

1 INFLUENCE OF BIOGAS DIGESTERS ON FAECAL INDICATOR ORGANISMS IN 2 DIGESTATE AND AROUND HOMESTEADS IN ETHIOPIA

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12 **Abstract:**

13 It is important, prior to application of organic wastes to land, that pathogen loads are reduced
14 sufficiently to minimize dissemination to the wider environment. Anaerobic digestion for biogas
15 production is a low-cost method to reduce pathogens in agricultural wastes that provides the added
16 benefits of energy generation. There have been claims of pathogen reduction following installation
17 of biogas digesters in homesteads in Sub-Saharan Africa. Homestead pathogen levels following
18 installation of the different designs of biogas digester were monitored using faecal indicator
19 organisms within small rural farms in Ethiopia. However, different designs of digesters have
20 registered varying levels of success. Of the three digester designs considered, fixed dome, floating
21 drum and flexible balloon, the fixed dome design achieved the highest reductions in indicator
22 organisms (coliforms, *Escherichia coli* and *Enterococci*) from feedstock to digestate; this is likely
23 to be due its longer hydraulic retention time. Households with biogas digesters installed had
24 significantly ($p < 0.05$) lower levels of coliforms, *E. coli* and *Enterococci* detected in the
25 environment inside and outside the house area. However, in the same households, there was a
26 significant ($p < 0.05$) increase in indicator organisms on the surface and handles of doors,
27 indicating the potential for spread of pathogens on contaminated hands to door handles from
28 handling the animal waste feedstock. This therefore, suggests the need for proper hand washing
29 regimes after handling of feedstocks for biogas digesters.

30 31 **Introduction**

32 In recent years, there has been a drastic decline in the availability of biomass fuels across many
33 parts of Africa arising from increasing deforestation, partly due to collection of wood for fuel (Eleri
34 & Eleri, 2009). In Ethiopia, where forest cover is already less than 5% of the total country's land
35 area, the remaining area of forest decreases by 5% each year (FAO, 2015). With an effort to
36 preserve tree cover, different energy sources are being investigated, including liquefied petroleum
37 gas, electricity, solar energy and biogas. The main focus has been on electricity production, which

38 is favored in Ethiopia because of the high value attached to it by local communities. However,
39 there remains poor access to electricity in rural areas where capital investment costs are high
40 (Luijten & Kerkhof, 2011). Due to the cost of of installation and production, very few people
41 outside urban areas (2%) have access to it (IEA -International Energy Agency, 2011). The rate of
42 use of petroleum hydrocarbons in homesteads in Ethiopia remains very low in rural areas, where
43 only 1.3% of households, accounting for 93 million people, rely on petroleum hydrocarbon
44 sourced energy (Eleri & Eleri, 2009). Consequently, wood and charcoal are still widely used, and
45 in some cases, this will contribute to increased deforestation, biomass loss and associated land
46 degradation (Hoffmann, 2016).

47

48 In recent years, there has also been an increased emphasis on solar and biogas technologies (Gu et
49 al., 2016). The potential for adoption of biogas in rural Africa is high with 36% of the population
50 in the rural areas (over 6 billion people) owning livestock (Staal et al., 2009). In Africa, Ethiopia
51 has the greatest number of livestock; 60 million cattle, 60 million sheep and goats, 52 million
52 chickens & 4.5 million camels (Tegegne & Gatachew, 2020). Therefore, there is great potential
53 for development of biogas technology using animal manures (Gebreegziabher et al., 2014). Of
54 agricultural households, 77% are cattle owners, although this varies greatly from region to region
55 (Lindfors, 2010). In China, it has been suggested that biogas from cattle manure could be a major
56 substitute for wood fuel and electricity to meet the energy needs of the rural population (Gu et al.,
57 2016). With the high numbers of livestock, there may be similar potential for biogas production in
58 Ethiopia. Most importantly, the production of biogas is relatively simple and can operate under
59 both tropical and temperate conditions (Itodo et al., 2007).

60 Biogas is produced by a simple process, that uses anaerobic digestion of animal and plant wastes
61 to produce a gas containing 60-70% methane that can be used for cooking and lighting (Itodo et
62 al., 2007). Anaerobic digestion also offers the potential of a low-cost method to reduce pathogens
63 in agriculture wastes (Avery et al., 2014), and so is a possible method to manage organic waste
64 with the added benefits of energy generation. Organic wastes from animal, human and plant
65 sources are suitable feedstocks for biogas production. However, these wastes also contain
66 numerous protozoa, bacteria, fungi and viruses, a number of which can cause diseases to humans
67 (Nelson & Murray, 2008) and animals (Gannon et al., 2012).The major species of concern are
68 enteric bacteria, such as *Salmonella sp*, *Listeria spp*, *Escherichia coli*, *Bacillus spp*,

69 *Campylobacter spp*, *Mycobacteria spp*, *Clostridia spp*, *Brucella spp* and *Yersinia spp* (Sobsey,
70 2006). These pathogens may enter into the food chain via contaminated crops and meat products
71 or may infect humans and livestock populations via direct contact with manure.

72 Manures are widely handled in Ethiopia where they are formed into patties and used as cooking
73 fuel, a custom which inevitably presents a major route of exposure to humans (Johannessen et al.,
74 2004). The process of anaerobic digestion may reduce pathogen loads in the digestate relative to
75 the feedstock (Avery et al., 2014). This could mitigate some risks of high concentrations of
76 pathogens being disseminated to the wider environment and into the food chain. Adoption of
77 biogas as a cooking fuel may reduce the use of dung patties, so reducing direct contact exposure.
78 However, alternative manure handling practices are likely to arise as manures must be moved from
79 source into the biogas digester. As observed by Tumwesige et al. (2013) during previous studies
80 in Uganda, this may also result in households changing their livestock husbandry practices; for
81 example keeping livestock corralled closer to the house for easier manure collection for the
82 digester. Spread of manure close to the house may also result from spillage while conveying and
83 loading it into the digester. It is not clear what effect this will have on the overall burden of
84 pathogens in and around the home, and the overall impact on exposure to pathogens.

85
86 Pathogen die-off during anaerobic digestion arises primarily due to raised temperatures, increased
87 competition for microbial food sources and conditions that are non-ideal for particular species
88 (Rajendran et al., 2012). However, it is not known whether small scale household anaerobic
89 digestion reduces pathogen levels (widely evaluated using faecal indicator organisms — FIOs) to
90 acceptable standards. In the US, these standards correspond to 100,000, 10,000 and 100 coliform
91 forming units (CFU) per ml for *Enterococci*, coliforms and *E.coli* respectively (United States
92 Environmental Protection Agency, 2013).

93
94 Biogas digester technology is advancing in Ethiopia, and several designs have been developed and
95 are available on the market. These include flexible balloon, floating drum and fixed dome digesters
96 (Semple et al., 2014). However, there is little or no provision of guidelines for handling of animal
97 wastes, and the digesters themselves are likely to have different levels of efficacy in digesting and
98 sanitizing wastes.

99 This study uses standard FIOs to determine the capacity of a range of small-scale biogas digester
100 designs to reduce pathogen loads in feedstock as well as in and around households in rural
101 Ethiopia. The study aimed to determine whether introduction of different designs of biogas
102 digesters to small rural farms in Ethiopia is likely to increase or reduce overall exposure of farmers
103 and householders to zoonotic pathogens.

104 **Materials and methods**

105 **Description of the study area**

106 The study was conducted in Kumbursa village in the Ada'a district of Ethiopia. Kumbursa is
107 located at distance of approximately 55.5 km East of Addis Ababa between 8° 41'1" (0.13 km)"
108 and 8° 42'49" (14.05 m)"N, and 39° 00'29" (0.74 m)" and 39° 01'44" (1.42 m)"E, covering an area of
109 ~640 ha (Figure 1) The village is at an altitude of 1888-1992 m above sea level and is characterized
110 as "Woina dega" traditional agroclimatic zone Precipitation in Kumbursa originates from the
111 South-West equatorial air stream. The rainfall distribution pattern is uni-modal, with a peak
112 between June and September (74 % of the mean annual precipitation), and a total annual average
113 of 800 mm (Minase et al., 2016) . The annual mean temperature of the area ranges from 16.3 –
114 19.7 °C, with a ten-year mean of 18.1 °C and the hottest months in March, April and May (Minase
115 et al., 2016).

116 **Study design**

117 Twelve households in Kumbursa were purposively selected on the basis that they have the same
118 number of livestock (at least 10 cows), similar agricultural activities and similar homestead
119 structures. The main criterion for selection was households that can generate the critical mass of
120 feedstock of at least 80 kg per day required to sustain biogas generation (Semple et al., 2014). The
121 households were randomly divided into four equal groups. Each group was fitted with digesters of
122 one of three designs; (i) fixed dome, (ii) flexible balloon or (iii) floating drum, with the other group
123 without biogas digesters installed serving as control households. A randomized complete block
124 design was used, with households in the location using the same water source but with one of each
125 different biogas digester designs installed within the block. This was done to minimize any bias
126 due to water source. As is normal practice, fixed dome digesters with a volume of 16 m³ were
127 installed underground to maintain stable temperatures, flexible balloon digesters with a volume of

128 12 m³ were installed in a ditch and under the shade, and floating drum digesters with a volume of
129 8 m³ were installed above ground.

130 **Sampling for Faecal Indicator Organisms**

131 **Sampling of feedstock and digestate**

132 Digesters were fed daily throughout the experimental period with manure mixed with 100 – 120
133 dm³ of water in a ratio of 1:1. Assuming the typical feedstock to gas ratio of 5:1 (Smith et al.,
134 2013), this gives hydraulic retention times of 60, 45 and 30 days for the fixed dome, flexible
135 balloon and floating drum digesters, respectively; the ratio of hydraulic retention times is 2.0 (fixed
136 dome) : 1.5 (flexible balloon) : 1.0 (floating drum). Once biogas generation had begun, 10 ml
137 samples of feedstock (at the time of feeding which was carried out daily) and digestate (at the
138 HRT) were obtained from each of the digesters concurrently with the household environmental
139 sampling, within the five months period of the experiment. Samples were collected in sterile vials
140 and kept cool as described above until analysis on return to the laboratory.

141 **Sampling around households**

142 In this study total coliforms, *E. coli* and *Enterococci* were used as FIOs for pathogen load and
143 sanitization. These organisms are used internationally in environmental standards and, therefore,
144 this allows findings to be compared widely with other work. After installation of the biogas
145 digesters, FIO loads around the inside and outside floor environments of the household were tested
146 on two occasions. This was done using boot swabs (Bowden and Knights, UK) which were worn
147 on top of the boot to obtain environmental samples. On each occasion, fifty paces were taken
148 outside the house in the yard area with one boot swab and 20 inside the house with a second boot
149 swab. This approach has been previously demonstrated to provide a suitable method for swabbing
150 a complex environment where intensive sampling regimes are not logistically possible (Public
151 Health England, 2013). Boot swabs were removed and placed in individual filter bags and kept
152 cool in a cool box with ice for 2 hours during transit to the laboratory. On each occasion, a further
153 sample was obtained from the door handles using a wet swab which was then placed in an isotonic
154 transport diluent (Amies clear, plastic shaft) (Fisher, UK). Samples were obtained from all twelve
155 households (with and without biogas digesters) in the study over a 3-month period during the
156 hottest period of the year (March – July 2015).

157

158 **Analysis of Faecal Indicator Organisms**

159 Enumeration of FIOs from boot swabs taken outside and inside the house was performed using a
160 Colilert and Enterolert most probable number (MPN) methods (Idexx, UK) according to
161 manufacturer's instructions. To generate the inoculum, the swabs were washed in 100 ml of sterile
162 water in a filter bag, divided into 50 ml aliquots, diluted 1:2 (volume/volume (V/V)) and shaken
163 (25 revolutions per minute (rpm); 30 seconds). Further dilutions were carried out as required before
164 inoculating quanti-trays. For the door knob samples, 5 ml of transport diluent was added to 95 ml
165 of sterile water, shaken as above and then further diluted before inoculating quanti-trays.

166

167 Samples from the feedstock and digestate were analysed by first performing a ten-fold dilution
168 (V/V); 10 ml of the sample was placed in the filter bag to which 90 ml sterile water was added.
169 Further ten-fold dilutions were prepared as necessary to obtain readable counts Coliforms, *E. coli*
170 and *Enterococci* were enumerated by the methods described above.

171

172 **Data analysis**

173 All microbial counts were normalized by \log_{10} — transformation prior to analysis of variance
174 (ANOVA) by biogas digester type (Gen stat 12th edition). One way ANOVA was also performed
175 to the effect of treatments (households at which biogas digesters were installed; control houses
176 without digesters) for all FIO combined. The means were compared using the least significant
177 differences (LSD) at $p < 0.05$. Finally, the mean number of each FIO detected in digestate was
178 compared with acceptable levels of waste disposal, which were assumed to be 100,000, 10,000
179 and 100 CFU / ml (5, 4 and 2 \log_{10} CFU / ml) for *Enterococci*, coliforms and *E.coli* respectively
180 for safe disposal of digestate (US EPA, 2013).

181

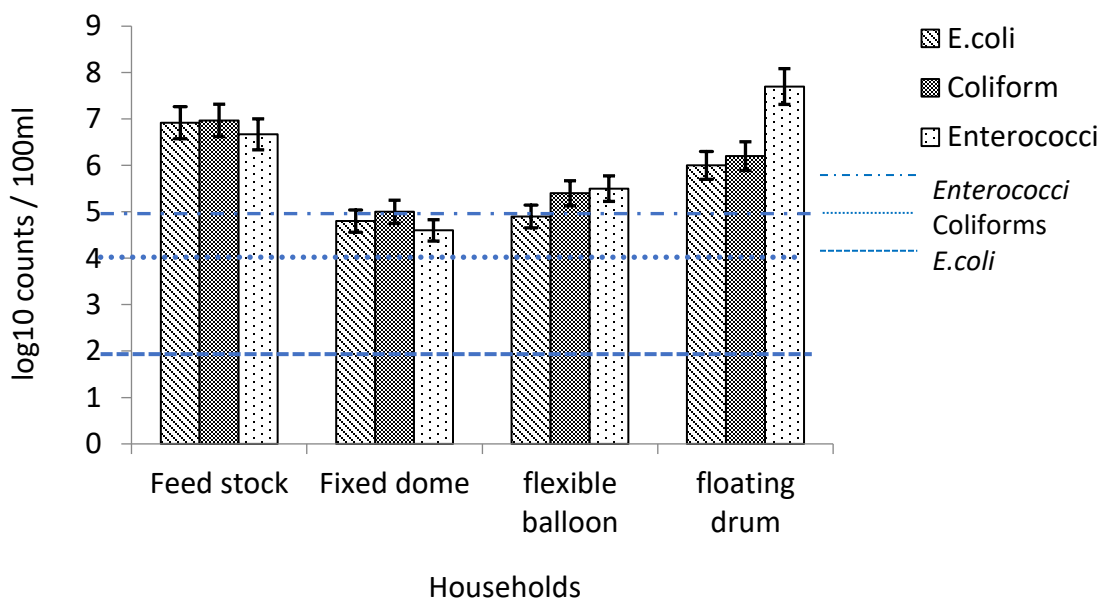
182 **Results**

183 **Sanitization of manure: Changes in Faecal Indicator Organisms counts during digestion**

184 In the feedstock, the FIO load was above the US EPA (2013) acceptable standards for waste
185 disposal for all organisms except *Enterococci*. Therefore, the feedstock would require sanitization
186 before safe disposal. All the biogas digester designs significantly ($p < 0.05$) reduced the population
187 of total coliforms and *E. coli* compared to the feedstock. However, while the *Enterococci*

188 population was reduced to 2.5 (standard error (\pm) 0.2) \log_{10} and 3.5 (\pm 0.2) \log_{10} CFU / ml in the
 189 fixed dome and flexible balloon digesters respectively, it increased to above the acceptable level
 190 of 5.0 \log_{10} CFU / ml to 5.5 (\pm 0.2) \log_{10} CFU / ml in the floating drum biogas digester design (Fig.
 191 1). The reductions between the fixed dome and flexible balloon were not significantly different,
 192 suggesting no additional benefit in sanitation was achieved by increasing the hydraulic retention
 193 time to above the 45 days in the flexible balloon digester to 60 days in the fixed dome digester.
 194 The fixed dome and flexible balloon digesters reduced *Enterococci* and coliforms to below US
 195 EPA standards of 5 and 4 \log_{10} CFU / ml respectively, but the reduction in *E.coli* was insufficient
 196 for safe disposal. The floating drum design, with its shorter hydraulic retention time, did not attain
 197 acceptable standards for disposal for any of the FIOs (Fig. 1).

198



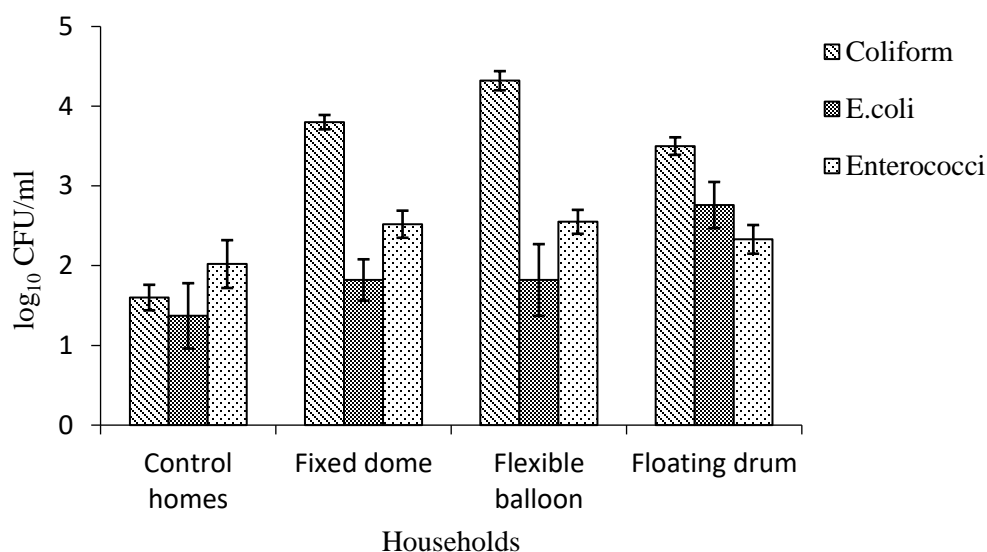
199

200 **Figure 1:** Concentrations of coliforms, *E. coli* and *Enterococci* in feedstock and digestate
 201 from different biogas digester designs installed in Households of Kumbursa,
 Ethiopia

202

203 **Faecal Indicator Organisms counts on door handles of households with and without biogas**
 204 **digesters**

205 Installation of biogas digesters in the household significantly ($p < 0.05$) increased counts of
 206 coliforms and *E. coli* on door swabs (Fig. 2). Coliforms significantly increased by $1.7 (\pm 0.2) \log_{10}$
 207 CFU / ml for floating drum digesters by $2.8 (\pm 0.2) \log_{10}$ CFU / ml for flexible balloon digesters
 208 and by $2.2 (\pm 0.2) \log_{10}$ CFU / ml for fixed dome. There was only a significant increase in *E. coli*
 209 ($2.8 (\pm 0.4) \log_{10}$ CFU / ml) (on door knobs where households had floating drum digesters. at 2.8
 210 ($\pm 0.4) \log_{10}$ CFU / ml; this increased *E. coli* to above safe levels ($2 \log_{10}$ CFU / ml). The trends
 211 did not align with the performance or hydraulic retention times of the digesters, suggesting that
 212 contamination occurs on feeding the digester, rather than on handling the digestate.



213 **Figure 2:** Concentration of coliforms, *E.coli* and *Enterococci* from door swabs of with and
 214 without biogas digester from households in Kumbursa, Ethiopia

215

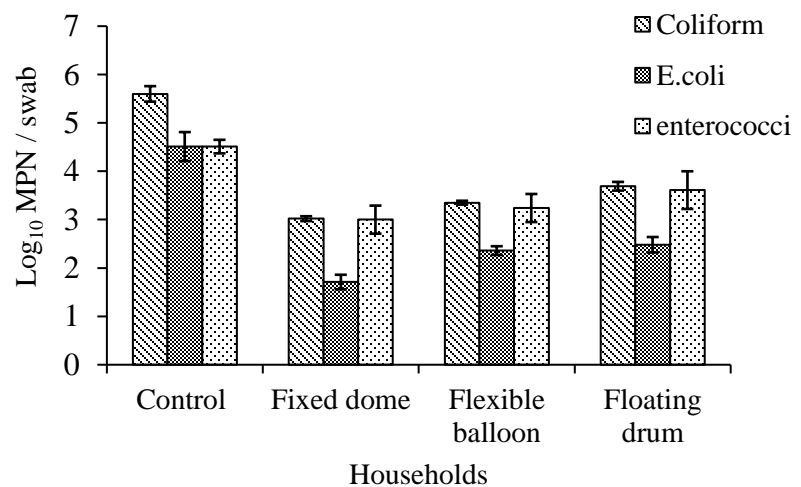
216 **Faecal Indicator Organisms counts inside and outside households with and without biogas**
 217 **digesters**

218 The counts inside the households of all FIO considered were significantly ($p < 0.05$) reduced in
 219 homes where biogas digesters had been installed (Fig.3). Coliforms were reduced by all types of
 220 digester from $5.6 (\pm 0.1) \log_{10}$ MPN / swab in households without biogas digesters to $3.7 (\pm 0.05)$;
 221 $3.4 (\pm ?)$ and $3.0 (\pm 0.05) \log_{10}$ MPN / swab in households with floating drum digesters; flexible
 222 balloon digesters and fixed dome digesters respectively. *E. coli* and *Enterococci* followed a similar
 223 pattern, *E. coli* was reduced from $4.5 (\pm 0.4) \log_{10}$ MPN / swab in the control to $2.5 (\pm 0.2)$; 1.4

224 (\pm ?) and $1.8 (\pm 0.2) \log_{10}$ MPN / swab in households with floating drum digesters; flexible balloon
 225 digesters and fixed dome digesters respectively. *Enterococci* were reduced from $4.5 (\pm 0.2) \log_{10}$
 226 MPN / swab in the control to $3.5 (\pm 0.4)$; $3.2 (\pm ?)$ and $3.6 (\pm ?) \log_{10}$ MPN / swab in households with
 227 floating drum digesters; flexible balloon digesters and fixed dome digesters respectively. The
 228 broad correspondence between the reduction in FIO inside households and digester performance
 229 of the different designs of digesters suggests that the lower counts inside the household is due to
 230 the reduction in organisms in the digestate.

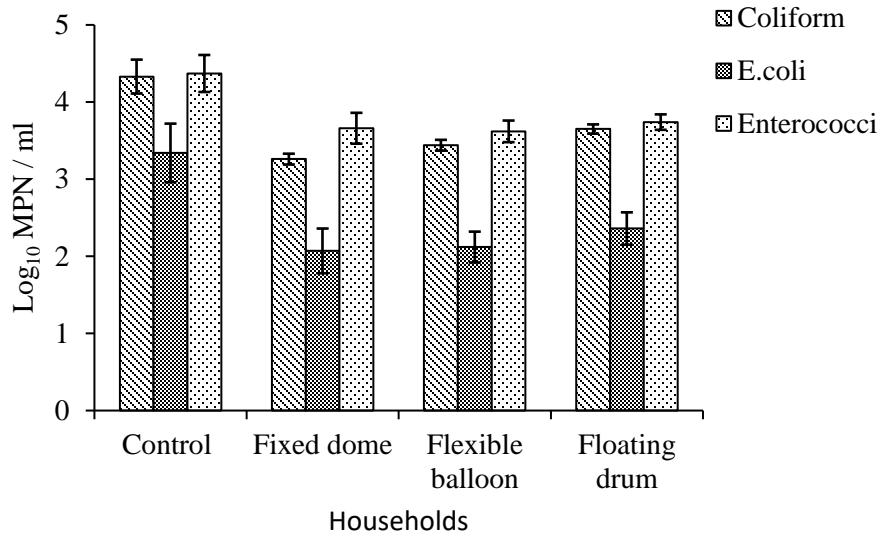
231 Installation of biogas digesters also significantly ($p < 0.05$) reduced the population of FIOs outside
 232 the house in the same order: fixed dome > flexible balloon > floating drum > (Fig.4). *E. coli* counts
 233 were reduced the most from $3.3 (\pm 0.4) \log_{10}$ MPN / swab in households without biogas digesters
 234 to between $2.0 (\pm 0.4)$ MPN / swab for the fixed dome and $2.4 (\pm 0.4) \log_{10}$ MPN / swab for the
 235 floating drum digester. *Enterococci* were reduced from $4.4 (\pm 0.2) \log_{10}$ MPN / swab in the control
 236 households to $3.0 (\pm 0.4)$ (fixed dome) to $3.6 (\pm 0.4) \log_{10}$ MPN / swab (floating drum). Coliforms
 237 were reduced from $4.2 (\pm 0.4)$ MPN / swab in the control to between $3.2 (\pm 0.1)$ MPN / swab in the
 238 fixed dome to $3.6 (\pm 0.1)$ MPN / swab in the floating drum.

239



240

241 **Figure 3:** Population of coliforms, E.coli and Enterococci inside the households with and without biogas digester



242

243 **Figure 4** Population of coliforms, *E.coli* and *Enterococci* from outside of houses of the
 244 household of Kumbursa, Ethiopia

244

245 **Discussion**

246 **Reductions in faecal indicator organisms counts from feedstock to digestate**

247 All three biogas digester designs significantly reduced the population of total coliforms and *E. coli*
 248 during digestion of feedstock. As mesophiles, coliforms and *E. coli* are sensitive to elevated
 249 temperatures of 45-60 °C and pH values below 6 and above 9. In a review of the literature, Avery
 250 et al. (2014) reported typical removals of 2 log₁₀ CFU / ml during mesophilic digestion. Production
 251 of organic acids during methanogenesis reduces the pH of the digestate (Chaudhry & Mukherjee,
 252 2016). The *E. coli* and total coliform counts were similar in feedstock and digestate of the different
 253 digesters indicating that *E. coli* being a subclass of coliforms, it dominated the coliform
 254 populations.

255

256 The digesters used in this study were selected because they are the most commonly adopted small
 257 scale biogas digester designs in Africa, and are locally available in Ethiopia. These designs have
 258 different hydraulic retention times, and the results obtained reflect this. The hydraulic retention
 259 time of feedstock or digestate in the biogas digester is well known to influence FIO die-off (Horan
 260 et al., 2004). The retention times of the floating drum, flexible balloon and fixed dome used in the
 261 study were 30, 45 and 60 days respectively, which broadly corresponded to increasing FIO

262 reductions. The biogas digesters' failure in reduction of *E. coli* to below acceptable standards
263 indicates that the HRT of 60 days and below is not sufficient enough to reduce *E. coli*. (Manyi-
264 Loh et al., 2016) suggested that even combination of two or more methods of treatment might be
265 the most effective way to control spread of pathogens from animal manures, for example biogas
266 digesters followed by composting.

267
268 *Enterococci* have been shown to be more resistant than *E. coli* during anaerobic digestion,
269 particularly at higher temperatures (Watcharasukarn et al., 2009). The capability of *Enterococci* to
270 withstand harsh conditions for some period of time contributed to its populations remaining
271 significantly higher than coliforms and *E. coli* in all types of digester (Fig.4). In the floating drum
272 digester, where the hydraulic retention time was only 30 days, not only did *Enterococci* not
273 decrease, but they actually increased, indicating that the conditions were suitable for growth.
274 *Enterococci* are indeed resistant to mesophilic temperatures (35-45°C) and have been shown to be
275 more resistant than faecal coliforms to biosolids treatment processes, including mesophilic
276 anaerobic digestion and composting (Bonjoch & Blanch, 2009; Viau & Peccia, 2009) (Martínez
277 et al., 2003; Watcharasukarn et al., 2009).

278 **Reductions in faecal indicator organisms counts in homesteads environments with and** 279 **without biogas digesters**

280 The presence of fixed dome, floating drum and flexible balloon designs of biogas digester
281 significantly reduced the populations of FIOs both inside and outside in the house. By contrast,
282 the door handles of households with biogas digesters installed had significantly higher levels of
283 FIOs than those without. This is likely to be attributable to contamination of hands from handling
284 manure during collection from the compound and mixing while feeding the biogas digester. It was
285 observed that during the collection and mixing of feedstock manure, most household members
286 used their hands without protective gloves, so there is a clear pathway of contamination from
287 manure to hands to door handles. The variation of FIO load across different biogas digester designs
288 is explained by the floating and fixed dome digesters having larger volumes compared to the
289 floating drum design. This required increased feeding frequency increasing the time of handling
290 manure and hence resulted in greater transfer of pathogens to hands, which were then transferred
291 to door handles. Our results indicate that while the ground was less contaminated, likely as a result
292 of gathering manure for digestion, the handling of manure for this purpose could potentially

293 increase human exposure to pathogens. Therefore, there is an important hygiene message which
294 must be clearly emphasized when introducing digesters into rural communities. To reduce
295 likelihood of spreading pathogens, it is critical for householders to wash their hands thoroughly
296 after handling feedstock manures. Since biogas digesters do not completely eliminate the FIOs and
297 pathogens, this also applies to handling the digestate.

298
299 It was noted that having a biogas digester at the homesteads significantly reduced coliform, *E. coli*
300 and Enterococci abundance on the indoor boot swabs (Fig. 3). This is likely to be a direct effect of
301 the reduction in FIOs on the ground around the outside of the homes, which may have been due to
302 gathering of dung from these areas to feed digesters, or changes in animal husbandry practices to
303 facilitate easier collection of manures. Dung is a vehicle for transmission of pathogenic to the
304 wider environment (Nelson & Murray, 2008). Therefore, when waste accumulates in the
305 compound or outside the household, this provides a reservoir for transmission into the household.
306 Households with biogas digesters collect and process the dung for biogas, so avoiding manure
307 accumulation and maintaining a lower load of manure-derived microorganisms around the
308 homestead. This is supported by the fact that of the three FIOs, it was the *E. coli* load that was
309 most reduced both during digestion, and in the indoor and outdoor floor samples. One confounding
310 factor in interpreting FIO loadings in the household environment in Ethiopia is that farming
311 householders' use dung to build and cement their houses, which may contribute to the observed
312 FIOs when dung used for building is still fresh. In contrast to this study Harroff et al., (2011)
313 reported that households with biogas digesters in Tiribogo, Central Uganda, had higher levels of
314 FIO inside the houses than those without biogas digesters. This could be attributed to difference
315 in the behavior of Ugandan farmers compared to farmers in Ethiopia. In Uganda most farmers
316 keep livestock, such as goats inside their houses, which is not the case in Ethiopia.

317
318 It is also likely that the reduction of pathogens in the biogas digester sanitization played a direct
319 role in reducing the pathogen inputs to the wider environment surrounding digester homesteads
320 (Tate et al., 2006). Of the three designs, the reduction in pathogens inside and outside the
321 household was least in the floating drum digester, which has the lowest hydraulic retention time.
322 This suggests that handling of the digestate could also have contributed to pathogen levels around
323 the home. Counts of *E. coli* on door handles increased the most in homes fitted with a floating

324 drum digester; this being the design of digester that least reduces the levels of *E. coli* in the
325 digestate. However, the ratio of coliforms to *E. coli* is higher than that in the feedstock and
326 digestate, indicating that there is a further source of coliforms or the coliforms survive better than
327 *E. coli* on the surfaces tested.

328

329 Conclusions

330 i. Fixed dome and flexible balloon digesters reduced FIOs to a similar degree and were
331 both more effective than the floating drum digesters. This is likely to be due to the
332 longer hydraulic retention time in these digesters (> 45 days).

333

334 ii. The biogas digester designs in this study did not reduce all pathogens to below
335 acceptable levels (US EPA, 2013). Therefore sanitisation of digestate could be
336 improved in rural Ethiopia by application of a supplemental treatment such as
337 composting.

338

339 iii. While digestion of animal manure is likely to reduce the pathogen burden in digestate
340 for application to land and can reduce environmental contamination inside and outside
341 the household, associated changes in practices may lead to increased exposure to
342 microorganisms derived from dung through increased handling of manure; this was
343 demonstrated by elevated FIO counts on door handles of homes with digesters.

344

345 iv. Improved education in hygiene of handling both manures and digestate is needed to
346 maximize the potential benefits of pathogen reduction through anaerobic digestion in
347 rural Africa

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