

1 Research paper – *Global Ecology and Biogeography*

2 **Functional biogeography of Neotropical moist forests: trait-climate**  
3 **relationships and assembly patterns of tree communities**

4

5 **Running title:** Functional biogeography of Neotropical forests

6

7 Bruno X. Pinho<sup>1\*</sup>, Marcelo Tabarelli<sup>1</sup>, Cajo ter Braak<sup>2</sup>, S. Joseph Wright<sup>3</sup>, Víctor Arroyo-  
8 Rodríguez<sup>4,5</sup>, Maíra Benchimol<sup>6</sup>, Bettina M. J. Engelbrecht<sup>7,3</sup>, Simon Pierce<sup>8</sup>, Peter Hietz<sup>9</sup>,  
9 Bráulio A. Santos<sup>10</sup>, Carlos A. Peres<sup>10,11</sup>, Sandra C. Müller<sup>12</sup>, Ian J. Wright<sup>13</sup>, Frans Bongers<sup>14</sup>,  
10 Madelon Lohbeck<sup>14,15</sup>, Ülo Niinemets<sup>16</sup>, Martijn Slot<sup>3</sup>, Steven Jansen<sup>17</sup>, Davi Jamelli<sup>1</sup>, Renato  
11 A.F. de Lima<sup>18,19</sup>, Nathan Swenson<sup>20</sup>, Richard Condit<sup>21,22</sup>, Jos Barlow<sup>23</sup>, Ferry Slik<sup>24</sup>, Manuel A.  
12 Hernández-Ruedas<sup>4</sup>, Gabriel Mendes<sup>1</sup>, Miguel Martínez-Ramos<sup>4</sup>, Nigel Pitman<sup>25</sup>, Nathan Kraft<sup>26</sup>,  
13 Nancy Garwood<sup>27</sup>, Juan Ernesto Guevara Andino<sup>28</sup>, Deborah Faria<sup>6</sup>, Eduardo Chacón<sup>29</sup>, Eduardo  
14 Mariano-Neto<sup>30</sup>, Valdecir Júnior<sup>10</sup>, Jens Kattge<sup>31,32</sup>, and Felipe P.L. Melo<sup>1</sup>

15 <sup>1</sup>Departamento de Botânica, Universidade Federal de Pernambuco, Recife, Brazil

16 <sup>2</sup>Biometris, Wageningen University & Research, Wageningen, The Netherlands

17 <sup>3</sup>Smithsonian Tropical Research Institute, Balboa, Republic of Panama

18 <sup>4</sup>Instituto de Investigaciones en Ecosistemas y Sustentabilidad, Universidad Nacional Autónoma  
19 de México, Morelia, Mexico

20 <sup>5</sup>Escuela Nacional de Estudios Superiores, Universidad Nacional Autónoma de México, Mérida,  
21 Mexico

22 <sup>6</sup>Departamento de Ciências Biológicas, Universidade Estadual de Santa Cruz, Ilhéus, Brazil

23 <sup>7</sup>Department of Plant Ecology, Bayreuth Center of Ecology and Environmental Research,  
24 University of Bayreuth, Bayreuth, Germany

25 <sup>8</sup>Department of Agricultural and Environmental Sciences, University of Milan, Milan, Italy

26 <sup>9</sup>Institute of Botany, University of Natural Resources and Life Sciences, Vienna, Austria

27 <sup>10</sup>Departamento de Sistemática e Ecologia, Universidade Federal da Paraíba, João Pessoa, Brazil

28 <sup>11</sup>School of Environmental Sciences, University of East Anglia, Norwich Research Park,  
29 Norwich, UK

30 <sup>12</sup>Departamento de Ecologia, Universidade Federal do Rio Grande do Sul, Porto Alegre, Brazil

31 <sup>13</sup>Department of Biological Sciences, Macquarie University, Sydney, Australia

32 <sup>14</sup>Forest Ecology and Forest Management group, Wageningen University & Research,  
33 Wageningen, The Netherlands

34 <sup>15</sup>World Agroforestry (ICRAF), Nairobi, Kenya

35 <sup>16</sup>Estonian University of Life Sciences, Tartu, Estonia

36 <sup>17</sup>Institute of Systematic Botany and Ecology, Ulm University, Ulm, Germany

37 <sup>18</sup>Departamento de Ecologia, Universidade de São Paulo, São Paulo, Brasil

38 <sup>19</sup>Tropical Botany, Naturalis Biodiversity Center, Leiden, The Netherlands

39 <sup>20</sup>Department of Biology, University of Maryland, Maryland, U.S.A.

40 <sup>21</sup>Field Museum of Natural History, Chicago, USA

41 <sup>22</sup>Morton Arboretum, Illinois, USA

42 <sup>23</sup>Lancaster Environment Centre, Lancaster University, United Kingdom

43 <sup>24</sup>Faculty of Science, Universiti Brunei Darusallam, Gadong, Brunei

44 <sup>25</sup>Department of Botany, Duke University, North Carolina, U.S.A.

45 <sup>26</sup>Department of Ecology and Evolutionary Biology, University of California, Los Angeles,  
46 U.S.A.

47 <sup>27</sup>School of Biological Sciences, Southern Illinois University, Carbondale, U.S.A.

48 <sup>28</sup>Grupo de Investigación en Biodiversidad Medio Ambiente y Salud, Universidad de las  
49 Américas, Quito, Ecuador.

50 <sup>29</sup>Escuela de Biología, Universidad de Costa Rica, San José, Costa Rica

51 <sup>30</sup>Instituto de Biologia, Universidade Federal da Bahia, Salvador, Brazil

52 <sup>31</sup>Max Planck Institute for Biogeochemistry, Jena, Germany

53 <sup>32</sup>German Centre for Integrative Biodiversity Research, Halle-Jena-Leipzig, Germany

54

55 \*Corresponding author: [bxpinho@hotmail.com](mailto:bxpinho@hotmail.com)

56

## 57 **Abstract**

58 **Aim:** Here we examine the functional profile of regional tree species pools across the latitudinal  
59 distribution of Neotropical moist forests, and test trait-climate relationships among local  
60 communities. We expected opportunistic strategies (acquisitive traits, small seeds) to be  
61 overrepresented in species pools further from the equator, but also in terms of abundance in local  
62 communities in currently wetter, warmer and more seasonal climates.

63 **Location:** Neotropics.

64 **Time period:** Recent.

65 **Major taxa studied:** Trees.

66 **Methods:** We obtained abundance data from 471 plots across nine Neotropical regions,  
67 including ~100,000 trees of 3,417 species, in addition to six functional traits. We compared

68 occurrence-based trait distributions among regional species pools, and evaluated single trait-  
69 climate relationships across local communities using community abundance-weighted means  
70 (CWM). Multivariate trait-climate relationships were assessed by a double-constrained  
71 correspondence analysis that tests both how CWMs relate to climate and how species  
72 distributions, parameterized by niche centroids in climate space, relate to their traits.

73 **Results:** Regional species pools were undistinguished in functional terms, but opportunistic  
74 strategies dominated local communities further from the equator, particularly in the northern  
75 hemisphere. Climate explained up to 57% of the variation in CWM traits, with increasing  
76 prevalence of lower-statured, light-wooded and softer-leaved species bearing smaller seeds in  
77 more seasonal, wetter and warmer climates. Species distributions were significantly but weakly  
78 related to functional traits.

79 **Main conclusions:** Neotropical moist forest regions share similar sets of functional strategies,  
80 from which local assembly processes, driven by current climatic conditions, select for species  
81 with different functional strategies. We can thus expect functional responses to climate change  
82 driven by changes in relative abundances of species already present regionally. Particularly,  
83 equatorial forests holding the most conservative traits and large seeds are likely to experience the  
84 most severe changes if climate change triggers the proliferation of opportunistic tree species.

85

86 *Key-words:* climate change, climate seasonality, community assembly, functional composition,  
87 functional traits, latitude, precipitation, species pool, temperature.

## 88 **Introduction**

89 Plants have evolved a broad range of functional strategies to cope with diverse environmental  
90 conditions (Díaz et al., 2016; Pierce et al., 2017). The functional assembly of plant communities  
91 results from the interplay among eco-evolutionary forces operating at different spatial-temporal  
92 scales (Kraft & Ackerly, 2014). At regional scales, the diverse functional strategies found in any  
93 given species pool reflect long-term speciation, dispersal and extinction filters (Mittelbach &  
94 Schemske, 2015). For instance, long-term climatic instability and natural disturbance regimes,  
95 such as hurricanes, storms and forest expansion-retraction dynamics due to glacial cycles, may  
96 select for functional profiles that favor population persistence under unstable conditions, while  
97 disturbance-sensitive species may be rare or even absent from regions under such conditions  
98 (Balmford, 1996; Betts et al., 2019). At local scales, the functional profile of plant communities  
99 depends on the filtering of regionally available species across varying current climate regimes  
100 (Swenson et al., 2012; Cadotte & Tucker, 2017). Assessing changes in functional composition of  
101 regional species pools and local communities along wide (bio)geographic and climatic gradients  
102 can help to understand potential responses to climate change and other human-caused disturb-  
103 ances (Violle et al., 2014). For instance, global climate change will soon bring unprecedented ex-  
104 treme climates to the Neotropics (Mora et al., 2013). Therefore, assessing how tree communities  
105 are functionally structured by trait-climate relationships helps predict the future of Neotropical  
106 forests in a rapidly changing world.

107       The advent of global plant trait databases in recent decades has enabled numerous  
108 investigations of patterns of trait variation and their relationships with climatic and  
109 biogeographic gradients (e.g., Swenson et al., 2012). These studies have revealed intriguing  
110 patterns, such as the tendency of plant species in warmer and less seasonal sites (closer to the

111 equator) to be taller and bear larger and softer leaves, larger seeds and denser woods (Wright et  
112 al., 2004, 2017; Moles et al., 2007, 2009; Swenson et al., 2012). However, these large-scale trait  
113 patterns were described mostly from species occurrence data across spatial grid-cells or  
114 latitudinal bands, and therefore failed to account for ecological processes operating at local  
115 scales that govern abundance of species and ultimately the functional profile of plant  
116 assemblages. On the other hand, studies that have assessed variation in abundance of species and  
117 their traits in local communities are based on either a single regional flora (e.g., van der Sande et  
118 al., 2016) or a single trait (e.g., Swenson & Enquist, 2007), and thus are unable to capture species  
119 assembly processes along wide biogeographic and climatic gradients. Scaling up abundance-  
120 based analyses of local communities to biogeographic scales can improve our understanding  
121 about climatic effects on local trait dominance, which ultimately drives ecosystem functioning  
122 (Poorter et al., 2017). In this way, Bruelheide et al. (2018) recently used a large dataset to  
123 examine global trait-environment relationships at the local community level (including  
124 abundance data), and found only weak support ( $R^2 < 0.1$ ) for trait-climate relationships. These  
125 global-scale analyses, though insightful, can mask relevant patterns within biotas that share a  
126 relatively common (but diverse) biogeographic history, such as Neotropical moist forests.

127 Neotropical moist forests extend from southern Mexico to northern Argentina and  
128 represent an enormous variation in past and current climatic conditions (Frierson et al., 2013;  
129 Blonder et al., 2018) and biogeographic histories (Gentry, 1982; Burnham & Graham, 1999).  
130 Overall, these differences clearly result in distinct taxonomic and phylogenetic composition more  
131 or less packed into biogeographic provinces. For instance, tropical moist forests of Meso-  
132 America (including Mexico) are taxonomically distinct from those in South America; the flora of  
133 the latter being mostly of Gondwanan origin while the northern Neotropics supports many plant

134 lineages with Laurasian affinity (Gentry, 1982; Graham, 1999). Also, palynological evidence  
135 points to a higher frequency of past disturbance events and faster recovery of tropical forests in  
136 northern Central America compared to South American counterparts (Cole et al., 2014). Mexican  
137 forests are the northern limit of the Neotropical forest distribution and experienced repeated  
138 expansion-retraction cycles due to Pleistocene glaciations (Burnham & Graham, 1999; Graham,  
139 1999), compared to South America, where many large blocks of forests remained stable during  
140 the last glacial and the influence of the Andes and the South American dry diagonal corridor is  
141 remarkable (Colinvaux et al., 2000; Hoorn et al., 2010; Leite et al., 2016).

142         While assessing relationships between traits and current climate is straightforward,  
143 addressing the effects of biogeographic history is challenging. Historical contingencies such as  
144 speciation/extinction dynamics and dispersal events must have affected the functional structure  
145 of current species pools and different drivers might act across localities (Fukami, 2015). Distance  
146 from the equator is related to current climate seasonality but has also been used as a proxy of  
147 biogeographic history, from plants to mammals, given its correlation to past cycles of climate  
148 change (Dynesius & Jansson, 2000; Betts et al., 2019). It is thus reasonable to expect that  
149 tropical biotas far from the equator experienced, currently and in the past, more shifting climates  
150 than their equatorial counterparts (Blonder et al., 2018; Betts et al., 2019). Such instability might  
151 select for opportunistic strategies related to fast growth and high dispersal ability. For instance,  
152 northern forests are mostly composed of broad-ranged plant species due to short- and long-term  
153 climatic instability, while small-ranged species concentrate under stable climates in Central  
154 America, Amazonia and Atlantic forests (Morueta-Holme et al., 2013), which are relatively  
155 equatorial regions that also support higher phylogenetic endemism (Sandel et al., 2020). In  
156 contrast, extreme southern Neotropical vegetation has developed under relatively low and

157 seasonal temperature and precipitation levels (Oliveira-Filho et al., 2013). The extent to which  
158 such historical contingencies can induce distinct signatures on the functional composition of  
159 Neotropical moist forests is yet to be fully understood.

160         Several key aspects of community functional composition can be expressed through the  
161 ‘global spectrum of plant form and function’ (Díaz et al., 2016). Specifically, plants well adapted  
162 to resource-poor/stressful environments with low disturbance regimes tend to grow slowly (i.e.,  
163 low metabolic resource demand) and invest in dense, durable tissues (i.e., conservative traits). In  
164 contrast, acquisitive resource-use traits (e.g., low-density woods, soft leaves) favor hydraulic  
165 efficiency and rapid plant growth, allowing resource pre-emption in productive habitats such as  
166 those under wetter and warmer climates (Westoby et al., 2002; Reich, 2014). Such opportunistic  
167 strategy can also benefit under more seasonal climates by optimizing carbon gain during the  
168 growing season in the more open forests that allow more light to reach the understory (Kobe,  
169 1999; Kikuzawa et al., 2013). Regarding size-related traits, increasing leaf area favors light  
170 capture, but limits heat exchange with the surrounding air, and leads to a higher daytime  
171 transpirational water loss, thereby being favored in warm, moist and sunny environments  
172 (Wright et al., 2017). Also, larger seeds may promote higher seedling performance under low  
173 resource availability (Leishman & Westoby, 1994; Muller-Landau, 2010), while smaller-seeded  
174 species have greater seed output that favors dispersal to recently disturbed sites and seeds that  
175 are more likely to exhibit dormancy, which favors survival under variable climates (de Casas et  
176 al., 2017). Finally, larger trees tend to have greater access to light and belowground resources,  
177 but are more prone to hydraulic failure during drought (Brum et al., 2019). Combinations of  
178 these traits define ecological strategies that influence plant responses to environmental  
179 conditions (Grime & Pierce, 2012).



180 Here we test two mutually compatible effects of biogeographic history and current climate  
181 as structuring drivers of the functional organization of tree communities across Neotropical moist  
182 forests. If historical contingency prevails, then we should expect functional dissimilarities among  
183 regional tree species pools, which could lead to differences in functional composition of local  
184 communities occurring in similar climates at different regions. If current climate represents a  
185 prevailing force, functional differences should emerge at the local scale due to changes in trait  
186 dominance in response to climatic conditions. In particular, we expected that regional species  
187 pools should be composed of different sets of functional strategies, with higher prevalence of  
188 species with opportunistic ecological strategies (i.e., low-density tissues, small seeds) in regions  
189 further from the equator due to long-term instability that selects for fast-growth and high disper-  
190 sal ability. Across local communities, more seasonal, wetter and warmer climates should favor  
191 dominance of opportunistic strategies. We additionally assessed the consistency of trait-climate  
192 relationships by evaluating to what extent the distribution of species, expressed as the abun-  
193 dance-weighted mean climatic conditions at which they are found (i.e., niche centroids), is medi-  
194 ated by functional traits. We discuss our results in terms of how useful they are for the under-  
195 standing of both community assembly patterns and potential responses of Neotropical tree floras  
196 to climate change and anthropogenic disturbances in human-dominated landscapes.

197

## 198 **Methods**

### 199 *Study regions and plots*

200 We studied 471 forest plots from nine biogeographic regions distributed across the Neotropics,  
201 covering the whole latitudinal distribution of Neotropical moist forests (Fig. 1; see Table S1 in  
202 Supporting Information for details on sampling across regions). All plots were located in lowland

203 (up to 800 m a.s.l.), old-growth forests within a variable matrix of land uses. Mean annual  
204 precipitation ranged from 1,154 to 7,068 mm, and mean annual temperature from *c.* 17 to 28 °C  
205 (Fig. S1). Temperature seasonality increases with distance from equator (Wright et al., 2009),  
206 while average temperature and precipitation are typically higher towards the northern Neotropics  
207 due to northward heat transport by ocean circulation (Frierson et al., 2013; Fig. 1).

208

### 209 *Vegetation data*

210 We used data from 96,290 live adult trees (stems with diameter at breast height, DBH  $\geq$  10 cm;  
211 excluding lianas and palms) belonging to 3,417 species. Tree inventories were carried out by the  
212 authors as described elsewhere (Pitman et al., 2001; Santos et al., 2008; Arroyo-Rodríguez et al.,  
213 2009; Faria et al., 2009; Hernández-Ruedas et al., 2014; Benchimol & Peres, 2015; Orihuela et  
214 al., 2015; Pinho et al., 2018) or compiled from the “Salvias” database through the Botanical  
215 Information and Ecology Network - ‘BIEN’ R package (Maitner et al., 2017), which includes the  
216 Gentry plots (Gentry, 1988). The sampled area and total number of individuals and species  
217 sampled by region are summarized in Table S1. The slight difference in sampling methods (e.g.,  
218 plot sizes) should not affect our results as we focus on the *relative* dominance of functional traits  
219 and strategies within communities and the resulting variation across the Neotropics.

220

### 221 *Functional traits*

222 A comprehensive set of six traits was measured in the field (following Pérez-Harguindeguy et  
223 al., 2013) and compiled from global databases, such as ‘TRY’ (Kattge et al., 2020) and the ‘Seed  
224 Information Database’ – SID (Royal Botanic Gardens Kew, 2020). These traits are leaf area - LA  
225 [cm<sup>2</sup>], specific leaf area – SLA [cm<sup>2</sup>/g], leaf dry matter content –LDMC [mg/g], wood density –

226 WD [ $\text{g}/\text{cm}^3$ ], seed mass – SM [mg] and maximum height – Hmax [m]. We chose these traits  
227 because they are known to influence tree performance along climatic gradients (Westoby et al.,  
228 2002; Reich et al., 2014), and position species along the plant (and organ) economics and size-  
229 related traits spectra (Díaz et al., 2016; Pierce et al., 2017). For instance, the leaf and stem traits  
230 considered are expected to reflect a trade-off between rapid resource acquisition that enables  
231 growth in resource-rich environments (indicated by high SLA, low LDMC, low WD), and  
232 conservation of resources in well-protected tissues that ensure survival under low resource  
233 availability, indicated by the opposite traits (Reich, 2014).

234       For leaf traits in compound-leaved species, we considered leaflets as the sample unit.  
235 Although we recognize the importance of intraspecific trait variation in community assembly  
236 (Siefert et al., 2015), we used species' mean functional traits as we consider a meaningful  
237 approach for the purpose of this study due to the extensive species-level trait data and high  
238 species turnover among regions. Species-level trait data covered on average from 57 to 80% of  
239 total plot abundances across traits/regions (see Table S2 for a summary of trait coverage by  
240 region). For species with doubtful identification and/or no trait information, we first used  
241 average trait values at the genus level, then we imputed remaining missing values (for which no  
242 genus-level data were available) through multivariate trait imputation with chained equations by  
243 predictive mean matching, using the R package 'mice' (van Buuren & Groothuis-Oudshoorn,  
244 2011). The imputed trait data represented 3% or less of individuals in plots for 40 of the 54  
245 region-trait combinations and over 10% for just one combination (see Table S2 for a summary  
246 by region), and the distribution of the original and imputed datasets largely overlapped (Fig. S2).  
247 Genus-level trait means were well correlated with species mean traits (Table S3), and their  
248 inclusion led to similar distributions of CWM trait values (Fig. S3). These findings demonstrate

249 that our results are not due to spurious artefacts in the imputation of missing trait data. Also, the  
250 exclusion of trees with DBH < 10 cm should not represent a significant bias because adults  
251 covered the whole range of functional strategies evident among saplings, and abundance-  
252 weighted distributions of species trait values largely overlap when considering smaller trees (see  
253 example for Northern Meso-America, Fig. S4).

254

#### 255 *Climate data*

256 For each plot, we assessed the average of five key bioclimatic variables that are thought to drive  
257 trait distributions and vegetation patterns (Swenson et al., 2012; Moles, 2018). The five climatic  
258 variables include mean annual precipitation (MAP), mean annual temperature (MAT),  
259 precipitation seasonality (PS – coefficient of variation of monthly values), temperature  
260 seasonality (TS – standard deviation of monthly values, multiplied by 100) and potential  
261 evapotranspiration (PET – the amount of water expected to be removed by the atmosphere  
262 through evapotranspiration processes annually). The first variables were obtained from  
263 WorldClim version 2.0 (Fick & Hijmans, 2017), which is a high-resolution global geo-database  
264 (30 arc seconds or ~ 1 km at equator) of monthly average data from 1970 to 2000. The PET was  
265 calculated from a set of WorldClim variables (taken in the same timeframe as above), using an  
266 equation proposed by the Food and Agriculture Organization of the United Nations, which  
267 involves minimum, maximum and average temperature, solar radiation, wind speed and water  
268 vapor pressure (Trabucco & Zomer, 2018). Other climatic variables were considered but then  
269 excluded due to collinearity (see below, Table S4). The five climatic variables considered were  
270 weakly inter-correlated (Table S4), but were strongly related to latitude (i.e., south-north  
271 gradient) or degrees from equator (Fig. S1). Despite complex climate variability due to, for

272 example, ocean circulation and elevation (Frierson et al., 2013), in this dataset temperature  
273 seasonality was strongly positively correlated with degrees from equator, while other climatic  
274 variables (MAP, MAT, PET, PS) increased linearly from south to north (Fig. S1, see Fig. 1 for  
275 the overall climatic pattern across the Neotropics).

276

### 277 *Data analyses*

278 We log-transformed LA, SLA and SM values, and sqrt-transformed Hmax to reduce skewness in  
279 trait distributions. We also log-transformed MAP to reduce the influence of two exceptionally  
280 wet sites. Functional composition of regional species pools was described from distributions of  
281 the traits of species occurring in each region. Functional traits were scaled-up from the species-  
282 level to the plot-level by calculating the Community-Weighted Mean ('CWM' – i.e., species'  
283 trait values weighted by their relative abundances), which reflects the dominance of trait values  
284 in a community (Muscarella et al., 2017). CWM trait values were calculated using function  
285 'functcomp' from the 'FD' R package (Laliberté & Legendre, 2010). To examine trait co-  
286 variation patterns among species and communities, we applied Principal Component Analyses to  
287 the species and CWM trait matrices, using the 'prcomp' R function (Venables & Ripley, 2002).  
288 We also computed CWM of species scores on the first two principal component axes, which  
289 should reflect economics and size trade-offs in functional strategies (Díaz et al., 2016).

290 To assess changes in community functional composition in response to climate or  
291 geography, we constructed separate linear mixed-effects models, for the CWM of each  
292 functional trait or strategy (i.e., species scores on the PCA axes, see above). The fixed effects  
293 were either the five bioclimatic variables described above or the geographic variables, latitude  
294 (to describe south-to-north gradients) and degrees from equator (to describe gradients toward

295 higher latitudes in both hemispheres). The random effect ‘biogeographic region’ was included in  
296 all models to account for the nested structure of our sampling design, and to assess among-region  
297 variation not explained by latitude or climate. To avoid multicollinearity between climatic or  
298 geographic variables, we checked the variance inflation factor of each predictor in each model,  
299 using the ‘car’ package for R. All values were  $< 3$ , which allowed us to include all five climate  
300 variables or the two geographic variables in the models (Neter et al., 1996). After running a full  
301 model with each set of predictors (i.e., climatic and geographic variables) for each response  
302 variable (i.e., CWM of each trait and PCA axes scores) using the maximum likelihood method  
303 with the R package ‘lme4’ (Bates et al., 2015), we tested all possible combinations of predictors  
304 and performed a model selection procedure to select the best-fit models as those with lowest  
305 Akaike Information Criterion values ( $\Delta AICc < 2$ ). Then, we applied model averaging to make  
306 inferences on how individual climatic variables influence CWM of traits and strategies, using the  
307 ‘MuMin’ R package (Barton, 2014).

308       To assess the variance in CWM of traits and strategies (i.e., species scores on principal  
309 component axes, see above) among regions and the strength of their relationships with latitude or  
310 climate, we partitioned the  $R^2$  of each selected model into the total variance between-regions  
311 (“conditional  $R^2$ ”) and the component explained by climate or latitude (“marginal  $R^2$ ”;  
312 Nakagawa & Schielzeth, 2013), reporting the variance explained by the model with highest  
313 marginal  $R^2$  for each response variable. The difference between conditional and marginal  $R^2$   
314 values represents the variance between-regions not explained by climate/latitude (expressed as  
315 fraction of the total variance). The within-region component is the remaining unexplained  
316 variance (i.e.,  $1 - \text{conditional } R^2$ ). For this, we used the R package ‘piecewiseSEM’ (Lefcheck,  
317 2016).

318 To assess composite trait-climate relationships at both species- and community-level, we  
319 applied double constrained correspondence analysis (dc-CA; ter Braak et al., 2018). The dc-CA  
320 method is a new and powerful regression-based approach, similar to the covariance-based three-  
321 table ordination RLQ method used to assess multivariate trait-environment relationships in what  
322 is known as the fourth-corner problem (Dray & Legendre, 2008). Like RLQ, dc-CA uses three  
323 data tables (trait values of species, environmental conditions of sites, and abundances of species  
324 per site) to define the correlation between traits and environmental conditions (i.e., the fourth-  
325 corner correlation). The fourth-corner correlation has proved to be powerful to test trait-  
326 environment relationships (Peres-Neto et al, 2017; ter Braak 2017). dc-CA searches for linear  
327 combinations of traits and environmental variables that maximize their fourth-corner correlation,  
328 using weighted least-squares, where the weights for species and for sites are their total  
329 abundance. In contrast, RLQ maximizes covariance not correlation. By maximizing the fourth-  
330 corner correlation, dc-CA considers the influence of environmental conditions on community  
331 functional composition (i.e., CWM traits) in combination with how species (environmental)  
332 niche centroids (SNC) relate to their traits (ter Braak et al., 2018). SNCs represent the mean  
333 climatic conditions where species are found (weighted by abundances) and are related to species'  
334 traits to discover whether trait-mediated mechanisms influence species' distributions.  
335 Specifically, the SNC with respect to environmental variable  $e$  is a weighted mean, calculated as  
336  $u_j = \sum_{i=1}^n y_{ij} e_i / \sum_{i=1}^n y_{ij}$ , where  $y_{ij}$  refers to the abundance of the  $j^{th}$  species in the  $i^{th}$  site, and  $e_i$  is  
337 the value of the environmental variable at the  $i^{th}$  site.

338 We additionally applied dc-CA considering geographic gradients (latitude, longitude and  
339 degrees from equator) instead of climate variables as predictors, and performed variation  
340 partitioning to define the separate and shared effects of geographical and climatic gradients. We

341 used the dc-CA based max-test to check significance of the dc-CA axes (ter Braak et al., 2017).  
342 The max-test solves the problem of inflated type I error rate in the fourth-corner approach  
343 (Peres-Neto et al, 2017) by applying two independent permutations for testing species- (SNC ~  
344 traits) and community-level (CWM ~ climate) patterns, and selecting the highest p-value. We  
345 applied the max-test after aggregating plots separated by less than 50 km (Fig. S5) to avoid  
346 pseudo-replication caused by nearby plots. In the analyses using dc-CA, the issue of the two  
347 exceptionally wet sites was solved by replacing their MAP values with the value 4500 mm/year,  
348 slightly higher than the maximum in the data set; this gave a slightly higher fit than the log-  
349 transformation, but did not give qualitatively different results. We performed the dc-CA using  
350 the software Canoco 5.12 (ter Braak & Šmilauer, 2018).

351

## 352 **Results**

### 353 *Functional composition of Neotropical moist forest regions*

354 The functional composition of tree species pools largely overlapped across regions (Fig. 2a), but  
355 strong differences among regions emerged from abundance-weighted trait values at the local  
356 community-level (i.e., CWM) (Fig. 2b). Tree communities in forests near the equator (e.g.,  
357 Amazonia, Northeastern Atlantic forest) were dominated by taller species with larger seeds,  
358 harder woods and greater leaf dry matter content. Tree communities in regions further from the  
359 equator were dominated by lower-statured species with smaller seeds and lower LDMC,  
360 particularly in the northern hemisphere (Fig. 2b). CWM values of wood density and maximum  
361 height were, however, relatively high at the extreme south (i.e., southeastern Atlantic forests),  
362 where specific leaf area achieved the lowest values (Fig. 2b).



363           The first two principal component axes of species-level trait values captured 55% of the  
364 variation in the functional space composed by six traits (Fig. 3a). The first PC axis indicated a  
365 common spectrum of variation among economic- and regenerative-traits, varying from species  
366 with acquisitive traits (i.e., high SLA, low LDMC and WD) and small seeds (i.e., opportunistic  
367 strategies), to species with more conservative strategies (i.e., low SLA, high LDMC and WD)  
368 and larger seeds (Fig. 3a, Table S5). The second axis mainly reflects variation in leaf area and  
369 maximum height, which co-varied positively (Fig. 3a, Table S5). Variation in SLA was mostly  
370 captured by a third PC axis (Table S5).

371           The first two principal components of community-level trait values (i.e., CWM) captured  
372 more variation (74%) and revealed similar trade-offs (Fig. 3b), except maximum height was  
373 strongly related to the first PC axis (Table S5). Tree communities from different regions could be  
374 distinguished along the first two PC axes (Fig. 3b). Specifically, the first axis indicated a gradient  
375 from communities dominated by species with conservative traits (high WD and LDMC) in  
376 equatorial regions and in the extreme south of the Neotropical forest biome, to a more acquisitive  
377 (high SLA) community composition in Northern forests. In turn, the second community trait axis  
378 distinguished communities at the Southeastern Atlantic region and Caribbean Islands as  
379 dominated by species with smaller leaves compared to more equatorial forests, particularly those  
380 across North-Amazon and Southern Meso-America (Fig. 3b).

381

### 382 *Trait-climate relationships across Neotropical tree communities*

383 Current climate explained 16 to 57% of the variation in CWM trait values across Neotropical  
384 moist forests (Table 1a, Fig. 4). Temperature seasonality presented the strongest relationships  
385 with CWM traits, except for leaf area which was more strongly related to mean annual

386 precipitation, and SLA which was not responsive to climate variation (Table 1a). LDMC, SM,  
387 WD and Hmax decreased with increasing temperature seasonality, while mean annual  
388 temperature had similar (but much weaker) effects on the last three of these traits (Table 1a).  
389 Additionally, increasing annual precipitation was associated with increased dominance of tree  
390 species with lower wood density, larger organs and lower leaf dry matter content, while  
391 precipitation seasonality was negatively related to seed mass (Table 1a). The species functional  
392 strategies evident on the first two PC axes (Fig. 3a) also changed predictably in response to  
393 climatic variables (Table 1b): PC1 (acquisitive to conservative resource economy) was strongly  
394 negatively related to temperature seasonality, while PC2 (small to large plants and organs)  
395 increased mainly with mean annual precipitation (Fig. 4).

396         The first two dc-CA axes revealed significant (max-test,  $P = 0.001$ ) composite trait-climate  
397 relationships (Fig. 5, Table S6). The first axis describes a gradient from environments with  
398 relatively high seasonality in temperature and precipitation, combined with high annual  
399 precipitation and potential evapotranspiration, to less seasonal climates, along which there was a  
400 shift in dominance from shorter plants with relatively acquisitive traits and small seeds, to slow-  
401 growing species with conservative traits and large seeds (Fig. 5a). This first axis separated  
402 communities across Northern regions from those at more equatorial regions (Fig. 5b). The  
403 second dc-CA axis was mostly explained by variation in MAP and PET, reflecting a gradient  
404 from drier sites under high potential evapotranspiration to exceptionally wet sites (Fig. 5a). This  
405 second gradient explained the variation in dominance from small-leaved species with high  
406 woody density across southeastern Atlantic forests and Caribbean Islands, to large-leaved, soft-  
407 wooded species in communities across northern regions, particularly the Chocó bioregion and  
408 North-Western Amazon (Fig. 5b).

409 Climate variables were good predictors of taxonomic composition across communities  
410 (CCA eigenvalues of 0.8 and 0.7 for the first two axes), as well as of multivariate gradients in  
411 community abundance-weighted traits (46% of variance explained; Table S6). In turn, traits were  
412 weak predictors of the distribution of individual species abundances across communities,  
413 explaining only 4% of variation in species climate niche centroids (Table S6). Forward selection  
414 on climate variables revealed that three of the five climate variables (MAP, TS and PET) account  
415 for most variation in composite trait-climate gradients (Fig. S6).

416

#### 417 *Geographic gradients in community functional composition*

418 All community weighted mean traits were significantly related to either latitude or degrees from  
419 equator. LA, LDMC, WD and SM decreased with increasing degrees from equator, while SLA  
420 increased and Hmax decreased with latitude (i.e., from southern to northern forests; Fig. S7).  
421 Latitude was more strongly related to vegetation patterns (i.e., taxonomic turnover across  
422 communities), while degrees from equator explained relatively more of functional variation  
423 (Table S7). Variance partitioning revealed unique and shared effects of geography and climate,  
424 combining to explain 66% of the variation in CWM trait values (Table 2). Most of this explained  
425 variation (39%) resulted from shared effects of geographical gradients and climate variables,  
426 though there were also unique effects of similar size from both climate (15%) and geography  
427 (12%) (Table 2).

428

#### 429 **Discussion**

430 Species pools of Neotropical moist forest regions from southern Mexico to southern Brazil  
431 possess similar distributions of trait values. Long-term filters that can control for the functional

432 composition of regional species pools thus have little to no importance. In contrast, local tree  
433 communities are functionally structured along climatic and (bio)geographical gradients. Species  
434 assembly processes that govern local abundance of species in tree communities must have  
435 generated the documented functional dissimilarities. The novelty of our findings is that we found  
436 clear though complex trait-climate relationships across Neotropical moist forests, that are not  
437 simply driven by geography. The observed patterns suggest an increase in relative abundance of  
438 lower-statured, light-wooded and softer-leaved species bearing smaller seeds (i.e., opportunistic  
439 strategies) under more seasonal climates in communities further away from the equator,  
440 especially under wetter and warmer conditions across northern forests. In contrast, communities  
441 in more stable climates (mostly close to the equator) are dominated by species with large seeds  
442 and conservative traits (i.e., the typical functional profile of late-successional tree species).  
443 Trait-climate relationships at the species-level (i.e., Species Niche Centroids ~ traits) were also  
444 significant but weaker, suggesting either that changes in the abundance of dominant species are  
445 responsible for varying functional signatures across Neotropical forests, or that there is scope for  
446 improvement of the trait set. It is important to recognize that dominance of functional strategies  
447 differ between regions mainly due to local assembly processes related to climate rather than  
448 changes in species pools that would be the result of biogeographic history. This helps to  
449 understand how Neotropical forests may respond to climate change and other human-imposed  
450 disturbances.

451       Our results contrast with those of Bruelheide et al. (2018), in which functional composition  
452 of plant communities (abundance-based) was weakly related to climate at the global scale. At the  
453 continental/biome scale (i.e., within Neotropics), we found strong trait-climate relationships  
454 across Neotropical moist forest tree communities, with combinations of climatic variables

455 explaining up to 57% of variance in CWM trait values. Changes in trait dominance were mainly  
456 driven by the increase in temperature seasonality with distance from the equator and, to a lesser  
457 extent, by changes in precipitation regimes, average temperature and potential evapotranspiration  
458 (Table 1) that are less clear in geographic terms (Fig. 1). Slightly different from our findings,  
459 previous assessments of global trait distributions suggest that higher mean annual temperature  
460 and/or precipitation in equatorial regions leads to increased prevalence of conservative traits and  
461 larger seeds (e.g., Swenson et al., 2012; Moles et al., 2014; Bruelheide et al., 2018). This  
462 discrepancy may arise because global patterns of trait distribution may reflect major differences  
463 among predominant biomes across climatic zones. Also, global-scale studies usually include  
464 both woody and herbaceous species (e.g., Bruelheide et al., 2018), which respond differently to  
465 climate across the Neotropics (Šímová et al., 2018). Most importantly, global-scale studies  
466 include dry forests, where the combination of high temperatures with low and highly seasonal  
467 precipitation may represent a physical stress, favoring conservative strategies (Westoby et al.,  
468 2002). As our study focuses on Neotropical moist forests, higher temperatures and precipitation  
469 should actually favor acquisitive traits of trees, as we found, due for example to increased  
470 hydraulic efficiency (Zhang et al., 2013; Santiago et al., 2018). Our findings strengthen the  
471 notion that climatic conditions play a key role in trait filtering across Neotropical tree  
472 assemblages.

473         Despite differences in magnitude of climate effects, the direction of community-level trait-  
474 climate relationship (i.e., including abundance data) in the Neotropics generally agrees with that  
475 of species-level global trait-climate relationships (e.g., Swenson et al., 2012; Wright et al., 2017,  
476 see Moles, 2018 for a review of these relationships). For example, we found that seed mass tends  
477 to be lower in plants growing at more seasonal sites further from the equator, as observed

478 elsewhere (Moles et al., 2007; Swenson et al., 2012; Malhado et al., 2015), presumably because  
479 larger seeds require longer growing seasons for development and are less likely to exhibit  
480 dormancy that helps to survive under adverse seasonal conditions (Thompson et al., 1993; de  
481 Casas et al., 2017). Moreover, tree species in wetter sites tend to present larger leaves and softer  
482 wood and leaf tissues, as we found, because these traits maximize resource capture in productive  
483 environments (Westoby et al., 2002; Wright et al., 2017). Conversely, small leaves with low  
484 specific leaf area (SLA) characterize cold tolerance (Poorter et al., 2009; Wright et al., 2017) as  
485 we observed in the southeastern Atlantic forests, a region that experiences relatively low and  
486 seasonal temperatures (Fig. S1; Oliveira-Filho et al., 2013). Also in agreement with our findings,  
487 plant height tends to be limited under more seasonal climates (Moles et al., 2009; Swenson et al.,  
488 2012), and non-significant relationships between SLA and climatic gradients is more a rule than  
489 an exception (Moles, 2018). Finally, large seeds and conservative traits may provide advantages  
490 for species in shaded conditions of relatively closed-canopy forests near the equator (Leishman  
491 & Westoby, 1994; Kitajima & Poorter, 2010).

492       While trait-climate relationships we observed are well supported by theory, they might to  
493 some extent be driven by (bio)geography, as some climatic variables were strongly correlated to  
494 either latitude or distance from equator. For instance, temperature seasonality increases sharply  
495 with distance to equator ( $R^2 = 0.81$ ), which similarly explain community functional composition  
496 (*cf.*, Figs. 4 and S7). We can thus only speculate about the relative importance of climatic and  
497 geographic gradients, as they share the largest fraction of explained variance in CWM traits.  
498 However, we note that both climate and geography have also unique effects on community func-  
499 tional composition (Table 2). To illustrate this, the Southern Brazilian Atlantic forest does not  
500 fully follow the trend of increasing dominance of opportunistic strategies with distance from

501 equator, as it supports many large, hard-wooded trees. This can, however, be explained by com-  
502 binations of climatic factors, given the seasonal but relatively cold and dry climates in the south-  
503 ern (compared to the northern) extreme of the Neotropics (Fig. S1), which select for conservative  
504 strategies. In fact, most species in this southernmost region have tropical-subtropical ranges due  
505 to forest expansion over subtropical grasslands during the last glacial maximum (Oliveira-Filho  
506 et al., 2013; Costa et al., 2017). Also, the dominance of species with small leaves and conserva-  
507 tive traits under wet and warm climates in the Caribbean region deviates from the trait-climate  
508 relationships, but may reflect selection by hurricane-force winds for stronger structural support  
509 (Lugo, 2000).

510       There is a consensus that climate change will make tropical forests warmer, with more  
511 seasonal rainfall and temperature including more frequent droughts, more heavy rains and  
512 frequent heatwaves (IPCC, 2014). These are conditions currently found across northern  
513 Neotropical moist forest regions, where opportunistic strategies thrive. Therefore, it is reasonable  
514 to expect forests in northern Neotropical regions to be more resilient to predicted climate  
515 changes, unless climate change leads to drastic changes towards alternative ecosystem states, like  
516 dry forest or woodlands. In contrast, climate changes should lead southern and especially  
517 equatorial forests of the future to more closely resemble the today's northern Neotropical moist  
518 forests due to proliferation of opportunistic strategies already present in regional species pools  
519 (Fig. 2a). This can disrupt ecosystem services such as carbon sequestration and storage if forests  
520 once dominated by conservative traits experience the proliferation of more opportunistic  
521 strategies (Poorter et al., 2017), including soft-wooded species that grow fast but die young  
522 (Brienen et al., 2020).

523 In summary, tree communities across Neotropical moist forests are functionally distinct  
524 because particular traits are favored under particular climates. Such functional predictability  
525 permits insights into tropical forest responses to global changes and the consequences for  
526 biodiversity persistence, provision of ecosystem services and global sustainability (Díaz et al.,  
527 2007). In fact, as ecosystem functioning is determined by the dominant traits in tropical forests  
528 (Poorter et al., 2017), our findings can help to anticipate the impact of future climate change  
529 and/or human-induced disturbances (e.g., habitat loss, fragmentation) on the functioning of  
530 Neotropical forests. For instance, we must expect an increase in prevalence of low-statured tree  
531 species with relatively acquisitive traits and small seeds, based on scenarios of increasing climate  
532 seasonality (IPCC, 2014). This functional strategy is associated with faster growth and  
533 reproductive rates (Reich, 2014; Moles, 2018) and thrives in human-modified tropical landscapes  
534 worldwide (Laurance et al., 2006; Santos et al., 2008), but plays a limited role in crucial  
535 ecosystem services such as carbon and nutrient retention (Poorter et al., 2017). Overall, this  
536 changing functional structure of Neotropical moist forests is likely to confer varying degrees of  
537 resilience to human-caused disturbances. Neotropical moist forests vary widely in functional  
538 terms and one should be aware of these differences when it comes to understand the functional  
539 assembly of Neotropical tree communities.

540

#### 541 **Acknowledgements**

542 This study was carried out with the financial support of the Coordenação de Aperfeiçoamento de  
543 Pessoal de Nivel Superior – Brasil (CAPES) – Finance code 001; and the Conselho Nacional de  
544 Desenvolvimento Científico e Tecnológico – CNPQ (grant 403770/2012-2). BXP was funded by  
545 a PhD scholarship from CNPq (140260/2015-3). BXP collected functional data in the Mexican



546 regions with the financial support of the Catedra José Sarukhán and the Centro del Cambio  
547 Global y la Sustentabilidad en el Sureste. BXP finished this study with the support of a  
548 postdoctoral fellowship from Fundação de Amparo à Ciência e Tecnologia do Estado de  
549 Pernambuco (FACEPE). FPLM, MT and BAS thank CNPq for productivity grants.

550

#### 551 **Data Accessibility**

552 Data and R codes used for the analyses are available from the Dryad Digital Repository:  
553 <http://doi.org/10.5061/dryad.vq83bk3s3>. The dc-CA variation decomposition with statistical tests  
554 of Table 2 is available at <https://doi.org/10.6084/m9.figshare.13259534.v2>

555

#### 556 **Authorship**

557 BXP, MT and FPLM conceived the ideas and designed methodology; BXP collected, compiled  
558 and analysed the data, with the supervision of CtB on the dc-CA analysis; BXP wrote the first  
559 draft with significant contribution from FPLM. All authors contributed data and/or with critical  
560 insights that improved the manuscript.

561

#### 562 **References**

563 Arroyo-Rodríguez, V., Pineda, E., Escobar, F., & Benítez-Malvido, J. (2009). Value of small  
564 patches in the conservation of plant-species diversity in highly fragmented rainforest.

565 *Conservation Biology*, 23, 729-739.

566 Balmford, A. (1996). Extinction filters and current resilience: the significance of past selection

567 pressures for conservation biology. *Trends in Ecology and Evolution*, 11, 193-196.

568 Barton, K. (2014). MuMin: Multi-model inference. R package version 1.10.0. [https://cran.r-](https://cran.r-project.org/web/packages/MuMIn/index.html)  
569 [project.org/web/packages/MuMIn/index.html](https://cran.r-project.org/web/packages/MuMIn/index.html)

570 Bates, D., Maechler, M., Bolker, B., & Walker, S. (2015). Fitting Linear Mixed-Effects Models  
571 Using lme4. *Journal of Statistical Software*, 67, 1-48.

572 Benchimol, M., & Peres, C. A. (2015). Edge-mediated compositional and functional decay of  
573 tree assemblages in Amazonian forest islands after 26 years of isolation. *Journal of*  
574 *Ecology*, 103, 408–420.

575 Betts, M. G., Wolf, C., Pfeifer, M., Banks-Leite, C., Arroyo-Rodríguez, V., ... Ewers, R. M.  
576 (2019). Extinction filters mediate the global effects of habitat fragmentation on animals.  
577 *Science*, 366, 1236-1239.

578 Blonder, B., Enquist, B. J., Graae, B. J., Kattge, J., Maitner, B. S., Morueta-Holme, N., ... Violle,  
579 C. (2018). Late Quaternary climate legacies in contemporary plant functional composition.  
580 *Global Change Biology*, 24, 4827-4840.

581 Brienen, R. J. W., Caldwell, L., Duchesne, L., Voelker, S., Barichivich, J., Baliva, M., ... Gloor,  
582 E. (2020). Forest carbon sink neutralized by pervasive growth-lifespan trade-offs. *Nature*  
583 *Communications*, 11, 4241.

584 Brown, J. H. (2014). Why are there so many species in the tropics? *Journal of Biogeography*, 41,  
585 8-22.

586 Bruelheide, H., Dengler, J., Purschke, O., Lenoir, J., Jiménez-Alfaro, B., Hennekens, S. M., ...  
587 Jandt, U. (2018). Global trait–environment relationships of plant communities. *Nature*  
588 *Ecology & Evolution*, 2, 1906-1917.

589 Brum, M., Vadeboncoeur, M. A., Ivanov, V., Asbjornsen, H., Saleska, S., ... Oliveira, R.S.  
590 (2019). Hydrological niche segregation defines forest structure and drought tolerance

591 strategies in a seasonal Amazon forest. *Journal of Ecology*, 107, 318-333.

592 Burnham, R. J., & Graham, A. (1999). The History of Neotropical Vegetation: New  
593 Developments and Status. *Annals of the Missouri Botanical Garden*, 86, 546-589.

594 Cadotte, M. W., & Tucker, C. M. (2017). Should environmental filtering be abandoned? *Trends*  
595 *in Ecology and Evolution*, 32, 429-437.

596 Cole, L. E. S., Bhagwat, S. A., & Willis, K. J. (2014). Recovery and resilience of tropical forests  
597 after disturbance. *Nature Communications*, 5, 3906.

598 Colinvaux, P. A., De Oliveira, P. E., & Bush, M. B. (2000). Amazonian and neotropical plant  
599 communities on glacial time-scales: The failure of the aridity and refuge hypotheses.  
600 *Quaternary Science Reviews*, 19, 141–169.

601 de Casas, R., Willis, C. G., Pearse, W. D., Baskin, C. C., Baskin, J. M., & Cavender-Bares, J.  
602 (2017). Global biogeography of seed dormancy is determined by seasonality and seed size:  
603 A case study in the legumes. *New Phytologist*, 214, 1527–1536.

604 Díaz, S., Kattge, J., Cornelissen, J. H. C., Wright, I. J., Lavorel, S., Dray, S., ... Gorné, L. D.  
605 (2016). The global spectrum of plant form and function. *Nature*, 529, 167–171.

606 Díaz, S., Lavorel, S., de Bello, F., Quétier, F., Grigulis, K., Robson, T. M. (2007). Incorporating  
607 plant functional diversity effects in ecosystem service assessments. *Proceedings of the Na-*  
608 *tional Academy of Sciences*, 104, 20684-20689.

609 Dray, S., & Legendre, P. (2008). Testing the species traits–environment relationships: the fourth-  
610 corner problem revisited. *Ecology*, 89, 3400-3412.

611 Faria, D., Mariano-Neto, E., Martini, A. M. Z., Ortiz, J. V., Montingelli, R., Rosso, S., ...  
612 Baumgarten, J. (2009). Forest structure in a mosaic of rainforest sites: The effect of

613 fragmentation and recovery after clear cut. *Forest Ecology and Management*, 257, 2226–  
614 2234.

615 Fick, S. E., & Hijmans, R. J. (2017). Worldclim 2: New 1-km spatial resolution climate surfaces  
616 for global land areas. *International Journal of Climatology*, 37, 4302-4315.

617 Frierson, D., Hwang, Y., Fučkar, N., Seager, R., Kang, S. M., Donohoe, A., Battisti, D. S.  
618 (2013). Contribution of ocean overturning circulation to tropical rainfall peak in the  
619 Northern Hemisphere. *Nature Geoscience*, 6, 940–944.

620 Fukami, T. (2015). Historical contingency in community assembly: integrating niches, species  
621 pools, and priority effects. *Annual Review of Ecology, Evolution, and Systematics*, 46, 1-  
622 23.

623 Gentry, A. H. (1982). Neotropical Floristic Diversity: Phytogeographical Connections Between  
624 Central and South America, Pleistocene Climatic Fluctuations, or an Accident of the  
625 Andean Orogeny? *Annals of the Missouri Botanic Garden*, 69, 557-593.

626 Gentry, A. H. (1988). Changes in plant community diversity and floristic composition on  
627 environmental and geographical gradients. *Annals of the Missouri Botanic Garden*, 75, 1-  
628 34.

629 Graham, A. (1999). The Tertiary History of the Northern Temperate Element in the Northern  
630 Latin American Biota. *American Journal of Botany*, 86, 32-38.

631 Grime, J. P., & Pierce, S. (2012). *The Evolutionary Strategies that Shape Ecosystems*. Wiley-  
632 Blackwell, Chichester, UK.

633 Hernández-Ruedas, M. A., Arroyo-Rodríguez, V., Meave, J. A., Martínez-Ramos, M., Ibarra-  
634 Manríquez, G., Martínez, E., ... Santos, B.A. (2014). Conserving tropical tree diversity and

635 forest structure: The value of small rainforest patches in moderately-managed landscapes.  
636 *PLoS ONE*, 9, e98931.

637 IPCC (2014). *Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and*  
638 *III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*  
639 [Core Writing Team, R.K. Pachauri and L.A. Meyer (eds.)]. IPCC, Geneva, Switzerland

640 Kattge, J., Bönisch, G., Díaz, S., Lavorel, S., Prentice, I. C., ... Wirth, C. (2020). TRY plant trait  
641 database – enhanced coverage and open access. *Global Change Biology*, 26, 119–188.

642 Kikuzawa, K., Onoda, Y., Wright, I. J., & Reich, P. B. (2013). Mechanism of leaf longevity  
643 patterns. *Global Ecology and Biogeography*, 22, 982-993.

644 Kitajima, K., & Poorter, L. (2010). Tissue-level leaf toughness, but not lamina thickness,  
645 predicts sapling leaf lifespan and shade tolerance of tropical tree species. *New Phytologist*,  
646 186, 708-721.

647 Kobe, R. K. (1999). Light gradient partitioning among tropical tree species through differential  
648 seedling mortality and growth. *Ecology*, 80, 187-201.

649 Kraft, N. J. B., & Ackerly, D. D. (2014). Assembly of Plant Communities. In: Monson R.  
650 (eds) *Ecology and the Environment*. The Plant Sciences, vol 8. Springer, New York.

651 Laliberté, E., & Legendre, P. (2010). A distance-based framework for measuring functional  
652 diversity from multiple traits. *Ecology*, 91, 2099–3005.

653 Laurance, W. F., Nascimento, H. E. M., Laurance, S. G., Andrade, A. C., Fearnside, P. M.,  
654 Ribeiro, J. E. L., & Capretz, R. L. (2006). Rain forest fragmentation and the proliferation  
655 of successional trees. *Ecology*, 87, 469–482.

656 Lefcheck, J. S. (2016). piecewiseSEM: Piecewise structural equation modeling in R for ecology,  
657 evolution, and systematics. *Methods in Ecology and Evolution*, 7, 573-579.

658 Leishman, M. R., & Westoby, M. (1994). The role of large seed size in shaded conditions –  
659 Experimental evidence. *Functional Ecology*, 8, 205–214.

660 Leite, Y. L. R., Costa, L., Loss, A. C., Rocha, R. G., Batalha-Filho, H., Bastos, A., & Pardini, R.  
661 (2016). Neotropical forest expansion during the last glacial period challenges refuge  
662 hypothesis. *Proceedings of the National Academy of Sciences of the United States of*  
663 *America*, 113, 1008-1013.

664 Maitner, B. S., Boyle, B., Casler, N., Condit, R., Donoghue II, J., Durán, S. M., ... Enquist, B. J.  
665 (2018). The BIEN R package: A tool to access the Botanical Information and Ecology  
666 Network (BIEN) Database. *Methods in Ecology and Evolution*, 9, 373–379.

667 Malhado, A., Oliveira-Neto, J. A., Stropp, J., Strona, G., Dias, L. C., Pinto, L. B., & Ladle, R. J.  
668 (2015). Climatological correlates of seed size in Amazonian forest trees. *Journal of*  
669 *Vegetation Science*, 26, 956–963.

670 Mittelbach, G. G., & Schemske, D. W. (2015). Ecological and evolutionary perspectives on  
671 community assembly. *Trends in Ecology and Evolution*, 30, 241 – 247.

672 Moles, A. T. (2018). Being John Harper: Using evolutionary ideas to improve understanding of  
673 global patterns in plant traits. *Journal of Ecology*, 106, 1-18.

674 Moles, A. T., Ackerly, D. D., Tweddle, J. C., Dickie, J. B., Smith, R., Leishman, M. R., ...  
675 Westoby, M. (2007). Global patterns in seed size. *Global Ecology and Biogeography*, 16,  
676 109-116.

677 Moles, A. T., Perkins, S. E., Laffan, S. W., Flores-Moreno, H., Awasthy, M., Tindall, M. L., ...  
678 Bonser, S.P. (2014). Which is a better predictor of plant traits: temperature or  
679 precipitation? *Journal of Vegetation Science*, 25, 1167–1180.

680 Moles, A. T., Warton, D. I., Warman, L., Swenson, N. G., Laffan, S. W., Zanne, A. E., ...  
681 Leishman, M.R. (2009). Global patterns in plant height. *Journal of Ecology*, 97, 923–932.  
682 Mora, C., Frazier, A., Longman, R., Dacks, R. S., Walton, M. M., Tong, E. J., ... Giambelluca,  
683 T. W. (2013). The projected timing of climate departure from recent variability. *Nature*,  
684 502, 183-187.

685 Morueta-Holme, N., Enquist, B. J., McGill, B. J., Boyle, B., Jørgensen, P. M., Ott, J.E., ...  
686 Svenning, J-C. (2013). Habitat area and climate stability determine geographical variation  
687 in plant species range sizes. *Ecology Letters*, 16, 1446–1454.

688 Muller-Landau, H. C. (2010). The tolerance-fecundity trade-off and the maintenance of diversity  
689 in seed size. *Proceedings of the National Academy of Sciences of the United States of*  
690 *America*, 107, 4242-4247.

691 Muscarella, R., Lohbeck, M., Martínez-Ramos, M., Poorter, L., Rodríguez-Velázquez, J. E., van  
692 Breugel, M., & Bongers, F. (2017). Demographic drivers of functional composition dy-  
693 namics. *Ecology*, 98, 2743-2750.

694 Nakagawa, S., & Schielzeth, H. (2013). A general and simple method for obtaining R<sup>2</sup> from  
695 generalized linear mixed-effects models. *Methods in Ecology and Evolution*, 4, 133–142.

696 Neter, J., Kutner, M.H., Nachtsheim, C.J., & Wasserman, W. (1996). Applied linear statistical  
697 models. – McGraw- Hill/Irwin.

698 Oliveira-Filho, A. T., Budke, J. C., Jarenkow, J. A., Eisenlohr, P. V., Neves, D. R. M. (2013).  
699 Delving into the variations in tree species composition and richness across South American  
700 subtropical Atlantic and Pampean forests. *Journal of Plant Ecology*, 8, 242–260.

701 Orihuela, R. L. Z., Peres, C. A., Mendes, G., Jarencow, J. A., & Tabarelli, M. (2015). Markedly  
702 divergent tree assemblage responses to tropical forest loss and fragmentation across a  
703 strong seasonality gradient. *PLoS One*, 10, e0136018

704 Peres-Neto, P. R., Dray, S., & ter Braak, C. J. F. (2017). Linking trait variation to the  
705 environment: critical issues with community-weighted mean correlation resolved by the  
706 fourth-corner approach. *Ecography*, 40, 806-816.

707 Pérez-Harguindeguy, N., Díaz, S., Garnier, E., Lavorel, S., Poorter, H., Jaureguiberry, P., ...  
708 Cornelissen, J. H. C. (2013). New handbook for standardised measurement of plant  
709 functional traits worldwide. *Australian Journal of Botany*, 61, 167–234.

710 Pierce, S., Negreiros, D., Cerabolini, B. E. L., Kattge, J., Díaz, S., Kleyer, M., ... Tampucci, D.  
711 (2017). A global method for calculating plant CSR ecological strategies applied across  
712 biomes world-wide. *Functional Ecology*, 31, 444–457.

713 Pinho, B. X., de Melo, F. P. L., Arroyo-Rodríguez, V., Pierce, S., Lohbeck, M., & Tabarelli, M.  
714 (2018). Soil-mediated filtering organizes tree assemblages in regenerating tropical forests.  
715 *Journal of Ecology*, 106, 137–147.

716 Pitman, N. C. A., Terborgh, J. W., Silman, M. R., Núñez, V. P., Neill, D. A., Cerón, C. E.,  
717 Palacios, W.A., & Aulestia, M. (2001). Dominance and distribution of tree species in upper  
718 Amazonian terra firme forests. *Ecology*, 82, 2101-2117.

719 Poorter, H., Niinemets, Ü., Poorter, L., Wright, I. J., & Villar, R. (2009). Causes and  
720 consequences of variation in leaf mass per area (LMA): a meta-analysis. *New Phytologist*,  
721 182, 565–588.

722 Poorter, L., van der Sande, M. T., Arets, E. J. M. M., Ascarrunz, N., Enquist, B., Finegan, B., ...  
723 Peña-Claros, M. (2017). Biodiversity and climate determine the functioning of Neotropical



724 forests. *Global Ecology and Biogeography*, 26, 1423–1434.

725 Reich, P. B. (2014). The world-wide ‘fast-slow’ plant economics spectrum: a traits manifesto.  
726 *Journal of Ecology*, 102, 275–301.

727 Royal Botanic Gardens Kew (2020). Seed Information Database (SID). Version 7.1. Available  
728 from: <http://data.kew.org/sid/>

729 Santiago, L. S., De Guzman, M. E., Baraloto, C., Vogenberg, J. E., Brodie, M., Hérault, B.,  
730 Fortunel, C., & Bonal, D. (2018). Coordination and trade-offs among hydraulic safety,  
731 efficiency and drought avoidance traits in Amazonian rainforest canopy tree species. *New*  
732 *Phytologist*, 218, 1015-1024.

733 Sandel, B., Weigelt, P., Kreft, H., Keppel, G., van der Sande, M. T., Levin, S., Smith, S., Craven,  
734 D., & Knight, T. M. (2020). Current climate, isolation and history drive global patterns of  
735 tree phylogenetic endemism. *Global Ecology and Biogeography*; 29, 4-15.

736 Santos, B. A., Peres, C. A., Oliveira, M. A., Grillo, A., Alves-Costa, C. P., & Tabarelli, M.  
737 (2008). Drastic erosion in functional attributes of tree assemblages in Atlantic forest  
738 fragments of northeastern Brazil. *Biological Conservation*, 141, 249–260.

739 Siefert, A., Violle, C., Chalmandrier, L., Albert, C. H., Taudiere, A., Fajardo, A., ... Wardle, D.  
740 A. (2015). A global meta-analysis of the relative extent of intraspecific trait variation in  
741 plant communities. *Ecology Letters*, 18, 1406–1419.

742 Šímová, I, Violle, C, Svenning, J-C, Kattge, J., Engemann, K., Sandel, B., ... Enquist, B.J.  
743 (2018). Spatial patterns and climate relationships of major plant traits in the New World  
744 differ between woody and herbaceous species. *Journal of Biogeography*, 45, 895-916.

745 Swenson, N. G., & Enquist, B. J. (2007). Ecological and evolutionary determinants of a key  
746 plant functional trait: Wood density and its community-wide variation across latitude and  
747 elevation. *American Journal of Botany*, 94, 451–459.

748 Swenson, N. G., Enquist, B. J., Pither, J., Kerkhoff, A. J., Boyle, B., Weiser, M. D., ... Nolting,  
749 K. (2012). The biogeography and filtering of woody plant functional diversity in North and  
750 South America. *Global Ecology and Biogeography*, 21, 798–808.

751 ter Braak, C. J. F. (2017). Fourth-corner correlation is a score test statistic in a log-linear trait–  
752 environment model that is useful in permutation testing. *Environmental and Ecological*  
753 *Statistics*, 24, 219–242.

754 ter Braak, C. J. F., & Šmilauer, P. (2018). *Canoco reference manual and user's guide: software*  
755 *for ordination* (version 5.10). Microcomputer Power, Ithaca, USA, 536 pp.

756 ter Braak, C. J. F., Šmilauer, P., & Dray, S. (2018). Algorithms and biplots for double  
757 constrained correspondence analysis. *Environmental and Ecological Statistics*, 25, 171-  
758 197.

759 Thompson, K., Band, S. R., & Hodgson, J. G. (1993). Seed size and shape predict persistence in  
760 the soil. *Functional Ecology*, 7, 236-241.

761 Trabucco, A., & Zomer, R. J. (2018). Global Aridity Index and Potential Evapo-Transpiration  
762 (ET0) Climate Database v2. CGIAR Consortium for Spatial Information (CGIAR-CSI).  
763 Published online, available from the CGIAR-CSI GeoPortal at <https://cgiarcsi.community>

764 van Buuren, S., & Groothuis-Oudshoorn, K. (2011). mice: Multivariate Imputation by Chained  
765 Equations in R. *Journal of Statistical Software*, 45, 1-67.

766 van der Sande, M., Arets, E., Peña-Claros, M., De Avila, A., Roopsind, A., Mazzei, L., ... Poorter, L.  
767 (2016). Old-growth Neotropical forests are shifting in species and trait composition. *Ecological*  
768 *Monographs*, 86, 228-243.

769 Venables, W. N., & Ripley, B. D. (2002). *Modern Applied Statistics with S*, Springer-Verlag.

770 Violle, C., Reich, P. B., Pacala, S. W., Enquist, B. J., & Kattge, J. (2014). The emergence and  
771 promise of functional biogeography. *Proceedings of the National Academy of Sciences of*  
772 *the United States of America*, 111, 13690–13696.

773 Westoby, M., Falster, D. S., Moles, A. T., Vesk, P. A., & Wright, I. J. (2002). Plant Ecological  
774 Strategies: Some Leading Dimensions of Variation Between Species. *Annual Review of*  
775 *Ecology and Systematics*, 33, 125–159.

776 Wright, I. J., Dong, N., Maire, V., Prentice, I. C., Westoby, M., Díaz, S., ... Wilf, P. (2017).  
777 Global climatic drivers of leaf size. *Science*, 357, 917–921.

778 Wright, S. J., Muller-Landau, H. C., & Schipper, J. (2009). The Future of Tropical Species on a  
779 Warmer Planet. *Conservation Biology*, 23, 1418-1426.

780 Wright, I. J., Reich, P. B., Westoby, M., Ackerly, D. D., Baruch, Z., Bongers, F., ..., Villar, R.  
781 (2004). The world-wide leaf economics spectrum. *Nature*, 428, 821–827.

782 Zhang, S. B., Cao, K. F., Fan, Z. X., & Zhang, J. L. (2013). Potential hydraulic efficiency in  
783 angiosperm trees increases with growth-site temperature but has no trade-off with  
784 mechanical strength. *Global Ecology and Biogeography*, 22, 971–981.

785 **Table 1.** Results of averaging of the best-fitted mixed-effects models ( $\Delta AICc < 2$ ) analysing the effects of climatic variables on the  
786 abundance-weighted community mean of (a) functional traits and (b) functional strategies, across 471 forest plots in nine  
787 biogeographic regions distributed throughout the Neotropics (see Fig. 1). For each variable retained in a best-fit model, we indicated  
788 the mean coefficient ( $\beta$ ), the standard error (SE), the 95% confidence intervals (95% CI) and the *p*-value. *P*-values of significant  
789 variables (according to 95% confidence intervals) are in bold. The predictors were standardized and thus the coefficients indicate their  
790 relative contribution for each response variable. The marginal  $R^2$  (variance explained by climatic factors) and conditional  $R^2$  (the  
791 former plus additional among-regions variance not explained by climatic factors) values of the full model are also shown. Units of  
792 climatic variables: MAP (mm), PET (mm), MAT ( $^{\circ}C$ ), PS (coefficient of variation of monthly values), TS (standard deviation of  
793 monthly values multiplied by 100).

Model factors	$\beta$	SE	95% CI		p-value	Models R <sup>2</sup> (marginal/conditional)
			Lower	Upper		
<b>(a) Functional traits</b>						
<b>Log (leaf area [cm<sup>2</sup>])</b>						0.24/0.29
Annual precipitation (MAP)	0.096	0.017	0.062	0.130	<b>&lt;2e-16</b>	
Potential evapotranspiration (PET)	-0.124	0.024	-0.171	-0.076	<b>4.00E-07</b>	
Annual temperature (MAT)	0.059	0.026	0.008	0.109	<b>0.023</b>	

Precipitation seasonality (PS)	0.043	0.028	0.015	0.092	0.126	
Temperature seasonality (TS)	0.024	0.031	-0.016	0.102	0.443	
<b>Log (specific leaf area [cm<sup>2</sup>/g])</b>						0.09/0.32
Annual precipitation (MAP)	0.010	0.007	0.000	0.024	0.160	
Annual temperature (MAT)	0.010	0.007	-0.001	0.025	0.155	
Precipitation seasonality (PS)	0.010	0.010	0.000	0.030	0.284	
Potential evapotranspiration (PET)	-0.001	0.005	-0.029	0.011	0.795	
Temperature seasonality (TS)	-0.003	0.007	-0.030	0.009	0.695	
<b>Leaf dry matter content (g/g)</b>						0.47/0.70
Temperature seasonality (TS)	-0.025	0.003	-0.031	-0.020	<b>&lt;2e-16</b>	
Annual precipitation (MAP)	-0.003	0.002	-0.006	0.000	<b>0.026</b>	
Annual temperature (MAT)	-0.003	0.003	-0.008	0.000	0.224	
Potential evapotranspiration (PET)	-0.002	0.003	-0.010	0.003	0.576	
Precipitation seasonality (PS)	0.0001	0.001	-0.005	0.004	0.894	
<b>Log (seed mass [mg]+1)</b>						0.57/0.68
Temperature seasonality (TS)	-0.653	0.067	-0.783	-0.522	<b>&lt;2e-16</b>	

Annual temperature (MAT)	-0.161	0.056	-0.270	-0.051	<b>0.003</b>
Precipitation seasonality (PS)	-0.107	0.047	-0.188	-0.016	<b>0.022</b>
Potential evapotranspiration (PET)	-0.032	0.058	-0.220	0.035	0.577
Annual precipitation (MAP)	-0.003	0.017	-0.088	0.062	0.869

**Wood density (g/cm<sup>3</sup>)**

0.39/0.62

Temperature seasonality (TS)	-0.052	0.006	-0.064	-0.040	<b>&lt;2e-16</b>
Annual precipitation (MAP)	-0.018	0.003	-0.024	-0.011	<b>3.00E-07</b>
Annual temperature (MAT)	-0.016	0.005	-0.026	-0.005	<b>0.003</b>
Potential evapotranspiration (PET)	0.009	0.007	0.000	0.024	0.207
Precipitation seasonality (PS)	-0.003	0.005	-0.016	0.001	0.523

**Sqrt (maximum height [m])**

0.16/0.41

Temperature seasonality (TS)	-0.190	0.034	-0.256	-0.125	<b>&lt;2e-16</b>
Annual temperature (MAT)	-0.173	0.033	-0.239	-0.108	<b>2.00E-07</b>
Potential evapotranspiration (PET)	0.057	0.045	0.005	0.145	0.204
Precipitation seasonality (PS)	-0.004	0.014	-0.065	0.030	0.767

**(b) Functional strategies**

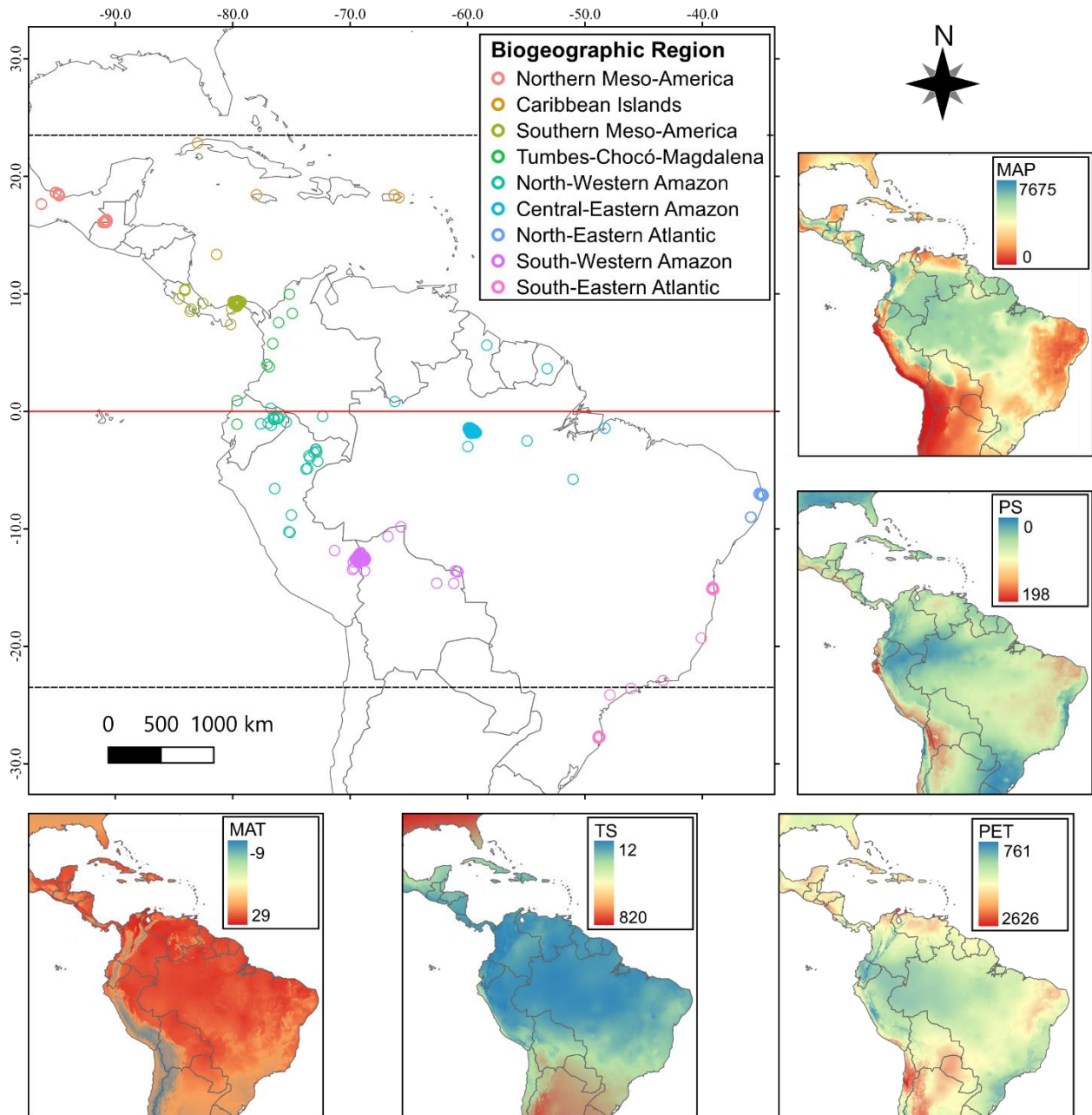
<b>PC1 (economics spectrum)</b>						0.40/0.69
Annual precipitation (MAP)	-0.096	0.032	0.034	0.158	<b>0.002</b>	
Annual temperature (MAT)	-0.159	0.043	0.075	0.243	<b>0.0002</b>	
Temperature seasonality (TS)	-0.560	0.057	0.448	0.673	<b>&lt;2e-16</b>	
Precipitation seasonality (PS)	-0.017	0.035	-0.033	0.140	0.634	
Potential evapotranspiration (PET)	0.013	0.037	-0.174	0.068	0.739	
<b>PC2 (size spectrum)</b>						0.28/0.28
Annual precipitation (MAP)	0.120	0.013	-0.145	-0.094	<b>&lt;2e-16</b>	
Potential evapotranspiration (PET)	-0.125	0.017	0.092	0.158	<b>&lt;2e-16</b>	
Precipitation seasonality (PS)	0.060	0.015	-0.090	-0.031	<b>7.15E-05</b>	
Temperature seasonality (TS)	-0.024	0.021	0.001	0.068	0.255	
Annual temperature (MAT)	0.022	0.023	-0.073	0.002	0.325	

795 **Table 2.** Variation partitioning of the trait-structured variation in the dc-CA with all traits, showing the unique and shared effects of  
796 geography (latitude, longitude, and degrees from equator) and climate (MAP, MAT, TS, PS, PET) in aggregated samples of tree  
797 communities across Neotropical moist forests (N = 59; see Fig. S5). The trait-structured variation is a weighted variance of the CWMs  
798 with respect to orthonormalized traits with the sample total as weight.

<b>Component</b>	<b>Variation (Adj R2)</b>	<b>% of Explained</b>	<b>DF</b>	<b>Mean Square</b>	<b>F</b>	<b>P</b>
Climate (unique)	0.15	22.7	5	0.03	5.9	0.0005
Geography (unique)	0.12	17.8	3	0.04	7.2	0.0045
Shared	0.39	59.5	--	--		--
Total Explained	0.66	100	8	0.09	15.3	0.0005

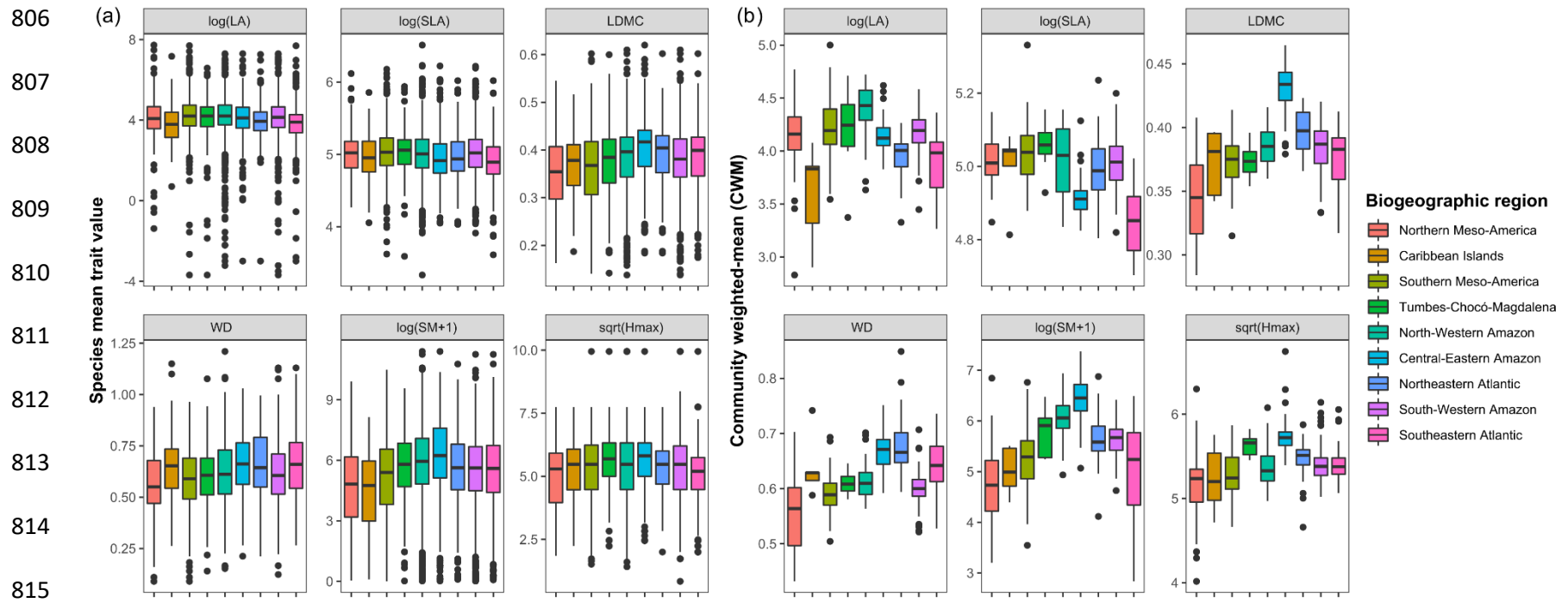
799



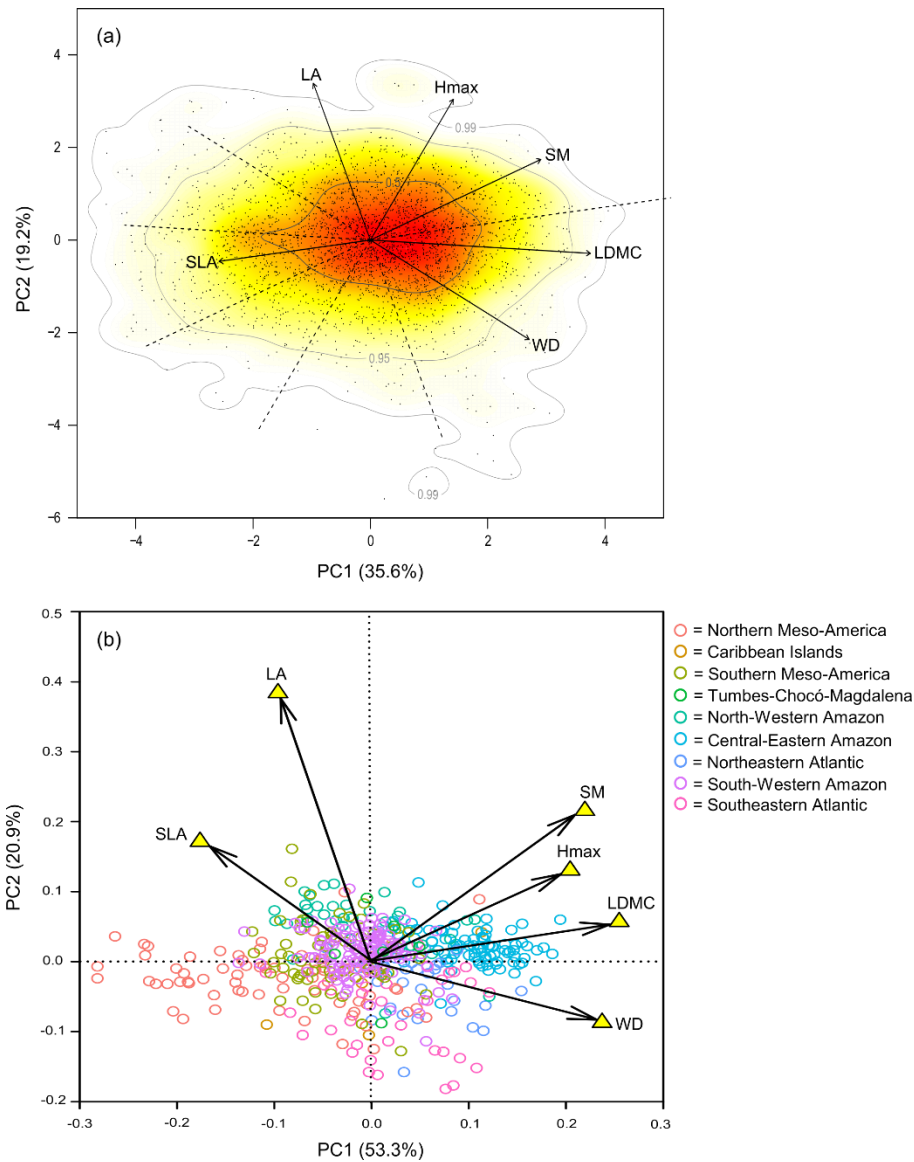


800

801 **Figure 1.** Location of the 471 forest plots studied in nine biogeographic regions, and variation in  
 802 the five climatic variables considered throughout the Neotropics. MAP, mean annual  
 803 precipitation (mm); MAT, mean annual temperature (°C); PS, precipitation seasonality  
 804 (coefficient of variation of monthly precipitation values); TS, temperature seasonality (standard  
 805 deviation of monthly temperature values); PET, potential evapotranspiration (mm).

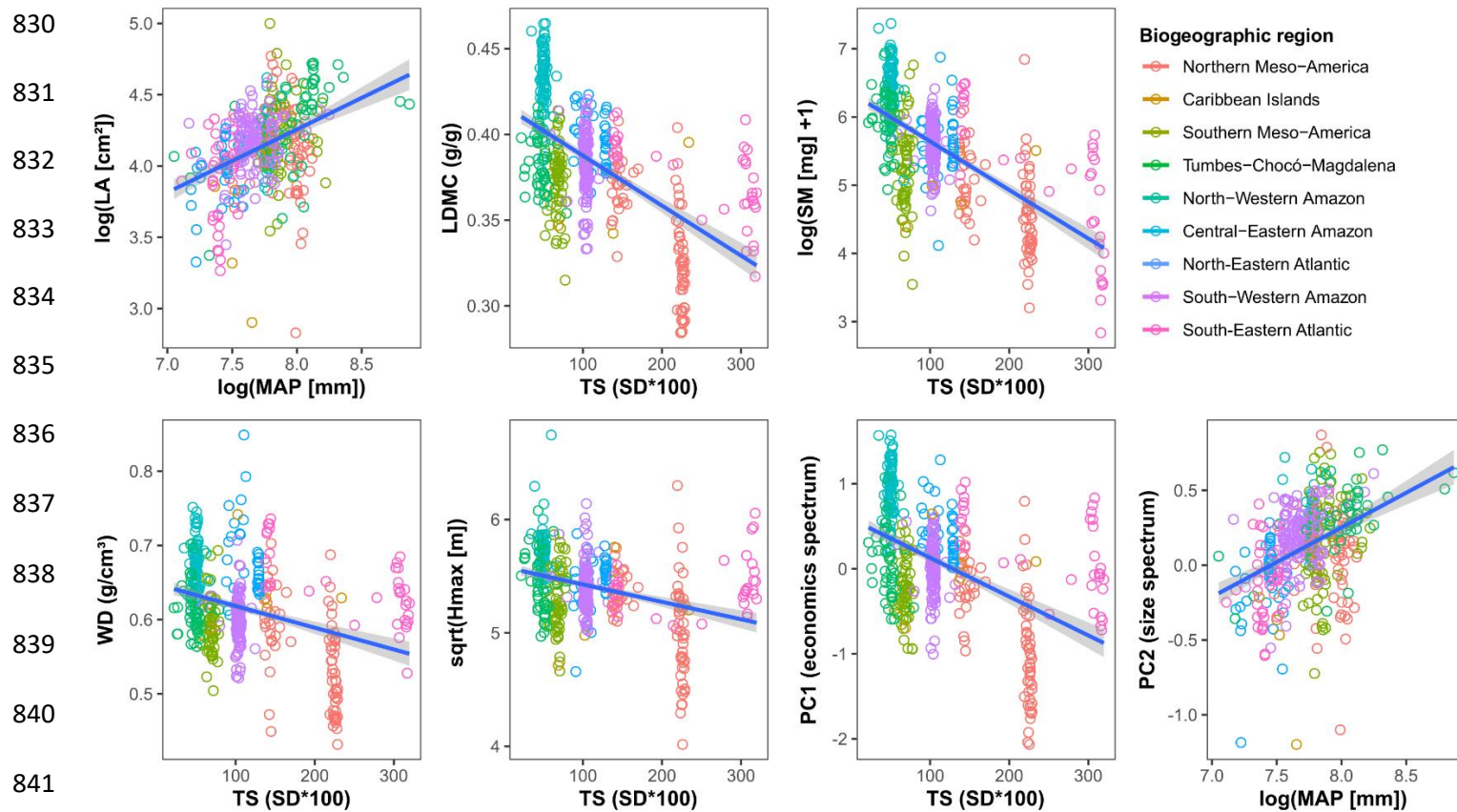


816 **Figure 2.** Differences between regions in functional trait values of (a) species present in each regional species pool (unweighted by  
 817 abundance), and (b) the communities (i.e., abundance-weighted means – CWM) in each region, for 3,417 tree species distributed in  
 818 471 forest plots across nine Neotropical moist forest regions. Boxplots indicate the median (center line), 25-75% quartiles (box edges),  
 819 < 1.5 times the inter quartile range (whiskers), and extreme values (dots). The boxplots are organized from the northernmost (left) to  
 820 the southernmost region (right) along the distribution of the Neotropical moist forest biome. LA, leaf area (cm<sup>2</sup>); SLA, specific leaf  
 821 area (cm<sup>2</sup>/g); LDMC, leaf dry matter content (g/g); WD, wood density (g/cm<sup>3</sup>); SM, seed mass (mg); Hmax, maximum height (m).

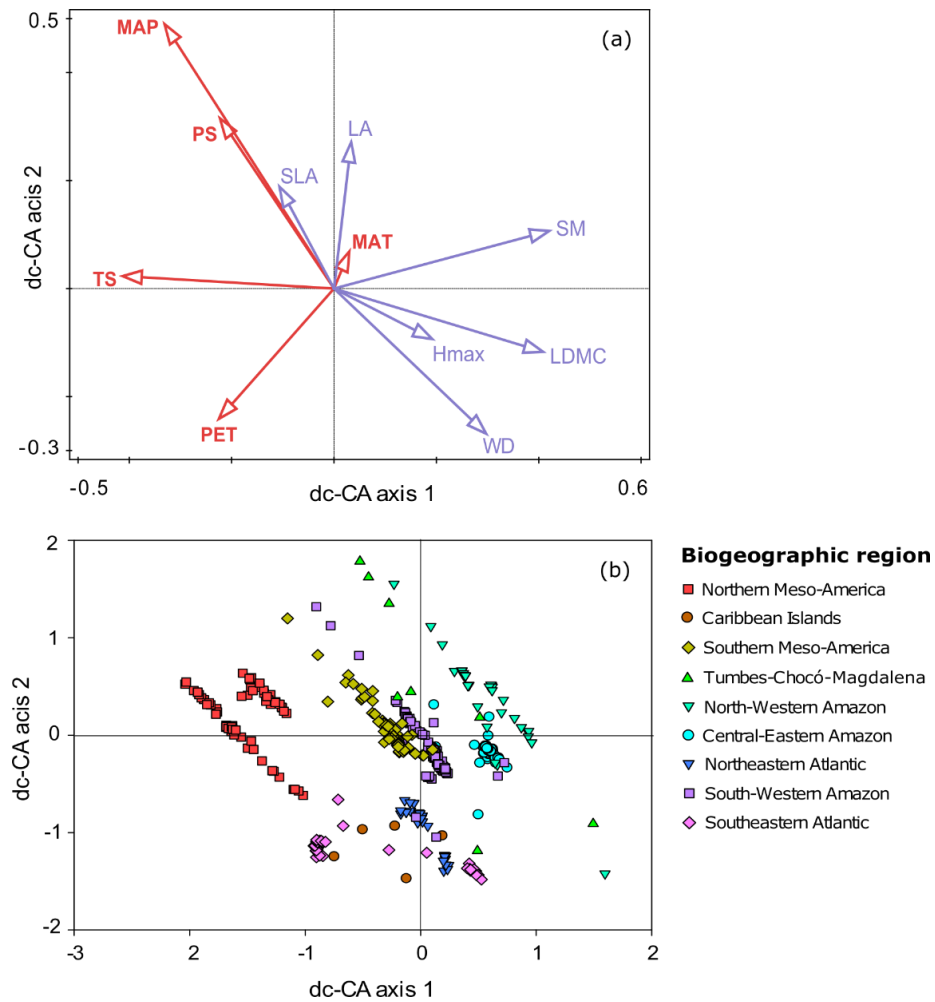


822

823 **Figure 3.** Ordination diagram of the first two axes of principal component analysis (PCA) of (a)  
 824 Neotropical tree species trait values ( $n = 3,417$  species); and (b) community-weighted mean trait  
 825 values of tree communities ( $n = 471$  plots) distributed across nine Neotropical moist forest re-  
 826 gions. The occurrence probability of species in the trait space is illustrated in (a) by color gradi-  
 827 ents from highest (red) to lowest (white) kernel density, with contour lines indicating 0.5, 0.95  
 828 and 0.99 quantiles. LA, leaf area; SLA, specific leaf area; LDMC, leaf dry matter content; WD,  
 829 wood density; SM, seed mass; Hmax, maximum height.



842 **Figure 4.** Significant relationships between climatic variables and community weighted-mean (CWM) of functional traits and  
 843 strategies (i.e., species scores on the two first axes of PCA on functional traits, see Fig. 3a), for 471 tree communities distributed  
 844 across nine Neotropical moist forest regions (see Table 1 for details on the best models). LA, leaf area; LDMC, leaf dry matter  
 845 content; WD, wood density; SM, seed mass; Hmax, maximum height. MAP, mean annual precipitation; TS, temperature seasonality;  
 846 SD, standard deviation. The relationship with highest slope (i.e., estimate) is shown for each trait/strategy, except for LA and PC2, for  
 847 which we selected the relationship with the lowest error and p-value from two relationships with similar slopes.



**Figure 5.** Ordination diagrams from double-constrained correspondence analysis (dc-CA) for 3,417 species across 471 Neotropical moist forest plots, showing (a) biplot of canonical weights of climate variables and scaled correlations of traits summarizing the coefficients of the multiple regressions of all CWMs of traits on the climate predictors; and (b) position (constrained scores) of samples (plots) in the dc-CA biplot. Graphs (a) and (b) form a biplot of the CWMs of all plots and traits. The significance of dc-CA results has been tested by aggregating community data by spatial clusters of plots ( $N = 59$ ; see Fig. S5) to avoid pseudo-replication (see Table 2 for related statistics). The position of the 30 species which contribute most to the first two dc-CA axes is showed in Fig. S8.

**Supporting information** – Pinho et al. *Global Ecology and Biogeography*

**Table S1.** Summary of sample sizes for each Neotropical moist forest region studied.

<b>Region</b>	<b>N plots</b>	<b>Sampled area (ha)</b>	<b>N species</b>	<b>N trees</b>
Northern Meso-America	78	7.8	329	5,174
Caribbean Islands	5	0.50	96	412
Southern Meso-America	62	48.24	890	38,050
Tumbes-Chocó-Magdalena	8	0.8	259	542
North-Western Amazon	37	27.52	1605	20,008
Central-Eastern Amazon	85	20.3	493	9,831
Northeastern Atlantic	38	3.17	198	2,775
South-Western Amazon	123	26.66	1029	16,378
Southeastern Atlantic	35	3.28	494	3,120
<b><i>Neotropics</i></b>	471	138.27	3,417	96,290

**Table S2.** Mean trait data coverage of total plot abundances across the study regions, represented by (a) species-level data, (b) genus average, and (c) imputation. For species-level trait data, we show in parenthesis the proportion covered by authors' data. The difference between the total proportion of species-level and of the authors' data represent the proportion covered by data compiled from global databases.

<b>Trait data</b>	N Meso- America	Caribbean Islands	S Meso- America	Tumbes- Chocó	NW- Amazon	CE- Amazon	NE- Atlantic	SW- Amazon	SE- Atlantic
<i>Species-level</i>									
(author's data)									
LA	0.94 (0.92)	0.72 (0.61)	0.83 (0.8)	0.51 (0.48)	0.73 (0.71)	0.65 (0.55)	0.78 (0.73)	0.84 (0.83)	0.74 (0.71)
SLA	0.94 (0.93)	0.55 (0.49)	0.8 (0.74)	0.5 (0.43)	0.69 (0.55)	0.66 (0.42)	0.76 (0.72)	0.78 (0.64)	0.72 (0.69)
LDMC	0.94 (0.93)	0.45 (0.18)	0.75 (0.66)	0.29 (0.47)	0.25 (0.41)	0.46 (0.11)	0.7 (0.7)	0.4 (0.35)	0.67 (0.68)
WD	0.88 (0.83)	0.79 (0.48)	0.85 (0.71)	0.52 (0.38)	0.67 (0.51)	0.8 (0.33)	0.84 (0.3)	0.78 (0.65)	0.79 (0.62)
SM	0.94 (0.93)	0.62 (0.51)	0.78 (0.7)	0.46 (0.41)	0.6 (0.56)	0.46 (0.25)	0.72 (0.35)	0.75 (0.7)	0.73 (0.62)
Hmax	0.87 (0.81)	0.51 (0.33)	0.83 (0.68)	0.46 (0.37)	0.66 (0.58)	0.66 (0.34)	0.89 (0.33)	0.8 (0.75)	0.86 (0.19)
<i>Genus average</i>									
LA	0.06	0.27	0.16	0.44	0.23	0.34	0.22	0.14	0.24
SLA	0.06	0.45	0.19	0.47	0.28	0.33	0.24	0.20	0.26
LDMC	0.05	0.42	0.23	0.64	0.66	0.50	0.27	0.55	0.29

WD	0.12	0.19	0.15	0.44	0.30	0.19	0.16	0.21	0.20
SM	0.06	0.32	0.20	0.48	0.34	0.53	0.28	0.23	0.24
Hmax	0.12	0.47	0.17	0.49	0.30	0.32	0.10	0.18	0.12
<i>Imputation</i>									
LA	0.00	0.01	0.01	0.05	0.04	0.01	0.00	0.02	0.02
SLA	0.00	0.00	0.01	0.03	0.03	0.01	0.00	0.02	0.02
LDMC	0.01	0.13	0.02	0.07	0.09	0.04	0.03	0.05	0.04
WD	0.00	0.02	0.00	0.04	0.03	0.01	0.00	0.01	0.01
SM	0.00	0.06	0.02	0.06	0.06	0.01	0.00	0.02	0.03
Hmax	0.01	0.02	0.00	0.05	0.04	0.02	0.01	0.02	0.02

---



**Table S3.** Correlation between species and genus mean in the trait database studied, considering genus with at least two species with available trait data (N). All correlations were significant after adjustment for multiple comparisons ( $p < 0.001$ ).

<b>Functional trait</b>	<b>Correlation (r-pearson) species-genus mean</b>
Leaf area (N = 1,718)	0.74
Specific leaf area (N = 1,555)	0.61
Leaf dry matter content (N = 845)	0.84
Wood density (N = 1,643)	0.86
Seed mass (N = 1,297)	0.73
Maximum height (N = 1,653)	0.66



**Table S5.** Trait loadings of principal component analysis (PCA) on Neotropical tree species mean traits (n = 3,417) and community-weighted mean traits (n = 471). Traits were standardized to Z units (mean = 0, SD = 1) after being transformed. The axis with highest loading for each trait is highlighted in bold.

Functional trait	Species-level			Community-level		
	PC1 (35.6%)	PC2 (19.1%)	PC3 (15.0%)	PC1 (53.3%)	PC2 (20.9%)	PC3 (10.8%)
Leaf area	-0.193	<b>0.672</b>	0.513	-0.191	<b>0.765</b>	-0.301
Specific leaf area	-0.391	-0.070	<b>-0.643</b>	-0.350	0.347	<b>0.771</b>
Leaf dry matter content	<b>0.564</b>	-0.043	0.009	<b>0.510</b>	0.113	0.136
Wood density	<b>0.449</b>	-0.356	0.108	<b>0.475</b>	-0.174	0.374
Seed mass	<b>0.466</b>	0.279	-0.125	<b>0.433</b>	0.428	0.226
Maximum height	0.270	<b>0.580</b>	-0.544	<b>0.409</b>	0.261	-0.325

1 **Table S6.** Statistics from double constrained correspondence analysis (dc-CA) of species  
 2 abundance matrix from 59 aggregated samples (clusters of 471 forest plots, connected by  
 3 maximum 50-km, see Fig. S5). Samples have been aggregated to spatial cluster means to avoid  
 4 pseudo-replication. E = environmental (climate) variables; T = functional traits; CWM =  
 5 community weighted mean traits; SNC = species niche centroids. Max-test consider the highest  
 6 p-value across species- and community-level tests.

<b>Statistic</b>	<b>Axis 1</b>	<b>Axis 2</b>	<b>Axis 3</b>
dc-CA eigenvalues	0.10	0.04	0.02
fourth-corner correlations (rFC)	0.31	0.20	0.12
% Explained fitted variation (cum.)	61	87	97
% CWM variation expl. by E (adj R2)	30	46	53
% SNC variation expl. by T (adj R2)	3	4	4
Max test (p-value)	0.001	0.004	0.152

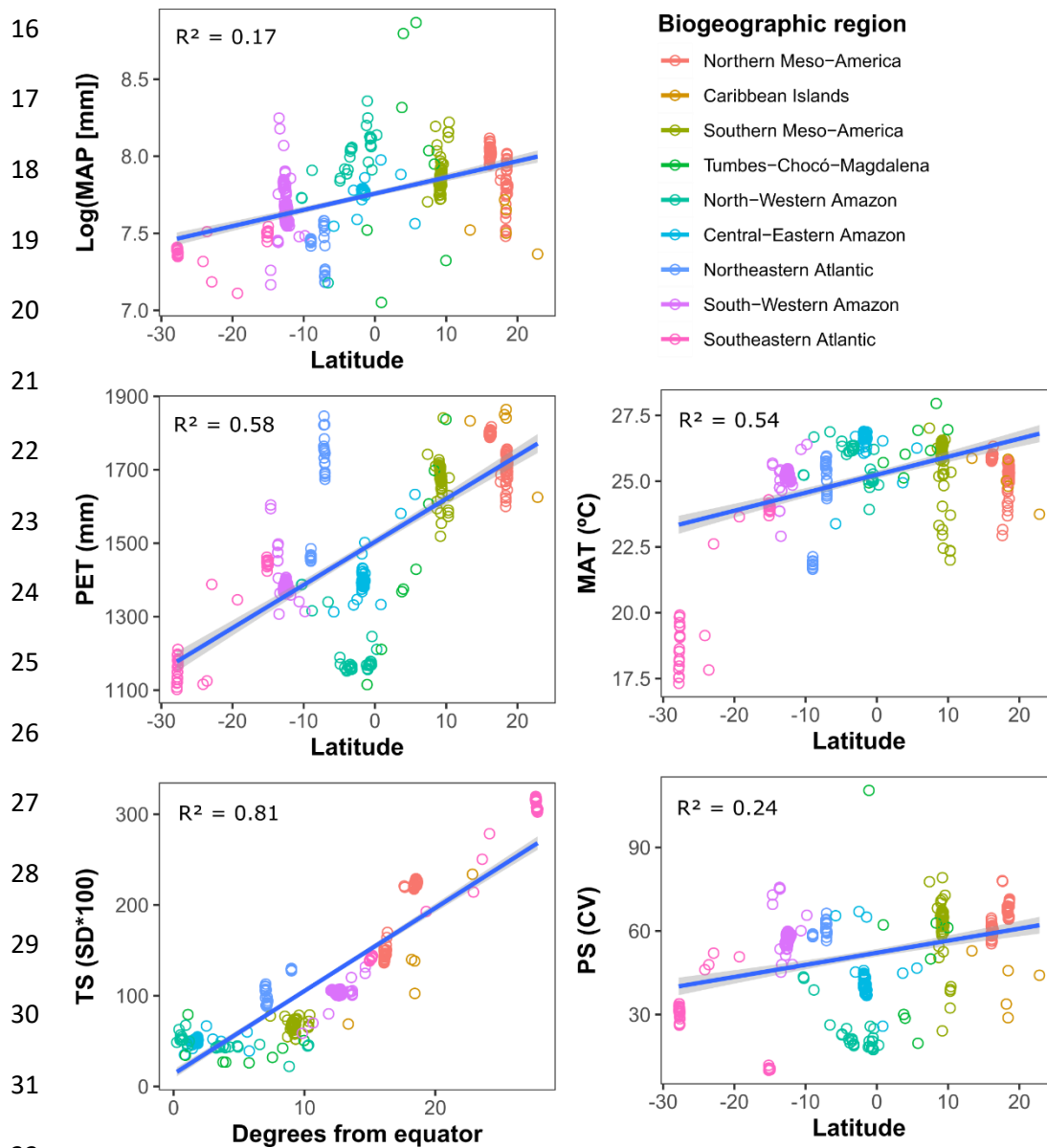
8 **Table S7.** Relative effects of geographical and climate variables (the three most important  
9 according to forward selection) in driving taxonomic (CCA) and functional (dc-CA) structure on  
10 the species abundance table by aggregated samples (i.e., plots connected by maximum 50-km,  
11 see Fig. S5). The ratio represents the fraction of the trait-structured environmental variation in  
12 relation to the environment-structured variation.

Species	Explained inertia of count (species)		
	unconstrained	Trait-constrained	Ratio
Source	CCA	dc-CA	(dc-CA:CCA)
<b><i>Geography</i></b>			
Latitude	0.8	0.03	0.04
Longitude	0.74	0.06	0.08
Degrees from equator	0.67	0.07	0.10
<b><i>Climate</i></b>			
Temperature seasonality	0.71	0.05	0.07
Potential evapotranspiration	0.77	0.05	0.06
Mean annual precipitation	0.54	0.03	0.06

14 **Table S8.** Names of tree species associated with the eight-letters codes showed in Figure S7.

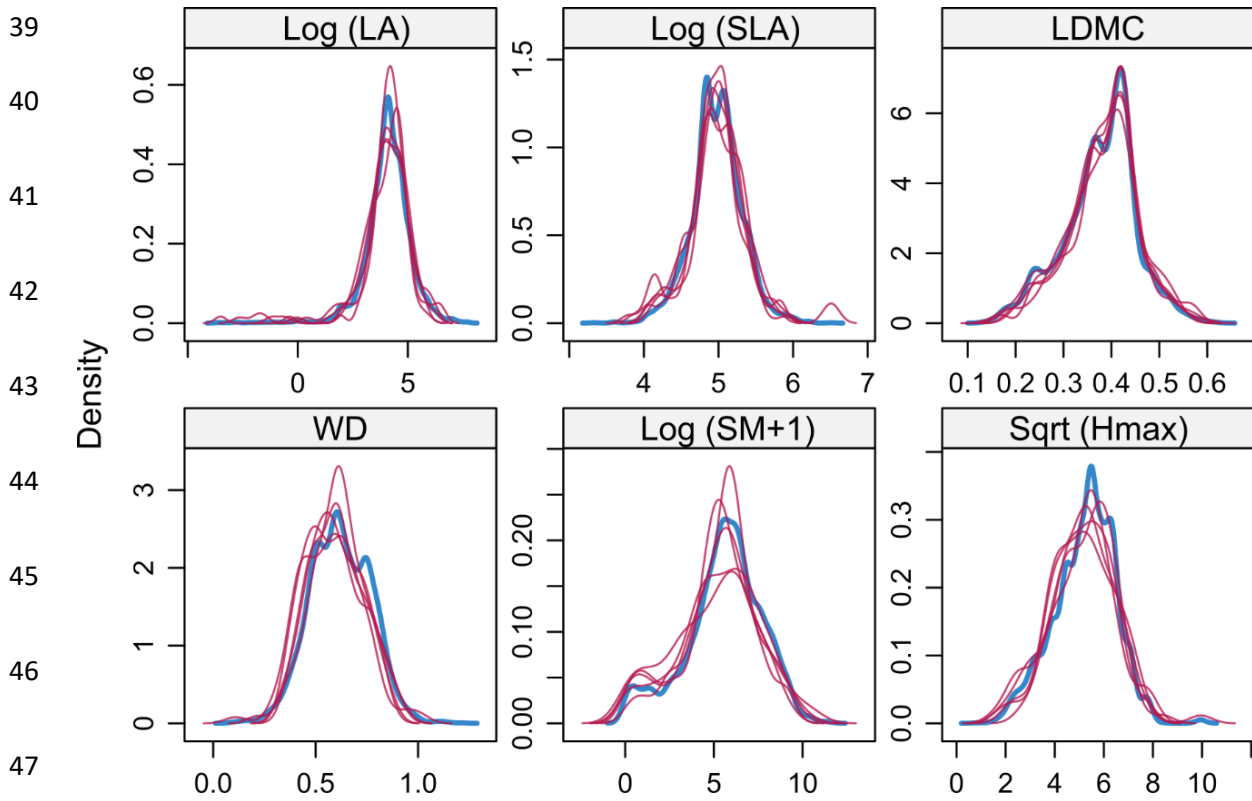
<b>Code</b>	<b>Species</b>
CrotSchi	<i>Croton schiedeanus</i>
TabeDonn	<i>Tabernaemontana donnell-smithii</i>
PseuGlab	<i>Pseudolmedia glabrata</i>
RollMuco	<i>Rollinia mucosa</i>
MabeOcci	<i>Mabea occidentalis</i>
PeraGlab	<i>Pera glabrata</i>
VochGuat	<i>Vochysia guatemalensis</i>
OrthObla	<i>Orthion oblanceolatum</i>
CymbBail	<i>Cymbopetalum baillonii</i>
HyerAlch	<i>Hyeronima alchorneoides</i>
ProtHebe	<i>Protium hebetatum</i>
EschOvat	<i>Eschweilera ovata</i>
SipaAndi	<i>Siparuna andina</i>
GuapOppo	<i>Guapira opposita</i>
MatiMala	<i>Matisia malacocalyx</i>
HampNutr	<i>Hampea nutricia</i>
SapiLate	<i>Sapium lateriflorum</i>
MicrElat	<i>Micrandra elata</i>
EschCori	<i>Eschweilera coriacea</i>
DendArbo	<i>Dendropanax arboreus</i>
CecrObtu	<i>Cecropia obtusifolia</i>

MicoCabu	<i>Miconia cabucu</i>
LunaMexi	<i>Lunania mexicana</i>
GuarGlab	<i>Guarea glabra</i>
PogoScho	<i>Pogonophora schomburgkiana</i>
PipeSanc	<i>Piper sanctum</i>
BursSima	<i>Bursera simaruba</i>
EschTrun	<i>Eschweilera truncata</i>
HirtHebe	<i>Hirtella hebeclada</i>
AspiAust	<i>Aspidosperma australe</i>

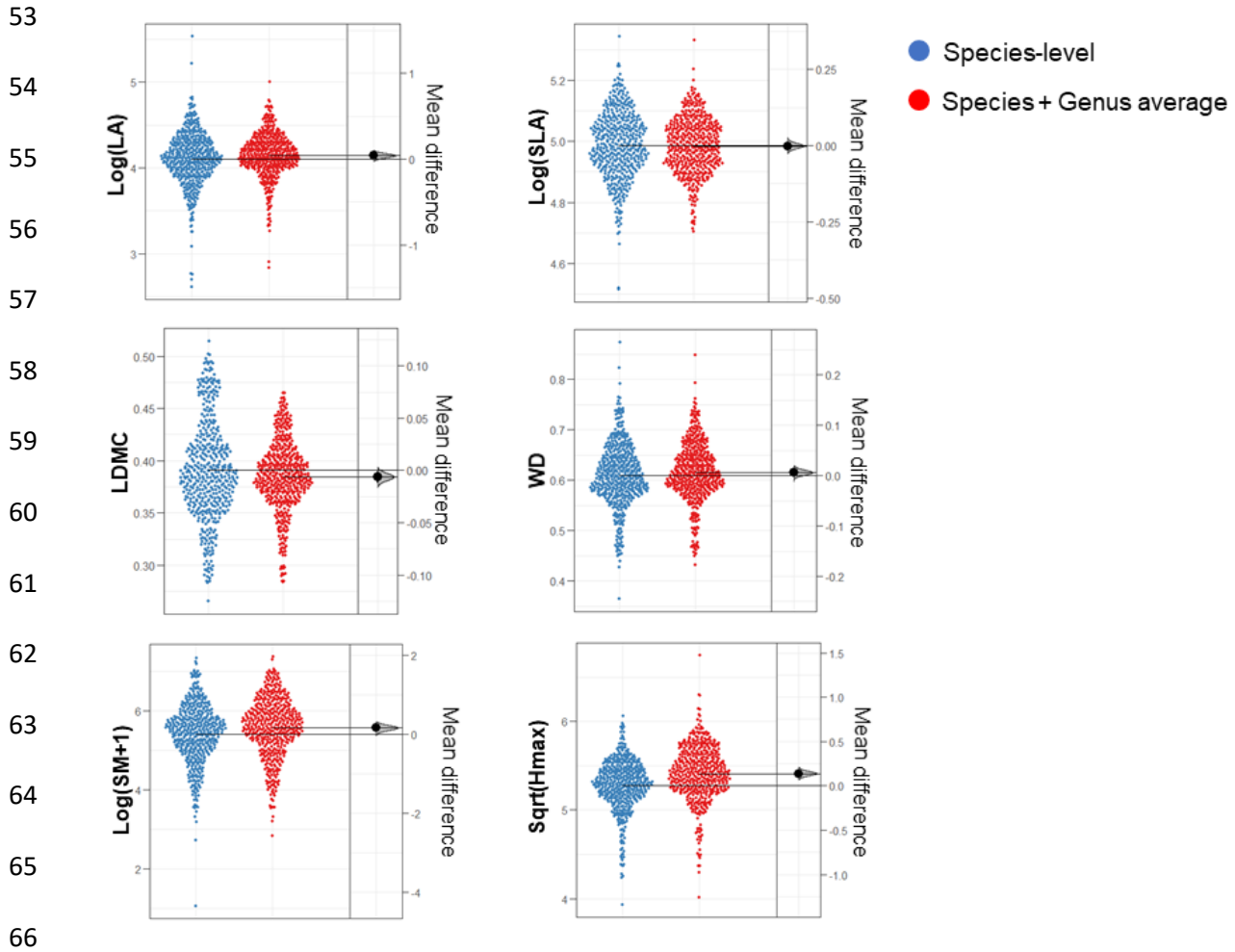


33 **Figure S1.** Relationships between climatic variables and latitude across 471 Neotropical moist  
 34 forest plots (see Fig. 1). The variance explained is the marginal R-square from mixed-effects  
 35 models with ‘biogeographic region’ as random factor. MAP, mean annual precipitation; MAT,  
 36 mean annual temperature; PET, potential evapotranspiration; PS, precipitation seasonality; TS,  
 37 temperature seasonality; SD, standard deviation of average monthly values; CV, coefficient of  
 38 variation of average monthly values.



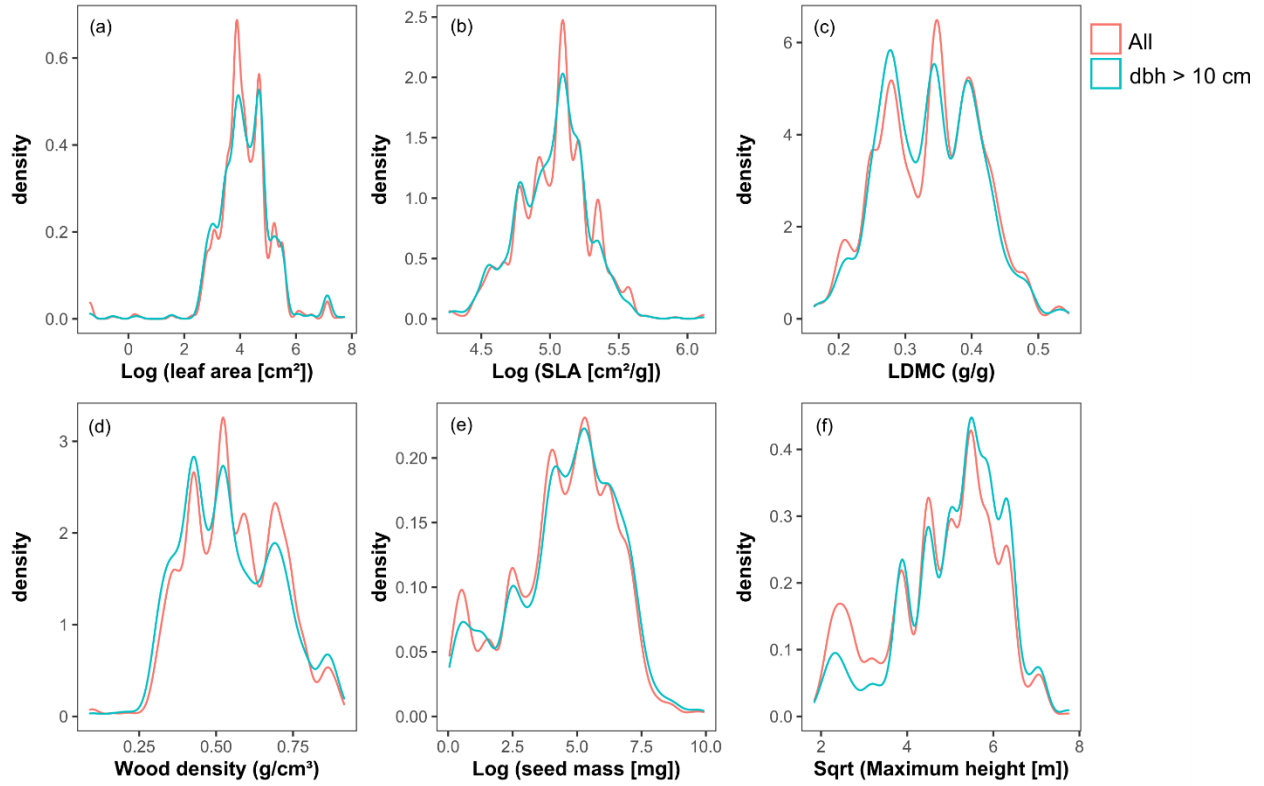


48 **Figure S2.** Distribution of species trait values for the original data (blue line) and five imputed  
 49 datasets (red). Trait imputation was performed through chained equations by predictive mean  
 50 matching, using R package ‘mice’. LA, leaf area (cm<sup>2</sup>); SLA, specific leaf area (cm<sup>2</sup>/g); LDMC,  
 51 leaf dry matter content (g/g); WD, wood density (g/cm<sup>3</sup>); SM, seed mass (mg); Hmax, maximum  
 52 height (m).



67 **Figure S3.** Distribution and mean differences ( $p > 0.05$  according to paired-tests) of CWM trait  
 68 values calculated from species-level data only (blue) and considering genus trait averages (red).  
 69 LA, leaf area ( $\text{cm}^2$ ); SLA, specific leaf area ( $\text{cm}^2/\text{g}$ ); LDMC, leaf dry matter content ( $\text{g}/\text{g}$ ); WD, wood  
 70 density ( $\text{g}/\text{cm}^3$ ); SM, seed mass (mg); Hmax, maximum height (m).

71



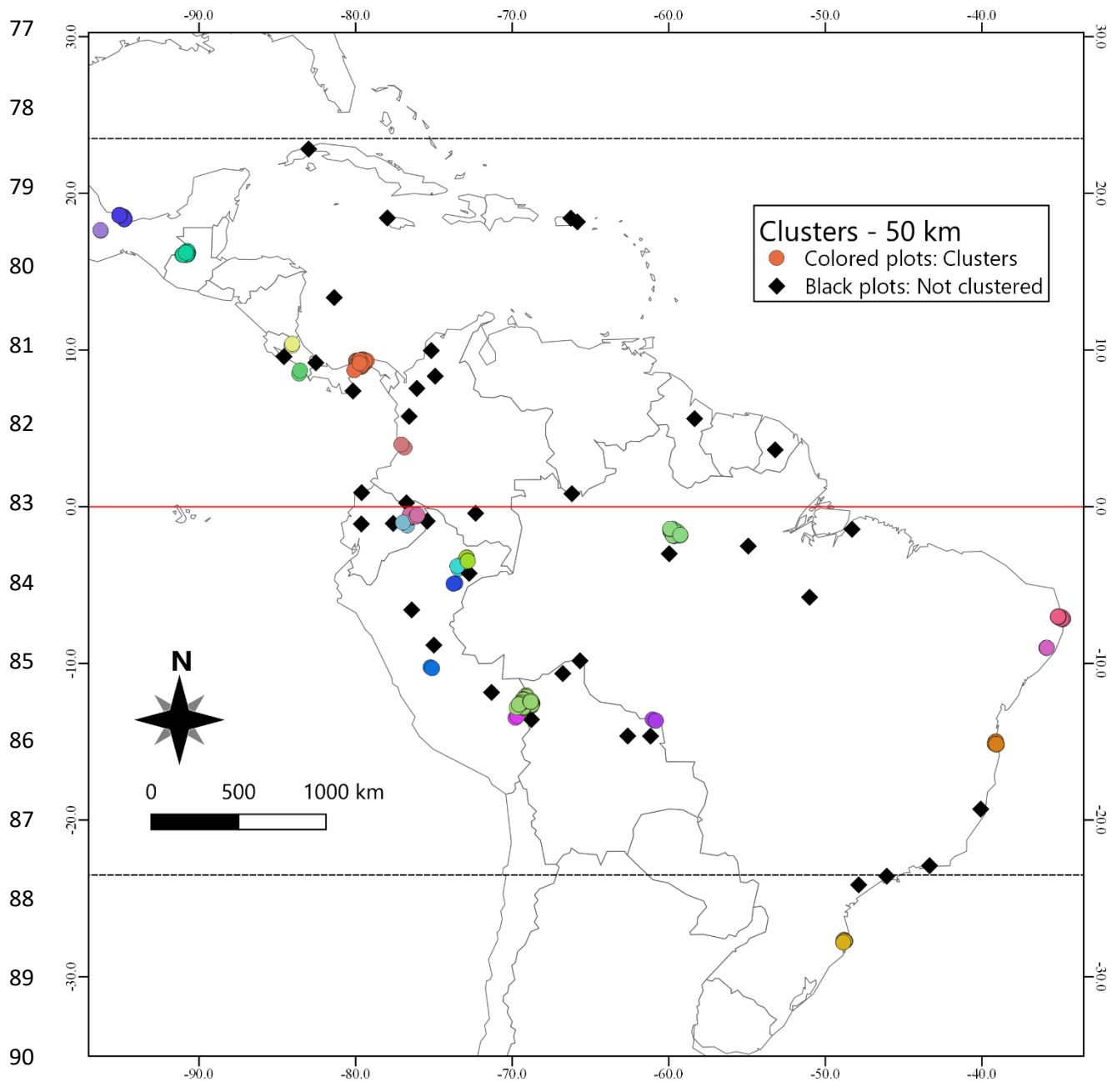
72

73 **Figure S4.** Abundance-weighted distributions of tree species trait values for all individuals with

74 diameter at breast height (DBH) from 2.5 cm (red), and for only adults with DBH > 10 cm

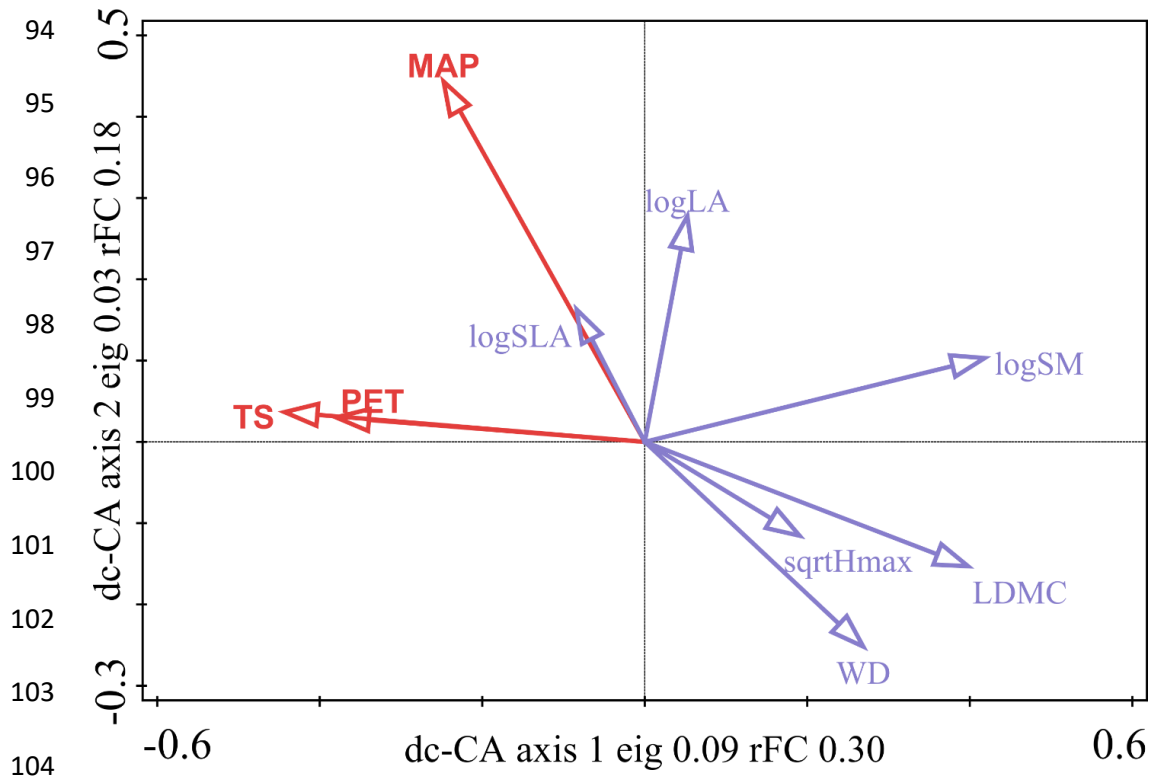
75 (blue), at the Northern Meso-America region.

76



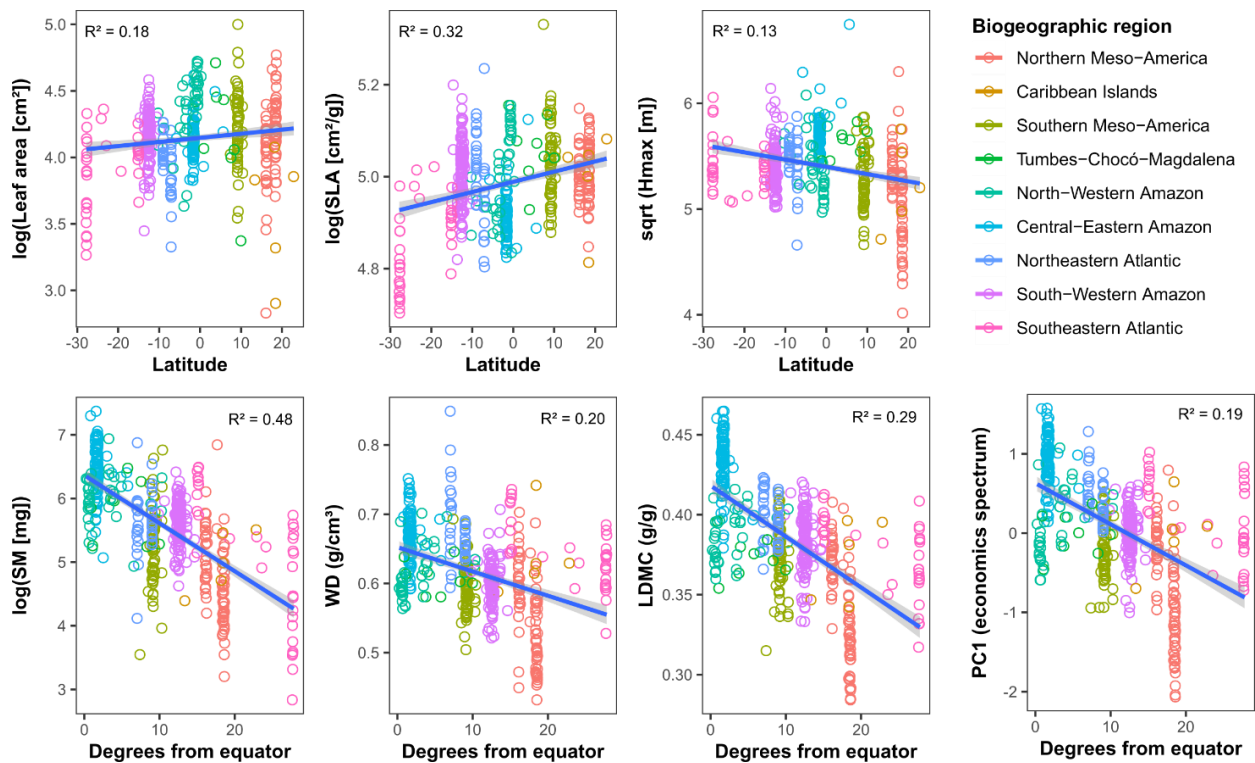
91 **Figure S5.** Spatial clusters (plots connected by maximum 50-km distance) that represented ag-  
 92 gregated samples (N = 59) in double constrained correspondence analysis (dc-CA).

93



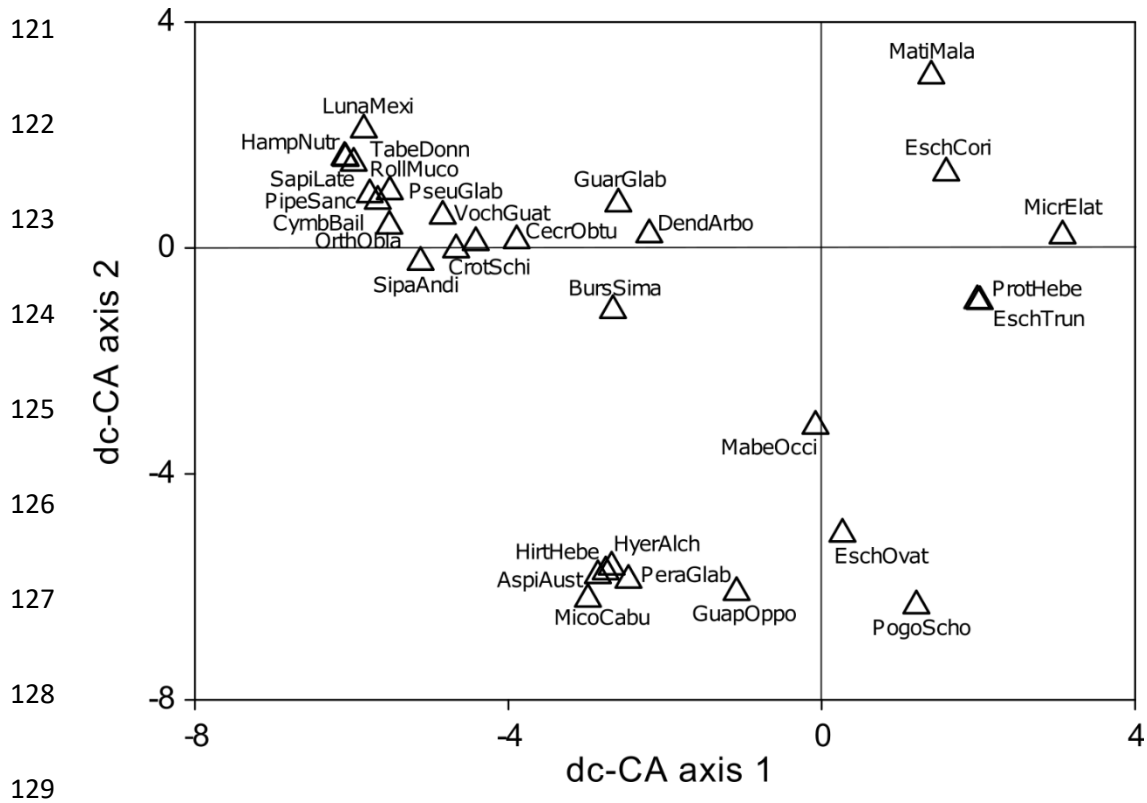
105 **Figure S6.** Biplot of the coefficients of the regressions of CWMs of functional traits on to the  
 106 climate variables (similar to Fig. 5a), with only the three most relevant climate variables,  
 107 according to forward selection in dc-CA analysis. ‘eig’ = eigenvalue; ‘rFC’ = fourth-corner  
 108 correlation. MAP, mean annual precipitation (mm); PET, potential evapotranspiration (mm); TS,  
 109 temperature seasonality (standard deviation of monthly values); LA, leaf area (cm<sup>2</sup>); SLA,  
 110 specific leaf area (cm<sup>2</sup>/g); LDMC, leaf dry matter content (g/g); WD, wood density (g/cm<sup>3</sup>); SM,  
 111 seed mass (mg); Hmax, maximum height.

112



113

114 **Figure S7.** Significant relationships from mixed-effects models with latitude (south-north gradi-  
 115 ent) and degrees from equator (the strongest for each trait/strategy) as predictor of abundance-  
 116 weighted community mean (CWM) of functional traits and strategies, for 471 tree communities  
 117 distributed across nine Neotropical moist forest regions. The variance explained is the marginal  
 118 R-square from mixed-effects models with 'biogeographic region' as random factor. LA, leaf  
 119 area; SLA, specific leaf area; LDMC, leaf dry matter content; WD, wood density; SM, seed  
 120 mass; Hmax, maximum height.



130 **Figure S8.** Position (unconstrained scores) on the dc-CA biplot (Fig. 5) of the 30 species which  
 131 contribute most to the first two dc-CA axes. The species names associated with the eight-letter  
 132 codes in are shown in Table S8.