

# *Transfer of $^{13}\text{C}$ between paired Douglas-fir seedlings reveals plant kinship effects and uptake of exudates by ectomycorrhizas*

Article

Accepted Version

Pickles, B. J., Wilhelm, R., Asay, A. K., Hahn, A. S., Simard, S. W. and Mohn, W. W. (2017) Transfer of  $^{13}\text{C}$  between paired Douglas-fir seedlings reveals plant kinship effects and uptake of exudates by ectomycorrhizas. *New Phytologist*, 214 (1). pp. 400-411. ISSN 1469-8137 doi: <https://doi.org/10.1111/nph.14325>  
Available at <http://centaur.reading.ac.uk/67532/>

It is advisable to refer to the publisher's version if you intend to cite from the work.

To link to this article DOI: <http://dx.doi.org/10.1111/nph.14325>

Publisher: Wiley

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1 **Transfer of  $^{13}\text{C}$  between paired Douglas-fir seedlings reveals plant kinship**  
2 **effects and uptake of exudates by ectomycorrhizas**

3

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17 Word counts – Summary 197. Total: approx. 5030. Introduction 1027, Methods 1661, Results  
18 1054, Discussion: 1203, Acknowledgements: 85.

19 Tables: 3

20 Figures: 4 [Colour figures: 1, 4]

21 Supporting information: Methods 2, Figures 6

22 Running title: Kinship and mycorrhizal effects on carbon transfer

23 Subject categories: plant-soil interactions, multi-trophic systems, mycorrhizas [Fungal paper;  
24 Plant-relevant soil processes]

25

26 Conflict of interest statement: The authors declare no conflict of interest.

27 **Summary**

28

29 • Processes governing the fixation, partitioning, and mineralization of carbon in soils are  
30 under increasing scrutiny as we develop a more comprehensive understanding of global  
31 carbon cycling. Here we examined fixation by Douglas-fir seedlings and transfer to  
32 associated ectomycorrhizal fungi, soil microbes, and full-sibling or non-sibling  
33 neighbouring seedlings.

34

35 • Stable isotope probing with 99%  $^{13}\text{C}$ -CO<sub>2</sub> was applied to trace  $^{13}\text{C}$ -labelled  
36 photosynthate throughout plants, fungi, and soil microbes in an experiment designed to  
37 assess the effect of relatedness on  $^{13}\text{C}$ -transfer between plant pairs. The fixation and  
38 transfer of  $^{13}\text{C}$ -label to plant, fungal, and soil microbial tissue was examined in biomass  
39 and PLFAs.

40

41 • After a 6-day chase period, approximately 26.8% of the  $^{13}\text{C}$  remaining in the system  
42 was translocated belowground. Enrichment was proportionally greatest in  
43 ectomycorrhizal biomass. The presence of mesh barriers (0.5 or 35  $\mu\text{m}$ ) between  
44 seedlings did not restrict  $^{13}\text{C}$ -transfer.

45

46 • Fungi were the primary recipients of  $^{13}\text{C}$ -labelled photosynthate throughout the system,  
47 representing 60–70% of total  $^{13}\text{C}$ -enriched phospholipids. Full-sibling pairs exhibited  
48 significantly greater  $^{13}\text{C}$ -transfer to recipient roots in two of four Douglas-fir families,  
49 representing 3- and 4-fold increases (+ approx. 4  $\mu\text{g}$  excess  $^{13}\text{C}$ ) compared to non-  
50 sibling pairs. The existence of a root/mycorrhizal exudation – hyphal uptake pathway  
51 was supported.

52

53 **Key words** (5-8): Carbon allocation / ectomycorrhizas / host relatedness / interior Douglas-fir /  
54 PLFA / stable-isotope probing.

## 55 Introduction

56 Accurate estimates of belowground carbon cycling are critical to linking terrestrial  
57 ecosystems with biogeochemical processes and making useful predictions about how these  
58 may change under future climates (Richter & Billings, 2015). In temperate forests, the  
59 estimated quantities of atmospheric carbon sequestered via fixation are globally-relevant (~73-  
60 159 Pg C aboveground, and 153-195 Pg C belowground; Reichstein, 2007; Lorenz & Lal,  
61 2010), with humid and warm evergreen forests displaying the highest gross primary production  
62 of any temperate or boreal forest types (Luyssaert *et al.*, 2007). Photosynthate is mainly  
63 incorporated into plant biomass, but also supports a diverse microbial soil community either  
64 directly, via mycorrhizal fungi, or more generally, via scavenging of root exudates (Nehls *et*  
65 *al.*, 2007; Phillips *et al.*, 2011) and rhizodeposits (Jones *et al.*, 2009). Studying the  
66 belowground ecology governing carbon cycling is challenging due to the complexity of these  
67 communities and soil systems in general (De Deyn *et al.*, 2008; Bardgett *et al.*, 2013), and  
68 requires quantitative data on the allocation of photosynthate to plant biomass and its transfer to  
69 mycorrhizal fungi and microbes (Kaiser *et al.*, 2015). Here we undertook a multifaceted  
70 exploration of carbon allocation within and between paired interior Douglas-fir (*Pseudotsuga*  
71 *menziesii* var. *glauca*) seedlings, their mycorrhizal symbionts, and soil microbiota.

72 While the majority of plants form mycorrhizas, less than 5% of plant species are  
73 estimated to associate with ectomycorrhizal fungi (EMF). However, globally, these interactions  
74 are abundant in all forest biomes (Brundrett, 2009). Soils in ecosystems where ectomycorrhizal  
75 plants dominate exhibit higher C:N ratios compared to soils where they do not (Averill *et al.*,  
76 2014), and ectomycorrhizal hosts have been found to allocate 10% – 30% of their  
77 photosynthate to mycorrhizas (Söderström, 1992; Leake *et al.*, 2006; Högberg & Read, 2006).  
78 EMF incorporate photosynthate into their biomass, enhancing carbon sequestration by  
79 synthesising recalcitrant carbon compounds like chitin (Clemmensen *et al.*, 2013; Kashian *et*  
80 *al.*, 2013), especially in poorly oxygenated soil where decomposition is slow (Wallander *et al.*,  
81 2001). Additionally, EMF add mineral nutrients to soils through the breakdown of rock with  
82 organic acids (Hoffland *et al.*, 2003; Plassard & Fransson, 2009), and promote soil aggregate  
83 formation and carbon sequestration by exuding extracellular proteins and compounds that bind  
84 mineral particles (Rillig & Mummey, 2006; Graf & Frei, 2013). However, not all mycorrhizal  
85 activity is a carbon sink. In addition to respiration of host-derived photosynthate, some EMF  
86 species can decompose plant litter to acquire limiting nutrients, thus releasing soil carbon back  
87 into the atmosphere at rates comparable to saprotrophic fungi (Talbot *et al.*, 2008; Rineau *et*  
88 *al.*, 2013; Phillips *et al.*, 2014). Therefore, to assess how these processes jointly influence the

89 dynamics of soil carbon cycling in complex belowground systems, quantification of the fate of  
90 photosynthate in EMF-dominated systems is required.

91 An emergent property of mycorrhizal systems is the common mycorrhizal network,  
92 which arises when a fungal mycelium connects multiple plant hosts belowground across scales  
93 of cm<sup>2</sup> to at least tens of m<sup>2</sup> (Selosse *et al.*, 2006; Beiler *et al.*, 2010; Simard *et al.*, 2012).  
94 Ectomycorrhizal networks have been demonstrated to transfer water, nitrogen, and small  
95 quantities of carbon between interior Douglas-fir hosts (Simard *et al.*, 1997a; Teste *et al.*,  
96 2009; Bingham & Simard, 2011). Evidence that EMF display trait heritability based on host  
97 and fungal genotype (Rosado *et al.*, 1994a,b; Karst *et al.*, 2008), and that closely related plants  
98 display greater arbuscular mycorrhizal network size and root colonisation (File *et al.*, 2012;  
99 Dudley *et al.*, 2013), raises the possibility of preferential connectivity of kin through  
100 compatibility of parent-mycorrhiza-offspring genotypes. This has not previously been explored  
101 in conifers. In arbuscular mycorrhizal grassland plant species, root exudates play a role in ‘kin  
102 recognition’ (Dudley *et al.*, 2013) by moderating intra- and inter-specific plant root behaviour  
103 (Semchenko *et al.*, 2014), suggesting that plant relatedness may influence nutrient uptake or  
104 transfer by altering root growth and hence mycorrhizal formation. The finding that root  
105 exudates are important in kin recognition suggests that mycorrhizas are involved in recognition  
106 mechanisms in temperate forests, where trees are comprehensively mycorrhizal. If nutrient  
107 transfer through ectomycorrhizal networks can also differ with host relatedness, then fitness,  
108 and thereby forest stand composition, may be altered (Simard, 2009). Thus, to assess the  
109 potential for host relatedness effects in the experimental system, the extent of carbon transfer  
110 between ‘kin’ (full sibling) and ‘non-kin’ (no shared parent) seedling pairs was quantified.

111 Although mycorrhizal fungi are known to transfer labelled carbon between plants  
112 (Finlay & Read, 1986; Simard *et al.*, 1997b), interpretations are split between those suggesting  
113 retention of labelled carbon by fungi within their biomass (Graves *et al.*, 1997; Fitter *et al.*,  
114 1998; Wu *et al.*, 2001) and those indicating small (Simard *et al.*, 1997a; Teste *et al.*, 2010;  
115 Philip *et al.*, 2010) or large (Klein *et al.*, 2016) degrees of carbon transfer through fungal  
116 mycelium and into plant biomass. The transfer of carbon between plants may also occur via  
117 uptake of root exudates along an ‘exudation-dissolved organic carbon-mycorrhizal hyphae’  
118 pathway (Robinson & Fitter, 1999), and these hypotheses regarding the mechanism of transfer  
119 need not be mutually exclusive. In this experiment we established a size-hierarchy between  
120 seedlings of different ages together in the same pot, separated by a mesh barrier to prevent  
121 direct root interaction, and labelling the larger (older) seedling with <sup>13</sup>C-CO<sub>2</sub>.

122 A small number of stable isotope probing (SIP) studies, in which a plant is exposed to  
123  $^{13}\text{C}$ -labelled  $\text{CO}_2$  and the distribution of  $^{13}\text{C}$ -photosynthate is examined after a specific time  
124 period, have investigated interconnections between the Pinaceae and their EMF symbionts (see  
125 Epron *et al.*, 2012). Here, for the first time in a paired seedling model system, we investigated  
126 the fungal and microbial communities active in plant photosynthate assimilation and transfer,  
127 applying SIP methods to quantify  $^{13}\text{C}$  allocation to plant and fungal biomass and the  
128 phospholipid fatty acids (PLFA) of plants, fungi, and bacteria (Boschker *et al.*, 1998). Our  
129 study examined paired interior Douglas-fir seedlings with the following objectives: (i) to  
130 quantify the distribution of  $^{13}\text{C}$ -labeled photosynthate throughout plant and soil carbon pools  
131 and within microbial biomass, (ii) to determine the proportion of carbon transferred (if any)  
132 between seedlings by EMF symbionts, and (iii) to determine whether relatedness has a role in  
133 carbon transfer between conifer seedlings.

134

## 135 **Materials and methods**

### 136 *Seed and soil*

137 Seeds from four ‘families’ (cross-bred from four different pairs of known parents) of  
138 interior Douglas-fir (*Pseudotsuga menziesii* (Mirb.) Franco var. *glauca* (Beissn.) Franco) were  
139 sourced from the B.C. Ministry of Forests, Lands, and Natural Resource Operations  
140 (Kalamalka Research Station, B.C.). To encourage EMF colonization, seeds were grown in a  
141 soil mixture with the following specifications: 1:1 mix (approximately 3.4 L total) of  
142 autoclaved potting soil (1.7 L 1:1 mix of peat to perlite) and forest soil (1.7 L). Forest soil was  
143 classified as Dystric Brunisol (Inceptisol in U.S. soil taxonomy) with moderate humus and  
144 sandy loam texture (Soil Classification Working Group, 1998). Soil was collected from two  
145 sub-locations within a mono-specific interior Douglas-fir stand (120.58°W, 49.43°N) in the  
146 Dry, Cool Interior Douglas-fir (IDFdk) biogeoclimatic subzone (Pojar *et al.*, 1987). Following  
147 the removal of the litter layer, the fermentation layer, humus layer, and mineral soil were  
148 sampled to a total depth of 10-15 cm. Large debris was removed during collection and soil  
149 was homogenised in clean conditions before sub-sampling to create experimental units.

150

### 151 *Experimental design and seedling growth*

152 Each experimental unit was a 3.8 L pot containing a pair of seedlings of different ages  
153 to establish a carbon gradient, which were spaced approximately 8 cm apart and separated by a  
154 nylon mesh bag (Plastok® Meshes and Filtration Ltd., Birkenhead, UK). One ‘donor’ seedling  
155 was established from seed planted in March 2012 in the 805 g (11.9 s.e.m.) dry weight soil

156 outside the mesh bag. One ‘recipient’ seedling was established from seed planted in November  
157 2012 inside the 8 x 18 cm mesh bag containing 403 g (5.9 s.e.m.) dry weight soil. Seedlings  
158 were grown in a glasshouse without supplementary light or fertiliser, in order to encourage  
159 mycorrhizal formation. Pots were watered to field capacity once per week following an early  
160 germination period of light daily watering. A fine gravel layer was applied to soil surfaces to  
161 discourage infection and mortality by pathogenic soil fungi (e.g. *Fusarium*, *Phytophthora*,  
162 *Pythium*, *Rhizoctonia*; colloquially called ‘damping off’ fungi). Harvesting took place in the  
163 first week of February 2013, when donor and recipient seedlings were 11 and 3 months old,  
164 respectively (Asay, 2013).

165 Seedling relatedness was manipulated through seed origin: seedling pairs were either  
166 full siblings (same parents; ‘kin’) or non-siblings (different parents; ‘non-kin’). Kin seed was  
167 obtained from four pairs of parent trees to assess whether any relatedness effects were a  
168 general observation or family specific. Sufficient seed was provided to establish 10 kin pairs  
169 from each of the four sets of parents (n=40). Non-kin seed was more readily available so  
170 additional pairs were planted (n=60). Hyphal connection was manipulated via mesh bags:  
171 seedling pairs were separated by a root-blocking barrier that either allowed the passage of EMF  
172 hyphae (35  $\mu\text{m}$  mesh) or was expected to restrict or prevent the passage of EMF hyphae (0.5  
173  $\mu\text{m}$ ) (Teste *et al.* (2006). Mesh sizes were divided equally among the relatedness treatments  
174 resulting in the following experimental factors: kin-unrestricted (n=20), kin-restricted (n=20),  
175 non-kin-unrestricted (n=30), non-kin-restricted (n=30).

176

#### 177 *<sup>13</sup>CO<sub>2</sub> isotope labelling*

178 Donor plants were pulse-labelled with 99 atom% <sup>13</sup>C-CO<sub>2</sub> eleven months after they  
179 were established in pots. Surviving pairs in each treatment were assigned to subsequent  
180 analyses as follows. To estimate the initial uptake and fixation of <sup>13</sup>C-label, we assessed  
181 incorporation of <sup>13</sup>C-photosynthate into plant and fungal biomass 1 day after labelling (“1-day  
182 chase”): non-kin-unrestricted (n=3), non-kin-restricted (n=3), non-labelled controls (n=3). To  
183 examine the transfer and incorporation of <sup>13</sup>C-photosynthate into biomass 6 days after labelling  
184 (“6-day chase”): kin-unrestricted (n=9), kin-restricted (n=8), non-kin-unrestricted (n=10), non-  
185 kin-restricted (n=10). Unlabelled controls consisted of kin-unrestricted (n=9), kin-restricted  
186 (n=8), non-kin-unrestricted (n=7), non-kin-restricted (n=7).

187 Immediately before labelling, all donor seedlings were sealed using Tuck® Contractors  
188 Sheathing Tape inside plastic Foodsaver® vacuum bags (6 L). Bags were fitted with an  
189 injection valve and inflated with ambient air. Non-labelled control seedlings were bagged in



190 the same manner and stored with a minimum separation distance of 4 m from the nearest  
191 labelled seedling. Three injections of  $^{13}\text{C}$ -CO<sub>2</sub> were received by labelled seedlings at equal  
192 time intervals through the 10 h pulse period, totalling 50 mL of  $^{13}\text{C}$ -CO<sub>2</sub> (with maximum  
193 concentrations of 2500-3000 ppm). An additional seedling was used to monitor bag CO<sub>2</sub>  
194 concentration using a portable infrared gas analyser (Qubit Systems, Kingston, Canada).  
195 Ambient greenhouse CO<sub>2</sub> levels (394 ppm) were checked prior to labelling, 5 h after labelling,  
196 and 10 h after labelling and showed no significant variation. After the final pulse, when bag  
197 CO<sub>2</sub> concentrations had dropped below 300 ppm, labelling bags were removed.

198

#### 199 *Sampling of plant and soil pools*

200 Eleven distinct plant and soil pools were examined in this experiment, with donor (D)  
201 and recipient (R) seedling samples collected as described in Figure 1. Due to growth stage  
202 differences between donor and recipient seedlings, root sections were collected from donor  
203 transport fine roots (McCormack *et al.*, 2015) and the recipient's main tap root, which were of  
204 equivalent diameter and structure (i.e. no absorptive fine roots were included in these samples).  
205 Samples were kept on dry ice after weighing, and stored at -80 °C. All fine root tips were  
206 sampled from each seedling, morphotyped based on ectomycorrhizal structures or their absence  
207 (Goodman *et al.*, 1998), then counted and weighed prior to subsampling for fungal  
208 identification via amplified internal transcribed spacer (ITS) sequences. All root tips from six  
209 plant pairs were kept on dry ice after weighing, and stored at -80 °C. Soil for PLFA analyses  
210 was immediately frozen in liquid nitrogen before storage at -80 °C. Samples were lyophilized  
211 prior to DNA or PLFA extraction and isotopic analysis. Remaining plant biomass from each  
212 pool was oven-dried and weighed.

213

#### 214 *Isotopic measurements*

215 Total carbon and nitrogen content and carbon isotopic composition of samples were  
216 measured with combustion analysis using an elemental analyzer (Elementar, Hanau, Germany)  
217 in C, N mode, interfaced with an isotope-ratio mass spectrometer (IRMS; Isoprime, Cheadle,  
218 UK). Samples were considered enriched if their  $\delta^{13}\text{C}$  value was greater than the upper 99%  
219 confidence interval of the control mean  $\delta^{13}\text{C}$  (natural abundance) and all control sample  $\delta^{13}\text{C}$   
220 values. Atom %<sup>13</sup>C excess was calculated for each pool as per Leake *et al.* (2006). Teste *et al.*'s  
221 (2009) modification of Boutton's (1991) isotopic calculations was applied to convert  $\delta^{13}\text{C}$  into  
222 "excess <sup>13</sup>C" as <sup>12</sup>C-equivalent (mg), the mass of labelled carbon compensating for the one  
223 Dalton difference in mass of <sup>12</sup>C.

224 Isotopic composition of PLFAs was analysed using gas chromatography-IRMS,  
225 following extraction according to Bligh & Dyer (1959), as detailed in Churchland *et al.* (2013),  
226 with the following exceptions: (i) methyl undecanoate (c11:0) was the internal standard, and  
227 (ii) quantitation was performed based on average values derived from serial dilution of  
228 undecanoate, nonadecanoate (c19:0), and methyl cis-13-docosenoate (c22:1 $\omega$ 9). Peak  
229 identification was based on retention time compared to two reference standards: bacterial acid  
230 methyl-ester standard 47080-0 (Sigma–Aldrich, St. Louis, USA) and a 37-Component fatty  
231 acid methyl-ester mix (47885-U). Unidentifiable <sup>13</sup>C-enriched peaks were included in analysis  
232 if they met the following conditions: i) detection in > 3 samples, ii) average  $\delta^{13}\text{C} > +50 \text{‰}$ .  
233 Taxonomic affiliations of specific PLFAs were assigned as per Högberg *et al.* (2013), with  
234 c18:1 $\omega$ 9 and c18:3 $\omega$ 6 added as fungal markers, according to Ruess & Chamberlain (2010).

235

#### 236 *Fine root tip fungal DNA extraction-sequencing*

237 Fungal DNA was extracted from multiple representative root tips of each EMF  
238 morphotype, from which adhering soil had been carefully removed, using ITS1 (White *et al.*,  
239 1990) and ITS4/ITS4B primers (Gardes & Bruns, 1993) following the protocol provided in  
240 Supporting Methods S1. Raw sequence data were analysed using SEQUENCHER Version 3.0  
241 (Gene Codes Corp., Ann Arbor, USA) and converted into FASTA format prior to comparison  
242 with the UNITE database (Kõljalg *et al.*, 2013), using the BLAST algorithm to identify each  
243 fungal species. Sequence data were deposited in the GenBank database as accession numbers  
244 KT314836 to KT314861. Three samples from the rhizosphere soil partition were selected for  
245 metatranscriptomic sequencing to assess activity of root-associated communities (Supporting  
246 Methods S2).

247

#### 248 *Statistical analyses*

249 All analyses were performed using R 3.2.3 (R Core Team, 2015) unless otherwise  
250 stated. Data was square root or  $\log_{10}$  transformed where necessary to meet parametric  
251 assumptions, with highly influential data points ( $\geq 3$  st. dev. from the treatment median) treated  
252 as statistical outliers and removed prior to analysis. Differences between treatments in the  
253 excess <sup>13</sup>C content of recipient pools were assessed by fitting linear mixed models in R-  
254 packages “nlme” (Pinheiro *et al.*, 2016) and “lme4” (Bates *et al.*, 2015). The fixed factors in  
255 each model were: seedling relatedness (kin or non-kin), hyphal restriction (35  $\mu\text{m}$  or 0.5  $\mu\text{m}$ ),  
256 recipient family (A, B, C, or D), and their two-way interactions. In all models, donor family  
257 was included as a random factor. The response variables examined were “excess <sup>13</sup>C as <sup>12</sup>C

258 equivalent” of the  $R_{\text{bulk}}$ ,  $R_{\text{rhizosphere}}$ ,  $R_{\text{ECM}}$ , and  $R_{\text{root}}$  pools. Model fit was determined using R-  
259 package “piecewiseSEM” (Lefcheck, 2016). Standardised coefficients, a measure of  
260 standardised effect size (SES) (Schielzeth, 2010), were estimated for each model (regression  
261 coefficients divided by two times their standard deviation).

262 Linear models were used to assess whether excess  $^{13}\text{C}$  in kin and non-kin recipient  
263 pools displayed different relationships to potentially explanatory biological factors ( $D_{\text{shoot}}$   $^{13}\text{C}$ -  
264 enrichment,  $D_{\text{ECM}}$   $^{13}\text{C}$ -enrichment,  $D_{\text{ECM}}$  abundance). SES was measured using Cohen’s *d*  
265 (Cohen, 1988). False discovery rate (FDR) correction (Verhoeven *et al.*, 2005) was applied  
266 where data was regressed against multiple factors.

267 Enriched fungal biomass ( $^{13}\text{C}$  per g tissue dry weight) was calculated by converting  
268 from mg enriched fungal PLFAs, using the conversion factor provided by Joergensen &  
269 Wichern (2008). Average carbon incorporation into fungal biomass was calculated for  $D_{\text{ECM}}$   
270 (reflecting the hartig net, mantle, and extramatrical mycelium of EMF) and  $D_{\text{root}}$  (to account for  
271 the presence of fungal endophytes and any potential extension of EMF hyphae into transport  
272 fine roots; see Kaiser *et al.*, 2010).

273 Seedling EMF community data was examined with the Sørensen (Bray-Curtis) distance  
274 measure using nonparametric multi-dimensional scaling (NMDS) and multi-response  
275 permutation procedures (MRPP) in PC-Ord 5 (MjM Software, Gleneden Beach, USA).

276 All data used in this analysis, along with a custom script for processing SIP-PLFA data,  
277 can be found at the stable URL: <https://github.com/roli-wilhelm>.

278

## 279 **Results**

### 280 *Partitioning of $^{13}\text{C}$ -labeled photosynthate*

281 Every plant and soil pool exhibited elevated levels of  $^{13}\text{C}$ -labelled carbon relative to the  
282 natural abundance in unlabelled controls (Table 1). No significant differences in  $\delta^{13}\text{C}$  were  
283 observed among unlabelled controls.  $^{13}\text{C}$ -enrichment of  $R_{\text{shoot}}$ , the most distant pool from donor  
284 plants, was significant in 4 kin and 2 non-kin samples (6/37). The decrease in  $^{13}\text{C}$ -labelled  
285 carbon, from  $D_{\text{shoot}}$  to  $R_{\text{shoot}}$ , revealed the scope and scale of carbon flow through the  
286 belowground system (Table 1; Supplementary Figure S1a). As expected, in each plant  $D_{\text{ECM}}$   
287 and  $R_{\text{ECM}}$  contained significantly more  $^{13}\text{C}$ -labelled carbon as a percentage of their total carbon  
288 content compared to all other pools, illustrating their assimilation of this carbon  
289 (Supplementary Figure S1b).

290 Of the total mass of pulsed  $^{13}\text{C}$ -labelled carbon (29.02 mg), approximately 75.4%  
291 (21.88 mg) was fixed in donor plant tissue after the 1-day chase. Following the 6-day chase

292 period, approximately 44.1% (9.64 mg) of the fixed label (33.2% of the pulse) was detected  
 293 across all biomass pools. The total transfer to measured belowground and recipient pools was  
 294 approximately 12% of the  $^{13}\text{C}$ -label fixed, amounting to 26.8% of the total  $^{13}\text{C}$ -label detected  
 295 in the 6-day chase. For an account of all individual pools see Table 1. The remaining  
 296 unaccounted  $^{13}\text{C}$ -label was either not fixed or fixed and respired during the labelling period.

297 A strong inverse relationship was observed between  $D_{\text{shoot}} \delta^{13}\text{C}$  and donor biomass ( $r^2 =$   
 298 0.66,  $P < 0.001$ ), while the total mass of excess  $^{13}\text{C}$  in  $D_{\text{shoot}}$  did not vary significantly with  
 299 donor biomass ( $r^2 < 0.01$ ,  $P = 0.29$ ). Thus larger plants had a lower  $^{13}\text{C}$ -content relative to total  
 300 seedling biomass than smaller plants.

301

### 302 *Role of seedling relatedness and ectomycorrhizal hyphae in carbon transfer*

303 Analysis using linear mixed models revealed a seedling relatedness effect on  $R_{\text{root}}$   
 304 excess  $^{13}\text{C}$ , with the significant model interaction term revealing variation between the four  
 305 recipient families (Table 2). Significantly greater  $R_{\text{root}}$  excess  $^{13}\text{C}$  was observed in kin  
 306 recipients from families A and B than in non-kin recipients (Fig. 2a), whereas there was no  
 307 significant difference based on relatedness in families C and D. In terms of  $\mu\text{g}$  excess  $^{13}\text{C}$  this  
 308 represented a 3-fold increase in kin of family A (kin mean 5.7  $\mu\text{g}$ ; non-kin mean 1.9  $\mu\text{g}$ ), and a  
 309 4-fold increase in kin of family B (kin mean 5.6  $\mu\text{g}$ ; non-kin mean 1.4  $\mu\text{g}$ ). Carbon transfer to  
 310 other recipient pools was not significantly different between relatedness treatments (Table 2).  
 311 Linear regression analysis revealed that  $R_{\text{root}} \delta^{13}\text{C}$  enrichment increased with increasing  $D_{\text{ECM}}$   
 312 abundance in kin pairs only (Fig. 3a). In both kin and non-kin pairs,  $R_{\text{bulk}} \delta^{13}\text{C}$  enrichment  
 313 increased with increasing  $D_{\text{ECM}}$  abundance (both:  $r^2 = 0.34$ , kin:  $P = 0.014$ ; non-kin:  $P =$   
 314 0.007). In kin pairs only,  $R_{\text{ECM}}$  excess  $^{13}\text{C}$  was positively associated with that of both  $D_{\text{ECM}}$   
 315 (Fig. 3b) and  $D_{\text{shoot}}$  (Fig. 3c).

316 Hyphal exclusion did not reduce colonisation of recipient roots, or significantly reduce  
 317 overall sub-surface carbon transfer. However, analysis using linear mixed models (Table 2)  
 318 indicated a reduction in  $R_{\text{bulk}}$  excess  $^{13}\text{C}$  with hyphal exclusion in recipient family D (SES =  
 319 0.60-0.66), and reduced  $R_{\text{rhizosphere}}$  excess  $^{13}\text{C}$  among kin recipients in families C (SES = 0.58)  
 320 and D (SES = 0.62). Conversely, increased  $R_{\text{root}}$  excess  $^{13}\text{C}$  with hyphal exclusion was  
 321 observed for kin in recipient family A and regardless of relatedness in family D (Fig. 2b).  
 322 Carbon transfer to other recipient pools did not differ between mesh treatments (Table 2).

323 To assess whether the observed relatedness effects could be due to differences in  
 324 belowground biomass allocation between families, the same linear mixed models were  
 325 performed for donor and recipient root: shoot ratio, and the biomass of  $D_{\text{root}}$ ,  $R_{\text{root}}$ ,  $D_{\text{ECM}}$  and

326 R<sub>ECM</sub>. The only significant factor across these models was that family C exhibited a lower root:  
327 shoot ratio than the other three families. No other fixed or interactive factors proved to be  
328 significant (data not shown).

329

### 330 *Assimilation of <sup>13</sup>C-photosynthate by the microbial community*

331 Total <sup>13</sup>C-incorporation into PLFAs in all pools amounted to 1% of belowground  
332 carbon transfer (0.023 mg <sup>13</sup>C in PLFAs / 2.31 mg total <sup>13</sup>C transferred). Fungal PLFA <sup>13</sup>C-  
333 enrichment was 5.57 mg fungal C g<sup>-1</sup> dry weight in D<sub>ECM</sub> and 1.97 mg fungal C g<sup>-1</sup> dry weight  
334 in D<sub>root</sub>. Estimated D<sub>ECM</sub> PLFA <sup>13</sup>C-enrichment was strongly correlated with <sup>13</sup>C-enrichment of  
335 D<sub>ECM</sub> tissue ( $r^2 = 0.844$ ,  $P < 0.001$ ), and D<sub>ECM</sub> PLFAs contained the highest total excess <sup>13</sup>C  
336 measured (Supplementary Figure S2). The D<sub>root</sub> pool contained the next highest level of PLFA  
337 <sup>13</sup>C-enrichment, originating from the plant root and associated fungi and bacteria. Fungal  
338 PLFA markers were the most <sup>13</sup>C-enriched in all pools, containing ~70% of assimilated <sup>13</sup>C-  
339 label. The second most <sup>13</sup>C-enriched taxonomic group were the 'higher eukaryotes,' a  
340 heterogeneous category of long-chain fatty acids, indistinguishable between fungi, plants, and  
341 other eukaryotic species.

342 Microbes closely associated with host roots were less diverse than those in soil (~27  
343 PLFAs), based on the average number of enriched PLFAs (D<sub>root</sub>: 12, D<sub>ECM</sub>: 9, R<sub>ECM</sub>: 13).  
344 Differences between root-associated and soil-associated communities were evident in <sup>13</sup>C-  
345 enriched PLFA profiles (Supplementary Figure S3) and unidentified fatty acid profiles  
346 (Supplementary Figure S4). The D<sub>coarse</sub> community was the most distinct, displaying increased  
347 <sup>13</sup>C-enrichment of medium-length fatty acids between c14:0 – c16:1ω9. Fungi assimilated the  
348 vast majority of photosynthate based on total <sup>13</sup>C-enrichment of PLFAs; however, other  
349 taxonomic groups in the rhizosphere exhibited substantial assimilation rates (Fig. 4). In the  
350 D<sub>root</sub> pool bacteria incorporated <sup>13</sup>C-label at rates comparable to that of fungi (i.e. relative to  
351 their biomass). Gram-negative bacteria assimilated significant amounts of <sup>13</sup>C-exudate across  
352 all donor pools in every sample assayed, whereas gram-positive bacteria did not assimilate  
353 detectable <sup>13</sup>C-exudate in ectomycorrhizal pools.

354

### 355 *Fungal root-tip community*

356 Seedlings were primarily colonised by *Rhizopogon vinicolor*, and an ectomycorrhizal  
357 Pyrenomataceae sp. (Table 3), both of which were also detected in a preliminary  
358 metatranscriptomic analysis of three rhizosphere soil samples (Supporting Methods S1 and Fig.  
359 S5). The abundance of these EMF species on recipient seedlings was positively related to their

360 abundance on donor seedlings regardless of treatment. MRPP analysis following NMDS  
361 ordination (Supporting Fig. S6) revealed that the only significant difference between EMF  
362 communities was weak and occurred between donor and recipient seedlings, rather than  
363 treatments ( $A = 0.131$ ,  $P < 0.001$ ; where  $A > 0.3$  is considered an ecologically relevant effect).  
364 Notably, in the sole case where a plant lacked *Rhizopogon* sp. ectomycorrhizas,  $^{13}\text{C}$ -enrichment  
365 was not detected in recipient pools.

366

## 367 Discussion

368 Fungi dominated the assimilation of photosynthetic carbon in all belowground  
369 experimental pools, with ectomycorrhizal fungi serving as major agents of carbon transfer.  
370 EMF incorporation of photosynthate from host plants is hypothesised to be a major factor in  
371 carbon sequestration in coniferous forests (Clemmensen *et al.*, 2013). An estimated 26.8% of  
372 the  $^{13}\text{C}$ -label remaining in the system was recovered from belowground pools (primarily donor  
373 roots), of which 6.3% was assimilated by, or transferred through, EMF. Carbon transfer from  
374 donor to recipient seedlings was significantly greater to sibling roots than non-sibling roots in  
375 two of the Douglas-fir families, indicating a host relatedness effect that was most likely  
376 mediated by EMF. The transfer of  $^{13}\text{C}$ -label in the presence of a hyphae-restricting mesh  
377 implies that labile  $^{13}\text{C}$ -compounds were exuded into soil by donor roots and/or EMF, before  
378 being taken up by recipient hyphae. Overall we observed that a diverse microbial community  
379 was actively assimilating  $^{13}\text{C}$ -labeled photosynthate.

380

### 381 *Scale and significance of belowground partitioning of Douglas-fir photosynthate*

382 Carbon allocation to donor plant root and ectomycorrhizal root tip biomass (~23% of  
383 the total recovered) was within the range of previously estimated allocation values for  
384 ectomycorrhizal seedlings of other species: Norway spruce, Scots pine, and silver birch (13-  
385 24%; Pumpanen *et al.*, 2008), Scots pine (31%; Heinonsalo *et al.*, 2010), and willow (47%;  
386 Durall *et al.*, 1994). Furthermore,  $^{13}\text{C}$ -enrichment of fungal-specific PLFAs indicated  
387 significant carbon allocation to fungi within donor transport fine roots, most likely attributable  
388 to intra-root EMF biomass (Kaiser *et al.*, 2010) and/or fungal endophytes.

389 Mycorrhizal networks in mature forests can be extensive (Beiler *et al.*, 2010), offering  
390 the potential for substantial carbon transfer among plants. Yet the net benefit of seedling-to-  
391 seedling carbon transfer remains poorly understood. Previous research indicates that EMF and  
392 their mycorrhizal networks mediate the transfer of variable amounts of carbon (Simard *et al.*,  
393 1997a; Teste *et al.*, 2009; Philip *et al.*, 2010), water (Allen, 2007; Plamboeck *et al.*, 2007;

394 Bingham & Simard, 2011), and nitrogen (He *et al.*, 2003; Read & Perez-Moreno, 2003; Teste  
395 *et al.*, 2009) between plants. The quantities of seedling-seedling carbon transfer measured in  
396 our experiment (~0.1% of total recovered  $^{13}\text{C}$ -label across all pools) and elsewhere (Teste *et*  
397 *al.*, 2009, 2010) are small, and unlikely to represent a substantial nutritional source. However,  
398 research on the EMF host-symbiont interface reveals that EMF produce signalling compounds,  
399 which are translocated into plant cell nuclei. For example, *Laccaria bicolor* produces the  
400 MiSSP7 protein, which alters the host transcriptome to promote ectomycorrhizal formation and  
401 reduce jasmonic acid production (Plett *et al.*, 2011, 2014). Our observations may therefore  
402 represent the transfer of signalling compounds through a fungal mycelium, or their uptake by  
403 roots or fungal hyphae following exudation. The stimulation of physiological responses in  
404 recipient plants by potential signalling compounds has previously been observed in arbuscular  
405 mycorrhizal systems (Babikova *et al.*, 2013; Song *et al.*, 2014) and EMF systems involving  
406 interior Douglas-fir and ponderosa pine (Song *et al.*, 2015).

407

408 *Host relatedness*

409 Intriguingly, kin pairs exhibited increased carbon transfer to the  $R_{\text{root}}$  pool in two of the  
410 four Douglas-fir families, with excess  $^{13}\text{C}$  in those families 3 to 4-fold greater than in non-kin  
411 pairings. The absolute quantities of increased excess  $^{13}\text{C}$  involved in this relatedness effect  
412 were small (+ 4  $\mu\text{g}$ ), but represented a large proportion of total recipient plant excess  $^{13}\text{C}$   
413 content (overall recipient plant mean: 6.4  $\mu\text{g}$ ). The strong positive relationship between  $R_{\text{root}}$   
414  $\delta^{13}\text{C}$  and  $D_{\text{ECM}}$  abundance in kin pairs demonstrated that the genetic relatedness effect on  
415 carbon transfer involved donor plant EMF. This was further reinforced by the positive  
416 relationships between excess  $^{13}\text{C}$  in the  $R_{\text{ECM}}$  pool and both the  $D_{\text{shoot}}$  and  $D_{\text{ECM}}$  pools in kin  
417 pairs. Further investigation is required, but we propose that the establishment of an  
418 ectomycorrhizal symbiosis between an individual fungal mycelium and a host plant may  
419 increase both the likelihood that the fungus will successfully colonise other hosts of a similar  
420 genotype, and the efficiency of carbon transfer through its mycelium. In our system,  
421 *Rhizopogon* spp. and the ectomycorrhizal Pyronemataceae sp. are the most likely candidate  
422 fungi due to their abundance on seedling roots and their detected activity in soil  
423 metatranscriptomes. We further hypothesise that increased carbon transfer among kin  
424 seedlings may have resulted from (i) increased inter-root EMF biomass between compatible  
425 host genotypes (Rosado *et al.*, 1994a,b; Dudley *et al.*, 2013), and/or (ii) increased inter-root  
426 activity due to increased transfer of signalling compounds and/or micronutrients (Plett *et al.*,  
427 2011; Babikova *et al.*, 2013). Alternative explanations for the observed carbon transfer are

428 certainly possible, but would have to account for the significant interaction between relatedness  
429 and family (e.g., differences in belowground carbon allocation between families would not  
430 explain why increased transfer of  $^{13}\text{C}$  to roots in families A and B only occurred in the  
431 presence of kin).

432

#### 433 *Hyphal restriction*

434         Transfer of  $^{13}\text{C}$ -carbon to recipient pools occurred regardless of hyphal exclusion,  
435 demonstrating that  $^{13}\text{C}$ -compounds can be transferred in the absence of a direct linkage  
436 between mycorrhizal hyphae (e.g. Robinson & Fitter, 1999). Carbon transfer was clearly  
437 associated with mycorrhizas, since: (i) exclusion reduced transfer to bulk soil in some of the  
438 recipient families; (ii) donor and recipient EMF abundance was positively associated for  
439 several EMF species; (iii)  $D_{\text{ECM}}$  abundance was associated with enrichment of recipient pools;  
440 and (iv)  $D_{\text{ECM}}$  and  $R_{\text{ECM}}$  biomass contained proportionally more  $^{13}\text{C}$ -label than all other pools.  
441 Previous studies similarly reported small quantities of  $^{13}\text{C}$  transfer across a 0.5  $\mu\text{m}$  mesh in  
442 both ectomycorrhizal (Teste *et al.*, 2009; Philip *et al.*, 2010; Deslippe & Simard, 2011) and  
443 arbuscular mycorrhizal (Fitter *et al.*, 1998) systems. Thus, the mesh-bagging treatment may not  
444 be effective for preventing mycorrhizal-mediated carbon transfer. This is potentially due to: (i)  
445 recipient EMF hyphae scavenging donor  $^{13}\text{C}$ -exudates that diffused through the mesh  
446 (Robinson & Fitter, 1999; Johnson & Gilbert, 2015), (ii) hyphae fusing across the mesh, or (iii)  
447 hyphae breaching the mesh, possibly degrading it via secreted organic acids (Plassard &  
448 Fransson, 2009). No breaches were observed in our experiment, but consistent with  
449 possibilities (i) and (ii), there were regions of mesh with adjacent patches of hyphae on either  
450 side. Future experiments could employ the in-growth core rotation method (Johnson *et al.*,  
451 2001) to reduce the possibility of (ii) and (iii), although it is unlikely to prevent (i), which  
452 represents an alternative belowground transfer pathway in natural systems (Simard *et al.*  
453 1997b; Robinson & Fitter, 1999; Philip *et al.*, 2010; Deslippe & Simard, 2011).

454

#### 455 *Conclusions*

456         Our stable isotope approach successfully elucidated the pattern and scale of  
457 mycorrhiza-mediated carbon transfer between interior Douglas-fir seedlings, and the  
458 incorporation of enriched carbon into microbial biomass. EMF symbionts, specifically  
459 *Rhizopogon* spp. and Pyronemataceae sp., were the primary external beneficiaries of host-  
460 derived photosynthate, and were able to take it up despite the presence of a hyphae-restricting  
461 mesh. The small quantities of carbon transferred between seedlings suggest that it is unlikely



462 to be an important nutritional source, although the timing and transfer of micronutrients or  
463 signalling compounds may have a substantial ecological impact. We report evidence that  
464 relatedness influences carbon transfer between donor and recipient plants, and that the presence  
465 of this effect varied between families, raising the possibility of a mosaic of relatedness effects  
466 at larger scales. These findings require further exploration in the field, however, the  
467 implications for forest ecology are substantial.

468

#### 469 **Conflict of interest statement**

470 The authors declare no conflict of interest.

471

#### 472 **Acknowledgements**

473 We thank Melissa Dergousoff and Scott Besse for sustained lab assistance, and Matthew  
474 Zustovic, Aaron Zwiebel, Hilary Leung, and Julia Maddison for lab and greenhouse  
475 contributions; Alice Chang provided advice on EA-IRMS analysis; Spencer Reitenbach  
476 provided the seeds used in this experiment. We also thank Ian Dickie and three anonymous  
477 referees for their comments on an earlier version of this manuscript. This project was funded by  
478 an NSERC Strategic Project grant, with individual support provided to A.S.H., R.W., and  
479 A.K.A., by NSERC and UBC scholarships.

480

#### 481 **Author contributions**

482 A.K.A. and S.W.S. planned the seedling kin-recognition experiment, which A.K.A. performed.  
483 B.J.P., R.W., and S.W.S. designed the SIP experiment, with B.J.P. analysing the excess  $^{13}\text{C}$  data  
484 and implementing all experimental models, and R.W. analysing the  $\delta^{13}\text{C}$  data. R.W. and  
485 W.W.M. planned the analysis of  $^{13}\text{C}$ -labelled PLFA, which R.W. performed and analysed.  
486 B.J.P. and A.K.A. collected the EMF colonisation data, which B.J.P. analysed. A.S.H. generated  
487 the metatranscriptome data, which A.S.H. and R.W. analysed. A.K.A., B.J.P., R.W., and A.S.H.  
488 performed the experiments. B.J.P. led the writing of the manuscript with significant  
489 contributions from R.W. All authors reviewed and commented on the manuscript.

490

#### 491 **Supporting Information**

492 Additional supporting information may be found in the online version of this article.

493 **Supporting Methods S1.** EMF DNA extraction and Sanger sequencing details.

494 **Supporting Methods S2.** Metatranscriptome analysis extraction-sequencing details.

495 **Supporting Figure S1.** Boxplots of excess  $^{13}\text{C}$ ; a) biomass, b) proportion.

- 496 **Supporting Figure S2.** Incorporation of  $^{13}\text{C}$ -label by microbial groups.  
497 **Supporting Figure S3.** PLFA profiles of root and soil pools.  
498 **Supporting Figure S4.** Profiles of unidentified PLFAs in root and soil pools.  
499 **Supporting Figure S5.** Metatranscriptomic data based on Silva taxonomy.  
500 **Supporting Figure S6.** NMDS ordination of seedling root tip fungal communities.

501

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722



723 **Figure headings**

724 **Figure 1.** Illustration of the plant and soil carbon pools sampled in each experimental unit,  
 725 which consisted of (a.) one donor (D) and one full-sibling or unrelated recipient (R)  
 726 *Pseudotsuga menziesii* var. *glauca* seedling in one pot separated by a nylon mesh bag (dark grey  
 727 dashed cylinder) with holes of either 0.5  $\mu\text{m}$  or 35  $\mu\text{m}$ . In panels b-f, red and blue fill represents  
 728 the donor and recipient pools, respectively. **b.** Above-ground plant biomass, divided into donor  
 729 shoot pool ( $D_{\text{shoot}}$ ) and recipient shoot pool ( $R_{\text{shoot}}$ ). **c.** Below-ground plant biomass, divided into  
 730 donor root pool ( $D_{\text{root}}$ ) and recipient root pool ( $R_{\text{root}}$ ). **d.** Below-ground plant-fungal biomass,  
 731 divided into donor ectomycorrhizal root tip pool ( $D_{\text{ECM}}$ ) and recipient ectomycorrhizal root tip  
 732 pool ( $R_{\text{ECM}}$ ). **e.** Soil attached to roots following their removal from soil and light shaking,  
 733 divided into donor rhizosphere pool ( $D_{\text{rhizosphere}}$ ) and recipient rhizosphere pool ( $R_{\text{rhizosphere}}$ ); the  
 734 donor rhizoplane pool ( $D_{\text{rhizoplane}}$ ), very tightly adhering soil washed from root surface following  
 735 vigorous vortexing in sterile water and subsequently pelleted by centrifugation, is approximated  
 736 with solid black lines. **f.** Soil not attached to roots following their removal from soil and light  
 737 shaking, divided into donor coarse soil pool ( $D_{\text{coarse}}$ ) and recipient bulk soil pool ( $R_{\text{bulk}}$ ).

738 **Figure 2.** Relatedness and hyphal restriction effects on *Pseudotsuga menziesii* var. *glauca*  
 739 seedling  $R_{\text{root}}$  pool enrichment (excess  $^{13}\text{C}$  as  $^{12}\text{C}$ -equivalent), and their variation among  
 740 recipient families. **a-b,** Standardised regression coefficients for linear mixed models (refer to  
 741 Table 2), illustrating the differences between treatment levels in each recipient family in terms of  
 742 their standardised effect size (SES). **a.** difference between kin and non-kin within mesh  
 743 treatment (positive values indicate greater enrichment in kin pairs). **b.** difference between 35  
 744  $\mu\text{m}$  and 0.5  $\mu\text{m}$  mesh within relatedness treatment (positive values indicate greater enrichment in  
 745 pairs separated by 35  $\mu\text{m}$  mesh). Circles indicate average estimates, lines are 95% confidence  
 746 intervals. Filled circles indicate significant difference between treatment levels, open points  
 747 indicate no significant difference.

748 **Figure 3.** Linear regression of kin (left panels) and non-kin (right panels)  $^{13}\text{C}$ -enrichment in  
 749 *Pseudotsuga menziesii* var. *glauca* seedlings, revealing the positive relationships between: **a.**  
 750  $R_{\text{root}} \delta^{13}\text{C}$  and  $D_{\text{ECM}}$  abundance (kin significant), **b.**  $R_{\text{ECM}}$  and  $D_{\text{ECM}}$  excess  $^{13}\text{C}$  as  $^{12}\text{C}$ -equivalent  
 751 (mg) (kin significant), and **c.**  $R_{\text{ECM}}$  and  $D_{\text{shoot}}$  excess  $^{13}\text{C}$  as  $^{12}\text{C}$ -equivalent (mg) (kin significant).  
 752 Circles represent data points. Black lines indicate significant linear relationship, grey lines  
 753 indicate non-significant relationship. Note that kin and non-kin panels are scaled separately.

754 **Figure 4.** Incorporation of  $^{13}\text{C}$ -label into PLFAs by different taxonomic groups in each biomass  
 755 pool based on the average  $\delta^{13}\text{C}$  of PLFAs. Error bars correspond to standard error. Dotted red  
 756 line indicates natural abundance value of  $\delta^{13}\text{C}$ . Dashed vertical black line indicates separation

757 between donor and recipient pools by mesh bag. Text indicates whether biomass in the pool was  
758 primarily derived from plant material (*Pseudotsuga menziesii* var. *glauca*), plant and fungal  
759 material, the interface between plant/fungi and soil, or soil alone. PLFA identities are provided  
760 in Supporting Figures S2 and S3.  
761

Pickles et al.

762 **Table 1.** Partitioning of plant-assimilated  $^{13}\text{C}$ -labelled carbon in aboveground and belowground pools of interior Douglas-fir associated with  
 763 ectomycorrhizal fungi six days after the start of a 10-h  $^{13}\text{CO}_2$  labelling period

Pool	$^{13}\text{C}$ atom% excess (APE) APE (s.e.m.)	Total amount of excess $^{13}\text{C}$ ( $\mu\text{g plant}^{-1}$ )				Excess $^{13}\text{C}$ incorporated of the mean total fixed (%)			
		Total $\mu\text{g}$ (s.e.m.)	n	Kin $\mu\text{g}$ (s.e.m.)	n	Non-kin $\mu\text{g}$ (s.e.m.)	n	Kin % (s.e.m.)	Non-kin % (s.e.m.)
D <sub>shoot</sub>	0.439 (0.037)	7050.67 (325.77)	35	6858.59 (521.51)	16	7212.41 (418.62)	19	31.345 (2.383)	32.962 (1.913)
D <sub>root</sub>	0.238 (0.020)	1980.24 (159.76)	35	1878.00 (229.32)	17	2076.80 (226.67)	18	8.583 (1.048)	9.491 (1.036)
D <sub>ECM</sub>	0.979 (0.068)	136.21 (27.57)	31	156.72 (56.54)	14	119.31 (20.52)	17	0.716 (0.258)	0.545 (0.094)
D <sub>rhizoplane</sub>	0.076 (0.010)	231.81 (21.54)	36	241.61 (33.26)	17	223.04 (28.65)	19	1.104 (0.152)	1.019 (0.131)
D <sub>rhizosphere</sub>	0.035 (0.004)	167.07 (20.08)	35	122.88 (22.96)	16	204.29 (29.39)	19	0.562 (0.105)	0.934 (0.134)
D <sub>coarse</sub>	0.014 (0.002)	53.74 (11.29)	33	60.44 (18.81)	15	48.15 (13.69)	18	0.276 (0.086)	0.220 (0.063)
R <sub>bulk</sub>	0.001 (0.000)	3.96 (0.50)	27	5.35 (0.83)	13	2.68 (0.39)	14	0.024 (0.004)	0.012 (0.002)
R <sub>rhizosphere</sub>	0.004 (0.001)	4.90 (1.06)	33	4.80 (1.67)	16	4.98 (1.39)	19	0.022 (0.008)	0.023 (0.006)
R <sub>ECM</sub>	0.119 (0.021)	2.43 (0.48)	34	2.48 (0.86)	16	2.39 (0.52)	18	0.011 (0.004)	0.011 (0.002)
R <sub>root</sub>	0.013 (0.001)	3.76 (0.47)	33	4.96 (0.80)	16	2.64 (0.35)	17	0.023 (0.004)	0.012 (0.002)
R <sub>shoot</sub>	0.006 (0.001)	2.76 (1.11)	6	3.46 (1.59)	4	1.37 (0.59)	2	0.016 (0.007)	0.006 (0.003)
<b>Plant and soil mean values</b>									
D <sub>plant</sub>		9167.11		8893.31		9408.52		40.64	43.00
D <sub>soil</sub>		452.62		424.93		475.48		1.94	2.17
R <sub>soil</sub>		8.86		10.15		7.66		0.05	0.03
R <sub>plant</sub>		8.95		10.89		6.40		0.05	0.03

764 APE = atom percent excess; n = number of significantly enriched samples; Total amount added during pulse = 29 020  $\mu\text{g}$  excess  $^{13}\text{C}$ ; Mean  
 765 amount fixed after 1-day chase = 21 881.08  $\mu\text{g}$  excess  $^{13}\text{C}$

766 **Table 2.** Effects of hyphal restriction and donor relatedness treatments on excess  $^{13}\text{C}$  as  $^{12}\text{C}$  equivalent ( $\mu\text{g}$ ) of recipient *Pseudotsuga menziesii*  
 767 var. *glauca* seedling biomass pools, using maximum likelihood analysis of linear mixed-effects models<sup>1</sup> and likelihood ratio tests

Pool	$\mathbf{R}_{\text{bulk}}$			$\mathbf{R}_{\text{rhizosphere}}$			$\mathbf{R}_{\text{ECM}}$			$\mathbf{R}_{\text{root}}$		
Data transformation	-			sqrt			-			log <sub>10</sub>		
Marginal $R^2$	0.605			0.565			0.325			0.673		
Factor	df	F-value	P-value	df	F-value	P-value	df	F-value	P-value	df	F-value	P-value
Intercept	1,15	<b>97.53</b>	<b>&lt;0.0001</b>	1,15	<b>283.15</b>	<b>&lt;0.0001</b>	1,15	<b>50.41</b>	<b>&lt;0.0001</b>	1,15	<b>110.65</b>	<b>&lt;0.0001</b>
Relatedness (Re)	1,15	1.81	0.199	1,15	0.15	0.708	1,15	0.08	0.787	1,15	<b>6.47</b>	<b>0.020</b>
Hyphal (Hy)	1,15	3.65	0.077	1,15	3.29	0.090	1,15	0.79	0.387	1,15	<b>5.38</b>	<b>0.032</b>
Recipient Family (Rfam)	3,15	2.79	0.079	3,15	2.14	0.138	3,15	0.31	0.816	3,15	0.50	0.687
Re*Hy	1,15	0.14	0.710	1,15	0.31	0.585	1,15	1.82	0.197	1,15	2.34	0.143
Re*Rfam	3,15	0.37	0.775	3,15	0.61	0.619	3,15	1.55	0.243	3,15	<b>4.59</b>	<b>0.015</b>
Hy*Rfam	3,15	<b>3.36</b>	<b>0.050</b>	3,15	<b>3.54</b>	<b>0.041</b>	3,15	0.03	0.991	3,15	<b>4.14</b>	<b>0.021</b>

768 Values in bold are significant at the  $\alpha < 0.05$  level.

769 Marginal  $R^2$  = model fit based on fixed factors alone. Inclusion of random factor did not increase model fit.

770 <sup>1</sup>Model form: (Pool) ~ (Re)\*(Hy) + (Re)\*(Rfam) + (Hy)\*(Rfam), random ~ 1| (Donor family)

Pickles et al.

771 **Table 3.** Identity and relative abundance of root-associated fungi on kin and non-kin *Pseudotsuga menziesii* var. *glauca* seedlings based on root  
772 tip morphotyping and ITS region taxonomic classification. No significant differences in colonisation were observed between mesh sizes.

Fungal morphotype ID	Seedling	Relative abundance		Accession	Sequence match (NCBI accession)	% Identity
		Kin	Non-kin			
<i>Rhizopogon vinicolor</i> (Morphotype 1)	Donor	0.46	0.58	KT314836	<i>Rhizopogon vinicolor</i> (AF263933)	652/656 (99%)
	Recipient	0.44	0.57			
<i>Rhizopogon vinicolor</i> (Morphotype 2)	Donor	0.24	0.11	KT314840	<i>Rhizopogon vinicolor</i> (HQ385848)	529/535 (99%)
	Recipient	0.14	0.03			
Pyronemataceae sp.	Donor	0.12	0.09	KT314854	uncultured Pyronemataceae (GU452518)	524/524 (100%)
	Recipient	0.26	0.29			
<i>Wilcoxina</i> sp.	Donor	0.08	0.05	-	Sequencing failed – Taxonomic ID	-
	Recipient	0.13	0.09			
Uncolonised root tips	Donor	0.05	0.15	N/A	N/A	N/A
	Recipient	0.00	<0.01			
<i>Fusarium</i> sp.	Donor	0.03	0.01	KT314859	<i>Fusarium acuminatum</i> (KP068924)	478/478 (100%)
	Recipient	0.03	0.02	KT314860	<i>Fusarium oxysporum</i> (KP132221)	451/451 (100%)
<i>Rhizopogon</i> sp.	Donor	0.01	0.01	KT314850	<i>Rhizopogon fragrans</i> (AM085523)	619/621 (99%)
	Recipient	< 0.01	< 0.01			
<i>Rhizopogon ochraceisporus</i>	Donor	0.01	0.00	KT314851	<i>Rhizopogon ochraceisporus</i> (AF366389)	603/609 (99%)
	Recipient	0.00	0.00			

773 N/A = not applicable (uncolonised root tips which did not generate fungal DNA).

774

New Phytologist









