In 2004, parabens were firstly related to human breast cancer by Darbre *et al.* [3],  
49 fact that was further confirmed by Dagher *et al.* in 2012 [4]. Since the first publication,  
50 numerous studies have been reported about environment outcomes and the effects on human  
51 health resulted from the exposure to those substances [1, 5-10]. Those studies described  
52 potential estrogenic and carcinogenic activity of parabens, which classifies them as endocrine  
53 disrupting chemicals (EDCs).

54 Although parabens are still used, mainly in cosmetic products, the concern regarding the  
55 adverse effects on human health has led to a gradually restriction on their usage. Besides the  
56 fact that most countries do not have regulations to parabens usage, several organizations, such  
57 as US Food and Drug Administration (FDA), World Health Organization (WHO), European  
58 Union (EU) and Scientific Society on Costumer Safety (SCCS), have established low  
59 concentration limits of those compounds in the case of some types of products. For instance,  
60 the content of ethyl and butyl parabens in cosmetics was changed from 0.4% to 0.14% in all  
61 EU members in 2014, and these values were also applied in several South America countries  
62 [10-12]. In this context, several industries have been replacing parabens to other preservatives  
63 with a police of “paraben-free” products. However, parabens are still largely detected in  
64 various environmental matrices [12-15], which has led to the contamination of potable water  
65 and increased the exposure of humans to those compounds.

66 In order to remove emerging contaminants, such as parabens, from water and wastewater, the  
67 scientific community has invested in the development and enhancement of advanced oxidation  
68 processes (AOPs) [16-18]. These processes are based on the generation of hydroxyl radicals  
69 ( $\bullet\text{OH}$ ), which are highly reactive ( $E^0 = 2.8 \text{ V}$ ) and promote the degradation of organic  
70 pollutants. Fenton process, photolysis, ozonation and other AOPs are very efficient for the  
71 removal of certain pollutants, however not necessarily effect for the complete depletion of  
72 organic matter on the treatment of more recalcitrant substances. For that purpose,  
73 electrochemical AOPs have shown promising results on the total removal of organic  
74 contaminants from water [19-21]. For instance, Wang *et al.* [22], Dantas *et al.* [23], Li *et al.*  
75 [24] have studied the degradation of the antibiotic sulfamethoxazole (SMX) by Fenton,

76 ozonation and electrochemical processes, respectively, and found that total SMX can be  
77 removed by the applied treatments. However, removal of TOC (total organic carbon) is  
78 different for each case: with Fenton reaction 83% of TOC was removed, ozonation promoted  
79 only 10% of TOC removal and electrochemical process achieved 91% of mineralization.  
80 Additionally, Amorim *et al.* [25] studied electrochemical degradation of SMX in the presence  
81 of trimethoprim, being both pollutant completely removed and 90% of TOC reduced.

82 On electrochemical process, organic compounds can be directly oxidized by electron transfer  
83 on the surface of the electrode, or indirectly oxidized by  $\bullet\text{OH}$  or other oxidizing agents. In the  
84 second case, oxidants are electrochemically produced on the anode and can react with organic  
85 compounds at the mediations of the electrode or extend the oxidation to the bulk of the solution.  
86 In most scenarios both oxidation mechanisms simultaneously occur and can be explored  
87 according to the objective of the treatment. Most recently, researchers have been exploring the  
88 coupling of technologies, in order to improve the efficiency of electrochemical treatments.  
89 Electro-Fenton process [21, 26], electrolysis irradiated by UV light [27-29] and  
90 electrochemistry combined to sonochemistry [30-32] can be highlighted, where in most cases  
91 better results are observed for the coupled processes due to a synergistic effect. As a result of  
92 the coupling, more oxidant agents can be produced and, most important, it can be promoted  
93 the activation of such species, which means that highly reactive species can be formed from  
94 poorly reactive oxidants.

95 Sonochemical processes are based on the irradiation of ultrasound through an aqueous medium.  
96 The resultant effect is the so-called cavitation phenomena, which consists on the formation,  
97 expansion and implosion of bubbles. The energy released in this cavitation process is the  
98 principle of sonochemistry, in which the most accepted theory states that at the point of bubble  
99 implosion, conditions of thousands of kelvins and hundreds of atm are produced (hot-spot  
100 theory). This conditions are very short-lived but sufficient to promote the generation of  
101 radicals, such as hydroxyl ( $\bullet\text{OH}$ ), hydrogen ( $\bullet\text{H}$ ) and hydroperoxyl ( $\text{HO}_2\bullet$ ) [33-35]. Coupling  
102 sonochemical to electrochemical technology it is possible to improve the formation of radicals,  
103 activate oxidants and also favor the mass-transfer process [30, 36, 37].

104 However, ultrasound irradiation as a wastewater treatment is still considered an emergent  
105 technology and more need to be known about frequency and power effects, pH, influence of  
106 complex matrices and set-up of the sonolytic system used [38, 39]. According to that, this study  
107 aims to enhance electrolysis by the coupling of ultrasound, but more importantly, it focuses on  
108 the understanding of cavitation effect on electrogenerated oxidant species. For that, the

109 degradation of methyl paraben, as a model molecule, was carried out under the presence of  
110 different supporting electrolytes and different conditions of ultrasound frequency.

111

## 112 **Materials and methods**

### 113 *Chemicals*

114 All solutions were prepared using high-purity water obtained from a Millipore Milli-Q system  
115 (resistivity >18M cm at 25 °C). Methyl paraben and the reactants for determination of oxidants  
116 (KI, Na<sub>2</sub>S<sub>2</sub>O<sub>3</sub>, NaOH, As<sub>2</sub>O<sub>3</sub> and starch solution (1%)) were purchased from Sigma-Aldrich.  
117 Supporting electrolytes Na<sub>2</sub>SO<sub>4</sub> and NaCl were obtained from Panreac, as well as, H<sub>2</sub>SO<sub>4</sub> used  
118 for pH adjustment. Acetonitrile, from VWR Chemicals, was used as mobile phase for HPLC  
119 analysis. All reactants were used as received.

### 120 *Experimental set-up*

121 Experiments were carried out in a bench scale flow system operating in a discontinuous mode.  
122 Electrochemical reactor is coupled to a peristaltic pump (PERCOM-I, JP Selecta, Barcelona,  
123 Spain) and to a jacketed glass vessel, used to store 1 L of solution. The cell is equipped with a  
124 plate of boron doped diamond (BDD) as anode and a plate of stainless steel as cathode (both  
125 with surface area of 75 cm<sup>2</sup>) and a HQ-Power PS3010 power supply (0-30 V, 0-10 A) provided  
126 the applied current. A thermostatic bath (Digiterm 100, JP Selecta) was used to maintain the  
127 solution temperature at 25 °C. Tests using ultrasound were carried out at three different acoustic  
128 frequencies: 20 kHz (UP200S, Hielscher Ultrasonics GmbH), 1 MHz and 10 MHz (EPOCH  
129 650, Olympus). The ultrasound probe was immersed in the glass tank directly acting in the  
130 working solution.

### 131 *Analytical techniques*

132 High performance liquid chromatography (Agilent Technologies 1260 chromatographer) was  
133 used to monitor the concentration of methyl paraben with an Eclipse Plus C18 column (3.5µm  
134 4.6 x 100 mm), at 25 °C, and UV detector set in 280 nm. Mobile phase was acetonitrile and  
135 water (60:40 v/v) with 1 mL min<sup>-1</sup> of flow rate. Total organic carbon (TOC) was determined  
136 in a Multi N/C 3100 Analytik Jena analyzer. Measurements of pH and conductivity were  
137 carried out in a GLP22 CRISON and GLP 31 CRISON, respectively. Oxidizing power of the  
138 solution was determined by iodometric titration. Inorganic anions (Cl<sup>-</sup>, ClO<sub>3</sub><sup>-</sup> and ClO<sub>4</sub><sup>-</sup>) were  
139 measured by ion chromatography using a 930 Compact IC Flex (Metrohm) and hypochlorite  
140 was determined by titration, with a solution of As<sub>2</sub>O<sub>3</sub> in 2.0 M NaOH.

141 *Synergistic index*

142 Synergistic index (SI), **Eq. 1**, was used to evaluate the effect of coupling ultrasound to  
143 electrolysis. Parameters  $k_{p1-p2}$ ,  $k_{p1}$  and  $k_{p2}$  are the kinetic constants for the coupled process and  
144 the individual processes 1 and 2, respectively.

$$145 \quad SI = \frac{k_{p1-p2}}{k_{p1}+k_{p2}} \quad (1.)$$

146

## 147 **Results and discussion**

148 **Fig. 1** shows the removal of MeP obtained in sulfate and chloride medium by the individual  
149 and coupled processes. Sonochemical process (SCh) showed a very low efficiency to the  
150 oxidation of MeP, achieving 26 and 30% of removal in chloride and sulfate media,  
151 respectively. The propagation of ultrasound in a liquid occurs by the vibration of the molecules  
152 in the medium, which produces small variations of pressures and cycles of alternate expansions  
153 and compressions [40]. In the case of water saturated with air, the acoustic cavitation effect  
154 will be produced, creating favored conditions to the formation of radical oxidants [33, 39].

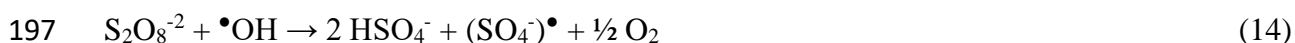
155 Under the studied conditions it is expected the formation of  $\bullet\text{OH}$ ,  $\text{H}_2\text{O}_2$  and  $\text{O}_3$  from water  
156 sonolysis according to the **Eqs. 2-9** (where the symbol  $\bullet$ ) represents the ultrasound waves).  
157 Moreover, in chloride medium, it is also possible the formation of active chlorine species (**Eqs.**  
158 **10-12**) [35, 41]. Hence, the faster MeP removal observed in chloride medium, for SCh process,  
159 can be explained by the higher concentration of oxidants and the rapid production of  
160 organochlorinated compounds (characteristic of oxidation by chlorine species).





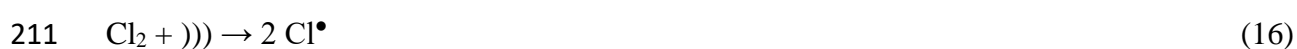
172 Electrochemical processes, individually (ECh) and coupled to sonolysis (SECh), completely  
 173 removed MeP. However, under the applied conditions, it was not possible to observe an  
 174 improvement on the efficiency for the coupled process. Different characteristics for the MeP  
 175 degradation in sulfate and chloride media were expected, as observed for diamond electrolysis  
 176 in a previous work [27]. The main species generated in each system are  $\text{SO}_4^{\bullet-}$  and  $\text{ClO}^-$ , which  
 177 results in different mechanism of oxidation. Sulfate radical has a redox potential comparable  
 178 to that of  $\bullet\text{OH}$  (2.6 and 2.8 V vs. normal hydrogen electrode, NHE, respectively) and its main  
 179 oxidation mechanism is by electron transfer, which promotes the formation of organic radicals  
 180 and its further rupture of bonds, resulting in high elimination of organic matter [42, 43]. On the  
 181 other hand, hypochlorite anions will rapidly convert the pollutant into organochlorinated  
 182 compounds by the addition of Cl atoms to its structure [44].

183 Although no improvement was observed for MeP removal, the SECh process accelerated the  
 184 oxidation of total organic matter, as it is shown in **Fig. 2** and **Tab. 1**. Two kinetic regions of  
 185 *pseudo*-first order can be observed in the case of electrolytic processes, with  $k_1$  and  $k_2$  as the  
 186 kinetic constants for each region. This transition is explained by the nature of the generated  
 187 intermediates, which seem to be easily oxidized by the present oxidant agents than the initial  
 188 pollutant [27]. No removal of TOC was achieved with SCh during the 8 hours of treatment (the  
 189 negative value of kinetic constant is attributed to the dispersion of the points), but complete  
 190 mineralization was attained for ECh and SECh in less than 5 hours. The coupled process  
 191 showed an improvement on the oxidation of organic matter due to the sonochemical activation  
 192 of oxidants. In sulfate medium,  $\text{SO}_4^{\bullet-}$  is produced by the homogeneous cleavage of the  
 193 electrogenerated persulfate (**Eq. 13**), which is more efficient than persulfate itself.  
 194 Additionally, the radicals produced by water sonolysis can also react with persulfate and  
 195 contribute to radical sulfate yield (**Eqs. 14 and 15**) [45, 46].



199 On the other hand, an interesting behavior is observed for chloride medium: in a first moment,  
 200 TOC removal is highly favored. However, after 1 hour of treatment, the removal rate decreases.

201 In order to better understand this result, it is necessary to take into account the nature of  
 202 oxidants species formed and the possible effect of ultrasound on it. It is well known that in an  
 203 electrochemical bath, chloride anions are oxidized to chlorine gas. As it was mentioned, the  
 204 effects of ultrasound are related to the gas dissolved into the water [34, 41, 46]. Higher  
 205 concentration of gases, may improve the cavitation effect, however after a certain time the  
 206 solution will be saturated and this improvement will decrease its significance. In addition to  
 207 this, ultrasound can promote the formation of  $\text{Cl}^\bullet$  and  $\text{Cl}_2^{\bullet-}$  species (**Eq. 16** and **17**), which are  
 208 strong oxidants for many organic compounds [47]. The slower rate in the second kinetic region  
 209 could be attributed to the nature of the intermediates formed, which can be less reactive with  
 210 relation to the oxidants present in the medium.



213 Considering that two regions of kinetics can be observed, it is not possible to determine one  
 214 synergistic index (SI) for all the treatment (**inset of Fig. 2**). In a first moment chloride medium  
 215 is better, with an SI 4.5 times higher than sulfate. On the other hand, for the second region,  
 216 sulfate medium presented an SI 5.6 times higher than chloride. These results are important to  
 217 guess the conditions in which ultrasound may have more influence and this can be an important  
 218 step for future studies, where an optimization of those conditions could be evaluated.

219 It is well-known and described the formation of chlorate and perchlorate in electrolysis of  
 220 chlorine using BDD electrodes (**Eqs. 18-21**) [21, 48, 49]. In addition to the fact that those  
 221 species have a lower oxidant potential, which decrease the process efficiency, they are also  
 222 related to serious health problems and, thus, are not desired to be produced. **Fig. 3** shows the  
 223 speciation of the oxochloranions detect in ECh and SECh processes. Although no formation of  
 224 oxochloranions were verified for sonolysis, it is possible to observe a decrease of the chloride  
 225 concentration, which is attributed to the formation of  $\text{Cl}_2$  and in agreement with the results and  
 226 discussion of **Fig. 1**. The behavior of production / consumption of chlorine species are similar  
 227 for ECh and SCEh; however, it is important to note that the coupling of ultrasound delays the  
 228 formation of perchlorate and that this only starts after all TOC was removed, at 4 hours of  
 229 treatment (**Fig. 2**). Hence, considering the batch system studied, seems feasible the application  
 230 of electrochemical process coupled to ultrasound during 4 hours, minimizing the adverse  
 231 production of chlorate and perchlorate.







236 Considering that the coupling of sonolysis to electrochemical treatment showed faster  
237 mineralization of organic matter with a decrease on the formation of chlorate species, it was  
238 investigated the effect of ultrasound frequency on the system. The frequency applied has an  
239 important role on the critical size of the bubbles and, thus, on the cavitation phenomena. Three  
240 acoustic frequencies were studied: 20 kHz (power ultrasound), with 20% and 75% of power  
241 (conditions referred as L20 and L75), 1 MHz (high frequency ultrasound) (referred as H1) and  
242 10 MHz (diagnostic ultrasound) (referred as H10).

243 **Fig. 4** shows the removal of MeP and TOC obtained for SECh process. Single sonolytic process  
244 resulted in low removal of MeP and none removal of TOC. Negative values observed for 10  
245 MHz (**Fig. 4a**) are attributed to error of the method, meaning that no removal was achieved  
246 under this condition. Lower frequency showed better MeP removals: at 20 kHz approximately  
247 30% was attained for both power percentages, while at 1 MHz only 7% was removed. SECh  
248 treatment seems to slightly enhance ECh results, however it was not possible to observe great  
249 variations of efficiency for different frequencies, as it was expected. Frequency does not seem  
250 to affect the MeP removal (**Fig. 4a**), in which total oxidation is achieved with approximately 4  
251 hours for all electrochemical processes. However, at the second kinetic region of TOC removal  
252 (**Fig. 4b**), it is possible to observe different rates: at L75 and at H1, complete mineralization is  
253 attained with 3 hours, while for other conditions 4 hours were necessary. Moreover, the kinetic  
254 constant calculated for L75 is 1.5 times higher than for H1 (1.42 compared to 0.93 h<sup>-1</sup>).

255 As it was already stated, ultrasound will directly influence the gas dissolved in the working  
256 solution, because of the cavitation phenomena. Cavitation bubbles have three sites in which  
257 chemical reactions can take place: inside the bubble, where temperatures are very high and  
258 thermal reactions can occur; the gas-liquid interface, in which hydrophobic compounds are  
259 more concentrated; and the bulk of the solution [39, 50, 51]. If volatile intermediates are  
260 produced, they can be oxidized inside the bubbles and, thus, the effect of the ultrasound will  
261 be more remarkable. As a matter of fact, phenol is a typical product of the degradation of MeP  
262 [52] and that is the main reason why the differences between applied frequencies can only be  
263 seen in the second region of TOC kinetics.

264 Synergistic index was calculated for each applied frequency in terms of MeP removal and TOC  
265 mineralization (**Fig. 5**). First, it is important to highlight that SI does not consider the time  
266 needed to remove all MeP or organic matter. This means that although SI is higher for L20 and  
267 H10 conditions, at L75 and H1 complete mineralization was achieved 1 hour faster than in the  
268 other conditions. Hence the order of efficiency can be summarized as  $L75 > H1 > L20 > H10$ .  
269 These results indicate that there is an optimal ultrasound frequency in which the degradation is  
270 favored, as it was previously observed in the literature [53-55].

271 At lower frequencies, ultrasound promotes the so-called transient cavitation, in which the  
272 bubbles are very unstable and the collapses are more violent, leading to sites of high  
273 temperature and pressure. At this point, it can be expected high production of radicals by  
274 sonolysis and the activation of oxidants. On the other hand, at high frequencies, the cavitation  
275 bubbles are stable and its collapse is less violent. Besides the lower cavitation intensity, radicals  
276 are still produced inside the bubbles, and its longer stability permits that those radicals escape  
277 to the bulk solution and react with organic compounds [34, 35, 53]. It is estimated that under  
278 transient cavitation, 10% of the radicals escape to the bulk solution; under stable cavitation this  
279 percentage could be improved [34]. Therefore, the optimal frequency will be a balance between  
280 the effects of transient and stable cavitation. It is plausible to consider that under the applied  
281 conditions, the high frequencies effect compensates the low frequencies one and thus, a very  
282 small difference is observed in the final result.

283 As an attempt to clarify the role of ultrasound on the mediated oxidation, sonochemical tests  
284 were carried out in the presence of persulfate at the different frequencies. For that,  $12.5 \text{ g L}^{-1}$   
285 of sodium persulfate were added to the synthetic wastewater studied before ( $100 \text{ mg L}^{-1}$  of  
286 MeP +  $3.0 \text{ g L}^{-1}$  of  $\text{Na}_2\text{SO}_4$ ) and ultrasound was applied for 8 hours, point in which the  
287 concentration of MeP and TOC were measured. **Tab. 2** shows that persulfate itself is capable  
288 to remove approximately 60% of MeP and 35% of organic matter. The irradiation of low  
289 frequency ultrasound enhances the oxidation and almost total mineralization is achieved. In the  
290 absence of ultrasound, persulfate is the only oxidant in the medium; on the other hand, in the  
291 sonolytic process  $\bullet\text{OH}$  is generated and  $\text{SO}_4\bullet$  is produced by the activation of persulfate, as it  
292 was already discussed. However, for 1 MHz and 10 MHz, the efficiency of the process  
293 significantly decreases compared to persulfate alone. This result suggests that, under these  
294 conditions, ultrasound is not only not activating the persulfate, but also inhibiting this oxidant,  
295 which could be explained by **Eqs. 22** and **23** [56]. Water sonolysis occurs even at high  
296 frequencies, thus, hydrogen peroxide is produced from the recombination of radicals (**Eqs. 4**,

297 **6** and **7**) and will react with persulfate. Additionally, even though the main source of sulfate  
298 radical is suppressed (persulfate sonolysis, **Eq. 13**), it may still be produced, at smaller  
299 concentrations, by the reaction of persulfate with  $\bullet\text{OH}$  and  $\bullet\text{H}$  (**Eqs. 14** and **15**); however, they  
300 can also be consumed by  $\text{H}_2\text{O}_2$ .



303 In conclusion, low frequency ultrasound (transient cavitation) activates persulfate, producing  
304  $\text{SO}_4^- \bullet$ , that in addition to  $\bullet\text{OH}$ , from water sonolysis, improve organic matter degradation. On  
305 the other hand, high frequency ultrasound (stable cavitation) is not able to activate persulfate  
306 and also can inhibit its action by the  $\text{H}_2\text{O}_2$  produced from water sonolysis. These is in  
307 agreement with the observations of Ferkous *et al.* [45], which studied the persulfate oxidation  
308 of naphthol blue black enhanced by ultrasound and stated that persulfate-enhanced effect was  
309 more remarkable at 20 kHz than at 585 kHz. In addition, most studies of ultrasound activation  
310 of persulfate only studied low frequency, mainly 20 kHz, obtaining very promising results [57-  
311 60].

312 At this point, it is important to remember that electrochemical cell behaves as a cocktail of  
313 species and not only persulfate and hydroxyl radicals are acting; all species can suffer the  
314 effects of violent and/or stable cavitation in different ways. Irradiation of ultrasound will not  
315 only produce more oxidants, but also influences on the kinetics of all reactions, increasing or  
316 decreasing their kinetic rates. Furthermore, physical effects can take place other than the  
317 chemical ones. For instance, violent cavitation, in transient regime, can also improve the mass-  
318 transport of species, which will improve the electrochemical process. These points reinforce  
319 the theory that the optimal ultrasound frequency to enhance an electrochemical system will be  
320 a balance between the effects of transient and stable cavitation, which represents a combination  
321 of all factors involved. Although it was expected higher improvement for the coupled process  
322 and greater differences between the applied frequencies, the results presented and discussed  
323 give an essential indication of how ultrasound acts in the studied system. It is also important to  
324 take into account that the system studied is simple and MeP is a model molecule, with a  
325 relatively easy oxidation.

326

327 **Conclusions**

328 The coupling of ultrasound to electrolysis was studied for the degradation of methyl paraben  
329 synthetic wastewater, where the influence of supporting electrolyte and acoustic frequencies  
330 were evaluated. From that, the following conclusions can be drawn:

- 331 • The target pollutant was completely mineralized by single electrolysis with diamond anodes  
332 and coupled to ultrasound. However, while removal of MeP is faster in the presence of  
333 chloride, its mineralization is more efficient in sulfate medium, due to the nature of the  
334 oxidant species generated in each condition.
- 335 • The mineralization of the wastewater by single electrolysis and coupled to ultrasound fits  
336 well to *pseudo*-first order kinetics, although two regions can be distinguished. The transition  
337 point between regions and the relation between  $k_1$  and  $k_2$  will depend on the reactivity of  
338 the intermediates formed and the nature of the main oxidant specie in the medium.
- 339 • The combination of electrolysis and sonolysis has a synergistic effect on the abatement of  
340 organic matter, in both chloride and sulfate media, due to the improvement on mediated  
341 oxidation.
- 342 • Ultrasound promotes different effects of cavitation at low and high frequencies. Hence,  
343 there is an optimal frequency for different systems, in which those effects are balanced. It  
344 is possible that, under specific conditions, the high frequencies effect compensates the low  
345 frequencies one and thus, a very small difference is observed in the final result.
- 346 • Sonochemical tests showed that low frequency ultrasound can activate persulfate oxidants  
347 to form radical species, which highly improves the oxidation. However, it also suggests that  
348 high frequencies are not efficient on the activation of persulfate and also inhibit its oxidative  
349 action by converting it into sulfate anions.

350

351 **References**

352 [1] L.E. Dodge, J.W. Choi, K.E. Kelley, S. Hernández-Díaz, R. Hauser, Medications as a potential  
353 source of exposure to parabens in the U.S. population, *Environmental Research*, 164 (2018)  
354 580-584.

355 [2] M.G. Soni, S.L. Taylor, N.A. Greenberg, G.A. Burdock, Evaluation of the health aspects of  
356 methyl paraben: a review of the published literature, *Food and Chemical Toxicology*, 40  
357 (2002) 1335-1373.

358 [3] P.D. Darbre, A. Aljarrah, W.R. Miller, N.G. Coldham, M.J. Sauer, G.S. Pope, Concentrations  
359 of parabens in human breast tumours, *Journal of Applied Toxicology*, 24 (2004) 5-13.

360 [4] Z. Dagher, M. Borgie, J. Magdalou, R. Chahine, H. Greige-Gerges, p-Hydroxybenzoate  
361 esters metabolism in MCF7 breast cancer cells, *Food and Chemical Toxicology*, 50 (2012)  
362 4109-4114.

363 [5] M.G. Soni, I.G. Carabin, G.A. Burdock, Safety assessment of esters of p-hydroxybenzoic  
364 acid (parabens), *Food and Chemical Toxicology*, 43 (2005) 985-1015.

365 [6] J. Boberg, C. Taxvig, S. Christiansen, U. Hass, Possible endocrine disrupting effects of  
366 parabens and their metabolites, *Reproductive Toxicology*, 30 (2010) 301-312.

367 [7] N. Aubert, T. Ameller, J.-J. Legrand, Systemic exposure to parabens: Pharmacokinetics,  
368 tissue distribution, excretion balance and plasma metabolites of [14C]-methyl-, propyl- and  
369 butylparaben in rats after oral, topical or subcutaneous administration, *Food and Chemical*  
370 *Toxicology*, 50 (2012) 445-454.

371 [8] D. Bledzka, J. Gromadzinska, W. Wasowicz, Parabens. From environmental studies to  
372 human health, *Environment International*, 67 (2014) 27-42.

373 [9] J.C. Xue, K. Kannan, Accumulation profiles of parabens and their metabolites in fish, black  
374 bear, and birds, including bald eagles and albatrosses, *Environment International*, 94 (2016)  
375 546-553.

376 [10] K. Nowak, W. Ratajczak-Wrona, M. Górka, E. Jabłońska, Parabens and their effects on  
377 the endocrine system, *Molecular and Cellular Endocrinology*, (2018).

378 [11] T. Velegriaki, E. Hapeshi, D. Fatta-Kassinos, I. Poulios, Solar-induced heterogeneous  
379 photocatalytic degradation of methyl-paraben, *Applied Catalysis B-Environmental*, 178 (2015)  
380 2-11.

381 [12] A.V. Marta-Sanchez, S.S. Caldas, A. Schneider, S.M.V.S. Cardoso, E.G. Primel, Trace  
382 analysis of parabens preservatives in drinking water treatment sludge, treated, and mineral  
383 water samples, *Environmental Science and Pollution Research*, (2018).

384 [13] W.H. Li, Y.L. Shi, L.H. Gao, J.M. Liu, Y.Q. Cai, Occurrence, fate and risk assessment of  
385 parabens and their chlorinated derivatives in an advanced wastewater treatment plant,  
386 *Journal of Hazardous Materials*, 300 (2015) 29-38.

387 [14] J. Chen, B.F.G. Pycke, B.J. Brownawell, C.A. Kinney, E.T. Furlong, D.W. Kolpin, R.U. Halden,  
388 Occurrence, temporal variation, and estrogenic burden of five parabens in sewage sludge  
389 collected across the United States, *Science of the Total Environment*, 593 (2017) 368-374.

390 [15] M. Honda, M. Robinson, K. Kannan, Parabens in human urine from several Asian  
391 countries, Greece, and the United States, *Chemosphere*, 201 (2018) 13-19.

392 [16] J.L. Wang, S.Z. Wang, Removal of pharmaceuticals and personal care products (PPCPs)  
393 from wastewater: A review, *Journal of Environmental Management*, 182 (2016) 620-640.

394 [17] M.A. Quiroz, E.R. Bandala, C.A. Martínez-Huitle, Advanced Oxidation Processes (AOPs)  
395 for Removal of Pesticides from Aqueous Media, in: M. Stoytcheva (Ed.) *Pesticides -*  
396 *Formulations, Effects, Fate*, InTech, Rijeka, 2011, pp. 808.

397 [18] M. Klavarioti, D. Mantzavinos, D. Kassinos, Removal of residual pharmaceuticals from  
398 aqueous systems by advanced oxidation processes, *Environment International*, 35 (2009)  
399 402-417.

400 [19] I. Sires, E. Brillas, M.A. Oturan, M.A. Rodrigo, M. Panizza, Electrochemical advanced  
401 oxidation processes: today and tomorrow. A review, *Environmental Science and Pollution*  
402 *Research*, 21 (2014) 8336-8367.

403 [20] H. Sarkka, A. Bhatnagar, M. Sillanpaa, Recent developments of electro-oxidation in water  
404 treatment - A review, *Journal of Electroanalytical Chemistry*, 754 (2015) 46-56.

405 [21] F.C. Moreira, R.A.R. Boaventura, E. Brillas, V.J.P. Vilar, Electrochemical advanced  
406 oxidation processes: A review on their application to synthetic and real wastewaters, *Applied*  
407 *Catalysis B-Environmental*, 202 (2017) 217-261.

408 [22] S.Z. Wang, J.L. Wang, Comparative study on sulfamethoxazole degradation by Fenton  
409 and Fe(II)-activated persulfate process, *Rsc Advances*, 7 (2017) 48670-48677.

410 [23] R.F. Dantas, S. Contreras, C. Sans, S. Esplugas, Sulfamethoxazole abatement by means of  
411 ozonation, *Journal of Hazardous Materials*, 150 (2008) 790-794.

412 [24] S. Li, D. Bejan, M.S. McDowell, N.J. Bunce, Mixed first and zero order kinetics in the  
413 electrooxidation of sulfamethoxazole at a boron-doped diamond (BDD) anode, *Journal of*  
414 *Applied Electrochemistry*, 38 (2008) 151-159.

415 [25] K.P. de Amorim, L.L. Romualdo, L.S. Andrade, Electrochemical degradation of  
416 sulfamethoxazole and trimethoprim at boron-doped diamond electrode: Performance,  
417 kinetics and reaction pathway, *Separation and Purification Technology*, 120 (2013) 319-327.

418 [26] R.C. Burgos-Castillo, I. Sires, M. Sillanpaa, E. Brillas, Application of electrochemical  
419 advanced oxidation to bisphenol A degradation in water. Effect of sulfate and chloride ions,  
420 *Chemosphere*, 194 (2018) 812-820.

421 [27] D. Dionisio, A.J. Motheo, C. Sáez, M.A. Rodrigo, Effect of the electrolyte on the  
422 electrolysis and photoelectrolysis of synthetic methyl paraben polluted wastewater,  
423 *Separation and Purification Technology*, (2018).

424 [28] J.M. Aquino, D.W. Miwa, M.A. Rodrigo, A.J. Motheo, Treatment of actual effluents  
425 produced in the manufacturing of atrazine by a photo-electrolytic process, *Chemosphere*, 172  
426 (2017) 185-192.

427 [29] S. Hussain, J.R. Steter, S. Gul, A.J. Motheo, Photo-assisted electrochemical degradation  
428 of sulfamethoxazole using a Ti/Ru<sub>0.3</sub>Ti<sub>0.7</sub>O<sub>2</sub> anode: Mechanistic and kinetic features of the  
429 process, *Journal of Environmental Management*, 201 (2017) 153-162.

430 [30] J.R. Steter, D. Dionisio, M.R.V. Lanza, A.J. Motheo, Electrochemical and  
431 sonoelectrochemical processes applied to the degradation of the endocrine disruptor methyl  
432 paraben, *Journal of Applied Electrochemistry*, 44 (2014) 1317-1325.

433 [31] F.L. Souza, C. Saez, P. Canizares, A.J. Motheo, M.A. Rodrigo, Coupling photo and sono  
434 technologies to improve efficiencies in conductive diamond electrochemical oxidation,  
435 *Applied Catalysis B-Environmental*, 144 (2014) 121-128.

436 [32] J.R. Steter, M.H. Kossuga, A.J. Motheo, Mechanistic proposal for the electrochemical and  
437 sonoelectrochemical oxidation of thiram on a boron-doped diamond anode, *Ultrasonics*  
438 *Sonochemistry*, 28 (2016) 21-30.

439 [33] Y.G. Adewuyi, *Sonochemistry: Environmental science and engineering applications*,  
440 *Industrial & Engineering Chemistry Research*, 40 (2001) 4681-4715.

441 [34] L.H. Thompson, L.K. Doraiswamy, *Sonochemistry: Science and engineering*, *Industrial &*  
442 *Engineering Chemistry Research*, 38 (1999) 1215-1249.

443 [35] T.Y. Wu, N. Guo, C.Y. Teh, J.X.W. Hay, Theory and Fundamentals of Ultrasound, Advances  
444 in Ultrasound Technology for Environmental Remediation, Springer Netherlands, Dordrecht,  
445 2013, pp. 5-12.

446 [36] G.S. Garbellini, G.R. Salazar-Banda, L.A. Avaca, Aplicação do ultra-som em sistemas  
447 eletroquímicos: considerações teóricas e experimentais, Química Nova, 31 (2008) 123-133.

448 [37] M.J.M. de Vidales, S. Barba, C. Saez, P. Canizares, M.A. Rodrigo, Coupling ultraviolet light  
449 and ultrasound irradiation with Conductive-Diamond Electrochemical Oxidation for the  
450 removal of progesterone, Electrochimica Acta, 140 (2014) 20-26.

451 [38] M. Gagol, A. Przyjazny, G. Boczkaj, Wastewater treatment by means of advanced  
452 oxidation processes based on cavitation - A review, Chemical Engineering Journal, 338 (2018)  
453 599-627.

454 [39] J. Lifka, B. Ondruschka, J. Hofmann, The Use of Ultrasound for the Degradation of  
455 Pollutants in Water: Aquasonolysis - A Review, Engineering in Life Sciences, 3 (2003) 253-262.

456 [40] A. Weissler, H.W. Cooper, S. Snyder, Chemical effect of ultrasonic waves - oxidation of  
457 potassium iodide solution by carbon tetrachloride, Journal of the American Chemical Society,  
458 72 (1950) 1769-1775.

459 [41] Pankaj, Aqueous Inorganic Sonochemistry, Theoretical and Experimental Sonochemistry  
460 Involving Inorganic Systems, Springer Netherlands, Dordrecht, 2011, pp. 213-271.

461 [42] H. Lutze, Sulfate radical based oxidation in water treatment, Institut für Instrumentelle  
462 Analytische Chemie, Universität Duisburg-Essen, Essen, Germany, 2013, pp. 163.

463 [43] Y. Yang, J. Jiang, X.L. Lu, J. Ma, Y.Z. Liu, Production of Sulfate Radical and Hydroxyl Radical  
464 by Reaction of Ozone with Peroxymonosulfate: A Novel Advanced Oxidation Process,  
465 Environmental Science & Technology, 49 (2015) 7330-7339.

466 [44] M. Mascia, A. Vacca, S. Palmas, Fixed bed reactors with three dimensional electrodes for  
467 electrochemical treatment of waters for disinfection, Chemical Engineering Journal, 211  
468 (2012) 479-487.

469 [45] H. Ferkous, S. Merouani, O. Hamdaoui, C. Pétrier, Persulfate-enhanced sonochemical  
470 degradation of naphthol blue black in water: Evidence of sulfate radical formation, Ultrasonics  
471 Sonochemistry, 34 (2017) 580-587.

472 [46] D. Zhao, X.Y. Liao, X.L. Yan, S.G. Huling, T.Y. Chai, H. Tao, Effect and mechanism of  
473 persulfate activated by different methods for PAHs removal in soil, Journal of Hazardous  
474 Materials, 254 (2013) 228-235.

475 [47] M.L. Alegre, M. Geronés, J.A. Rosso, S.G. Bertolotti, A.M. Braun, D.O. Mártire, M.C.  
476 Gonzalez, Kinetic study of the reactions of chlorine atoms and  $\text{Cl}_2 \bullet^-$  radical anions in  
477 aqueous solutions. 1. Reaction with benzene, The Journal of Physical Chemistry A, 104 (2000)  
478 3117-3125.

479 [48] C.d.N. Brito, D.M. de Araújo, C.A. Martínez-Huitle, M.A. Rodrigo, Understanding active  
480 chlorine species production using boron doped diamond films with lower and higher sp<sup>3</sup>/sp<sup>2</sup>  
481 ratio, Electrochemistry Communications, 55 (2015) 34-38.

482 [49] O. Azizi, D. Hubler, G. Schrader, J. Farrell, B.P. Chaplin, Mechanism of Perchlorate  
483 Formation on Boron-Doped Diamond Film Anodes, Environmental Science & Technology, 45  
484 (2011) 10582-10590.

485 [50] A. DeVisscher, P. VanEenoo, D. Drijvers, H. VanLangenhove, Kinetic model for the  
486 sonochemical degradation of monocyclic aromatic compounds in aqueous solution, Journal  
487 of Physical Chemistry, 100 (1996) 11636-11642.

488 [51] P. Chowdhury, T. Viraraghavan, Sonochemical degradation of chlorinated organic  
489 compounds, phenolic compounds and organic dyes - A review, *Science of the Total*  
490 *Environment*, 407 (2009) 2474-2492.

491 [52] J.R. Steter, R.S. Rocha, D. Dionisio, M.R.V. Lanza, A.J. Motheo, Electrochemical oxidation  
492 route of methyl paraben on a boron-doped diamond anode, *Electrochimica Acta*, 117 (2014)  
493 127-133.

494 [53] K. Kim, E. Cho, B. Thokchom, M. Cui, M. Jang, J. Khim, Synergistic sonoelectrochemical  
495 removal of substituted phenols: Implications of ultrasonic parameters and physicochemical  
496 properties, *Ultrasonics Sonochemistry*, 24 (2015) 172-177.

497 [54] M. Capocelli, E. Joyce, A. Lancia, T.J. Mason, D. Musmarra, M. Prisciandaro, Sonochemical  
498 degradation of estradiols: Incidence of ultrasonic frequency, *Chemical Engineering Journal*,  
499 210 (2012) 9-17.

500 [55] J.W. Kang, H.M. Hung, A. Lin, M.R. Hoffmann, Sonolytic destruction of methyl tert-butyl  
501 ether by ultrasonic irradiation: The role of O<sub>2</sub>, H<sub>2</sub>O<sub>2</sub>, frequency, and power density,  
502 *Environmental Science & Technology*, 33 (1999) 3199-3205.

503 [56] Y. Yang, H.G. Guo, Y.L. Zhang, Q.Z. Deng, J. Zhang, Degradation of Bisphenol A Using  
504 Ozone/Persulfate Process: Kinetics and Mechanism, *Water Air and Soil Pollution*, 227 (2016).

505 [57] Z.S. Wei, F.A. Villamena, L.K. Weavers, Kinetics and Mechanism of Ultrasonic Activation  
506 of Persulfate: An in Situ EPR Spin Trapping Study, *Environmental Science & Technology*, 51  
507 (2017) 3410-3417.

508 [58] C.H. Weng, K.L. Tsai, Ultrasound and heat enhanced persulfate oxidation activated with  
509 Fe<sup>0</sup> aggregate for the decolorization of CI Direct Red 23, *Ultrasonics Sonochemistry*, 29  
510 (2016) 11-18.

511 [59] C. Cai, H. Zhang, X. Zhong, L.W. Hou, Ultrasound enhanced heterogeneous activation of  
512 peroxymonosulfate by a bimetallic Fe-Co/SBA-15 catalyst for the degradation of Orange II in  
513 water, *Journal of Hazardous Materials*, 283 (2015) 70-79.

514 [60] L.W. Hou, H. Zhang, X.F. Xue, Ultrasound enhanced heterogeneous activation of  
515 peroxydisulfate by magnetite catalyst for the degradation of tetracycline in water, *Separation*  
516 *and Purification Technology*, 84 (2012) 147-152.

517