



Pedotransfer functions for Irish soils – estimation of bulk density (ρ_b) per horizon type

B. Reidy, I. Simo, P. Sills, and R. E. Creamer

Teagasc, Johnstown Castle, Environment Research Centre, Co. Wexford, Ireland

Correspondence to: B. Reidy (brianj.reidy@yahoo.ie)

Received: 20 September 2015 – Published in SOIL Discuss.: 9 October 2015

Accepted: 28 December 2015 – Published: 18 January 2016

Abstract. Soil bulk density is a key property in defining soil characteristics. It describes the packing structure of the soil and is also essential for the measurement of soil carbon stock and nutrient assessment. In many older surveys this property was neglected and in many modern surveys this property is omitted due to cost both in laboratory and labour and in cases where the core method cannot be applied. To overcome these oversights pedotransfer functions are applied using other known soil properties to estimate bulk density. Pedotransfer functions have been derived from large international data sets across many studies, with their own inherent biases, many ignoring horization and depth variances. Initially pedotransfer functions from the literature were used to predict different horizon type bulk densities using local known bulk density data sets. Then the best performing of the pedotransfer functions were selected to recalibrate and then were validated again using the known data. The predicted co-efficient of determination was 0.5 or greater in 12 of the 17 horizon types studied. These new equations allowed gap filling where bulk density data were missing in part or whole soil profiles. This then allowed the development of an indicative soil bulk density map for Ireland at 0–30 and 30–50 cm horizon depths. In general the horizons with the largest known data sets had the best predictions, using the recalibrated and validated pedotransfer functions.

1 Introduction

Soils are a vital global resource providing a range of ecosystem services, upon which we depend. Such services include the platform on which we produce food, fibre and raw materials, purifying and regulating water, cycling of carbon and nutrients, and providing a habitat for biodiversity (EU, 2002). To understand many of the processes on-going in soils that deliver these ecosystem services, we must quantify soil characteristics, as these vary considerably according to soil type. Bulk density (ρ_b) is defined as the oven-dry mass per unit volume of a soil (IUSS 20 Working Group, 2006). This is an integral soil property, as it not only describes the packing structure of soils (Dexter, 1988), but is essential for the measurement of soil carbon and nutrient stock assessment (Ellert and Bettany, 1995). Bulk density measures can also describe the permeability of a soil, whereby it defines drainage characteristics (Arya and Paris, 1981) and is used in pedotransfer functions that model soil hydraulic characteristics (Murphy

et al., 2003; Van Alphen et al., 2001; Minasny and McBratney, 2007). Bulk density can also indicate compacted layers resulting from machinery or animal trafficking (Saffih-Hdadi, 2009), which can then impact the nutrient availability in soils (Douglas and Crawford, 1998).

Furthermore bulk density (ρ_b) is a critical soil characteristic for soil carbon studies and modelling, it can indicate the amount and/or volume rather than the concentration of carbon at a given point. Soil organic carbon (SOC) pool stock calculation depends upon suitable data in terms of organic carbon content and soil bulk density, and on the methods used to upscale point data to comprehensive spatial estimates (Vanguelova et al., 2016). The lack of appropriate bulk density documentation is problematic for statistical confidence assessments. Historically, ρ_b measurements are commonly missing from databases for reasons that include omission due to sampling and/or budgetary constraints and laboratory mishandling and/or conflicting methodologies (Batjes,

2009). Pedotransfer functions (PTF) based on readily measured soil attributes, such as organic carbon and clay content, show strong potential to replace ρ_b measurements as their direct measurement are not feasible or lacking from historical records.

However, bulk density has been found to vary with depth (Leonavičiūtė, 2000) and soil type (Manrique and Jones, 1991), while the use of generic pedotransfer functions can result in large errors in the calculation of SOC stocks. In saying this, De Vos indicates there is a need for specific PTF to be calibrated and validated on a regional basis (De Vos et al., 2005). Others take this further and report that PTF should be developed for particular horizon types or designations (Suster et al., 2011). Correlation with international data sets can be employed to generate PTF where local information is lacking. There is information available from large international soil survey databases (Hollis et al., 2006; Batjes, 2005, 2009), but in many cases bulk density is poorly documented. In these instances the use of splines or models of bulk density are then used with their own inherent variances, which can be problematic without large validation data sets (Letten et al., 2005).

With the launch of the Irish Soil Information System (Irish SIS) and the publication of the 3rd edition of the Irish soil map, there is the opportunity to measure, interpolate, and map bulk density values on a national scale. The latest soil map for Ireland has been published online by the Irish soil information system (Creamer et al., 2014).

The research presented in this paper will use new data generated by the Irish SIS to provide primary data for the calculation of PTF at the soil horizon level. This was done using soil bulk density measurements which were available for 15.9 % of the soil profiles described in Ireland in the last 40 years. In addition to this, PTF from the literature were used with known texture and organic carbon data to develop the calculations for bulk density. These PTF were then recalibrated for Irish soil horizons, where ρ_b was measured. The PTF were then applied to the soil horizons with unknown ρ_b . This allowed the calculation of soil bulk density to a depth of 50 cm for all soil profiles described. Using the PTF, bulk density is now known at different horizon designations. This has led to an indicative map of soil bulk density in Ireland being developed.

2 Materials and methods

2.1 Soil profiles

From 2012 to 2014 the Irish SIS sampled 246 soil pits as part of its field survey. The pits were selected by using an extensive auger survey of the Irish SIS. Pits were dug in areas where a high density of augers were found representing a particular soil type. From a practical position multiple pits were selected within a 10 km \times 10 km area when possible. This allowed excavation costs to be reduced greatly. The pits

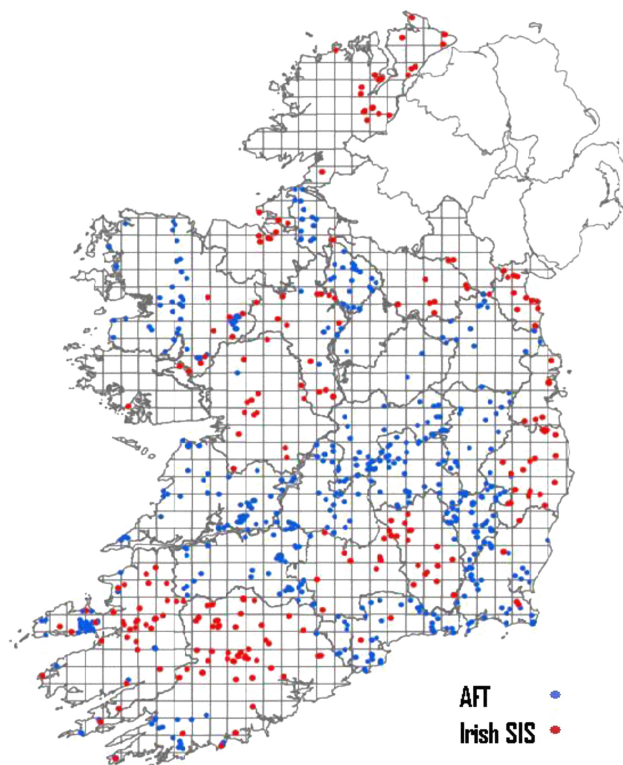


Figure 1. Location of Irish soil information system (Irish SIS) and An Foras Talúntais (AFT) soil profile pits. The blue circles correspond to AFT and the red circles correspond to Irish SIS.

were distributed across 16 counties in Ireland (Fig. 1). At each site a pit was excavated to approximately 1 m, where this was not possible, it was excavated to the depth of underlying bedrock preceding this. The pit face was at least 1 m wide. In total there were 1028 soil horizons identified (Simo et al., 2014). Within these pits, 470 horizons were sampled for bulk density (ρ_b). The remainder could not be measured for bulk density as the stainless steel rings were unusable due to coarse fragments. Therefore these horizons (528) required ρ_b predictions and pedotransfer functions were developed for this, detailed below.

2.2 Legacy data

In addition, detailed descriptions of 560 soil profiles were available from legacy data collected under the An Foras Talúntais soil survey (AFT) conducted between the 1960s and 1990s (An Foras Talúntais Staff, 1963, 1969, 1973; Conry, 1987; Conry and Ryan, 1967; Conry et al., 1970; Diamond and Sills, 2011; Finch and Ryan, 1966; Finch et al., 1971, 1983; Finch and Gardiner, 1977, 1993; Gardiner and Ryan, 1964; Gardiner and Radford, 1980; Hammond and Brennan, 2003; Kiely et al., 1974). However, very few bulk density measurements were taken as part of this survey, but detailed descriptions of soil horizons did exist, along with an-

alytical data for a number of soil parameters, such as texture and SOC. In total there were 2950 horizons described across 809 soil profiles located across the whole of Ireland (Fig. 1).

2.3 Field sampling

In the centre of each horizon, a smooth undisturbed vertical soil surface was prepared for ρ_b sampling. Three 50 mm \times 50 mm stainless steel rings were hammered into place. When possible, the rings were taken at 25, 50 and 75 cm from the edge of the pit wall. Care was taken to just fill the ring and not compact the soil. The ring plus soil was then removed from the surface of the soil matrix with as little disturbance as possible using a flat sided trowel. Any excess soil was trimmed from the ring edges before being placed in a sealed plastic bag. Also if protruding coarse fractions were present, they were marked and retained for cutting in the laboratory. For other soil parameters (texture, SOC, pH, cation exchange capacity, Fe/Al content), within the same horizon 2 kg of soil was sampled with a trowel into plastic bags and then sealed.

2.4 Bulk density analysis

The laboratory method followed that of the method applied during the few sites collected during the An Foras Talúntais survey (Massey et al., 2014). This method corresponds to ISO 11272:1998 – Soil Quality Part 5: Physical methods Sect. 5.6 – Determination of dry bulk density. The primary difference between the ISO and An Foras Talúntais methodologies is that the ISO does not account for stone mass and volume in its core method, whereas the methodology applied here does include this Eq. (1).

To calculate bulk density (stone-free):

$$\rho_b \left(\text{g cm}^{-3} \right) = (M_d - M_s) / (V - V_s), \quad (1)$$

where M_d = oven dry soil material weight (g), M_s = oven dry stone weight (g), V = volume of soil core (cm^{-3}), V_s = volume of stones (mL). The resulting ρ_b values were the mean of three field replicate samples.

2.5 Pedotransfer functions review and selection

Following a detailed review of the literature, 22 pedotransfer functions (PTF) were collated (Alexander, 1980; Adams, 1973; Rawls and Brakensiek, 1985; Honeysett and Ratkowsky, 1990; Federer, 1983; Huntington et al., 1989; Manrique and Jones, 1991; Bernoux et al., 1998; Leonavičiūtė, 2000; Kaur et al., 2002; Jeffrey, 1970; Harrison and Bocock, 1981; Tamminen and Starr, 1994). A first stage assessment was conducted using the Irish SIS data where ρ_b information was available for a range of soil horizon types. At this stage several ($n = 10$) PTFs were removed as negative and/or extremely low or high values were obtained and

the PTF did not appear to suit Irish data sets. The best remaining 12 PTFs for the various horizon types were then selected for use in further investigation (Table 2). These PTFs were chosen from the particular papers due to their development using high sample number ($n > 100$); sampling depth to at least 80 cm; wide range of soils covered and statistical evaluation (R^2). In most cases topsoils and subsoils were investigated and in others particular horizon types were investigated. For mineral soils eight PTFs were applied: Manrique and Jones (1991), Bernoux et al. (1998), Leonavičiūtė (2000) (x4), Kaur et al. (2002) (x2). For organic soils four PTFs were applied: Jeffrey (1970), Harrison and Bocock (1981), Manrique and Jones (1991), Tamminen and Starr (1994) (Table 2). As these PTF required soil organic carbon data, soil texture data and loss on ignition data, the methods below were applied to samples from the field campaign.

2.6 Soil organic carbon analysis

The soil was placed on aluminium trays and placed in an oven at 40 °C for 4 days. The dry weight was recorded and the soil sieved to 2 mm and stored. A LECO TrueSpec CN elemental analyser was used to measure SOC. Concentrated hydrochloric acid was used to remove inorganic carbon. The method followed that of Massey et al. (2014), which is an adaptation of Organic Application Note of the analysis of Carbon and Nitrogen in Soil and Sediment (LECO Corporation). This method corresponds to ISO 10694: 1995 – Soil quality Part 3: Chemical methods Sect. 3.8 – Determination of organic and total carbon after dry combustion (elemental analysis). The soils in the AFT survey had organic carbon estimated by the Walkley-Black dichromate oxidation method as described by Jackson (1958) and modified for colorimetric estimation. A comparison of archive samples using both methods was comparable with an R^2 of 97 %.

2.7 Soil texture analysis

The different particle sizes in the soil (sand, silt, clay) were determined via the pipette method. The premise of this method is based on Stokes' Law where the relationship between particle grain size and settling velocity in a fluid medium is predictable. A subsample of 2 mm dried and sieved soil was initially treated with hydrogen peroxide to remove all organic matter. Then it was suspended in a dispersant, sodium hexametaphosphate. Then finally 25 mL of the suspension was removed at exact time periods following shaking to represent silt and then clay fractions. This method of Massey et al. (2014) followed the methodology stated by An Foras Talúntais, National soil survey (Culleton, 1972). The work was conducted by an external laboratory following USDA texture guidelines. An inter-laboratory study was conducted to ensure continuity in the methodology between Teagasc and the external lab, where 50 soil samples were

analysed by both laboratories (85.4 % of soil samples were in agreement in textural class).

2.8 Loss on ignition

The soil organic matter content was estimated via loss on ignition (LOI) of any sample found to be over 10 % organic carbon via the elemental analyser. A subsample of the 2 mm dried and sieved soil was dried initially at 105 °C, cooled, and reweighed and then placed in a muffle furnace at 550 °C for 16 h. The difference in mass was equivalent to the organic matter content. This method is described in detail by Massey et al. (2014), which corresponds to BS EN 13039:2000 – Soil improvers and growing media – Determination of organic matter content and ash.

2.9 AFT and Irish SIS horizons

The horizon designations in the AFT survey were correlated to modern Irish Soil Information System definitions (Table 1). The Irish SIS designations are similar to the World Reference Base (WRB) system except for O, AB and Cr horizons which are equivalent to H, BA and CR in the WRB. The AFT designations were based on the soil horizon classification of soil survey staff, USDA (1960). When the equivalent horizon designation was identified the newly derived PTF could be applied to all horizons of this type. The soil horizon designation Ah indicating a lack of cultivation had no equivalent in the AFT records. The AFT survey did not record a non-cultivated A horizon.

2.10 Evaluation of PTFs

The individual ρ_b values were grouped together based on horizon designation. Each individual observed ρ_b value was predicted by each of the eight PTF in the case of mineral soils and the four PTF in the case of organic soils. A polynomial regression equation was generated for observed versus predicted ρ_b within each horizon type per PTF. The coefficient of determination (R^2) was compared across the PTF (Fig. 2a and Table 4).

The same data points were then compared using complementary prediction quality indices (De Vos et al., 2005). Here the quality of the prediction was determined via Eq. (2), the mean predicted error (MPE); Eq. (3), the standard deviation of the prediction error (SDPE); Eq. (4), the root mean squared prediction error (RMSPE); and Eq. (5) and the prediction coefficient of determination (R_p^2). These are defined as

$$\text{MPE} = \frac{1}{n} \sum_{i=1}^n (P\hat{b}, i - Pb, i) \quad (2)$$

$$\text{SDPE} = \sqrt{\frac{1}{n-1} \sum_{i=1}^n ((P\hat{b}, i - Pb, i) \text{MPE})^2} \quad (3)$$

Table 1. Irish SIS horizon designations used in this study and equivalent horizon titles used in the national soil survey by An Foras Talúntais.

Irish SIS	An Foras Talúntais
O	O, Oh
Ap	A, A1
Ap1	A11
Ap2	A12, A13
Ap _g	A/C, A11 _g , A12 _g , A13 _g
Ah	N/A
AB	A/B, A3, A14 _g
Bw	B, B1, B2, B21, B21h, B22, B3
B _g	B1 _g ,
B _s	Bsh,
B _t	Bth, Bts, Btc
B _{t_g}	Btgh, Btgs, Btgc
BC	BC _{t_g} , Bct, Bcg
BC _g	B2ca, 2Bca, Bca1
C _g	A/C _g
C/Ck/Cr	C1, C2, C3
E	A2, A21, A22, A23m, III1, II2

$$\text{RMSPE} = \sqrt{\frac{1}{n} \sum_{i=1}^n (P\hat{b}, i - Pb, i)^2} \quad (4)$$

$$R_p^2 = \frac{[\text{cov} Pb, i, P\hat{b}, i]^2}{\text{var}(Pb, i) - \text{var}(P\hat{b}, i)} \quad (5)$$

where Pb, i , and $P\hat{b}, i$ are the observed and predicted ρ_b values, respectively; n the number of observations; and var and cov, variance and the covariance function, respectively. MPE allows the evaluation of the bias of the PTF. The SDPE shows the random variation of the predictions after correction for global bias. The RMSPE is the overall error of the prediction. R_p^2 is a measure of the strength of the linear relationship between measurements and predictions, and indicates the fraction of the variation that is shared between them. The PTF generating the various R_p^2 values were compared (Table 5).

2.11 Calibration of the PTF

Using the prediction quality indices, the PTF selected per horizon was determined based on the highest R_p^2 value (Table 6). Once, the PTF was selected, it was updated using Irish data. For this, all data were divided into two groups, using 80 % of the data for the calibration process and 20 % for the validation model. These two groups were randomly selected. The validation data set is independent of the calibration data set but both are representative of the same soils. This is due to both data sets having the same sampling and analysis methods used, therefore the validation can be considered internal.

A particular PTF was then recalibrated using 80 % of the observed data points, randomly selected to generate a new model equation for that particular horizon type. Coefficients

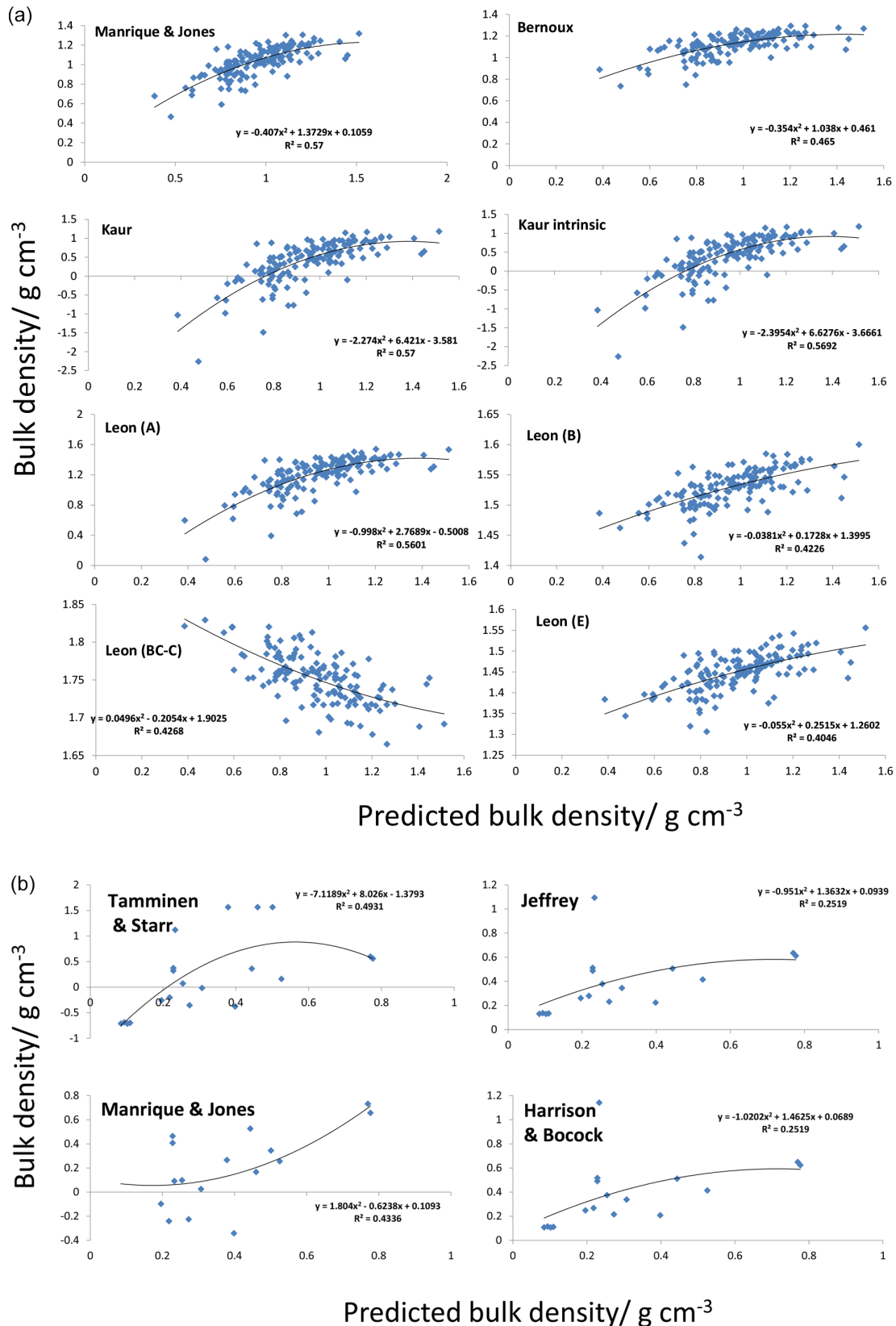


Figure 2. (a) Observed bulk density values for horizon Ap compared to prediction for original PTF formulae used indicating coefficient of variation equation and R^2 values. (b) Observed bulk density values for horizon O compared to prediction for original PTF formulae used indicating coefficient of variation equation and R^2 values.

Table 2. Published pedotransfer functions with corresponding authors used in this study. OC is organic carbon. ρ_b is bulk density in g cm^{-3} .

Author(s)	Pedotransfer function
Manrique and Jones (1991)	$\rho_b = 1.660 - 0.318(\text{OC})^{0.5}$
Bernoux et al. (1998)	$\rho_b = 1.398 - 0.0047(\text{Clay}) - 0.042(\text{OC})$
Kaur intrinsic	$\ln(\rho_b) = 0.313 - 0.191(\text{OC}) + 0.02102(\text{Clay}) - 0.000476(\text{Clay})^2 - 0.00432(\text{Silt})$
Kaur et al. (2002)	$\rho_b = 1.506 - 0.266(\text{OC}) + 0.004517(\text{Clay}) - 0.00352(\text{Silt})$
Leonavičiūtė (2000) A	$\rho_b = 1.70398 - 0.00313(\text{Silt}) + 0.00261(\text{Clay}) - 0.11245(\text{OC})$
Leonavičiūtė (2000) B	$\rho_b = 1.07256 + 0.032732 \ln(\text{Silt}) + 0.038753 \ln(\text{Clay}) + 0.078886 \ln(\text{Sand}) - 0.054309 \ln(\text{OC})$
Leonavičiūtė (2000) BC	$\rho_b = 1.06727 + 0.01074 \ln(\text{Silt}) + 0.08068 \ln(\text{Clay}) + 0.08759 \ln(\text{Sand}) + 0.05647 \ln(\text{OC})$
Leonavičiūtė (2000) E	$\rho_b = 0.99915 - 0.00592 \ln(\text{Silt}) + 0.07712 \ln(\text{Clay}) + 0.09371 \ln(\text{Sand}) - 0.08415 \ln(\text{OC})$
Jeffrey (1970)	$\rho_b = 1.482 - 0.6786 \log_{10}(\text{LOI})$
Harrison and Bocock (1981) – topsoil	$\rho_b = 1.558 - 0.728 \log_{10}(\text{LOI})$
Harrison and Bocock (1981) – subsoil	$\rho_b = 1.729 - 0.769 \log_{10}(\text{LOI})$
Tamminen and Starr (1994)	$\rho_b = 1.565 - 0.2298 (\text{LOI})^{0.5}$

Table 3. Statistics of observed bulk density, ρ_b (g cm^{-3}) for each horizon type, used in the development of pedotransfer functions.

Hz type	<i>N</i>	Mean ρ_b observed	Standard deviation	Co-efficient of variation	Min	Max	Variance
E	9	1.347	0.090	6.682	0.911	1.687	0.077
Ap	111	0.976	0.071	7.275	0.475	1.514	0.039
Ap1	28	1.044	0.061	5.843	0.386	1.289	0.035
Ap2	16	1.072	0.069	6.437	0.817	1.331	0.014
Ap _g	18	1.180	0.047	3.983	0.626	1.789	0.076
Ah	16	0.879	0.043	4.892	0.624	1.483	0.037
AB	12	1.014	0.075	7.396	0.881	1.373	0.020
O	20	0.329	0.039	11.854	0.196	0.777	0.032
Bw	52	1.147	0.094	8.195	0.758	1.844	0.053
B _g	56	1.381	0.080	5.793	0.902	1.762	0.035
B _s	7	1.086	0.058	5.341	0.710	1.353	0.052
B _t	8	1.307	0.036	2.754	0.907	1.501	0.058
B _{tg}	15	1.521	0.072	4.734	1.131	1.770	0.033
BC	15	1.444	0.084	5.817	0.770	1.754	0.051
BC _g	15	1.498	0.067	4.473	1.146	1.859	0.044
C/C _k /C _r	21	1.396	0.088	6.304	0.487	1.833	0.089
C _g	12	1.566	0.067	4.278	1.146	1.949	0.049

of the selected PTF were updated using multiple regression analysis (Table 7).

2.12 Model validation

After the recalibration the validation process was applied, using 20 % of the observed data points, again randomly selected. In some cases there were too few data points when 20 % of the observations were extracted. In this instance no validation could be performed, this affected four horizons (Bs, Bt, C/C_k/C_r and E, Table 7).

2.13 Digital Soil Mapping (DSM) techniques

The application of PTF has facilitated the prediction of soil bulk density for each genetic horizon for a total of 809 soil

profiles. The availability of this bulk density data allowed the development of maps derived upon these data points. Depths of the horizons were recorded, but these were not consistent across all sites as indicated earlier. Therefore, to obtain the bulk density at the different depths the horizon average was used (average of horizons that fall within the depth criterion). The horizon average was used for estimating bulk density at 0–30 and 30–50 cm depths (Fig. 4a and b). The DSM technique applied was a model which utilized the Universal Kriging method in R software. This involved the development of surface grids from the above profile bulk density data using spatial analyst interpolation.

Universal Kriging was the final model applied for the development of the indicative bulk density maps. Covariables used within the universal kriging approach included a land use map (O’Sullivan et al., 2015), slope data and a Digital El-

Table 4. Co-efficient of determination values (R^2) when comparing original bulk density values to predicted values for each horizon type, using the listed pedotransfer functions. Bold indicates the highest R^2 value for a particular horizon type.

Author HORIZON	Bernoux (1998)	Kaur et al. (2002)	Kaur et al. (2002) intrinsic	Leonaviciuté				Manrique and Jones (1991)	Jeffrey (1970)	Harrison and Boccock (1981) Topsoil	Tamminen and Starr (1994)	N
				(2000) (A)	(2000) (B)	(2000) (BC-C)	(2000) (E)					
Ap	0.46	0.57	0.57	0.56	0.42	0.43	0.40	0.57				111
Ap1	0.57	0.74	0.60	0.74	0.54	0.52	0.52	0.70				29
Ap2	0.48	0.36	0.25	0.36	0.30	0.35	0.26	0.36				16
Apg	0.59	0.69	0.50	0.69	0.59	0.55	0.55	0.69				18
Ah	0.34	0.34	0.42	0.36	0.13	0.17	0.43	0.31				16
AB	0.34	0.59	0.38	0.58	0.60	0.59	0.55	0.63				12
Bw	0.09	0.32	0.21	0.35	0.33	0.10	0.36	0.28				52
Bg	0.21	0.22	0.04	0.19	0.25	0.19	0.25	0.32				56
Bs	0.36	0.64	0.43	0.79	0.50	0.65	0.57	0.31				7
Bt	0.99	0.99	0.96	0.96	0.82	0.84	0.78	0.96				8
Btg	0.57	0.59	0.21	0.40	0.65	0.18	0.63	0.69				15
BC	0.09	0.55	0.26	0.41	0.58	0.28	0.55	0.59				15
C/Ck/Cr	0.06	0.22	0.12	0.09	0.33	0.03	0.27	0.34				21
Cg	0.02	0.71	0.52	0.63	0.39	0.33	0.41	0.64				12
BCg	0.41	0.07	0.24	0.09	0.26	0.33	0.26	0.19				15
E	0.48	0.61	0.78	0.62	0.52	0.44	0.57	0.49				9
O								0.43	0.25	0.25	0.49	20

Table 5. Co-efficient of determination values (R_p^2) when comparing original bulk density values to predicted values for each horizon type, using complimentary prediction quality indices (De Vos et al., 2005). Bold indicates the highest R_p^2 value for a particular horizon type.

Author HORIZON	Bernoux (1998)	Kaur et al. (2002)	Kaur et al. (2002) intrinsic	Leonaviciuté				Manrique and Jones (1991)	Jeffrey (1970)	Harrison and Boccock (1981) Topsoil	Tamminen Starr (1994)
				(2000) (A)	(2000) (B)	(2000) (BC-C)	(2000) (E)				
Ap	0.43	0.48	0.42	0.47	0.41	0.45	0.38	0.53			
Ap1	0.43	0.62	0.56	0.62	0.50	0.30	0.50	0.60			
Ap2	0.02	0.06	0.08	0.06	0.02	0.04	0.03	0.14			
Apg	0.52	0.61	0.47	0.60	0.56	0.55	0.53	0.64			
Ah	0.01	0.10	0.37	0.10	0.01	0.05	0.00	0.07			
AB	0.46	0.54	0.20	0.52	0.66	0.45	0.60	0.59			
Bw	0.05	0.27	0.21	0.28	0.29	0.06	0.32	0.23			
Bg	0.12	0.20	0.00	0.19	0.15	0.08	0.17	0.20			
Bs	0.01	0.46	0.17	0.55	0.12	0.14	0.54	0.17			
Bt	0.61	0.59	0.96	0.58	0.50	0.37	0.51	0.69			
Btg	0.48	0.37	0.20	0.29	0.53	0.02	0.49	0.42			
BC	0.00	0.43	0.25	0.26	0.52	0.25	0.48	0.50			
C/Ck/Cr	0.05	0.22	0.11	0.08	0.28	0.01	0.25	0.26			
Cg	0.02	0.34	0.19	0.21	0.35	0.32	0.31	0.47			
BCg	0.01	0.07	0.00	0.09	0.15	0.02	0.17	0.06			
E	0.10	0.53	0.29	0.52	0.51	0.06	0.52	0.48			
O								0.00	0.24	0.24	0.31

evaluation Model (DEM 20 m resolution). Land-use data were applied as this reflected the soil management types, in terms of compaction and/or poaching etc, which are major drivers of soil bulk density. The DEM provided information on altitude and slope degree, these data types were selected as they represent natural changes in bulk density as a result of the major topographical features and provide an indicator of the

climatic influence on soils at high altitudes (colder, wetter more acidic conditions). The soil association map was not included in this analysis, as this map is also a predicted product, SIS Final Technical Report 5, which uses the co-variants described within the prediction (Mayr et al., 2014).

The mask is the result of a number of updates that were made to the original post-processing, which was verified with

Table 6. The mean predicted error (MPE, g cm^{-3}); the standard deviation of the prediction error (SDPE, g cm^{-3}); the root mean squared prediction error (RMSPE, g cm^{-3}); and the prediction coefficient of determination (R_p^2) using complimentary prediction quality indices (De Vos et al., 2005) for each horizon type and selected pedotransfer function type.

Horizon	Selected PTF	MPE	SDPE	RMSPE	R_p^2
Ap	Manrique and Jones	0.067	0.132	0.148	0.532
Ap1	Leonavičiūtė A	0.246	0.137	0.280	0.619
Ap2	Manrique and Jones	0.110	0.117	0.158	0.142
Apg	Manrique and Jones	-0.058	0.174	0.179	0.640
Ah	Kaur (intrinsic)	-0.164	0.173	0.234	0.367
AB	Leonavičiūtė B	0.538	0.151	0.557	0.660
Bw	Leonavičiūtė E	0.425	0.206	0.471	0.318
Bg	Manrique and Jones	0.055	0.169	0.176	0.199
Bs	Leonavičiūtė A	0.488	0.172	0.513	0.551
Bt	Kaur (intrinsic)	0.375	0.128	0.393	0.957
Btg	Leonavičiūtė B	0.119	0.134	0.176	0.525
BC	Leonavičiūtė B	0.232	0.189	0.295	0.516
C/Ck/Cr	Leonavičiūtė B	0.275	0.158	0.315	0.276
Cg	Manrique and Jones	-0.085	0.159	0.175	0.471
BCg	Leonavičiūtė E	0.173	0.262	0.307	0.169
E	Kaur	0.067	0.050	0.082	0.529
O	Tamminen and Starr	-0.117	0.682	0.666	0.315

soil profile pit descriptions. This includes areas of peat, rock, alluvium, water and Sand. A matrix was compiled based on the legend of dunes, tidal marshes, and urban areas (Creamer et al., 2014).

2.14 Map validation methodology

For the validation of the map, independent data were used from the SoilH project having 72 locations sampled for bulk density (Kiely, 2015). The De Vos indexes (De Vos et al., 2005, covered in Sect. 2.10 above) were applied to establish the prediction quality of the Universal Kriging of the indicative bulk density maps. The map validation methodology is covered in detail in the SIS Final Technical Report 18 (Simo et al., 2015).

2.15 Mapping confidence

The validation applied indicated low confidence for both bulk density maps (for 0–30 and 30–50 cm, having an $R^2 = 0.32$ and $R^2 = 0.25$, respectively). The main problem is that the data used for mapping bulk density were not taken with this purpose in mind. Bulk density is a soil property that it is strongly influenced by the management practices and the sampling point strategy could influence directly the map product. Some features of the distribution may reflect regional variations in land use and management practices as well as the underlying soil properties, and the analysis may be influenced by sampling density across land use types. Therefore, these maps should be considered as indicative maps, guarantees cannot be made that the map gives the full

actual picture, hence the bulk density could vary in a particular location, thus the map legend shows ranges and not unique single values (Simo et al., 2015).

3 Results

3.1 Bulk density

The observed ρ_b values were grouped together based on horizon designation (Ap, Ap1, Ap2, Apg, Ah, O, E, AB, Bw, Bg, Bs, Bt, Btg, BC, BCg, C/Ck/Cr and Cg) and statistics applied in preparation for PTF application (Table 3). The minimum number of replicates per horizon type was seven for the Bs horizon and the maximum number of replicates per horizon was 111 for Ap. Horizons Ap1 and Ap2 are generally considered unique to Ap, this reflects the adoption of shallow till ploughing in some areas, however the bulk densities of both were similar, 1.044 and 1.072 g cm^{-3} , respectively. These designations were not unfounded as Ap horizons were generally lower (0.976 g cm^{-3}) when compared to Ap1 and Ap2 horizons. The largest bulk density was in Cg horizons (1.566 g cm^{-3}) and the lowest in the O horizons (0.329 g cm^{-3}). The Bt horizons had the lowest standard deviation and co-efficient of variation, 0.036 and 2.75 %, respectively. The O horizons had the largest co-efficient of variation at 11.854 %.

3.2 Application of pedotransfer functions

The selected eight mineral PTF and four organic PTF were applied to all horizon types (Table 4). The coefficient of determination for each PTF used is presented in Table 4. Those highlighted in bold indicate the highest R^2 value for a particular horizon type. This may span multiple PTF, for example horizon Ap has an R^2 value of 0.57 using the Kaur, Kaur intrinsic, and Manrique and Jones equations. The highest selected R^2 value from all the PTF was for horizon Bt at 0.99, this was for both Bernoux and Kaur PTFs. The lowest selected R^2 value for a specific horizon was the Bg with 0.32 using Manrique and Jones PTF. The highest R^2 value for O horizons was 0.49 using the Tamminen and Starr PTF.

Using the Ap horizon as an example, the plot of observed versus predicted ρ_b values for all mineral PTFs are presented in Fig. 2a. For O horizons the plot of observed versus predicted ρ_b values are presented in Fig. 2b. In both cases the regression equations and coefficients of determination are included in the plot. In the case of the Ap horizon, the Manrique and Jones PTF has all values positive for the predictions. For Kaur many of the predicted data points are negative as are those for the Kaur intrinsic PTF. Coupled with the R^2 value of 0.57 Manrique and Jones appears to be the best fit PTF. The same principles were applied to the rest of the mineral horizon PTF. For the O horizons Tamminen and Starr had the best R^2 value at 0.493, however this range contained negative values therefore the next highest R^2 value of 0.433

Table 7. Recalibrated pedotransfer functions (PTF) using Irish input data compared to measured bulk density. R_p^2 is the prediction coefficient of determination, R_v^2 is the validation coefficient of determination, both based on prediction quality indices (De Vos et al., 2005). ρ_b is bulk density (g cm^{-3}). OC is organic carbon.

Horizon	Original PTF	PTF recalibrated	R_p^2	R_v^2
Ap	Manrique and Jones	$\text{Db} = 1.5228 - 0.2806(\text{OC} \wedge 0.5)$	0.544	0.540
Ap1	Leonavičiūtė A	$\text{Db} = 1.26841 - 0.0010264(\text{silt}) + 0.004514(\text{clay}) - 0.092491(\text{OC})$	0.709	0.553
Ap2	Manrique and Jones	$\text{Db} = 1.3377 - 0.16927(\text{OC} \wedge 0.5)$	0.137	0.931
Apg	Manrique and Jones	$\text{Db} = 1.705925 - 0.342497(\text{OC} \wedge 0.5)$	0.758	0.899
Ah	Kaur (intrinsic)	$\text{Ln}(\text{Db}) = 0.228477 - 0.089759(\text{OC}) + 0.0064201(\text{Clay}) + 0.0004778(\text{clay} \wedge 2) - 0.00963(\text{Silt})$	0.621	0.744
AB	Manrique and Jones	$\text{Db} = 1.3966572 - 0.256208(\text{OC} \wedge 0.5)$	0.531	0.957
Bw	Leonavičiūtė E	$\text{Db} = -3.255 + 0.1517(\text{ln}(\text{Silt})) + 0.4519(\text{ln}(\text{Clay})) + 0.667(\text{ln}(\text{Sand})) - 0.183(\text{ln}(\text{OC}))$	0.472	0.560
Bg	Manrique and Jones	$\text{Db} = 1.588 - 0.302(\text{OC} \wedge 0.5)$	0.158	0.527
Bs	Leonavičiūtė A	$\text{Db} = 1.4809 - 0.0116(\text{Silt}) + 0.02937(\text{Clay}) - 0.64738(\text{OC})$	0.788	n/a
Bt	Kaur (intrinsic)	$\text{ln}(\text{Db}) = 0.208123 - 0.00139(\text{Silt}) + 0.002082(\text{Clay}) + 0.000343(\text{Clay} \wedge 2) - 0.1867 \cdot \text{OC}$	0.974	n/a
Btg	Leonavičiūtė B	$\text{Db} = 1.241791 - 0.02586(\text{ln}(\text{Silt})) - 0.01709(\text{ln}(\text{Sand})) - 0.07708(\text{ln}(\text{OC}))$	0.594	0.471
BC	Manrique and Jones	$\text{Db} = 1.8618 - 0.839(\text{OC} \wedge 0.5)$	0.580	0.257
C/Ck/Cr	Manrique and Jones	$\text{Db} = 1.773479 - 0.832265(\text{OC} \wedge 0.5)$	0.329	n/a
Cg	Manrique and Jones	$\text{Db} = 1.859853 - 0.477253(\text{OC} \wedge 0.5)$	0.668	0.994
BCg	Leonavičiūtė E	$\text{Db} = 1.6969 + 0.2297(\text{ln}(\text{Silt})) - 0.1102(\text{ln}(\text{Clay})) - 0.1303(\text{ln}(\text{Sand})) + \text{ln}(\text{OC})$	0.522	0.987
E	Leonavičiūtė E	$\text{Db} = -9.74290 + 1.282390(\text{ln}(\text{Silt})) + 0.6351(\text{ln}(\text{Clay})) + 1.222(\text{ln}(\text{Sand})) - 0.30286(\text{ln}(\text{OC}))$	0.562	n/a
O	Tamminen and Starr	$\text{Db} = 0.715618 - 0.05471(\text{LOI} \wedge 0.5)$	0.453	0.821

n/a = not applicable.

generated using Manrique and Jones was considered. Again on inspection this PTF also had generated negative values. The R^2 values of 0.251 for both Jeffrey and Harrison and Boccock were deemed too low to pursue even with all positive values. Tamminen and Starr was finally selected as the PTF for further investigation.

3.3 Selection of the best PTF

The performance of the selected PTF were further scrutinized using the prediction quality indices. The first of the indices to be examined was the prediction coefficient of determination, R_p^2 , across the eight mineral and four organic PTF. In many cases where the R^2 was the same across two or more PTF (Table 4), there was a clear R_p^2 value, larger than the others (bold, Table 5). For example Ap, where Manrique and Jones (0.53) is greater than Kaur and Kaur intrinsic at 0.48 and 0.42, respectively. The same situation occurred for horizon Ap1 (Leonavičiūtė A) and Apg (Manrique and Jones). The best performing PTF based on R^2 value, changed for horizons Ap2, Ah, Bt, Btg, BC, BCg, Cg, C/Ck/Cr and E, due to a higher R_p^2 value with a different PTF. For horizons AB, Bw, Bg, Bs and O the original best performing PTF based on highest R^2 value, was still appropriate, displaying the highest R_p^2 value also.

In Table 6 other indices were applied (MPE, SDPE and RMSPE) to support the most appropriate PTF selection. In general, the results show a positive MPE indicating an overestimation of ρ_b values (Table 6). However, horizons Apg, Ah, Cg, and O displayed a negative MPE indicating an underestimation of ρ_b values. The Bg horizon displayed the highest accuracy with a low MPE value of 0.055 g cm^{-3} ,

whereas the AB horizon had the poorest level of accuracy (0.538 g cm^{-3}).

RMSPE is the overall prediction error; this was highest with horizon O, 0.666 g cm^{-3} , and lowest for horizon E, 0.082 g cm^{-3} (Table 6). The prediction coefficient of determination (R_p^2) had a large range from 0.142 (Ap2) to 0.957 (Bt) and a median of 0.516 (BC). This was indicating that for horizons Ap2, Bg, and BCg there was low correlation and hence an unstable prediction. The SDPE value was converging to RMSPE value for horizons Ap, Apg, Bg, Cg, and O, therefore overall predictive error was due to precision error (SDPE). In contrast the total error was due to accuracy in the case of AB horizons with the large difference between the SDPE value and RMSPE value (0.406 g cm^{-3}). There was no pattern where low or high levels of MPE, SDPE or RMSPE or combinations thereof, resulted in a higher R_p^2 value.

The observed and predicted ρ_b values are presented in a box and whisker plot in Fig. 3. These predicted values are calculated using the selected PTF based on R_p^2 values of Table 6. The horizons with low accuracy (MPE) are evident in the case of AB, Bs, Bt, and C. Furthermore there is no overlap in the position of the interquartile ranges of the observed and predicted box and whisker plots. Those with good accuracy Apg, Bg, Cg, and E are evident as the red (observed) and blue (predicted) median bars are closer in position. In most cases for deeper and normally denser horizons, the interquartile range of ρ_b values are generally greater in the predictions than the observed. The max and min spread of the data (between 0.2 to 0.3 g cm^{-3}) is much narrower than the observed data ranges for horizons Bs, Bt, Btg, BC, BCg and C.

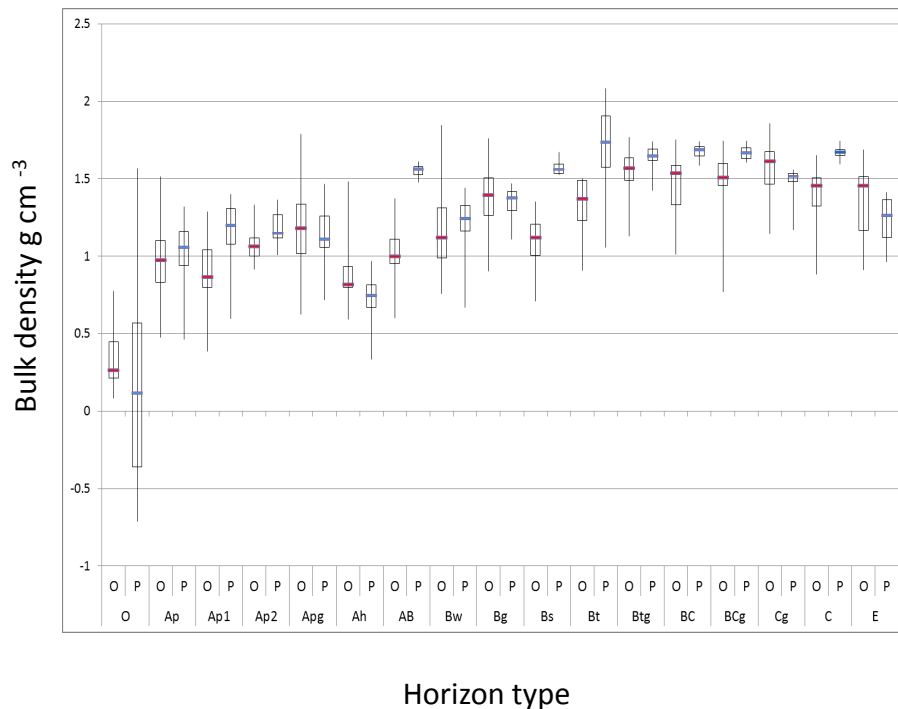


Figure 3. Observed bulk density (O) and predicted bulk density (P) g cm^{-3} , for each horizon type. Prediction based on selected PTF with best $R_p^2 R$ following prediction quality indices.

3.4 Recalibration of the selected PTF

Having selected the best performing PTF for each horizon type using the prediction quality indices, 80% of the observed data set was randomly selected for the recalibration of the PTF. The recalibrated PTF are presented in Table 7. For Ap, Ap2, AB, and Bg the Manrique and Jones intercept and coefficients have decreased due to lower densities in the data set. The intercept and coefficients increased with this PTF for Apg, BC, C/Ck/Cr, and Cg indicating higher densities in the data set. Leonavičiūtė A (Ap1), Kaur intrinsic (Ah and Bt) and Leonavičiūtė E (Bw), have decreased intercept and coefficients. Leonavičiūtė B increased intercept and coefficients, in both the cases of recalibration for Btg and BC. Leonavičiūtė E increased the coefficients and intercepts in the case of BCg and decreased in the case of E horizons.

The R_p^2 values have increased in most cases following recalibration (Table 7 compared to Table 6), especially in the case of Ah, Bs, and BCg (0.254, 0.237 and 0.353) however, there was a slight decrease for Ag and Bg horizons (0.129 and 0.041).

3.5 Validation of the recalibrated PTF

Validation has improved the coefficient of determination once again (Table 7), where 20% of the observed values were again randomly selected and R^2 generated. There have been increases in the R^2 validation values in comparison to the

R_p^2 values of 0.3 or more for Ap2, AB, Bg, Cg, BCg, and O. There was a large decrease for BC (0.323) and a small decrease for Ap1 and Btg (0.156 and 0.123). Except for horizon BC all other horizons have an R^2 of at least 0.47 or higher. Horizon BC with a low correlation (0.257) would have an unstable predictability. For horizons Bs, Bt, C, and E there were not enough data points in the validation data set of 20% to generate any validation indices.

3.6 Indicative soil bulk density map

Having bulk density data measured per horizon allowed the prediction of ρ_b in horizons where there were no measurements. This allowed gap filling in the Irish SIS and AFT profile data. In combination with mapping units from the latest edition of the Irish soil map and the methodology described above, a ρ_b map of Ireland was produced (Fig. 4). These maps highlight that lower bulk densities are found at the surface (0–30 cm) which is consistent with expected findings in relation to soil types and management, due in principle to higher soil organic carbon in these soils. The bulk density ranges from < 0.79 to $> 1.1 \text{ g cm}^{-3}$ (Fig. 4a). At increasing depths, 30–50 cm, higher bulk density values are likely to be found (< 1.0 to $> 1.4 \text{ g cm}^{-3}$). In general the bulk densities are lower in mountainous and hill areas and higher in lowland areas for both depth ranges.

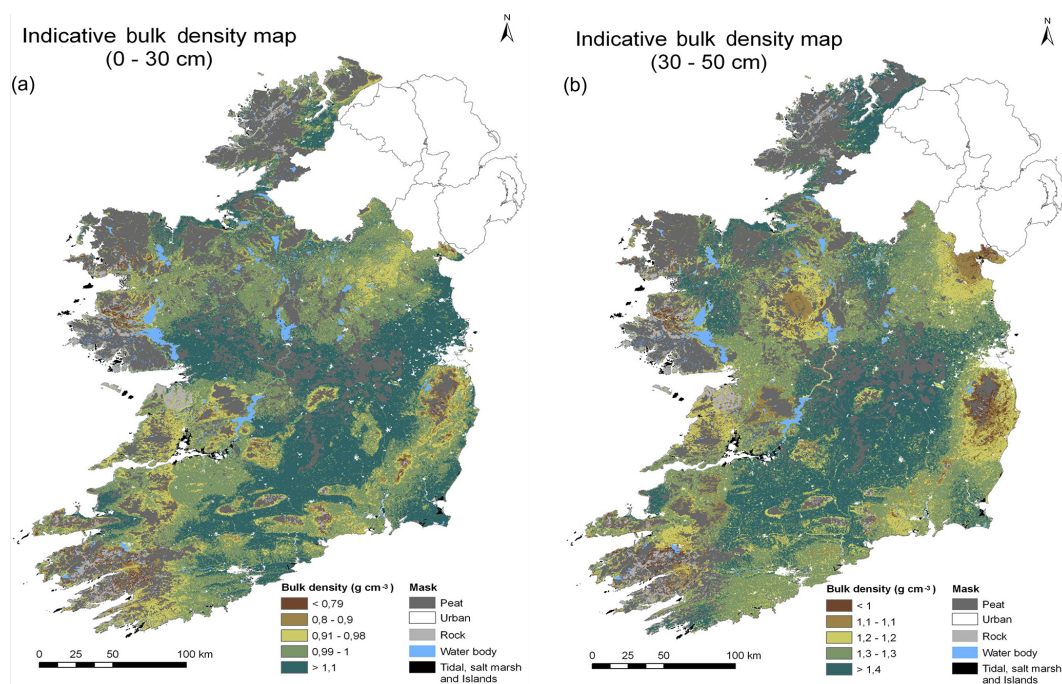


Figure 4. (a) Indicative soil bulk density distribution map for Ireland (0–30 cm, g cm^{-3}). (b) Indicative soil bulk density distribution map (30–50 cm, g cm^{-3}).

4 Discussion

4.1 Observed ρ_b values

The observed ρ_b values across all horizons have a mean of 1.187 g cm^{-3} with a standard deviation of 0.305 g cm^{-3} . Removing the O horizon value of 0.329 g cm^{-3} , the mean and standard deviation are 1.214 and 0.217 g cm^{-3} , respectively. This mean value compares favourably to Manrique and Jones (1991) on a range of agricultural soils $1.2\text{--}1.5 \text{ g cm}^{-3}$. The ForSite study of De Vos et al. (2005) reported another comparable value of 1.23 g cm^{-3} for topsoil. This value also compares well to the subsurface soils of Harrison and Bockock (1991), 1.29 g cm^{-3} , and forest soils of Tamminen and Starr, 1.19 g cm^{-3} .

Kiely et al. (2010), looking in particular at Irish soils to 50 cm depth found bulk densities for Brown Earths in the range of 1.02 to 1.22 g cm^{-3} , Brown Podzolics 0.94 to 1.07 g cm^{-3} , Gleys and Grey Brown Podzolics (Luvvisols) 0.86 to 1.3 g cm^{-3} and Podzols 0.53 to 1.23 g cm^{-3} . Reidy and Bolger (2013) reported ρ_b values of 1.018 to 1.063 g cm^{-3} on Gley soils in the Irish midlands to 30 cm depth. The generally higher levels in this study may be attributable to the greater depth studied and reported ρ_b increase with depth. This study's measured ρ_b values are well within the general ranges reported nationally and internationally. The O horizon value of 0.329 g cm^{-3} in this study appears to be greater than those reported in the literature. Wellock et al. (2011) report ρ_b values for Irish raised, high-

and low-level blanket peats of 0.133 , 0.118 and 0.125 g cm^{-3} and Kiely et al. (2010) report values of 0.15 to 0.25 g cm^{-3} for Irish peat soils. It should be noted that the O horizons in this present study included only horizons with greater than 12 % organic carbon. It is likely that these other studies, which indicate lower ρ_b values, are due to the peats having at least 40 % organic carbon content.

Looking at the mean values per horizon, the use of this approach appears justified with the large differences between surface horizons and sub-surface horizons (Ap, 0.976 g cm^{-3} , and Cg, 1.566 g cm^{-3} , Table 3). The difference between each type of surface horizon is also notable, where O horizons are 0.329 , and Ap1 and Ap2 (while close together at 1.044 and 1.072 g cm^{-3}) are different from Ap, reflecting differences in organic matter content and management, respectively. Therefore where possible predictions for soil bulk density should be at horizon level rather than topsoil or subsoil categorization.

To support this thinking, De Vos et al. (2005) noted that because of differences in topsoil and subsoil ρ_b values, PTFs developed using topsoil parameters only, which are being used to indicate ρ_b values in the subsoil, may lead to an underestimation. For this reason they developed topsoil and subsoil PTFs. An extension of this logic would be to use horizon specific PTFs, as applied in this paper. Because it was found that there were clearly significant differences in the PTF used according to the horizon type and this should be recognized in studies applying ρ_b down a profile to a specific depth.

The practice of splitting the bulk density of a singular profile into horizons has other advantages, especially when modelling systems. Many studies note that high levels of SOC are found at the surface, particularly at 0–30 cm depth. However more SOC could be found in the 30–100 cm range where the soils are denser. Adhikari et al. (2014) modelled ρ_b values using quadratic splines, when different horizon data were not available. This is a method to reflect the changes of ρ_b in soil profiles by using discrete soil depths. It was noted that accurate quantification of SOC stocks required a depth function. Tranter et al. (2007) also included a depth function when describing PTF based on soil mineral packing structures and soil structure. However it should also be noted that the fitting quality of splines to profile data depends on smoothing parameters, which introduces another source of error (Malone et al., 2009). In this study the data have been directly measured across the various horizons which avoids this error.

4.2 Application of literature PTF

The decision was made to apply our data set to PTF derived from the literature and then recalibrate. De Vos et al. (2005) indicated that the global predictive capacity of these functions appeared to be amenable to further improvement. Martin et al. (2011) stated that recalibration of existing PTF is worthwhile as the PTF itself defining more generally a function type, may be valid across several regions. However caution is required as parameters obtained under the given conditions can be too dependent on the data set characteristics. Generating new PTF from limited data could be prone to propagation of errors. In the Khalil et al. (2013) study for particular Great Groups, in Ireland, there was only SOC data to 10 cm available. The SOC had to be predicted to 50 cm and this predicted value was used once again to predict ρ_b values to 50 cm. This process was then repeated to generate values to 100 cm.

Nevertheless compartmentalization of bulk density data also has its merits; Heuscher et al. (2005) who analysed 47 000 measurements in the USDA survey improved the ρ_b predictions of their soils by placing the soils into suborders and then applying modelling techniques. The R^2 value improved from 0.45 to 0.62 in this process. Similar results were found by Manrique and Jones (1991) when they developed and applied the predictions within soil orders. This highlights an area for further investigation with data from the Irish SIS.

4.3 Recalibration of literature PTF

When recalibrating the PTF, it allowed the refinement of the equations for the Irish scenario. To date this is the most comprehensive model of Irish soils using the largest available data set, with soil profile, soil horizon, and depth coverage. The use of 80 % of the data points also followed the accepted De Vos et al. (2005) method. Where the categorization into

horizon PTF is justified and the R^2 values increased or are equalled for 14 out of the 17 horizons studied (Table 6).

The study of Xu et al. (2011) desired more data for deeper soils and greater site number (in the Irish context) to calibrate that studies PTF. They had used 0–10 cm soil depth carbon values to predict, firstly carbon content to 50 cm depth and then to predict soil bulk density to 50 cm depth. The use of sequential empirical regressions in developing PTF can propagate errors (Meersmans et al., 2008). The use of a singular PTF for peat and mineral soils in the Xu et al. (2011) study is also unlikely to be useful once actual peat ρ_b and SOC estimations at depth are required. This present study had both the depth and sample number data to calculate different PTF for various horizon types. The data generated in this study will avoid the propagation of errors described above and allow more accurate SOC calculation.

4.4 Validation of the recalibrated PTF

De Vos et al. (2005) emphasized the need for recalibration and local validation. This would aid the decision making process with reference to the level of what prediction error is acceptable. Getting this right is crucial as it has been recognized that correction factors led to an increase in the Belgian SOC prediction by 22 %, which also affected their projections due to landuse change and climate change (Letens et al., 2007). Although prediction errors between 10 and 20 % were deemed acceptable in the study of Prévost (2004). Huang et al. (2003) state that model acceptance would require between 10 and 20 % of the variance observed. For horizons with many replicates such as Ap ($n = 111$), the MPE falls within this criteria 0.067 g cm^{-3} , or 6.8 % of 0.976 g cm^{-3} . However this is not the case for many other horizon types which clearly need more replicates for example Bs ($n = 7$) MPE is 0.488 g cm^{-3} , or 44 % of 1.086 g cm^{-3} . Though, in most cases where a validation could be performed the predicted coefficient of variation was equalled or improved (R_v^2 , Table 7).

4.5 Mapping application

With the bulk density maps to 0–30 and 30–50 cm depth, the potential of these pedotransfer functions is realized. In Ireland there currently is no national map of soil carbon values, primarily due to the lack of bulk density data and also depth coverage. The National Soil Database project (2001-CD/S2-M2) measured 1365 points for organic carbon to 10 cm, however it did not measure bulk density. The SoilC project (Kiely et al., 2010) measured bulk density and organic carbon to 50 cm depth although this project was limited on number of sites ($n = 62$). Any studies deeper than 10 cm were in localized areas which did not allow extrapolation to the national area. Forest soils were covered in CARBiFOR 1 (Black and Farrell, 2006) and CARBiFOR 2 (CARBiFOR, 2015) projects, where soils were surveyed to 50 cm depth.

ρ_b was measured but site number was restrictive ($n = 44$). However, in both cases mapping criteria were not developed for greater areas. Most SOC studies and inventories are confined to 30 cm soil depth but the amount of SOC stored below 30 cm is of relevance in many ecosystems (Adhikari et al., 2014).

The PTF developed in this study allows the estimation of national organic carbon coverage of all soil types to 1 m depth with bulk density. This deficit of data was recognized with the initial development and is now further realized because of the recent availability of the Irish soil information system and its carbon data (Creamer et al., 2014). The same set of principles of method development of the PTF and mapping application could be applied to any national data set lacking in bulk density coverage.

5 Conclusions

The ρ_b values reported for horizon type allowed a greater range of soils in the Irish SIS to have ρ_b values allocated in the cases where there are omissions and to depth (recommended 1 m). The same process was applied to the AFT samples that did not have ρ_b values measured in the field. This paper covers the methodology of producing soil horizon PTF given the measured data available. Related predictions are based on the best data available after screening for accuracy and precision of PTF; they were then recalibrated and eventually validated within the Irish scenario. The methodology enabled the researcher to return to the Irish SIS to produce a validated ρ_b map at two depths, 0–30 cm and 30–50 cm (details of validation of map are given in Simo et al., 2015). Now that a ρ_b value is available for the different soil depths, values could be attributed to each soil mapping unit using Irish SIS into the future. Potentially this data could then be combined with known carbon data to produce a soil carbon map to 1 m. The data could also be used to produce a drainage map for the country. Another area for potential use would be the PTF used in hydrology studies, which use bulk density values. Furthermore, where nutrient management is a concern in soils, areas prone to compaction can be identified via this map. The PTF produced are valid for some horizons (with large R^2 values) and have limited success with other horizons. It is hoped in time as the sample number of these rarer horizons increases that the accuracy of the prediction increases. In general the greater sample number the better the prediction and validation.

Acknowledgements. This work was conducted as part of the Irish Soil Information System Project, managed by Teagasc (the Irish Agriculture and Food Development Authority) and co-funded by the Environmental Protection Agency (EPA) of Ireland through their Science, Technology, Research and Innovation for the Environment (STRIVE) Programme, as part of the National Development Plan 2007–2013. In addition, the authors would like to thank the lab staff of Johnstown Castle for their contribution to

this work, the student interns working in the field and the many landowners who gave access freely to their land.

Edited by: H. Reuter

References

- Adams, W. A.: The effect of organic matter on the bulk and true densities of some uncultivated podzolic soils, *J. Soil Sci.*, 24, 10–17, 1973.
- Adhikari, K., Hartemink, A. E., Minasny, B., BouKheir, R., Greve, M. B., and Greve, M. H.: Digital Mapping of Soil Organic Carbon Contents and Stocks in Denmark, *PLoS ONE*, 9, e105519, doi:10.1371/journal.pone.0105519, 2014.
- Alexander, E. B.: Bulk densities of California soils in relation to other soil properties, *Soil Sci. Soc. Am. J.*, 44, 689–692, 1980.
- An Foras Talúntais staff: West Cork Resource Survey, *Soil Survey Bulletin*, Dublin, 1963.
- An Foras Talúntais staff: West Donegal Resource Survey, Part 1 Soils and other physical resources, *Soil Survey Bulletin*, Dublin, 1969.
- An Foras Talúntais staff: County Leitrim Resource Survey, *Soil Survey Bulletin* No. 29, Dublin, 1973, 1973.
- Arya, L. M. and Paris, J. F.: A physicoempirical model to predict the soil moisture characteristic from particle-size distribution and bulk density data, *Soil Sci. Soc. Am. J.*, 45, 1023–1030, 1981.
- Batjes, N.: SOTER-Based Soil Parameter Estimates for Central and Eastern Europe (Version 2.0), Report 2005/4, ISRIC World Soil Information, Wageningen, 2005.
- Batjes, N. H.: Harmonized soil profile data for applications at global and continental scales: updates to the WISE database, *Soil Use Manage.*, 25, 124–127, 2009.
- Bernoux, M., Arrouays, D., Cerri, C., and Bourennane, H.: Modeling vertical distribution of carbon in oxisols of the western Brazilian Amazon (Rondonia), *Soil Sci.*, 163, 941–951, 1998.
- Black, K. G. and Farrell, E. P. (Eds.): Carbon sequestration and Irish forest ecosystems, COFORD, Dublin, 2006.
- CARB/FOR: II: Carbon Sequestration in Irish Forest Ecosystems, available at: <http://www.ucd.ie/carbifor>, last access: 14 May 2015.
- Conry, M. J. and Ryan, P.: Soils of Co. Carlow, *Soil Survey Bulletin* No. 17, Dublin, 1967.
- Conry, M. J., Hammond, R. F., and O'Shea, T.: Soils of County Kildare, *Soil Survey Bulletin*, Dublin, 1970.
- Conry, M. J.: Soils of Co. Laois, *Soil Survey Bulletin* No. 41, Dublin, 1987.
- Creamer, R., Simo, I., Reidy, B., Carvalho, J., Fealy, R., Hallett, S., Jones, R., Holden, A., Holden, N., Hannam, J., Massey, P., Mayr, T., McDonald, E., O'Rourke, S., Sills, P., Truckell, I., Zawadzka, J., and Schulte, R.: Irish Soil 501 Information System., 2007-S-CD-1-S1, EPA STRIVE Programme 2007–2013, EPA, Dublin, 2014.
- Culleton, E.: Laboratory analyses in soil survey investigations: theory and techniques, An Foras Talúntais National Soil Survey, Johnstown Castle, Wexford, Republic of Ireland, 1972.
- De Vos, B., Meirvenne, M. V., Quataert, P., Deckers, J., and Muys, B.: Predictive quality of pedotransfer functions for estimating bulk density of forest soils, *Soil Sci. Soc. Am. J.*, 69, 500–510, 2005.

- Dexter, A. R.: Advances in characterization of soil structure, *Soil Till. Res.*, 11, 199–238, 1988.
- Diamond, J. and Sills, P.: Soils of Co. Waterford, National Soil Survey of Ireland, Soil Survey Bulletin No. 44, Teagasc, Oak park, Carlow, Co. Carlow, Grehan Printers, Dublin, 314 pp., 2011.
- Douglas, J. T. and Crawford, C. E.: Soil compaction effects on utilization of nitrogen from livestock slurry applied to grassland, *Grass Forage Sci.*, 53, 31–40, 1998.
- Ellert, B. H. and Bettany, J. R.: Calculation of organic matter and nutrient stored in soils under constrain management regimes, *Can. J. Soil Sci.*, 75, 529–538, 1995.
- EU – European Union: Towards a Thematic Strategy for Soil Protection, Communication from the Commission to the Council, the European Parliament, the Economic and Social Committee and the Committee of the Regions, COM (2002) 179 final, Brussels, 35 pp., 2002.
- Federer, C. A.: Nitrogen mineralization and nitrification: Depth variation in four New England forest soils, *Soil Sci. Soc. Am. J.*, 47, 1008–1014, 1983.
- Finch, T. F. and Gardiner, M. J.: Soils of Co. Westmeath, Soil Survey Bulletin No. 37, Dublin, 1977.
- Finch, T. P. and Gardiner, M. J.: Soils of Tipperary North Riding, Soil Survey Bulletin No. 42, Dublin, 1993.
- Finch, T. F. and Ryan, P.: Soils of Co. Limerick, Soil Survey Bulletin No. 16, Dublin, 1966.
- Finch, T. F., Culleton, E., and Diamond, S.: Soil Survey Bulletin No. 23, Dublin, 1971.
- Finch, T. F., Gardiner, M. J., Comey, A., and Radford, T.: Soils of Co. Meath, Soil Survey Bulletin No. 37, Dublin, 1983.
- Gardiner, M. J. and Radford, T.: Soil Associations of Ireland and their land use potential, Soil Survey Bulletin No. 36, Dublin, 1980.
- Gardiner, M. J. and Ryan, P.: Soils of Co. Wexford, Soil Survey Bulletin No. 1, Dublin, 1964.
- Hammond, R. F. and Brennan, L. E.: Soils of Co. Offaly, Soil Survey Bulletin No. 43, Dublin, 2003.
- Harrison, A. F. and Bocoock, K. L.: Estimation of soil bulk-density 476–481, from loss-on-ignition values, *J. Appl. Ecol.*, 8, 919–927, 1981.
- Heuscher, S. A., Brandt, C. C., and Jardine, P. M.: Using soil physical and chemical properties to estimate bulk density, *Soil Sci. Soc. Am. J.*, 69, 51–56, 2005.
- Hollis, J., Jones, R., Marshall, C., Holden, A., van de Veen, J., and Montanarella, L.: SPADE 2: The Soil Profile Database for Europe version 2.0, European Soil Bureau Research Report No. 19, EUR 22127 EN, Office for the Official Publications of the European Communities, Luxembourg, 2006.
- Honeysett, J. L. and Ratkowsky, D. A.: The use of ignition loss to estimate bulk density of forest soils, *J. Soil Sci.*, 40, 299–308, 1989.
- Huang, S., Yang, Y., and Wang, Y.: A critical look at procedures for validating growth and yield models, in: Modelling forest systems, edited by: Amaro, A., Reed, D., and Soares, P., CAB International, Wallingford, UK, 271–293, 2003.
- Huntington, T. G., Johnson, C. E., Johnson, A. H., Siccama, T. G., and Ryan, D. F.: Carbon, organic matter, and bulk density relationships in a forested spodosol, *Soil Sci.*, 148, 380–386, 1989.
- IUSS Working Group WRB: World Reference Base for soil resources, World Soil Resources Report No. 103, FAO, Rome, 128 pp., 2006.
- Jackson, M. L.: Soil Chemical Analysis, Prentice-Hall, Inc., New Jersey, 1958.
- Jeffrey, D. W.: A note on the use of ignition loss as a means for the approximate estimation of soil bulk density, *J. Ecol.*, 58, 297–299, 1970.
- Kaur, R., Sanjeev, K., and Gurung, H.: Apedo-transfer function (PTF) for estimating soil bulk density from basic soil data and its comparison with existing PTFs, *Aust. J. Soil Res.*, 40, 847–857, 2002.
- Khalil, M. I., Kiely, G., O'Brien, P., and Müller, C.: Organic carbon stocks in agricultural soils in Ireland using combined empirical and GIS approaches, *Geoderma*, 193–194, 222–235, 2013.
- Kiely, G.: Interactions of Soil Hydrology, Land Use and Climate Change and their Impact on Soil Quality, Datasets Available at: Secure Archive for Environmental Research Data managed by Environmental Protection Agency Ireland, <http://erc.epa.ie/safer/resource?id=e6df65fa-e1b8-11e3-b233-005056ae0019>, last access: 6 October 2015.
- Kiely, G., Eaton, J., McGoff, N., Xu, X., and Leahy, P.: SoilC measurement and modelling of soil carbon stocks and stock changes in Irish soils. Final Report, Environmental Protection Agency, Johnstown Castle, Ireland, 2010.
- Kiely, J., Diamond, S., Burke, P. J., and Collins, T.: Soil Map of West Mayo, An Foras Talúntais, Dublin, 1974.
- Leonavičiūtė, N.: Predicting soil bulk and particle densities by pedotransfer functions from existing soil data in Lithuania, *Geografijos metraštis*, 33, 317–330, 2000.
- Letten, S., Van Orshoven, J., Van Wesemael, B., De Vos, B., and Muys, B.: Stocks and fluxes of soil organic carbon for landscape units in Belgium derived from heterogeneous data sets for 1990 and 2000, *Geoderma*, 127, 11–23, 2005.
- Letten, S., De Vos B., Quataert, P., Van Wesemael, B., Muys, B., and Van Orshoven, J.: Variable carbon recovery of Walkley-Black analysis and implications for national soil organic carbon accounting, *Eur. J. Soil Sci.*, 58, 1244–1253, 2007.
- Malone, B. P., McBratney, A. B., Minasny, B., and Laslett, G. M.: Mapping continuous depth functions of soil carbon storage and available water capacity, *Geoderma*, 154, 138–152, 2009.
- Manrique, L. A. and Jones, C. A.: Bulk density of soils in relation to soil physical and chemical properties, *Soil Sci. Soc. Am. J.*, 55, 476–481, 1991.
- Martin, M. P., Wattenbach, M., Smith, P., Meersmans, J., Jolivet, C., Boulonne, L., and Arrouays, D.: Spatial distribution of soil organic carbon stocks in France, *Biogeosciences*, 8, 1053–1065, doi:10.5194/bg-8-1053-2011, 2011.
- Massey, P., O'Connor, C., Sills, P., Fenelon, A., Maloney-Finn, L., Reidy, B., and Creamer, R.: Irish Soil Information System: Laboratory Standard Operating Procedures (2007-S-CD-1-S1), EPA STRIVE Programme 2007–2013, Report 7, EPA, Dublin, 2014.
- Mayr, T. R., Zawadzka, J. E., Corstanje, R., Lo Papa, G., and Fealy, R.: Predictive Mapping, Final Technical Report 5, Irish Soil Information System (2007-S-CD-1-S1), EPA STRIVE Programme 2007–2013, EPA, Dublin, 2014.
- Meersmans, J., De Ridder, F., Canters, F., De Baets, S., and Van Molle, M.: A multiple regression approach to assess the spatial

- distribution of Soil Organic Carbon (SOC) at the regional scale (Flanders, Belgium), *Geoderma*, 143, 1–13, 2008.
- Minasny, B. and McBratney, A.: Estimating the Water Retention Shape Parameter from Sand and Clay Content, *Soil Sci. Soc. Am. J.*, 71, 1105–1110, 2007.
- Murphy, B., Geeves, G., Miller, M., Summerell, G., Southwell, P., and Rankin, M.: The application of pedotransfer functions with existing soil maps to predict soil hydraulic properties for catchment-scale hydrologic and salinity modelling, Proceedings MODSIM Conference, Townsville, Australia, 2003.
- O'Sullivan, L., Creamer, R. E., Fealy, R., Lanigan, G., Simo, I., Fenton, O., Carfrae, J., and Schulte, R. P. O.: Functional land management for managing soil-based ecosystem services: A case-study of the trade-off between primary productivity and carbon storage in response to the intervention of drainage systems in Ireland, *Land Use Policy*, 47, 42–54, 2015.
- Prévost, M.: Predicting soil properties from organic matter content following mechanical site preparation of forest soils, *Soil Sci. Soc. Am. J.*, 68, 943–949, 2004.
- Rawls, W. J. and Brakensiek, D. L.: Prediction of soil water properties for hydrologic modelling, in: *Watershed Management in the Eighties*, edited by: Jones, E. and Ward, T. J., Proc of a Symp. ASCE, 30 April–2 May 1985, Denver, CO, ASCE, New York, 293–299, 1985.
- Reidy, B. and Bolger, T.: Soil carbon stocks in a Sitka spruce chronosequence following afforestation, *Irish Forestry*, 70, 200–219, 2013.
- Saffih-Hdadi, K., Défossez, P., Richard, G., Cui, Y. J., Tang, A. M., and Chaplain, V.: A method for predicting soil susceptibility to the compaction of surface layers as a function of water content and bulk density, *Soil Till. Res.*, 105, 96–103, 2009.
- Simo, I., Creamer, R., Reidy, B., Jahns, G., Massey, P., Hamilton, B., Hannam, J. A., Jones, R. J. A., McDonald, E., Sills, P., and Spaargaren, O.: Irish Soil Information System Soil profile handbook (2007-S-CD-1-S1), EPA STRIVE Programme 2007–2013, Report 10, EPA, Dublin, 2014.
- Simo, I., Creamer, R. E., O'Sullivan, L., Reidy, B., Schulte, R. P. O., and Fealy, R. M.: Final Technical Report 18, Irish Soil Information System (2007-S-CD-1-S1) EPA STRIVE Programme 2007–2013, EPA, Dublin, 2015.
- Suuster, E., Ritz, C., Roostalu, H., Reintam, E., Kõlli, R., and Astover, A.: Soil bulk density pedotransfer functions of the humus horizon in arable soils, *Geoderma*, 163, 74–82, 2011.
- Tamminen, P. and Starr, M.: Bulk density of forested mineral soils, *Silva Fennica*, 28, 53–60, 1994.
- Tranter, G., Minasny, B., McBratney, A. B., Murphy, B., McKenzie, N. J., Grundy, M., and Brough, D.: Building and testing conceptual and empirical models for predicting soil bulk density, *Soil Use Manage.*, 23, 437–443, 2007.
- USDA: Soil survey staff of the United States Department of Agriculture: Soil Classification – A Comprehensive System, 7th Approximation, Soil Conservation Service, US Department of Agriculture, Washington, D.C., p. 265, 1960.
- Van Alphen, B. J., Booltink, H. W. G., and Bouma, J.: Combining pedotransfer functions with physical measurements to improve the estimation of soil hydraulic properties, *Geoderma*, 103, 133–147, 2001.
- Vanguelova, E. I., Bonifacio, E., De Vos, B., Hoosbeek, M. R., Bergere, T. W., Armolaitis, K., Celi, L., Dinka, L., Kjønaas, O. J., Pavlenda, P., Pumpanen, J., Püttsepp, Ü., Reidy, B., Simončič, P., Tobin, B., Vesterdal, L., and Zhiyanski, M.: Forest soil carbon stocks at different scales – uncertainties and recommendations, *Environ. Monit. Assess.*, submitted, 2016.
- Wellock, M. L., Reidy, B. J., LaPerle, C. M., Bolger, T., and Kiely, G.: Soil carbon stocks of afforested peat in Ireland, *Forestry*, 84, 441–451, 2011.
- Xu, X., Liu, W., Zhang, C., and Kiely, G.: Estimation of soil organic carbon stock and its spatial distribution in the Republic of Ireland, *Soil Use Manage.*, 27, 156–162, 2011.