Improving the pollinator pantry: restoration and management of open farmland ponds enhances the complexity of plant-pollinator networks

Highlights

- Pond restoration improves plant-pollinator network complexity.
- Pond management and restoration increased interactions and inter-species links.
- Interactions at restored and managed ponds involved a greater number of plant species.
- Pond management effectively promotes plant-pollinator network diversity.











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Figure S1. Satellite photo of ponds. (a) Overgrover recently restored ponds near the villages of Bodha Baconsthorpe. **(b)** Long-term managed ponds near of Briston.



(b)



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Supplementary Figure S2. Network level metrics produce ponds over 2016-2017. Metrics are presented in three gra Differences between management treatments are labelle ponds, b – between recently restored and overgrown pon restored ponds. The degree of significance is marked wit



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ed by long-term managed, recently restored, and overgrown aphs due to differences in metric measurements. d as a – between long-term managed and overgrown ids, and c – between long-term managed and recently h asterisks: '**' P < 0.01, '***' P < 0.001.





	J
(ISA)	
Network Metric	

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Table S1. Agricultural pond study sites used for plant-pollinatornetwork analysis.Pond names, codes, and map coordinates are givenfor the nine study ponds.

Pond Code	Coordinates
CHFA2	52°54'18" N 001°08'57" E
NROAD	52°53'40" N 001°09'47" E
BAWO2	52°54'02" N 001°09'48" E
WADD10	52°50'36" N 001°02'16" E
WADD17	52°50'39" N 001°02'44" E
WADD23	52°51'40" N 001°03'16" E
SHOOT	52°53'47" N 001°08'25" E
BECK	52°53'42" N 001°08'12" E
SABA	52°54'42" N 001°09'40" E
	Pond Code CHFA2 NROAD BAWO2 WADD10 WADD17 WADD23 SHOOT BECK SABA

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Pond	Category	Hedgerow Length Mean
BAWO2	Overgrown	218.56
CHFA2	Overgrown	292.70
NROAD	Overgrown	272.571
WADD23	Long-term Managed	333.05
WADD17	Long-term Managed	273.52
WADD10	Long-term Managed	456.52
SHOOT	Recently Restored	296.20
BECK	Recently Restored	344.28
SABA	Recently Restored	206.59
Pond	Category	Hedgerow Length Mean
Pond BAWO2	Category Overgrown	Hedgerow Length Mean 335.52
Pond BAWO2 CHFA2	Category Overgrown Overgrown	Hedgerow Length Mean 335.52 315.01
Pond BAWO2 CHFA2 NROAD	Category Overgrown Overgrown Overgrown	Hedgerow Length Mean 335.52 315.01 342.22
Pond BAWO2 CHFA2 NROAD WADD23	Category Overgrown Overgrown Overgrown Long-term Managed	Hedgerow Length Mean 335.52 315.01 342.22 313.74
Pond BAWO2 CHFA2 NROAD WADD23 WADD17	Category Overgrown Overgrown Overgrown Long-term Managed Long-term Managed	Hedgerow Length Mean 335.52 315.01 342.22 313.74 312.15
Pond BAWO2 CHFA2 NROAD WADD23 WADD17 WADD10	Category Overgrown Overgrown Overgrown Long-term Managed Long-term Managed	Hedgerow Length Mean 335.52 315.01 342.22 313.74 312.15 344.10
Pond BAWO2 CHFA2 NROAD WADD23 WADD17 WADD10 SHOOT	Category Overgrown Overgrown Overgrown Long-term Managed Long-term Managed Recently Restored	Hedgerow Length Mean 335.52 315.01 342.22 313.74 312.15 344.10 348.27
Pond BAWO2 CHFA2 NROAD WADD23 WADD17 WADD10 SHOOT BECK	Category Overgrown Overgrown Overgrown Long-term Managed Long-term Managed Recently Restored Recently Restored	Hedgerow Length Mean 335.52 315.01 342.22 313.74 312.15 344.10 348.27 338.35
Pond BAWO2 CHFA2 NROAD WADD23 WADD17 WADD10 SHOOT BECK SABA	Category Overgrown Overgrown Overgrown Long-term Managed Long-term Managed Recently Restored Recently Restored Recently Restored	Hedgerow Length Mean 335.52 315.01 342.22 313.74 312.15 344.10 348.27 338.35 343.83

		Degrees of freedom
Hedgerow Longth (E00 m)	Recently Restored	2
Hedgerow Length (500 m)	Overgrown	2
Grassland Area (500 m)	Recently Restored	2
Glassiand Alea (500 m)	Overgrown	2
Woodland Area (500 m)	Recently Restored	2
	Overgrown	2
Other Freshwater Features	Recently Restored	2
Area (500 m)	Overgrown	2
Arable Field/Pasture Area	Recently Restored	2
(500 m)	Overgrown	2

Human Settlement Area (500	Recently Restored	2
m)	Overgrown	2
Hadgaraw Langth (1000 m)	Recently Restored	2
	Overgrown	2
Grassland Area (1000 m)	Recently Restored	2
Grassiand Area (1000 III)	Overgrown	2
Woodland Area (1000 m)	Recently Restored	2
	Overgrown	2
Other Freshwater Features	Recently Restored	2
Area (1000 m)	Overgrown	2
Arable Field/Pasture Area	Recently Restored	2
(1000 m)	Overgrown	2
Human Settlement Area	Recently Restored	2
(1000 m)	Overgrown	2

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Table S2. Comparison betwee

ponds. Landscape elements for radius from each study pond, an Bonferroni correction.

Table S2a (500 m radius)		
Grassland Area Mean	Woodland Area Mean	Other Freshwater Features Mean
3125.79	48689.91	2296.33
3092.99	34578.20	1525.54
2750.55	8768.28	686.58
2644.16	2606.86	947.36
2058.10	27632.82	1441.57
1687.70	16872.35	1682.42
4985.40	3374.75	1358.66
3392.25	2211.8	1626.04
4524.72	2471.46	1987.62

Table S2b (1000 m radius)

Grassland Area Mean	Woodland Area Mean	Other Freshwater Features Mean
15041.31	51086.25	2386.37
9560.56	68974.87	2515.34
12627.06	58208.48	2403.10
5663.35	97392.37	1759.24
10757.96	25391.74	2057.40
8751.46	30457.13	2168.34
11693.22	15078.92	2128.25
7745.41	8155.51	2256.25
9745.86	3236.32	2378.33

Table S2c (Landscape comparison)

Long-term Managed		
t-value	p-value	Degrees of freedom
-0.764	0.78	NA
-1.562	0.48	2
4.910	0.01	NA
4.543	0.331	2
-1.714	0.87	NA
0.927	0.69	2
4.583	1	NA
0.215	1	2
-0.232	1	NA
-1.886	0.61	2

1.898	0.25	NA
3.588	0.37	2
1.921	0.35	NA
1.05	1	2
0.511	1	NA
1.316	0.27	2
-2.088	0.22	NA
0.301	1	2
4.718	0.225	NA
3.872	0.032	2
0.265	1	NA
-1.931	1	2
-0.407	1	NA
-0.829	1	2

n landcape elements present around Norfolk study

und in **(a)** 500 m radius from each study pond, **(b)** 1000 m id **(c)** management comparisons using pairwise t-tests with

Arable Field/Pasture Area Mean	Human Settlement Area Mean
59492.19	12044.62
44016.56	37875.18
32266.14	32504.34
65618.50	0
59598.95	1509.28
77861.03	1509.28
72560.83	18131.20
85240.93	12537.71
29940.06	63992.22

Arable Field/Pasture Area Mean	Human Settlement Area Mean
44675.32	18554.58
51329.69	20740.74
39440.80	21173.00
58269.36	93529.23
51011.41	22016.04
53698.35	11198.77
70065.09	36750.43
66740.82	29175.63
34848.13	31675.18

Recently Restored	
t-value	p-value
NA	NA
-0.477	1
NA	NA
-2.588	0.087
NA	NA
2.440	0.14
NA	NA
-0.239	1
NA	NA
-1.359	0.94

NA	NA
-0.248	1
NA	NA
-2.003	0.9
NA	NA
5.911	0.69
NA	NA
6.757	0.13
NA	NA
2.318	0.553
NA	NA
-1.369	0.78
NA	NA
-4.168	1

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	Aegopodium podagraria	Alisma plantago- aquatica	Anthriscus sylvestris
Anasimyia spp.	0	4	0
Ancistrocerus parietum	0	0	0
Ancistrocerus trifasciatus	0	0	0
Andrena bicolor	0	0	0
Andrena dorsata	0	1	0
Andrena nigroaenea	0	0	0
Andrena/Lasioglossum spp.	0	1	0
Aphantopus hyperantus	0	0	0
Apis mellifera	6	0	0
Autographa gamma	0	0	0
Bombus hortorum	0	0	1
Bombus hypnorum	0	0	0
Bombus lapidarius	0	0	0
Bombus lucorum/terrestris	0	0	0
Bombus pascuorum	0	0	0
Bombus spp.	0	0	0
Bombus sylvestris	0	0	0
Bombus vestalis	0	0	0
Buathra laborator	0	0	0
Cheilosia illustrata	0	0	0
Cheilosia pagana	0	0	0
Cheilosia spp.	0	0	0
Chrysotoxum bicinctum	0	0	0
Elophila nympheata	0	0	0
Epistrophe diaphana	0	0	0
Epistrophe grossulariae	0	0	0
Episyrphus balteatus	1	12	1

Eristalis abusivus	0	0	0
Eristalis arbustorum	0	0	0
Eristalis horticola	0	0	0
Eristalis interruptus	0	1	0
Eristalis nemorum	0	0	0
Eristalis pertinax	0	0	0
Eristalis spp.	0	0	0
Eupeodes corollae	0	0	0
Gymnomerus laevipes	0	0	0
Helophilus hybridus	0	0	0
Helophilus pendulus	0	0	0
Helophilus trivittatus	0	0	0
Ichneumon sarcitorius	0	0	0
Ichneumon xanthorius	0	0	0
Maniola jurtina	0	0	0
Melangyna umbellatarum	0	0	0
Meliscaeva auricollis	0	0	0
Nymphalis urticae	0	0	0
Ochlodes sylvanus	0	0	0
Pararge aegeria	0	0	0
Pieris brassicae	1	0	0
Pieris rapae	0	0	0
Pipiza austriaca	0	0	0
<i>Pipiza</i> spp.	0	0	0
Platycheirus manicatus	0	0	0
Platycheirus scutatus	0	0	0
Polygonia c-album	0	0	0
Pyronia tithonus	0	0	0
Sphaerophoria scripta	0	0	0
Symmorphus gracilis	0	0	0
Thymelicus lineola	0	0	0
Thymelicus spp.	0	0	0
Thymelicus sylvestris	0	0	0
Vanessa atalanta	1	0	0
Vanessa cardui	0	0	0
Vespula germanica	0	0	0

0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	1
0	0	0
	0 0 0 0 0 0 0	0 0 0 0 0 0 0 0 0 0 0 0 0 0

Table S3. Interaction table of plants an
restored ponds. Columns are plant specspecies. Numbers indicate the number of
given species was observed to directly vis
species.

Arctium minus	Brassica oleracea	Cardamine amara	Chamerion angustifolium	Cirsium arvense	
0	0	0	0	0	-
0	0	0	0	1	
0	0	0	0	1	
0	0	0	0	1	
0	0	0	0	0	
0	0	0	0	0	
0	0	0	0	1	
0	0	0	0	1	
0	0	2	0	12	
0	0	0	0	0	
0	0	3	0	6	
0	0	0	0	1	
0	0	9	0	13	
0	0	0	0	0	
2	0	1	1	5	
0	0	0	0	0	
0	0	0	0	0	
0	0	1	0	0	
0	0	0	0	1	
0	0	0	0	0	
0	0	0	0	0	
0	0	0	0	0	
0	0	0	0	0	
0	0	0	0	0	
0	0	0	0	0	
0	0	0	0	1	
0	0	1	0	3	

0	0	0	0	0
0	0	0	0	5
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	1	0	0
0	0	0	0	1
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	1
0	0	0	0	0
0	0	0	0	1
0	0	0	0	2
0	0	0	0	1
0	0	0	0	1
0	0	0	0	5
0	0	0	0	0
0	0	0	0	0
0	1	0	0	1
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	1
0	0	0	0	0
0	0	0	0	1
0	0	0	0	0
0	0	0	0	2
0	0	0	0	0
0	0	0	0	2
0	0	0	0	4
0	0	0	0	1
0	0	0	0	0

0	0	0	0	0
0	0	0	0	0
0	0	2	0	0
0	0	0	0	0
0	0	0	0	1
0	0	0	0	0
0	0	0	0	0

d pollinators at recently ies and rows are pollinator times a pollinator from a sit the flower face of a plant

Cirsium vulgare	Epilobium hirsutum	Epilobium montanatum	Glechoma hederacea	Heracleum sphondylium
0	0	0	0	0
0	0	0	0	0
0	0	0	0	1
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
1	3	0	0	0
0	0	0	0	0
13	2	0	0	0
0	0	0	0	0
7	6	0	0	0
0	0	0	0	0
16	8	0	0	0
14	3	0	0	1
0	0	0	0	0
5	0	0	0	0
0	0	0	0	1
0	0	0	0	0
0	0	0	0	0
0	0	0	0	1
0	0	0	0	0
0	0	0	0	0
0	0	0	0	2
0	0	0	0	0
7	3	2	0	36

0	0	0	0	0
0	0	0	0	0
1	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	4	0	0	4
0	0	0	0	0
1	1	0	0	0
1	1	0	0	0
0	1	0	0	0
0	0	0	0	1
0	0	0	0	1
1	0	0	0	0
0	0	0	0	0
0	1	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
1	0	0	0	0
1	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	1	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
1	1	0	0	0
0	0	0	0	1
1	0	0	0	0
2	0	0	0	0
1	0	0	0	0
1	0	0	0	0
0	0	0	0	0
0	0	0	0	0

0	0	0	0	0
0	1	0	0	1
0	0	0	0	0
0	0	0	0	5
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0

Hieracium agg.	Hypericum perforatum	Lathyrus pratensis	Lycopus europaeus	Mentha aquatica
0	0	0	5	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	1
0	0	0	0	1
0	0	0	0	0
2	0	0	2	2
0	0	0	0	0
0	0	0	9	24
0	0	0	0	1
0	1	0	0	9
0	0	0	0	0
0	3	1	0	5
0	0	0	0	0
1	7	22	5	37
0	2	0	0	3
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	1	0
0	0	0	0	1
0	0	0	0	0
0	0	0	0	0
0	0	0	7	0
0	0	0	0	0
0	0	0	0	3
0	24	0	0	0

0	0	0	9	12
0	0	0	1	4
0	0	0	1	17
0	0	0	1	18
0	0	0	0	4
0	0	0	0	1
0	0	0	0	0
0	0	0	0	0
1	0	0	0	0
0	0	0	1	3
0	0	0	5	1
0	0	0	2	2
0	0	0	0	0
0	0	0	0	0
2	0	0	0	0
0	0	0	0	0
0	1	0	7	1
0	0	0	2	0
0	0	0	0	0
0	0	0	0	2
2	1	0	0	28
0	0	0	1	1
0	0	0	0	0
0	0	0	0	3
0	0	0	0	0
0	0	0	0	10
0	0	0	0	0
0	0	0	1	1
2	0	0	0	0
0	0	0	0	0
1	0	0	0	1
0	0	0	0	0
0	0	0	0	0
0	0	0	0	1
0	0	0	0	0
0	0	0	1	0

0	0	0	1	0
1	0	0	1	1
0	0	0	0	0
0	0	0	0	17
0	0	0	0	0
0	0	0	0	6
0	0	0	1	0

Mentha spicata	Prunus spinosa	Pulicaria dysenterica	Ranunculus aquatilis	Ranunculus repens
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	4
0	0	0	0	0
0	0	0	0	1
0	0	0	0	0
0	0	0	0	2
0	0	0	0	0
0	0	0	0	3
0	0	0	0	0
1	0	0	0	5
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	2
0	0	0	0	45

0	0	0	0	3
0	0	0	0	3
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	2
0	0	2	0	0
0	0	0	0	0
0	0	0	0	0
1	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	1
0	0	0	0	0
0	0	0	0	0
0	0	0	0	3
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	1	0	0
0	0	0	0	0
0	0	0	1	1
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	18	0	0
0	0	0	0	0
0	0	0	0	0
0	0	5	0	0
0	0	0	0	0
0	0	1	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	
---	----------------------------	---	-----------------------------	
0	0	0	0	
0	0	0	0	
0	0	0	0	
0	0	0	0	
0	0	0	1	
0	0	0	1	
	0 0 0 0 0 0	0 0 0 0 0 0 0 0 0 0 0 0 0 0	000000000000000000000000000	

Rosa canina	Rubus fruticosus agg.	Salix cinerea agg.	Senecio jacobaea	Silene dioica
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	1
0	1	0	0	1
0	4	0	0	0
2	25	1	5	0
0	0	0	0	0
0	12	0	4	0
0	0	0	0	0
0	7	0	2	0
0	0	0	0	0
0	5	0	0	0
0	13	0	1	0
0	2	0	0	0
0	0	0	0	0
0	0	0	2	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	1	0
0	0	0	0	0
0	0	0	0	0
0	1	0	0	0
0	85	0	1	4

0	0	0	1	0
0	0	0	1	0
0	0	0	6	0
0	0	0	9	0
0	0	0	0	0
0	0	0	2	0
0	0	0	1	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	2	0
0	0	0	1	0
0	0	0	0	0
0	0	0	0	0
0	5	0	3	0
0	0	0	0	0
0	0	0	0	0
0	1	0	0	0
0	0	0	0	0
0	0	0	0	0
0	1	0	0	1
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	2	0
0	0	0	0	0
0	2	0	5	0
0	0	0	0	0
0	0	0	0	0
0	0	0	1	0
0	0	0	0	0
0	0	0	0	0

0	0	0	0	0
0	0	0	0	1
0	1	0	1	0
0	0	0	0	0
0	0	0	0	0
0	1	0	0	0
0	0	0	0	0

Stachys sylvatica	Stellaria spp.	Taraxacum agg.	Trifolium dubium	Trifolium repens
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	1	0	0
0	0	0	0	0
1	0	2	0	0
0	0	0	0	0
0	0	0	0	11
0	0	0	0	0
3	0	0	0	3
0	0	0	0	0
4	0	0	0	3
0	0	0	0	11
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	1	0
0	1	0	0	0

0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	1	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	1
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0

0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0

Tripleurosper		
mum	Vicia cracca	Vicia lutea
inodorum		
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
1	4	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
2	0	0

0	0	0
0	0	0
0	0	0
1	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	1
0	0	0
0	0	0
0	0	0
0	1	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0
0	0	0

0	0	0
2	0	0
0	0	0
0	0	0
0	0	0
0	0	0
1	0	0

Improving the pollinator pantry: restoration and management of open farmland ponds enhances the complexity of plant-pollinator networks

Authors: Richard E. Walton, Carl D. Sayer, Helen Bennion & Jan C.

	Anthriscus sylvestris	Arctium minus	Ballota nigra
Andrena bicolor	0	0	0
Andrena dorsata	0	0	0
Andrena nigroaenea	0	0	0
Andrena/Lasioglossum spp.	0	0	0
Aphantopus hyperantus	0	0	0
Apis mellifera	0	2	0
Bombus campestris	0	0	0
Bombus hortorum	0	1	0
Bombus hypnorum	0	0	0
Bombus lapidarius	0	0	0
Bombus lucorum/terrestris	0	4	0
Bombus pascuorum	0	0	4
Bombus spp.	0	0	0
Bombus sylvestris	0	0	0
Cheilosia illustrata	0	0	0
Cheilosia spp.	0	0	0
Chrysotoxum bicinctum	2	0	0
Dolichovespula sylvestris	0	0	0
Elophila nympheata	0	0	0
Epistrophe diaphana	0	0	0
Epistrophe grossulariae	0	0	0
Episyrphus balteatus	0	0	0
Eristalis abusivus	0	0	0
Eristalis arbustorum	0	0	0
Eristalis horticola	0	0	0
Eristalis interruptus	0	0	0
Eristalis intricarius	0	0	0
Eristalis nemorum	0	0	0

Eristalis pertinax	0	0	0
Eristalis spp.	0	0	0
Eupeodes corollae	0	0	0
Gymnomerus laevipes	0	0	0
Helophilus hybridus	0	0	0
Helophilus pendulus	0	0	0
Helophilus trivittatus	0	0	0
Ichneumon sarcitorius	0	0	0
Ichneumon xanthorius	0	0	0
Maniola jurtina	0	0	0
Megachile centuncularis	0	0	0
Melangyna umbellatarum	2	0	0
Meliscaeva auricollis	0	0	0
Neoascia podagrica	0	0	0
Nymphalis urticae	0	0	0
Pieris rapae	0	0	0
<i>Pipiza</i> spp.	0	0	0
Platycheirus scutatus	0	0	0
Polygonia c-album	0	0	0
Pyronia tithonus	0	0	0
Scaeva pyrasti	0	0	0
Sphaerophoria scripta	0	0	0
Thymelicus lineola	0	1	0
Thymelicus sylvestris	0	0	0
Tropidia scita	0	0	0
Vanessa atalanta	0	0	0
Vanessa cardui	0	0	0
Vespula vulgaris	0	0	0
Volucella bombylans	0	0	0
Volucella inanis	0	0	0
Volucella pellucens	0	0	0
Xylota segnis	0	0	0

Table S4. Interaction table of plants and polmanaged ponds. Columns are plant species aspecies. Numbers indicate the number of timesspecies was observed to directly visit the flowe

Centaurea nigra	Chamerion angustifolium	Cirsium arvense	Cirsium vulgare	Convolvulus arvensis
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	3	0	0
0	0	0	0	0
1	6	15	0	0
1	0	0	0	0
3	2	5	2	0
0	0	0	0	0
29	8	19	3	0
7	12	16	5	0
2	4	5	2	0
0	0	0	0	0
0	0	1	0	0
0	0	0	0	0
0	0	0	0	0
0	0	4	0	0
0	0	1	0	0
0	0	1	0	0
0	0	0	0	0
0	0	0	0	0
0	4	1	0	1
0	0	0	0	0
0	0	2	0	0
0	0	1	0	0
0	0	2	0	0
0	0	1	0	0
0	0	0	0	0

0	0	3	0	0
0	0	0	0	0
0	0	1	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	2	0
0	0	0	0	0
0	0	0	0	0
0	0	1	0	0
0	1	5	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	1	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
1	0	1	0	0
0	0	0	0	0
0	0	1	0	1
0	0	5	0	0
1	0	2	3	1
0	0	0	0	0
0	0	1	0	0
0	0	0	0	0
0	0	0	0	0
0	0	1	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0

linators at long-term and rows are pollinator s a pollinator from a given r face of a plant species.

Epilobium hirsutum	Eupatorium cannabinum	Ficaria verna	Galeopsis tetrahit	Glechoma hederacea
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
26	0	1	0	0
0	0	0	0	0
3	0	0	0	0
0	0	0	0	0
3	0	0	1	0
21	1	0	0	1
5	0	0	0	1
1	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
2	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0

1	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
2	0	0	0	0
0	0	0	0	0
1	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
3	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	1	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0

Heracleum sphondylium	Hyacinthoides non-scripta	Hypericum perforatum	Iris psuedacorus	Lycopus europaeus
0	0	0	0	0
1	0	1	0	0
0	0	0	0	0
1	0	2	0	0
1	0	0	0	0
3	0	1	4	2
0	0	0	0	0
1	0	0	6	0
0	0	0	0	0
0	0	1	4	0
2	1	1	19	0
0	2	12	2	2
0	0	3	0	0
0	0	0	0	0
6	0	0	0	0
10	0	0	0	0
4	0	0	0	0
0	0	0	0	0
0	0	0	0	1
0	0	0	0	0
0	0	1	0	1
97	0	5	0	0
0	0	0	0	0
0	0	0	0	0
2	0	1	0	0
4	0	0	0	0
0	0	0	0	0
0	0	0	0	0

0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
1	0	0	0	0
1	0	0	0	0
0	0	0	0	0
1	0	0	0	0
1	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	2	0	1
0	0	2	0	0
1	0	0	0	0
0	0	0	0	1
0	0	0	0	0
1	0	2	0	0
0	0	0	0	0
0	0	0	0	0
1	0	0	0	0
0	0	1	0	1
0	0	0	0	0
0	0	1	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
2	0	0	0	0
1	0	0	0	0
0	0	0	0	0
1	0	3	0	1

Lythrum salicaria	Matricaria chamomila	Mentha aquatica	Oenanthe aquatica	Plantago lanceolata
0	0	0	0	0
0	0	4	0	0
0	0	0	0	1
0	0	8	0	1
0	0	0	3	0
1	0	98	0	0
0	0	2	0	0
2	0	14	0	0
0	0	1	0	0
0	0	5	0	0
0	0	45	0	0
2	0	53	0	0
0	0	11	0	0
0	0	0	0	0
0	0	1	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	1	0	0
0	12	1	0	0
0	2	12	0	0
0	2	14	0	0
0	0	11	1	0
0	0	20	0	0
0	0	2	0	0
0	0	4	0	0

0	0	6	0	0
0	0	0	0	0
1	0	4	0	0
0	0	1	0	0
0	0	3	0	0
0	0	6	0	0
0	0	3	0	0
0	0	0	0	0
0	1	0	0	0
0	0	0	0	0
0	0	1	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	1	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	1	0	0
0	0	12	0	0
0	0	1	0	0
0	0	1	0	0
2	0	2	0	0
0	0	0	0	0
0	0	13	0	0
0	0	2	2	0

Potentilla spp.	Prunella vulgaris	Prunus spinosa	Pulicaria dysenterica	Ranunculus repens
0	0	0	0	0
1	0	0	0	0
0	0	0	0	1
1	0	1	0	2
0	0	0	0	0
0	0	0	0	1
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	4
0	1	1	1	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	1
1	0	0	0	0
0	0	0	0	9
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0

0	0	0	0	0
0	0	0	0	0
0	0	0	0	1
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	1	0	1
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	3
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0

Rubus fruticosus agg.	Sambuca nigra	Senecio jacobaea	Silene dioica	Solanum dulcamara
1	0	0	0	0
0	0	0	0	0
0	0	0	0	0
3	0	1	0	0
3	0	0	1	0
18	0	15	1	0
1	0	0	0	0
6	0	1	0	0
0	0	0	0	0
14	0	13	1	0
20	0	1	3	0
22	0	0	2	2
10	0	2	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
1	0	0	2	0
25	0	0	17	0
0	0	1	0	0
0	1	3	0	0
0	0	14	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0

0	0	0	0	0
0	0	1	0	0
1	0	0	0	0
0	0	0	0	0
0	0	4	0	0
0	0	0	0	0
0	0	1	0	0
0	0	0	0	0
0	0	0	0	0
2	0	0	0	0
0	0	0	0	0
1	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	4	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	1	0	0
0	0	1	0	0
1	0	0	1	0
0	0	0	0	0
1	0	0	0	0
0	0	0	0	0
0	0	0	0	0
1	0	0	1	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0

Stachys palustris	Stachys sylvatica	Tripleurosperm um inodorum	Viburnum opulus	Vicia cracca
0	0	0	0	0
0	0	0	4	0
0	3	0	0	0
0	3	0	4	0
0	1	0	0	0
0	2	0	8	3
0	0	0	0	0
0	6	0	0	6
0	0	0	0	0
0	6	0	0	1
0	8	0	0	3
0	9	0	0	25
0	0	0	0	7
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	1
0	1	0	0	10
0	0	0	0	0
0	0	0	0	0
0	0	0	19	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0

0	0	1	9	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	1	1
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	3	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	1
0	0	0	0	0
0	0	0	0	0
0	0	0	1	0
0	0	0	0	0
0	0	0	0	0
0	1	0	0	0
0	0	0	0	0
0	0	0	0	0
0	0	0	0	2

Improving the pollinator pantry: restoration and management farmland ponds enhances the complexity of plant-pollinator r Authors: Richard E. Walton, Carl D. Sayer, Helen Bennion & Jan (

Anthriscus Arctium minus sylvestris Alomya debellator Andrena dorsata Andrena/Lasioglossum spp. Aphantopus hyperantus Apis mellifera Bombus hortorum Bombus hypnorum **Bombus** lapidarius Bombus lucorum/terrestris Bombus pascuorum Bombus spp. Bombus vestalis Buathra laborator Cheilosia illustrata Cheilosia spp. *Chrysotoxum bicinctum* Epistrophe diaphana Epistrophe elegans *Epistrophe grossulariae Episyrphus balteatus* Eristalis abusivus Eristalis arbustorum Eristalis horticola Eristalis interruptus Eristalis pertinax Eupeodes corollae Eupeodes luniger Helophilus hybridus Helophilus pendulus Ichneumon xanthorius

0	0
0	0
0	0
0	0
0	0
0	0
0	0
0	0
0	0
0	0
0	0
0	0
0	0
0	0
1	0
0	0
0	0
0	0
0	0
0	0
0	0

of open tetworks C. Axmacher

Table S5. Interactplant species and rpollinator from a giv

Centaurea nigra	Cirsium arvense	Cirsium vulgare	Convolvulus arvensis
0	0	0	0
0	0	0	0
0	1	0	1
0	7	0	0
0	1	0	0
0	0	8	0
0	0	0	0
0	7	26	0
0	2	23	0
1	6	7	0
0	0	0	0
0	1	0	0
0	0	1	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	2	2	1
0	0	0	0
0	1	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	1	0	0
0	0	0	1
0	0	0	0

•	•	•	•
0	0	0	0
0	3	0	0
0	0	0	0
0	1	0	0
0	2	2	0
0	0	0	0
1	2	1	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	1	0
0	0	0	0
0	1	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	1	0	0

ion table of plants and pollinators at overgrown ponds. Columns are ows are pollinator species. Numbers indicate the number of times a ven species was observed to directly visit the flower face of a plant species.

Crataegus monogyna	Glechoma hederacea	Hedera helix	Heracleum sphondylium
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	46	2
0	0	0	0
0	0	0	0
0	0	0	0
0	1	0	0
0	0	0	0
0	0	0	2
0	0	0	0
0	0	0	0
0	0	0	1
0	0	0	6
0	0	0	0
0	0	0	4
0	0	0	1
0	0	1	0
0	0	0	37
0	0	0	0
0	0	1	0
0	0	0	0
0	0	0	0
0	0	2	0
0	0	0	4
1	0	0	0
0	0	0	0
0	0	0	3
0	0	0	4

0	0	0	1
0	0	0	0
0	0	0	1
0	0	1	0
0	0	0	0
0	0	3	0
0	0	0	0
0	0	0	1
0	0	0	0
0	0	0	0
1	0	0	0
0	0	0	0
0	0	0	0
0	0	0	1
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	2
0	0	5	0
0	0	0	2
0	0	0	0

Hieracium agg.	Papaver rhoeas	Ranunculus repens	Rubus fruticosus agg.
0	0	0	0
0	0	0	0
0	0	1	0
0	0	0	3
1	0	2	6
0	0	0	18
0	0	0	1
3	0	0	1
0	1	1	7
0	0	0	11
0	0	1	4
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
1	0	0	1
0	0	0	0
0	0	0	0
0	0	0	1
16	0	6	29
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	1	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0

0	0	0	0
0	0	1	4
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	1
2	0	0	0
0	0	0	0
0	0	0	0
1	0	0	0
0	0	0	0
6	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	1
0	0	0	0

Salix cinerea agg.	Sambuca nigra	Senecio jacobaea	Silene dioica
0	0	0	0
0	0	1	0
0	0	2	0
0	0	1	0
0	0	2	1
0	0	0	0
0	0	0	0
0	0	0	0
1	0	4	0
0	0	0	0
0	0	1	0
0	0	0	0
0	0	0	0
0	0	1	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	2	0
0	0	0	0
0	0	2	0
0	0	3	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0

0	0	0	0
0	1	5	0
0	0	0	0
0	0	0	0
0	0	1	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	3	0
0	0	1	0
0	0	0	0
0	0	2	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
Sisymbrium officinale	Taraxacum agg.	Trifolium dubium	Trifolium repens
--------------------------	----------------	------------------	------------------
0	0	0	0
1	0	0	0
1	0	0	0
0	0	0	1
0	0	0	0
0	0	0	12
0	0	0	0
0	0	2	9
0	0	0	10
0	0	0	2
0	0	1	23
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	1
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	1	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0

0	0	0	0
0	0	0	1
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
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0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	1
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0

Tripleurospermum inodorum	Veronica persica	Vicia cracca	Vicia hirsuta
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	5	2
0	0	0	0
0	0	0	0
0	0	4	0
0	1	5	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	1	0
0	0	0	0
0	0	0	0
1	0	0	0
0	0	0	0
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0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0

0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
1	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	4	0
0	0	1	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0
0	0	0	0

Improving the pollinator pantry: restoration and management of open farr complexity of plant-pollinator networks

Authors: Richard E. Walton, Carl D. Sayer, Helen Bennion & Jan C. Axmacher

	species.degree		normalised.	species.stre	interaction.p
Anasimyia spp.		2	0.05555556	0.30696799	-0.346516
Ancistrocerus parietum		1	0.02777778	0.01219512	-0.9878049
Ancistrocerus trifasciatus		2	0.05555556	0.03037694	-0.4848115
Andrena bicolor		2	0.05555556	0.0163445	-0.4918278
Andrena dorsata		2	0.05555556	0.05970493	-0.4701475
Andrena nigroaenea		1	0.02777778	0.1	-0.9
Andrena/Lasioglossum spp.		5	0.13888889	0.48015655	-0.1039687
Aphantopus hyperantus		2	0.05555556	0.03302846	-0.4834858
Apis mellifera		14	0.38888889	3.71677079	0.19405506
Bombus hortorum		11	0.30555556	1.28135763	0.02557797
Bombus hypnorum		1	0.02777778	0.01219512	-0.9878049
Bombus lapidarius		12	0.33333333	1.37864521	0.03155377
Bombus lucorum/terrestris		18	0.5	4.95903559	0.21994642
Bombus pascuorum		18	0.5	5.70471733	0.26137319
Bombus spp.		8	0.22222222	0.74095282	-0.0323809
Bombus sylvestris		1	0.02777778	0.01041667	-0.9895833
Bombus vestalis		2	0.05555556	0.09319149	-0.4534043
Buathra laborator		3	0.08333333	0.06263501	-0.312455
Cheilosia illustrata		1	0.02777778	0.01694915	-0.9830508
Cheilosia pagana		1	0.02777778	0.00414938	-0.9958506
Cheilosia spp.		1	0.02777778	0.01818182	-0.9818182
Chrysotoxum bicinctum		1	0.02777778	0.01612903	-0.983871
Epistrophe diaphana		1	0.02777778	0.03636364	-0.9636364
Epistrophe grossulariae		5	0.13888889	1.05394797	0.0107896
Episyrphus balteatus		16	0.4444444	6.29024353	0.33064022
Eristalis abusivus		4	0.11111111	0.25460852	-0.1863479
Eristalis arbustorum		5	0.13888889	0.14679588	-0.1706408
Eristalis horticola		4	0.11111111	0.19490106	-0.2012747
Eristalis interruptus		5	0.13888889	0.43521194	-0.1129576
Eristalis nemorum		1	0.02777778	0.01659751	-0.9834025
Eristalis pertinax		4	0.11111111	0.10050383	-0.224874
<i>Eristalis</i> spp.		3	0.08333333	0.10239823	-0.2992006
Eupeodes corollae		3	0.08333333	0.41796537	-0.1940115
Gymnomerus laevipes		1	0.02777778	0.07142857	-0.9285714
Helophilus hybridus		5	0.13888889	0.56384511	-0.087231
Helophilus pendulus		5	0.13888889	0.15560103	-0.1688798
Helophilus trivittatus		5	0.13888889	0.09433074	-0.1811339

Ichneumon sarcitorius	1	0.02777778	0.01818182	-0.9818182
Ichneumon xanthorius	3	0.08333333	0.04242513	-0.3191916
Maniola jurtina	6	0.16666667	0.75231445	-0.0412809
Melangyna umbellatarum	1	0.02777778	0.01219512	-0.9878049
Meliscaeva auricollis	6	0.16666667	0.2205837	-0.1299027
Nymphalis urticae	3	0.08333333	0.10008225	-0.2999726
Ochlodes sylvanus	1	0.02777778	0.14285714	-0.8571429
Pararge aegeria	1	0.02777778	0.00829876	-0.9917012
Pieris brassicae	10	0.27777778	1.56087064	0.05608706
Pieris rapae	3	0.08333333	0.03173683	-0.3227544
Pipiza austriaca	2	0.05555556	1.01204819	0.0060241
Pipiza spp.	1	0.02777778	0.01244813	-0.9875519
Platycheirus manicatus	1	0.02777778	0.02380952	-0.9761905
Platycheirus scutatus	1	0.02777778	0.04149378	-0.9585062
Polygonia c-album	1	0.02777778	0.01219512	-0.9878049
Pyronia tithonus	4	0.11111111	0.72002326	-0.0699942
Sphaerophoria scripta	5	0.13888889	0.22175815	-0.1556484
Symmorphus gracilis	1	0.02777778	0.01818182	-0.9818182
Thymelicus lineola	7	0.19444444	0.3868535	-0.0875924
Thymelicus spp.	1	0.02777778	0.0212766	-0.9787234
Thymelicus sylvestris	3	0.08333333	0.07206558	-0.3093115
Vanessa atalanta	5	0.13888889	0.19080831	-0.1618383
Vanessa cardui	1	0.02777778	0.01219512	-0.9878049
Vespula germanica	1	0.02777778	0.01694915	-0.9830508
Vespula spp.	1	0.02777778	0.01694915	-0.9830508
Vespula vulgaris	7	0.19444444	0.52023273	-0.0685382
Volucella bombylans	3	0.08333333	0.10133737	-0.2995542
Volucella inanis	2	0.05555556	0.16144851	-0.4192757
Volucella inflata	1	0.02777778	0.01219512	-0.9878049
Volucella pellucens	4	0.11111111	0.37548613	-0.1561285
Xylota segnis	3	0.08333333	0.17185449	-0.2760485

	species.degree		normalised.c	species.stre	interaction.p
Aegopodium podagraria		4	0.05882353	0.21953154	-0.1951171
Alisima plantago-aquatica		4	0.05882353	1.03040936	0.00760234
Anthriscus sylvestris		3	0.04411765	0.13112208	-0.289626
Arctium minus		1	0.01470588	0.015625	-0.984375
Brassica oleracea		1	0.01470588	0.02631579	-0.9736842
Cardamine amara		9	0.13235294	1.09982829	0.01109203
Chamerion angustifolium		1	0.01470588	0.0078125	-0.9921875
Cirsium arvense		29	0.42647059	11.3169268	0.3557561
Cirsium vulgare		19	0.27941176	4.05372149	0.16072218

Epilobium hirsutum	15	0.22058824	2.51672175	0.10111478
Epilobium montanatum	1	0.01470588	0.00877193	-0.9912281
Glechoma hederacea	1	0.01470588	0.03278689	-0.9672131
Heracleum sphondylium	12	0.17647059	6.05877858	0.42156488
Hieracium agg.	9	0.13235294	1.88923248	0.09880361
Hypericum perforatum	7	0.10294118	0.36337378	-0.0909466
Lathyrus pratensis	2	0.02941176	0.18800403	-0.405998
Lycopus europaeus	21	0.30882353	6.92937523	0.2823512
Mentha aquatica	31	0.45588235	13.3557836	0.39857366
Mentha spicata	2	0.02941176	0.15066964	-0.4246652
Prunus spinosa	1	0.01470588	0.00819672	-0.9918033
Pulicaria dysenterica	5	0.07352941	1.92757629	0.18551526
Ranunculus aquatilis	1	0.01470588	0.5	-0.5
Ranunculus repens	16	0.23529412	3.18556918	0.13659807
Rosa canina	2	0.02941176	0.02947332	-0.4852633
Rubus fruticosus agg.	17	0.25	4.47439509	0.20437618
Salix cinerea agg.	1	0.01470588	0.0106383	-0.9893617
Senecio jacobaea	21	0.30882353	4.51006948	0.16714617
Silene dioica	5	0.07352941	1.19344981	0.03868996
Stachys sylvatica	4	0.05882353	0.10666884	-0.2233328
Stellaria spp.	1	0.01470588	0.00438597	-0.995614
Taraxacum agg.	3	0.04411765	0.23238771	-0.2558708
Trifolium dubium	1	0.01470588	0.125	-0.875
Trifolium repens	6	0.08823529	0.53687875	-0.0771869
Tripleurospermum inodorum	5	0.07352941	0.6332511	-0.0733498
Vicia cracca	3	0.04411765	1.04764344	0.01588115
Vicia lutea	2	0.02941176	0.07962529	-0.4601874

mland ponds enhances the

Table S6. Species-Level Metric ResPollinator species (a) and plant specielisted alongside species-level metrics

Pollinator Species (Table S6a) nestedrank PDI species.specresource.ran PSI node.specia betweennes: 0.9771429 0.9714286 0.7014724 0.14584641 1.671642 0.00049314 0.53731343 1 0.01219512 0.74626866 1 1 1.58209 0 0.58208955 0.9714286 0.9714286 0.6969321 0.01518847 1.462687 0.01849867 0.6969321 0.59701493 0.9714286 0.9714286 0.00817225 1.313433 0.01052996 0.6119403 0.9714286 0.9714286 0.6969321 0.02985247 1.522388 0.00108019 1 0.76119403 1 1 1.940299 0 0.1 0.08014903 0.20895522 0.9571429 0.8857143 0.4887301 1.432836 0.01400524 0.56716418 0.9928571 0.9714286 0.8190587 0.01910569 1.507463 0.00293465 0.04477612 0.9211429 0.6285714 0.3839737 0.18093362 1.104478 0.04531183 0.07462687 0.8879121 0.3443582 0.13041706 0.03180009 0.7142857 1.164179 0.7761194 1 1 1 0.01219512 1.58209 0 0.05970149 0.3030493 0.13993319 0.8923077 0.6857143 1.164179 0.03180009 0.01492537 0.8945055 0.5142857 0.3170462 0.18381229 1.059701 0.10047529 0 0.9297297 0.5142857 0.3469011 0.32279643 1.089552 0.07934991 0.10447761 0.9306122 0.8 0.4462142 0.15118845 1.223881 0.0512628 0.68656716 1 1 1 0.01041667 1.761194 0 0.55223881 0.9942857 0.9714286 0.8451543 0.05099291 1.701493 0.00026832 0.9714286 0.9428571 0.5976143 0.02372327 0.02831236 0.41791045 1.358209 0.79104478 1 1 0.01694915 1.701493 0 1 1 0.00414938 0 0.80597015 1 1 1.552239 1 1 0 0.82089552 0.01818182 1 1.850746 0.8358209 1 1 0.01612903 1.701493 0 1 0.70149254 1 1 1 0.03636364 1.850746 0 0.23880597 0.952381 0.8857143 0.4780914 0.13786758 1.238806 0.01901045 0.4429905 0.02985075 0.9519328 0.5714286 0.50385082 1.164179 0.07747641 0.08379818 0.31343284 0.9690476 0.9142857 0.5984743 1.313433 0.01419557 0.19402985 0.9485714 0.8857143 0.4942822 0.03662714 1.179104 0.02931687 0.9865546 0.7138467 0.07229611 0.01578245 0.32835821 0.9142857 1.298507 0.17910448 0.8857143 0.6615998 0.09554039 1.343284 0.9809524 0.01086759 0.65671642 1 1 1 0.01659751 1.552239 0 0.9428571 0.5070926 0.02614305 0.37313433 0.9142857 1.373134 0.00854091 0.43283582 0.9714286 0.9428571 0.5976143 0.04411808 1.447761 0.00444626 0.9428571 0.6248809 0.10242905 0.00787462 0.3880597 0.9642857 1.656716 0.07142857 0.85074627 1 1 1 1.880597 0 0.26865672 0.9619048 0.8857143 0.4942822 0.08410591 1.328358 0.01557888 0.2238806 0.9714286 0.8857143 0.5482439 0.05268421 1.268657 0.0208562 0.28358209 0.9285714 0.8857143 0.4498137 0.01950397 1.164179 0.03218183

0.86567164	1	1	1	0.01818182	1.850746	0
0.47761194	0.9428571	0.9428571	0.5606119	0.01414171	1.373134	0.02925323
0.14925373	0.9485714	0.8571429	0.4498137	0.08003591	1.328358	0.01634184
0.88059701	1	1	1	0.01219512	1.58209	0
0.1641791	0.9714286	0.8571429	0.5447871	0.07176695	1.179104	0.03144734
0.40298507	0.9828571	0.9428571	0.6734771	0.04723537	1.313433	0.01720292
0.89552239	1	1	1	0.14285714	2.029851	0
0.71641791	1	1	1	0.00829876	1.552239	0
0.08955224	0.9897959	0.7428571	0.7337745	0.12738572	1.208955	0.03767413
0.49253731	0.9428571	0.9428571	0.5606119	0.01057894	1.358209	0.01108597
0.62686567	0.9714286	0.9714286	0.6969321	0.5060241	1.776119	0
0.67164179	1	1	1	0.01244813	1.552239	0
0.91044776	1	1	1	0.02380952	1.791045	0
0.64179104	1	1	1	0.04149378	1.552239	0
0.92537313	1	1	1	0.01219512	1.58209	0
0.34328358	0.9952381	0.9142857	0.8568027	0.57396936	1.41791	0.00684927
0.29850746	0.9285714	0.8857143	0.4498137	0.05669619	1.328358	0.01802478
0.94029851	1	1	1	0.01818182	1.850746	0
0.11940299	0.9314286	0.8285714	0.4342026	0.08735187	1.208955	0.02711094
0.73134328	1	1	1	0.0212766	1.731343	0
0.44776119	0.9714286	0.9428571	0.5976143	0.02411396	1.462687	0.00492576
0.25373134	0.9714286	0.8857143	0.5411628	0.04214372	1.253731	0.01714973
0.95522388	1	1	1	0.01219512	1.58209	0
0.97014925	1	1	1	0.01694915	1.701493	0
0.98507463	1	1	1	0.01694915	1.701493	0
0.13432836	0.9142857	0.8285714	0.3635146	0.10074338	1.238806	0.07458497
0.46268657	0.9714286	0.9428571	0.5976143	0.04533434	1.567164	0.00228881
0.52238806	0.9915966	0.9714286	0.7992027	0.07516889	1.41791	0.02374581
1	1	1	1	0.01219512	1.58209	0
0.35820896	0.9857143	0.9142857	0.6831301	0.05555194	1.402985	0.00721803
0.50746269	0.9428571	0.9428571	0.5606119	0.05728483	1.58209	0.00284587

Plant Species (Table S6b)

estedrank	PDI	species.speci	resource.ran P	SI	node.specia	betweennes:
0.51428571	0.9925373	0.6882895	0.9552239	1	1.314286	0.02813892
0.45714286	0.9925373	0.70181	0.9552239	1	1.514286	0.00147113
0.6	0.9701493	0.5686678	0.9701493	1	1.514286	0.00176724
0.77142857	1	1	1	1	1.514286	0
0.82857143	1	1	1	1	1.771429	0
0.25714286	0.973466	0.4373871	0.880597	1	1.114286	0.04921332
0.85714286	1	1	1	1	1.514286	0
0.02857143	0.9207807	0.2503161	0.5820896	1	1.028571	0.10559086
0.11428571	0.9410841	0.337225	0.7313433	1	1.057143	0.07387696

0.2	0.9365672	0.3072328	0.7910448	1	1.114286	0.04921332
0.8	1	1	1	1	1.571429	0
0.74285714	1	1	1	1	1.514286	0
0.22857143	0.9921227	0.6615746	0.8358209	1	1.428571	0.00680921
0.28571429	0.9104478	0.3306829	0.880597	1	1.2	0.03830471
0.31428571	0.9906716	0.6424935	0.9104478	1	1.228571	0.03700208
0.62857143	0.9993216	0.9568609	0.9850746	1	1.485714	0.00080972
0.08571429	0.9170813	0.2762849	0.7014925	1	1.171429	0.03834277
0	0.9177088	0.2569691	0.5522388	1	1.085714	0.08650489
0.68571429	0.9850746	0.70181	0.9850746	1	1.514286	0
0.88571429	1	1	1	1	1.514286	0
0.37142857	0.9925373	0.6923231	0.9402985	1	1.685714	0.00163068
0.91428571	1	1	1	1	2.028571	0
0.17142857	0.9873964	0.5500555	0.7761194	1	1.057143	0.21587999
0.65714286	0.9925373	0.7408927	0.9850746	1	1.428571	0.00586833
0.14285714	0.9812116	0.4816411	0.761194	1	1.028571	0.10559086
0.94285714	1	1	1	1	1.628571	0
0.05714286	0.9308005	0.2745403	0.7014925	1	1.171429	0.04175981
0.4	0.9776119	0.5188966	0.9402985	1	1.228571	0.03660822
0.48571429	0.9776119	0.5381006	0.9552239	1	1.257143	0.02140428
0.97142857	1	1	1	1	1.571429	0
0.57142857	0.9850746	0.6047079	0.9701493	1	1.571429	0.0012217
1	1	1	1	1	1.885714	0
0.34285714	0.972863	0.5147279	0.9253731	1	1.257143	0.0242102
0.42857143	0.9626866	0.4614272	0.9402985	1	1.285714	0.01926441
0.54285714	0.988806	0.648107	0.9701493	1	1.342857	0.00951639
0.71428571	0.9850746	0.70181	0.9850746	1	1.514286	0

ults of Bipartite Interactions at Recently Restored Ponds.

s (b) involved in mutualistic interactions during 2016-2017 are computed through the bipartite package in R.

weighted.be	closeness	weighted.clc F	isher.alpha partner.dive	effective.par	proportional
0	0.01283284	0.0092977 /	<i>IA</i> 0.6869616	1.987667	0.13458555
0	0.01369797	0.00170178 /	<i>IA</i> 0	1	0.06771031
0	0.01485149	0.00320647 /	<i>IA</i> 0.6931472	2	0.13542063
0	0.01629338	0.00321133 /	<i>IA</i> 0.6931472	2	0.13542063
0	0.01427473	0.00302493 /	<i>IA</i> 0.6931472	2	0.13542063
0	0.01023743	0.00168523 /	<i>IA</i> 0	1	0.06771031
0.00046342	0.01513986	0.01093297 /	IA 1.4708085	4.352753	0.29472626
0	0.01441892	0.00717045 /	<i>IA</i> 0.5004024	1.649385	0.11168037
0.06611514	0.01831203	0.03063997 /	<i>IA</i> 2.0689165	7.916242	0.5360112
0.00473722	0.01773527	0.02715063 /	<i>IA</i> 2.1013948	8.177568	0.55370569
0	0.01369797	0.00170178 /	<i>IA</i> 0	1	0.06771031
0.06302564	0.01773527	0.02809714 /	IA 2.2927877	9.902505	0.6705017
0.26384264	0.01874459	0.0324109 /	<i>IA</i> 2.3841898	10.850268	0.73467506
0.21214514	0.01845622	0.03146787 /	IA 2.3193235	10.168793	0.68853216
0	0.01715851	0.02483893 /	IA 1.6910768	5.42532	0.3673501
0	0.0119677	0.0032312 /	<i>IA</i> 0	1	0.06771031
0	0.01254446	0.00830181 /	<i>IA</i> 0.4505612	1.569193	0.10625052
0	0.01586081	0.00587594 /	IA 1.0397208	2.828427	0.19151369
0	0.01254446	0.00169206 /	<i>IA</i> 0	1	0.06771031
0	0.01398635	0.00168903 /	<i>IA</i> 0	1	0.06771031
0	0.01119869	0.00168283 /	<i>IA</i> 0	1	0.06771031
0	0.01254446	0.00168997 /	<i>IA</i> 0	1	0.06771031
0	0.01119869	0.0032234 /	<i>IA</i> 0	1	0.06771031
0	0.01701432	0.00912983 /	IA 1.4941751	4.45566	0.30169412
0.31348048	0.01773527	0.0352174 /	IA 1.8655086	6.45922	0.43735584
0.00370739	0.01629338	0.02024287 /	IA 1.1032863	3.014055	0.2040826
0	0.01759108	0.01429727 /	IA 1.4327571	4.190236	0.28372219
0	0.01643757	0.0201731 /	IA 0.8622685	2.368528	0.16037374
0.02656963	0.016005	0.02126416 /	IA 1.007807	2.739586	0.18549825
0	0.01398635	0.00602238 /	<i>IA</i> 0	1	0.06771031
0	0.01571662	0.00804969 /	IA 1.3296613	3.779763	0.25592895
0	0.01499567	0.00458473 /	IA 1.0397208	2.828427	0.19151369
0	0.01297703	0.01011263 /	IA 0.9649629	2.62469	0.17771861
0	0.01081419	0.00168812 /	<i>IA</i> 0	1	0.06771031
0	0.01614919	0.00886556 /	IA 1.4750763	4.371369	0.29598679
0	0.01672594	0.01156355 /	IA 1.3592367	3.89322	0.26361118
0	0.01773527	0.00899116 /	IA 1.549826	4.710651	0.31895963

0	0.01119869	0.00168283 NA	0	1	0.06771031
0	0.01571662	0.00459494 NA	1.0986123	3	0.20313094
0.00054924	0.01614919	0.01415647 NA	1.6307991	5.107955	0.34586123
0	0.01369797	0.00170178 NA	0	1	0.06771031
0	0.01759108	0.01436876 NA	1.4306853	4.181564	0.28313501
0	0.01629338	0.0100619 NA	0.9002561	2.460233	0.16658315
0	0.0097568	0.00167967 NA	0	1	0.06771031
0	0.01398635	0.00324623 NA	0	1	0.06771031
0.03038001	0.0173027	0.02392799 NA	1.1457961	3.144944	0.21294516
0	0.01586081	0.00459147 NA	1.0986123	3	0.20313094
0	0.01182351	0.00168614 NA	0.6931472	2	0.13542063
0	0.01398635	0.00468644 NA	0	1	0.06771031
0	0.01167932	0.00168873 NA	0	1	0.06771031
0	0.01398635	0.01236921 NA	0	1	0.06771031
0	0.01369797	0.00170178 NA	0	1	0.06771031
0.0018537	0.01528405	0.01431461 NA	0.5670609	1.763078	0.11937854
0.00053208	0.01614919	0.00904206 NA	1.549826	4.710651	0.31895963
0	0.01119869	0.00168283 NA	0	1	0.06771031
0.01259826	0.0173027	0.01296031 NA	1.7233917	5.603502	0.37941486
0	0.01225608	0.00324629 NA	0	1	0.06771031
0	0.01485149	0.00546702 NA	1.0397208	2.828427	0.19151369
0	0.01687013	0.00966953 NA	1.3862944	4	0.27084125
0	0.01369797	0.00170178 NA	0	1	0.06771031
0	0.01254446	0.00169206 NA	0	1	0.06771031
0	0.01254446	0.00169206 NA	0	1	0.06771031
0	0.01701432	0.00791868 NA	1.9061547	6.727171	0.45549888
0	0.01384216	0.00591798 NA	1.0397208	2.828427	0.19151369
0	0.01528405	0.01747316 NA	0.5359599	1.709088	0.11572288
0	0.01369797	0.00170178 NA	0	1	0.06771031
0	0.01542824	0.01015929 NA	1.0027183	2.725681	0.18455671
0	0.01369797	0.00452985 NA	1.0986123	3	0.20313094

weighted.be	closeness	weighted.clc	Fisher.alpha p	oartner.dive	effective.par	proportional
0	0.02908314	0.01044189	2.76E+00	1.0027183	2.725681	0.2319049
0	0.0256326	0.01339469	1.59E+00	0.9257019	2.523639	0.23209664
0	0.0256326	0.00470085	5.37E+08	1.0986123	3	0.25970664
0	0.0256326	0.00335008	7.96E-01	0	1	0.11044004
0	0.02152481	0.00174744	1.34E+08	0	1	0.03278689
0	0.03253368	0.01863875	5.04E+00	1.8632504	6.44465	0.38778257
0	0.0256326	0.00175215	1.34E+08	0	1	0.11044004
0	0.03401249	0.02587812	1.60E+01	2.9212826	18.565085	0.48232286
0.03514938	0.03351955	0.02789455	7.18E+00	2.3382615	10.363204	0.53404439

0	0.03253368	0.02160994	8.35E+00	2.4292023	11.349824	0.55544599
0	0.02464673	0.00331442	7.96E-01	0	1	0.19672131
0	0.0256326	0.00617073	4.28E-01	0	1	0.10526316
0	0.0271114	0.02263107	4.73E+00	1.3894179	4.012514	0.26063221
0	0.03105488	0.01169582	1.09E+01	2.1439522	8.533096	0.25865894
0	0.03056195	0.02210903	2.49E+00	1.2385156	3.450488	0.46543218
0	0.02612553	0.01993611	5.26E-01	0.1788449	1.195835	0.1539183
0.02899824	0.03154781	0.02326272	1.17E+01	2.6731619	14.485699	0.35499627
0.36643234	0.03302662	0.03192892	9.46E+00	2.820419	16.783882	0.56948149
0	0.0256326	0.00330244	2.68E+08	0.6931472	2	0.11647972
0	0.0256326	0.00175317	1.34E+08	0	1	0.10526316
0	0.02300361	0.01869626	1.81E+00	1.0195341	2.771903	0.07247627
0	0.01741702	0.0017636	1.34E+08	0	1	0.00172563
0.05975395	0.03351955	0.02623124	5.90E+00	1.8527567	6.377376	0.48264499
0	0.0271114	0.00476428	2.62E+00	0.6365142	1.889882	0.18636756
0.50966608	0.03401249	0.03165348	4.50E+00	1.9033053	6.70803	0.60025255
0	0.02366086	0.00174977	1.34E+08	0	1	0.0811044
0	0.03154781	0.02392139	1.12E+01	2.6953716	14.811022	0.45205823
0	0.03056195	0.01007828	3.98E+00	1.4184837	4.130852	0.34253667
0	0.03006901	0.01200045	2.47E+00	1.2798542	3.596115	0.35030198
0	0.02464673	0.00174235	1.34E+08	0	1	0.19672131
0	0.02464673	0.00471855	5.45E+00	1.0397208	2.828427	0.09749784
0	0.01922445	0.00175862	1.34E+08	0	1	0.0069025
0	0.03006901	0.02123541	2.22E+00	1.4749013	4.370604	0.32953881
0	0.02957608	0.00663212	7.82E+00	1.549826	4.710651	0.34253667
0	0.02859021	0.00840937	1.99E+00	0.9556999	2.60049	0.21656601
0	0.0256326	0.00330369	2.68E+08	0.6931472	2	0.11734254

proportional d				
0.06643658	0.545867			
0.07075065	0.19655745			
0.11820535	0.21673585			
0.27868853	0.07919875			
0.22346851	0.22035554			
0.00862813	0.58018772			
0.23805004	0.33545545			
0.2364107	0.17500717			
0.64621005	0.19478036			
0.64066544	0.16989287			
0.07075065	0.19655745			
0.56901111	0.19395459			
0.72174288	0.13880833			
0.59910888	0.26876881			
0.44756615	0.27709412			
0.16566005	0.15020965			
0.10267472	0.36652801			
0.17169974	0.22443691			
0.05090595	0.25657458			
0.20793788	0			
0.0474547	0.26937438			
0.05349439	0.24753196			
0.0474547	0.38296484			
0.4761648	0.16861954			
0.42392867	0.50633677			
0.37045729	0.27968832			
0.45470233	0.17303201			
0.34143227	0.27215637			
0.31633592	0.31559259			
0.20793788	0.16800373			
0.31334484	0.18468714			
0.14754098	0.30273372			
0.08714409	0.47364968			
0.01207938	0.51884138			
0.37791199	0.18946167			
0.32174288	0.25554058			
0.41932701	0.11482838			

0.0474547	0.26937438
0.18981881	0.16045796
0.37513867	0.25249111
0.07075065	0.19655745
0.33458647	0.28155993
0.2466566	0.28564692
0.00603969	0.64521747
0.20793788	0.10789064
0.36583261	0.29223904
0.33994823	0.07719415
0.07247627	0.58865012
0.20793788	0.13915286
0.03623814	0.31854002
0.20793788	0.28155189
0.07075065	0.19655745
0.14528124	0.74983888
0.25366695	0.27489764
0.0474547	0.26937438
0.39491448	0.26740186
0.0811044	0.28317951
0.17515099	0.22523628
0.33811475	0.21914925
0.07075065	0.19655745
0.05090595	0.25657458
0.05090595	0.25657458
0.28634599	0.32019863
0.24072476	0.26736837
0.25539258	0.32182426
0.07075065	0.19655745
0.39325089	0.18525915
0.1285591	0.31685397

proportional d

0.12335037	0.29926515
0.11420698	0.38279864
0.13576464	0.21875087
0.04525488	0.19098706
0.04525488	0.33001389
0.29165188	0.26674059
0.04525488	0.1063324
0.8401607	0.30651857
0.46898558	0.2366185

0.51363496	0.17044411
0.04525488	0.07868392
0.04525488	0.3035243
0.18158582	0.45897919
0.38616423	0.40712338
0.15615142	0.19440153
0.05411738	0.48830918
0.65554857	0.44465236
0.75955257	0.38660622
0.09050976	0.33881024
0.04525488	0.11517495
0.12544214	0.83832454
0.04525488	0.8723332
0.28860738	0.25634849
0.08552637	0.17109796
0.30357109	0.26041123
0.04525488	0.16319662
0.67027104	0.30506602
0.18694121	0.23218746
0.16274178	0.16918855
0.04525488	0
0.12800013	0.37845553
0.04525488	0.6169996
0.19779118	0.36114741
0.21317994	0.32420716
0.11768487	0.30894121
0.09050976	0.27606203

Improving the pollinator pantry: restoration and management of open f complexity of plant-pollinator networks

Authors: Richard E. Walton, Carl D. Sayer, Helen Bennion & Jan C. Axmach

	species.degree	normalised.c	species.stre	interaction.p
Andrena bicolor	1	0.02702703	0.00769231	-0.9923077
Andrena dorsata	5	0.13513514	0.63335804	-0.0733284
Andrena nigroaenea	3	0.08108108	1.12653563	0.04217854
Andrena/Lasioglossum spp.	7	0.18918919	0.47808466	-0.0745593
Aphantopus hyperantus	5	0.13513514	0.59162896	-0.0816742
Apis mellifera	19	0.51351351	3.48929717	0.13101564
Bombus campestris	3	0.08108108	0.03547009	-0.32151
Bombus hortorum	14	0.37837838	1.25005457	0.01786104
Bombus hypnorum	1	0.02702703	0.00277778	-0.9972222
Bombus lapidarius	15	0.40540541	3.13878812	0.14258588
Bombus lucorum/terrestris	21	0.56756757	6.66039413	0.26954258
Bombus pascuorum	18	0.48648649	5.45189907	0.24732773
Bombus spp.	6	0.16216216	0.34954458	-0.1084092
Bombus sylvestris	1	0.02702703	0.01020408	-0.9897959
Cheilosia illustrata	2	0.05405405	0.0450313	-0.4774844
Cheilosia spp.	1	0.02702703	0.07042254	-0.9295775
Chrysotoxum bicinctum	3	0.08108108	0.56898534	-0.1436716
Dolichovespula sylvestris	1	0.02702703	0.01020408	-0.9897959
Epistrophe diaphana	1	0.02702703	0.04545455	-0.9545455
Epistrophe grossulariae	7	0.18918919	0.73352917	-0.0380673
Episyrphus balteatus	12	0.32432432	3.35628238	0.19635687
Eristalis abusivus	3	0.08108108	0.16685341	-0.2777155
Eristalis arbustorum	5	0.13513514	1.22456316	0.04491263
Eristalis horticola	8	0.21621622	0.92168281	-0.0097896
Eristalis interruptus	3	0.08108108	0.10413273	-0.2986224
Eristalis intricarius	2	0.05405405	0.01575964	-0.4921202
Eristalis nemorum	1	0.02702703	0.01111111	-0.9888889
Eristalis pertinax	5	0.13513514	1.26198479	0.05239696
<i>Eristalis</i> spp.	1	0.02702703	0.01587302	-0.984127
Eupeodes corollae	5	0.13513514	0.19946205	-0.1601076
Gymnomerus laevipes	1	0.02702703	0.00277778	-0.9972222
Helophilus hybridus	3	0.08108108	0.07886765	-0.3070441
Helophilus pendulus	3	0.08108108	0.14135598	-0.2862147
Helophilus trivittatus	2	0.05405405	0.02420635	-0.4878968
Ichneumon sarcitorius	1	0.02702703	0.00704225	-0.9929577
Ichneumon xanthorius	3	0.08108108	0.07606987	-0.3079767

Maniola jurtina	6	0.16216216	0.1617327	-0.1397112
Megachile centuncularis	1	0.02702703	0.00277778	-0.9972222
Melangyna umbellatarum	3	0.08108108	0.52239819	-0.1592006
Meliscaeva auricollis	4	0.10810811	0.54253057	-0.1143674
Neoascia podagrica	1	0.02702703	0.05263158	-0.9473684
Nymphalis urticae	3	0.08108108	0.083913	-0.3053623
Pieris rapae	1	0.02702703	0.11111111	-0.8888889
Pipiza spp.	1	0.02702703	0.06349206	-0.9365079
Platycheirus scutatus	2	0.05405405	0.05967383	-0.4701631
Polygonia c-album	1	0.02702703	0.00277778	-0.9972222
Pyronia tithonus	2	0.05405405	0.0324263	-0.4837868
Scaeva pyrasti	1	0.02702703	0.00704225	-0.9929577
Sphaerophoria scripta	6	0.16216216	0.54095498	-0.0765075
Thymelicus lineola	4	0.10810811	0.20856009	-0.19786
Thymelicus sylvestris	7	0.18918919	0.62092516	-0.0541535
Tropidia scita	1	0.02702703	0.00277778	-0.9972222
Vanessa atalanta	5	0.13513514	0.57345195	-0.0853096
Vanessa cardui	1	0.02702703	0.00277778	-0.9972222
Vespula vulgaris	1	0.02702703	0.00277778	-0.9972222
Volucella bombylans	8	0.21621622	0.48540987	-0.0643238
Volucella inanis	1	0.02702703	0.00704225	-0.9929577
Volucella pellucens	1	0.02702703	0.03611111	-0.9638889
Xylota segnis	6	0.16216216	0.56932296	-0.0717795

	species.degree	normalised.c	species.stre	interaction.p
Anthriscus sylvestris	2	0.03389831	0.7	-0.15
Arctium minus	4	0.06779661	0.17497815	-0.2062555
Ballota nigra	1	0.01694915	0.02564103	-0.974359
Centaurea nigra	8	0.13559322	1.21874334	0.02734292
Chamerion angustifolium	7	0.11864407	0.33495393	-0.0950066
Cirsium arvense	25	0.42372881	6.4239321	0.21695728
Cirsium vulgare	6	0.10169492	0.62521294	-0.0624645
Convolvulus arvensis	3	0.05084746	0.23046448	-0.2565118
Epilobium hirsutum	11	0.18644068	1.26724254	0.02429478
Eupatorium cannabinum	2	0.03389831	0.06828035	-0.4658598
Ficaria verna	1	0.01694915	0.00480769	-0.9951923
Galeopsis tetrahit	1	0.01694915	0.00892857	-0.9910714
Glechoma hederacea	2	0.03389831	0.0121906	-0.4939047
Heracleum sphondylium	21	0.3559322	7.57303772	0.3130018
Hyacinthoides non-scripta	2	0.03389831	0.01860086	-0.4906996
Hypericum perforatum	16	0.27118644	3.08883134	0.13055196
Iris psuedacorus	5	0.08474576	0.28104043	-0.1437919

Lycopus europaeus	7	0.11864407	1.56334499	0.08047786
Lythrum salicaria	5	0.08474576	0.33095712	-0.1338086
Matricaria chamomila	4	0.06779661	0.62314953	-0.0942126
Mentha aquatica	34	0.57627119	18.8904933	0.52619098
Oenanthe aquatica	3	0.05084746	0.53475936	-0.1550802
Plantago lanceolata	1	0.01694915	0.2	-0.8
Potentilla spp.	2	0.03389831	0.21590909	-0.3920455
Prunella vulgaris	1	0.01694915	0.00578035	-0.9942197
Prunus spinosa	3	0.05084746	0.28270342	-0.2390989
Pulicaria dysenterica	1	0.01694915	0.00578035	-0.9942197
Ranunculus repens	9	0.15254237	1.92239461	0.10248829
Rubus fruticosus agg.	18	0.30508475	3.64561822	0.14697879
Sambucus nigra	1	0.01694915	0.04545455	-0.9545455
Senecio jacobaea	15	0.25423729	3.82449525	0.18829968
Silene dioica	9	0.15254237	0.67482818	-0.0361302
Solanum dulcamara	1	0.01694915	0.01282051	-0.9871795
Stachys sylvatica	9	0.15254237	1.06406884	0.00711876
Tripleurospermum inodorum	1	0.01694915	0.05	-0.95
Viburnum opulus	7	0.11864407	1.97048026	0.13864004
Vicia cracca	11	0.18644068	1.08007605	0.00727964

armland ponds enhances the

Table S7. Species-Level Metric ResPonds. Pollinator species (a) and plan2017 are listed alongside species-leve

Pollinator Species (Table S7a)

nestedrank	PDI	resource.ran	species.spe	PSI	node.specia	betweennes
0.74137931	1	1	1	0.00769231	1.706897	0
0.31034483	0.9513889	0.8888889	0.5191467	0.08485073	1.189655	0.02668798
0.48275862	0.9814815	0.9444444	0.651494	0.25773956	1.793103	0.00097773
0.13793103	0.9375	0.8333333	0.4158276	0.04523285	1.086207	0.0564203
0.32758621	0.9444444	0.8888889	0.4885521	0.18197587	1.431034	0.01240427
0.01724138	0.9688209	0.5	0.4882945	0.24206902	1	0.09864381
0.51724138	0.9722222	0.9444444	0.5980292	0.01025641	1.327586	0.00857476
0.06896552	0.9126984	0.6388889	0.3063008	0.08623594	1.068966	0.05463178
0.75862069	1	1	1	0.00277778	1.431034	0
0.05172414	0.9204981	0.6111111	0.3395271	0.2880231	1.086207	0.0564203
0	0.9209877	0.4444444	0.3210825	0.24561975	1.034483	0.06513332
0.03448276	0.9460168	0.5277778	0.3899353	0.24570861	1.12069	0.04895781
0.18965517	0.9419192	0.8611111	0.4740477	0.06579555	1.224138	0.02564773
0.77586207	1	1	1	0.01020408	1.586207	0
0.56896552	0.9953704	0.9722222	0.865043	0.03661413	1.275862	0.0148817
0.67241379	1	1	1	0.07042254	1.655172	0
0.43103448	0.9583333	0.9444444	0.5849976	0.12759414	1.37931	0.01158256
0.79310345	1	1	1	0.01020408	1.586207	0
0.81034483	1	1	1	0.04545455	1.862069	0
0.17241379	0.9166667	0.8333333	0.3644345	0.10031184	1.241379	0.0288016
0.0862069	0.9753723	0.6944444	0.5480789	0.52631976	1.051724	0.0637543
0.4137931	0.9930556	0.9444444	0.8079848	0.04341114	1.344828	0.00858167
0.25862069	0.984127	0.8888889	0.6531867	0.089246	1.224138	0.01942532
0.10344828	0.9532164	0.8055556	0.4944478	0.23058085	1.068966	0.05241736
0.39655172	0.9916667	0.9444444	0.7815257	0.0486386	1.155172	0.02764994
0.60344828	0.9861111	0.9722222	0.7370277	0.00710506	1.258621	0.01118808
0.68965517	1	1	1	0.01111111	1.431034	0
0.27586207	0.9660494	0.8888889	0.5487359	0.15032713	1.241379	0.01343097
0.82758621	1	1	1	0.01587302	1.758621	0
0.34482759	0.9722222	0.8888889	0.5416667	0.02909942	1.155172	0.03689722
0.84482759	1	1	1	0.00277778	1.431034	0
0.46551724	0.9722222	0.944444	0.6243052	0.03575131	1.206897	0.02740245
0.44827586	0.9861111	0.944444	0.7017517	0.03803738	1.258621	0.01615336
0.5862069	0.9907407	0.9722222	0.7839537	0.01021825	1.362069	0.0074389
0.86206897	1	1	1	0.00704225	1.655172	0
0.55172414	0.9444444	0.9444444	0.5610836	0.02535662	1.37931	0.01068211

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0.20689655	0.9611111	0.8611111	0.4787136	0.03421756	1.448276	0.00476633
0.87931034	1	1	1	0.00277778	1.431034	0
0.53448276	0.9722222	0.9444444	0.5980292	0.25559955	1.637931	0.00073227
0.37931034	0.9583333	0.9166667	0.509902	0.11903243	1.655172	0.00821142
0.72413793	1	1	1	0.05263158	1.741379	0
0.5	0.9814815	0.9444444	0.651494	0.04344927	1.396552	0.00957528
0.89655172	1	1	1	0.11111111	1.896552	0
0.70689655	1	1	1	0.06349206	1.758621	0
0.62068966	0.9861111	0.9722222	0.7370277	0.03743514	1.5	0.00698516
0.9137931	1	1	1	0.00277778	1.431034	0
0.63793103	0.9722222	0.9722222	0.6972167	0.01621315	1.568966	0.00046563
0.93103448	1	1	1	0.00704225	1.655172	0
0.24137931	0.9537037	0.8611111	0.4439016	0.07864878	1.344828	0.0265733
0.36206897	0.9833333	0.9166667	0.6495191	0.05158022	1.448276	0.00715936
0.15517241	0.9351852	0.8333333	0.3965126	0.09942745	1.37931	0.01112683
0.94827586	1	1	1	0.00277778	1.431034	0
0.29310345	0.9907407	0.8888889	0.7525997	0.05875741	1.206897	0.0179449
0.96551724	1	1	1	0.00277778	1.431034	0
0.98275862	1	1	1	0.00277778	1.431034	0
0.12068966	0.9074074	0.8055556	0.3525058	0.07905979	1.068966	0.05805367
1	1	1	1	0.00704225	1.655172	0
0.65517241	1	1	1	0.03611111	1.431034	0
0.22413793	0.9259259	0.8611111	0.4093714	0.09994908	1.155172	0.04361847

Plant Species (Table S7b)

nestedrank	PDI	species.spec	resource.ran P	SI	node.specia	betweennes
0.63888889	0.9827586	0.7009845	0.9827586	1	1.916667	0
0.52777778	0.9827586	0.5765721	0.9482759	1	1.305556	0.00574702
0.77777778	1	1	1	1	1.527778	0
0.30555556	0.9904875	0.6621044	0.8793103	1	1.194444	0.01901597
0.36111111	0.9640805	0.4376721	0.8965517	1	1.166667	0.0256144
0.02777778	0.9283122	0.2977871	0.5862069	1	1.055556	0.13575396
0.41666667	0.9586207	0.419942	0.9137931	1	1.25	0.01357474
0.58333333	0.9655172	0.5673086	0.9655172	1	1.555556	0.00127227
0.16666667	0.9721485	0.4922893	0.8275862	1	1.111111	0.08483831
0.69444444	0.9827586	0.7009845	0.9827586	1	1.416667	0.0013945
0.83333333	1	1	1	1	1.5	0
0.86111111	1	1	1	1	1.611111	0
0.72222222	0.9827586	0.7009845	0.9827586	1	1.333333	0.00506291
0.05555556	0.9920014	0.6837892	0.6551724	1	1.166667	0.05791233
0.66666667	0.9913793	0.7401978	0.9827586	1	1.333333	0.00506291
0.11111111	0.9626437	0.3595643	0.7413793	1	1.111111	0.04580031
0.44444444	0.9854809	0.5850832	0.9310345	1	1.222222	0.01509711

0.02386543	1.25	1	0.8965517	0.3821251	0.9396552	0.38888889
0.00493774	1.361111	1	0.9310345	0.4530785	0.9482759	0.47222222
0.00954199	1.638889	1	0.9482759	0.7220098	0.9928161	0.5
0.12022779	1.055556	1	0.4310345	0.3307503	0.9539057	0
0	1.666667	1	0.9655172	0.6151036	0.9827586	0.55555556
0	2.055556	1	1	1	1	0.88888889
0	1.805556	1	0.9827586	0.7009845	0.9827586	0.75
0	1.444444	1	1	1	1	0.91666667
0.00175068	1.388889	1	0.9655172	0.5673086	0.9655172	0.61111111
0	1.444444	1	1	1	1	0.9444444
0.05826167	1.333333	1	0.862069	0.4670689	0.9750958	0.27777778
0.07665272	1.083333	1	0.7068966	0.3384947	0.9275862	0.08333333
0	1.916667	1	1	1	1	0.97222222
0.05518267	1.166667	1	0.7586207	0.384132	0.9448276	0.13888889
0.04580031	1.111111	1	0.862069	0.5991096	0.9878296	0.25
0	1.527778	1	1	1	1	0.80555556
0.11253513	1.111111	1	0.862069	0.3948294	0.9463602	0.22222222
0	1.916667	1	1	1	1	1
0.02929686	1.388889	1	0.8965517	0.5005063	0.9764065	0.33333333
0.04580031	1.111111	1	0.8275862	0.467958	0.9758621	0.1944444

ults of Bipartite Interactions at Long-term Managed nt species (b) involved in mutualistic interactions during 2016-al metrics computed through the bipartite package in R.

weighted.be	closeness	weighted.clc	Fisher	alpha partner.dive.	effective.par	proportional
0	0.01363636	0.00145968	NA	0	1	0.07140841
0	0.01909091	0.01153891	NA	1.3896812	4.01357	0.28660266
0	0.01272727	0.00542009	NA	0.9502705	2.586409	0.18469137
0	0.02018182	0.01329821	NA	1.7782333	5.919389	0.42269417
0	0.01654545	0.00768647	NA	1.4648164	4.326749	0.30896623
0.63529978	0.02109091	0.05001214	NA	1.9193297	6.816388	0.4867474
0	0.01763636	0.00536732	NA	1.0397208	2.828427	0.20197347
0	0.02036364	0.03194663	NA	2.3579124	10.568865	0.75470579
0	0.01654545	0.00146204	NA	0	1	0.07140841
0.07646682	0.02018182	0.04009838	NA	2.2452634	9.442903	0.67430263
0.03109971	0.02072727	0.044125	NA	2.3954806	10.97347	0.78359803
0.03991664	0.01981818	0.04457001	NA	2.1859503	8.899101	0.63547064
0	0.01872727	0.02603965	NA	1.5350017	4.641333	0.33143021
0	0.01490909	0.00146329	NA	0	1	0.07140841
0	0.01818182	0.00881028	NA	0.4101163	1.506993	0.10761197
0	0.01418182	0.01192655	NA	0	1	0.07140841
0	0.01709091	0.00982685	NA	1.0549202	2.871746	0.2050668
0	0.01490909	0.00146329	NA	0	1	0.07140841
0	0.012	0.00146296	NA	0	1	0.07140841
0	0.01854545	0.00868197	NA	1.9061547	6.727171	0.48037658
0.21721706	0.02054545	0.04682113	NA	1.6106031	5.005829	0.35745828
0	0.01745455	0.01482314	NA	0.6277053	1.873307	0.13376986
0	0.01872727	0.01860966	NA	1.1358048	3.113678	0.22234282
0	0.02036364	0.03035549	NA	1.5388731	4.659337	0.3327158
0	0.01945455	0.02279526	NA	0.687092	1.987926	0.14195465
0	0.01836364	0.00415941	NA	0.6365142	1.889882	0.13495343
0	0.01654545	0.00547477	NA	0	1	0.07140841
0	0.01854545	0.018189	NA	1.3046615	3.686441	0.26324289
0	0.01309091	0.00145929	NA	0	1	0.07140841
0	0.01945455	0.00967319	NA	1.3862944	4	0.28563363
0	0.01654545	0.00146204	NA	0	1	0.07140841
0	0.01890909	0.00968522	NA	0.9743148	2.649351	0.18918594
0	0.01836364	0.01034381	NA	0.8486856	2.336574	0.16685099
0	0.01727273	0.00537611	NA	0.5623351	1.754765	0.125305
0	0.01418182	0.00145894	NA	0	1	0.07140841
0	0.01709091	0.00409734	NA	1.0986123	3	0.21422522

0	0.01636364	0.01323706 NA	1.5832585	4.870801	0.34781616
0	0.01654545	0.00146204 NA	0	1	0.07140841
0	0.01436364	0.00283417 NA	1.0397208	2.828427	0.20197347
0	0.01418182	0.00537244 NA	1.332179	3.789291	0.27058726
0	0.01327273	0.00285827 NA	0	1	0.07140841
0	0.01690909	0.00651827 NA	0.9502705	2.586409	0.18469137
0	0.01163636	0.00146794 NA	0	1	0.07140841
0	0.01309091	0.00543635 NA	0	1	0.07140841
0	0.01581818	0.0041447 NA	0.6365142	1.889882	0.13495343
0	0.01654545	0.00146204 NA	0	1	0.07140841
0	0.01509091	0.00283202 NA	0.6931472	2	0.14281681
0	0.01418182	0.00145894 NA	0	1	0.07140841
0	0.01745455	0.00867386 NA	1.6674619	5.298702	0.37837188
0	0.01636364	0.0097003 NA	1.0735428	2.925727	0.20892147
0	0.01709091	0.01045846 NA	1.834372	6.261201	0.44710236
0	0.01654545	0.00146204 NA	0	1	0.07140841
0	0.01890909	0.01563824 NA	0.9089087	2.481613	0.17720803
0	0.01654545	0.00146204 NA	0	1	0.07140841
0	0.01654545	0.00146204 NA	0	1	0.07140841
0	0.02036364	0.01401946 NA	1.9915094	7.326584	0.52317968
0	0.01418182	0.00145894 NA	0	1	0.07140841
0	0.01654545	0.01493263 NA	0	1	0.07140841
0	0.01945455	0.01062059 NA	1.7201935	5.585609	0.39885943
	000000000000000000000000000000000000000	0.01636364 0.01654545 0.01436364 0.01436364 0.01418182 0.01327273 0.01690909 0.01163636 0.01327273 0.01690909 0.01163636 0.01309091 0.01581818 0.01509091 0.01509091 0.01418182 0.01745455 0.01636364 0.01709091 0.01654545 0.01654545 0.01654545 0.01654545 0.01654545 0.01654545 0.01654545 0.01654545 0.01654545 0.01654545 0.01654545 0.01654545 0.01654545 0.01654545 0.01654545 0.01654545 0.01654545	0 0.01636364 0.01323706 NA 0 0.01654545 0.00146204 NA 0 0.01436364 0.00283417 NA 0 0.01418182 0.00537244 NA 0 0.01327273 0.00285827 NA 0 0.01690909 0.00651827 NA 0 0.01690909 0.00651827 NA 0 0.01163636 0.00146794 NA 0 0.01309091 0.00543635 NA 0 0.01581818 0.0041447 NA 0 0.01654545 0.00146204 NA 0 0.01654545 0.00145894 NA 0 0.01745455 0.00867386 NA 0 0.01636364 0.0097003 NA 0 0.01654545 0.00146204 NA 0 </td <td>0 0.01636364 0.01323706 NA 1.5832585 0 0.01654545 0.00146204 NA 0 0 0.01436364 0.00283417 NA 1.0397208 0 0.01418182 0.00537244 NA 1.332179 0 0.01327273 0.00285827 NA 0 0 0.01690909 0.00651827 NA 0 0 0.01163636 0.00146794 NA 0 0 0.01309091 0.00543635 NA 0 0 0.01581818 0.00146794 NA 0 0 0.01581818 0.00146204 NA 0 0 0.0159091 0.00283202 NA 0.6931472 0 0.01745455 0.00145894 NA 0 0 0.01745455 0.00867386 NA 1.6674619 0 0.01654545 0.00145204 NA 0 0 0.01654545 0.00146204 NA 0 0 0.01654545 0.00146204 NA 0 0</td> <td>0 0.01636364 0.01323706 NA 1.5832585 4.870801 0 0.01654545 0.00146204 NA 0 1 0 0.01436364 0.00283417 NA 1.0397208 2.828427 0 0.01438664 0.00283417 NA 1.332179 3.789291 0 0.01327273 0.00285827 NA 0 1 0 0.01690909 0.00651827 NA 0.9502705 2.586409 0 0.01690901 0.00543635 NA 0 1 0 0.01309091 0.00543635 NA 0 1 0 0.01581818 0.0041447 NA 0.6365142 1.889882 0 0.01654545 0.00146204 NA 0 1 0 0.0159091 0.00283202 NA 0.6931472 2 0 0.01745455 0.00867386 NA 1.6674619 5.298702 0 0.01654545 0.00146204 NA 1</td>	0 0.01636364 0.01323706 NA 1.5832585 0 0.01654545 0.00146204 NA 0 0 0.01436364 0.00283417 NA 1.0397208 0 0.01418182 0.00537244 NA 1.332179 0 0.01327273 0.00285827 NA 0 0 0.01690909 0.00651827 NA 0 0 0.01163636 0.00146794 NA 0 0 0.01309091 0.00543635 NA 0 0 0.01581818 0.00146794 NA 0 0 0.01581818 0.00146204 NA 0 0 0.0159091 0.00283202 NA 0.6931472 0 0.01745455 0.00145894 NA 0 0 0.01745455 0.00867386 NA 1.6674619 0 0.01654545 0.00145204 NA 0 0 0.01654545 0.00146204 NA 0 0 0.01654545 0.00146204 NA 0 0	0 0.01636364 0.01323706 NA 1.5832585 4.870801 0 0.01654545 0.00146204 NA 0 1 0 0.01436364 0.00283417 NA 1.0397208 2.828427 0 0.01438664 0.00283417 NA 1.332179 3.789291 0 0.01327273 0.00285827 NA 0 1 0 0.01690909 0.00651827 NA 0.9502705 2.586409 0 0.01690901 0.00543635 NA 0 1 0 0.01309091 0.00543635 NA 0 1 0 0.01581818 0.0041447 NA 0.6365142 1.889882 0 0.01654545 0.00146204 NA 0 1 0 0.0159091 0.00283202 NA 0.6931472 2 0 0.01745455 0.00867386 NA 1.6674619 5.298702 0 0.01654545 0.00146204 NA 1

weighted.be	closeness	weighted.clc	Fisher.alpha p	partner.dive	effective.par	proportional
0	0.01848975	0.00268879	1.59E+00	0.6931472	2	0.01071976
0	0.02843381	0.00832083	3.18E+00	1.2130076	3.363586	0.34226646
0	0.02470479	0.00507429	4.28E-01	0	1	0.1194487
0.02619761	0.03029832	0.02457853	2.83E+00	1.2298829	3.420829	0.34155181
0	0.03076445	0.02240834	2.56E+00	1.7276309	5.627307	0.64730351
0.06923653	0.03262896	0.02905655	1.08E+01	2.6522021	14.185242	0.62366784
0	0.02936607	0.01338367	3.31E+00	1.7274636	5.626365	0.3948293
0	0.02423866	0.00377519	5.37E+08	1.0986123	3	0.15390505
0	0.03169671	0.02762626	3.72E+00	1.7290258	5.635161	0.53069543
0	0.0265693	0.00268374	2.68E+08	0.6931472	2	0.14471669
0	0.02517091	0.00139367	1.34E+08	0	1	0.15926493
0	0.0233064	0.00139278	1.34E+08	0	1	0.08575804
0	0.02796768	0.00269157	2.68E+08	0.6931472	2	0.25191424
0.07859281	0.03076445	0.03283858	6.81E+00	1.4623873	4.316251	0.29170891
0	0.02796768	0.00390826	2.62E+00	0.6365142	1.889882	0.25191424
0.05239521	0.03169671	0.02031933	1.04E+01	2.3582013	10.571918	0.43668897
0	0.02983219	0.02224706	1.60E+00	1.2933039	3.644809	0.43406257

0	0.02936607	0.00739684	1.45E+01	1.8891592	6.613805	0.30398162
0	0.02750155	0.00888092	5.71E+00	1.5595812	4.756828	0.30493874
0	0.02284027	0.01443842	1.65E+00	0.9160681	2.499444	0.17075038
0.64446108	0.03262896	0.03820249	9.21E+00	2.6237096	13.786772	0.67264761
0	0.02237415	0.00690643	2.39E+00	1.0114043	2.749459	0.05436447
0	0.01709136	0.001365	1.34E+08	0	1	0.00382848
0	0.02035426	0.00270838	2.68E+08	0.6931472	2	0.01454824
0	0.02610317	0.0013958	1.34E+08	0	1	0.13246554
0	0.02703543	0.0035778	5.37E+08	1.0986123	3	0.14624809
0	0.02610317	0.0013958	1.34E+08	0	1	0.13246554
0	0.02796768	0.01606197	5.69E+00	1.7903136	5.991331	0.30579145
0.02582335	0.03216283	0.03207384	5.67E+00	2.2736283	9.714584	0.75491813
0	0.01848975	0.00139225	1.34E+08	0	1	0.01684533
0.0007485	0.03076445	0.02612836	6.23E+00	2.1322715	8.434003	0.40540606
0	0.03169671	0.0194851	4.47E+00	1.4971906	4.469116	0.41212969
0	0.02470479	0.00269859	7.96E-01	0	1	0.1194487
0.05164671	0.03169671	0.02182932	3.79E+00	1.9191936	6.81546	0.48383759
0	0.01848975	0.00139402	1.34E+08	0	1	0.01531394
0.0508982	0.02703543	0.02406181	2.32E+00	1.5578645	4.74867	0.24732006
0	0.03169671	0.02641976	3.95E+00	1.8302172	6.235241	0.47654416

proportional	d
0.09954058	0.17304648
0.43164416	0.24311166
0.04594181	0.57135374
0.5467075	0.13076629
0.26339969	0.38584513
0.69291142	0.15662455
0.40964778	0.06082074
0.69129218	0.12352572
0.27565084	0
0.46957723	0.3326196
0.69081341	0.19249955
0.64102564	0.22313489
0.52788037	0.1949446
0.07503829	0.22105225
0.25158609	0.24991417
0.10872894	0.41084176
0.18683002	0.368059
0.07503829	0.22105225
0.01684533	0.47485766
0.33020674	0.28259447
0.37414957	0.57432459
0.33690659	0.22261236
0.41271057	0.23450979
0.42062277	0.37520472
0.45941807	0.19475996
0.35068913	0.04701524
0.27565084	0.11309478
0.43591118	0.33416311
0.0482389	0.29611595
0.47320061	0.12426449
0.27565084	0
0.43261868	0.1853782
0.39739663	0.18011551
0.32388974	0.08814353
0.10872894	0.1580463
0.19678407	0.21232152

0.33537519	0.2036566
0.27565084	0
0.15467075	0.48393057
0.05513017	0.50489485
0.02909648	0.45684234
0.21822358	0.28753843
0.00689127	0.62670986
0.0482389	0.45663121
0.13782542	0.30967066
0.27565084	0
0.10949464	0.22600228
0.10872894	0.1580463
0.2136294	0.32823126
0.17534456	0.32365807
0.27565084	0.3390586
0.27565084	0
0.43663859	0.18144864
0.27565084	0
0.27565084	0
0.48804335	0.21426531
0.10872894	0.1580463
0.27565084	0.23060407
0.35925101	0.30068129

proportional d

0.10959387	0.77806593
0.18431418	0.18484695
0.05479693	0.2937771
0.18745093	0.38154253
0.30835914	0.08639643
0.77730774	0.20303011
0.30830754	0.26336211
0.1643908	0.38908791
0.30878956	0.19406729
0.10959387	0.24851976
0.05479693	0
0.05479693	0.11597842
0.10959387	0.03260414
0.23651734	0.59741682
0.10355971	0.11678788
0.57930869	0.26534318
0.19972434	0.21635168

0.36241624	0.36538916
0.26065961	0.28877197
0.13696184	0.38012353
0.75547281	0.30748793
0.15066194	0.63977386
0.05479693	0.69846812
0.10959387	0.57548815
0.05479693	0.034519
0.1643908	0.40589228
0.05479693	0.034519
0.32830655	0.309304
0.53232941	0.11584263
0.05479693	0.42088611
0.46215749	0.34628884
0.24489384	0.2162788
0.05479693	0.17385061
0.37346633	0.17765137
0.05479693	0.43874269
0.26021254	0.56305053
0.34167207	0.21504902

Improving the pollinator pantry: restoration and management of open the complexity of plant-pollinator networks

Authors: Richard E. Walton, Carl D. Sayer, Helen Bennion & Jan C. Axmac

	species.degree	normalised.c	species.stre	interaction.p
Alomya debellator	1	0.04	0.33333333	-0.6666667
Andrena dorsata	2	0.08	1.03333333	0.01666667
Andrena/Lasioglossum spp.	4	0.16	0.46923077	-0.1326923
Aphantopus hyperantus	4	0.16	0.26357809	-0.1841055
Apis mellifera	8	0.32	2.15510779	0.14438847
Bombus hortorum	5	0.2	1.76722151	0.1534443
Bombus hypnorum	1	0.04	0.01136364	-0.9886364
Bombus lapidarius	6	0.24	1.47371467	0.07895244
Bombus lucorum/terrestris	10	0.4	4.03169424	0.30316942
Bombus pascuorum	7	0.28	2.16077104	0.16582443
Bombus spp.	6	0.24	0.9001554	-0.0166408
Bombus vestalis	1	0.04	0.02564103	-0.974359
Buathra laborator	2	0.08	0.34741784	-0.3262911
Cheilosia illustrata	2	0.08	0.04722222	-0.4763889
Cheilosia spp.	1	0.04	0.08333333	-0.9166667
Chrysotoxum bicinctum	2	0.08	0.04469697	-0.4776515
Epistrophe diaphana	1	0.04	0.05555556	-0.9444444
Epistrophe elegans	1	0.04	0.01388889	-0.9861111
Epistrophe grossulariae	3	0.12	0.07831279	-0.3072291
Episyrphus balteatus	8	0.32	2.2677572	0.15846965
Eristalis abusivus	1	0.04	0.06666667	-0.9333333
Eristalis arbustorum	3	0.12	0.54259018	-0.1524699
Eristalis horticola	1	0.04	0.06666667	-0.9333333
Eristalis interruptus	1	0.04	0.1	-0.9
Eristalis pertinax	2	0.08	1.03389831	0.01694915
Eupeodes corollae	2	0.08	0.13247863	-0.4337607
Eupeodes luniger	1	0.04	0.5	-0.5
Helophilus hybridus	1	0.04	0.02564103	-0.974359
Helophilus pendulus	2	0.08	0.375	-0.3125
Ichneumon xanthorius	1	0.04	0.05555556	-0.9444444
Leucozona laternaria	1	0.04	0.01388889	-0.9861111
Maniola jurtina	6	0.24	1.38263403	0.06377234
Meliscaeva auricollis	1	0.04	0.01388889	-0.9861111
Nymphalis urticae	2	0.08	0.04259018	-0.4787049
Ochlodes sylvanus	3	0.12	0.1127844	-0.2957385
Pararge aegeria	1	0.04	0.05084746	-0.9491525
Pieris brassicae	3	0.12	0.56536656	-0.1448778

Pieris rapae	2	0.08	0.02525253	-0.4873737
Platycheirus scutatus	1	0.04	0.06666667	-0.9333333
Pyronia tithonus	1	0.04	0.5	-0.5
Sericomyia silentis	1	0.04	0.5	-0.5
Sphaerophoria scripta	2	0.08	0.13333333	-0.4333333
Syritta pipiens	1	0.04	0.03333333	-0.9666667
Syrphus vetripennis	2	0.08	0.21388889	-0.3930556
Thymelicus lineola	4	0.16	0.61408451	-0.0964789
Thymelicus sylvestris	2	0.08	0.06666667	-0.4666667
Vanessa cardui	1	0.04	0.02564103	-0.974359
Vespula rufa	1	0.04	0.02777778	-0.9722222
Vespula vulgaris	1	0.04	0.08474576	-0.9152542
Volucella bombylans	2	0.08	0.03914141	-0.4804293
Xylota segnis	1	0.04	0.02564103	-0.974359

	species.degree	normalised.c	species.stre	interaction.p
Anthriscus sylvestris	3	0.05882353	1.625	0.20833333
Centaurea nigra	2	0.03921569	0.28030303	-0.3598485
Cirsium arvense	16	0.31372549	7.16902526	0.38556408
Cirsium vulgare	9	0.17647059	2.65376818	0.18375202
Convolvulus arvensis	3	0.05882353	0.5106383	-0.1631206
Crataegus monogyna	2	0.03921569	2	0.5
Glechoma hederacea	1	0.01960784	0.01851852	-0.9814815
Hedera helix	7	0.1372549	4.58743169	0.51249024
Heracleum sphondylium	16	0.31372549	10.8484277	0.61552673
Hieracium agg.	7	0.1372549	2.85624907	0.26517844
Papaver rhoeas	1	0.01960784	0.01851852	-0.9814815
Ranunculus repens	7	0.1372549	0.66305186	-0.0481355
Rubus fruticosus agg.	14	0.2745098	4.59900092	0.25707149
Salix cinerea agg.	1	0.01960784	0.01851852	-0.9814815
Sambucus nigra	1	0.01960784	0.06666667	-0.9333333
Senecio jacobaea	15	0.29411765	7.00477763	0.40031851
Silene dioica	1	0.01960784	0.01639344	-0.9836066
Sisymbrium officinale	1	0.01960784	0.5	-0.5
<i>Taraxacum</i> agg.	1	0.01960784	0.33333333	-0.6666667
Trifolium dubium	2	0.03921569	0.07291667	-0.4635417
Trifolium repens	9	0.17647059	2.07934621	0.11992736
Tripleurospermum inodorum	2	0.03921569	1.33333333	0.16666667
Veronica persica	1	0.01960784	0.03030303	-0.969697
Vicia cracca	6	0.11764706	1.67003367	0.11167228
Vicia hirsuta	1	0.01960784	0.0444444	-0.9555556

farmland ponds enhances

Table S8. Species-Level Metric ResPollinator species (a) and plant specielisted alongside species-level metrics

cher

Po	llinato	r Spe	cies	(Tabl	e S8a
	innato		50103	1 1 4 1	e oba

nestedrank	PDI	resource.ran	species.spec	PSI	node.specia	betweennes
0.76	1	1	1	0.33333333	2.5	0
0.42	0.9583333	0.9583333	0.6922187	0.51666667	1.729167	0
0.2	0.875	0.875	0.4677072	0.11730769	1.458333	0.0269559
0.16	0.9702381	0.875	0.6263873	0.11739025	1.354167	0.03708567
0.04	0.986413	0.7083333	0.7516443	0.62014841	1.083333	0.24090708
0.14	0.9375	0.8333333	0.496904	0.22740504	1.645833	0.00600488
0.78	1	1	1	0.01136364	1.770833	0
0.08	0.9647436	0.7916667	0.5736422	0.28692154	1.416667	0.03726255
0	0.9438406	0.625	0.4634874	0.2627224	1.3125	0.05099994
0.06	0.9166667	0.75	0.4301536	0.17590559	1.479167	0.01949724
0.1	0.9836957	0.7916667	0.7210109	0.29680094	1.3125	0.06998877
0.8	1	1	1	0.02564103	1.708333	0
0.44	0.9583333	0.9583333	0.6922187	0.17370892	1.895833	0.01234336
0.46	0.9583333	0.9583333	0.6922187	0.02361111	1.458333	0.03987558
0.56	1	1	1	0.08333333	1.729167	0
0.48	0.9583333	0.9583333	0.6922187	0.02234848	1.708333	0.00467713
0.6	1	1	1	0.05555556	1.729167	0
0.82	1	1	1	0.01388889	1.729167	0
0.26	0.9166667	0.9166667	0.5527708	0.02610426	1.583333	0.03096872
0.02	0.9358108	0.7083333	0.5044505	0.42959759	1.208333	0.13176855
0.68	1	1	1	0.06666667	1.729167	0
0.28	0.9166667	0.9166667	0.5527708	0.18086339	1.604167	0.07621107
0.7	1	1	1	0.06666667	1.729167	0
0.64	1	1	1	0.1	1.729167	0
0.38	0.9791667	0.9583333	0.7328281	0.3559322	1.916667	0
0.32	0.9895833	0.9583333	0.8164966	0.05982906	1.666667	0.00499171
0.84	1	1	1	0.5	1	0
0.86	1	1	1	0.02564103	1.708333	0
0.34	0.9861111	0.9583333	0.7806247	0.11458333	1.708333	0.00187466
0.62	1	1	1	0.05555556	1.729167	0
0.88	1	1	1	0.01388889	1.729167	0
0.12	0.9166667	0.7916667	0.4513355	0.15596737	1.333333	0.04230953
0.9	1	1	1	0.01388889	1.729167	0
0.5	0.9583333	0.9583333	0.6922187	0.02129509	1.625	0.01641464
0.22	0.9375	0.9166667	0.5773503	0.03844709	1.458333	0.02417162
0.66	1	1	1	0.05084746	1.916667	0
0.24	0.9583333	0.9166667	0.590727	0.15416215	1.625	0.0055874

0.52	0.9583333	0.9583333	0.6922187	0.01262626	1.541667	0.0190284
0.72	1	1	1	0.06666667	1.916667	0
0.92	1	1	1	0.5	2.583333	0
0.94	1	1	1	0.5	1	0
0.36	0.9861111	0.9583333	0.7806247	0.08333333	1.625	0.0114607
0.96	1	1	1	0.03333333	1.729167	0
0.3	0.9930556	0.9583333	0.8630747	0.1734127	1.645833	0.01115041
0.18	0.9583333	0.875	0.5625	0.1600939	1.541667	0.05943606
0.54	0.9583333	0.9583333	0.6922187	0.03333333	1.8125	0
0.98	1	1	1	0.02564103	1.708333	0
0.74	1	1	1	0.02777778	1.729167	0
0.58	1	1	1	0.08474576	1.916667	0
0.4	0.9791667	0.9583333	0.7328281	0.0223064	1.541667	0.0190284
1	1	1	1	0.02564103	1.708333	0

Plant Species (Table S8b)

betweennes	node.specia		resource.ran PSI	species.spe	PDI	nestedrank
0	1.956522	1	0.96	0.5656854	0.96	0.41666667
0	1.826087	1	0.98	0.7	0.98	0.54166667
0.17786376	1.173913	1	0.7	0.303302	0.9085714	0.04166667
0.09404101	1.304348	1	0.84	0.4989097	0.9653846	0.16666667
0	1.695652	1	0.96	0.5656854	0.96	0.45833333
0	NaN	1	0.98	0.7	0.98	0.58333333
0	1.652174	1	1	1	1	0.70833333
0.15121693	1.565217	1	0.88	0.7823315	0.9943478	0.25
0.01127646	1.521739	1	0.7	0.5215273	0.9810811	0
0.01127646	1.521739	1	0.88	0.5736433	0.9825	0.29166667
0	1.652174	1	1	1	1	0.75
0.05193783	1.304348	1	0.88	0.5015951	0.9766667	0.33333333
0.12869709	1.173913	1	0.74	0.4083769	0.9593103	0.125
0	1.652174	1	1	1	1	0.79166667
0	1.826087	1	1	1	1	0.83333333
0.21712302	1.217391	1	0.72	0.2700617	0.9	0.08333333
0	1.695652	1	1	1	1	0.875
0	2.173913	1	1	1	1	0.91666667
0	2.521739	1	1	1	1	0.95833333
0	1.695652	1	0.98	0.7393691	0.99	0.5
0.07700397	1.304348	1	0.84	0.4735328	0.9678261	0.20833333
0	2.043478	1	0.98	0.7	0.98	0.625
0	1.826087	1	1	1	1	1
0.07956349	1.391304	1	0.9	0.4406813	0.94	0.375
0	1.956522	1	1	1	1	0.66666667

ults of Bipartite Interactions at Overgrown Ponds.

⇒s (b) involved in mutualistic interactions during 2016-2017 are computed through the bipartite package in R.

weighted.be	closeness	weighted.clc	Fisher	alpha partner.dive.	effective.par	proportional
0	0.01292432	0.00300588	NA	0	1	0.09168909
0	0.01920482	0.00344555	NA	0.6931472	2	0.18337818
0	0.02304578	0.00862733	NA	1.3862944	4	0.36675636
0.01192843	0.02460293	0.0211389	NA	1.0751393	2.930401	0.26868582
0.18290258	0.02865151	0.02159387	NA	0.9792846	2.662551	0.24412686
0.00662691	0.02045053	0.03018337	NA	1.4085605	4.090064	0.37501422
0	0.01878958	0.00339731	NA	0	1	0.09168909
0.06527502	0.02366864	0.03119746	NA	1.3130961	3.717666	0.34086945
0.16269052	0.02522579	0.03075121	NA	1.7437945	5.719003	0.52437016
0.01689861	0.02273435	0.0273561	NA	1.6728027	5.327077	0.48843486
0.02584493	0.02543341	0.0285932	NA	0.9954909	2.706053	0.24811549
0	0.01951625	0.00344	NA	0	1	0.09168909
0.00049702	0.01733624	0.00498146	NA	0.6931472	2	0.18337818
0	0.0232534	0.00597434	NA	0.6931472	2	0.18337818
0	0.01941244	0.01401996	NA	0	1	0.09168909
0	0.01972387	0.00618121	NA	0.6931472	2	0.18337818
0	0.01941244	0.01066277	NA	0	1	0.09168909
0	0.01941244	0.00337951	NA	0	1	0.09168909
0	0.0211772	0.0060635	NA	1.0986123	3	0.27506727
0.38121272	0.02678293	0.03468217	NA	1.4673362	4.337665	0.39771656
0	0.01920482	0.0064315	NA	0	1	0.09168909
0.03114645	0.02107339	0.00562151	NA	1.0986123	3	0.27506727
0	0.01920482	0.0064315	NA	0	1	0.09168909
0	0.01920482	0.00904405	NA	0	1	0.09168909
0	0.01660957	0.00554873	NA	0.6365142	1.889882	0.17328152
0	0.02034672	0.01220292	NA	0.5004024	1.649385	0.1512306
0	0.00062286	NA	NA	0	1	0.09168909
0	0.01951625	0.00344	NA	0	1	0.09168909
0	0.01972387	0.01041041	NA	0.5623351	1.754765	0.16089284
0	0.01941244	0.01066277	NA	0	1	0.09168909
0	0.01941244	0.00337951	NA	0	1	0.09168909
0.06974818	0.02491436	0.02185271	NA	1.5821699	4.865502	0.44611347
0	0.01941244	0.00337951	NA	0	1	0.09168909
0	0.02076196	0.00553417	NA	0.6931472	2	0.18337818
0	0.02304578	0.01159859	NA	1.0549202	2.871746	0.26330777
0	0.01660957	0.00739063	NA	0	1	0.09168909
0	0.02055434	0.00965801	NA	1.0397208	2.828427	0.25933591

0	0.0222153	0.00614422	NA	0.6931472	2	0.18337818
0	0.01660957	0.00621223	NA	0	1	0.09168909
0	0.01230146	0.00226366	NA	0	1	0.09168909
0	0.00062286	NA	NA	0	1	0.09168909
0	0.02076196	0.00954958	NA	0.5623351	1.754765	0.16089284
0	0.01920482	0.00344555	NA	0	1	0.09168909
0	0.02065815	0.01508473	NA	0.4101163	1.506993	0.13817482
0.04522863	0.02200768	0.01431574	NA	1.2130076	3.363586	0.30840411
0	0.0179591	0.00601737	NA	0.6931472	2	0.18337818
0	0.01951625	0.00344	NA	0	1	0.09168909
0	0.01941244	0.00620515	NA	0	1	0.09168909
0	0.01660957	0.01006292	NA	0	1	0.09168909
0	0.0222153	0.00848609	NA	0.6365142	1.889882	0.17328152
0	0.01951625	0.00344	NA	0	1	0.09168909
weighted.be	closeness	weighted.clc	Fisher.alpha	partner.dive	effective.par	proportional
0	0.03292894	0.00577848	5.37E+08	1.0986123	3	0.02169625
0	0.03682842	0.00572796	2.68E+08	0.6931472	2	0.0729783
0.03165584	0.05459272	0.01839003	1.01E+01	2.4626818	11.736244	0.38461539
0.16883117	0.05155979	0.02090685	2.73E+00	1.5886465	4.897116	0.41897936
0	0.03986135	0.00563557	5.37E+08	1.0986123	3	0.20118343
0	0	NA	2.68E+08	0.6931472	2	0.00394477
0	0.04116118	0.00335019	1.34E+08	0	1	0.10650888
0.13717532	0.04289428	0.0226076	2.07E+00	0.8767438	2.403062	0.15779093
0.03896104	0.04419411	0.01990311	6.38E+00	1.9179013	6.806658	0.32380013
0	0.04419411	0.01755391	2.87E+00	1.4080605	4.088019	0.34299803
0	0.04116118	0.00335019	1.34E+08	0	1	0.10650888
0	0.05069324	0.01439714	6.18E+00	1.6313454	5.110746	0.49309665
0.41396104	0.05459272	0.02132426	4.69E+00	2.0362773	7.662033	0.61874664
0	0.04116118	0.00335019	1.34E+08	0	1	0.10650888
0	0.03596187	0.00334455	1.34E+08	0	1	0.0295858
0.10551948	0.05329289	0.01591988	1.19E+01	2.5435568	12.72485	0.32485207
0	0.0389948	0.00338053	1.34E+08	0	1	0.12031558
0	0.02902946	0.00324404	1.34E+08	0	1	0.00394477
0	0.02556326	0.00338691	1.34E+08	0	1	0.00591716
0	0.03986135	0.00761295	2.62E+00	0.6365142	1.889882	0.15779093
0.06818182	0.05155979	0.02021792	2.94E+00	1.658971	5.253902	0.44033531
0	0.03119584	0.0034549	2.68E+08	0.6931472	2	0.00788955
0	0.03682842	0.0033434	1.34E+08	0	1	0.06508876
0.03571429	0.04809359	0.01655657	2.91E+00	1.6364956	5.137135	0.28599606
0	0.03379549	0.0057947	7.96E-01	0	1	0.0887574

proportional	d
0.00591716	0.75462818
0.06114398	0.61146947
0.16765286	0.37301715
0.39299803	0.3183957
0.34070553	0.59552336
0.47534517	0.3218333
0.17357002	0
0.42122781	0.39091343
0.54810432	0.28349088
0.4965035	0.26755039
0.36865138	0.43852997
0.07692308	0.18175429
0.14595661	0.38756107
0.20118343	0.12293113
0.14201183	0.40007388
0.23274162	0.10000782
0.14201183	0.32432537
0.14201183	0.04481921
0.32938856	0.11263056
0.45950313	0.42044749
0.0591716	0.38130014
0.19723866	0.35204334
0.0591716	0.38130014
0.0591716	0.46931523
0.1183432	0.4799919
0.16765286	0.33329905
0.00394477	0.84518762
0.07692308	0.18175429
0.14792899	0.37860109
0.14201183	0.32432537
0.14201183	0.04481921
0.40394477	0.30690604
0.14201183	0.04481921
0.19329389	0.11570751
0.27613412	0.22217046
0.11637081	0.3134368
0.2209073	0.36317933

0.31558185	0
0.0591716	0.38130014
0.00394477	0.84518762
0.00394477	0.84518762
0.1183432	0.39752579
0.0591716	0.24035258
0.20118343	0.50684409
0.22953649	0.45615528
0.15779093	0.19007603
0.07692308	0.18175429
0.14201183	0.1812845
0.11637081	0.4175869
0.31558185	0.10542687
0.07692308	0.18175429

proportional d

0.16279753	0.78070432
0.10853168	0.43578614
0.63687715	0.38185164
0.26574613	0.4388886
0.16279753	0.42135607
0.10853168	1
0.05426584	0.12200633
0.13040419	0.85950906
0.36936905	0.60245828
0.22183978	0.43959949
0.05426584	0.12200633
0.27733894	0.16266073
0.41578664	0.20702149
0.05426584	0.12200633
0.05426584	0.40394574
0.6905247	0.52996042
0.05426584	0.09517783
0.05426584	0.84743513
0.05426584	0.7581904
0.10255601	0.22314918
0.28510741	0.40410838
0.10853168	0.87305398
0.05426584	0.23040266
0.27877097	0.40272947
0.05426584	0.28045883
Improving the pollinator pantry: restoration and managemer complexity of plant-pollinator networks

Authors: Richard E. Walton, Carl D. Sayer, Helen Bennion & Jar

a)	
	Patefield
Nestedness	0.395
Links per species	<0.001
Connectance	<0.001
Linkage Density	<0.001
Fisher's Alpha	<0.001
Shannon's Diversity	<0.001
H2'	<0.001

nt of open farmland ponds enhances the

ι C. Axmacher

Long-term Managed		
Vazquez	Swap	Patefield
0.678	0.859	0.351
NaN	NaN	<0.001
NaN	NaN	<0.001
<0.001	<0.001	<0.001
NaN	NaN	<0.001
<0.001	<0.001	<0.001
<0.001	<0.001	<0.001

Table S9. Results from three null model tests ("Patefield", "Vasquez", network-level metric indicating if a significant difference exists between the metric from a randomly-generated network. NaN indicates "not a number" a were not calculable. Significant differences from random network structure

Recently Restored		
Vazquez	Swap	Patefield
0.677	0.937	0.968
NaN	NaN	<0.001
NaN	NaN	<0.001
<0.001	<0.001	<0.001
NaN	NaN	<0.001
<0.001	<0.001	<0.001
<0.001	<0.001	<0.001

and "Swap"). P-values are given for each given networks metric and that of the same and that some or all iterations of 500 model runs are bolded.

Overgrown	
Vazquez	Swap
0.314	0.023
NaN	NaN
NaN	NaN
<0.001	<0.001
NaN	NaN
<0.001	<0.001
<0.001	<0.001

Declaration of interests

 \boxtimes The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

□The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

ponds enhances the complexity of plant-pollinator networks
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Improving the pollinator pantry: restoration and management of open farmland

17 Abstract

In line with general biodiversity losses across agricultural landscapes, insect 18 19 pollinators have experienced recent sharp declines. A range of conservation 20 measures have been developed to address these declines, with plant-pollinator 21 interaction networks providing key insights into the effectiveness of these measures. 22 For the first time, we studied interactions between three diurnal pollinator groups 23 (bees, hoverflies, and butterflies) and insect-pollinated plants to understand how they 24 are affected by pond management and restoration. Major network contributors were 25 identified, and important network-level parameters compared at nine farmland ponds 26 under different management strategies to assess management effects on plant-27 pollinator interactions: three 'overgrown' tree-covered ponds, three 'long-term 28 managed ponds' kept in an open-canopy, early- to mid-successional state by periodic 29 interventions involving tree and sediment removal, and three 'recently restored ponds', 30 initially heavily overgrown with woody vegetation, and subsequently rapidly 31 transformed into an early succession state through major tree and sediment removal. 32 Interaction complexity, as measured by the metrics 'links per species', 'linkage 33 density', Fisher's alpha and Shannon's Diversity, was higher for both long-term 34 managed and recently restored ponds compared to overgrown ponds. Several 35 network-level parameters indicated that highest complexity levels were found at 36 recently restored ponds due to their substantially higher plant diversity. Bipartite 37 interaction analysis suggests major benefits of pond management and restoration for 38 agricultural pollinator assemblages. We strongly advocate the inclusion of ponds in 39 conservation strategies and policies aimed at pollinators - ponds should be part of the 40 pollinator pantry.

41 Keywords: Agricultural landscapes, agro-ecosystems, biodiversity conservation,
 42 ecosystem services, plant-pollinator relationships, pond management
 43

44 1. Introduction

45

46 Numerous plant species rely strongly on insect visitors to successfully transfer pollen 47 and thus produce viable seeds (Ollerton et al., 2011). However, dramatic declines in 48 pollinating insect populations within agricultural landscapes raise serious concerns 49 for the pollination services they provide (Potts et al., 2010; Ollerton et al., 2014; Potts 50 et al., 2016). Insect pollinators can form complex networks with the flowering plants 51 they visit in search of nectar and pollen (Memmott, 1999; Cole et al., 2017). 52 Complexity in this context can be understood as the number of plant and pollinator 53 species involved in network interactions and the number of specific plant-pollinator 54 links (MacArthur, 1955; Pimm, 1980; Landi et al., 2018). Although complexity may in 55 some instances decrease network stability (Allesina and Tang, 2012), recent studies 56 have indicated that complexity can also increase resilience to disturbances linked to 57 changing environmental conditions and/or local species extirpations (Vilà et al., 58 2009; Goldstein and Zych, 2016; Martínez-Núñez et al., 2019). Increased complexity 59 is believed to allow plant-pollinator networks to adjust in ways that preserve a high 60 number of links between network components resulting particularly from many 61 different pollinator species visiting the same plant species.

62

Although bottom-up effects driven by flower richness and abundance exert a great
influence on plant-pollinator network structure (Spiesman and Inouye, 2013; Cole et
al., 2017), top-down effects linked, for example, to reductions in pollinating insect
populations can trigger pollination gaps in plant species strongly reliant on flower
visitors for reproduction (Blüthgen and Klein, 2011; Mathiasson and Rehan, 2020).
Furthermore, losses of highly specialised pollinators may cause strong declines in
'partner' plant species in agricultural landscapes (Biesmeijer et al., 2006; Weiner et

70 al., 2014). This can result in negative feedback loops, where plants with 71 anemochorous and other non-zoochorous syndromes increasingly dominate in the 72 landscape, while strictly insect pollinator-dependent plants decline (Biesmeijer et al., 73 2006). Reductions in plant visitors have also led to decreases in seed set of some 74 insect-pollinated plant species (Pauw, 2007; Hermansen et al., 2017; Nabors et al., 75 2018), triggering declines in recruitment, population sizes and fitness. These findings 76 accentuate the need not only for conservation of diverse forage resources, but for 77 understanding the structure of plant-pollinator networks and their prominent players. 78 specifically within the agricultural landscape where these networks are under 79 considerable stress due to habitat modification and landscape homogenisation 80 (Vanbergen, 2014; Moreira et al., 2015; Martínez-Núñez et al., 2019).

81

82 Plant-pollinator networks also represent important models for ecologists and 83 conservationists to understand the interactions that drive a system towards 84 enhanced complexity or simplification. Observing and analysing actual plant-85 pollinator interactions, therefore, can provide useful insights into plants and plant communities that are attracting the greatest number of pollinator specimens and 86 87 species (Memmott, 1999; Dicks et al., 2002; Carvell et al., 2006; Walton et al., 2020). 88 In turn, such knowledge enables conservation efforts to become more focused 89 toward plant communities that best promote local pollinating insect abundance and 90 diversity, thereby enhancing the provision of pollination services.

91

In agricultural landscapes, where plant-pollinator networks are increasingly trending
toward high levels of generalism and homogenisation (Biesmeijer et al., 2006;
Martínez-Núñez et al., 2019), it is important to understand how management

95 strategies focused on remaining semi-natural habitat patches might help to promote 96 increased complexity in plant-pollinator networks. A key semi-natural farmland 97 habitat widely disregarded in this context is the pond. In many temperate agricultural 98 landscapes, ponds are highly abundant (Biggs et al., 2005; Declerck et al., 2006; 99 Ruggiero et al., 2008). Nonetheless, in many areas of Northern Europe, due to a 100 general cessation of management practices involving periodic woody vegetation and 101 sediment removal over recent decades, large numbers of ponds have become 102 heavily overgrown by encroaching trees and shrubs (Saver et al., 2012; Janssen et 103 al., 2018). As a result, early-successional open-canopy ponds that support diverse 104 communities of wetland plants have become relatively rare in the landscape (Sayer 105 et al., 2013; Sayer and Greaves, 2020). Both pond restoration, that removes large 106 parts of the encroaching woody vegetation and accumulated pond sediment, and 107 pond management, that maintains ponds in an open-canopy state, have been shown 108 to strongly promote communities of aquatic and semi-aquatic plants and 109 invertebrates (Sayer et al., 2012; Sayer et al., 2013), whilst also enhancing avian 110 abundance and diversity (Davies et al., 2016; Lewis-Phillips et al., 2019a, 2019b, 111 2020). Furthermore, pond management and restoration has been shown to 112 significantly improve flower resources for pollinating insects (Walton et al., 2021) by 113 creating an array of different micro-habitats both within and around ponds (Sayer et 114 al., 2013; Walton et al. 2021). Ponds may therefore afford both edge and centre-of-115 field food resources for foraging adult pollinators, and there is evidence of linked 116 benefits also for crop pollination (Stewart et al., 2017). However, while positive 117 pollinator population responses to pond management and restoration have been 118 observed (Walton et al. 2021), the overall impacts of these improvements on

interaction networks between terrestrial pollinators and pond margin vegetation havebeen little explored.

121

In this study we analyse the effects of pond restoration and management on plantpollinator networks. Focusing on links between the species-richness and abundance
of flowering herbaceous species found in and around pond margins we test the
hypothesis that pond restoration and regular pond management result in plantpollinator networks that are more complex than at overgrown tree-covered ponds.
We then discuss the implications of our results for both pond and pollinator
conservation.

129

130 2. Materials and methods

131

132 2.1 Study sites

133 We selected nine farmland ponds located within reasonable proximity (≤ 14 km 134 between ponds) to each other in North Norfolk, eastern England (Supplementary 135 Information Table S1). Three ponds were subjected to light-to-moderate removal of 136 woody vegetation about twice every decade over the last 50 years, a management 137 strategy known to result in the long-term persistence of open-canopy, macrophyte-138 dominant early successional ponds hereby referred to as 'long-term managed ponds' 139 (Fig. 1a - see Sayer et al., 2012). A further three ponds had been recently restored 140 (1 in 2011 and 2 in 2014) by major removal of woody vegetation and sediment 141 resulting in macrophyte-dominated, early successional conditions. These are hereby 142 referred to as 'recently restored ponds' (Fig. 1b). The remaining three ponds 143 represent the shrub- and tree-dominated "late successional" state resulting from at 144 least 30-40 years of abandonment and are hereby referred to as 'overgrown ponds'. 145 (Fig. 1c). The three different management treatments allowed us to determine the

impacts of pond restoration and long-term pond management on plant-pollinatorinteractions.

148

149 The study region is characterised by chalk bedrock overlain by glacial deposits of 150 sand, silt and gravel (Prince, 1964; Brenchley and Rawson, 2006). The agricultural 151 landscape in this area, predominantly arable cropland (with wheat, barley, sugar 152 beet, rapeseed, and broad beans as main crops) enclosed by hedgerows, some 153 pastureland, and scattered semi-natural habitats including managed woodland and 154 grassland (LUC, 2018), supports large numbers of farmland ponds dug as watering ponds for livestock or, especially during the 17th and 19th centuries, to extract marl 155 156 deposits used to improve the soil (Prince, 1964). The long-term managed ponds 157 were located at Manor Farm (mean distance between 3 ponds = 1.5 km) in the 158 village of Briston which supports a mosaic of about 40 ponds representing differing 159 stages of succession, with most of these ponds subject to regular rotational woody 160 vegetation and sediment management (Sayer et al., 2012; Sayer et al., 2013). The 161 three recently restored farmland ponds and the three unmanaged, overgrown ponds 162 were located in a similar farmland matrix in the vicinity of the villages of Bodham and 163 Baconsthorpe, about 14 km NE of Manor Farm. Recently restored and overgrown 164 ponds were separated by a mean distance of 1.5 km. While we acknowledge the 165 potential implications of separation between long-term managed and the other pond 166 treatments owing to the rarity of managed and restored ponds, and of the overall low 167 number of replicates per treatment, great efforts were made to select ponds with 168 highly similar geological, agricultural, and environmental settings to ensure 169 comparability of the results (see Supplementary Information Fig. S1). Additionally, no 170 managed honey bee hives were located within 500 m of the ponds.

171

172 All ponds in this study were small (< 475 m², range: 121-455 m²), shallow (\leq 1.3 m 173 depth) and surrounded by uncropped margins (mean: 8.7 m, range: 5-17.2 m) 174 consisting of rough grassland and woody vegetation. Typical of Norfolk farmland 175 ponds in general, the ponds were found in arable fields in close proximity to 176 hedgerows and small patches of deciduous woodland. The landscape matrix was 177 analysed using Google Earth Pro (earth.google.com) with a focus on land-use and 178 habitat elements within 500 m and 1000 m radii around each pond – distances that 179 reflect the foraging ranges of most nesting pollinator species observed in this study 180 (Gathmann and Tscharntke, 2002; Knight et al., 2005; Zurbuchen et al., 2010). 181 Landscape elements were measured with Google Earth Pro tools and included mean 182 hedgerow length as well as the mean area (m²) of grassland, woodland, freshwater 183 habitat features (primarily other ponds), arable fields, pasture, and human 184 settlements.

185

186 2.2 Survey methods

187 Insect-pollinated plants and pollinating invertebrates were surveyed approximately 188 one day per month at all ponds between March and October 2016-2017 - resulting 189 in 12 surveys per pond. Flowering plants were surveyed on the same day as each 190 pollinator survey by recording any insect-pollinated plant found in flower that either 191 emerged above the pond water column or grew in the uncropped margin between 192 the pond and surrounding cropland. The survey was terminated once no new 193 species could be identified. All insect-pollinated plant species in flower were 194 identified by morphological features to the lowest taxonomic level using Rose (2006). 195

196 Pollinating insects visiting flowers were surveyed using two different approaches, 197 time-lapse photography and visual observation. Surveying pollinators using time-198 lapse photography has been shown to be an effective method when used 199 concurrently with other survey methods (Edwards et al., 2015; Georgian et al., 200 2015), and was thus paired with direct flower-visiting observations made during the 201 entire surveying event. Two Timelapse Cam 8.0 camera systems (© EBSCO 202 Industries, Inc., Birmingham, AL) were set within randomly-chosen flower patches to 203 reduce biases of highly visited plants. Flower patches were photographed at 30-204 second intervals between the hours of 07:30 and 18:00, the main activity period for 205 diurnal pollinators (Campbell et al., 2014) to obtain a total of 12 surveys per pond. 206 The cameras were set between 30-50 cm from flower patches including multiple 207 plant species to ease insect identification and to have ~ 75 cm in viewing width, with 208 cameras photographing the chosen flower patches throughout the duration of the 209 survey day. All ensuing photographs were visually assessed for the presence of a 210 flower visitor, and those with a clearly identifiable visitor (identified to genus or 211 species level depending on photo resolution and angle) to the flower face were 212 included in plant-pollinator network construction and analyses.

213

A standardised visual observation survey was used to obtain a more complete idea of plant-pollinator interactions at each pond margin. The entire pond margin of each pond was intensively observed during walks that followed irregular transects for a standardised time of 30 minutes by a single observer during each monthly survey. Interactions were counted when a flower visitor directly interacted with the flower face. Multiple species on the same plant were only counted if the interaction time with the flower allowed for confident identification. Due to the small size of the habitat

patches formed by the pond margins, visual observation was undertaken by one
observer circumnavigating the entire outer edge of the pond and its associated
margin; matching the same area as the flowering plant survey.

224

225 Time-lapse and visual surveys were standardised by only conducting surveys when 226 ambient air temperature was \ge 8°C and wind speeds were < 25 km/h. Surveys were 227 undertaken regardless of the presence of any precipitation during each respective 228 day. These environmental conditions depart from those more commonly used (see 229 Pollard and Yates, 1993), but allowed us to effectively sample all ponds throughout 230 the entire study period (March-October), and to document any interactions occurring 231 in the cooler months of spring and autumn (Corbet et al., 1993; Stubbs and Falk, 232 2002; Falk, 2015). For both surveying techniques, all specimens observed or 233 recorded were identified to the lowest level possible using taxonomic keys (Stubbs 234 and Falk, 2002; Owens and Richmond, 2012; Archer, 2014; Thomas and Lewington, 235 2014; Falk, 2015; Yeo and Corbet, 2015). Solitary bee specimens from the genus 236 Andrena that could be safely identified as Andrena bicolor (Fabricius 1775), Andrena 237 dorsata (Kirby 1802) and Andrena nigroaenea (Kirby 1802) were catergorised at the 238 species level. However, due to difficulties of ascertaining correct identification in the 239 field, smaller specimens in the genera Andrena and Lasioglossum were combined 240 into one aggregate "taxon", each, in the analysis. We similarly combined specimens 241 of the two bumblebee species Bombus lucorum and B. terrestris.

242

243 2.3 Statistical analysis

Data were preliminarily checked for potential spatial autocorrelation by calculating
the Bray-Curtis similarity index (Magurran, 2004; Fortin and Dale, 2005) from the

246 number of interactions made between pollinators and flowers at all sites. The 247 distance between all sites was determined by calculating the haversine distance 248 (Sinnott, 1984; Fortin and Dale, 2005). The resulting matrices were subjected to a 249 Mantel test using 1000 permutations (Fortin and Dale, 2005), but no significant 250 spatial autocorrelation signal was detected (P = 0.83). Overall measurements of 251 landscape elements contained within the 500 m and 1000 m radii around each pond 252 were compared according to pond management using a pairwise t-test with 253 Bonferroni correction (Gotelli and Ellison, 2004).

254

255 Interaction networks for each pond were checked for sampling completeness to 256 establish potential differences in sampling effort. Sampling completeness was 257 calculated by dividing the pairwise interaction richness with the Chao estimator of 258 interaction richness (Chacoff et al., 2012; Macgregor et al., 2017; Martínez-Núñez et 259 al., 2019). Estimated completeness was $63 \pm 11\%$ (mean \pm SD), a value that is 260 similar to previous studies of plant-pollinator networks (Grass et al., 2018; Martínez-261 Núñez et al., 2019). To assess the effect of management on sampling completeness, 262 we used Bonferroni-corrected pairwise t-tests (Gotelli and Ellison, 2004). These 263 revealed that, despite a standardisation of sampling effort, long-term managed 264 ponds showed a trend towards stronger undersampling when compared to 265 overgrown ponds ($t_{2,4}$ = 17.32, P = 0.01). No major difference for sampling 266 completeness existed between recently restored and overgrown ponds.

267

We assessed the effect of management regime on the number of network-level interactions and on the number of pollinator and plant species involved in network interactions using MANOVA. The total data for each pond was checked for normality

(Huberty and Petoskey, 2000) prior to running the MANOVA. After calculating the
MANOVA, pairwise t-tests using a Bonferroni correction (see Gotelli and Ellison,
2004) were conducted to specify which management categories differed significantly
in their interaction indices (Wright, 1992; Cabin and Mitchell, 2000).

275

276 Bipartite networks of flowers visited by diurnal pollinator species were constructed by 277 combining the data for each pond in each of the three pond management categories. 278 Data were combined in this instance to help visualise trends between pond 279 categories more clearly, as this approach simplified the number of networks and 280 associated metrics produced. Several network-level metrics were computed for the 281 three mutualistic networks. These included nestedness, which describes a network 282 of interactions where specialist species (species only showing interactions with a 283 single 'partner') can only interact with a subset of species that also have interactions 284 with generalist species (Landi et al., 2018). Nestedness relates to network 285 complexity, as high nestedness in a network indicates a high cohesion amongst 286 species forming the core of interacting generalists, whereas a network with low 287 nestedness has a low cohesion amongst interacting generalist species (Bascompte 288 and Jordano, 2006; Burgos et al., 2007; Tylianakis et al., 2010). Connectance is a 289 further common indicator of complexity that measures the fraction of the realised 290 interactions in relation to all theoretically possible interactions (Dunne et al., 2002; 291 Morris et al., 2014; Landi et al., 2018). Interaction strength asymmetry (ISA) can 292 indicate if there is a loss in specialist-specialist and specialist-generalist interactions 293 (Bascompte et al., 2006; Aizen et al., 2012; Soares et al., 2017), further leading ISA 294 to act as a measure of stability in complex observed networks (Neutel and Thorne, 295 2018), since interacting pollinators become less dependent on a specific plant

296 species, and plant species become less dependent on a specific pollinator species, 297 as network complexity increases. The metric H₂', while not specifically related to 298 network complexity, is important to understanding whether an observed network is 299 dominated more strongly by specialist or generalist species, allowing for useful 300 comparisons between networks with regards to proportions of generalist versus 301 specialist interactions (Blüthgen et al., 2006; Morris et al., 2014; Soares et al., 2017). 302 Finally, the metrics 'links per species', 'linkage density', 'Fisher's alpha', and 303 'Shannon's Diversity' of the networks characterise different aspects of complexity in 304 interaction diversity, or richness of interactions, between plants and pollinators 305 (Dormann et al., 2008; Ebeling et al., 2011; Fründ et al., 2013). To identify those 306 flowering plant and pollinator species that provided major contributions to the 307 constructed networks, the species-level metric 'degree' was assessed for each 308 species to summarise the number of links with visits/visitors, with any species having 309 \geq 10 interactions included.

310

311 Null models were run to determine if trends observed in the networks could be 312 explained by random occurrences such as those produced by sampling effects 313 (Gotelli and Graves, 1996). For our analysis, three null models were tested against 314 the observed networks. The "Patefield" null model was used as it is based on 315 marginal totals of the observed network, thereby recreating an ecological equivalent 316 of common and rare species involved in interactions (Patefield, 1981; Dormann et 317 al., 2009). The "Vazquez" null model was used as it does not strictly constrain 318 marginal totals of connectance for the observed network by recreating network 319 interactions based on species' relative abundances (i.e. more abundant species are 320 likely to show more interactions with target species) (Vázquez et al., 2005, 2007).

Finally, the "Swap" null model was used as it constrains connectance to those values
found in the observed network and thereby takes into account ecological or
evolutionary processes that constrain interactions, such as morphological pollination
barriers (Stang et al., 2006; Dormann et al., 2009). Null models were run using 500
simulations, and results were compared with generated network-level metrics.

326

327 Differences in network-level metrics between the three observed interaction 328 networks were then tested for significance by comparing observed and null-modelled 329 model outputs and by calculating the share of null model values exceeding observed 330 values (see Dormann, 2020). This method avoids standardising values using z-331 scores, which can be disadvantageous when used with the calculated network 332 metrics, as the standard deviation can be an unsuitable measure of spread 333 (Dormann, 2020). Instead, this approach uses a method similar to bootstrapping, 334 comparing the observed metric values with values obtained from a null model run 335 several thousand times, and using the statistical outputs of observed and null-336 modelled networks by counting the proportion of null-modelled values that exceed 337 observed values to obtain a p-value (Efron and Tibshirani, 1993; Dormann, 2020). 338 Network-level metrics were used as response variables to calculate the mean 339 difference between each management category, and Patefield null models, used 340 here as this represents a conservative approach to compare against the observed 341 networks (Dormann, 2020), were run 5000 times to check for significant differences 342 between the categories for the metric of interest (Dormann, 2020). Flower-visitation 343 networks and other analyses were calculated in R (Version 3.5.1 GUI El Capitan 344 build, © 2016) using the bipartite (Dormann et al., 2008) and vegan packages 345 (Oksanen et al., 2018).

346

347 3. **Results**

348

349 Comparison of land use between management categories indicated that the ponds 350 were situated in very similar landscapes. There was a significantly higher area of 351 grassland within 500 m of recently restored ponds compared to the long-term 352 managed ponds ($t_{2,4}$ = 4.91, P = 0.01) and a significantly higher area of other 353 freshwater features within 1000 m of long-term managed and overgrown ponds (t_{2,4} 354 = 3.87, P = 0.03). Nonetheless, no other significant differences in landscape 355 components were observed between the management categories within the 500 and 356 1000 m radii (see Supplementary Information Table 2).

357

358 The flower-visiting network generated for recently restored ponds contained the 359 highest number of diurnal pollinator species (n = 68) visiting flowering plant species 360 (n = 36) (Fig. 2a, Table 1, Supplementary Information Fig. S2, Table S3). Long-term 361 managed ponds had the second highest number of pollinator species (n = 59) 362 interacting with a slightly higher number (n = 37) of flowering plant species (Fig. 2b, 363 Table 1, Supplementary Information Fig. S2, Table S4). Overgrown ponds harboured 364 the lowest number of diurnal pollinator species (n = 51) that furthermore visited the 365 lowest number of flowering plant species (n = 25) (Fig. 2c, Table 1, Supplementary 366 Information Fig. S2, Table S5).

367

368 Six insect species: the hoverfly *Episyrphus balteatus* (De Geer 1776), the

bumblebees *Bombus pascuorum* (Scopoli 1763), *B. hortorum* (Linnaeus 1761), *B.*

370 *lapidarius* (Linnaeus 1758), *B. lucorum/terrestris* (Linnaeus 1758), and the honey

371 bee Apis mellifera (Linnaeus 1758), constituted a sizeable proportion of pollinator-

372 flower interactions at both recently restored and long-term managed ponds (see

Supplementary Table S6 & S7 for number of interactions for each species). At
recently restored ponds, the large white butterfly, *Pieris brassicae* (Linnaeus 1758),
was also a major contributing pollinator in the network. Only *B. lucorum/terrestris*was found to be an important contributing pollinator species at overgrown ponds
(Supplementary Table S8).

378

379 The insect-pollinated plant species that were major contributors towards network 380 interactions in the overall recently restored plant-pollinator network were Mentha 381 aquatica L., Cirsium arvense (L.) Scop., Lycopus europaeus L. (Lamiaceae), 382 Senecio jacobaea L. (Asteraceae), C. vulgare (Savi) Ten., Rubus fruticosus agg. L. 383 (Rosaceae), Ranunculus repens L. (Ranunculaceae), Epilobium hirsutum L. 384 (Onagraceae) and Heracleum sphondylium L. (Apiaceae) (Supplementary Table S6). 385 The most important plant species for the overall network at long-term managed 386 ponds again included M. aquatica, C. arvense, H. sphondylium, R. fruticosus agg., S. 387 jacobaea, and E. hirsutum, while Hypericum perforatum L. (Hypericaceae) and Vicia 388 cracca L. (Fabaceae) were also important (Supplementary Table S7). Major 389 contributing flowering plant species in the overgrown ponds network included C. 390 arvense, H. sphondylium, S. jacobaea, and R. fruticosus agg. (Supplementary Table 391 S8).

392

Results from the MANOVA revealed a strong association between the number of plant species involved in network interactions and pond management (df = 2, F = 19.11, P = 0.002) (Table 2). Further analysis with Bonferroni-corrected pairwise ttests revealed that both recently restored and long-term managed ponds had significantly more plant species involved in network interactions than overgrown

398 ponds ($t_{2,4} = -8.69$, P = 0.009 & $t_{2,4} = -5.03$, P = 0.004, respectively). The number of 399 network interactions and pollinator species involved in network interactions was 400 greater at both recently restored and long-term managed ponds, when compared 401 with overgrown ponds, but MANOVA and pairwise analysis showed that these 402 differences were not significant.

403

The plant-pollinator network recorded at the recently restored ponds showed the highest linkage density and the highest interaction diversity (Table 2, Supplementary Information Fig. S2). Long-term managed pond communities showed the highest degree of nestedness but displayed intermediate values in network metrics when compared to restored and overgrown ponds. Overgrown ponds had the highest specialisation but were otherwise characterised by the lowest network metric values.

411 Null model results showed each constructed network to be largely non-random. 412 Nestedness formed an exception, as results were non-significant for the 'Patefield, 413 Vazquez, and Swap' null models for all pond categories (Supplementary Information 414 Table S9). All other network-level metrics were shown to be significantly different (P 415 < 0.001 in all cases) from a random network structure for each management 416 category (Supplementary Information, Table S9). Comparison of network-level 417 metrics between categories using the null model approach revealed that both long-418 term managed and recently restored ponds differed significantly from overgrown 419 ponds for several key metrics. Links per species, interaction diversity (Fisher's 420 alpha), and ISA were significantly higher at both long-term managed and recently 421 restored ponds, while specialisation was significantly higher at overgrown ponds 422 (Table 3), when compared with long-term managed and recently restored ponds.

Furthermore, the networks at recently restored ponds had a significantly higher
interaction diversity (Fishers's alpha, Shannon Diversity) than at long-term managed
ponds (Table 3).

426

427 4. Discussion

428 429 This study provides important and novel insights into the flower-visitation networks of 430 diurnal pollinators at farmland ponds with different management histories. Our 431 research reveals complex flower-visitation networks at these semi-natural habitats, 432 strongly suggesting that ponds afford important food-supplying localities for 433 pollinators in the agricultural landscape. In line with our initial hypothesis, this study 434 shows that pond restoration and subsequent management, which lead to increases 435 in the extent and diversity of pond and pond-marginal communities (Walton et al. 436 2021), has a highly significant positive impact on the forage networks of pollinators 437 inhabiting agricultural landscapes with more plant species involved in the managed 438 and restored networks. Pond restoration by tree and sediment removal and 439 subsequent long-term management to maintain open-canopy conditions, therefore, 440 has significant benefits that clearly extend into terrestrial plant-pollinator networks. 441

In general, overgrown ponds did not harbour plant species that were absent from
neighbouring farmland habitats, suggesting that, in contrast to recently restored and
managed ponds, they did not offer unique floral resources to pollinator communities.
This interpretation is further strengthened by our observation that restored and
managed pond networks had significantly more plant species involved in mutualistic
networks than overgrown ponds. A phenomenon common to all pond management
treatments was the presence of a few key high-quantity or high-quality nectar

449 producing flowering plant species that were heavily foraged, especially thistles 450 (Cirsium spp.) and bramble (R. fruticosus agg.). Cirsium spp. produce low quantities 451 of nectar with a high sugar content (Hicks et al., 2016), while for *R. fruticosus* agg. 452 nectar is provided in large quantities with a low sugar content (Fowler et al., 2016). 453 Higher abundances of these plant species at managed and restored ponds likely 454 served as an attracting force to pollinators known to congregate amongst the most 455 abundant floral resources in the foraging landscape (Gathmann and Tscharntke, 456 2002; Fowler et al., 2016; Lucas et al., 2017). The greater abundance of R. 457 *fruticosus* agg. and *Cirsium* spp. recorded at recently restored and long-term 458 managed ponds (see Walton et al., 2021) in turn is likely a function of the plants' 459 positive response to the openness and ground disturbance (Decocq et al., 2004; 460 Gardiner and Vaughan, 2008) associated with woody vegetation and sediment 461 removal at pond margins. However, the quality and especially the strength of these 462 links will require further verification in future studies. The greater abundances of 463 plant species found at all pond management treatments (Walton et al., 2021) 464 furthermore coincided with the regular occurrence of plant species found exclusively 465 at recently restored and long-term managed ponds, such as *M. aquatica* and *E.* 466 *hirsutum*. In combination, these trends seemed to result in greater food resources 467 being provided for farmland pollinators.

468

Although some plants, such as *R. fruticosus* agg., *Cirsium* spp., and *H. sphondylium*,
were major species within each network, the important role of water mint, *M. aquatica*, is particularly noteworthy since this species only occurred at two ponds,
one recently restored pond and one long-term managed pond. The strong presence
of *M. aquatica*, and to some extent also *L. europaeus*, in the observed plant-

474 pollinator networks strongly suggests that these plants are favoured by pollinators 475 that visit ponds. *Mentha aquatica* flowers late in the summer, making this plant an 476 important food source for pollinators that have few alternative flowering plant species 477 to choose from at ponds during this part of the year, providing high-quality pollen and 478 nectar (Ebeling et al., 2011; Kulloli et al., 2011) which is especially important for 479 pollinators preparing to overwinter (Alford, 1969; Seeley and Visscher, 1985). Due to 480 their semi-aquatic nature, *M. aquatica* and *L. europaeus* are closely associated with 481 open-canopy ponds, and both are otherwise uncommon in the wider agricultural 482 landscape. Results from the overgrown ponds suggests that pollinator species 483 visiting this habitat rely on common plant species that are abundant in the wider 484 agricultural landscape, such as *T. repens*, a species well known to provide pollen by 485 foraging hymenopterans (Hanley et al., 2008; Kleijn and Raemakers, 2008).

486

487 The enhanced plant-pollinator network-level metric measures constructed for long-488 term managed and recently restored ponds, especially interaction diversity, supports 489 previous work on network structure by Morris et al. (2014), who found increased 490 taxonomic richness to improve key interaction metrics, while not automatically 491 improving specialisation. The significantly higher specialisation found at overgrown 492 ponds, described further by a higher ISA, is likely a function of fewer and less 493 abundant plant species being available at these sites, resulting in some generalist 494 pollinators regularly visiting less competitive plants (Blüthgen et al., 2007; Ebeling et 495 al., 2011), as well as a greater abundance of early-spring flowering shrubs such as 496 P. spinosa or S. cinerea agg. that are exclusively accessed by early flying pollinator 497 species. Furthermore, the significantly lower ISA metrics at restored and managed 498 ponds when compared to overgrown ponds indicates that the two former networks

499 are more stable than the latter as there is an improved ratio between plant and 500 pollinator species involved in the interactions (Neutel and Thorne, 2018). Other 501 metrics that did not show significant differences between management categories, 502 but were generally improved at canopy-managed ponds, may still highlight potential 503 benefits of pond restoration and management for plant-pollinator networks. For 504 example, nestedness, although considered to be a measure of cohesion amongst 505 generalist species in a network (Soares et al., 2017), to a degree also measures 506 robustness, as more nested communities are more resistant to extinction events 507 (Burgos et al., 2007). Although we observed no significant differences between 508 treatments, general improvements to nestedness in our results at managed and 509 restored ponds indicate that a greater number of pollinator and plant contributors 510 may lead to more robust flower-visitation networks at these sites, but further 511 research is required to confirm this trend. Such management-linked improvements in 512 network structure are increasingly important since flowering plant and pollinator 513 communities are known to be faltering in many agricultural landscapes across the 514 globe (Biesmeijer et al., 2006; Carey et al., 2008; Potts et al., 2010; Raven and 515 Wagner, 2021), making open-canopy pond habitats potentially highly important 516 pollinator conservation sites within farmed landscapes.

517

518 The networks produced in this study will have missed out on recording especially 519 some of the rarer interactions due to time constraints of the surveys and also a 520 potential slight observer bias towards large, easily identifiable species. Nonetheless, 521 given the highly standardised recording approach used, these limitations will have 522 affected samples similarly across the different treatments. The resulting network 523 incompleteness was particularly apparent in long-term managed ponds in

524 comparison to overgrown ponds. Despite these limitations, our results clearly show 525 that interaction complexity increases with both pond management and restoration. 526 This trend toward greater complexity was most evident with the significantly higher 527 metrics links per species and interaction diversity (Fisher's alpha). These two 528 specific metrics have been shown to increase in complexity with an increase in 529 species diversity and abundance (Warren, 1990; Ebeling et al., 2011; Rzanny and 530 Voigt, 2012). The significant improvements we observed in these metrics are likely 531 related to an increased richness and abundance of the flowering plant species found 532 at both, long-term managed and recently restored ponds (Walton et al., 2021). These 533 increases in turn allow for a greater number of plant species being involved in the 534 observed interactions at these ponds. The lack of significant differences for other 535 complexity metrics such as connectance and linkage density could be resulting from 536 dampening effects linked to neighbouring habitats and the overall landscape 537 structure (Fründ et al., 2013; Martínez-Núñez et al., 2019; Marja et al., 2021). The 538 complexity of the wider landscape, for example, can affect associations between 539 network complexity and some of the network metrics (Martínez-Núñez et al., 2019), 540 with these associations being dampened with increasing landscape homogeneity. 541 Furthermore, connectance does not always improve significantly in line with 542 increasing flowering plant diversity (Fründ et al., 2013). Overall, the metrics used in 543 this study consistently indicate higher plant-pollinator network complexity is achieved 544 by both pond management and restoration, with the matrix structure of the 545 surrounding agricultural landscape apparently interacting with the degree of actual 546 management- and restoration-linked changes observed in individual metrics. 547

548 5. Conclusions

549

550 Our analysis highlights the ability of interaction network analysis to reveal the wider 551 ecological effects of habitat management beyond pure species richness and 552 abundance. We show that diurnal plant-pollinator networks at farmland ponds are 553 not only diverse and complex, but also exhibit strong variation according to pond 554 successional stage and management history. Ponds subject to restoration and 555 management by major woody vegetation and sediment removal, resulting in pond 556 margins rich in flowering plants, had plant-pollinator networks that exhibited greater 557 complexity than network interactions observed at highly overgrown, late-558 successional ponds. Variations in the plants and pollinators involved in interactions 559 across the studied pond management categories also suggest that a landscape 560 supporting a mosaic of ponds, at different stages of succession, may improve the 561 overall complexity and diversity of plant-pollinator interactions at the landscape scale 562 (see also Walton et al., 2021). Such improvements, which encourage greater stability 563 in plant-pollinator networks at farmland ponds and the surrounding agricultural 564 landscape, are only likely if the ratio of overgrown to managed or restored ponds is 565 narrowed (Boothby and Hull, 1997; Sayer et al., 2012). Our study helps to highlight 566 the multi-functional value of ponds, that extends beyond aquatic and semi-aquatic 567 species into the terrestrial sphere with clear, positive implications for pollinator 568 abundance and diversity. Due to their often-ubiguitous nature in agricultural 569 landscapes, we strongly recommend that farmland ponds receive greater attention in 570 agri-environmental and conservation strategies directed at pollinators.

571

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- 586

587 **References** 588

- Aizen, M.A., Sabatino, M., Tylianakis, J.M., 2012. Specialization and rarity predict nonrandom loss of interactions from mutualist networks. Science 335, 1486– 1489. https://doi.org/10.1126/science.1215320
- Alford, D.V., 1969. A study of the hibernation of bumblebees
 (Hymenoptera:Bombidae) in Southern England. J. Anim. Ecol. 38, 149–170.
 https://doi.org/10.2307/2743
- 595 Allesina, S., Tang, S., 2012. Stability criteria for complex ecosystems. Nature 483, 205–208. https://doi.org/10.1038/nature10832
- Archer, M.E., 2014. The Vespoid Wasps: (Tiphiidae, Mutillidae, Sapygidae, Scoliidae
 and Vespidae) of the British Isles, Handbooks for the identification of British
 insects. Royal Entomological Society, London, UK.
- Bascompte, J., Jordano, P., 2006. The structure of plant-animal mutualistic
 networks, in: Pascual, M., Dunne, J. (Eds.), Ecological Networks. Oxford
 University Press, Oxford, pp. 143–159.
- Bascompte, J., Jordano, P., Olesen, J.M., 2006. Asymmetric coevolutionary
 networks facilitate biodiversity maintenance. Science 312, 431–433.
 https://doi.org/10.1126/science.1123412
- Biesmeijer, J.C., Roberts, S.P.M., Reemer, M., Ohlemüller, R., Edwards, M.,
 Peeters, T., Schaffers, A.P., Potts, S.G., Kleukers, R., Thomas, C.D., Settele,
 J., Kunin, W.E., 2006. Parallel declines in pollinators and insect-pollinated
 plants in Britain and The Netherlands. Science 313, 351–354.
 https://doi.org/10.1126/science.1127863
- Biggs, J., Williams, P., Whitfield, M., Nicolet, P., Weatherby, A., 2005. 15 years of
 pond assessment in Britain: results and lessons learned from the work of
 Pond Conservation. Aquat. Conserv.-Mar. Freshw. Ecosyst. 15, 693–714.
 https://doi.org/10.1002/agc.745

- Blüthgen, N., Klein, A.-M., 2011. Functional complementarity and specialisation: The
 role of biodiversity in plant-pollinator interactions. Basic Appl. Ecol. 12, 282–
 291. https://doi.org/10.1016/j.baae.2010.11.001
- 618 Blüthgen, Nico, Menzel, F., Blüthgen, Nils, 2006. Measuring specialization in species 619 interaction networks. BMC ecology 6, 9. https://doi.org/10.1186/1472-6785-6-620 9
- Blüthgen, Nico, Menzel, F., Hovestadt, T., Fiala, B., Blüthgen, Nils, 2007.
 Specialization, constraints, and conflicting interests in mutualistic networks.
 Current Biology 17, 341–346. https://doi.org/10.1016/j.cub.2006.12.039
- 624 Boothby, J., Hull, A.P., 1997. A census of ponds in Cheshire, North West England. 625 Aquat. Conserv.-Mar. Freshw. Ecosyst. 7, 75–79.
- Brenchley, P.J., Rawson, P.F., 2006. The Geology of England and Wales.
 Geological Society of London, London.
- Burgos, E., Ceva, H., Perazzo, R.P.J., Devoto, M., Medan, D., Zimmermann, M.,
 María Delbue, A., 2007. Why nestedness in mutualistic networks? J. Theor.
 Biol. 249, 307–313. https://doi.org/10.1016/j.jtbi.2007.07.030
- Cabin, R.J., Mitchell, R.J., 2000. To Bonferroni or not to Bonferroni: when and how
 are the questions. Bulletin of the Ecological Society of America 81, 246–248.
- Campbell, J.W., Starring, A.M., Smith, G.L., 2014. Flower Visitors of Hymenocallis
 coronaria (Rocky Shoals Spider-lily) of Landsford Canal State Park South
 Carolina, USA. Natural Areas Journal 34, 332–337.
 https://doi.org/10.3375/043.034.0316
- 636 https://doi.org/10.3375/043.034.0316
- 637 Carey, P.D., Wallis, S., Chamberlain, P.M., Cooper, A., Emmett, B.A., Maskell, L.C.,
 638 McCann, T., Murphy, J., Norton, L.R., Reynolds, B., Scott, W.A., Simpson,
 639 I.C., Smart, S.M., Ullyett, J.M., 2008. Countryside Survey: UK Results from
 640 2007. NERC/Centre for Ecology & Hydrology, Wallingford, UK.
- 641 Carvell, C., Roy, D.B., Smart, S.M., Pywell, R.F., Preston, C.D., Goulson, D., 2006.
 642 Declines in forage availability for bumblebees at a national scale. Biol.
 643 Conserv. 132, 481–489. https://doi.org/10.1016/j.biocon.2006.05.008
- 644 Chacoff, N.P., Vázquez, D.P., Lomáscolo, S.B., Stevani, E.L., Dorado, J., Padrón,
 645 B., 2012. Evaluating sampling completeness in a desert plant–pollinator
 646 network. Journal of Animal Ecology 81, 190–200.
 647 https://doi.org/10.1111/j.1365-2656.2011.01883.x
- Cole, L.J., Brocklehurst, S., Robertson, D., Harrison, W., McCracken, D.I., 2017.
 Exploring the interactions between resource availability and the utilisation of semi-natural habitats by insect pollinators in an intensive agricultural landscape. Agric. Ecosyst. Environ. 246, 157–167.
- https://doi.org/10.1016/j.agee.2017.05.007
 Corbet, S.A., Fussell, M., Ake, R., Fraser, A., Gunson, C., Savage, A., Smith, K.,
 - 1993. Temperature and the pollinating activity of social bees. Ecol. Entomol. 18, 17–30. https://doi.org/10.1111/j.1365-2311.1993.tb01075.x
- Davies, S.R., Sayer, C.D., Greaves, H., Siriwardena, G.M., Axmacher, J.C., 2016. A
 new role for pond management in farmland bird conservation. Agric. Ecosyst.
 Environ. 233, 179–191. https://doi.org/10.1016/j.agee.2016.09.005
- Declerck, S., De Bie, T., Ercken, D., Hampel, H., Schrijvers, S., Van Wichelen, J.,
 Gillard, V., Mandiki, R., Losson, B., Bauwens, D., Keijers, S., Vyverman, W.,
 Goddeeris, B., De Meester, L., Brendonck, L., Martens, K., 2006. Ecological
 characteristics of small farmland ponds: Associations with land use practices
 at multiple spatial scales. Biol. Conserv. 131, 523–532.
- 664 https://doi.org/10.1016/j.biocon.2006.02.024

- Decocq, G., Aubert, M., Dupont, F., Alard, D., Saguez, R., Wattez-Franger, A.,
 Foucault, B.D., Delelis-Dusollier, A., Bardat, J., 2004. Plant diversity in a
 managed temperate deciduous forest: understorey response to two
 silvicultural systems. J. Appl. Ecol. 41, 1065–1079.
 https://doi.org/10.1111/j.0021-8901.2004.00960.x
- Dicks, L.V., Corbet, S.A., Pywell, R.F., 2002. Compartmentalization in plant-insect
 flower visitor webs. J. Anim. Ecol. 71, 32–43. https://doi.org/10.1046/j.00218790.2001.00572.x
- 673 Dormann, C.F., 2020. Using bipartite to describe and plot two-mode networks in R
 674 28. https://cran.r-
- 675 project.org/web/packages/bipartite/vignettes/Intro2bipartite.pdf (accessed 12 676 October 2020)
- Dormann, C.F., Fründ, J., Blüthgen, N., Gruber, B., 2009. Indices, graphs and null
 models: analyzing bipartite ecological networks. The Open Ecology Journal 2,
 7–24. http://dx.doi.org/10.2174/1874213000902010007
- Dormann, C.F., Gruber, B., Fründ, J., 2008. Introducing the bipartite Package:
 analysing ecological networks. R news, R news Vol 8/2, 8 11. 8/2, 8–11.
- Dunne, J.A., Williams, R.J., Martinez, N.D., 2002. Network structure and biodiversity
 loss in food webs: robustness increases with connectance. Ecol. Lett. 5, 558–
 567. https://doi.org/10.1046/j.1461-0248.2002.00354.x
- Ebeling, A., Klein, A.-M., Tscharntke, T., 2011. Plant-flower visitor interaction webs:
 Temporal stability and pollinator specialization increases along an
 experimental plant diversity gradient. Basic Appl. Ecol. 12, 300–309.
 https://doi.org/10.1016/j.baae.2011.04.005
- Edwards, J., Smith, G.P., McEntee, M.H.F., 2015. Long-term time-lapse video
 provides near complete records of floral visitation. J. Pollinat. Ecol. 16, 91–
 100. https://doi.org/10.26786/1920-7603%282015%2916
- 692 Efron, B., Tibshirani, R.J., 1993. An Introduction to the Bootstrap. Chapman & Hall, 693 London.
- Falk, S., 2015. Field Guide to Bees of Great Britain and Ireland. British Wildlife
 Publishing, Totnes, UK.
- Fortin, M.-J., Dale, M.R.T., 2005. Spatial Analysis: A Guide for Ecologists.
 Cambridge University Press, Cambridge.
- Fowler, R.E., Rotheray, E.L., Goulson, D., 2016. Floral abundance and resource
 quality influence pollinator choice. Insect Conserv. Divers. 9, 481–494.
 https://doi.org/10.1111/icad.12197
- Fründ, J., Dormann, C.F., Holzschuh, A., Tscharntke, T., 2013. Bee diversity effects
 on pollination depend on functional complementarity and niche shifts. Ecology
 94, 2042–2054. https://doi.org/10.1890/12-1620.1
- Gardiner, T., Vaughan, A., 2008. Responses of ground flora and insect assemblages
 to tree felling and soil scraping as an initial step to heathland restoration at
 Norton Heath Common, Essex, England. Conservation Evidence 5, 95–100.
- Gathmann, A., Tscharntke, T., 2002. Foraging ranges of solitary bees. J. Anim. Ecol.
 708 71, 757–764. https://doi.org/10.1046/j.1365-2656.2002.00641.x
- Georgian, E., Fang, Z., Emshwiller, E., Pidgeon, A., 2015. The pollination ecology of
 Rhododendron floccigerum Franchet (Ericaceae) in Weixi, Yunnan Province,
 China. J. Pollinat. Ecol. 16, 72–81. https://doi.org/10.26786/19207603%282015%2911
- Goldstein, J., Zych, M., 2016. What if we lose a hub? Experimental testing of
 pollination network resilience to removal of keystone floral resources.

- 715Arthropod-Plant Interactions 10, 263–271. https://doi.org/10.1007/s11829-716016-9431-2
- Gotelli, N.J., Ellison, A.M., 2004. A Primer of Ecological Statistics. Sinauer
 Associates Publishers, Sunderland, MA, USA.
- Gotelli, N.J., Graves, G.R., 1996. Null models in ecology. Smithsonian Institution
 Press, Washington, D. C.
- Grass, I., Jauker, B., Steffan-Dewenter, I., Tscharntke, T., Jauker, F., 2018. Past and potential future effects of habitat fragmentation on structure and stability of plant–pollinator and host–parasitoid networks. Nat. Ecol. Evol. 2, 1408–1417. https://doi.org/10.1038/s41559-018-0631-2
- Hanley, M.E., Franco, M., Pichon, S., Darvill, B., Goulson, D., 2008. Breeding
 system, pollinator choice and variation in pollen quality in British herbaceous
 plants. Funct. Ecol. 22, 592–598. https://doi.org/10.1111/j.13652435.2008.01415.x
- Hermansen, T.D., Minchinton, T.E., Ayre, D.J., 2017. Habitat fragmentation leads to
 reduced pollinator visitation, fruit production and recruitment in urban
 mangrove forests. Oecologia 185, 221–231. https://doi.org/10.1007/s00442017-3941-1
- Hicks, D.M., Ouvrard, P., Baldock, K.C.R., Baude, M., Goddard, M.A., Kunin, W.E.,
 Mitschunas, N., Memmott, J., Morse, H., Nikolitsi, M., Osgathorpe, L.M.,
 Potts, S.G., Robertson, K.M., Scott, A.V., Sinclair, F., Westbury, D.B., Stone,
 G.N., 2016. Food for pollinators: quantifying the nectar and pollen resources
 of urban flower meadows. PLoS One 11, e0158117.
 https://doi.org/10.1371/journal.pone.0158117
- Huberty, C.J., Petoskey, M.D., 2000. 7 Multivariate Analysis of Variance and
 Covariance, in: Tinsley, H.E.A., Brown, S.D. (Eds.), Handbook of Applied
 Multivariate Statistics and Mathematical Modeling. Academic Press, San
 Diego, pp. 183–208. https://doi.org/10.1016/B978-012691360-6/50008-2
- Janssen, A., Hunger, H., Konold, W., Pufal, G., Staab, M., 2018. Simple pond
 restoration measures increase dragonfly (Insecta: Odonata) diversity.
 Biodivers. Conserv. 27, 2311–2328. https://doi.org/10.1007/s10531-018-15395
- Kleijn, D., Raemakers, I., 2008. A retrospective analysis of pollen host plant use by
 stable and declining bumble bee species. Ecology 89, 1811–1823.
 https://doi.org/10.1890/07-1275.1
- Knight, M.E., Martin, A.P., Bishop, S., Osborne, J.L., Hale, R.J., Sanderson, A.,
 Goulson, D., 2005. An interspecific comparison of foraging range and nest
 density of four bumblebee (Bombus) species. Molecular Ecology 14, 1811–
 1820. https://doi.org/10.1111/j.1365-294X.2005.02540.x
- Kulloli, S.K., Chandore, A.N., Aitawade, M.M., 2011. Nectar dynamics and pollination
 studies in three species of Lamiaceae. Current Science 100, 509–516.
- Landi, P., Minoarivelo, H.O., Brännström, Å., Hui, C., Dieckmann, U., 2018.
 Complexity and stability of ecological networks: a review of the theory. Popul.
 Ecol. 60, 319–345. https://doi.org/10.1007/s10144-018-0628-3
- Lewis-Phillips, J., Brooks, S., Sayer, C.D., McCrea, R., Siriwardena, G., Axmacher,
 J.C., 2019a. Pond management enhances the local abundance and species
 richness of farmland bird communities. Agric. Ecosyst. Environ. 273, 130–
 140. https://doi.org/10.1016/j.agee.2018.12.015
- Lewis-Phillips, J., Brooks, S.J., Sayer, C.D., McCrea, R., Siriwardena, G., Robson,
 H., Harrison, A.L., Axmacher, J.C., 2019b. Seasonal benefits of farmland

765 pond management for birds. Bird Study 66, 342–352. https://doi.org/10.1080/00063657.2019.1688762 766 Lewis-Phillips, J., Brooks, S.J., Sayer, C.D., Patmore, I.R., Hilton, G.M., Harrison, A., 767 Robson, H., Axmacher, J.C., 2020. Ponds as insect chimneys: Restoring 768 769 overgrown farmland ponds benefits birds through elevated productivity of 770 emerging aquatic insects. Biol. Conserv. 241, 108253. 771 https://doi.org/10.1016/j.biocon.2019.108253 772 LUC, 2018. North Norfolk Landscape Character Assessment. North Norfolk District 773 Council, Cromer. 774 Lucas, A., Bull, J.C., de Vere, N., Neyland, P.J., Forman, D.W., 2017. Flower 775 resource and land management drives hoverfly communities and bee 776 abundance in seminatural and agricultural grasslands. Ecol. Evol. 7, 8073-777 8086. https://doi.org/10.1002/ece3.3303 778 MacArthur, R., 1955. Fluctuations of Animal Populations and a Measure of 779 Community Stability. Ecology 36, 533-536. https://doi.org/10.2307/1929601 780 Macgregor, C.J., Evans, D.M., Pocock, M.J.O., 2017. Estimating sampling 781 completeness of interactions in quantitative bipartite ecological networks: 782 incorporating variation in species' specialisation. bioRxiv 195917. 783 https://doi.org/10.1101/195917 784 Magurran, A.E., 2004. Measuring biological diversity. Blackwell Publishing, Ltd, 785 Oxford. 786 Marja, R., Klein, A.-M., Viik, E., Batáry, P., 2021. Environmentally-friendly and 787 organic management practices enable complementary diversification of plant-788 bumblebee food webs. Basic Appl. Ecol. 53, 164–174. 789 https://doi.org/10.1016/j.baae.2021.03.013 790 Martínez-Núñez, C., Manzaneda, A.J., Lendínez, S., Pérez, A.J., Ruiz-Valenzuela, 791 L., Rey, P.J., 2019. Interacting effects of landscape and management on 792 plant-solitary bee networks in olive orchards. Funct. Ecol. 33, 2316–2326. 793 https://doi.org/10.1111/1365-2435.13465 794 Mathiasson, M.E., Rehan, S.M., 2020. Wild bee declines linked to plant-pollinator 795 network changes and plant species introductions. Insect Conserv. Divers. 13, 796 595-605. https://doi.org/10.1111/icad.12429 797 Memmott, J., 1999. The structure of a plant-pollinator food web. Ecol. Lett. 2, 276-798 280. https://doi.org/10.1046/j.1461-0248.1999.00087.x 799 Moreira, E.F., Boscolo, D., Viana, B.F., 2015. Spatial Heterogeneity Regulates Plant-Pollinator Networks across Multiple Landscape Scales. PLOS ONE 10, 800 801 e0123628. https://doi.org/10.1371/journal.pone.0123628 Morris, R.J., Gripenberg, S., Lewis, O.T., Roslin, T., 2014. Antagonistic interaction 802 803 networks are structured independently of latitude and host guild. Ecol. Lett. 804 17, 340-349. https://doi.org/10.1111/ele.12235 805 Nabors, A.J., Cen, H.J., Hung, K.-L.J., Kohn, J.R., Holway, D.A., 2018. The effect of 806 removing numerically dominant, non-native honey bees on seed set of a 807 native plant. Oecologia 186, 281-289. https://doi.org/10.1007/s00442-017-808 4009-y 809 Neutel, A.-M., Thorne, M.A.S., 2018. Symmetry, asymmetry, and beyond: the crucial 810 role of interaction strength in the complexity-stability debate, in: Moore, J.C., 811 de Ruiter, P.C., McCann, K.S., Wolters, V. (Eds.), Adaptive Food Webs: 812 Stability and Transitions of Real and Model Ecosystems. Cambridge 813 University Press, Cambridge, pp. 31-44.

814 Oksanen, J., Blanchet, F.G., Friendly, M., Kindt, R., Legendre, P., McGlinn, D., 815 Minchin, P.R., O'Hara, R.B., Simpson, G.L., Solymos, P., Stevens, M.H.H., 816 Szoecs, E., Wagner, H., 2018. vegan: Community Ecology Package. 817 Ollerton, J., Erenler, H., Edwards, M., Crockett, R., 2014. Extinctions of aculeate 818 pollinators in Britain and the role of large-scale agricultural changes. Science 819 346, 1360-1362. https://doi.org/10.1126/science.1257259 820 Ollerton, J., Winfree, R., Tarrant, S., 2011. How many flowering plants are pollinated 821 by animals? Oikos 120, 321-326. https://doi.org/10.1111/j.1600-822 0706.2010.18644.x Owens, N., Richmond, D., 2012. Bumblebees of Norfolk. Norfolk & Norwich 823 824 Naturalists' Society, Norwich, UK. 825 Patefield, W.M., 1981. An efficient method of generating random R × C Tables with 826 given row and column totals. Journal of the Royal Statistical Society: Series C 827 (Applied Statistics) 30, 91-97. https://doi.org/10.2307/2346669 828 Pauw, A., 2007. Collapse of a pollination web in small conservation areas. Ecology 829 88, 1759-1769. https://doi.org/10.1890/06-1383.1 830 Pimm, S.L., 1980. Properties of food webs. Ecology 61, 219–225. 831 https://doi.org/10.2307/1935177 832 Pollard, E., Yates, T.J., 1993. Monitoring Butterflies for Ecology and Conservation: 833 the British Butterfly Monitoring Scheme., Conservation Biology Series. 834 Chapman & Hall, London, UK. 835 Potts, S.G., Biesmeijer, J.C., Kremen, C., Neumann, P., Schweiger, O., Kunin, W.E., 836 2010. Global pollinator declines: trends, impacts and drivers. Trends Ecol. 837 Evol. 25, 345-353. https://doi.org/10.1016/j.tree.2010.01.007 838 Potts, S.G., Imperatriz-Fonseca, V., Hien, T., Biesmeijer, J.C., Breeze, T.D., Dicks, 839 L.V., Garibaldi, L.A., Hill, R., Settele, J., Vanbergen, A.J., 2016. The 840 assessment report on pollinators, pollination and food production: summary 841 for policymakers. Secretariat of the Intergovernmental Science-Policy 842 Platform on Biodiversity and Ecosystem Services, Bonn, Germanv. 843 Prince, H.C., 1964. The origins of pits and depressions in Norfolk. Journal of the 844 Geographic Association 49, 15-32. 845 Raven, P.H., Wagner, D.L., 2021. Agricultural intensification and climate change are 846 rapidly decreasing insect biodiversity. PNAS 118. 847 https://doi.org/10.1073/pnas.2002548117 848 Rose, F., 2006. The Wild Flower Key: How to identify wild flowers, trees, and shrubs 849 in Britain and Ireland, 2nd ed. Penguin Group, London, UK. 850 Ruggiero, A., Céréghino, R., Figuerola, J., Marty, P., Angélibert, S., 2008. Farm 851 ponds make a contribution to the biodiversity of aquatic insects in a French 852 agricultural landscape. Comptes rendus - Biologies 331, 298-308. 853 https://doi.org/10.1016/j.crvi.2008.01.009 854 Rzanny, M., Voigt, W., 2012, Complexity of multitrophic interactions in a grassland 855 ecosystem depends on plant species diversity. J. Anim. Ecol. 81, 614-627. 856 https://doi.org/10.1111/j.1365-2656.2012.01951.x 857 Sayer, C.D., Andrews, K., Shilland, E., Edmonds, N., Edmonds- brown, R., Patmore, 858 I., Emson, D., Axmacher, J., 2012. The role of pond management for 859 biodiversity conservation in an agricultural landscape. Aquat. Conserv.-Mar. 860 Freshw. Ecosyst. 22, 626-638. https://doi.org/10.1002/aqc.2254 861 Sayer, C., Shilland, E., Greaves, H., Dawson, B., Patmore, I., Emson, D., Alderton, E., Robinson, P., Andrews, K., Axmacher, J., Wiik, E., 2013. Managing 862

863 **Britain's ponds** - conservation lessons from a Norfolk farm. British Wildlife 25, 864 21-28. 865 Sayer, C.D., Greaves, H.M., 2020. Making an impact on UK farmland pond 866 conservation. Aquat. Conserv.-Mar. Freshw. Ecosyst. 30, 1821–1828. 867 https://doi.org/10.1002/agc.3375 868 Seeley, T.D., Visscher, P.K., 1985. Survival of honeybees in cold climates - the 869 critical timing of colony growth and reproduction. Ecol. Entomol. 10, 81-88. 870 https://doi.org/10.1111/j.1365-2311.1985.tb00537.x 871 Sinnott, R.W., 1984. Virtues of the Haversine. Sky Telesc. 68, 158. 872 Soares, R.G.S., Ferreira, P.A., Lopes, L.E., 2017. Can plant-pollinator network 873 metrics indicate environmental guality? Ecol. Indic. 78, 361-370. 874 https://doi.org/10.1016/j.ecolind.2017.03.037 875 Spiesman, B.J., Inouye, B.D., 2013. Habitat loss alters the architecture of plant-876 pollinator interaction networks. Ecology 94, 2688-2696. 877 https://doi.org/10.1890/13-0977.1 878 Stang, M., Klinkhamer, P.G.L., van der Meijden, E., 2006. Size constraints and 879 flower abundance determine the number of interactions in a plant-flower 880 visitor web. Oikos 112, 111-121. https://doi.org/10.1111/j.0030-881 1299.2006.14199.x 882 Stewart, R.I.A., Andersson, G.K.S., Brönmark, C., Klatt, B.K., Hansson, L.-A., 883 Zülsdorff, V., Smith, H.G., 2017. Ecosystem services across the aquatic-884 terrestrial boundary: linking ponds to pollination. Basic Appl. Ecol. 18, 13-20. 885 https://doi.org/10.1016/j.baae.2016.09.006 886 Stubbs, A., Falk, S., 2002. British Hoverflies: An Illustrated Identification Guide, 2nd 887 ed. British Entomological & Natural History Society, Reading, UK. 888 Thomas, J., Lewington, R., 2014. The Butterflies of Britain and Ireland, 3rd ed. 889 British Wildlife Publishing, Totnes, UK. 890 Tylianakis, J.M., Laliberte, E., Nielsen, A., Bascompte, J., 2010. Conservation of 891 species interaction networks. Biol. Conserv. 143, 2270-2279. 892 https://doi.org/10.1016/j.biocon.2009.12.004 893 Vanbergen, A.J., 2014. Landscape alteration and habitat modification: impacts on 894 plant-pollinator systems. Current Opinion in Insect Science 5, 44-49. 895 https://doi.org/10.1016/j.cois.2014.09.004 896 Vázquez, D.P., Melián, C.J., Williams, N.M., Blüthgen, N., Krasnov, B.R., Poulin, R., 897 2007. Species abundance and asymmetric interaction strength in ecological 898 networks. Oikos 116, 1120-1127. https://doi.org/10.1111/j.0030-899 1299.2007.15828.x 900 Vázquez, D.P., Poulin, R., Krasnov, B.R., Shenbrot, G.I., 2005. Species abundance 901 and the distribution of specialization in host-parasite interaction networks. 902 Journal of Animal Ecology 74, 946–955. https://doi.org/10.1111/j.1365-903 2656.2005.00992.x 904 Vilà, M., Bartomeus, I., Dietzsch, A.C., Petanidou, T., Steffan-Dewenter, I., Stout, 905 J.C., Tscheulin, T., 2009. Invasive plant integration into native plant-pollinator 906 networks across Europe. Proc. R. Soc. B-Biol. Sci. 276, 3887-3893. 907 https://doi.org/10.1098/rspb.2009.1076 908 Walton, R.E., Sayer, C.D., Bennion, H., Axmacher, J.C., 2020. Nocturnal pollinators 909 strongly contribute to pollen transport of wild flowers in an agricultural 910 landscape. Biol. Lett. 16, 20190877. https://doi.org/10.1098/rsbl.2019.0877 911 Walton, R.E., Sayer, C.D., Bennion, H., Axmacher, J.C., 2021. Open-canopy ponds 912 benefit diurnal pollinator communities in an agricultural landscape:

- 913 implications for farmland pond management. Insect Conserv. Divers. 14, 307–
 914 324. https://doi.org/10.1111/icad.12452
- Warren, P.H., 1990. Variation in food-web structure: the determinants of
 connectance. The American Naturalist 136, 689–700.
 https://doi.org/10.1086/285123
- Weiner, C.N., Werner, M., Linsenmair, K.E., Bluethgen, N., 2014. Land-use impacts
 on plant-pollinator networks: interaction strength and specialization predict
 pollinator declines. Ecology 95, 466–474. https://doi.org/10.1890/13-0436.1
- Wright, S.P., 1992. Adjusted p-values for simultaneous inference. Biometrics 48, 1005–1013. https://doi.org/10.2307/2532694
- 923 Yeo, P.F., Corbet, S.A., 2015. Solitary Wasps, 2nd ed. Pelagic Publishing, Exeter, 924 UK.
- Zurbuchen, A., Landert, L., Klaiber, J., Müller, A., Hein, S., Dorn, S., 2010. Maximum
 foraging ranges in solitary bees: only few individuals have the capability to
 cover long foraging distances. Biol. Conserv. 143, 669–676.
- 928 https://doi.org/10.1016/j.biocon.2009.12.003
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- 931 Figures
- 932

933 **Figure 1.** Photographs of three ponds in North Norfolk, UK used to study plant-

934 pollinator interactions which depict three management categories: (**a**) long-term

- managed pond (WADD17) near the village of Briston, (**b**) recently restored pond
- 936 (SHOOT) near the village of Baconsthorpe, and (**c**) overgrown pond (BAWO2)
- 937 between the villages of Baconsthorpe and Bodham.
- 938

939 **Figure 2**. Flower-visitation networks for (**a**) long-term managed ponds, (**b**) recently

- restored ponds, and (**c**) overgrown ponds during 2016-2017. Flower species are on
- 941 the left-hand side of the networks and diurnal pollinator species are on the right-hand
- side. The width of corresponding boxes and connecting lines is directly proportional
- 943 to the recorded number of visits for each species. Species with a high species-level
- 944 'degree' of interaction are named.

945

Table 1. Summary of network metrics for diurnal pollinators for the three pond management categories.

- 949

	Long-term	Recently Restored	Overgrown
	Managed		
Number of	59	68	51
Pollinating Insect			
Species			
Number of Plant	37	36	25
Species Visited			
Nestedness	4.74	4.82	7.29
Connectance	0.12	0.11	0.10
Links per Species	2.70	2.56	1.70
Linkage Density	7.75	8.29	4.89
Interaction Strength	0.04	0.05	0.14
Asymmetry (ISA)			
Fisher's Alpha	96.91	108.05	55.83
Shannon's Diversity	4.67	4.80	4.09
Specialisation (H ₂ ')	0.32	0.33	0.46

Table 2. MANOVA results on the effect of management on Number of Network
Interactions, Number of Pollinator Species Involved in Interactions, and Number of
Plant Species Involved in Interactions. Key: (Df) degrees of freedom, (Sum Sq) sum
of squares, (Mean Sq) mean of squares (Sum Sq/Df), (F value) F-statistic, and
(Pr(>F)) p-value.

~)	Df	Sum Sq	Mean Sq	F value	Pr(>F)
Response: Number of					-
Network Interactions					
Management	2	123587	61793	3.01	0.12
Residuals	6	123249	20541		
Response: Number of					
Pollinator Species Involved					
in Interactions					
Management	2	366.22	183.11	2.65	0.15
Residuals	6	415.33	69.22		
Response: Number of Plant					
Species Involved in					
Interactions					
Management	2	148.67	74.33	19.11	0.002**
Residuals	6	23.33	3.89		

- **Table 3.** Comparison of network-level metrics between three pond management
 963 categories using p-values from Patefield null model analysis. Significant differences
 964 are tied to the first management category in the column header except for
 965 Specialisation which is in favour of Overgrown ponds. Significant differences where
 966 P < 0.05 are bolded.

	Long-term Managed versus Overgrown	Recently Restored versus Overgrown	Recently Restored versus Long-term Managed
Nestedness	0.55	0.61	0.52
Connectance	1	1	0.90
Links per Species	<0.001	<0.001	0.66
Linkage Densitv	1	0.83	0.11
Fisher's Alpha	0.005	<0.001	0.001
Shannon's Diversity	1	1	0.02
Interaction Strength Asymmetry (ISA)	<0.001	<0.001	0.89
Specialisation (H ₂ ')	<0.001	<0.001	0.59