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Microbial degradation of dimethylsulfide and related C₁-sulfur compounds: organisms and pathways controlling fluxes of sulfur in the biosphere

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1 **Abstract**

2 **Dimethylsulfide (DMS) plays a major role in the global sulfur cycle. It has**
3 **important implications for atmospheric chemistry, climate regulation, and**
4 **sulfur transport from the marine to the atmospheric and terrestrial**
5 **environment. In addition, DMS acts as an info-chemical for a wide range of**
6 **organisms ranging from microorganisms to mammals. Microorganisms that**
7 **cycle DMS are widely distributed in a range of environments, for instance oxic**
8 **and anoxic marine, freshwater and terrestrial habitats. Despite the importance**
9 **of DMS that has been unearthed by many studies since the early 1970s, the**
10 **understanding of the biochemistry, genetics and ecology of DMS-degrading**
11 **microorganisms is still limited. This review examines current knowledge on the**
12 **microbial cycling of DMS and points out areas for future research that should**
13 **shed more light on the role of organisms degrading DMS and related**
14 **compounds in the biosphere.**

15

16 **DMS and related organic sulfur compounds**

17 Volatile sulfur compounds play an important role within the biogeochemical cycle of
18 sulfur. In being able to transfer from the liquid into the gas phase and vice versa,
19 reduced volatile sulfur compounds have particular importance for affecting the
20 composition and chemistry of the atmosphere. Although carbonyl sulfide (COS) has
21 the highest concentration of the reduced volatile sulfur compounds in the atmosphere,
22 dimethylsulfide (DMS) has the highest source strength (Watts 2000) and is thought of
23 as a climate cooling gas (Charlson *et al.* 1987). DMS is produced by a variety of
24 chemical and biological processes, both natural and man-made, and it is itself subject
25 to a wide variety of chemical and biological transformations in the environment.
26 Some aspects of the microbial metabolism of the related compounds (see Table 1)

27 dimethylsulfoniopropionate (DMSP), dimethylsulfoxide (DMSO), dimethylsulfone
28 (DMSO₂), methanethiol (MT) and methanesulfonic acid (MSA) are also considered
29 where appropriate as these occur as precursors and/or degradation products of DMS.
30

31 **Industrial roles of DMS and related compounds.**

32 From an anthropocentric point of view, DMS and the related compounds DMSO and
33 DMSO₂ are of particular interest in terms of their roles as flavour compounds and
34 their industrial applications. DMS is a colourless liquid with a boiling point of 41°C.
35 and has a disagreeable odour akin to that of rotting cabbage. In our daily lives it is
36 often present at low concentrations as an important flavour compound in a wide range
37 of foods, including raw and processed fruits and vegetables such as tomatoes,
38 sweetcorn, grapes, asparagus and brassicas (Bills and Keenan 1968; Buttery *et al.*
39 2002; Kubec *et al.* 1998; Miers 1966; Segurel *et al.* 2004; Ulrich *et al.* 2001; Wong
40 and Carson 1966), cheeses (McGugan 2002; Milo and Reineccius 1997) honey (de la
41 Fuente *et al.* 2007), and truffles (Talou *et al.* 1987), for instance. DMS is equally
42 important as a flavour compound in a variety of beverages including beers (Meilgaard
43 2002), wines (e.g. Segurel *et al.* 2004), orange and grapefruit juice (Shaw *et al.* 1980),
44 and is also found in roast coffee (Rhoades 2002) and processed milk (Keenan and
45 Lindsay 1968). DMS can be part of the essential aroma profile but also be of concern
46 as it can contribute to off-notes.

47 DMSO is a water-soluble polar organic solvent that is useful in a range of
48 industries, and is also relevant as a pharmaceutical drug delivery agent that can
49 facilitate the movement of various compounds across lipid membranes (Leake 1967).
50 Both DMSO and DMSO₂ are found in a wide range of foods including milk (Pearson
51 *et al.* 1981). Humans excrete 4-11mg of DMSO₂ per day via urine. Marketed as

52 methylsulfonylmethane, it is also a constituent of some dietary supplements (see
53 Parcell 2002 for a review).

54

55 **Environmental significance of DMS and related compounds**

56 The roles of C₁-sulfur compounds in an industrial and human context described above
57 are eclipsed by the major functions of these compounds in the environment, which
58 have stimulated a substantial body of research over the last three decades. Chemical
59 weathering of rock and the water solubility of sulfate lead to loss of sulfur from the
60 continents due to surface water runoff to the oceans. The oceans are rich in sulfur,
61 having a sulfate concentration of approximately 28mM. Emission of sulfur species
62 from the marine environment into the atmosphere, their atmospheric transport and
63 subsequent deposition by wet and dry deposition on the continents are thus an
64 important link in the sulfur cycle, affording sulfur transport from the oceans to the
65 continents (compare Figure 1). Prior to the work by Lovelock and colleagues it was
66 assumed that hydrogen sulfide was the volatile sulfur compound emitted into the
67 atmosphere that provided a precursor for sulfate aerosols in marine air (Saltzman and
68 Cooper 1989), however Lovelock and colleagues showed that dimethylsulfide was
69 much more abundant in the marine boundary layer than hydrogen sulfide (Lovelock *et*
70 *al.* 1972). Based on these findings it was realised that DMS provides a route for sulfur
71 transport between the oceans and the terrestrial environment (Nguyen *et al.* 1978). It
72 is now well established that DMS is the most abundant form of biogenic sulfur input
73 into the atmosphere; estimates range from 19 to 50 Tg of sulfur that are emitted as
74 DMS from the marine environment per annum (Andreae 1990), which translates to
75 around 200 million tons of sulfur, or roughly to 0.66 tons of sulfur emitted per km² of
76 ocean surface on average.

77

78 **Atmospheric oxidation of DMS and the CLAW hypothesis**

79 In the atmosphere, DMS is subject to chemical and photochemical oxidation resulting
80 in a range of organic and inorganic sulfur species, mainly sulfate, sulfur dioxide and
81 methanesulfonic acid (MSA) (Hatakeyama *et al.* 1982; Panter and Penzhorn 1980;
82 Pham *et al.* 1995), but DMSO and DMSO₂ are also formed (Harvey and Lang 1986;
83 Zhu *et al.* 2003), and DMSO has been detected in rain water (Kiene and Gerard 1994;
84 Ridgeway *et al.* 1992; Sciare *et al.* 1998). The atmospheric residence time of DMS is
85 short, only about a day, and the main atmospheric sinks are believed to be the daytime
86 oxidation with hydroxyl radicals and reaction with nitrate radicals during the night;
87 however, it appears that the reactions removing DMS and their rate constants are
88 complex and not yet well understood in detail (see Barnes *et al.* 2006 for a review).
89 As indicated above, the atmospheric transport and subsequent dry and wet deposition
90 of these sulfur compounds on the continents provide an important link in the global
91 sulfur cycle. In soils, atmospherically derived sulfur contributes to the pool of sulfur
92 available for assimilation as a plant nutrient, directly as sulfate, or indirectly after
93 microbial regeneration of sulfate from organic sulfur compounds such as MSA,
94 DMSO and DMSO₂ (see Kertesz 2000 for a review). The atmospheric oxidation
95 products of DMS form aerosol particles which have direct and indirect effects that
96 lead to negative temperature forcing of the Earth-atmosphere system, directly
97 reflecting solar radiation and indirectly by providing particles that can act as cloud
98 condensation nuclei (CCN) in the atmosphere. An increase in the number of CCN
99 facilitates the formation of clouds that have a higher number of relatively smaller
100 water droplets, thereby increasing the cloud albedo and decreasing the amount of solar
101 radiation to reach the Earth surface. Hence, atmospheric DMS has been linked to
102 climate regulation and is considered as a climate-cooling gas (Charlson *et al.* 1987).

103 Charlson and colleagues hypothesised that production of DMSP by phytoplankton in
104 the oceans was the basis of a geophysiological feedback loop that regulates global
105 climate, also known as the CLAW hypothesis according to the first letters of the
106 authors' surnames (Charlson *et al.* 1987). The CLAW hypothesis states that an
107 increase in solar irradiation and climate warming stimulates phytoplankton growth in
108 the oceans and leads to an increased production of DMSP in the surface ocean causing
109 a greater flux of DMS into the atmosphere. The associated increase of DMS-derived
110 aerosol particles in the atmosphere causes more solar radiation to be reflected, either
111 directly by aerosols or indirectly through intensified formation of high albedo clouds;
112 ultimately these consequences of DMS emission are predicted to cause a cooling of
113 the Earth's climate. Climate cooling and reduction of the amount of
114 photosynthetically active radiation reaching the ocean surface, due to increased
115 albedo, cause a decrease in phytoplankton growth and lead to a reduction of DMSP
116 production in the ocean, a concomitant decrease in DMS emission and therefore an
117 easing of the aforementioned negative temperature forcing; the phytoplankton
118 DMSP/DMS system is therefore suggested to form a negative feedback loop
119 (Charlson *et al.* 1987).

120 Vallina and Simó found that marine DMS concentrations are positively
121 correlated with solar radiation dose (Vallina and Simó 2007) which might lend
122 support to the CLAW hypothesis as an increase of solar radiation would be expected
123 to cause climate warming and increased DMS emission. Different approaches of
124 modelling the expected increase of marine DMS production under global warming
125 scenarios, however, have suggested only a modest 1-2% increase in DMS production,
126 which is much weaker than observable seasonal variations of DMS (Bopp *et al.* 2003;
127 Vallina *et al.* 2007). Nevertheless, studies have confirmed that DMS-derived aerosol

128 can be a significant source of CCN especially in the remote marine atmosphere that
129 receives little dust and aerosol from the continents (Ayers *et al.* 1991; Vallina *et al.*
130 2006), but the interactions and pathways in atmospheric DMS oxidation are complex
131 and not fully understood precluding quantitative modelling (Ayers *et al.* 1997). The
132 view that emissions of DMS from the marine environment have implications for
133 climate and atmospheric chemistry is widely supported, but there is as yet no
134 unambiguous evidence for the validity of the CLAW hypothesis.

135

136 **Sources of DMS**

137 *Marine environment*

138 Various estimates of the flux of DMS to the atmosphere have been made (range of 15-
139 109 Tg a⁻¹) but a review of the sources of DMS suggests to adopt a figure of
140 approximately 24.49 +/- 5.3 Tg a⁻¹ (Watts 2000). The strength of the marine
141 environment as a source of DMS has been estimated at around 21 Tg a⁻¹ and is
142 therefore by far the most important source totalling around 80% of the total DMS
143 flux, the remaining 20% originate from vegetation, salt marshes and estuaries, soils,
144 wetlands and also include anthropogenic sources (Watts 2000).

145 Dimethylsulfoniopropionate (DMSP) is the main source of DMS in the marine
146 environment. DMSP is a metabolite of certain species of macroalgae (Challenger and
147 Simpson 1948; Van Alstyne and Puglisi 2007) and phytoplankton, in particular in
148 dinoflagellates and in species such as the Haptophytes *Emilinia huxleyi* and
149 *Phaeocystis* (Liss *et al.* 1994; Malin and Kirst 1997). Algae can accumulate DMSP to
150 high internal concentrations reaching to hundreds of mM (reviewed in Stefels 2000;
151 Yoch 2002). Corals and their zooxanthellae also contain large amounts of DMSP (Hill
152 *et al.* 1995), which can be the source of high local DMS concentrations (approx 1µM)

153 in coral reefs, for instance in coral mucus ropes (Broadbent and Jones 2004). It has
154 been suggested that DMSP has a role as an osmolyte (Kiene *et al.* 2000; Stefels 2000),
155 an antifreeze compound (Kirst *et al.* 1991) or an antioxidant (Sunda *et al.* 2002), but
156 its exact role remains unresolved and it is possible that it serves distinct roles in
157 different organisms (Otte *et al.* 2004). Some vascular plants also contain DMSP, for
158 instance some halophytes of the genus *Spartina* and *Wollastonia biflora* contain
159 significant amounts of DMSP, and the molecule has also been detected in sugar cane
160 (see Otte *et al.* 2004 for a review).

161 Upon lysis of DMSP-containing organisms, for instance by viral attack (Malin
162 *et al.* 1998) or zooplankton grazing (Wolfe *et al.* 1994; Wolfe and Steinke 1996),
163 DMSP becomes dissolved in seawater. Microbial degradation of dissolved DMSP
164 occurs through a number of different pathways (Howard *et al.* 2006; Johnston *et al.*
165 2008) (compare Figure 2), and the majority of DMSP is not degraded to DMS
166 (González *et al.* 1999; Kiene *et al.* 2000; Moran *et al.* 2003; Yoch 2002). Until
167 recently, the enzyme cleaving DMSP was generally referred to as “DMSP lyase”, but
168 the exact mechanisms by which DMS is formed from DMSP had not been
169 investigated in any detail. Using genetic analysis of bacteria that form DMS from
170 DMSP, Johnston and coworkers have described three different pathways of DMSP-
171 dependent DMS formation that involve enzymes that are members of different
172 enzyme families (Curson *et al.* 2008; Johnston *et al.* 2008; Todd *et al.* 2009; Todd *et*
173 *al.* 2007).

174 Dissolved DMSP and/or DMS derived from it has been shown to be a
175 powerful signalling molecule that attracts certain bacteria, e.g. chemotaxis by
176 *Silicibacter* TM1040 (Miller *et al.* 2004), but also affects the swimming (copepods,
177 harbour seals, coral reef fish) and flying (petrels, shearwaters) behaviour of a range of

178 organisms presumably as a foraging cue (see review by Johnston *et al.* 2008). A role
179 of DMS as an info-chemical is also indicated by studies demonstrating that it allowed
180 dogs and pigs to detect truffles in soil (Talou *et al.* 1990) and a study that showed the
181 importance of volatile organic sulfur compounds including DMS, DMDS and
182 dimethyltrisulfide (DMTS) in the “*bouquet of death*” that attracted burying beetles to
183 carcasses of mice (Kalinová *et al.* 2009).

184 The majority of DMS emission is from open ocean environments, but microbial
185 mats and intertidal sediments are also important sources of DMS (Stuedler and
186 Peterson 1984). Several studies have investigated the cycling of DMS and related
187 compounds in such ecosystems (Jonkers *et al.* 1998; Kiene 1988; 1990; Kiene and
188 Capone 1988; Lymio *et al.* 2009; Visscher *et al.* 2003; Visscher *et al.* 1991). DMSP-
189 producing plants and macroalgae, e.g. the salt marsh cord grass *Spartina alterniflora*
190 or the green algal seaweed *Ulva* spp., can contribute to the production of DMS in such
191 ecosystems (Kiene and Capone 1988 and references therein), however, other
192 pathways of DMS formation may be more important in anoxic sediments, including
193 reduction of DMSO, metabolism of sulfur-containing amino acids, and methylation of
194 sulfide (Jonkers *et al.* 1996; Kiene and Capone 1988; Lomans *et al.* 1997; Visscher *et*
195 *al.* 2003; Visscher *et al.* 1991).

196

197 *Terrestrial sources of DMS*

198 DMS formation also occurs in terrestrial and freshwater environments, and, with
199 exceptions (see below), DMS formation in these environments is not due to DMSP
200 degradation. As noted above for coastal sediments, the processes involved are
201 respiratory reduction of DMSO (Zinder and Brock 1978c), degradation of sulfur-
202 containing amino acids (Kadota and Ishida 1972; Kiene and Capone 1988), and

203 anaerobic degradation of methoxylated aromatic compounds (Bak *et al.* 1992;
204 Lomans *et al.* 2001). Methylation of sulfide in aerobic microorganisms due to the
205 action of thiol-S methyltransferase has been demonstrated and predominantly gives
206 rise to MT (Drotar *et al.* 1987).

207 Overall, the emission of DMS from terrestrial and freshwater sources has not
208 been studied as intensively as that from the marine environment and as yet there is not
209 a clear view of the relative importance of different production mechanisms. Wetland
210 emission rates of volatile sulfur compounds, including DMS, were subject to diel
211 variations and an influence of plant communities was noted; in most wetlands,
212 emission rates were insignificant compared to those measured in intertidal sediments
213 dominated by *Spartina* (Cooper *et al.* 1989). *Sphagnum*-dominated peat bogs were
214 shown to evolve both DMS and MT, the formation of both compounds was
215 biological, and methylation of MT was the main source of DMS (Kiene and Hines
216 1995). Soils may also emit volatile organic sulfur compounds including DMS and
217 fluxes can be enhanced by waterlogging (Banwart and Bremner 1976), but soils are
218 not considered a major source of atmospheric sulfur (Andreae 1990; Watts 2000) and
219 volatilisation of sulfur compounds is not thought to contribute significantly to loss of
220 sulfur from soils (Banwart and Bremner 1976). Recently, DMS formation and
221 degradation was observed in deeper layers (mainly below 1 m depth) along the profile
222 of an agricultural soil in Australia. The so-called agricultural sulfate soil investigated
223 in that study is in close proximity to a tidal inlet, may receive sporadic inputs of
224 seawater and thus is characterised by relatively high sulfate concentrations. It was
225 suggested that DMS might be a potential source of the SO₂ emissions that have been
226 observed from this type of soils (Kinsela *et al.* 2007). The decomposition of plant
227 residues in soil, especially those of crucifer species with a high content of sulfur-

228 containing glucosinolates, can generate a number of volatile sulfur compounds
229 including DMS, MT and DMDS (Lewis and Papvizas 1970). Such locally enhanced
230 production of volatile sulfur compounds after amendment of soils with crucifer
231 residues can be exploited in controlling soil borne phytopathogenic fungi. The bio-
232 fumigant effects of crucifer tissue amendment were suggested to be mainly due to
233 isothiocyanates (e.g. Gamliel and Stapleton 1993) with additional contribution by
234 less toxic volatile sulfur species such as DMDS (Bending and Lincoln 1999). A
235 combination of isothiocyanates and DMS was potent in inhibiting the activity of soil
236 nitrifying bacteria (Bending and Lincoln 2000).

237 In freshwater environments, DMS and MT production may occur in anoxic
238 regions of stratified lakes and their sediments, as a result of sulfide methylation and/or
239 degradation of methoxylated aromatic compounds (Fritz and Bachofen 2000; Lomans
240 *et al.* 2001b; Lomans *et al.* 1997; Richards *et al.* 1991), but production of DMS has
241 also been detected in oxic layers of freshwater lakes (e.g. Richards *et al.* 1991). DMS
242 in oxic freshwater lakes may be derived from phytoplankton and DMS release by
243 phytoplankton cultures was stimulated by methionine (Caron and Kramer 1994).
244 Although DMSP is not generally considered a major DMS precursor in freshwater
245 environments, DMS production in Lake Kinneret (Israel) appeared to be due to
246 blooms of the DMSP-containing freshwater dinoflagellate *Peridinium gatunense*
247 (Ginzburg *et al.* 1998). In a study of freshwater river sediments, Yoch and colleagues
248 found that DMS was produced upon addition of DMSP to sediment slurries and
249 identified DMS-producing Gram-positive bacteria (Yoch *et al.* 2001), demonstrating
250 that the genetic potential for DMSP degradation was present far away from the marine
251 environment, although it was noted by the authors that the enzyme systems
252 responsible for DMS production could have cognate substrates other than DMSP.

253

254 *DMS production by plants*

255 Plants may be the main source of DMS in the terrestrial environment with a source
256 strength estimated at 3.2 Tg a^{-1} , of which half is thought to be derived from tropical
257 forests (Watts 2000). Plants emit a range of volatile sulfur compounds including H_2S ,
258 DMS, MT, COS, and CS_2 , with H_2S and DMS usually the dominant species, but
259 emission rates are variable and dependent on many factors (reviewed by Schröder
260 1993). In a study of environmental conditions that affect volatile sulfur emissions
261 from plants, Fall and coworkers (Fall *et al.* 1988) showed that DMS was the dominant
262 sulfur compound emitted by a range of crops including corn, alfalfa and wheat. Sulfur
263 fluxes were positively correlated with temperature and light intensity but were
264 independent of the pCO_2 (Fall *et al.* 1988). A similar correlation of DMS emission
265 rates and temperatures was found in a study of the gas exchange of DMS and COS of
266 trees, but DMS emission was not a universal feature across the tree species tested and
267 it was concluded that the contribution of tree-derived DMS to the global sulfur budget
268 is negligible in temperate regions (Geng and Mu 2006).

269

270 *Anthropogenic sources of DMS*

271 In an industrial context, DMS and other reduced sulfur compounds such as
272 methanethiol, dimethyldisulfide (DMDS) and hydrogen sulfide are products in the
273 wood pulping process, e.g. in the paper industry, and can occur in significant amounts
274 in liquors of the so-called Kraft process where it is a byproduct of the Swern oxidation
275 of alcohols to aldehydes (Omura and Swern 1978). The food and brewing industry,
276 agriculture and animal farming are also responsible for DMS emissions (Kim *et al.*
277 2007; Rappert and Müller 2005). Anthropogenic sources of DMS are thought to be

278 responsible for less than 1% of the total sources, but the emission of volatile sulfur
279 compounds can be significant at the local scale. Due to the low odour thresholds of
280 volatile organic sulfur compounds these can be a cause of nuisance odours (Zhu *et al.*
281 2002), for instance from wastewater treatment of paper manufacture (Catalan *et al.*
282 2008), or in the treatment of other sewage with high DMSO concentrations, caused by
283 reduction of DMSO to DMS under anaerobic conditions (Glindemann *et al.* 2006).
284 Industrial operations providing composts for mushroom production (Derikx *et al.*
285 1990; Noble *et al.* 2001), field spreading of manure and application of biosolids, as
286 well as livestock operations are further DMS sources linked to the agriculture and
287 farming industries (Rappert and Müller 2005). DMS is also emitted from landfills, but
288 is less abundant than hydrogen sulfide (Kim *et al.* 2005).

289

290

291 **Sinks for DMS and related compounds**

292 Given the role ascribed to DMS in affecting atmospheric chemistry and climate, it is
293 of interest to understand the factors that control the flux of DMS to the atmosphere. In
294 surface seawater the DMS concentration is determined by the rate of production
295 (mainly) from DMSP, and a variety of loss terms. Sea-to-air transport is dependent on
296 hydrological and meteorological parameters, for instance wind speed (Liss and
297 Merlivat 1986) and wave action (Watson *et al.* 1991). DMS is also photochemically
298 oxidised in surface water to DMSO (Brimblecombe and Shooter 1986). Although
299 large quantities of DMS are produced in the upper mixed layer of the oceans, only a
300 small fraction of DMS escapes to the atmosphere, while the majority (estimated at
301 ~90%) is degraded in the mixed surface layer due to microbial processes, including its
302 use as either a carbon or sulfur source, or its biological degradation to DMSO (Archer

303 *et al.* 2002; Hatton *et al.* 2004; Kiene and Bates 1990). The microorganisms and the
304 microbial metabolism of DMS are discussed below.

305

306 *DMS-degrading microorganisms*

307 The first insights into the microbiology of DMS-degrading organisms were obtained
308 by studies of *Thiobacillus* and *Hyphomicrobium* species beginning in the 1970s with
309 the isolation of *Thiobacillus* strains from a pine bark biofilter that was used to remove
310 odorous compounds such as H₂S, MT, DMS and DMDS from effluents of a paper
311 pulp factory in Finland where these compounds were produced from methoxy groups
312 of lignin in the paper pulping process (Sivelä and Sundman 1975). Further
313 *Thiobacillus* species and isolates of *Hyphomicrobium* were then obtained that grew on
314 DMS as sole carbon source (De Bont *et al.* 1981; Kanagawa and Kelly 1986; Pol *et*
315 *al.* 1994; Smith and Kelly 1988; Suylen and Kuenen 1986). A diverse range of
316 microorganisms able to degrade DMS has since been isolated from a wide variety of
317 environments, including soils, plant rhizospheres, activated sludge, biofiltration
318 operations, seawater, cultures of marine algae, marine and freshwater sediments,
319 microbial mats and also humans from which DMS degraders have been isolated from
320 feet and mouth samples. Table 2 lists species that have been shown to grow at the
321 expense of DMS, while Figure 3 illustrates the identity of DMS-degrading organisms
322 in a phylogenetic context for representative strains with known 16S rRNA genes.

323

324 *Microbial metabolism of DMS*

325 There are numerous biological pathways that contribute to DMS degradation in the
326 environment; in principal these serve (i) the utilisation of DMS as a carbon and
327 energy source, (ii) its oxidation to DMSO by phototrophic or heterotrophic organisms,

328 (iii) and its utilisation as a sulfur source. Various types of DMS degradation pathways
329 have been reported in the literature, some of these featuring MT and/or H₂S as
330 intermediates, while other pathways do not give rise to volatile sulfur compounds.
331 The scheme in Figure 4 provides an overview of the conversions of DMS and related
332 C₁-sulfur compounds that occur in a wide range of different organisms. Details of
333 specific biochemical conversions of DMS and microorganisms carrying them out are
334 presented below.

335

336 *Utilisation of DMS as a carbon and energy source for bacterial growth.*

337 Utilisation of DMS as a carbon and energy source is thought to occur by one of two
338 pathways that have been suggested which contain either a DMS monooxygenase (De
339 Bont *et al.* 1981) or a presumed methyltransferase (Visscher and Taylor 1993b)
340 carrying out the initial oxidation of DMS. It has been suggested that the
341 methyltransferase is inhibited by chloroform while the DMS monooxygenase was
342 suggested to be inhibited by methyl-tert butyl ether (Visscher and Taylor 1993b).

343

344 **DMS monooxygenase pathway.** The work by De Bont and colleagues suggested that
345 DMS metabolism in *Hyphomicrobium* S involved an initial NAD(P)H dependent step
346 of DMS oxidation by a DMS monooxygenase (DMO), yielding formaldehyde and
347 methanethiol (De Bont *et al.* 1981). DMO has also been suggested to be responsible
348 for initial DMS degradation in some *Thiobacillus* strains (Visscher and Taylor
349 1993b). Formaldehyde is either directly assimilated into biomass or further oxidised
350 via formate to CO₂ in order to provide reducing power. Assimilation of the
351 formaldehyde produced during DMS and MT degradation in methylotrophic bacteria
352 is accomplished by the serine or ribulose monophosphate cycles (e.g. Anthony 1982;

353 De Bont *et al.* 1981), while in DMS-degrading autotrophs that have been analysed
354 formaldehyde is oxidised to CO₂, part of which is then assimilated into biomass via
355 the Calvin-Benson-Bassham cycle (Kelly and Baker 1990). Methanethiol produced by
356 DMS monooxygenase in the first step is degraded by MT oxidase to formaldehyde,
357 hydrogen peroxide and sulfide (Gould and Kanagawa 1992; Suylen *et al.* 1987).
358 Formaldehyde is again either assimilated directly into biomass or oxidised to CO₂
359 while sulfide is converted to sulfite either by methanethiol oxidase (in the case of
360 *Hyphomicrobium* spp.) or sulfide oxygenase (in case of *Thiobacillus* spp.) which is
361 then oxidised to sulfate (via sulfite oxidase). Hydrogen peroxide is reduced to water
362 and oxygen by catalase and the growth on DMS of organisms utilising MT oxidase is
363 usually inhibited by the catalase inhibitor 3-amino-1,2,4-triazole.

364 The biochemistry and genetic basis of DMS and methanethiol degradation in
365 these isolates has remained largely uncharacterised, although methanethiol oxidase
366 was purified from several species including *Hyphomicrobium* EG (Suylen *et al.*
367 1987), *Thiobacillus thioparus* Tk-m (Gould and Kanagawa 1992), and *Rhodococcus*
368 *rhodochrous* (Kim *et al.* 2000). MT oxidase from *Hyphomicrobium* strain EG (Suylen
369 *et al.* 1987) was reported not to require any co-factors for activity. The insensitivity of
370 this MT oxidase towards the metal-chelating agents EDTA and neocuproine
371 suggested that the enzyme did not contain metal ions or haem co-factors. It was
372 suggested that the native *Hyphomicrobium* enzyme was a monomer with a molecular
373 weight of 40-50 kDa, but MT oxidase from *Thiobacillus thioparus* sp. Tk-m (Gould
374 and Kanagawa 1992) appeared to be a monomer of 29-40 kDa. Two more recent
375 studies reported purification of MT oxidase from *Rhodococcus rhodochrous* (Kim *et*
376 *al.* 2000) and a reassessment of the MT oxidase from *Thiobacillus thioparus* Tk-m
377 (Lee *et al.* 2002), giving molecular weights for these enzymes of ~61 kDa. It is not

378 clear whether different forms of methanethiol oxidase with different molecular weight
379 may exist; in any case there is still a considerable lack of understanding of the
380 biochemistry of methanethiol oxidation in bacteria.

381 Although the activity of DMS monooxygenase in methylotrophs and autotrophs
382 degrading DMS under aerobic conditions was reported in a number of studies (Anesti
383 *et al.* 2005; Anesti *et al.* 2004; Borodina *et al.* 2000; De Bont *et al.* 1981; Moosvi *et*
384 *al.* 2005), further information about the enzyme has not been forthcoming as it
385 appeared to be unstable and no purification has been achieved. No genes encoding a
386 DMS monooxygenase have been identified.

387

388 **Methyltransferase pathway.** *Thiobacillus* ASN-1 used an alternative initial step of
389 DMS degradation which was independent of oxygen and which was suggested to be
390 due to a methyltransferase (Visscher and Taylor 1993b). It was suggested that the
391 methyl group was transferred to an acceptor molecule and then further oxidised via
392 folate-bound intermediates. The methyl accepting factor was suggested to be
393 cobalamin related although it was not identified (Visscher and Taylor 1993a; b).
394 Further oxidation of the remaining methanethiol appeared to follow the same scheme
395 as in the DMS monooxygenase pathway described above.

396

397 **DMSO₂ and DMSO oxidation via DMS.** In the initial study of *Hyphomicrobium* X
398 by De Bont and colleagues (De Bont *et al.* 1981) one of the substrates for growth of
399 the strain was DMSO, which was reduced to DMS and thus fed into the DMS
400 monooxygenase pathway. Subsequently, it was shown that DMSO₂ could also be
401 degraded by some methylotrophs via DMS, as enzyme activities for DMSO₂
402 reductase, DMSO reductase and DMS monooxygenase were detected in cell-free

403 extracts of *Hyphomicrobium sulfonivorans* and *Arthrobacter sulfonivorans* growing
404 on these compounds (Borodina *et al.* 2000; Borodina *et al.* 2002).

405

406 *Growth on DMS under anoxic conditions*

407 Several bacterial and archaeal strains able to degrade DMS and MT under anoxic
408 conditions have been isolated (Finster *et al.* 1992; Kiene *et al.* 1986; Lomans *et al.*
409 1999b; Lyimo *et al.* 2000; Ni and Boone 1991; Tanimoto and Bak 1994; Visscher and
410 Taylor 1993a). The thermodynamic aspects of growth of SRB and methanogens on
411 methylated sulfur compounds have been reviewed in detail elsewhere (Scholten *et al.*
412 2003). SRB and methanogens are thought to be responsible for anaerobic DMS
413 oxidation in anoxic sediments of coastal salt marshes, estuaries, and freshwater
414 sediments (Kiene and Capone 1988; Kiene *et al.* 1986; Lomans *et al.* 1999a; Zinder
415 and Brock 1978b), but the degradation of DMS has also been reported with nitrate as
416 electron acceptor (Haaijer *et al.* 2008; Tanimoto and Bak 1994; Visscher and Taylor
417 1993a). The characteristics of methanogenic Archaea growing on DMS and MT have
418 been reviewed previously, isolates belonged to the genera *Methanolobus*,
419 *Methanomethylovorans*, *Methanosarcina* and *Methanosalsus* (Lomans *et al.* 2002).
420 Compared to methanogens, relatively few SRB growing on DMS have been isolated.
421 Tanimoto and Bak (1994) obtained Gram positive, spore-forming SRB from
422 thermophilic fermenter sludge which they classified as *Desulfotomaculum* species.
423 These isolates were also able to grow on DMS using nitrate as electron acceptor
424 (Tanimoto and Bak 1994). Based on slurry incubations with tungstate and
425 bromoethanesulfonate addition to selectively inhibit SRB and methanogens,
426 respectively, Lyimo and coworkers found that the degradation of DMS and MT in
427 anoxic mangrove sediments was dominated by SRB (Lyimo *et al.* 2009). A strain was

428 isolated, the first SRB from a marine environment, which was closely related to
429 *Desulfosarcina* sp. and exhibited very slow growth rates on DMS, but which had a
430 high affinity for DMS. The authors concluded that due to the extremely slow growth
431 observed, such SRB might be outcompeted by methanogens in enrichments and slurry
432 incubations when relatively high DMS concentrations are used since methane
433 production increased exponentially during slurry incubations.
434 The biochemical and genetic basis of DMS degradation in SRB remains
435 uncharacterised. More data are available for methanogens. It was shown that during
436 growth on acetate of the methanogen *Methanosarcina barkeri* the cells also converted
437 DMS and methylmercaptopropionate (MMPA) to methane and a corrinoid protein
438 functioned as a co-enzyme M methylase capable of DMS and MMPA degradation
439 (Tallant and Krzycki 1997). Fused corrinoid/methyl transfer proteins have been
440 implicated in methyl sulphide metabolism in *Methanosarcina acetivorans*
441 (Oelgeschlaeger and Rother 2009).

442

443 *Oxidation of DMS to DMSO*

444 In phototrophic bacteria, the oxidation of DMS to DMSO can be used to provide
445 electron donors for carbon dioxide fixation as suggested by a study of DMS
446 degradation by a culture of an anoxygenic phototrophic purple sulfur bacterium that
447 converted DMS stoichiometrically to DMSO (Zeyer *et al.* 1987). Similarly, DMS can
448 be utilised by certain phototrophic green sulfur bacteria when growing on reduced
449 sulfur compounds such as thiosulfate and hydrogen sulfide (Vogt *et al.* 1997).

450 DMS to DMSO conversion by heterotrophic bacteria was first described by
451 Zhang *et al.* (Zhang *et al.* 1991) in *Pseudomonas acidovorans* DMR-11 (reclassified as
452 *Delftia acidovorans*). In this strain DMSO was stoichiometrically formed from DMS

453 as a product of co-oxidation during heterotrophic metabolism, for instance during
454 growth on a range of organic compounds, but no carbon from DMS was assimilated.
455 DMS removal in cell free extracts of strain DMR-11 was dependent on the presence
456 of NADPH, which could not be replaced by NADH. Complete conversion of DMS to
457 DMSO was also shown in the marine heterotrophic bacterium *Sagittula stellata* E-37
458 (González *et al.* 1997) in cells grown on glucose, irrespective of additional organic
459 carbon being added during the assay. The enzymes responsible for the conversion of
460 DMS to DMSO in both *Sagittula stellata* and *Delftia acidovorans* are unknown.

461

462 **DMS dehydrogenase.** The biochemistry and genetics of DMS to DMSO oxidation in
463 phototrophic metabolism in which DMS serves as an H donor have been studied in
464 detail in *Rhodovulum sulfidophilum* (Hanlon *et al.* 1996; McDevitt *et al.* 2002). In this
465 strain, DMS-dependent DMSO formation is mediated by DMS dehydrogenase
466 (DMSDH), a heterotrimeric enzyme comprising three subunits (DdhABC) in which a
467 molybdopterin co-factor is bound to the A subunit (Hanlon *et al.* 1996). The enzyme
468 is encoded by the *ddh* operon containing the genes *ddhABCD*, which encode the A, B
469 (containing putative [Fe-S] clusters) and C (containing a *b*-type haem) subunits, and
470 *ddhD* is thought to encode a polypeptide that could be responsible for the maturation
471 of the molybdopterin-containing enzyme (McDevitt *et al.* 2002).

472

473 **Oxidation of DMS to DMSO by methanotrophs and nitrifying bacteria.** DMS
474 oxidation has also been observed in resting cell suspensions of methane-grown
475 methanotrophic isolates of *Methylomicrobium* (Fuse *et al.* 1998; Sorokin *et al.* 2000)
476 and in *Methylomicrobium pelagicum* the product was identified as DMSO. The
477 nitrifying bacteria *Nitrosomonas europaea* and *Nitrosococcus oceani* (Juliette *et al.*

478 1993) also converted DMS to DMSO and some evidence suggests that ammonia
479 monooxygenase (AMO) is the enzyme co-oxidising DMS to DMSO in these bacteria.
480 While the co-oxidation of MT by purified methane monooxygenase (MMO), the key
481 enzyme in aerobic methanotrophic bacteria, has been reported (Colby *et al.* 1977), it
482 is still unclear whether DMS is co-oxidised by MMO, although this seems likely
483 given the close evolutionary relationship of AMO and particulate MMO (Holmes *et*
484 *al.* 1995).

485

486 **Reduction of DMSO to DMS by DMSO reductase.** A range of microorganisms can
487 couple the oxidation of organic carbon compounds to respiratory reduction of DMSO
488 to DMS under anoxic conditions (Zinder and Brock 1978a). The enzyme
489 dimethylsulfoxide reductase, which reduces DMSO to DMS, was first purified and
490 characterised from *Rhodobacter sphaeroides*. In this strain, it is a soluble periplasmic
491 single subunit enzyme of 82 kDa that contains a molybdopterin co-factor (Sato and
492 Kurihara 1987), which can also reduce trimethylamine oxide (Styrvold and Strom
493 1984). It is encoded by the gene *dmsA* (Yamamoto *et al.* 1995). A similar enzyme was
494 purified from *Rhodobacter capsulatus* (McEwan *et al.* 1991). The DMSO reductase in
495 *E. coli* is rather different. It is a heterotrimeric enzyme expressed under anaerobic
496 conditions, which is anchored in the periplasmic membrane. It is encoded by the
497 operon *dmsABC* (Bilous *et al.* 1988), in which the genes encode the active catalytic
498 subunit DmsA (82 kDa) that contains the molybdopterin co-factor, an electron
499 transfer protein DmsB (23.6 kDa) and a membrane anchor DmsC (22.7 kDa)
500 (Sambasivarao *et al.* 1990). Despite the differences in enzyme structure, the catalytic
501 subunits of *R. sphaeroides* and *E. coli* share 29% sequence identity at the amino acid
502 level (Yamamoto *et al.* 1995). In *Hyphomicrobium sulfonivorans* a membrane-bound

503 DMSO reductase that reduced DMSO to DMS was expressed during aerobic growth
504 on DMSO₂, thus not having a role in anaerobic respiration under these conditions.
505 Only a weak cross-reaction was reported for the immunoblotting of *H. sulfonivorans*
506 membrane fraction with an antibody against the *R. capsulatus* enzyme (Borodina *et*
507 *al.* 2002). The observation that DMSO reductase activity was present in the
508 membrane fraction would suggest that it might be similar to the *E. coli* type DMSO
509 reductase, but that it is regulated differently to the *E. coli* enzyme.

510 DMSO reductase may carry out the reverse reaction in which DMS is reduced
511 to DMS, so it might be a candidate for DMS degradation in the environment.
512 However, although the enzyme from *R. capsulatus* can carry out the reverse reaction
513 *in vitro*, its K_s for DMS is high (1 mM) and DMSO strongly inhibits this reaction
514 (Adams *et al.* 1999), so it would appear unlikely to be relevant under physiological
515 conditions. The *E. coli* enzyme is expressed constitutively under anaerobic conditions
516 (Weiner *et al.* 1992). Overall, at this point there is little support to suggest that DMSO
517 reductases could provide a route of DMS degradation in, for instance, the oxic mixed
518 surface layer of the oceans.

519

520 *Assimilation of C₁ sulfur compounds as a sulfur source*

521 In addition to serving as substrate for growth of aerobic and anaerobic
522 microorganisms, DMSO and DMS can also be used as a source of sulfur. A strain of
523 *Marinobacter* was able to utilise DMS as a sulfur source with the aid of light,
524 probably using a flavoprotein (Fuse *et al.* 2000). *Pseudomonas aeruginosa* can grow
525 with methanesulfonate as a sole sulfur source, using the flavin-linked
526 methanesulfonate monooxygenase MsuED (Kertesz *et al.* 1999) that is repressed by
527 sulfide, sulfite and sulfate. It is closely related to the alkanesulfonate monooxygenase

528 (SsuED) that is induced during the sulfate-starvation response in *E. coli* (Eichhorn *et*
529 *al.* 1999). Bacterial sulfur assimilation by these enzymes has been reviewed in detail
530 (Kertesz 2000). In a strain of *Acinetobacter*, DMS degradation via DMSO which led
531 to the assimilation of sulfur was observed. The enzyme oxidising DMS to DMSO was
532 related to multi-component monooxygenases oxidising toluene and similar substrates.
533 It was termed DMS monooxygenase by the authors (Horinouchi *et al.* 1997), but this
534 is inappropriate as the degradation of DMS by this enzyme does not generate MT and
535 formaldehyde. Similarly *Rhodococcus* strain SY1 utilised DMS, DMSO and DMSO₂
536 as sulfur sources and in both strains the sequence of oxidation started with DMS
537 oxidation to DMSO which was oxidised to DMSO₂ and further to MSA (Omori *et al.*
538 1995). Work on *Pseudomonas putida* DS1 suggested the latter was then a substrate
539 for a SsuED type enzyme (Endoh *et al.* 2003).

540

541 *MSA catabolism*

542 A different kind of methanesulfonate monooxygenase exists in methylotrophic
543 bacteria such as *Methylosulfonomonas methylovora* which can grow on MSA as a sole
544 source of carbon and energy (Kelly and Murrell 1999). Its MSA monooxygenase is
545 composed of four distinct polypeptides. The hydroxylase subunit was composed of a
546 48 and 20 kDa subunits making up a native protein of around 210 kDa of a α_3/β_3
547 structure. Further components were identified as a ferredoxin (32 kDa) and a
548 reductase (38 kDa). The enzyme subunits are encoded by the genes *msmABCD* (De
549 Marco *et al.* 1999) and the closely linked *msmEFGH* operon encodes proteins
550 involved in transport of MSA (Jamshad *et al.* 2006). Transcriptional analysis showed
551 that *msmEFGH* operon was expressed constitutively while *msmABCD* was induced by
552 MSA (Jamshad *et al.* 2006).

553

554 **Ecology of microorganisms degrading DMS and related compounds**

555 Early studies suggested that microorganisms catabolising DMS mainly belonged to
556 the genera *Hyphomicrobium*, and *Thiobacillus*, additional isolation studies have
557 significantly extended the range of organisms able to grow on DMS (Table 1). In
558 addition to the shortcomings of microbial community analyses by cultivation-
559 dependent approaches, there are particular difficulties that are often encountered in
560 isolation of DMS-degrading bacteria (e.g. Smith and Kelly 1988; Suylen and Kuenen
561 1986). The diversity of cultivable DMS-oxidising bacteria still precludes delineation
562 of major patterns in their distribution. It is almost certain that the true extent of the
563 phylogenetic diversity of DMS-degrading organisms has not yet been identified,
564 either because organisms are recalcitrant to culturing conditions or due to the capacity
565 to degrade DMS being a phenotypic trait that is only rarely tested, even in studies of
566 methylotrophic bacteria. This is most likely due to the low attraction of working with
567 this smelly compound. The ability to degrade DMS is usually not conserved among
568 closely related species, i.e. there is no perfect correlation of phylotype and phenotype.
569 This largely negates the direct application of the widely used cultivation-independent
570 ribosomal RNA approach for studying DMS degrading microbial populations in the
571 environment. Nevertheless some investigations on relevant environments, using 16S
572 rRNA genes as markers have shown the presence of microbial populations that might
573 degrade DMS, based on their relatedness to known DMS-degrading strains. For
574 example, bacteria were found in marine DMS enrichment cultures (Vila-Costa *et al.*
575 2006) that were related to marine DMS degrading *Methylophaga* isolates (Schäfer
576 2007). Also, populations of related bacteria were detected in stable isotope probing
577 experiments with ¹³C-DMS following a DMSP-producing phytoplankton bloom of

578 *Emiliania huxleyi* in the English Channel (Neufeld *et al.* 2008). Further application of
579 SIP will allow improved definition of the phylogenetic diversity of DMS-degrading
580 microbial populations in environmental samples, but the approach can only detect
581 those organisms that assimilate the carbon from DMS. Additional tools that target key
582 enzymes of DMS metabolism will therefore be required to map the diversity and
583 activity of DMS degrading microorganisms. This will require new insights into the
584 metabolism of DMS at a molecular level including studying the biochemistry and
585 genetics of suitable model organisms in order to obtain a detailed understanding of the
586 enzymes and genes underpinning DMS degradation across a range of isolates.
587 Molecular methods targeting functional genes of DMS metabolism will not only allow
588 the elucidation of patterns in the distribution of DMS-degrading microorganisms in
589 nature independent of cultivation, but will also highlight particular microbial
590 populations for targeted isolation. Studying environmentally relevant model
591 organisms in more detail should also be useful in delineating the physiological
592 response of DMS degrading microorganisms and their potential to degrade DMS
593 under varying environmental conditions. Many of the known DMS-degrading bacteria
594 (compare Table 2) are able to grow on a range of substrates. DMS-degrading
595 *Methylophaga* species, for instance, also grow on methanol and methylated amines
596 (De Zwart *et al.* 1996; Schäfer 2007), two compounds which are present in the marine
597 environment in concentrations as high as 50-250nM in the case of methanol in the
598 tropical Atlantic (Williams *et al.* 2004). These concentrations are similar to or exceed
599 those of DMS which are typically in the low nanomolar range (Kettle *et al.* 1999).
600 Being presented with more than one growth substrate may have important effects and
601 the physiological and transcriptional responses of DMS-degrading organisms under
602 such conditions require further study.

603

604 **Interactions of DMS-degrading microorganisms and plants**

605 The focus of most research on the synthesis and catabolism of DMS has been on the
606 marine system. There is some evidence for production of DMS and other volatile
607 sulfur species by plants, but there are few data on emissions from vegetation in
608 temperate and boreal regions (Watts 2000). The association with plants of microbial
609 populations degrading DMS and related compounds is therefore of particular interest
610 for future study. Aboveground interactions of plants and bacteria occur in the
611 phyllosphere, which is the site of volatile sulfur emission. Previously, it was shown
612 that plants harbour diverse populations of epiphytic and endophytic
613 *Methylobacterium* species (e.g. Abanda-Nkpwatt *et al.* 2006; Knief *et al.* 2008),
614 which are thought to thrive on methanol released from pectin metabolism in the cell
615 wall (Galbally and Kirstine 2002). Similarly, it might be expected that DMS emission
616 from leaves could help to sustain populations able to degrade this substrate. Such
617 phyllosphere populations would likely affect the net flux of DMS and other volatile
618 sulfur compounds emitted from plants. Whatever the function is of volatile sulfur
619 release by plants, organisms degrading these compounds have the potential to affect
620 the functioning of the biological systems that might rely on volatile compounds.
621 Emission of volatile sulfur has been suggested as a route for removal of excess sulfur
622 (see review of Rennenberg 1984) or toxic HS⁻ ions (Saini *et al.* 1995). A recent report
623 suggests a role for H₂S emission as a plant defence signal in the context of sulfur
624 induced resistance of crops (Papenbrock *et al.* 2007). As a major volatile sulfur
625 species emitted by plants, DMS may have a role that has to be determined as yet.

626 There is also potential for interactions between plants and C₁-sulfur compound
627 degrading microorganisms belowground. The activity of soil microbial populations

628 involved in cycling of organic sulfur compounds is of particular importance for
629 contributing to soil fertility as the preferred sulfur source of plants is sulfate, but the
630 majority of sulfur in soils is bound in organic form (Kertesz and Mirleau 2004).
631 Recent improvements with respect to anthropogenic emissions of sulfur from fossil
632 fuel combustion have lead to a reduction in man-made sulfate aerosols in the
633 atmosphere and to a concomitant decrease in the rate of deposition of atmospheric
634 sulfur (Irwin *et al.* 2002). In some areas, the decrease in atmospheric S deposition is
635 leading to increasing incidences of sulfur deficiency for a range of agricultural crops,
636 such as oilseed rape (Schnug *et al.* 1995). Evidence for a decline of “natural” sulfur
637 fertilisation of soils derived from atmospheric sulfur due to fossil fuel combustion is
638 provided by changes of the sulfur isotope ratio in wheat straw (Zhao *et al.* 2003).
639 Consideration of future SO₂ emission rates (McGrath and Zhao 1995) or future
640 climate scenarios indicates that the potential for sulfur starvation in crops is likely to
641 increase (Hartmann *et al.* 2008) with important consequences for agricultural
642 productivity. Previous research has demonstrated that bacterial organosulfur
643 compound degrading populations in the rhizosphere play an important role in
644 regenerating sulfate for uptake by crop-plants for instance, but work has so far
645 focussed on the utilisation of alkane- and arylsulfonates and –sulfates as sulfur
646 sources for bacteria (Kertesz and Mirleau 2004; Schmalenberger *et al.* 2008;
647 Schmalenberger *et al.* 2009). Further work is needed to fully appreciate the role of
648 microbial populations degrading C₁-sulfur compounds such as DMSO, DMSO₂ and
649 MSA, and the utilisation of these compounds as both sulfur and carbon sources in the
650 rhizosphere needs to be investigated. The potential importance of DMSO₂ and DMSO
651 degrading methylotrophs in the rhizosphere of plants has been demonstrated by the
652 work of Borodina *et al.* (Borodina *et al.* 2000; Borodina *et al.* 2002).

653

654 **Outlook**

655 DMS-degrading microorganisms are widely distributed in the environment, but there
656 is still a lack of insight into their phylogenetic and functional diversity. The
657 development and application of functional gene probes and stable isotope probing
658 experiments will allow to decipher patterns in the distribution of DMS degrading
659 microorganisms in nature. Functional genetic markers based on key enzymes of DMS
660 metabolism and that of related compounds will also allow to investigate in more detail
661 the role of DMS degrading organisms in controlling fluxes of volatile sulfur to the
662 atmosphere and will help to assess their contribution to metabolising organically
663 bound sulfur and returning inorganic sulfur back to the environment. Clearly, the
664 emission of DMS from the marine environment is controlled significantly by the
665 activity of microorganisms. Microbial DMS metabolism affects the flux of DMS to
666 the atmosphere and thus the composition of the atmosphere and global climate,
667 therefore, the activity of marine microbial DMS-degrading microorganisms is
668 ultimately also an important factor that influences the amount of sulfur transported to
669 the continents where it affects the levels of sulfur in soils. Establishing the
670 phylogenetic affiliation of DMS degrading organisms in the environment and
671 identification of the pathways used by microbial populations to remove DMS from the
672 water column will help to identify the environmental regulation of marine microbial
673 DMS oxidation. This will contribute to gaining a better understanding of the complex
674 microbial processes involved in controlling the flux of sulfur from the oceans into the
675 atmosphere and should be useful to improve the prospects of modelling marine DMS
676 emissions under future climatic scenarios.

677

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683

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- 1311

1312 Table 1. DMS and related organic sulfur compounds

Compound	Formula
Dimethylsulfide (DMS)	$(\text{CH}_3)_2\text{-S}$
Dimethylsulfonio-propionic acid (DMSP)	$(\text{CH}_3)_2\text{-S-CH}_2\text{-CH}_2\text{-COOH}$
Dimethylsulfoxide (DMSO)	$(\text{CH}_3)_2\text{-SO}$
Dimethylsulfone (DMSO₂)	$(\text{CH}_3)_2\text{-SO}_2$
Methanethiol (MT)	$\text{CH}_3\text{-SH}$
Dimethyldisulfide (DMDS)	$\text{CH}_3\text{-S-S-CH}_3$
Methanesulfonic acid (MSA)	$\text{CH}_3\text{-SO}_3\text{H}$

1313 Table 2. Bacterial isolates capable of growth on DMS as a sole source of carbon and energy.

Species	Strain	[DMS] _{MAX}	Isolated from	Isolation substrate (concentration)	Reference
<i>Klebsiella pneumoniae</i> ¹	ATCC 9621	N.D.	Unknown ²	Unknown	(Rammler and Zafferoni 1967)
<i>Thiobacillus</i> sp.	MS1	2.4mM	<i>Pinus</i> sp. bark biofilter from a cellulose mill.	DMS (1.6mM)	(Sivelä and Sundman 1975)
<i>Hyphomicrobium</i> sp.	S	N.D.	Soil (Wageningen, Netherlands)	DMSO (12.8mM)	(De Bont <i>et al.</i> 1981)
<i>Thiobacillus thioparus</i>	Tk-m	2mM	Activated sludge	Thiometon (6mM)	(Kanagawa <i>et al.</i> 1982; Kanagawa and Kelly 1986)
<i>Hyphomicrobium</i> sp.	EG	0.1mM	Papermill biofilter	DMSO (10mM)	(Suylen and Kuenen 1986)
<i>Thiobacillus</i> sp.	E1	2mM	Commercial peat	DMS (2mM)	(Smith 1987)
<i>Thiobacillus</i> sp.	E3	2mM	Garden compost	DMS (2mM)	(Smith 1987)
<i>Thiobacillus</i> sp.	E4	2mM	Cattle manure	DMS (2mM)	(Smith 1987)
<i>Thiobacillus</i> sp.	E5	2mM	Marine mud (Plymouth, UK)	DMS (2mM)	(Smith 1987)
<i>Thiobacillus</i> sp.	E7	2mM	<i>Sphagnum</i> sp. moss from a deodorisation unit	DMDS (2mM)	(Smith 1987)
<i>Thiobacillus thioparus</i>	E6	2mM	Pond water (Coventry, UK)	DMDS (2mM)	(Smith and Kelly 1988)
<i>Hyphomicrobium</i> sp.	I55	N.D.	Peat biofilter	DMS (1mM)	(Zhang <i>et al.</i> 1991)
<i>Thiobacillus thioparus</i>	DW44	N.D.	Peat biofilter	Thiosulfate (20mM)	(Cho <i>et al.</i> 1991)
<i>Thiobacillus</i> sp.	K4	N.D.	Biofilter	CS ₂	(Plas <i>et al.</i> 1991)
<i>Thiobacillus</i> sp.	T5	1.3mM	Marine microbial mat (Texel, Netherlands)	Thiosulfate (10mM)	(Visscher <i>et al.</i> 1991)

¹ “*Aerobacter aerogenes*”.² Isolation details of this strain do not appear in the literature.

N.D. not determined

<i>Thiobacillus</i> sp.	ANS-1	N.D.	Tidal sediment (Georgia, USA)	DMS (0.5mM)	(Visscher and Taylor 1993b)
<i>Hyphomicrobium</i> sp.	VS	1mM	Activated sludge	DMS (15µM)	(Pol <i>et al.</i> 1994)
<i>Desulfotomaculum</i> sp.	TDS2	N.D.	Thermophilic fermenter sludge	DMS (5mM) and 10mM sulfate)	(Tanimoto and Bak 1994)
<i>Desulfotomaculum</i> sp.	SDN4	N.D.	Thermophilic fermenter sludge	DMS (5mM) and 5mM nitrate	(Tanimoto and Bak 1994)
<i>Methylophaga sulfidovorans</i>	RB-1	2.4mM	Marine microbial mat (Texel, Netherlands)	DMS (1.5mM)	(de Zwart <i>et al.</i> 1996)
<i>Hyphomicrobium</i> sp.	MS3	N.D.	Garden soil (Ghent, Belgium)	DMS/DMDS (1.4/1.1mM)	(Smet <i>et al.</i> 1996)
<i>Xanthobacter tagetidis</i>	TagT2C	2.5mM	<i>Tagetes patula</i> rhizosphere	T2C(2.5mM)	(Padden <i>et al.</i> 1997)
<i>Pseudonocardia asaccharolytica</i>	580	N.D.	Animal rendering plant biofilter	DMDS (1mM)	(Reichert <i>et al.</i> 1998)
<i>Pseudonocardia sulfidoxydans</i>	592	N.D.	Animal rendering plant biofilter	DMS (0.5mM)	(Reichert <i>et al.</i> 1998)
<i>Starkeya novella</i> ³	SRM	N.D.	Sewage (Kwangju, South Korea)	Thiosulfate (63mM)	(Cha <i>et al.</i> 1999)
<i>Thiocapsa roseopersicina</i>	M11	1mM	Marine microbial mat (Mellum, Germany)	Sulfide (1.6mM)	(Jonkers <i>et al.</i> 1999)
<i>Methylobacterium podarium</i>	FM1	N.D.	<i>Homo sapiens</i> foot	MMA (20mM)	(Vohra 2000)
<i>Hyphomicrobium sulfonivorans</i>	S1	N.D.	Garden soil (Warwickshire, UK)	DMSO ₂ (10mM)	(Borodina <i>et al.</i> 2002)
<i>Arthrobacter sulfonivorans</i>	ALL/A	N.D.	<i>Allium aflatumense</i> rhizosphere	DMSO ₂ (10mM)	(Borodina <i>et al.</i> 2002)
<i>Arthrobacter sulfonivorans</i>	ALL/B	N.D.	<i>Allium aflatumense</i> rhizosphere	DMSO ₂ (10mM)	(Borodina <i>et al.</i> 2002)
<i>Arthrobacter methylophilus</i>	TGA	N.D.	<i>Tagetes minuta</i> rhizosphere	DMSO ₂	(Borodina <i>et al.</i> 2002)

³ “*Thiobacillus novellus*”.

				(10mM)	
<i>Methylobacterium podarium</i>	FM4	1mM	<i>Homo sapiens</i> foot	MMA (20mM)	(Anesti <i>et al.</i> 2004)
<i>Hyphomicrobium sulfonivorans</i>	CT	N.D.	<i>Homo sapiens</i> teeth	DMS (1mM)	(Anesti <i>et al.</i> 2005)
<i>Hyphomicrobium sulfonivorans</i>	DTg	N.D.	<i>Homo sapiens</i> tongue	DMS (1mM)	(Anesti <i>et al.</i> 2005)
<i>Methylobacterium thiocyanatum</i>	MM4	N.D.	<i>Homo sapiens</i> tongue	MMA (20mM)	(Anesti <i>et al.</i> 2005)
<i>Methylobacterium extorquens</i>	MM9	N.D.	<i>Homo sapiens</i> tongue	Methionine (5mM)	(Anesti <i>et al.</i> 2005)
<i>Methylobacterium</i> sp.	MM10	N.D.	<i>Homo sapiens</i> tongue	Cysteine (5mM)	(Anesti <i>et al.</i> 2005)
<i>Micrococcus luteus</i>	MM7	N.D.	<i>Homo sapiens</i> teeth	MMA (20mM)	(Anesti <i>et al.</i> 2005)
<i>Bacillus licheniformis</i>	3S(b)	N.D.	<i>Homo sapiens</i> gingivae	DMS (1mM)	(Anesti <i>et al.</i> 2005)
<i>Bacillus licheniformis</i>	2Tgb	N.D.	<i>Homo sapiens</i> tongue	DMS (1mM)	(Anesti <i>et al.</i> 2005)
<i>Brevibacterium casei</i>	3Tg	N.D.	<i>Homo sapiens</i> tongue	DMS (1mM)	(Anesti <i>et al.</i> 2005)
<i>Brevibacterium casei</i>	3S(a)	N.D.	<i>Homo sapiens</i> gingivae	DMS (1mM)	(Anesti <i>et al.</i> 2005)
<i>Mycobacterium fluoranthenivorans</i>	DSQ3	N.D.	River sediment (London, UK)	DMA (10mM)	(Boden 2005; Boden <i>et al.</i> 2008)
<i>Methylophaga</i> sp.	DMS001	N.D.	<i>Emiliana huxleyi</i> culture	DMS (50µM)	(Schäfer 2007)
<i>Methylophaga</i> sp.	DMS002	N.D.	<i>Emiliana huxleyi</i> culture	DMS (50µM)	(Schäfer 2007)
<i>Methylophaga</i> sp.	DMS003	N.D.	<i>Emiliana huxleyi</i> culture	DMS (50µM)	(Schäfer 2007)
<i>Methylophaga</i> sp.	DMS004	N.D.	<i>Emiliana huxleyi</i> culture	DMS (50µM)	(Schäfer 2007)
<i>Methylophaga</i> sp.	DMS007	N.D.	<i>Emiliana huxleyi</i> culture	DMS (50µM)	(Schäfer 2007)
<i>Methylophaga</i> sp.	DMS009	N.D.	<i>Emiliana huxleyi</i> culture	DMS (50µM)	(Schäfer 2007)
“ <i>Methylophaga thiooxidans</i> ” ⁴	DMS010	N.D.	<i>Emiliana huxleyi</i> culture	DMS (50µM)	(Schäfer 2007)
<i>Methylophaga</i> sp.	DMS011	N.D.	<i>Emiliana huxleyi</i> culture	DMS (50µM)	(Schäfer 2007)
<i>Methylophaga</i> sp.	DMS021	N.D.	Rock pool water (Coral Beach, UK)	DMS (50µM)	(Schäfer 2007)
<i>Methylophaga</i> sp.	DMS026	N.D.	Sea water (English channel)	DMS (50µM)	(Schäfer 2007)
<i>Methylophaga</i> sp.	DMS039	N.D.	Sea water (Achmelvich, UK)	DMS (50µM)	(Schäfer 2007)
<i>Methylophaga</i> sp.	DMS040	N.D.	Sea water (Achmelvich, UK)	DMS (50µM)	(Schäfer 2007)
<i>Methylophaga</i> sp.	DMS043	N.D.	Sea water (Achmelvich, UK)	DMS (50µM)	(Schäfer 2007)

⁴ “*Methylophaga* sp. DMS010”.

<i>Methylophaga</i> sp.	DMS044	<i>N.D.</i>	Sea water (Achmelvich, UK)	DMS (50µM)	(Schäfer 2007)
<i>Methylophaga</i> sp.	DMS048	<i>N.D.</i>	Rock pool water (Coral Beach, UK)	Formate (10mM)	(Schäfer 2007)
<i>Methylophaga aminisulfidovorans</i> ⁵	MP*	<i>N.D.</i>	Sea water (Mokpo, South Korea)	Methanol (220mM)	(Kim <i>et al.</i> 2007)
<i>Hyphomicrobium facile</i>	-	<i>N.D.</i>	Marsh sediment (De Bruuk, Netherlands)	DMS (50µM)	(Haaijer <i>et al.</i> 2008)
<i>Microbacterium</i> sp.	NTUT26	<i>N.D.</i>	Wastewater sludge from a wood pulp factory (Taiwan)	DMS (1.6mM)	(Shu and Chen 2009)
<i>Desulfosarcina</i> sp.	SD1	<i>N.D.</i>	Mangrove sediment (Tanzania)	DMS (initially 20µM, additions rising to 100 µM)	(Lyimo <i>et al.</i> 2009)

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⁵ “*Methylophaga aminisulfidovorans*”

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Table 3. Bacterial isolates that are capable of oxidising DMS to DMSO

Species	Strain	DMS oxidation product	Isolated from	Isolation substrate	Metabolism producing DMSO	Reference
<i>Thiocystis</i>	A	DMSO	Salt Pond (MA, USA)	Sulfide	Anoxygenic photosynthesis	Zeyer et al. 1987
<i>Delftia acidovorans</i>	DMR-11	DMSO	Peat biofilter	Peptone	Anaerobic chemoheterotrophy	Zhang et al. (1991)
<i>Nitrosomonas europaea</i>		DMSO				Juliette et al.
<i>Methylomicrobium pelagicum</i>	NI	DMSO	Seawater (Japan)	Methane	Aerobic methane oxidation	Fuse 1998
<i>Sagittula stellata</i>	E-37	DMSO	Seawater enrichment culture on high molecular weight fraction of pulp mill effluent	Yeast extract/tryptone	Aerobic heterotrophic growth	Gonzalez et al. 1997
<i>Rhodovulum sulfidophilum</i>	SH1	DMSO	Seawater	Bicarbonate	Anoxygenic phototrophic growth	Hanlon et al. 1994
<i>Acinetobacter sp.</i>	20B	DMSO	Soil (Japan)	Succinate		Horinouchi et al. 1997
<i>Pseudomonas fluorescens</i>	76	DMSO	Unknown	Unknown	Heterotrophic growth	Ito et al. 2007
<i>Thiocapsa roseopersicina</i>	M1	DMSO	Marine microbial mat (Mellum, Germany)	Sulfide	Phototrophic growth	Visscher and van Gemerden 1991

1319 **Figure captions**

1320

1321 **Figure 1.** Simplified scheme of the major pathways of DMS production and
1322 transformation in the marine environment. DMS emission into the atmosphere is a
1323 source of heat-reflecting aerosols that can serve as cloud condensation nuclei and
1324 thereby affect the radiative balance of the Earth, thus linking DMS production to
1325 climate regulation. Atmospheric transport of DMS and its oxidation products and
1326 deposition in the terrestrial environment provides an important link in the global
1327 sulfur cycle. The role of microbes as sinks for DMS is discussed in the text.

1328

1329 **Figure 2.** Major pathways of dimethylsulfoniopropionate (DMSP) degradation.
1330 DMSP can be demethylated to methylmercaptopropionic acid (MMPA), which can be
1331 either demethylated to mercaptopropionic acid (MPA) or demethiolated to acrylate.
1332 The pathway leading to DMS from DMSP is also known as the ‘cleavage’ pathway,
1333 the responsible enzymes have been referred to as DMSP-lyases, but are in fact
1334 enzymes belonging to different protein families and exhibit different activities. These
1335 give rise to acrylate or 3-hydroxypropionate (3-HP). DMS can be oxidised by
1336 methyltransferases or DMS monooxygenases to methanethiol, or is oxidised to
1337 DMSO, for instance by DMS dehydrogenase. Refer to text for references.

1338

1339 **Figure 3.** Phylogenetic tree depicting the genetic diversity of bacterial isolates
1340 capable of assimilating carbon from DMS (overlaid in pink) or degrading DMS to
1341 DMSO (green). The tree is based on an alignment of small subunit ribosomal RNA
1342 gene sequences and was derived using the Neighbour joining option in MEGA4.
1343 Bootstrap values are of 100 replicates.

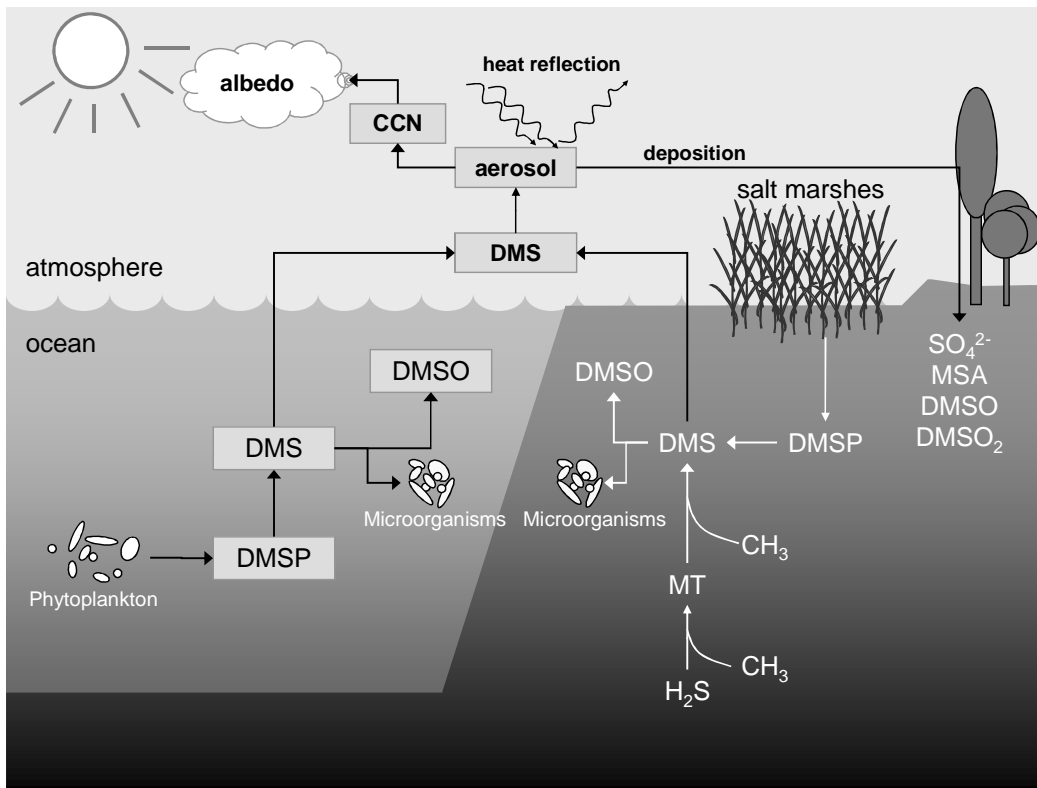
1344

1345 **Figure 4.** Scheme showing the biochemical and chemical interconversions of C₁-
1346 sulfur compounds and key intermediates in carbon and sulfur metabolism that have
1347 been observed across a wide range of microorganisms, refer to Table 1 for chemical
1348 formulae of the C₁-sulfur compounds. Either the enzymes/processes are given or an
1349 organism in which the conversion has been observed are given as an example, for
1350 further detail refer to text. 1, MSA monooxygenase; 2, FMNH₂-dependent DMSO₂
1351 monooxygenase (Endoh et al. 2005); 3, DMSO₂ dehydrogenase, 4, *Rhodococcus* SY1
1352 (Omori et al. 1995); 5 DMSO reductase; 6, DMS dehydrogenase; 7, DMS
1353 monooxygenase / DMS methyltransferase; 8 methylation of MT; 9, chemical
1354 oxidation of MT to DMDS; 10, DMDS reductase (Smith and Kelly 1988); 11, MT
1355 oxidase; 12, bacterial inorganic sulfur oxidation pathways; 13, sulfite oxidase; 14,
1356 formaldehyde oxidation (various enzymes); 15, formate dehydrogenase; 16, Calvin-
1357 Benson-Bassham cycle; 17, serine cycle or ribulose monophosphate cycle.

1358

1359 Figure 1

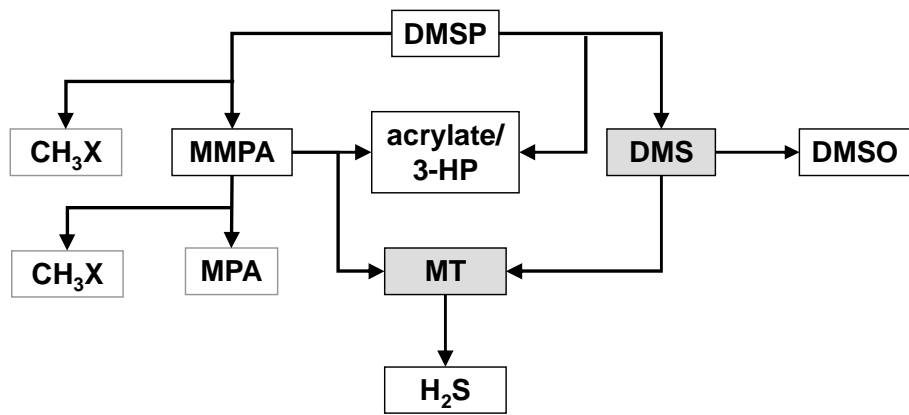
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1363 Figure 2

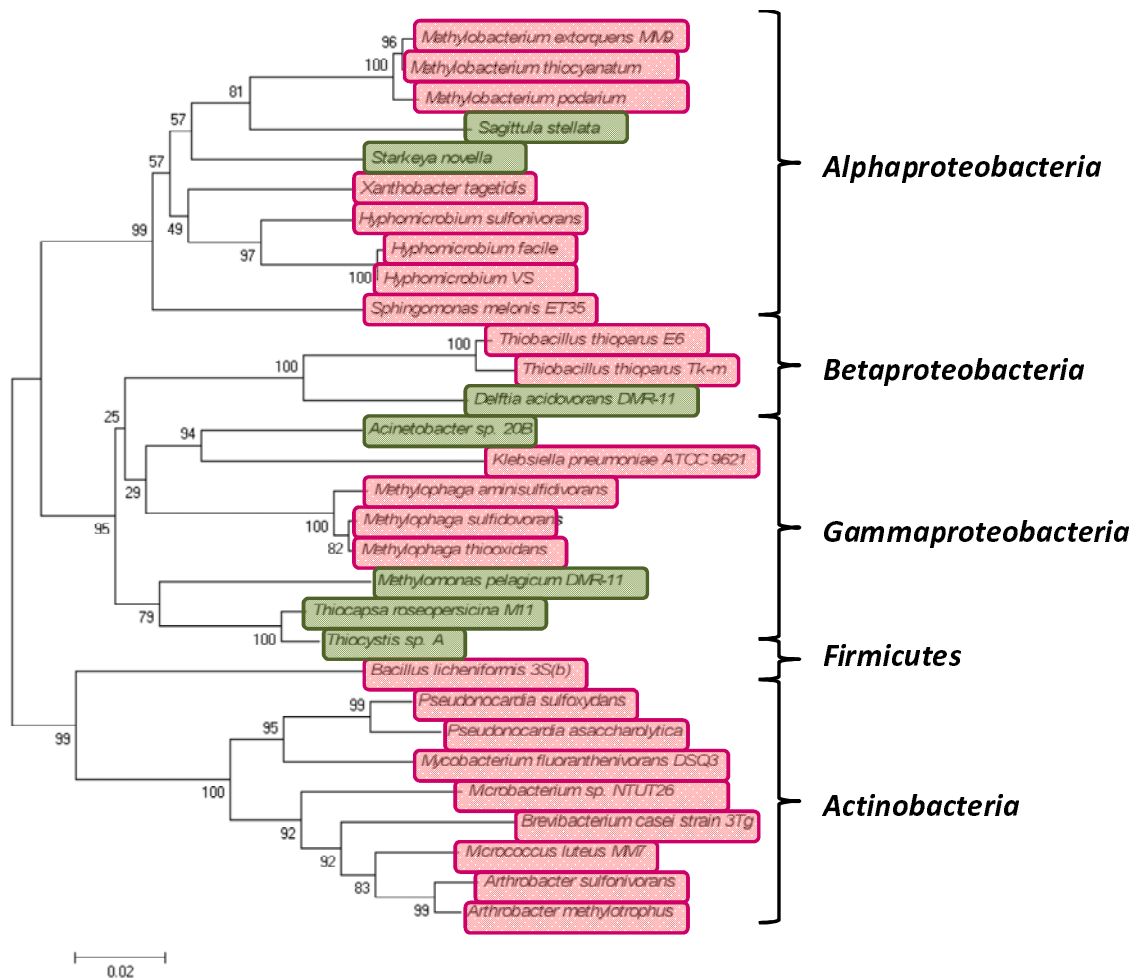


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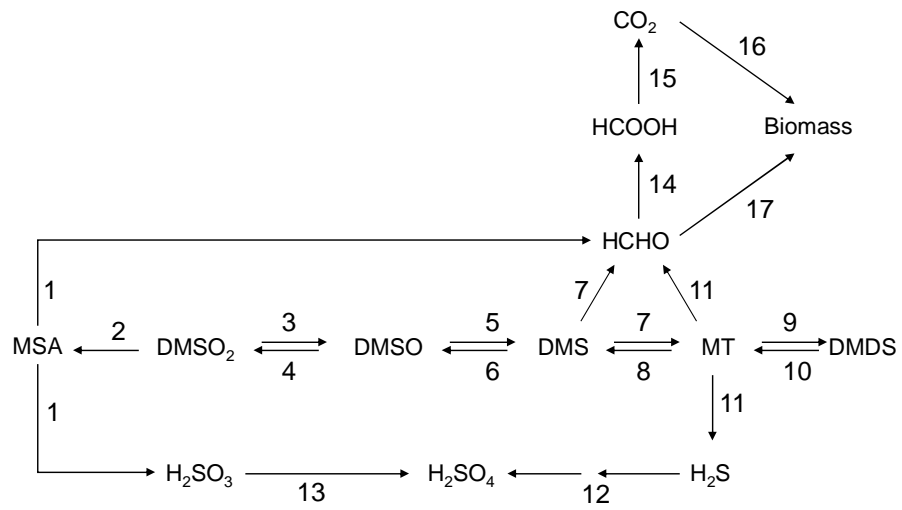
1367 Figure 3.



1368

1369

1370 Figure 4



1371