
This is the **accepted version** of the article:

Garcia Orellana, Jordi; Cochran, J. K.; Bokuniewicz, H.; [et al.]. «Evaluation of 224Ra as a tracer for submarine groundwater discharge in Long Island Sound (NY)». *Geochimica et Cosmochimica Acta*, Vol. 141 (September 2014), p. 314-330. DOI 10.1016/j.gca.2014.05.009

This version is available at <https://ddd.uab.cat/record/249149>

under the terms of the  license

1 Evaluation of ^{224}Ra as a tracer for submarine
2 groundwater discharge in
3 Long Island Sound (NY)
4

5 *Garcia-Orellana J.^{1,2,*}, Cochran J.K.³, Daniel J.W.R.³, Rodellas V.^{1,2},*
6 *Bokuniewicz H.³, Heilbrun C.³*
7

8 ¹Departament de Física, Universitat Autònoma de Barcelona,
9 E-08193 Bellaterra, Spain.

10
11 ²Institut de Ciència i Tecnologia Ambientals (ICTA), Universitat Autònoma de Barcelona,
12 E-08193 Bellaterra, Spain.

13
14 ³School of Marine and Atmospheric Sciences, Stony Brook University, Stony Brook, New
15 York, 11794-5000 USA.

16
17
18
19 *Corresponding author. Tel (+34) 935868285, Fax (+34) 935812155.

20 E-Mail address: jordi.garcia@uab.cat
21
22
23
24

25 **ABSTRACT**

26 The approach to quantify submarine groundwater discharge using Ra isotopes
27 generally involves developing a Ra mass balance in an estuary, bay or lagoon. In this work
28 we present a ^{224}Ra mass balance used to evaluate the importance of the submarine
29 groundwater discharge (SGD) in Long Island Sound (NY, US), the third most important
30 estuary in US, located between Long Island and Connecticut, and affected by summertime
31 hypoxia in the western basin. Three surveys were conducted between April 2009 and August
32 2010 where 25 water stations were sampled for Ra isotopes, oxygen and Mn. Stations were
33 oriented along 4 transects: one axial extending from the western to eastern Sound and three
34 longitudinal transects in the western, central and eastern Sound.

35 The inventory of ^{224}Ra in the water column in summer was circa 2 times greater than
36 in winter, suggesting an increased ^{224}Ra flux to the Sound in summer. A mass balance for
37 ^{224}Ra was constructed considering tidal exchange, inputs from rivers, desorption from
38 resuspended particles, diffusive fluxes (including bioirrigation) from bottom sediments and
39 radioactive decay in the water column. Fluxes of ^{224}Ra from bottom sediments were
40 measured by incubating cores under oxic conditions in a continuous flow mode such that the
41 overlying water was circulated through a Mn-oxide fiber to maintain a constant activity of
42 ^{224}Ra . Fluxes from muddy sediments (comprising ~67% of the Sound bottom) ranged from
43 127 to 312 $\text{dpm m}^{-2} \text{d}^{-1}$ and were ~60 $\text{dpm m}^{-2} \cdot \text{d}^{-1}$ in sandy sediments (33% of the Sound).
44 Incubations under hypoxic conditions showed variable fluxes depending on reduction and
45 mobilization of Mn. The ^{224}Ra mass balance shows a net input of Ra to the Sound of $106 \pm$
46 $50 \cdot 10^{12} \text{ dpm y}^{-1}$ in spring and $244 \pm 112 \cdot 10^{12} \text{ dpm y}^{-1}$ in the summer that is attributed to
47 SGD. Elevated ^{224}Ra values were observed near shore and in the pore fluids of the coarse
48 beach sands along the Long Island and Connecticut coasts, suggesting that SGD driven by

49 tidal recirculation through the beach face is a major source of ^{224}Ra to the Sound. Seasonal
50 variation in this source seems unlikely, and the calculated ^{224}Ra SGD fluxes for spring and
51 summer overlap within the uncertainties. Nevertheless we conclude that variations in the
52 ^{224}Ra water column inventories could be produced by seasonal changes in bioirrigation due
53 to the increase of the benthonic productivity in summer and/or redox cycling of Mn as well
54 as sediment resuspension and desorption of ^{224}Ra from resuspended particles, and that our
55 mass balance underestimates these terms, particularly in the summer. ^{224}Ra fluxes from
56 sediments in estuaries, especially those with significant areas of muddy sediments and
57 seasonal hypoxia, are important and should be well constrained in future uses of this isotope
58 as a tracer for SGD.
59

60 **1. INTRODUCTION**

61 The phenomenon of submarine groundwater discharge (SGD) has been shown to be an
62 important component of the hydrological cycle (Moore et al., 2008), but also a major source
63 of nutrients (e.g. Krest et al., 2000), trace metals (e.g. Windom et al., 2006) and radionuclides
64 (e.g. Garcia-Orellana et al., 2013) to the coastal ocean. There are many approaches to
65 quantifying SGD; hydrological models or balances (e.g. Pluhowski and Kantrowitz, 1964),
66 direct measurements via seepage meters (e.g. Cable et al., 1997) and chemical tracers such
67 as ^{222}Rn (e.g. Burnett and Dulaiova, 2003) or Ra isotopes (e.g. Moore, 1996) all have been
68 used. Several models or approaches have been used to determine the amount of SGD by using
69 Ra isotopes in coastal environments such as lagoons, bays, estuaries or open coastal areas:
70 Ra mass balance (Moore, 1996; Krest and Harvey, 2003), Ra endmembers mixing model
71 (Charette and Buesseler, 2004; Garcia-Solsona et al., 2010), eddy diffusion coefficient
72 (Moore, 2000) and modeling (Turner et al., 1997; Robinson et al., 2007; Garcia-Orellana et
73 al., 2010). The approach to quantifying SGD through a mass balance for Ra isotopes requires
74 a detailed evaluation of both sources and sinks of Ra in the system. Sources of Ra include
75 desorption from suspended riverine sediments, regeneration and release from bottom
76 sediments, and input associated with submarine groundwater discharge (Moore, 1997;
77 Hancock et al., 2000; Kelly and Moran, 2002). In locations with extensive marshland, tidal
78 pumping and percolation of water through marsh sediments can also be a significant source
79 of Ra, mainly short-lived isotopes, to the embayment (Bollinger and Moore, 1984; 1993). Ra
80 is lost from the system principally in association with tidal exchange through inlets and by
81 radioactive decay. The latter term is particularly important for the short-lived Ra isotopes
82 (^{223}Ra and ^{224}Ra) and usually less important for the long-lived (^{228}Ra and ^{226}Ra). Previous
83 studies involving mass balances of ^{223}Ra and ^{224}Ra in coastal systems have shown that

84 diffusive inputs and desorptive losses from sediments are often small, and the dominant terms
85 in the Ra balances are supply with SGD, exchange with the open ocean and decay (e.g.,
86 [Charette et al., 2003](#); [Beck et al., 2007, 2008](#); [Garcia-Solsona et al., 2008a](#)). However in
87 estuaries with significant areas of muddy sediments and in which the water column is
88 seasonally hypoxic, the flux of short-lived Ra isotopes from the sediments can be significant.

89

90 In this work we present the Ra mass balance used to evaluate the importance of the submarine
91 groundwater discharge (SGD) in Long Island Sound (NY, US), a major “urban” estuary
92 located in the northeastern USA, and affected by summertime hypoxia in the western basin.

93

94 **2. SETTING**

95 Long Island Sound (LIS) is the third most important estuary in US and it is located between
96 Long Island (NY) and Connecticut (Fig. 1a). The dimensions are 93 km length and 34 km
97 width with $6.2 \cdot 10^{10}$ m³ of water volume and $2.8 \cdot 10^9$ m² of surface sediment. It ranges
98 geographically from the East River in New York City to The Race at its eastern end, with an
99 average depth of 20 m and with a salinity range between 23 and 31. The dominant freshwater
100 input is from the Connecticut River, near the Sound’s eastern end. Tidal ranges in LIS differ
101 from west to east. In the western Sound the range is 2 – 3.5 m, whereas the range is smaller
102 along the eastern shores (0 – 1 m). Tidal currents are strong in LIS, exceeding 1 m s⁻¹ in the
103 East River – a tidal strait connecting western LIS to the lower Hudson River – and range
104 from ~5 m s⁻¹ in the central basin to 1 m s⁻¹ at the eastern end ([Blumberg and Pritchard, 1997](#);
105 [Vieira, 1990](#)). However, mean circulation is fairly weak, approximately 0.1 m s⁻¹ or less
106 ([Vieira, 2000](#)).

107 Historically, bottom water dissolved oxygen (DO) decreases during summer in Long
108 Island Sound (LIS), such that hypoxia ($\text{DO} < 3.0 \text{ mg L}^{-1}$) persists in the East River and
109 western Narrows (Parker and O'Reilly, 1991). This seasonal hypoxia is attributed to the
110 combined effects of phytoplankton and bacteria on biochemical oxygen demand (BOD)
111 coupled with maximum density stratification of the water column (Jensen et al., 1991; Lee
112 and Lwiza, 2008).

113 Seasonal hypoxia in LIS can affect trace element distributions. For example, Mn (IV)
114 is used as an alternate electron acceptor in the bacterial oxidation of organic matter. Its
115 reduction to the more soluble Mn(II) typically takes place in the muddy sediments of LIS,
116 but under hypoxic conditions in the western Sound, the Mn "redoxcline" moves to the
117 sediment water interface or even into the bottom water, resulting in an enhanced flux of
118 dissolved Mn^{2+} into the water column. As this Mn^{2+} mixes and contacts oxic water, it is
119 oxidized to Mn^{4+} , resulting in increased concentrations of particulate Mn. Thus, benthic
120 $\text{Mn}^{2+}(\text{aq})$ fluxes vary seasonally in LIS and are higher in the summer than winter/spring
121 (Aller, 1994).

122 Manganese oxide serves as an effective scavenger of radium. Thus under oxic
123 sediment conditions, zones of Mn oxides in muddy sediments may act as a control on the Ra
124 diffusive flux (Cochran, 1979; Torgersen et al., 1996). Conversely, reduction and
125 solubilization of Mn^{4+} as $\text{Mn}^{2+}(\text{aq})$, may release associated Ra into sediment pore water or
126 overlying water and augment the flux of radium, especially short-lived, from the sediments
127 due to diffusion and bioirrigation.

128 We hypothesize that in coastal environment such as Long Island Sound where the
129 SGD is mainly governed by recirculation of overlying water through the sediments and
130 seasonal hypoxia develops, seasonal variations of the ^{224}Ra inventories in the estuarine water

131 column are related to hypoxia-driven cycling of manganese in the sediments as well as
132 changes in bioirrigation by the benthic fauna. These processes must be evaluated in order to
133 determine the contribution of SGD to the ^{224}Ra balance.

134

135 **3. METHODS**

136 Samples were obtained from stations in Long Island Sound (LIS) aboard the R/V *Seawolf*
137 during spring (24 – 30 April) and summer (29 July – 04 August) 2009, and summer (03 – 12
138 August) 2010. Stations were oriented along 4 transects: one axial extending from the Narrows
139 to the Race and three cross-Sound transects in the western, central and eastern Sound (Fig.
140 1b; Table S1 - Supplemental Data).

141

142 **3.1 Water sampling and procedures**

143 Water samples comprising 60 L of seawater were acquired at two depths – surface and deep
144 – for every station. Surface samples were obtained with a submersible pump at approximately
145 0.5 m; deep samples were taken with Niskin bottles attached to the ship's rosette and tripped
146 ~1 m from the bottom. A CTD was deployed at each station to determine profiles of
147 temperature, salinity, and dissolved oxygen (DO). These water samples were stored in triple-
148 rinsed plastic carboys. The 60 L water samples were subsequently filtered on-board through
149 cartridges containing ~15 g Ra-adsorptive MnO_2 -impregnated acrylic fiber (Mn-fiber), with
150 untreated fiber acting as a pre-filter to eliminate particles. ^{223}Ra and ^{224}Ra were measured
151 using a Delayed-Coincidence Counter (RaDeCC; [Moore and Arnold, 1996](#); [Garcia-Solsona](#)
152 [et al., 2008b](#)). Following counting of the Mn fiber in the RaDeCC, the fiber was leached in
153 HCl and Ra was co-precipitated with BaSO_4 . The BaSO_4 was removed from the leachate by
154 centrifugation and sealed in glass vials for counting on a Canberra Intrinsic Ge well detector.

155 Counting efficiencies for the 352 keV ^{214}Pb (^{226}Ra) and 911 keV ^{228}Ac (^{228}Ra) peaks were
156 determined using the IAEA 300 Baltic Sea Sediment Standard.

157 A second set of surface and deep samples was also obtained for suspended particle
158 matter (SPM), Al(s), Mn(s) and Mn(aq) determination. Surface samples were taken with a
159 Rubbermaid® HDPE 14-liter bucket lowered over the side of the ship; deep samples were
160 taken from Niskin bottles. For SPM, 500 mL of each sample was transferred to an acid-
161 washed Nalgene® 500-mL LDPE bottle and stored in the dark. In the laboratory, samples
162 were vacuum-filtered through pre-weighed Whatman® Nuclepore membrane filters. Filters
163 were dried and re-weighed. After SPM determination, filters were then transferred to acid-
164 washed Falcon® Blue Max™ 15-mL polystyrene conical tubes where 10 mL 6 N trace
165 metal grade (TMG) HCl was added, the samples were vortexed, and leached for ~12 hours.
166 Samples were then centrifuged at 3,500 – 4,000 rpm for ~20 min and supernatant transferred
167 to acid-washed Wheaton® 8-ml HDPE bottles. All Mn(s) and Al(s) analyses were done on a
168 Perkin-Elmer AAnalyst 800 atomic absorption spectrometer equipped with graphite furnace
169 (GFAAS), employing Zeeman background correction. For Mn(aq), 30 mL of each water
170 sample was filtered through a Whatman® Puradisc™ 0.2 µm PES filter (Cat. No. 6780-
171 2502), acidified to 1.2 N with TMG concentrated HCl, and stored in acid-washed Nalgene®
172 30-mL HDPE bottles.

173

174 **3.2 Sediment sampling and procedures**

175 Sediment cores were taken at stations 8, 13, 16, and 120E (Fig. 1b) using a Soutar-type box
176 corer (25 cm × 25 cm). ST13 and ST120E cores were taken on 29 July 2009 and ST8 and
177 ST16 cores were taken on 09 August 2010 and 12 August 2010, respectively. A core also

178 was collected by hand from a shallow, muddy sand station in Stony Brook (West Meadow
179 Beach) on 24 July 2009.

180 For the LIS cores collected in 2010, two box cores were taken at each station. From
181 one box-core, a subcore (9.5 cm diameter) was taken by pushing a butyrate acetate tube into
182 the sediment, two subcores (7.5 cm diameter) were taken in the same manner, and a
183 rectangular subcore was taken for x-radiography. From the other box core, one large subcore
184 (20 cm diameter) was taken for measurement of $^{224,223}\text{Ra}$ fluxes. In 2009, only the large 20-
185 cm diameter subcores were taken from a single box core at each station. All subcores were 10
186 – 20 cm in length.

187

188 **3.2.1 Ra flux** Ra-flux incubations were carried out using the 20 cm-diameter subcores
189 (modified Nalgene[®] polycarbonate multipurpose jars). The cores were covered with between
190 2.6 and 2.9 L of overlying Ra-free water from the bottom water taken at each site. For all
191 cores (2009 and 2010) Ra fluxes were determined under oxic conditions by aerating the
192 overlying water and circulating it continuously *via* a peristaltic pump connected to an in-line
193 cartridge filled with ~15 g of Mn-fiber, following the scheme of [Rodellas et al. \(2012\)](#). At
194 intervals ranging from 0.5 – 3 days, the cartridge was removed and the fiber was analyzed
195 for ^{223}Ra and ^{224}Ra by delayed-coincidence counting methods ('RaDeCC,' [Moore and](#)
196 [Arnold, 1996](#); [Beck et al., 2007](#)). Ra concentrations were determined five times over the first
197 ~250 hours after core collection under these aerated conditions. At the conclusion of the
198 experiment, the cores were sectioned in 5 cm intervals and pore water was separated from
199 the sediment by centrifugation. ^{224}Ra was measured in the pore water samples according to
200 the methods described above.

201 For cores collected in 2010, we attempted to simulate the transition from summer
202 oxic to hypoxic conditions and determine the effect on the ^{224}Ra flux. We began the hypoxic
203 phase by bubbling $\text{N}_2(\text{g})$ into the overlying water of the core, which continued to circulate
204 through Mn fiber and back into the core. However, in order to extract the diffused Ra onto
205 Mn fiber (which is not stable under low oxygen conditions), the water was circulated through
206 a reservoir that aerated it before it passed through the fiber. The water then circulated through
207 a de-aeration reservoir and back into the core. It proved difficult to balance the aeration and
208 de-aeration reservoirs to maintain a constant volume of overlying water in the flux core and
209 to maintain very low dissolved oxygen in the water (values ranged from 2-3 $\text{mg}\cdot\text{L}^{-1}$), and this
210 approach was discontinued after two days. After that point, the flux cores were filled with
211 de-aerated Ra-free LIS bottom water from the respective station, sealed, and the overlying
212 water was continuously circulated with a magnetically coupled stirrer modified from the
213 methods of [Mackin and Swider \(1989\)](#) and [Aller \(1994\)](#). Dissolved oxygen in the overlying
214 water during this period was $<1.5 \text{ mg}\cdot\text{L}^{-1}$. At intervals ranging from 1 – 3 days, the overlying
215 water of the flux core was carefully removed (and immediately replaced with de-aerated Ra-
216 free water for the next flux measurement), aerated, filtered through Mn-fiber, and analyzed
217 for ^{224}Ra by RaDeCC.

218 ^{224}Ra fluxes under oxic conditions were estimated by plotting ^{224}Ra concentrations
219 (dpm L^{-1}) vs time and determining the slope of the best-fit line. The slopes were multiplied
220 by the overlying water volume and divided by the surface area of the polycarbonate
221 containers to obtain fluxes in $\text{dpm m}^{-2} \text{ d}^{-1}$. For the 2010 cores in which we tried to simulate
222 the transition from summer oxic-to-hypoxic conditions, individual fluxes for each time point
223 were calculated using equation (1):

224

225
$$J_{Ra} = \frac{C_{Ra(\Delta t)} \cdot V}{A \cdot \Delta t} \quad (1)$$

226

227 where J_{Ra} = sediment Ra flux (dpm m⁻² d⁻¹), $C_{Ra(\Delta t)}$ = Ra concentration after time Δt (dpm L⁻¹), V = volume of the core overlying water (L), A = surface area of the core (m²), and Δt =
228 interval over which Ra is collected on the Mn fiber in oxic incubations or in which Ra is
229 allowed to accumulate in overlying water during the hypoxic phase (d).
230

231

232 **3.2.2 Mn(aq) flux.** Incubations to determine the flux of Mn(aq) from the sediments were
233 carried out in the 9.5 cm-diameter cores collected at stations 8 and 16 in 2010, adding as
234 overlying water ~700 mL of unfiltered LIS bottom water (collected in a Niskin bottle) from
235 the respective station. Two flux cores from each station were incubated in the dark at an *in*
236 *situ* summer temperature of 21 °C. The overlying water in one subcore was continuously
237 circulated to ensure aeration; the other was sealed and continuously circulated with a
238 magnetically coupled stirrer according to the methods of Mackin and Swider (1989) and Aller
239 (1994). These subsequently will be referred to as “summer oxic” and “summer hypoxic” Mn
240 flux cores respectively. At the close of these initial incubations, the summer hypoxic Mn flux
241 cores from each station were moved to a dark cold-room set to an *in situ* winter temperature
242 of 2.5 °C and allowed to acclimate for 24 hours. After acclimation the seal was broken in
243 order to allow gradual re-oxygenation. These are referred to as “winter oxic” Mn flux cores.

244 For all Mn flux core incubations a time series of multiple overlying water samples
245 were extracted into syringes at 6-hour intervals during, at least, the first 24 hours then at
246 various intervals thereafter based on observed DO. At each sampling time approximately 10
247 mL of overlying water was removed and simultaneously replaced with the aforementioned
248 bottom water. Samples were passed through a Whatman® Puradisc™ 0.2 µm PES filter (Cat.

249 No. 6780-2502) and DO concentrations were determined via a modified Winkler titration
250 method (Mackin et al., 1991). One hour later (after the modified Winkler titration was
251 performed) another 9 ml was extracted and replaced. These samples were also filtered,
252 acidified to 1.09 N with trace-metal grade (TMG) concentrated HCl, and analyzed for
253 dissolved Mn(aq). The dissolved Mn concentrations in the overlying water of the flux cores
254 were corrected for dilution due to replacement of seawater taken for DO measurements
255 (dilution factors were variable in the different cores due to differences in core overlying water
256 volumes). Because samples were passed through a 0.2 μm filter (thus excluding sediment), a
257 correction for particulate Mn contribution to the “dissolved” Mn was not necessary. All Mn
258 analyses were done on by GFAAS. Acidified seawater samples were diluted with de-ionized
259 water and analyzed in triplicate to determine Mn(aq) concentrations (in mmol L^{-1}).

260

261 **3.2.3 Sediment core solid phase analyses.** The 7.5 cm-diameter subcores were sampled to
262 obtain profiles of water content, organic matter content, and total leachable Mn. Cores were
263 sectioned at 0.5 cm intervals to a depth of 3 cm, and at 1 cm intervals from 3 cm to a depth
264 of 6 cm. Sections were dried at 60 °C for 24 h to determine water content. The core sections
265 were then pulverized to a powder with an agate mortar and pestle for the analyses. Sediment
266 (~1.5 g) from each core section was combusted in a muffle furnace at 450°C for 6 h and
267 organic matter content determined by loss-on-ignition. Approximately 150 to 200 mg of dry
268 sediment from each core section was leached with 10 mL 6 N trace metal grade (TMG) HCl
269 for 24 hours in acid-washed Falcon® Blue Max™ 15-mL polystyrene conical tubes. Samples
270 were then centrifuged at 3,500 – 4,000 RPM for ~20 min and the supernatant was transferred
271 to acid-washed Wheaton® 8-mL HDPE bottles. Total leachable Mn and Al were determined
272 via GFAAS.

273 Sediment from the Ra incubation cores was removed by sectioning each core in 1 cm
274 intervals after the flux experiments were completed. The sediment was dried and ground and
275 sealed into jars for measurement of ^{226}Ra , ^{228}Ra and ^{228}Th by gamma spectrometry with a
276 3800 mm² Canberra intrinsic Ge detector, using the gamma emissions at 352 keV (^{214}Pb),
277 911 keV (^{228}Ac) and 583 keV (^{208}Tl), respectively. Counting efficiencies were determined
278 using the IAEA 300 Baltic Sea Sediment Standard and no self-absorption corrections were
279 made due to the relatively high gamma energies.

280

281 **3.2.4 X-radiographs.** Rectangular subcores were stored in LIS water and aerated. Images
282 were taken in the days immediately following coring with a digital x-ray setup.

283

284 **4. RESULTS**

285 **4.1 Temperature, salinity and dissolved oxygen**

286 Figure 4 shows the temperature, salinity and DO profiles in stations along the axial transects
287 in spring (2009) and summer (2010). Table S1 (Supplemental Data) gives temperature,
288 salinity and DO of the water column at the surface and bottom layers during the spring and
289 summer sampling campaigns. During spring 2009, samples showed a gradient of
290 temperatures and salinities from The Race (ST 101) to the East River (ST 16) ranging from
291 a mean temperature of 6.4 and 9.9 °C and from a mean salinity of 30.5 and 25.6, respectively.
292 The distribution of DO in each station along LIS is constant with a mean value of 9.9 mg·L⁻¹,
293 except for ST 16 in the western Sound that was likely influenced by exchange with the East
294 River. Temperature and salinity showed relatively constant profiles in the spring, with a slight
295 decrease of temperature with depth for the eastern stations and an increase of salinity with
296 depth in The Race station (ST101).

297 During summer 2010, profiles showed that the water column was divided into two
298 major layers: a surface layer with high temperature, lower salinity and a deeper layer with
299 lower DO. These two layers were separated by a well-developed pycnocline layer. The
300 surface layer (0-6 m to 0-9 m) was homogenous with a temperature ranging from 22.8 to 23.7
301 °C, salinity between 23.2 and 26.3 and DO from 2.8 to 7.1 mg·L⁻¹. The dense bottom layer
302 occurred from 6-10 m to the bottom, depending on which parameter (T, S or DO) was
303 considered. Temperature showed the thermocline between 6 and 8 m with bottom
304 temperatures ranging from 19.5 and 22.9 °C from east to west. A halocline was evident from
305 6 to 10 m; bottom water salinities ranged between 26.6 and 30.3 with a gradient from west
306 to east. Finally, the oxycline occurred between 8 and 12 m, with concentrations in bottom
307 waters ranging from 2.7 and 6.7 mg·L⁻¹. DO showed a clear decrease in concentration from
308 The Race to East River in agreement with previous studies ([Parker and O'Reilly, 1991](#);
309 [Anderson and Taylor, 2001](#); [Lee and Lwiza, 2008](#))

310

311 **4.2 Ra activities in LIS**

312 ²²⁴Ra was well correlated with ²²³Ra for virtually all samples from LIS (Fig. S1 Supplemental
313 Data); therefore subsequent discussion of the distribution of short-lived Ra in LIS is focused
314 on ²²⁴Ra because data are more complete and precise. ²²⁴Ra activities along the central
315 transect were generally greater in the deep water than surface samples in all seasons (Fig. 3;
316 Table S2 - Supplemental Data). Summer 2009 and 2010 data showed markedly higher ²²⁴Ra
317 activities than spring 2009 with a general pattern of higher ²²⁴Ra in the western Sound and
318 greater activities in deep samples (Fig. 3; Table S2 - Supplemental Data). Mean ²²⁴Ra
319 activities of 10.9 ± 4.2 dpm·100L⁻¹ in summer 2009 and 2010 were more than 2-fold higher
320 than the average concentration of 5.4 ± 1.3 dpm·100L⁻¹ in spring 2009 (Table 1).

321 ^{224}Ra activities along the three cross-Sound transects in the western, central and
322 eastern Sound in spring 2009 and summer 2010 do not show a clear trend in spring 2009, but
323 in summer 2010 showed higher activities in stations close to the Connecticut and Long Island
324 shores (Fig. 4; Table S2 - Supplemental Data), in agreement with the previous results of
325 [Torgersen et al. \(1996\)](#) and [Bokuniewicz et al. \(submitted\)](#). Thus, the inventory of ^{224}Ra in
326 the water column in summer was two times higher than in spring, indicating an increased
327 ^{224}Ra flux to the Sound in summer (Table 3).

328 Although the data are not as extensive as those for ^{224}Ra , measurements of ^{226}Ra and
329 ^{228}Ra activities along the central LIS transect showed seasonal differences, especially for
330 ^{228}Ra , similar to those of ^{224}Ra (Fig. 5; Table S2 - Supplemental Data). Mean activities for
331 ^{226}Ra were $9 \pm 3 \text{ dpm}\cdot 100\text{L}^{-1}$ in spring 2009 and $11 \pm 2 \text{ dpm}\cdot 100\text{L}^{-1}$ for both summer
332 samplings (2009 and 2010). ^{228}Ra showed a larger seasonal difference, with a mean of $42 \pm$
333 $13 \text{ dpm}\cdot 100\text{L}^{-1}$ in spring 2009 and 66 ± 25 and $74 \pm 19 \text{ dpm}\cdot 100\text{L}^{-1}$ for summer 2009 and
334 2010, respectively. These values are comparable to those obtained by [Turekian et al. \(1996\)](#)
335 in summer 1991 and 1993: 16 ± 3 and $14 \pm 2 \text{ dpm}\cdot 100\text{L}^{-1}$ for ^{226}Ra in 1991 and 1993,
336 respectively, and $65 \pm 13 \text{ dpm}\cdot 100\text{L}^{-1}$ for ^{228}Ra in 1991.

337

338 **4.3 Suspended particulate matter (SPM)**

339 Suspended particle concentrations were similar in surface and deep water in spring 2009 with
340 a mean concentration of $1.5 \pm 1.3 \text{ mg}\cdot\text{L}^{-1}$. In summer 2010, SPM concentrations were
341 generally higher (mean $3.3 \pm 2.3 \text{ mg}\cdot\text{L}^{-1}$) than in spring 2009 and had higher concentrations
342 in deep samples (mean $4.7 \pm 2.6 \text{ mg}\cdot\text{L}^{-1}$) and a clear increasing gradient of SPM from east to
343 west in summer 2010 with maximum concentration of $13.5 \text{ mg}\cdot\text{L}^{-1}$ at ST16 (Table S1 -

344 Supplemental Data). SPM concentrations were comparable to the previous results of Kim
345 and Bokuniewicz (1991).

346

347 **4.4 Dissolved and particulate Mn**

348 Dissolved Mn concentrations (Mn_{aq}) were similar in spring 2009 throughout LIS with lower
349 concentrations in the central than in the western Sound and lowest concentrations at the Race
350 (Fig. 6). Summer 2009 Mn_{aq} were highest in deep samples throughout the Sound with the
351 exception of ST16 in the Narrows (Fig. 6). Compared with spring 2009, summer 2009 deep
352 Mn_{aq} concentrations within the western Sound were higher, and surface concentrations
353 throughout LIS were lower, with the exception of ST16 (Fig. 6). Summer 2010 Mn_{aq}
354 concentrations followed the pattern of summer 2009, but showed higher concentrations:
355 western LIS values were 168% higher than spring 2009 values and deep sample
356 concentrations in the central and eastern Sound were 158% higher than surface Mn_{aq} (Fig.
357 6). Overall, the 2010 data showed a general pattern of elevated deep Mn_{aq} concentrations
358 compared with those in surface waters and highest values in the western Sound and along the
359 Sound margins (Fig. 6; Table S1 - Supplemental Data).

360 Particulate Mn (Mn_p) concentrations in spring and summer 2009 showed no clear
361 trends. Summer 2010 data showed a pattern of elevated surface Mn_p concentrations relative
362 to deep with highest values occurring in the central Sound (Fig. 6).

363

364 **4.5 Core incubations**

365 **4.5.1 *Ra* fluxes.**

366 The ^{224}Ra fluxes from the sediment incubations under oxic, well-mixed conditions are
367 summarized in Table 2 (see also Fig. S2- Supplemental Data). Taking the two years as a

368 single set, the ^{224}Ra fluxes under fully aerated, summer temperature conditions displayed the
369 following relationship among the stations: $J_{\text{Ra}}^{\text{ST13}} > J_{\text{Ra}}^{\text{ST120E}} > J_{\text{Ra}}^{\text{ST8}} > J_{\text{Ra}}^{\text{ST16}} > J_{\text{Ra}}^{\text{STWM}}$.

370 In the incubation cores from 2010 (from ST8 and ST16), the two cores displayed
371 differences in Ra flux values, yet similar variations in flux with time as hypoxic conditions
372 were imposed (Fig. 7). DO decreased from $\sim 3 \text{ mg L}^{-1}$ to less than 1.5 mg L^{-1} during the
373 hypoxic phase. At ST8, ^{224}Ra fluxes decreased from the oxic phase flux (average 164 ± 13
374 $\text{dpm m}^{-2} \text{ d}^{-1}$) to $\sim 50 \text{ dpm m}^{-2} \text{ d}^{-1}$ for several days of hypoxic incubation (Fig. 7a). ^{224}Ra flux
375 then increased for the following two incubation periods, first peaking at $\sim 140 \text{ dpm m}^{-2} \text{ d}^{-1}$ for
376 a 1-day incubation then decreasing to $\sim 70 \text{ dpm m}^{-2} \text{ d}^{-1}$ for the final 2-day incubation (Fig.
377 7a).

378 A similar pattern was observed at ST16, although the hypoxic fluxes were higher
379 (Figure 7b). Fluxes under oxic conditions were $127 \pm 12 \text{ dpm m}^{-2} \text{ d}^{-1}$ and, under hypoxic
380 conditions, initially decreased to $\sim 75 \text{ dpm m}^{-2} \text{ d}^{-1}$, $\sim 60\%$ of the oxic flux. Fluxes then
381 increased to a maximum of $214 \text{ dpm m}^{-2} \text{ d}^{-1}$, approximately twice the oxic value and then
382 decreased to $\sim 170 \text{ dpm m}^{-2} \text{ d}^{-1}$ but remained greater than the oxic flux (Fig. 7b).

383

384 **4.6.2 Mn fluxes.** Net dissolved Mn fluxes, J_{Mn} ($\text{mmol m}^{-2} \text{ d}^{-1}$), were estimated according to
385 equation (2), modified from [Aller \(1994\)](#).

386

$$387 \quad J_{\text{Mn}} = \frac{[C_{\text{Mn}}(t) - C_{\text{Mn}}(t-\Delta t)] \cdot V(t)}{A \cdot \Delta t} \quad (2)$$

388

389 where $C_{\text{Mn}}(t)$ = Mn(aq) concentration at time t (mmol L^{-1})

390 $C_{\text{Mn}}(t - \Delta t)$ = Mn(aq) concentration in the previous sample (mmol L^{-1}),

391 $V(t)$ = volume of the core overlying water (L),

392 A = surface area of the core (m^2),

393 Δt = change in time since the previous sample (d).

394 The summer oxic Mn flux cores maintained saturated concentrations of DO and
395 Mn(aq) fluxes were small at both stations: ST8 showed fluxes less than $0.5 \text{ mmol m}^{-2} \text{ d}^{-1}$,
396 after an initial pulse ($1.6 \text{ mmol m}^{-2} \text{ d}^{-1}$, probably related to disturbance associated with core
397 collection) and ST16 had negative fluxes (into the sediments) (Figure 8a and b).

398 The summer hypoxic Mn flux core for ST8 showed fluxes near zero until DO reached
399 a value near 3.0 mg L^{-1} (Fig. 8c) after which the flux increased significantly over ~ 29 hours
400 and then decreased slightly for the final two sampling times (Fig. 8c). As this core was
401 allowed to transition to winter (2°C) oxic conditions, the DO increased rapidly and Mn was
402 removed from the overlying water (Fig. 8c).

403 Simulation of summer hypoxic conditions at ST16 produced hypoxia within ~ 6 hours
404 (Fig. 8) and more rapidly than for ST8 (~ 48 hours, Fig. 8d) – and the core displayed an
405 immediate high Mn flux of $2.9 \text{ mmol m}^{-2} \text{ d}^{-1}$, decreasing gradually thereafter (Fig. 8c).
406 Establishment of winter oxic conditions produced Mn fluxes into the sediment, as at ST8
407 (Fig. 8d).

408

409 **4.6.3 Sediment core physico-chemical analyses.** Water content was almost identical between
410 ST8 and ST16 and indicative of the muddy nature of the sediments (Table S3 - Supplemental
411 Data). ST16 had a slightly greater percentage of organic matter in the top 0.5 cm of sediments
412 than ST8; the same pattern emerged at sediment depths below 3.0 cm (Table S3 -
413 Supplemental Data). Sediment Mn concentrations were more than two-fold higher in the top
414 0.5 cm at ST8 than ST16. Below 1 cm, both subcores exhibited similar Mn concentrations,

415 although concentrations decreased gradually with depth at ST8 and increased gradually with
416 depth at ST16 (Table S3 - Supplemental Data).

417 The solid phase radionuclide measurements (Table S4 - Supplemental Data) showed
418 comparable activities of the ^{232}Th series nuclides at all the sites. The pore water ^{224}Ra
419 activities were ~10 times greater than those in the overlying water and show relatively little
420 variation with depth (Table S5 - Supplemental Data).

421

422 **4.6.4 X-radiographs.** Benthic infaunal organism abundances were higher at ST16 than ST8
423 due to the greater number of burrows and remnant shell (Fig. S3 - Supplemental Data).
424 However, structure and actual abundances may differ because the former can integrate over
425 time.

426

427 **5. DISCUSSION**

428 **5.1 Dissolved oxygen, manganese and ^{224}Ra activities**

429 DO in LIS during spring 2009 was uniform in surface and deep samples except in the
430 Narrows where surface DO concentrations were 14% and 21% higher compared with bottom
431 waters at ST16 and ST13, respectively. During summer 2009 and summer 2010, DO values
432 were not only lower overall compared with spring, but lower in deep samples than in surface
433 samples (Table S1 - Supplemental Data). The disparity between surface and deep DO values
434 during summer can be attributed to stratification of the water column, which occurred in the
435 central Sound (Fig. 2) and caused DO-depletion to be localized below the pycnocline. During
436 summer 2010, the extreme western Narrows (ST16) showed little to no stratification and
437 uniformly hypoxic or near-hypoxic conditions. The absence of stratification there resulted in
438 DO-depletion throughout the water column.

439 Both deep $Mn_{(aq)}$ and deep ^{224}Ra were strongly correlated with DO concentrations in
440 LIS during summer 2010 (Fig. 9a and b). ^{224}Ra concentrations were greatest during summer
441 sampling periods in regions of the LIS with low DO concentrations. In summer 2009 and
442 2010 in stations where hypoxia occurred, Ra showed a general pattern of higher deep
443 concentrations than surface concentrations, especially in the central and eastern Sound (Fig.
444 3). Stations nearest the Connecticut and LI shores showed less difference in surface and deep
445 concentrations, presumably due to shallower water and enhanced input of ^{224}Ra along the
446 Sound margins (Fig. 4; [Torgersen et al., 1996](#)). ST18 and ST16 in the western Narrows
447 exhibited little or no stratification, thus Ra concentrations were high in both surface and deep
448 water (Fig. 3a and b).

449 $Mn_{(aq)}$ concentrations also were highest during summer sampling periods in the
450 western LIS (Fig. 6). Furthermore, a pattern of higher deep $Mn_{(aq)}$ concentrations in the
451 central LIS was evident during summer 2009 and was observed again during summer 2010
452 (Fig. 6). This is in agreement with the seasonal pattern of Mn redox cycling observed by
453 [Aller \(1994\)](#). $Mn_{(aq)}$ concentrations in the western LIS were more uniform throughout the
454 water column presumably due to a greater reductant C_{org} flux and low DO. The relationships
455 among dissolved manganese, ^{224}Ra and DO suggest that the presence of a manganese oxide
456 “redox barrier” in the sediments can have a significant effect on the ^{224}Ra flux to the overlying
457 water. [Sun and Torgersen \(2001\)](#) modeled this effect and showed that ^{224}Ra was effectively
458 scavenged by manganese oxides within the sediments of LIS.

459

460 **5.2 The mass balance of ^{224}Ra in LIS**

461 When SGD is quantified by difference using a Ra mass balance, all other sources of Ra must
462 be accurately determined. Radium inputs to LIS include desorption from riverine sediments

463 entering and Sound and from resuspended sediments, the flux from bottom sediments,
464 exchange with the sea and New York Harbor and the input associated with SGD. On the
465 other hand, radium is lost from the system principally by radioactive decay of the short-lived
466 Ra isotopes (^{223}Ra and ^{224}Ra) and the exchange with low-Ra seawater through inlets.

467

468 ***5.2.1 Seawater exchange***

469 Long Island Sound is connected with the Atlantic Ocean (Block Island Sound) via The Race
470 (ST101). The sample collected in spring 2009 at ST101 showed the lowest ^{224}Ra activity of
471 all the sampling stations, $3.6 \text{ dpm}\cdot 100\text{L}^{-1}$. This value is lower than those reported by
472 [Torgensen et al. \(1996\)](#) in the same area ($^{224}\text{Ra}_{\text{surf}} = 8.9 \text{ dpm}\cdot 100\text{L}^{-1}$ and $^{224}\text{Ra}_{\text{deep}} = 12.9$
473 $\text{dpm}\cdot 100\text{L}^{-1}$) and those reported by [Beck et al. \(2008\)](#) for the mouth of Great South Bay (14.4
474 $\text{dpm}\cdot 100\text{L}^{-1}$, although this value was likely influenced by recirculated bay water).

475 Fluxes of water in and out of LIS through the Race were determined by [Crowley](#)
476 [\(2005\)](#) to be $5.92 \cdot 10^{14}$ and $5.74 \cdot 10^{14} \text{ L}\cdot\text{y}^{-1}$, respectively. Given the ^{224}Ra activity for The
477 Race ($^{224}\text{Ra} = 3.60 \pm 0.22 \text{ dpm}\cdot 100\text{L}^{-1}$) and the average Ra activities in the eastern Sound for
478 spring 2009 ($5.95 \text{ dpm}\cdot 100\text{L}^{-1}$) and summer 2009 and 2010 (7.30 and $7.18 \text{ dpm}\cdot 100\text{L}^{-1}$,
479 respectively), the ^{224}Ra flux into LIS from Block Island Sound is $(21.3 \pm 1.4) \cdot 10^{12} \text{ dpm y}^{-1}$
480 (assumed comparable for both spring and summer) and the fluxes from LIS to Block Island
481 Sound are $(34.2 \pm 8.0) \cdot 10^{12}$ and $(41.6 \pm 16.0) \cdot 10^{12} \text{ dpm y}^{-1}$ for the spring and summer,
482 respectively (Table 4).

483

484 ***5.2.2 East and Connecticut Rivers***

485 The two main rivers that supply freshwater to the LIS are the Connecticut and the East Rivers.
486 Estimated freshwater from the Connecticut River is $1.7 \cdot 10^{13} \text{ L}\cdot\text{y}^{-1}$ ([Dion, 1983](#)). The ^{224}Ra

487 activities of the Connecticut River (ST4) for spring 2009 and summer 2010 were 8 ± 1 and
488 12 ± 4 dpm·100L⁻¹, respectively, yielding ²²⁴Ra fluxes of $(1.4 \pm 0.1) \cdot 10^{12}$ and $(2.0 \pm 0.4) \cdot 10^{12}$
489 dpm y⁻¹ to LIS. Because ST4 is near the mouth of the Connecticut River and its salinity is
490 high enough (~ 28) to guarantee the total desorption of Ra isotopes from suspended particles,
491 these fluxes include both dissolved ²²⁴Ra and desorption of ²²⁴Ra from riverborne particles.

492 The East River serves as the western boundary of LIS and acts as an estuary with
493 exchange between LIS and New York Harbor. The flux of water from the East River to LIS
494 is $4.6 \cdot 10^{13}$ L·y⁻¹ and that from LIS to East River is $6.4 \cdot 10^{13}$ L·y⁻¹ (Robert Wilson, personal
495 communication). The ²²⁴Ra activity determined for the East River is 9 ± 2 dpm·100L⁻¹ (data
496 not shown). Therefore the ²²⁴Ra flux from the East River into LIS is $(4.1 \pm 0.9) \cdot 10^{12}$ dpm·y⁻¹.
497 Mean activities for ²²⁴Ra in the deep water of western LIS are 6.9 ± 1.3 dpm·100L⁻¹ and
498 18.2 ± 4.5 dpm·100L⁻¹ for spring 2009 and summer 2010, respectively. If we assume
499 estuarine circulation associated with the East River-LIS and use these activities to determine
500 the flux of ²²⁴Ra from LIS to NY Harbor via the East River, we obtain $(4.4 \pm 0.8) \cdot 10^{12}$ and
501 $(11.6 \pm 2.9) \cdot 10^{12}$ dpm y⁻¹ for spring 2009 and summer 2010, respectively.

502

503 ***5.2.3 Desorption from resuspended particles***

504 ²²⁴Ra also may be added to the water column through desorption from resuspended bottom
505 sediments. The short half-life of this isotope ensures a rapid production from decay of ²²⁸Th
506 on and in the bottom sediments. Our measurements of ²²⁴Ra in sediment pore water (0-2 cm)
507 after the incubation experiments (Table S5 - Supplemental Data) gave values of ~15 dpm·L⁻¹.
508 The K_d of Ra in the fine-grained sediments of LIS is ~50 L·kg⁻¹ (Cochran, 1979; Sun and
509 Torgersen, 1980). Thus, the adsorbed ²²⁴Ra at the sediment-water interface is ~0.75dpm·g⁻¹.
510 This radium can be desorbed as the surface sediments are resuspended into the overlying

511 water column (with relatively low particle concentrations compared with the solid/pore water
512 ratio in bottom sediments). Resuspension rates for the muddy sediments of LIS can be
513 estimated from our measured suspended sediment concentrations, with the assumption that
514 resuspension is tidally-mediated and thus occurs on a time scale of 0.5 d. The measured
515 suspended sediment concentrations of $1.5 \pm 1.3 \text{ mg}\cdot\text{L}^{-1}$ for spring and $\sim 4 \pm 2.5 \text{ mg}\cdot\text{L}^{-1}$ for
516 summer give standing crops of $3 \pm 2.6 \text{ mg}\cdot\text{cm}^{-2}$ in the spring and $8 \pm 5 \text{ mg}\cdot\text{cm}^{-2}$ in the summer
517 for an average depth of LIS of 20 m, and the resuspension fluxes required to support these
518 standing crops are thus $6 \pm 5.2 \text{ mg}\cdot\text{cm}^{-2}\cdot\text{d}^{-1}$ and $16 \pm 10 \text{ mg}\cdot\text{cm}^{-2}\cdot\text{d}^{-1}$. Sediment fluxes
519 measured in sediment traps in LIS showed a similar seasonal variation (McCall, 1977) but
520 are generally greater than our estimates ($\sim 70 \text{ mg}\cdot\text{cm}^{-2}\cdot\text{d}^{-1}$ in summer; Bokuniewicz et al.,
521 1991; 121 ± 11 and $250 \pm 15 \text{ mg}\cdot\text{cm}^{-2}\cdot\text{d}^{-1}$ in spring and summer, respectively; McCall, 1977).
522 However, the traps in both studies were deployed close to the sediment-water interface—10-
523 30 cm in the case of McCall (1977) and 1 m in the case of Bokuniewicz et al. (1991)—and
524 difference in fluxes in the two studies likely reflects this. We view these values as not
525 representative of the fluxes required to support the observed suspended sediment
526 concentrations in the water column of LIS and thus likely to release desorbed ^{224}Ra through
527 the water column.

528 If the sediment pore water ^{224}Ra is at steady state, such that tidal resuspension of
529 sediment continuously supplies desorbable ^{224}Ra to the water column, and if desorption is
530 rapid and is dominated by the fine-grained sediment characteristic of the western and central
531 basins of LIS ($\sim 67\%$ of the total area), we calculate ^{224}Ra fluxes of $(31 \pm 27)\cdot 10^{12} \text{ dpm}\cdot\text{y}^{-1}$
532 for the spring and $(82 \pm 51)\cdot 10^{12} \text{ dpm}\cdot\text{y}^{-1}$ for the summer (Table 4). These fluxes are
533 maximum estimates because, in addition to the assumptions given above, they assume that
534 all the filterable particles in the water column of LIS are derived from resuspended bottom

535 sediments and so neglect in situ production of biogenic particles. The continual input of Ra
536 derived from desorption from resuspended particles should not be as important for the long-
537 lived radium isotopes (^{228}Ra and ^{226}Ra) because regeneration of desorbable Ra is dependent
538 on their half-lives and is thus slow relative to resuspension driven by the tidal circulation in
539 the Sound.

540

541 **5.2.4 Decay**

542 Radioactive decay at steady state is estimated easily considering the ^{224}Ra inventory in LIS
543 and the decay constant of ^{224}Ra . The inventory is calculated from the average radium
544 activities in the surface and deep water samples in each basin (eastern, central and western)
545 and the volume in the corresponding area (Table 3). During the sampling periods, decay term
546 were $(205 \pm 34) \cdot 10^{12} \text{ dpm} \cdot \text{y}^{-1}$ for spring 2009 and $(447 \pm 88) \cdot 10^{12} \text{ dpm} \cdot \text{y}^{-1}$ for summer 2010
547 (Table 4).

548

549 **5.2.5 Sediment diffusion and bioirrigation**

550 The flux of ^{224}Ra from the sediments of LIS is a potentially important source of Ra to the
551 overlying water. As ^{224}Ra is produced in the sediments from ^{228}Th decay, a portion of the
552 produced Ra atoms are recoiled into the pore water. As noted above, they can adsorb onto
553 particle surfaces and be desorbed during resuspension. Dissolved Ra atoms are also able to
554 diffuse through the pore water and are subject to bioirrigation that facilitates transport into
555 the overlying water column. Although bioirrigation involves fluid flow across the sediment-
556 water interface, this processes is excluded from the definition of SGD (Moore, 2010b) and
557 we treat it separately in constructing the ^{224}Ra balance.

558 Our oxic core incubation data show that the ^{224}Ra flux mediated by diffusion and
559 bioirrigation is readily measurable in the laboratory over short time periods. To incorporate
560 this flux into the Ra mass balance requires estimates of fluxes under both spring and summer
561 conditions. Incubations were run under temperatures closer to summer (20 °C) and with fully
562 oxic overlying water. Under these conditions, the benthic faunal community is active and the
563 Mn redox barrier to Ra diffusion is present within the sediment. Thus these represent summer
564 conditions that might be expected in central and eastern LIS, which generally have an oxic
565 water column during the summer. The cores from central LIS have oxic ^{224}Ra fluxes of 164
566 ± 13 (ST8) and 170 ± 7 (ST120E) $\text{dpm m}^{-2} \text{d}^{-1}$ and 57 ± 3 $\text{dpm m}^{-2} \text{d}^{-1}$ in the sandy mud core
567 from West Meadow (Table 2).

568 Although the temperature was maintained at 20 °C throughout the experiment, the
569 incubation cores from 2010 can be used to obtain some idea of the magnitude of the ^{224}Ra
570 flux from the sediments at other times of the year and under other conditions. Following the
571 oxic incubation, the cores were allowed to become hypoxic. Initially the DO dropped to only
572 $\sim 3 \text{ mg}\cdot\text{L}^{-1}$. Bioirrigation was suppressed, but the data from the Mn flux cores (Fig. 8) suggest
573 that Mn had not yet been reduced and mobilized. The ^{224}Ra flux in both cores decreased to
574 $\sim 50 - 75 \text{ dpm m}^{-2} \text{d}^{-1}$ (Fig. 7). This likely represents the flux of ^{224}Ra without bioirrigation,
575 but with the Mn redox barrier in place. Indeed, a comparable flux ($\sim 50 \text{ dpm m}^{-2} \text{d}^{-1}$) may be
576 calculated using Fick's first law applied to the pore water ^{224}Ra data obtained for the 2009
577 cores at the end of the oxic incubations (Table S5 - Supplemental Data).

578 During the next phase of the hypoxic incubations in the 2010 cores, DO decreased to
579 $\sim 1 \text{ mg}\cdot\text{L}^{-1}$ and the ^{224}Ra flux increased in both cores. We attribute this pattern to the reduction
580 of Mn^{4+} in the sediment and release of associated ^{224}Ra . Fluxes then decreased, but in the

581 core from western LIS (ST16) remained elevated above those observed during the oxic
582 incubation (Fig. 7), likely due to the absence of the Mn redox barrier.

583 We have translated this pattern into a seasonal approximation of the ^{224}Ra flux from
584 LIS sediments as follows. We divide LIS into three sections: western, central and eastern
585 (Fig. 1b). The central and eastern regions experience minimal if any seasonal hypoxia and
586 the principal difference between them is the sediment type, with muddy sediments in the
587 central basin and sandy sediments in the eastern. We assume that the principal seasonal
588 difference in the ^{224}Ra flux in these areas is caused by variation in benthic faunal activity
589 producing bioirrigation (Cai et al., 2013), and that this process tracks water temperature
590 (REF). We take the minimum ^{224}Ra flux observed under mildly hypoxic conditions (~ 50 dpm
591 $\text{m}^{-2} \text{d}^{-1}$) as reflecting the absence of bioirrigation under conditions typical of the winter, and
592 consequently a ^{224}Ra flux dominated by molecular diffusion. In contrast, ^{224}Ra fluxes of ~ 167
593 $\text{dpm m}^{-2} \text{d}^{-1}$ (the average of ST8 and ST120E) and ~ 57 $\text{dpm m}^{-2} \text{d}^{-1}$ (the value for the sandy
594 core STWM), can be taken as representative of conditions of maximum bioirrigation and
595 summer conditions for the central and eastern Sound, respectively. Using the seasonal
596 temperature trend in LIS (Aller, 1977), we fit a 4th-order polynomial to the pattern including
597 the “summer” and “winter” ^{224}Ra fluxes as defined above and determine the appropriate flux
598 for our April and August samplings. For the eastern, sandy portion of LIS we scale the winter
599 fluxes determined for the muddy sediments to that of the sandy core. These fluxes give the
600 sediment Ra fluxes for our spring (April, 2009) and summer (August, 2009 and 2010)
601 samplings: 100 and 160 $\text{dpm m}^{-2} \text{d}^{-1}$ for central LIS and 35 and 55 $\text{dpm m}^{-2} \text{d}^{-1}$ for the eastern
602 sound, respectively.

603 For western LIS, which experiences strong summer hypoxia, we use the approach
604 described above to scale the seasonal change in bioirrigation driven by water temperature,

605 but use the pattern of DO and maximum ^{224}Ra fluxes observed in the incubation experiment
606 with the core from ST16 to include higher short-term ^{224}Ra fluxes associated with low DO
607 and Mn oxide reduction. This allows us to estimate fluxes for April and August for western
608 LIS to use in the ^{224}Ra mass balance: 100 and 220 $\text{dpm m}^{-2} \text{d}^{-1}$, respectively. The area of LIS
609 west of the Race (Fig. 1) is $2.8 \cdot 10^9 \text{ m}^2$, and western, central and eastern sub-basins each
610 comprise approximately 1/3 of the total. The resultant net fluxes of ^{224}Ra from the sediments
611 of LIS (in dpm y^{-1}) are given in Table 4.

612

613 ***5.2.6 ^{224}Ra mass balance model and estimates of ^{224}Ra supplied by submarine groundwater*** 614 ***discharge***

615 Table 4 shows the ^{224}Ra mass balance for spring and summer periods in Long Island
616 Sound. As all the Ra fluxes are determined, the Ra imbalance between inputs and outputs
617 can be ascribed to SGD. Thus, the estimated ^{224}Ra -derived SGD fluxes are $(106 \pm 50) \cdot 10^{12}$
618 and $(244 \pm 112) \cdot 10^{12} \text{ dpm} \cdot \text{y}^{-1}$ for spring and summer conditions, respectively.

619 We have included estimates of errors in the calculation of ^{224}Ra supplied by SGD, but
620 it is difficult to fully assess the uncertainties associated with all of the terms in the mass
621 balance. The most significant terms in the ^{224}Ra balance are the loss by decay in the water
622 column and gains from diffusion and bioirrigation in bottom sediments and desorption during
623 sediment resuspension. Error in the decay term can be estimated as the uncertainties on the
624 average ^{224}Ra inventories in LIS for the spring and summer samplings. These are of the order
625 $\pm 16\text{-}20\%$ (Table 4).

626 The flux from bottom sediments is based on relatively few measurements (4 muddy
627 cores from the western and central Sound, 1 sandy core), and using the standard deviation of
628 the mean seems unjustified. In addition, as described above, we have scaled the fluxes
629 according to temperature in the water for central and eastern LIS, and a composite of

630 temperature and dissolved oxygen for western LIS. We assign an uncertainty of $\pm 30\%$ to the
631 resultant fluxes. The release of ^{224}Ra by desorption from resuspended particles has
632 uncertainties ($\sim \pm 60 - 90\%$) that are based on those associated with the average suspended
633 sediment concentrations. The other terms in the ^{224}Ra mass balance are less important and
634 their errors do not contribute significantly to the ^{224}Ra flux attributed to SGD (Table 4).
635 Although the spring ^{224}Ra fluxes due to SGD are nominally less than those in the summer,
636 the values overlap within the uncertainties.

637

638 *5.2.7 Processes driving the SGD flux*

639 The most accepted definition of SGD was provided by [Burnett et al. \(2003\)](#), who defined
640 SGD as any flow of water out across the sea floor without regard to its composition (e.g.,
641 salinity), its origin, or the mechanism(s) driving the flow. [Taniguchi et al. \(2002\)](#) defined
642 SGD as the addition of two components: SFGD defined as the submarine fresh groundwater
643 discharge and the RSGD defined as the recirculated saline groundwater. Therefore, the
644 estimated total net SGD in our ^{224}Ra balance comprises both net fresh groundwater and
645 recirculated LIS water components.

646 Elevated ^{224}Ra concentrations in pore waters compared with surface (overlying)
647 waters along the LIS shore, as seen in our cross-Sound transects and in shore samples
648 reported by [Bokuniewicz et al. \(submitted\)](#), illustrate the process of Ra input from nearshore
649 sandy sediments ([Boehm et al., 2006](#)). As seawater percolates through the beach face
650 between tidal cycles, ^{224}Ra that accumulates in pore waters is transported to surface waters
651 offshore ([Urish and McKenna, 2004](#)). Subsequent dilution occurs due to wave action and
652 mixing over the course of the tidal cycle. A transect made in Smithtown Bay from the shore
653 out into the open LIS demonstrates the importance of this process for adding ^{224}Ra to LIS

654 (Bokuniewicz et al., submitted). Values decreased from ~ 55 dpm 100L^{-1} nearshore to ~ 15
655 dpm 100L^{-1} at 2 km offshore and decreased further to ~ 5 dpm 100L^{-1} in the open Sound
656 (ST120E). Salinity increased from ~ 24 to 26 over the same transect (Bokuniewicz et al.,
657 submitted).

658 The tidally-mediated flux of water through beach sands is a form of submarine
659 groundwater discharge *sensu* Burnett et al. (2003). The short half-life of ^{224}Ra ensures that it
660 is produced rapidly in the beach sediments and mobilized to the pore water via recoil. Indeed
661 we are able to confirm that production of ^{224}Ra over a tidal cycle can account for the observed
662 activities. We collected a core from Smithtown Bay and isolated it with pore water for several
663 weeks, long enough for ^{224}Ra to reach a steady state between production and decay. The
664 steady state pore water activity was 1100 ± 26 dpm 100L^{-1} . The water entering the sediment
665 on high tide had 60 dpm 100L^{-1} and the water seeping out of the beach face as the tide was
666 receding had 174 dpm 100L^{-1} . Using a simple equation of ingrowth balanced by decay, the
667 time to produce the difference between the incoming and outflowing water is ~ 0.6 d,
668 consistent with a tidal cycle (Bokuniewicz et al., submitted). Thus, the LIS nearshore data
669 demonstrate a marginal input of ^{224}Ra to LIS, especially in areas comprising sandy sediments.
670 This conclusion is in agreement with that of Torgersen et al. (1996), who found that the
671 distribution of surface ^{224}Ra in LIS was a function of cross-Sound (N – S) distance and eddy
672 dispersive mixing.

673 Bokuniewicz et al. (submitted) modeled the distribution of ^{224}Ra along a transect
674 extending from 1 m water depth nearshore to central LIS. The ^{224}Ra flux per meter of
675 shoreline required to support the distribution was $1.03 \cdot 10^8$ dpm $\text{m}^{-1} \text{y}^{-1}$. Extrapolating to the
676 shoreline lengths of the Long Island and Connecticut shores of LIS (Bokuniewicz et al.,
677 submitted) yields $\sim 98 - 208 \cdot 10^{12}$ dpm y^{-1} . This is quite comparable to the ^{224}Ra flux due to

678 SGD calculated from the ^{224}Ra mass balance (Table 4). We conclude that tidal percolation
679 through the coarse-grained sediment along the Long Island shore can supply the ^{224}Ra
680 attributed to SGD in the ^{224}Ra balance.

681 If the SGD flux of ^{224}Ra into LIS is caused principally by tidal percolation through
682 coarse-grained sands along the shores of the Sound, there is no clear reason for this process
683 to be seasonal and to be greater in the summer than in the spring, although other studies have
684 also reported seasonal differences between winter and summer (e.g. [Moore, 2010](#)). Thus,
685 seasonal differences in the ^{224}Ra balance, if real, must be attributed to other factors. The
686 supply of ^{224}Ra from bottom sediments represents the second largest source of this
687 radionuclide to the LIS water column. Our incubation experiments suggest that bioirrigation
688 and Mn redox cycling, both of which vary seasonally in LIS, are important in controlling the
689 ^{224}Ra flux from the bottom. It seems likely that our estimates of the ^{224}Ra flux from bottom
690 sediments do not adequately represent spatial or seasonal variations in that source of ^{224}Ra to
691 the overlying LIS water column, especially under summer hypoxic conditions. Moreover, the
692 flux of ^{224}Ra during the summer as hypoxia develops is likely to be temporally variable,
693 reflecting both the reduction of manganese oxides and release of associated Ra as well as
694 changes in the flux from the bottom with reduced bioirrigation and the absence of a
695 manganese redox barrier in the sediments. This implies that accurate balances for the short-
696 lived Ra isotopes must take into account the importance of muddy sediments, especially in
697 estuaries that experience seasonal hypoxia.

698 The ^{224}Ra SGD fluxes of $106 - 244 \cdot 10^{12}$ dpm y^{-1} (Table 4) can be converted to a
699 SGD flow by considering the ^{224}Ra activities of the water that is transferred out the coarse
700 sediment due to SGD. Analyses of beach pore water reported in [Bokuniewicz et al.](#)
701 [\(submitted\)](#) range from 120 to 680 dpm 100L^{-1} with a mean of 330 dpm 100L^{-1} . Therefore

702 the volumetric flux of SGD is $32 - 74 \cdot 10^{12} \text{ L y}^{-1}$. equivalent to 1.3 – 3.5 times the flux of
703 fresh water delivered to LIS by the Connecticut River. This comparison agrees with other
704 studies conducted in the western shore of the North Atlantic Ocean (e.g. [Moore, 2010](#)) or the
705 entire Atlantic Ocean ([Moore et al., 2008](#)), that concluded that the total SGD flux is up to ~3
706 times greater than the river flux. Compared with the fresh groundwater underflow for Nassau
707 and Suffolk counties estimated at $3.45 \cdot 10^{11} \text{ L} \cdot \text{y}^{-1}$ ([Monti and Scorca, 2003](#)), the SGD for LIS
708 calculated from the ^{224}Ra balance is ~100 times that of the fresh SGD. In other words,
709 recirculated seawater accounts for ~100% of the total SGD flow into LIS.

710

711 **5.4 Comparison of long-lived radionuclides (^{226}Ra and ^{228}Ra) with ^{224}Ra in LIS**

712 Our data for ^{226}Ra and ^{228}Ra in LIS are not detailed enough to permit mass balances
713 to be constructed for those Ra isotopes. However, samples from the central transect of LIS
714 taken in spring 2009 and summer 2010 show a seasonal difference with higher activities,
715 especially for ^{228}Ra , in summer (Fig. 5). The activities of the parent isotopes ^{230}Th and ^{232}Th ,
716 are approximately equal in LIS sediments ([Cochran, 1979](#)) and thus the production rate of
717 ^{226}Ra in the muddy sediment of central and western LIS and in the coarse-grained beach
718 sediment characteristic of the Long Island's north shore and the Connecticut coast is slow
719 compared with the other Ra isotopes. Pore water ^{226}Ra concentrations are accordingly low
720 ([Cochran 1979, 1985](#)). As a consequence, seasonal changes in ^{226}Ra in the LIS water column
721 are likely to be related to releases of ^{226}Ra associated with manganese oxides in the sediments
722 and thus varying with dissolved manganese.

723 In contrast, there are significant fluxes of ^{228}Ra from the sediments of LIS, with long-
724 term values of ~35 to 80 dpm $\text{m}^{-2} \text{ d}^{-1}$ based on the ^{228}Ra deficiency in muddy sediment cores
725 from LIS ([Cochran, 1979, 1985; Turekian et al. 1996](#)). Although the production rate of ^{228}Ra

726 is also slower than that of ^{224}Ra in the sediments of LIS, the flux of ^{228}Ra is likely to be
727 dependent on the intensity of bioirrigation and the redox cycle of manganese and thus
728 seasonally variable as is that for ^{224}Ra . Moreover, because the ^{228}Ra production rate is
729 significantly lower than that of ^{224}Ra , it also seems likely that its flux from tidal percolation
730 through coarse-grained sediments along shore will be less. Thus a detailed ^{228}Ra mass
731 balance in LIS may permit other terms, such as the flux from sediments and input of fresh
732 SGD to be better determined.

733

734 **6. Summary**

735 The results presented in this work demonstrate the difficulty in determining SGD
736 fluxes using the short-lived Ra isotopes in an estuary such as Long Island Sound (LIS), in
737 which manganese redox cycling is active due to seasonal development of hypoxia and there
738 is a flux of Ra from bottom sediments, mediated by diffusion and bioirrigation. Based on a
739 spatial survey of ^{224}Ra at ~25 stations in LIS conducted in spring 2009 and summer 2009 and
740 2010, the water column inventories of ^{224}Ra in LIS are higher in bottom waters than in surface
741 waters and, for the whole water column, are a factor of ~2 greater in the summer than in the
742 spring. These differences are likely due to variation in the flux of ^{224}Ra from bottom
743 sediments as a result of seasonal changes in bioirrigation and/or redox cycling of Mn. Season
744 variation in the rate of resuspension of bottom sediments and desorption of ^{224}Ra from this
745 material also may be a factor.

746 The ^{224}Ra flux from bottom sediments, measured in oxic core incubation experiments,
747 ranges from 127 – 312 $\text{dpm m}^{-2} \text{d}^{-1}$ in the muddy sediments of LIS and is ~60 $\text{dpm m}^{-2} \text{d}^{-1}$ in
748 sandy mud. Imposing hypoxic conditions on the flux cores produces temporally variable
749 fluxes that initially decrease from the oxic value but then increase and can exceed it. Inclusion

750 of the flux of ^{224}Ra from bottom sediments and from desorption from resuspended sediments
751 in a Ra mass balance shows a net input to LIS that we attribute to submarine groundwater
752 discharge. Higher activities of ^{224}Ra in stations taken nearshore and in pore waters of the
753 coarse beach sands along the Long Island and Connecticut shores of LIS (Bokuniewicz et al.,
754 submitted) suggest that tidal percolation of water through the beach face is responsible for
755 most of the input of ^{224}Ra from SGD to LIS. Estimates of the magnitude of this flux by
756 Bokuniewicz et al. (submitted) show it to be sufficient to account for the inferred SGD term
757 in the ^{224}Ra balance of LIS.

758 Our results suggest that processes such as bioirrigation and diffusion in bottom
759 sediments provide an important source of ^{224}Ra to the overlying water column. In estuarine
760 systems that experience seasonal hypoxia, the effects of low dissolved oxygen on the benthic
761 community and on the redox cycling of elements such as Mn can moderate the flux of ^{224}Ra
762 from the sediments. Attributing excesses of ^{224}Ra in the water column in such systems solely
763 to SGD is not justified, and Ra mass balances must take into account other sources, such as
764 bottom sediments, for an accurate estimation of SGD.

765

766 **7. Acknowledgments**

767 This project has been funded by New York Sea Grant (Project R/CCP-16) and the Ministerio
768 de Ciencia e Innovación (CTM2010-11232-E) and we are grateful for their support. Support
769 from the government of Spain and the Fulbright Commission for a post-doctoral fellowship
770 to J.G-O. (ref 2007-0516) is gratefully acknowledged. Portions of this work comprise the
771 Masters thesis of JWRD. We thank David Bowman and the crews of the R/V *Seawolf* for
772 their assistance in sampling, and Professor Robert Aller for helpful discussions.

773

774

775 **8. References**

776

777 Aller, R.C., 1977. Diagenetic processes near the sediment-water interface of Long Island
778 Sound. II. Fe and Mn. *Advances in Geophysics* 22, 351–415.

779 Aller, R.C., 1994. The sedimentary Mn cycle in Long Island Sound: Its role as intermediate
780 oxidant and the influence of bioturbation, O₂, and C_{org} flux on diagenetic reaction
781 balances. *Journal of Marine Research* 52, 259–295.

782 Anderson, T.H., Taylor, G.T., 2001. Nutrient pulses, plankton blooms, and seasonal hypoxia
783 in western Long Island Sound. *Estuaries* 24, 228–243.

784 Beck, A.J., Rapaglia, J.P., Cochran, J.K., Bokuniewicz, H.J., 2007. Radium mass-balance in
785 Jamaica Bay, NY: Evidence for a substantial flux of submarine groundwater. *Marine*
786 *Chemistry* 106, 419–441.

787 Beck, A.J., Rapaglia, J.P., Cochran, J.K., Bokuniewicz, H.J., Yang, S., 2008. Submarine
788 groundwater discharge to Great South Bay, NY, estimated using Ra isotopes. *Marine*
789 *Chemistry* 109, 279–291.

790 Benninger, L.K., 1976. The uranium-series radionuclides as tracers of geochemical processes
791 in Long Island Sound. Ph.D. Dissertation, Yale University. New Haven, CT.

792 Blumberg, A.F., Pritchard D.W., 1997. Estimates of the transport through the East River,
793 New York. *Journal of Geophysical Research* 102, 5685–5703.

794 Boehm, A.B., Paytan, A., Shellenbarger, G.G., Davis, K.A., 2006. Composition and flux of
795 groundwater from a California beach aquifer: Implications for nutrient supply to the surf
796 zone. *Continental Shelf Research* 26, 269–282.

797 Bokuniewicz H., McTiernan, L., Davis, W. 1991. Measurement of sediment resuspension
798 rates in Long Island Sound. *Geo-Marine Letters* 11, 159-161.

799 Bollinger, M.S., Moore, W.S., 1984. Radium fluxes from a salt marsh. *Nature* 309, 444 - 446.

800 Bollinger, M.S., Moore, W.S., 1993. Evaluation of salt marsh hydrology using radium as a
801 tracer. *Geochimica and Cosmochimica Acta* 57, 2203–2212.

802 Burnett, W.C., Bokuniewicz, H., Huettel, M., Moore, W.S., Taniguchi, M., 2003.
803 Groundwater and porewater inputs to the coastal zone. *Biogeochemistry* 66, 3-33.

804 Burnett, W.C., Dulaiova, H., 2003. Estimating the dynamics of groundwater input into the
805 coastal zone via continuous radon-222 measurements. *Journal of Environmental*
806 *Radioactivity* 69, 21–35.

807 Cable, J.E., Burnett, W.C., Chanton, J.P., Weatherly, G.L. 1996. Estimating groundwater
808 discharge into the northeastern Gulf of Mexico using radon-222. *Earth and Planetary*
809 *Science Letters* 144, 591–604.

810 Cai, P., Peng, S., Moore, W.S., Shi, X., Wang, G., Dai, M. 2013. ^{224}Ra : ^{228}Th disequilibrium
811 in coastal sediments: Implications for solute transfer across the sediment-water interface.
812 *Geochemica and Cosmochimica Acta*. Accepted Manuscript.

813 Charette, M.A., Splivallo, R., Herbold, C., Bollinger, M.S., Moore, W.S. 2003. Salt marsh
814 submarine groundwater discharge as traced by radium isotopes. *Marine Chemistry* 84,
815 113–121.

816 Charette M.A., Buesseler, K.O. 2004. Submarine groundwater discharge of nutrients and
817 copper to an urban subestuary of Chesapeake bay (Elizabeth River). *Limnology and*
818 *Oceanography* 49, 376–385.

819 Cochran, J.K. 1979. The Geochemistry of ^{226}Ra and ^{228}Ra in Marine Deposits, Ph.D. Thesis,
820 Yale University, New Haven, CT.

821 Dion, E.P. 1983. Trace elements and radionuclides in the Connecticut and Amazon river
822 estuary, Ph.D. Thesis, 233 pp., Yale Univ., New Haven, CT.

823 Garcia-Orellana, J., Cochran, J.K., Bokuniewicz, H., Yang, S., Beck, A.J. 2010. Time-series
824 sampling of ^{223}Ra and ^{224}Ra at the inlet to Great South Bay (New York): a strategy for
825 characterizing the dominant terms in the Ra budget of the bay. *Journal of Environmental*
826 *Radioactivity* 101, 582 - 588.

827 Garcia-Orellana, J., Rodellas, V., Casacuberta, N., López-Castillo, E., Vilarrasa, M., Moreno,
828 V., Garcia-Solsona, E., Masqué, P. 2013. Submarine Groundwater Discharge as a source
829 of natural radioactivity. *Marine Chemistry* 156, 61-72.

830 Garcia-Solsona, E., Masqué, P., Garcia-Orellana, J., Rapaglia, J., Beck, A.J., Cochran, J.K.,
831 Bokuniewicz, H.J., Zaggia, L., Collavini, F., 2008a. Estimating submarine groundwater
832 discharge around Isola La Cura, northern Venice Lagoon (Italy), by using the radium
833 quartet. *Marine Chemistry* 109, 292 - 306.

834 Garcia-Solsona, E., Garcia-Orellana, J., Masqué, P., Dulaiova, H. 2008b. Uncertainties
835 associated with ^{223}Ra and ^{224}Ra measurements in water via a delayed coincidence counter
836 (RaDeCC). *Marine Chemistry* 109, 198 - 219.

837 Garcia-Solsona, E., Garcia-Orellana, J., Masque, P., Garces, E., Radakovitch, O., Mayer, A.,
838 Estrade, S., Basterretxea, G., 2010. An assessment of karstic submarine groundwa- ter and
839 associated nutrient discharge to a Mediterranean coastal area (Balearic Islands, Spain)
840 using radium isotopes. *Biogeochemistry* 97 (2–3), 211–229.

841 Hancock, G.J., Webster, I.T., Ford P.W., Moore, W.S. 2000. Using Ra isotopes to examine
842 transport processes controlling benthic fluxes into a shallow estuarine lagoon. *Geochimica*
843 *and Cosmochimica Acta* 64, 3685–3699.

844 Jensen, L.M., Sand-Jensen, K., Marcher, S., Hansen, M. 1990. Plankton community
845 respiration along a nutrient gradient in a shallow Danish estuary. *Marine Ecology Progress*
846 *Series* 61, 75–85.

847 Kelly, R.P., Moran S.B. 2002. Seasonal changes in groundwater input to a well-mixed
848 estuary estimated using radium isotopes and implications for coastal nutrient budgets.
849 *Limnology and Oceanography* 47, 1796–1807.

850 Kim, B.H., and Bokuniewicz, H.J., 1991, Estimates of sediment fluxes in Long Island Sound.
851 *Estuaries* 14, 237-247.

852 Krest, J.M., Moore, W.S., Gardner, L.R., Morris, J.T. 2000. Marsh nutrient export supplied
853 by groundwater discharge: Evidence from radium measurements. *Global Biogeochemical*
854 *Cycles* 14, 167–76.

855 Krest J.M., Harvey J.W. 2003. Using natural distributions of short-lived radium isotopes to
856 quantify groundwater discharge and recharge. *Limnology and Oceanography* 48, 290–8.

857 Lee, Y.J., Lwiza, K.M.M. 2005. Interannual variability of temperature and salinity in shallow
858 water: Long Island Sound, New York. *Journal of Geophysical Research* 110, C09022.

859 Mackin, J., Aller, R., Vigil, H., Rude, P. 1991. Nutrient and dissolved oxygen fluxes across
860 the sediment-water interface. In *Long Island Sound Study Final Report, Sediment*
861 *Geochemistry and Biology*, US-EPA Contract CE 002870026. Section IV: 1–252.

862 Mackin, J.E., Swider, K.T. 1989. Organic matter decomposition pathways and oxygen
863 consumption in coastal marine sediments. *Journal of Marine Research* 47, 681–716.

864 McCall, P.L. 1977. Community patterns and adaptive strategies of the infaunal benthos of
865 Long Island Sound. *Journal of Marine Research* 35, 221-266

866 Moore, W.S., Arnold R. 1996. Measurement of ^{223}Ra and ^{224}Ra in coastal waters using a
867 delayed coincidence counter. *Journal of Geophysical Research* 101, 1321–1329.

868 Moore, W.S. 1997. High fluxes of radium and barium from the mouth of the Ganges-
869 Brahmaputra River during low river discharge suggest a large groundwater source. *Earth
870 and Planetary Science Letters* 150, 141-150.

871 Moore, W.S. 2000. Determining coastal mixing rates using radium isotopes. *Continental
872 Shelf Research* 20, 1993-2007.

873 Moore, W.S. 2003. Sources and fluxes of submarine groundwater discharge delineated by
874 radium isotopes. *Biogeochemistry* 66, 75–93.

875 Moore, W.S. 2010. The Effect of Submarine Groundwater Discharge on the Ocean. *Annual
876 Review of Marine Science* 2, 59–88.

877 Moore, W.S., Sarmiento, J.L., Key, R.M. 2008. Submarine groundwater discharge revealed
878 by ^{228}Ra distribution in the upper Atlantic Ocean. *Nature Geosciences* 1, 309–11.

879 Parker, C.A., O'Reilly, J.E. 1991. Oxygen depletion in Long Island Sound: a historical
880 perspective. *Estuaries* 14, 248–264.

881 Pluhowski, E.J., Kantrowitz, I.H. Hydrology of the Babylon–Islip Area, Suffolk County,
882 Long Island, New York. 1964. US Geological Survey Water Supply Paper 1768, 128 pp.

883 Robinson, C., Li L., Barry, D.A. 2007. Effect of tidal forcing on a subterranean estuary.
884 *Advances in Water Resources*. 30, 851- 865.

885 Rodellas, V., Garcia-Orellana, J., Garcia-Solsona, E., Masqué, P., Domínguez, J.A.,
886 Ballesteros, B.J., Mejías, M., Zarroca, M., 2012. Quantifying groundwater discharge from
887 different sources into a Mediterranean wetland by using ^{222}Rn and Ra isotopes. *Journal of
888 Hydrology* 466–467, 11–22.

889 Torgersen, T., Turekian, K.K., Turekian, V.C., Tanaka, N., DeAngelo, E., O'Donnell, J.
890 1996. ^{224}Ra distribution in surface and deep water of Long Island Sound: sources and
891 horizontal transport rates. *Continental Shelf Research* 16, 1545–1559.

892 Turner, I.L., Coates, B.P., Acworth R.I. 1997. Tides, waves, and the super-elevation of
893 groundwater at the coast. *Journal of Coastal Research* 13, 46-60.

894 Urish, D.W., McKenna, T.E. 2004. Tidal effects on ground water discharge through a sandy
895 marine beach. *Ground Water – Oceans* 42, 971–982.

- 896 Vieira, M.E.C. 1990. Observations of Currents, Temperature and Salinity in Long Island
897 Sound, 1988. A Data Report. Special Data Report #6, Marine Sciences Research Center,
898 State University of New York. Stony Brook, New York, USA.
- 899 Vieira, M.E.C. 2000. The long-term residual circulation in Long Island Sound. *Estuaries* 23,
900 199–207.
- 901 Windom, H.L., Moore, W.S., Niencheski, L.F.H., Jahnke, R.A. 2006. Submarine
902 groundwater discharge: a large, previously unrecognized source of dissolved iron to the
903 South Atlantic Ocean. *Marine Chemistry* 102, 252–266.
- 904

905 **TABLES**

906 Table 1. Mean ^{224}Ra concentrations in the different basins of Long Island Sound in spring
907 2009 and summer 2009 and 2010.

908 Table 2. ^{224}Ra fluxes from LIS sediment cores incubated under oxic, well-mixed conditions
909 and hypoxic conditions.

910 Table 3. ^{224}Ra inventories in western, central and eastern LIS in spring 2009 and summer
911 2010.

912 Table 4. ^{224}Ra mass balance in Long Island Sound.

913

914 **Supplemental Data - Tables**

915 Table S1. Location and physico-chemical data (T^a , S, DO, SPM, Mn_{aq} , Mn_s) for LIS
916 campaigns.

917 Table S2. Ra activities for LIS sampling campaigns in 2009 and 2010.

918 Table S3. Water and organic content and solid phase manganese concentrations in LIS
919 sediment cores

920 Table S4. Solid phase U/Th series activities from incubation cores sampled after
921 completion of the incubation experiments.

922 Table S5. Pore water ^{224}Ra and ^{223}Ra from incubation cores sampled after completion of the
923 incubation experiments.

924

925 **FIGURES**

926 Figure 1: (a) Map of Long Island Sound showing sub-basins and major geographic features,
927 (b) Sampling stations in Long Island Sound. Stations were oriented along 4 transects: one
928 axial from the Narrows to the Race and three cross-Sound transects in the delimited
929 western, central and eastern Sound.

930 Figure 2. Temperature, salinity and DO profiles in stations along the axial transects in
931 spring (2009) and summer (2010).

932 Figure 3. ^{224}Ra concentrations in surface and deep waters in the campaigns carried out in
933 April 2009, July 2010 and August 2010.

934 Figure 4. Surface and deep ^{224}Ra during spring 2009 and summer 2010 concentrations
935 along the three longitudinal transects.

936 Figure 5. $^{228,226}\text{Ra}$ activities in samples collected along the central transect of Long Island
937 Sound in 2009 and 2010.

938 Figure 6. Dissolved (Mn_{aq}) and particulate Mn (Mn_{p}) from surface and deep LIS samples.

939 Figure 7. ^{224}Ra fluxes ($\text{dpm}\cdot\text{m}^{-2}$) over time for sediment cores collected in 2010 and
940 incubated initially in an oxic mode then allowed to become hypoxic.

941 Figure 8. Mn(aq) fluxes and accompanying DO concentrations over time for summer oxic
942 Mn flux cores for ST8 (a) and ST16 (b) and for summer anoxic to anoxic winter oxic Mn
943 flux for ST8 (c) and ST16 (d). Core temperatures for summer and winter phases were 21°C
944 and 2.5°C , respectively.

945 Figure 9. a) LIS Summer 2010 surface and deep Mn(aq) concentrations plotted against
946 corresponding DO concentrations. b) LIS summer 2010 surface and deep ^{224}Ra
947 concentrations plotted against corresponding DO concentrations. A correlation between
948 deep Mn(aq) - ^{224}Ra and DO values is observed.

949 **Supplemental Figures**

950 Figure S1. Correlation between ^{224}Ra and ^{223}Ra showing an identical distribution of both
951 radionuclides in LIS.

952 Figure S2. ^{224}Ra accumulated in the overlying water of oxic incubation cores vs. time. Slope
953 of the best-fit lines gives the ^{224}Ra flux in $\text{dpm m}^{-2} \text{d}^{-1}$.

954 Figure S3. Radiographs of ST8 (a) and ST16 (b) subcores collected in Summer 2010.

955

956 Table 1.

Spring 2009	Surface	Deep
	(dpm·100L ⁻¹)	
Western Basin	3.40 ± 1.57	6.94 ± 1.35
Central Basin	4.56 ± 1.89	5.72 ± 2.02
Eastern Basin	6.26 ± 1.29	5.63 ± 1.03
Mean	4.7 ± 1.4	6.1 ± 0.7
Summer 2009		
Western Basin	12.65 ± 5.54	15.51 ± 3.82
Central Basin	7.29 ± 3.06	12.71 ± 3.30
Eastern Basin	7.34 ± 1.60	7.25 ± 0.87
Mean	9.1 ± 3.1	11.8 ± 4.2
Summer 2010		
Western Basin	11.91 ± 5.94	18.20 ± 4.50
Central Basin	7.39 ± 5.48	16.59 ± 6.60
Eastern Basin	6.14 ± 3.74	8.22 ± 2.81
Mean	8.5 ± 3.0	14.3 ± 5.4

957

958

Sediment	Core	²²⁴ Ra flux (dpm·m ⁻² ·d ⁻¹)	
		Oxic incubations	Hypoxic incubations**
Muddy	ST 16	127 ± 12*	75 → 214 → 170
	ST 13	313 ± 16	n.d.
	ST 120E	170 ± 7	n.d.
	ST 8	164 ± 13*	50 → 140 → 70
Sandy	STWM	57 ± 3	n.d.

960

n.d.: not determined

961

*Mean ± 1SD of individual fluxes over course of oxic incubation (see Fig. 7)

962

**Values indicate variation in flux with time (~7 days) after start of hypoxic

963

incubation (see Fig. 7)

964

965

966 Table 3.

	Basin	Water Volume (L)	Spring 2009		Summer 2010	
			Mean ^{224}Ra concentration (dpm·100L ⁻¹)	^{224}Ra Inventory (dpm)	Mean ^{224}Ra concentration (dpm·100L ⁻¹)	^{224}Ra Inventory (dpm)
Surface	Eastern basin	$3.40 \cdot 10^{12}$	6.08 ± 1.40	$(2.07 \pm 0.48) \cdot 10^{11}$	6.53 ± 4.04	$(2.22 \pm 1.37) \cdot 10^{11}$
	Central basin	$2.90 \cdot 10^{12}$	4.92 ± 1.37	$(1.43 \pm 0.40) \cdot 10^{11}$	7.50 ± 6.26	$(2.17 \pm 1.82) \cdot 10^{11}$
	Western basin	$1.87 \cdot 10^{12}$	4.73 ± 1.76	$(0.88 \pm 0.33) \cdot 10^{11}$	11.91 ± 5.94	$(2.23 \pm 1.11) \cdot 10^{11}$
Deep	Eastern basin	$1.88 \cdot 10^{13}$	5.10 ± 1.33	$(9.58 \pm 2.51) \cdot 10^{11}$	8.40 ± 3.10	$(15.80 \pm 5.82) \cdot 10^{11}$
	Central basin	$1.76 \cdot 10^{13}$	5.70 ± 2.26	$(10.02 \pm 3.98) \cdot 10^{11}$	15.68 ± 5.85	$(27.59 \pm 10.29) \cdot 10^{11}$
	Western basin	$7.42 \cdot 10^{12}$	6.92 ± 1.26	$(5.14 \pm 0.94) \cdot 10^{11}$	18.20 ± 4.50	$(13.50 \pm 3.34) \cdot 10^{11}$
Total		$5.20 \cdot 10^{13}$		$(29.1 \pm 4.8) \cdot 10^{11}$		$(63.5 \pm 12.5) \cdot 10^{11}$

967 Considered stations for: Eastern basin (1, 2, 3, 4, 101, 102)
968 Central basin (113E, 113W, 5, 6, 7, 8, 9, 117)
969 Western basin (10, 120E, 120W, 11, 12, 13, 14, 15, 16, 17, 18)

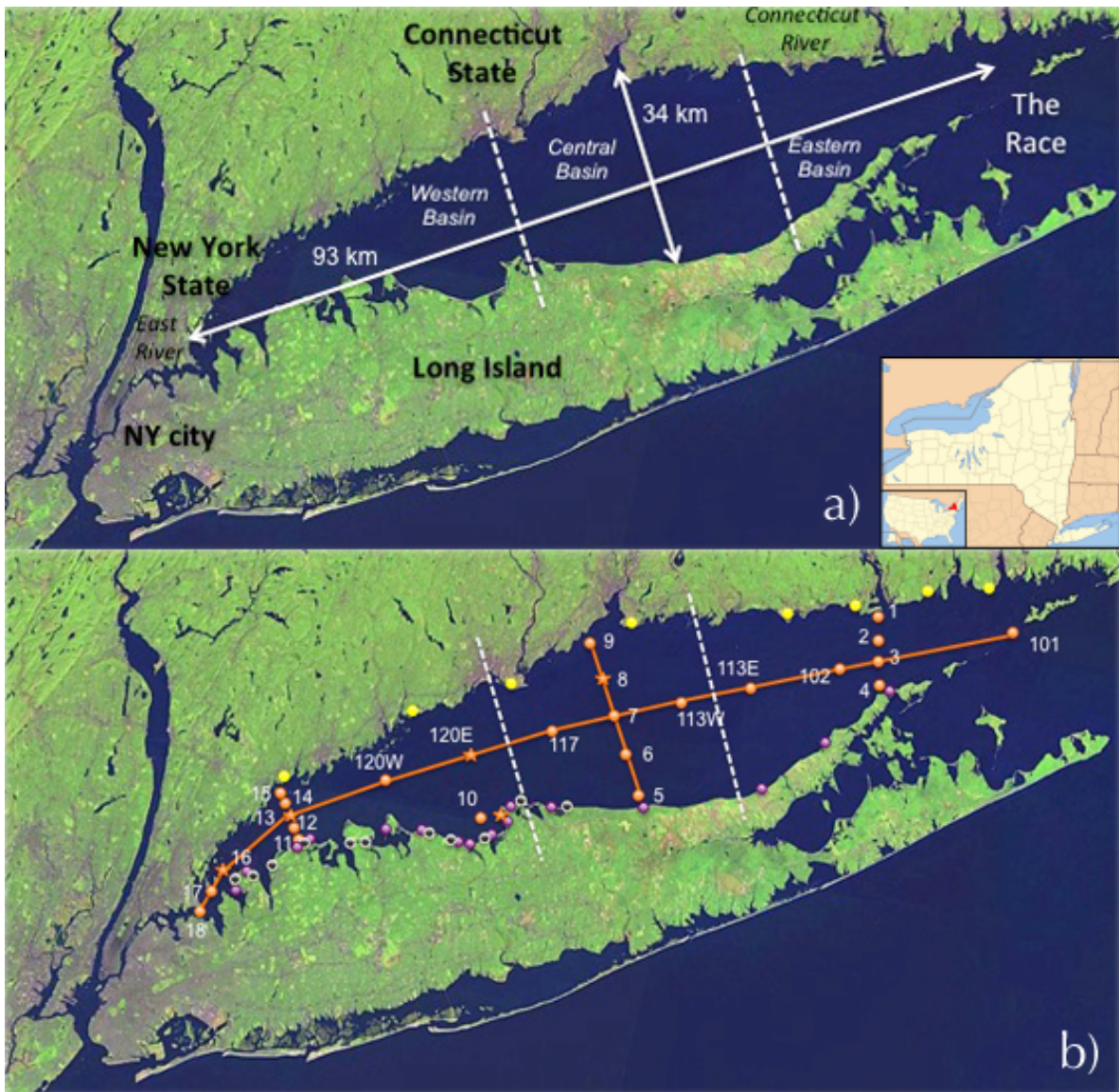
970

971 Table 4.

	²²⁴ Ra (·10 ¹² dpm·y ⁻¹)	
	SPRING 2009	SUMMER 2010
<i>Fluxes OUT</i>		
LIS to Block Island Sound	34.2 ± 8.0	41.6 ± 16.0
LIS to East river/NY Harbor	4.4 ± 1.0	11.6 ± 2.9
Decay	205 ± 34	447 ± 88
<i>Fluxes IN</i>		
Block Island Sound to LIS	21.3 ± 1.4	21.3 ± 1.4
East River to LIS	4.1 ± 0.9	4.1 ± 0.9
Connecticut River	1.4 ± 0.1	2.0 ± 0.4
Desorption during resuspension	31 ± 27	82 ± 51
Diffusion and bioirrigation from sediments	80 ± 24	147 ± 44
SGD	106 ± 50	244 ± 112
SGD from shoreline flux	173	

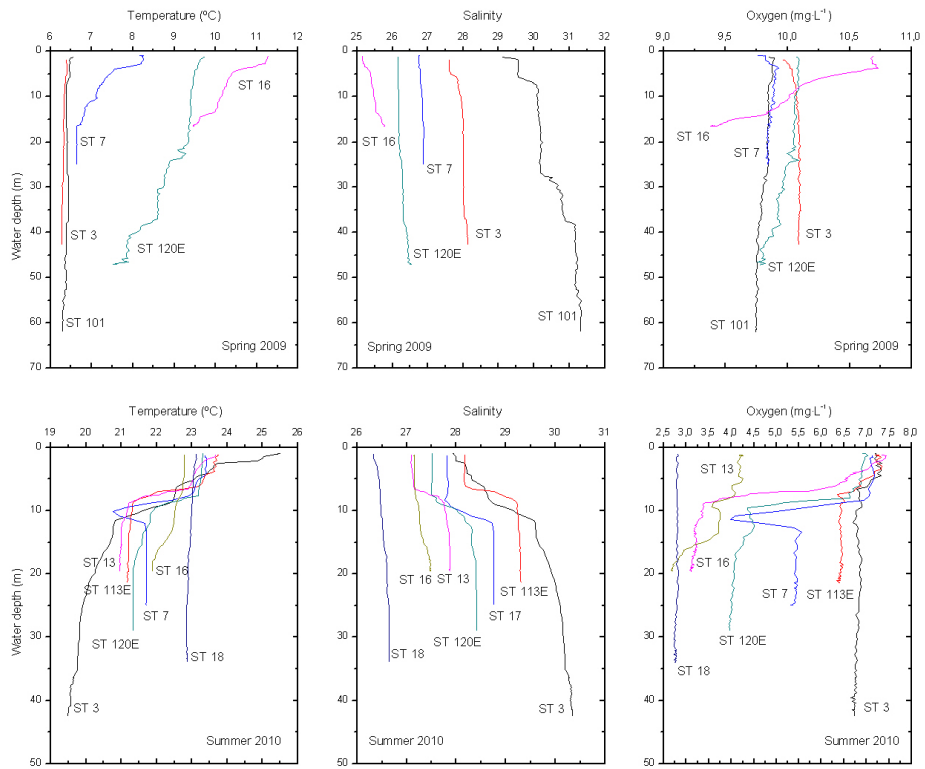
972

973 Figure 1

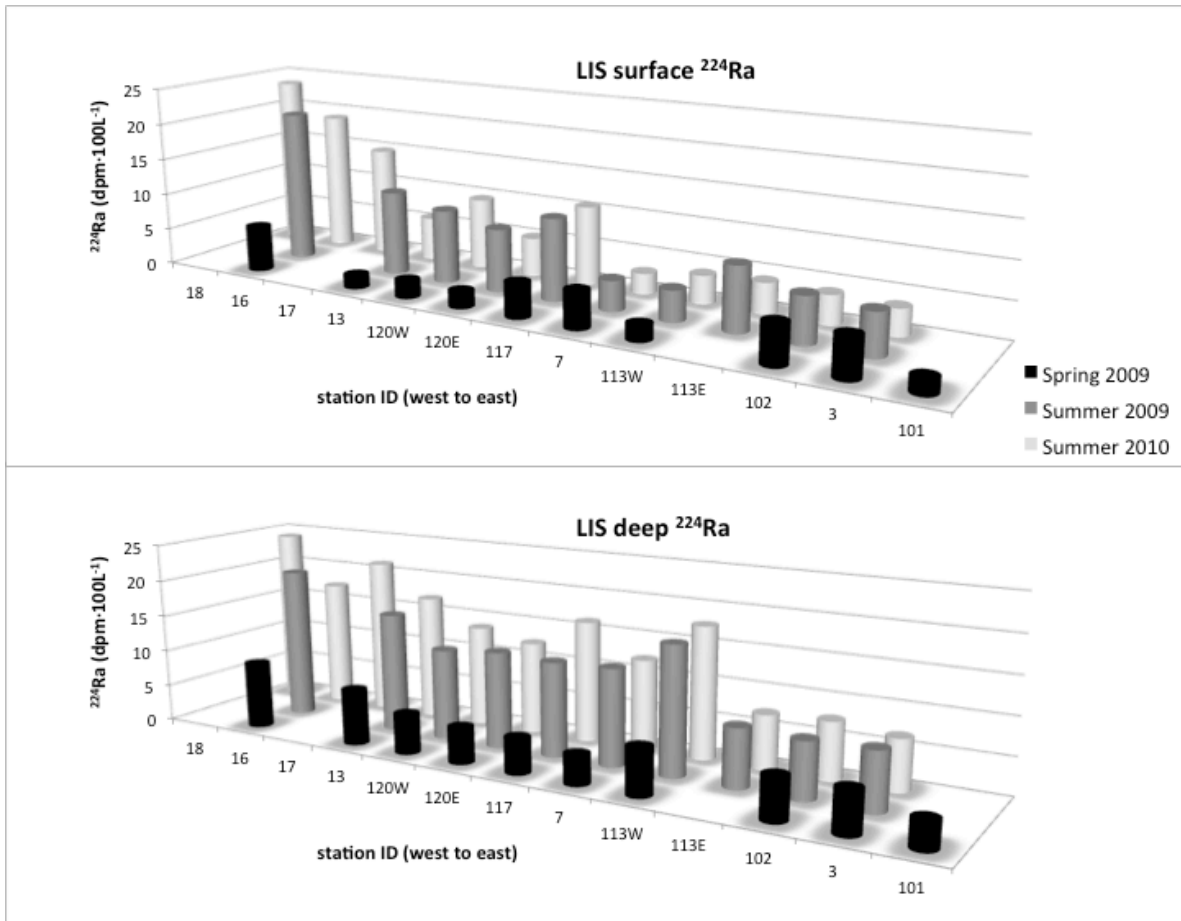


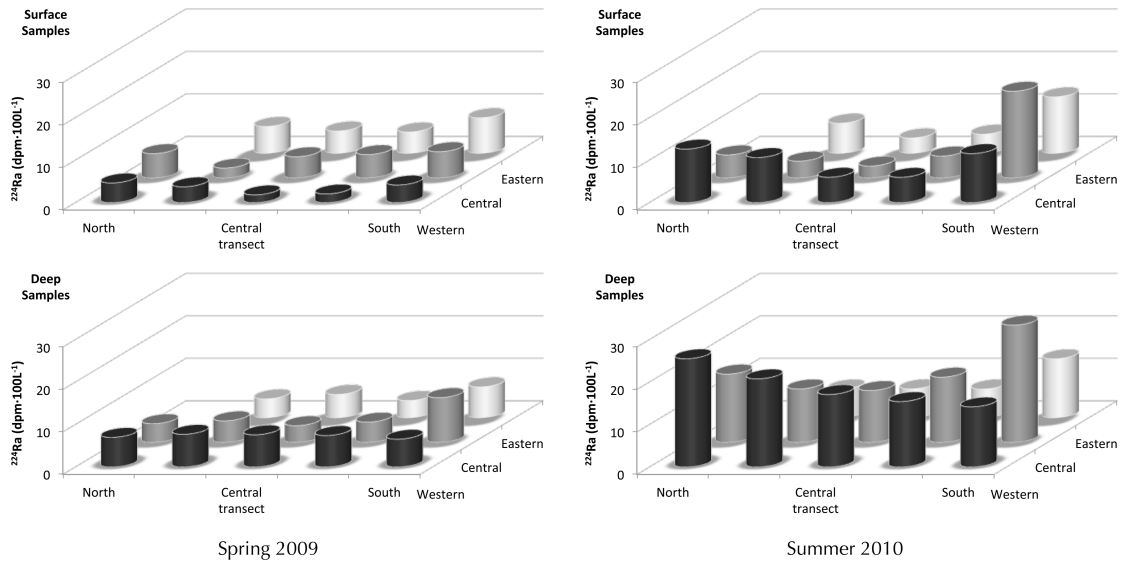
974

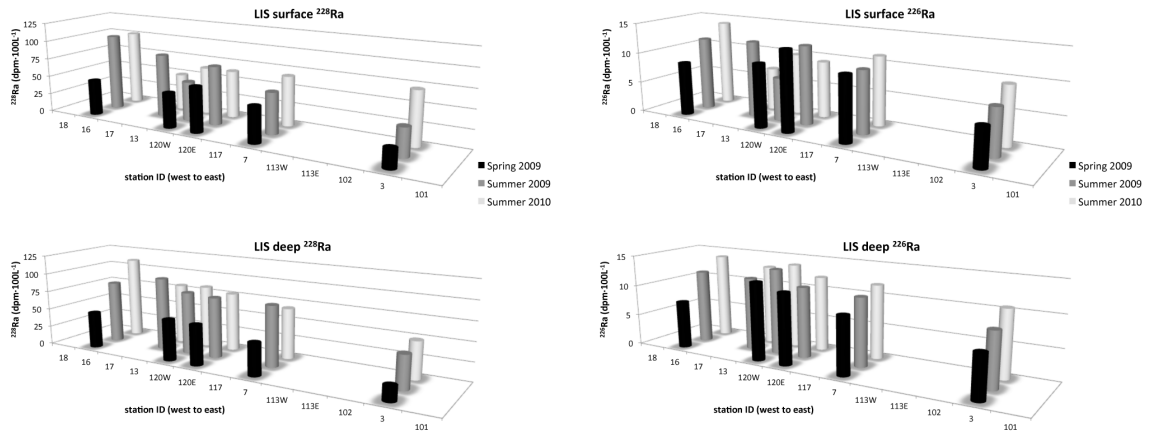
975 Figure 2



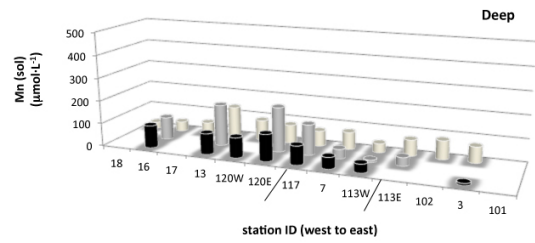
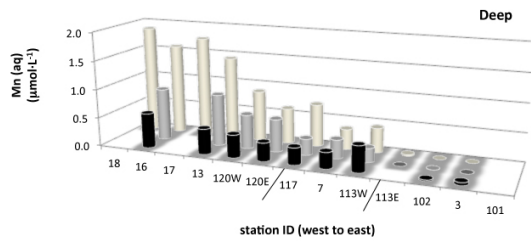
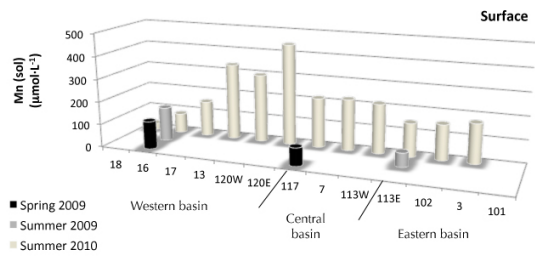
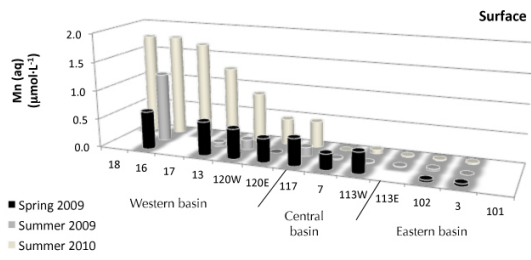
976





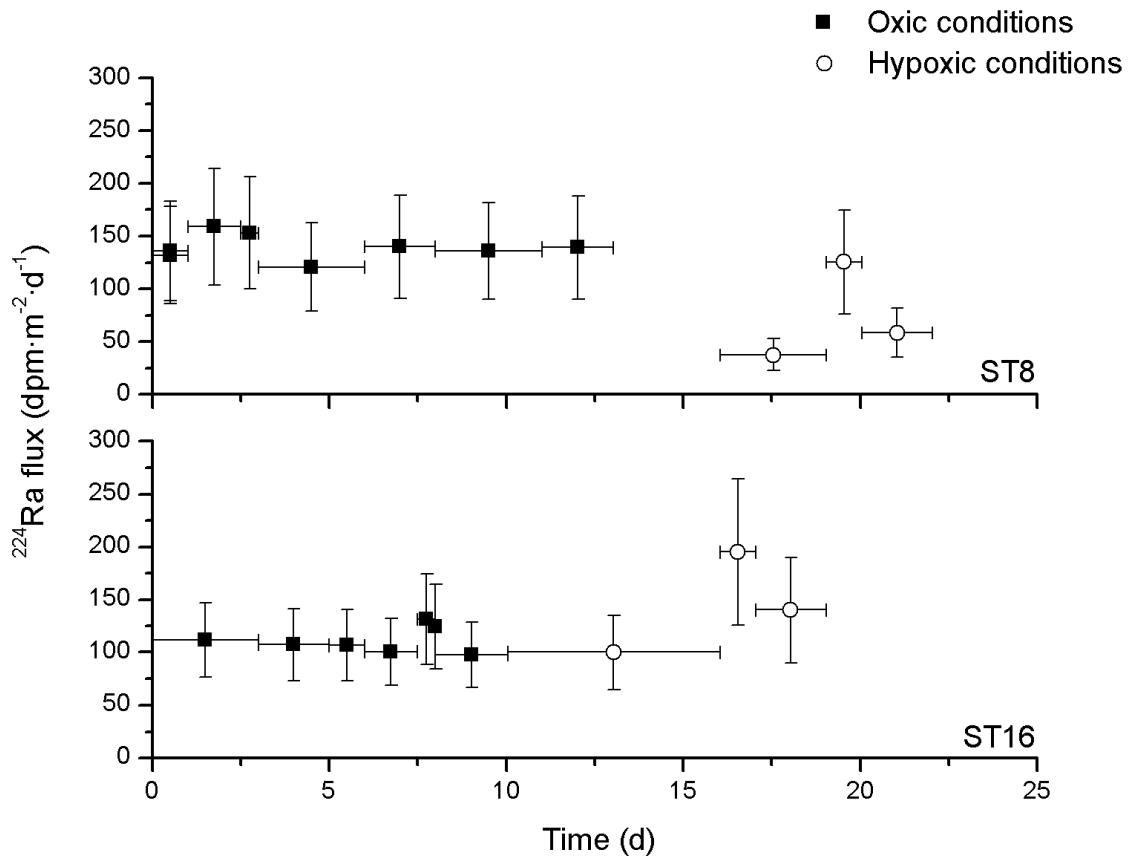


983 Figure 6

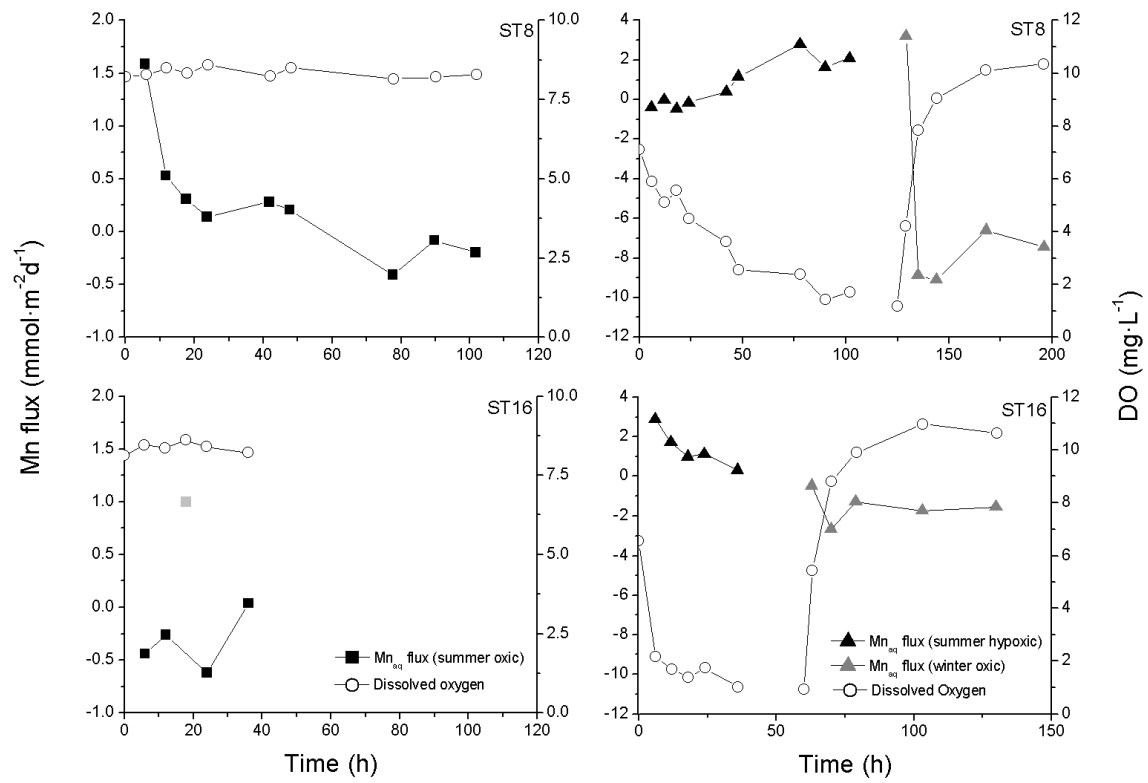


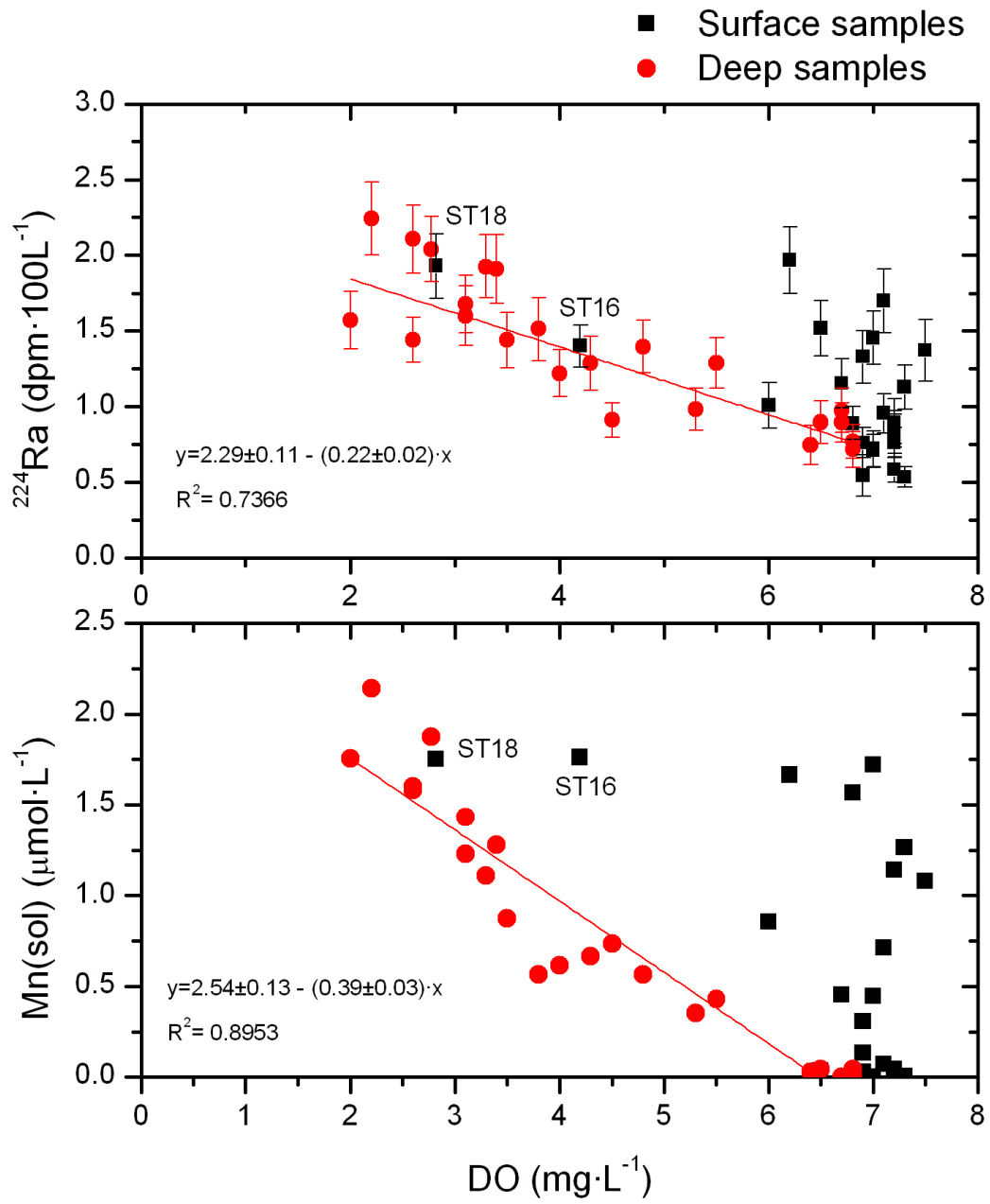
984

985 Figure 7



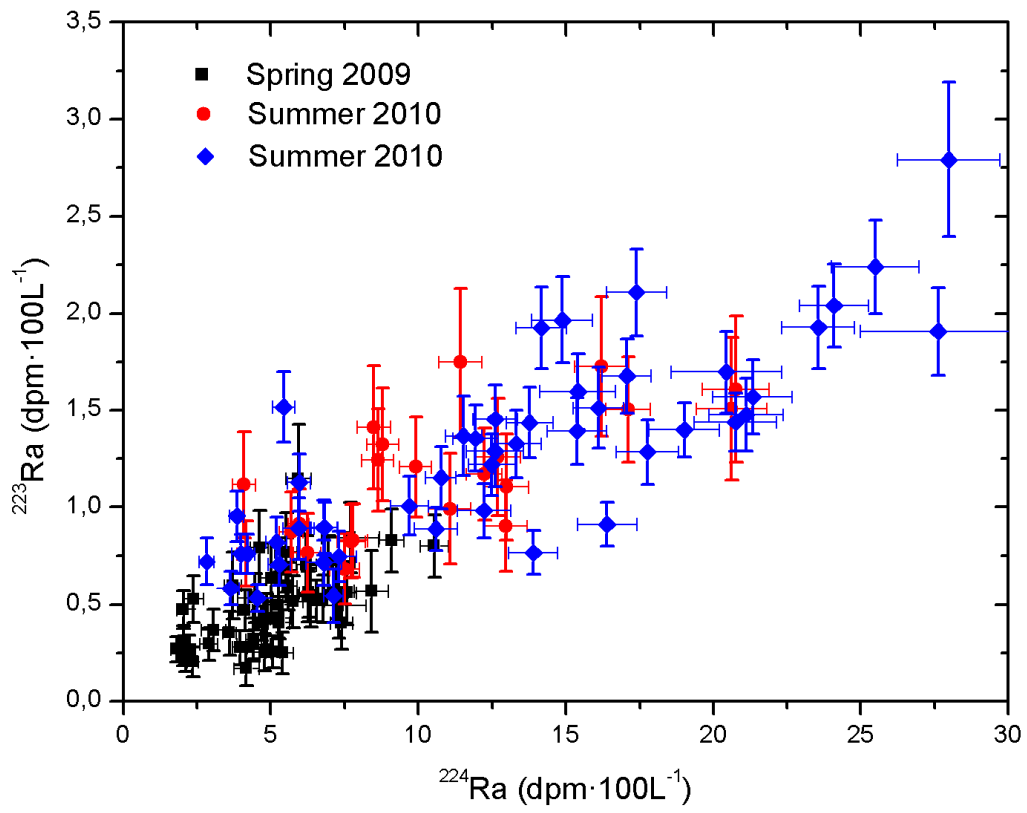
986

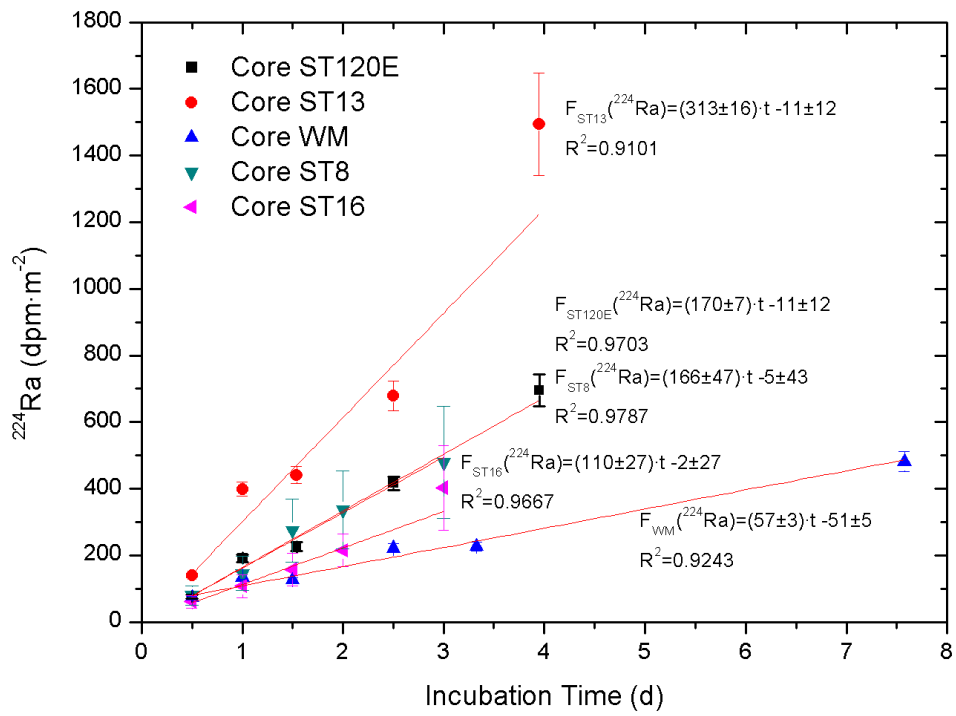




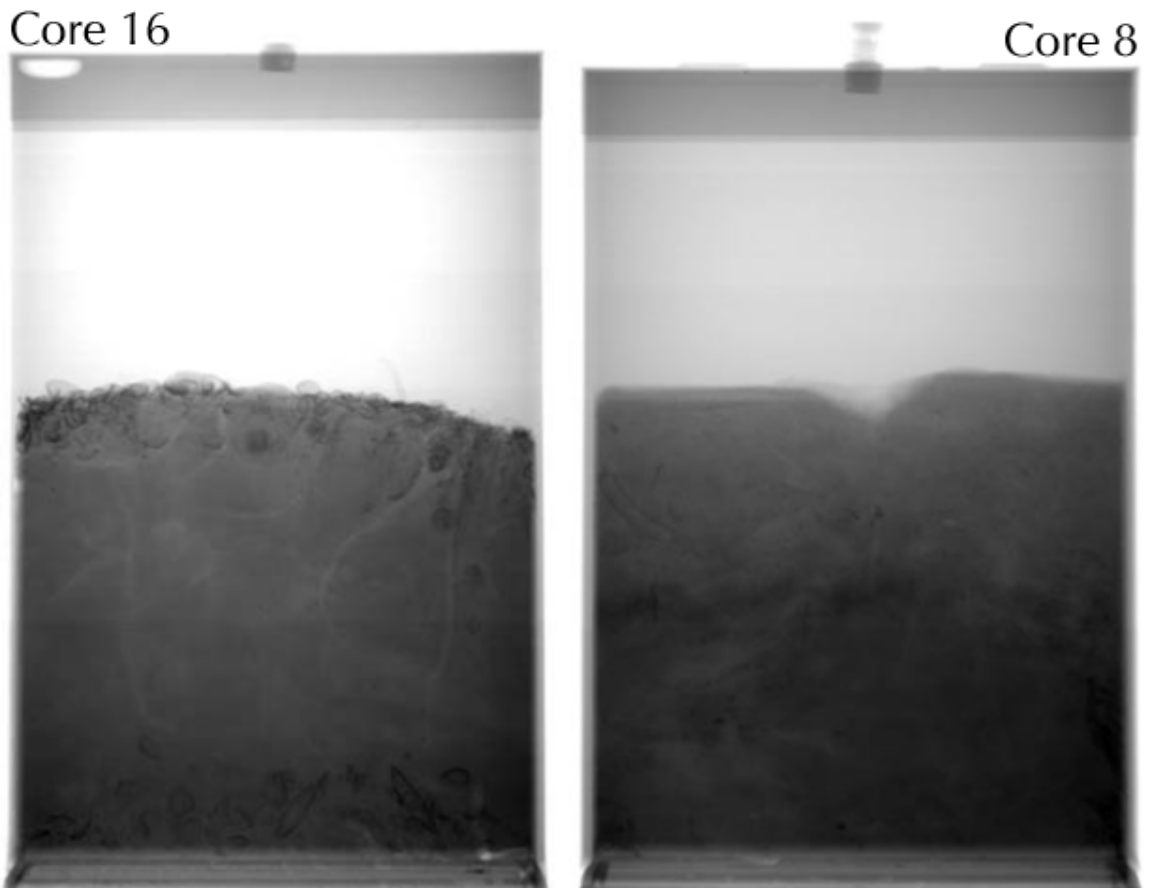
990

991





996 Figure S3



997

998

SPRING 2009	Sampling hour	Coordinates		Depth (m)	Sampling depth (m)	Salinity	T ^a (°C)	DO mg·L ⁻¹	SPM mg·L ⁻¹	Mn _{aq} μmol·L ⁻¹	Mn _p μmol·g ⁻¹
		Latitude (N)	Longitude (W)								
ST1 _{surf}	24/04/09 04:35	41° 15.02'	72° 20.03'	9	0.5	20.8	6.8	10.1	4.73		
ST1 _{deep}	24/04/09 04:25	41° 15.02'	72° 20.03'	9	8	27.3	6.3	10	8.50		
ST2 _{surf}	24/04/09 05:30	41° 13.17'	72° 19.56'	41	0.5	27.7	9.9	9.9	0.86		
ST2 _{deep}	24/04/09 05:30	41° 13.17'	72° 19.56'	41	40	29	7.7	9.9	0.74		
ST3 _{surf}	24/04/09 06:15	41° 11.51'	72° 20.01'	43	0.5	27.6	7.6	9.9	0.33	0.05	627
ST3 _{deep}	24/04/09 06:15	41° 11.51'	72° 20.01'	43	42	28.1	7.5	10	2.55	0.05	9
ST4 _{surf}	24/04/09 06:50	41° 09.532'	72° 20.080'	25	0.5	27.4	7.7	10.1	1.09		
ST4 _{deep}	24/04/09 06:50	41° 09.532'	72° 20.080'	25	22	27.9	7.7	10.1	2.13		
ST101 _{surf}	24/04/09 09:00	41° 14.209'	72° 03.594'	70	0.5	29.1	8.1	10.1	0.98		
ST101 _{deep}	24/04/09 09:00	41° 14.209'	72° 03.594'	70	61	31.3	7.2	9.9	1.56		
ST102 _{surf}	24/04/09 11:50	41° 09.848'	72° 25.740'	49	0.5	28.1	9.4	9.7	0.51	0.04	74
ST102 _{deep}	24/04/09 11:50	41° 09.848'	72° 25.740'	49	46	28.6	7.7	9.9	0.40	0.02	323
ST5 _{surf}	28/04/09 08:50	40° 59.018'	72° 52.028'	21	0.5	26.5	10.5	10	1.03		
ST5 _{deep}	28/04/09 08:50	40° 59.018'	72° 52.028'	21	18	26.5	9.5	10	1.11		
ST6 _{surf}	28/04/09 09:30	41° 02.329'	72° 53.610'	37	0.5	26.5	10.9	10	0.78		
ST6 _{deep}	28/04/09 09:30	41° 02.329'	72° 53.610'	37	35	26.9	7.9	9.9	1.51		
ST117 _{surf}	28/04/09 10:40	41° 04.993'	73° 03.270'	26	0.5	26.3	11.3	9.9	1.06	0.47	79
ST117 _{deep}	28/04/09 10:40	41° 04.993'	73° 03.270'	26	23	26.8	8.2	9.8	1.04	0.30	82
ST7 _{surf}	28/04/09 11:30	41° 05.826'	72° 55.303'	28	0.5	26.7	10.8	9.8	0.84	0.27	73
ST7 _{5m}	28/04/09 12:45	41° 05.826'	72° 55.303'	28	5	26.8	8.8	9.8	0.66		
ST7 _{15m}	28/04/09 12:05	41° 05.826'	72° 55.303'	28	14.5	26.9	9.4	9.9	0.97		
ST7 _{deep}	28/04/09 11:30	41° 05.826'	72° 55.303'	28	25	26.9	8.6	9.9	1.10	0.28	50

ST8 _{surf}	28/04/09 15:05	41° 09.166'	72° 56.800'	20.5	0.5	26.7	12.1	10	0.68		
ST8 _{deep}	28/04/09 15:25	41° 09.166'	72° 56.800'	20.5	17	26.7	8.9	10	1.58		
ST113 _{surf}	28/04/09 13:45	41° 07.329'	72° 45.422'	32	0.5	26.8	12.1	10	0.49	0.37	373
ST113 _{deep}	28/04/09 13:45	41° 07.329'	72° 45.422'	32	29	26.9	8.9	9.9	2.30	0.45	37
ST9 _{surf}	28/04/09 16:00	41° 12.446'	72° 58.714'	12	0.5	26.2	13.5	10.4	0.88		
ST9 _{deep}	28/04/09 16:20	41° 12.446'	72° 58.714'	12	9	26.2	11.3	10.3	1.56		
ST11 _{surf}	30/04/09 10:25	40° 55.529'	73° 34.271'	15	0.5	25.5	12.0	11.2	1.09		
ST11 _{deep}	30/04/09 10:25	40° 55.529'	73° 34.271'	15	12.5	26	9.9	9.2	3.15		
ST13 _{surf}	30/04/09 10:50	40° 57.469'	73° 35.187'	16	0.5	25.6	12.2	11.4	0.93	0.59	149
ST13 _{deep}	30/04/09 11:00	40° 57.469'	73° 35.187'	16	15	26.2	9.1	9.4	2.32	0.44	91
ST15 _{surf}	30/04/09 11:25	40° 58.904'	73° 36.499'	13.5	0.5	26	11.5	10.2	0.96		
ST15 _{deep}	30/04/09 11:25	40° 58.904'	73° 36.499'	13.5	10	26.1	10.0	9.9	1.23		
ST12 _{surf}	30/04/09 13:50	40° 56.512'	73° 34.614'	0.5	0.5	25.6	13.07	11.2	1.02		
ST12 _{deep}	30/04/09 13:50	40° 56.512'	73° 34.614'	17	15.5	26.1	10.16	9.2	2.22		
ST14 _{surf}	30/04/09 14:20	40° 58.364'	73° 35.762'	0.5	0.5	25.9	26.35	10.5	1.33		
ST14 _{deep}	30/04/09 14:20	40° 58.364'	73° 35.762'	18	15.8	26.3	26.7	9.5	1.68		
ST10 _{surf}	30/04/09 08:15	40° 56.275'	73° 11.576'	16	0.5	26.2	10.7	10	0.88		
ST10 _{deep}	30/04/09 08:15	40° 56.275'	73° 11.576'	16	14.2	26.3	9.6	9.6	1.73		
ST120W _{surf}	30/04/09 15:30	40° 59.680'	73° 25.231'	52	0.5	26.3	11.0	10	0.72	0.52	168
ST120W _{deep}	30/04/09 15:30	40° 59.680'	73° 25.231'	52	47	26.5	8.9	9.7	1.82	0.40	87
ST120E _{surf}	30/04/09 16:45	41° 01.975'	73° 13.515'	32	0.5	26.4	11.5	10	0.66	0.41	122
ST120E _{deep}	30/04/09 16:45	41° 01.975'	73° 13.515'	32	28	26.7	8.3	9.7	1.21	0.32	117
ST16 _{surf}	30/04/09 12:35	40° 52.500'	73° 44.944'	18	0.5	25.2	12.4	10.7	1.53	0.66	122
ST16 _{deep}	30/04/09 12:35	40° 52.500'	73° 44.944'	18	16.4	25.8	10.6	9.4	2.00	0.61	98

1000

1001 Supplemental Data - Table S1. T^a, S, DO, SPM, Mn_{aq}, Mn_s (continuation).

SUMMER 2009	Sampling hour	Coordinates		Depth (m)	Sampling depth (m)	Salinity	T ^a (°C)	DO mg·L ⁻¹	SPM mg·L ⁻¹	Mn _{aq} μmol·L ⁻¹	Mn _{sol} μmol·g ⁻¹
		Latitude (N)	Longitude (W)								
ST16 _{surf}	29/07/09 14:50	40° 52.463'	73° 44.969'	20	0.5	24.9		6.9	2.64	1.19	148
ST16 _{deep}	29/07/09 14:50	40° 52.463'	73° 44.969'	20	19	25.8		3.8	3.68	0.91	99
ST13 _{surf}	29/07/09 16:00	40° 57.439'	73° 35.365'	21	0.5	25.7		8.4	1.42	0.07	417
ST13 _{deep}	29/07/09 16:00	40° 57.439'	73° 35.365'	21	20	26.1		4.2	4.22	0.90	180
ST120W _{surf}	29/07/09 12:50	40° 59.699'	73° 25.201'	46	0.5	25.5		9.4	2	0.17	357
ST120W _{deep}	29/07/09 12:50	40° 59.699'	73° 25.201'	46	45	26.5		4.2	3.42	0.59	802
ST120E _{surf}	29/07/09 10:30	41° 01.960'	73° 13.500'	28	0.5	25.5		8	0.84	0.00	325
ST120E _{deep}	29/07/09 10:30	41° 01.960'	73° 13.500'	28	27	26.8		4.6	3.04	0.57	194
ST117 _{surf}	29/07/09 08:30	40° 04.994'	73° 03.253'	28	0.5	24.9		6.9	0.76	0.16	494
ST117 _{deep}	29/07/09 09:15	40° 04.994'	73° 03.253'	28	26.7	27		3.5	3.84	0.28	133
ST7 _{surf}	04/08/09 19:20	41° 05.925'	72° 55.454'	26	0.5	25.4		7	1.24	0.02	195
ST7 _{deep}	04/08/09 19:20	41° 05.925'	72° 55.454'	26	25	27.2		4.1	5.58	0.33	42
ST113W _{surf}	04/08/09 09:50	41° 07.363'	72° 45.583'	29	0.5	25.8		7.5	1.04	0.02	895
ST113W _{deep}	04/08/09 09:50	41° 07.363'	72° 45.583'	29	28	27.5		4.7	6.8	0.24	10
ST113E _{surf}	04/08/09 11:00	41° 08.653'	72° 35.206'	23	0.5	25.3		8.3	2.22	0.01	61
ST113E _{deep}	04/08/09 11:00	41° 08.653'	72° 35.206'	23	22	27.8		5.7	2.22	0.00	37
ST102 _{surf}	04/08/09 12:00	41° 10.139'	72° 27.721'	28	0.5	27		6.4	1.1	0.04	
ST102 _{deep}	04/08/09 12:00	41° 10.139'	72° 27.721'	28	27	28.9		5.4	1.14	0.00	
ST3 _{surf}	04/08/09 11:40	41° 11.556'	72° 19.919'	36	0.5	26.5		7	1.36	0.00	43
ST3 _{deep}	04/08/09 11:40	41° 11.556'	72° 19.919'	36	35	29.6		5.4	1.1	0.00	60
ST1 _{surf}	04/08/09 14:15	41° 14.948'	72° 19.928'	92	0.5	29		6.4	1.42		
ST1 _{deep}	04/08/09 14:15	41° 14.948'	72° 19.928'	92	90	31.1		5.7	0.8		

1003

1004 Supplemental Data - Table S1. T^a, S, DO, SPM, Mn_{aq}, Mn_s (continuation).

SUMMER 2010	Sampling hour	Coordinates		Depth (m)	Sampling depth (m)	Salinity	T ^a (°C)	DO mg·L ⁻¹	SPM mg·L ⁻¹	Mn _{aq} μmol·L ⁻¹	Mn _{sol} μmol·g ⁻¹
		Latitude (N)	Longitude (W)								
ST5 _{surf}	03/08/10 09:45	40° 59.055'	72° 51.867'	21	0.5	27.6	23.0	6.5	1.53		
ST5 _{deep}	03/08/10 09:45	40° 59.055'	72° 51.867'	21	20	28	21.8	3.8	3.64		
ST6 _{surf}	03/08/10 10:30	40° 02.299	72° 53.650'	36	0.5	27.6	23.4	6.9	1.19		
ST6 _{deep}	03/08/10 10:30	40° 02.299	72° 53.650'	36	35	28.6	20.8	4.3	2.63		
ST117 _{surf}	03/08/10 11:40	40° 04.994'	73° 03.253'	24	0.5	27.7	23.2	6.7	2.31	0.47	79
ST117 _{deep}	03/08/10 11:40	40° 04.994'	73° 03.253'	24	23	28.6	21.5	4.5	3.79	0.30	82
ST7 _{surf}	03/08/10 12:35	41° 05.925'	72° 55.454'	26	0.5	27.8	24.0	7	1.05	0.27	73
ST7 _{deep}	03/08/10 12:35	41° 05.925'	72° 55.454'	26	25	28.8	22.0	5.3	3.36	0.28	50
ST113W _{surf}	03/08/10 13:40	41° 07.363'	72° 45.583'	29	0.5	27.7	24.5	7.1	1.10	0.37	373
ST113W _{deep}	03/08/10 13:40	41° 07.363'	72° 45.583'	29	28	28.8	22.4	5.5	6.75	0.45	37
ST8 _{surf}	03/08/10 14:55	41° 09.136'	72° 56.808'	19	0.5	27.8	24.3	7.2	1.50		
ST8 _{deep}	03/08/10 14:55	41° 09.136'	72° 56.808'	19	18	28.6	22.2	4.8	3.64		
ST9 _{surf}	03/08/10 15:30	41° 12.434'	72° 58.669'	11	0.5	27.8	25.0	7.1	3.10		
ST9 _{deep}	03/08/10 15:30	41° 12.434'	72° 58.669'	11	10	28	22.9	3.4	7.01		
ST113E _{surf}	05/08/10 10:25	41° 08.653'	72° 35.206'	22	0.5	28.2	24.5	7.2	1.25		
ST113E _{deep}	05/08/10 10:25	41° 08.653'	72° 35.206'	22	21	29.3	21.6	6.4	2.32		
ST102 _{surf}	05/08/10 11:15	41° 10.139'	72° 27.721'	33	0.5	27.9	24.4	7.2	1.00	0.04	74
ST102 _{deep}	05/08/10 11:15	41° 10.139'	72° 27.721'	33	31	29.8	22.0	6.7	1.51	0.02	323
ST1 _{surf}	05/08/10 12:15	41° 14.948'	72° 19.928'	8	0.5	27.5	22.2	6.9	2.99		
ST1 _{deep}	05/08/10 12:15	41° 14.948'	72° 19.928'	8	7	29.2	21.9	6.8	4.49		
ST2 _{surf}	05/08/10 12:40	41° 13.092'	72° 19.920'	42	0.5	29	25.7	7.3	1.27		

ST2 _{deep}	05/08/10 12:40	41° 13.092'	72° 19.920'	42	40	30.4	20.4	6.8	2.28		
ST3 _{surf}	05/08/10 12:10	41° 11.556'	72° 19.919'	45	0.5	27.9	24.3	7.2	1.23	0.05	627
ST3 _{deep}	05/08/10 12:10	41° 11.556'	72° 19.919'	45	42	30.3	20.4	6.7	1.30	0.05	9
ST4 _{surf}	05/08/10 13:35	41° 09.547'	72° 19.922'	25	0.5	28	25.4	6.9	1.05		
ST4 _{deep}	05/08/10 13:35	41° 09.547'	72° 19.922'	25	23	29.9	22.5	6.5	2.54		
ST10 _{surf}	10/08/10 08:20	40° 56.326'	73° 11.609'	15	0.5	27.3	23.1	5.9	1.05		
ST10 _{deep}	10/08/10 08:20	40° 56.326'	73° 11.609'	15	14	28	21.9	3.2	5.48		
ST120E _{surf}	10/08/10 09:10	41° 01.960'	73° 13.500'	30	0.5	27.5	23.3	7.0	1.65	0.41	122
ST120E _{deep}	10/08/10 09:10	41° 01.960'	73° 13.500'	30	29	28.4	21.3	4.0	6.05	0.32	117
ST120W _{surf}	10/08/10 10:15	40° 59.699'	73° 25.201'	49	0.5	27.5	22.9	6.0	2.15	0.52	168
ST120W _{deep}	01/01/04 10:15	40° 59.699'	73° 25.201'	49	48	28.2	21.4	3.5	5.94	0.40	87
ST11 _{surf}	10/08/10 11:15	40° 55.777'	73° 34.280'	16	0.5	27.1	23.5	7.0	2.24		
ST11 _{deep}	10/08/10 11:15	40° 55.777'	73° 34.280'	16	15	27.4	21.4	2.2	3.26		
ST12 _{surf}	10/08/10 11:40	40° 56.510'	73° 34.618'	18	0.5	27.1	24.2	6.8	2.53		
ST12 _{deep}	10/08/10 11:40	40° 56.510'	73° 34.618'	18	17	27.7	21.4	2.6	4.97		
ST14 _{surf}	10/08/10 12:15	40° 58.898'	73° 36.487'	16	0.5	27.3	24.2	7.5	2.08		
ST14 _{deep}	10/08/10 12:15	40° 58.898'	73° 36.487'	16	15	27.8	22.0	3.3	5.23		
ST15 _{surf}	10/08/10 12:45	40° 58.303'	73° 35.739'	19	0.5	27.2	24.1	7.2	1.80		
ST15 _{deep}	10/08/10 12:45	40° 58.303'	73° 35.739'	19	18	27.8	21.5	3.1	5.26		
ST13 _{surf}	10/08/10 13:50	40° 57.439'	73° 35.365'	21	0.5	27.1	24.5	7.3	2.31	0.59	149
ST13 _{deep}	10/08/10 13:50	40° 57.439'	73° 35.365'	21	20	27.8	21.6	3.1	5.03	0.44	91
ST17 _{surf}	12/08/10 10:30	40° 55.109'	73° 40.243'	16	0.5	27.2	23.0	6.2	2.75		
ST17 _{deep}	12/08/10 10:30	40° 55.109'	73° 40.243'	16	15	27.5	21.2	2.0	7.08		
ST18 _{surf}	12/08/10 10:40	40° 48.020'	73° 47.356'	34	0.5	26.3	22.7	2.8	5.49		
ST18 _{deep}	12/08/10 10:40	40° 48.020'	73° 47.356'	34	33	26.6	22.5	2.8	6.43		
ST16 _{surf}	12/08/10 12:25	40° 52.463'	73° 44.969'	20	0.5	27.2	22.6	4.2	3.50	0.66	122
ST16 _{deep}	12/08/10 12:25	40° 52.463'	73° 44.969'	20	21	27.5	21.6	2.6	13.51	0.61	98

SPRING 2009	^{224}Ra	^{223}Ra	^{226}Ra	^{228}Ra
	dpm·100L ⁻¹	dpm·100L ⁻¹	dpm·100L ⁻¹	dpm·100L ⁻¹
ST1 _{surf}	8.4 ± 0.6	0.57 ± 0.21		
ST1 _{deep}	7.3 ± 1.0	0.49 ± 0.17		
ST2 _{surf}	5.9 ± 0.4	1.15 ± 0.28	5.7 ± 0.7	35 ± 4
ST2 _{deep}	5.3 ± 0.4	0.50 ± 0.17	5.6 ± 0.6	27 ± 5
ST3 _{surf}	5.3 ± 0.4	0.41 ± 0.13	5.9 ± 1.1	32 ± 4
ST3 _{deep}	5.5 ± 0.4	0.68 ± 0.16	6.7 ± 1.1	24 ± 4
ST4 _{surf}	6.4 ± 0.4	0.54 ± 0.16		
ST4 _{deep}	4.5 ± 0.6	0.53 ± 0.12		
ST101 _{surf}	3.7 ± 0.4	0.80 ± 0.16	5.8 ± 0.5	19 ± 3
ST101 _{deep}	3.6 ± 0.2	0.35 ± 0.11	5.3 ± 0.5	29 ± 3
ST102 _{surf}	5.3 ± 0.4	0.43 ± 0.11	6.2 ± 0.5	22 ± 4
ST102 _{deep}	5.6 ± 0.3	0.59 ± 0.12	6.3 ± 0.5	28 ± 4
ST5 _{surf}	5.7 ± 0.3	0.51 ± 0.13	8.5 ± 1.3	63 ± 5
ST5 _{deep}	4.4 ± 0.6	0.29 ± 0.08	8.9 ± 1.2	59 ± 5
ST6 _{surf}	2.4 ± 0.2	0.20 ± 0.08		
ST6 _{deep}	5.1 ± 0.2	0.25 ± 0.08		
ST117 _{surf}	4.7 ± 0.2	0.40 ± 0.09	10.8 ± 0.6	54 ± 5
ST117 _{deep}	4.8 ± 0.2	0.43 ± 0.08	11.4 ± 0.6	35 ± 5
ST7 _{surf}	5.0 ± 0.4	0.63 ± 0.18	10.1 ± 1.4	64 ± 6
ST7 _{5m}	3.1 ± 0.4	0.37 ± 0.11		
ST7 _{15m}	4.4 ± 0.3	0.32 ± 0.11		
ST7 _{deep}	4.0 ± 0.3	0.28 ± 0.09	8.9 ± 1.3	57 ± 5
ST8 _{surf}	5.5 ± 0.4	0.56 ± 0.13	9.1 ± 1.2	52 ± 5
ST8 _{deep}	4.8 ± 0.6	0.77 ± 0.20	9.7 ± 1.1	60 ± 4
ST113 _{surf}	3.7 ± 0.4	0.25 ± 0.09	10.7 ± 0.6	61 ± 5
ST113 _{deep}	6.3 ± 0.3	0.37 ± 0.11	10.6 ± 0.5	57 ± 4
ST9 _{surf}	6.2 ± 0.3	0.70 ± 0.15		
ST9 _{deep}	10.5 ± 0.5	0.80 ± 0.16		
ST11 _{surf}	4.6 ± 0.3	0.79 ± 0.19		
ST11 _{deep}	7.0 ± 0.9	0.71 ± 0.14		
ST13 _{surf}	3.3 ± 0.4	0.45 ± 0.11	8.4 ± 1.2	38 ± 6
ST13 _{deep}	7.6 ± 0.3	0.56 ± 0.16	9.5 ± 1.3	61 ± 5
ST15 _{surf}	4.1 ± 0.2	0.47 ± 0.10		
ST15 _{deep}	6.5 ± 0.3	0.52 ± 0.11		
ST12 _{surf}	3.7 ± 0.3	0.60 ± 0.17		
ST12 _{deep}	7.7 ± 1.0	0.84 ± 0.18		
ST14 _{surf}	3.7 ± 0.4	0.44 ± 0.13		
ST14 _{deep}	7.4 ± 0.4	0.40 ± 0.14		

ST10 _{surf}	7.2 ± 0.4	0.56 ± 0.14		
ST10 _{deep}	6.8 ± 0.4	0.53 ± 0.12		
ST120W _{surf}	4.3 ± 0.6	0.53 ± 0.12	10.0 ± 1.5	66 ± 7
ST120W _{deep}	5.4 ± 0.4	0.25 ± 0.11	12.0 ± 1.5	77 ± 8
ST120E _{surf}	4.0 ± 0.4	0.45 ± 0.12	12.6 ± 1.5	86 ± 7
ST120E _{deep}	4.9 ± 0.3	0.42 ± 0.13	11.1 ± 1.4	76 ± 6
ST16 _{surf}	6.3 ± 0.4	0.54 ± 0.13	8.5 ± 1.2	65 ± 5
ST16 _{deep}	9.1 ± 0.4	0.83 ± 0.16	7.5 ± 1.3	63 ± 5

1007

1008 Supplemental Data - Table S2 (continuation).

SUMMER 2009	²²⁴ Ra dpm·100L ⁻¹	²²³ Ra dpm·100L ⁻¹	²²⁶ Ra dpm·100L ⁻¹	²²⁸ Ra dpm·100L ⁻¹
ST1 _{surf}	8.8 ± 0.5	1.3 ± 0.3	9.4 ± 0.8	51 ± 5
ST1 _{deep}	6.0 ± 0.3	0.9 ± 0.2	10.5 ± 0.9	28 ± 6
ST3 _{surf}	5.7 ± 0.4	0.9 ± 0.2	8.2 ± 1.0	44 ± 6
ST3 _{deep}	7.7 ± 0.5	0.8 ± 0.2	9.9 ± 0.8	52 ± 5
ST102 _{surf}	6.2 ± 0.5	0.8 ± 0.2	10.0 ± 0.5	52 ± 3
ST102 _{deep}	7.5 ± 0.5	0.7 ± 0.2	8.9 ± 0.5	40 ± 3
ST113E _{surf}	8.6 ± 0.5	1.2 ± 0.3	9.2 ± 0.5	54 ± 4
ST113E _{deep}	7.8 ± 0.5	0.8 ± 0.2	9.2 ± 0.6	42 ± 4
ST113W _{surf}	4.2 ± 0.3	0.8 ± 0.2	8.3 ± 0.6	31 ± 4
ST113W _{deep}	17.1 ± 0.8	1.5 ± 0.3	8.1 ± 0.5	41 ± 3
ST7 _{surf}	4.1 ± 0.4	1.1 ± 0.3	9.7 ± 1.0	66 ± 7
ST7 _{deep}	13.0 ± 0.8	1.1 ± 0.3	11.4 ± 1.0	64 ± 8
ST117 _{surf}	11.1 ± 0.7	1.0 ± 0.3	11.5 ± 0.7	74 ± 5
ST117 _{deep}	12.7 ± 0.8	1.3 ± 0.3	10.0 ± 0.6	63 ± 5
ST120E _{surf}	8.5 ± 0.6	1.4 ± 0.3	13.8 ± 1.1	97 ± 5
ST120E _{deep}	13.0 ± 0.7	0.9 ± 0.2	12.1 ± 1.1	99 ± 8
ST120W _{surf}	9.9 ± 0.5	1.2 ± 0.3	7.9 ± 1.0	65 ± 6
ST120W _{deep}	12.2 ± 0.6	1.2 ± 0.2	15.2 ± 1.1	101 ± 8
ST13 _{surf}	11.4 ± 0.7	1.7 ± 0.4	13.2 ± 1.1	101 ± 8
ST13 _{deep}	16.2 ± 0.9	1.7 ± 0.4	15.0 ± 0.8	116 ± 6
ST16 _{surf}	20.8 ± 1.1	1.6 ± 0.4	16.7 ± 1.1	123 ± 8
ST16 _{deep}	20.6 ± 1.2	1.5 ± 0.4	13.9 ± 1.1	99 ± 7

1009

1010 Supplemental Data - Table S2 (continuation).

SUMMER 2010	²²⁴ Ra dpm·100L ⁻¹	²²³ Ra dpm·100L ⁻¹	²²⁶ Ra dpm·100L ⁻¹	²²⁸ Ra dpm·100L ⁻¹
------------------------	---	---	---	---

ST1 _{surf}	13.3 ± 0.8	1.33 ± 0.17		
ST1 _{deep}	13.9 ± 0.8	0.77 ± 0.11		
ST2 _{surf}	4.5 ± 0.3	0.53 ± 0.07		
ST2 _{deep}	6.8 ± 0.6	0.71 ± 0.12		
ST3 _{surf}	3.65 ± 0.29	0.58 ± 0.08	10.1 ± 1.6	77 ± 5
ST3 _{deep}	6.83 ± 0.45	0.89 ± 0.13	11.4 ± 1.5	53 ± 5
ST4 _{surf}	7.13 ± 0.32	0.54 ± 0.54		
ST4 _{deep}	6.80 ± 0.47	0.90 ± 0.90		
ST102 _{surf}	4.0 ± 0.2	0.86 ± 0.11		
ST102 _{deep}	7.7 ± 0.4	0.97 ± 0.15		
ST113E _{surf}	4.2 ± 0.3	0.76 ± 0.10		
ST113E _{deep}	7.3 ± 0.6	0.75 ± 0.13		
ST113W _{surf}	3.8 ± 0.2	0.95 ± 0.13		
ST113W _{deep}	17.8 ± 1.1	1.29 ± 0.17		
ST5 _{surf}	5.4 ± 0.4	1.52 ± 0.18		
ST5 _{deep}	16.1 ± 0.9	1.51 ± 0.21		
ST6 _{surf}	4.0 ± 0.3	0.76 ± 0.10		
ST6 _{deep}	12.6 ± 1.0	1.29 ± 0.18		
ST7 _{surf}	2.8 ± 0.2	0.72 ± 0.12	12.1 ± 1.3	72 ± 4
ST7 _{deep}	12.2 ± 0.9	0.98 ± 0.14	12.7 ± 1.5	73 ± 5
ST8 _{surf}	5.2 ± 0.4	0.82 ± 0.13		
ST8 _{deep}	15.4 ± 1.0	1.39 ± 0.17		
ST9 _{surf}	20.4 ± 1.9	1.70 ± 0.21		
ST9 _{deep}	27.6 ± 2.6	1.91 ± 0.23		
ST117 _{surf}	10.8 ± 0.5	1.15 ± 0.16		
ST117 _{deep}	16.4 ± 1.0	0.91 ± 0.11		
ST10 _{surf}	11.9 ± 0.8	1.36 ± 0.17		
ST10 _{deep}	28.0 ± 1.7	2.79 ± 0.40		
ST120E _{surf}	5.3 ± 0.4	0.71 ± 0.11	9.9 ± 1.5	71 ± 6
ST120E _{deep}	12.5 ± 0.8	1.22 ± 0.16	13.0 ± 1.4	86 ± 6
ST120W _{surf}	9.7 ± 0.6	1.01 ± 0.15	10.8 ± 1.5	70 ± 6
ST120W _{deep}	13.8 ± 0.8	1.44 ± 0.18	14.7 ± 1.9	91 ± 7
ST11 _{surf}	12.6 ± 0.8	1.45 ± 0.18		
ST11 _{deep}	25.5 ± 1.5	2.24 ± 0.24		
ST12 _{surf}	10.6 ± 0.7	0.89 ± 0.89		
ST12 _{deep}	20.8 ± 1.4	1.44 ± 1.44		
ST13 _{surf}	6.0 ± 0.4	1.13 ± 0.15	7.5 ± 1.4	53 ± 5
ST13 _{deep}	17.1 ± 0.8	1.68 ± 0.19	13.8 ± 1.7	87 ± 7
ST14 _{surf}	11.5 ± 0.6	0.89 ± 0.16		
ST14 _{deep}	14.2 ± 0.8	1.60 ± 0.20		
ST15 _{surf}	5.9 ± 0.5	1.37 ± 0.20		
ST15 _{deep}	15.4 ± 1.3	1.93 ± 0.21		
ST17 _{surf}	14.9 ± 1.0	1.97 ± 0.22		

ST17 _{deep}	21.3 ± 1.3	1.57 ± 0.19		
ST18 _{surf}	23.6 ± 1.2	1.93 ± 0.21		
ST18 _{deep}	24.1 ± 1.2	2.04 ± 0.21		
ST16 _{surf}	19.0 ± 1.2	1.40 ± 0.14	15.0 ± 1.8	109 ± 7
ST16 _{deep}	17.4 ± 1.0	2.11 ± 0.23	14.9 ± 1.5	119 ± 6

1011

1012 SD - Table 3

ST8				
Depth in core (cm)	Water content (%)	Sediment Organic Matter (%)		Mn (mmol·g ⁻¹)
0.5	77.4	6.0		27.1
1.0	68.4	6.1		8.5
1.5	68.3	6.3		10.1
2.0	66.5	6.1		9.1
2.5	65.7	6.1		7.9
3.0	64.6	5.6		7.5
4.0	63.6	5.7		6.5
5.0	64.0	6.0		5.9
6.0	65.1	5.9		5.4

ST16				
Depth in core (cm)	Water content (%)	Sediment Organic Matter (%)		Mn (mmol·g ⁻¹)
0.5	83.4	8.3		10.3
1.0	68.6	6.8		7.4
1.5	65.8	6.4		8.2
2.0	66.5	7.1		7.9
2.5	66.4	6.9		8.8
3.0	66.5	7.9		8.2
4.0	64.6	7.8		8.2
5.0	62.7	7.6		9.8
6.0	62.3	7.7		10.5

1013

1014

ST 8 (2010)			
Depth in core (cm)	^{226}Ra (dpm·g ⁻¹)	^{228}Ra (dpm·g ⁻¹)	^{228}Th (dpm·g ⁻¹)
0-2.0	1.35 ± 0.03	1.98 ± 0.10	2.13 ± 0.06
2.0-4.0	1.46 ± 0.04	1.80 ± 0.12	2.18 ± 0.07
4.0-5.0	1.37 ± 0.03	1.90 ± 0.09	2.04 ± 0.05
5.0-7.0	1.39 ± 0.04	1.91 ± 0.11	1.89 ± 0.06
7.0-9.0	1.24 ± 0.04	2.00 ± 0.10	1.81 ± 0.06
9.0-11.0	1.17 ± 0.04	2.32 ± 0.11	1.80 ± 0.06
11.0-13.0	1.35 ± 0.04	1.97 ± 0.11	1.83 ± 0.06
15.0-17.0	1.32 ± 0.04	2.20 ± 0.11	1.97 ± 0.07

ST 13 (2009)			
Depth in core (cm)	^{226}Ra (dpm·g ⁻¹)	^{228}Ra (dpm·g ⁻¹)	^{228}Th (dpm·g ⁻¹)
0-2.0	1.24 ± 0.03	1.64 ± 0.07	n.m.
2.0-4.0	1.12 ± 0.02	1.54 ± 0.06	n.m.
4.0-6.0	1.22 ± 0.02	1.88 ± 0.06	n.m.
6.0-8.0	1.16 ± 0.02	1.49 ± 0.06	n.m.
8.0-10.0	1.27 ± 0.03	1.70 ± 0.08	n.m.
10.0-12.0	1.40 ± 0.02	1.70 ± 0.05	n.m.
12.0-14.0	1.27 ± 0.03	1.18 ± 0.07	n.m.
14.0-16.0	1.43 ± 0.02	1.85 ± 0.05	n.m.

ST 16 (2010)			
Depth in core (cm)	^{226}Ra (dpm·g ⁻¹)	^{228}Ra (dpm·g ⁻¹)	^{228}Th (dpm·g ⁻¹)
0-1.5	1.09 ± 0.02	1.83 ± 0.02	2.13 ± 0.01
1.5-3.0	0.94 ± 0.04	1.65 ± 0.04	1.73 ± 0.02
3.0-5.0	1.05 ± 0.03	1.81 ± 0.03	2.18 ± 0.01
5.0-7.0	1.19 ± 0.03	2.00 ± 0.03	2.25 ± 0.01
7.0-9.0	1.08 ± 0.03	2.12 ± 0.02	2.11 ± 0.02
9.0-11.0	1.19 ± 0.02	2.10 ± 0.02	2.14 ± 0.01
11.5-13.5	1.13 ± 0.02	2.08 ± 0.02	2.15 ± 0.01
15.0-17.0	1.08 ± 0.03	2.24 ± 0.02	2.40 ± 0.01

ST 120E (2009)			
Depth in core (cm)	^{226}Ra (dpm·g ⁻¹)	^{228}Ra (dpm·g ⁻¹)	^{228}Th (dpm·g ⁻¹)
0-2.0	1.40 ± 0.03	2.50 ± 0.09	n.m.
2.0-4.0	1.34 ± 0.03	2.40 ± 0.08	n.m.
4.0-6.0	1.33 ± 0.02	1.91 ± 0.07	n.m.

6.0-8.0	1.29 ± 0.03	1.99 ± 0.08	n.m.
8.0-10.0	1.36 ± 0.03	1.70 ± 0.08	n.m.
10.0-12.0	1.33 ± 0.02	2.00 ± 0.05	n.m.
12.0-14.0	1.38 ± 0.03	2.40 ± 0.08	n.m.

ST MW
(2009)

Depth in core (cm)	²²⁶ Ra (dpm·g ⁻¹)	²²⁸ Ra (dpm·g ⁻¹)	²²⁸ Th (dpm·g ⁻¹)
0-1.0	0.30 ± 0.01	0.40 ± 0.02	n.m.
1.0-3.0	0.33 ± 0.01	0.34 ± 0.02	n.m.
3.0-5.0	0.36 ± 0.01	0.39 ± 0.02	n.m.
5.0-7.0	0.42 ± 0.01	0.46 ± 0.02	n.m.
7.0-9.0	0.53 ± 0.01	0.60 ± 0.02	n.m.
9.0-11.0	0.40 ± 0.01	0.48 ± 0.02	n.m.
11.0-13.0	0.37 ± 0.01	0.39 ± 0.02	n.m.
13.0-15.0	0.35 ± 0.01	0.38 ± 0.02	n.m.

1016

1017 SD - Table 5

ST 13 (2009)		
Depth in core (cm)	²²⁴ Ra (dpm·L ⁻¹)	²²³ Ra (dpm·L ⁻¹)
0-2.0	14.4 ± 0.7	0.44 ± 0.16
2.0-4.0	14.7 ± 0.9	0.90 ± 0.28
4.0-6.0	12.7 ± 0.9	1.28 ± 0.37
7.0-9.0	14.6 ± 1.1	1.21 ± 0.37
10.0-12.0	17.2 ± 1.0	0.96 ± 0.26
12.0-14.0	13.2 ± 1.0	0.31 ± 0.19
14.0-16.0	17.9 ± 1.1	0.67 ± 0.24
16.0-19.0	17.6 ± 0.7	1.22 ± 0.26

ST 120E
(2010)

Depth in core (cm)	²²⁴ Ra (dpm·L ⁻¹)	²²³ Ra (dpm·L ⁻¹)
0-2.0	16.0 ± 0.7	0.62 ± 0.19
2.0-4.0	22.4 ± 1.3	1.31 ± 0.40
4.0-6.0	16.4 ± 0.9	0.91 ± 0.26
8.0-10.0	15.3 ± 0.8	0.99 ± 0.26
10.0-12.0	14.0 ± 0.9	0.52 ± 0.22
12.0-14.0	19.3 ± 1.1	1.20 ± 0.34
14.0-16.0	17.2 ± 1.0	1.10 ± 0.33

1018

1019