1	Abundance changes of marsh plant species over 40 years are better
2	explained by niche position water level than functional traits
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21 Abstract

22 Understanding the factors that determine species' resistance to environmental change is of 23 utmost importance for biodiversity conservation. Here we investigated how the abundances of 24 marshland species are determined by niche properties and functional traits. We re-surveyed 25 150 vegetation plots that were first surveyed in 1973 in order to explore species abundance 26 changes over time. We found that the mean water level in the habitats of most studied species 27 decreased significantly from 1973 to 2012. Nine of 17 target species were identified as 28 abundance decreasing species and the other eight as abundance increasing species. The comparisons of seven plant characteristics (niche position water level, plant height, and five 29 30 leaf traits) showed that the decreasing species had a significantly higher value of optimum 31 water level and marginally significantly lower leaf N contents and specific leaf area (SLA) 32 than those in increasing species. The stepwise regression analysis showed that optimum water 33 level and leaf N were the best predictors of abundance changes of marsh plant species, as well 34 as that the effect of optimum water level was stronger than that of leaf N. Our findings 35 demonstrated that niche properties may be important for forecasting changes in wetland plant 36 communities over time.

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38 Keywords: species abundance; environmental change; optimum water level; leaf traits;
39 vegetation resurvey; wetland.

40 Introduction

41 Environmental changes are altering both the geographical distributions and the abundances of 42 plants worldwide (Beckage et al., 2008; Hughes, 2000; Sturm et al., 2001; Sala et al., 2000; 43 Walther et al., 2002). However, it is difficult to predict how plant communities will respond to 44 these changes because of the species-specific responses which occur among taxa (Grime, 45 1998; Lavorel & Garnier, 2002). Therefore, understanding which particular plant characteristics determine the performance of species under environmental change is of utmost 46 47 importance for biodiversity conservation as well as for predicting future changes in species distributions under climate change. 48

49 The responses of species to habitat change primarily depends on their biological and ecological characteristics (Joan et al., 2009; Broennimann et al., 2006). Biological 50 characteristics, e.g., establish new populations, are determined by morphological and 51 52 physiological traits (e.g., individual size, root depth, mycorrhizae) related to resource 53 acquisition or competitiveness (Lavergne et al., 2004; Van der Veken et al., 2007). Ecological 54 characteristics, such as niche properties, may also be shaped by environmental drivers (e.g., 55 water availability, temperature) (Devictor et al., 2010; Swihart et al., 2006; Williams et al., 56 2007; Alarcon & Cavieres, 2018). There is a growing interest in the use of functional traits 57 and niche properties of species as indicators of their response capacities in the assessment of 58 the impacts of environment change on plant communities (Foden et al., 2013; Triviño et al., 59 2013; Garcia et al., 2014; Pearson et al., 2014). However, whether these characteristics are 60 predictors of the performance responses of species to temporal environmental changes has

61 rarely been experimentally tested in the field.

62 Several key properties or traits have been proposed as determinants of species responses to 63 environmental change (Soudzilovskaia et al., 2013; Alarcon & Cavieres, 2018; Estrada et al., 64 2015; Broennimann et al., 2006), including niche position, leaf N contents and specific leaf 65 area et al.. Meanwhile, many hypotheses have been proposed, for example (1) species with 66 marginal niche position is more sensitive to environmental change than those with middle niche position (Johnson, 1998); (2) species with higher leaf N contents (nitrogen-demanding 67 68 species) are often comparably winner species (Diekmann et al., 1999); (3) species with higher 69 specific leaf area tended to be loser species compared to winner species (Naaf and Wolf, 70 2011). However, these hypotheses remain poorly explored under other conditions of 71 environmental changes besides climate warming.

72 Over the last several decades, the Sanjiang Plain in Northeast China has undergone 73 pronounced temperature increases and unprecedented land use changes with marked 74 decreases in wetland areas (Liu & Ma, 2002). These changes caused significant changes in 75 species composition and richness (Lou et al. 2015). Moreover, our previous studies have identified a significant impact of habitat change on niche properties (Lou et al, 2018) and 76 functional traits (Lou et al., 2012; 2016) of dominant marsh plants. However, these studies 77 78 mainly focused on either community-level changes or differences in functional traits of 79 individual species. It is not known how the abundance of individual species has responded to 80 habitat change over the recent decades. The studies by Lou et al. (2012; 2015; 2016; 2018) 81 provide the basis for exploring how abundance changes of marsh plant species to be related

82 with functional traits and niche properties.

This study aimed to investigate the influence of functional traits, niche properties and environmental changes on abundance changes of marsh plant species in Northeast China. Firstly, we identified increasing and decreasing species by comparing abundance of the dominant marshland species between the sampling year 1973 and 2012. Secondly, we tested whether the increasing and decreasing species differed in niche and trait attributes. Thirdly, we quantified the relative importance of niche properties, functional traits and environmental changes.

90 Materials and methods

91 Study area

92 The Sanjiang Plain in Northeast China is the alluvial plain of the Heilongjiang, Songhua, and Wusuli Rivers. It encompasses a total area of 108,829 km^2 and has an altitude of mostly <200 93 94 m a.s.l. (Fig. 1). The climate type of the area is temperate humid and sub-humid continental monsoon climate, with mean annual precipitation of 510-620 mm and mean annual 95 96 temperature of 2.1-4.5 °C. The vegetation is dominated by four plant communities: 97 permanently flooded emergent marshes (dominant species: Phragmites australis, Carex 98 lasiocarpa, and Carex pseudocuraica), seasonally flooded tussock marshes (Carex 99 appendiculata or C. meyeriana), occasionally flooded marsh meadows (C. appendiculata and 100 Calamagrostis angustifolia), and shrub marsh meadows (Betula fruticosa, Alnus sibirica, Salix brachypoda, and C. angustifolia). The soil types include Luvisols, Phaeozems, 101

102 Cambisols, and Histosols.

103 The study area includes the Sanjiang Plain Marsh Ecological Experimental Station (CERN) 104 and the International Long Term Ecological Research Network (ILTER, http:// 105 data.iter-europe.net/deims/site/iter-eap-cn-29). In recent decades, this area has undergone 106 significant climate warming and human reclamation activities (Liu & Ma, 2002) which 107 resulted in drying of the wetland habitats and drastic decreases in marsh areas from more than 108 50,000 km² in the 1940s to 9,100 km² in 2000 (Lou *et al.*, 2015).

109 Vegetation survey

110 In July and August of 1973, when the water level was the lowest in the whole growing season 111 (Lou et al., 2016), the first vegetation survey of the Sanjiang Plain was conducted by Zhao 112 Kuiyi, Li Chonghao, Yi Fuke et al. (Changchun Institute of Geography, CAS) with the aim of 113 mapping vegetation and describing vegetation types. In this historical study, vegetation was 114 surveyed using sampling plots of $1 \text{ m} \times 1 \text{ m}$. For each vascular plant species in each plot, the 115 percentage cover (0-100% vertical projection) was estimated. Furthermore, water level in 116 each plot was measured using a meter stick above the soil surface. In the above-mentioned 117 historical surveys, vegetation plots were not permanently marked, but the positions of the 118 plots were described in great detail so that they could be located again.

In July and August of 2012, we resurveyed 70 plots of *Carex lasiocarpa* formations and 80 plots of *Carex appendiculata* formations (a total of 150 plots) following the sampling protocols described in Zhao *et al* (1999). The descriptions of the sites and plot locations from the previous surveys (e.g., landform, proximity to river) were carefully studied and all available information was taken into account when we were relocating plots within similar vegetation types. Plots were placed as close as possible to their supposed original locations in each wetland site (estimated spatial errors <10 m). All the plots were located in national or provincial nature reserves, and no human disturbance (such as fire) occurred from 1973 to 2012. In 2012, the geographic position of every studied plot was determined using a GPS. Plant nomenclature followed Fu (1995).

129 Data analyses

130 Species abundance changes

131 Species abundance indicates the percentage cover on average for all plots where the plant 132 species was present. We calculated species abundance for both 1973 and 2012. If the changes 133 in abundance of a species from 1973 to 2012 were positive, the species was regarded as 134 'increasing species', in the opposite case, the species was regarded as 'decreasing species'. 135 Only species occurring in at least 15 plots in whether the original sampling or resampling 136 were considered for statistical analyses of species cover (n = 17). For each of the 17 species, 137 paired t-test was used to test the significance of the differences in species cover between the 138 two sampling years.

139 Analysis of the drivers of species abundance changes

140 In order to explain the observed differences in abundance between the two periods, we 141 assembled the data on environmental changes and species traits according to their ecology

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142 and biology.

As Lou *et al.* (2015) found that hydrology is the main factor controlling the changes along the marsh zonation in the study area, we used the change in water level as the indicator of environmental change. For each species in each sampling year, we calculated the average water level in all plots where the species occurred. For each species, the significance of the differences in the water level between the two sampling years was tested by paired t-test, and the change in the water level between the two sampling years was calculated by subtracting the water level value in 1973 from the water level value in 2012.

150 We used niche optimum along the water level gradient as an ecological property. For all the 151 selected species, the responses of each species to water level gradient were fitted by 152 Generalized Additive Models (GAM, Hastie and Tibshirani, 1990). In which, water level and 153 abundance data (represented by percent cover, 0-100%, as a continuous variable) from the 154 field survey in 1973 were used as the explanatory and dependent variable, and the identity and 155 Gaussian were selected as the link function and error term, respectively. The optimum of each 156 species was identified as the value of water level at which abundance of a species had its maximum. 157

For the analysis of biological traits, we measured plant height, leaf area (LA), specific leaf area (SLA), leaf N, P, and N:P ratios from the samples collected in the ten field sites during vegetation surveys in 2012. In each marsh site, aboveground parts of ca. 10 to 20 individuals or ramets per studied species were harvested and taken back to the laboratory. Plant height

162 was measured by ruler. Three healthy mature leaves were obtained from each individual and measured by a leaf area meter (LI3000C). Leaf samples were further dried in a drying oven 163 164 for 24 h at 70 °C, and their dry weights were measured by analytical balance with the 165 accuracy of 0.0001 g. Phosphorus and Nitrogen concentrations were determined using the 166 molybdenum blue method and the indophenol blue method, respectively, using Skalar 167 Auto-Analyzer. Boxplots were used to remove outliers, following which the mean values of 168 each trait in each studied species were calculated. We evaluated whether decreasing and 169 increasing species differed in their trait attributes using the t-test. We also tried to collect the 170 information on root traits of the studied species, but this task was not completed because of 171 sampling difficulties resulted from the clonality of the studied species in the field. For this 172 reason, root traits were not included in the analysis of biological traits.

173 Finally, we assessed how much of the variation in abundance could be explained by 174 environmental changes and plant ecological and biological properties using a linear model. 175 We conducted a stepwise regression analysis of the changes in abundance, and only 176 significant variables were selected as the predictors. Meanwhile, in order to clearly assess the 177 relationship between each trait or niche parameter and abundance changes, we also conducted 178 a linear regression analysis for each trait or niche parameter with change in cover as the 179 response variable. In addition, we also tested the effect of phylogeny on species abundance 180 changes and the selected significant variables, and found no correlation between them 181 (supporting information I). So the effect of phylogeny was excluded, and phylogeny was not 182 included as an driving factor of species abundance changes.

All independent variables were standardized prior to the regression analysis. The statistical
analyses were conducted in R version 3.4.1 (R Development Core Team 2010) with the R
package mgcv (Wood, 2006).

186 **Results**

187 Changes in water level in the investigated habitats

188 General climate conditions during the two sampling years were comparable. Mean daily air 189 temperature and monthly precipitation from January to December in 1973 were similar to 190 those in 2012 (Fig. 2) and were not significantly different (t test: F=0.481, p=0.495 for mean 191 air temperature; F=0.072, p=0.792 for monthly precipitation). Mean water level in the 150 192 plots decreased significantly from 9.7±0.69 cm (mean±SE) in 1973 to 5.3±0.41 cm in 2012 193 (paired t-test: t=5.477, p<0.001). Out of the 17 common species, 11 species experienced a 194 significant decline in the mean water level of their habitats of at least 4.5 cm from 1973 to 195 2012, and two species (Stachys baicalensis and Scutellaria baicalensis) experienced an 196 increase, but it was not significant (Table 1).

197 Changes in species abundance

198 The 17 dominant marsh species are all perennials. Of which, nine species decreased in their 199 abundance from 1973 to 2012 and were identified as decreasing species, and eight species 200 increased in their abundance and were identified as increasing species. Average abundance 201 changes between the sampling year 1973 and 2012 among decreasing and increasing species 202 ranged from -12.2% (Carex pseudocuraica) to +7.3% (Carex appendiculata). Two of the 203 decreasing species (Menyanthes trifoliata and Sanguisorba tenuifolia) and five of the increasing species (Lysimachia thyrsiflora, Glyceria spiculosa, Equisetum fluviatile, Iris 204 205 laevigata, and Carex pseudocuraica) were significant (Fig. 3). The group of decreasing and increasing species were both heterogeneous and were comprised of seven and five families,respectively (supporting information I).

208 Drivers of species abundance changes

The comparison of seven tested traits between the increasing and decreasing species showed that only the difference in optimum water level was significant (Fig. 4, t-test: t=3.811, p=0.002). The differences in leaf N and SLA between the increasing and decreasing species were marginally significant (t-test: t=-1.891, p=0.078; t=-2.095, p=0.054, respectively; Fig. 4). No significant differences were observed in the other traits (plant height, leaf P, N:P ratios and leaf area) between the increasing and decreasing species (Fig. 4).

215 Optimum water level and leaf N were good individual predictors of abundance change and 216 they were negatively and positively correlated with abundance changes, respectively (Table 2). 217 The best model based on multiple traits explained 68% of the variance and included only 218 optimum water level and leaf N. However, R^2 only marginally increased when leaf N was 219 added as a predictor in the model with niche position water level (Table 2, Fig. 5).

220 **Discussion**

In this study, we conducted a novel analysis of the driving factors of species abundance change over time at the local scale. Out of all tested factors, the most important factor for explaining changes in marsh plant abundance was the relative realized optimum position along the water level gradient, and leaf N also had a minor impact. i.e. species with a higher water level optimum and, at the same time, lower leaf N content decreased more dramatically in abundance between 1973 and 2012.

227 The two traits included into the best model (optimum water level and leaf N) represent the 228 environmental tolerance and resource capture capacity of a species, respectively, and these 229 traits explained more than half of the variation in cover change. The explanatory power of the 230 relative realized optimum position along the water level gradient was much greater than that 231 of leaf N. On the one hand, this indicates that the influence of environmental tolerance on 232 species sensitivity is greater than that of biological traits. On the other hand, this is a relatively 233 high value in comparison with those obtained in other studies on abundance dynamics 234 (Soudzilovskaia et al., 2013). This is likely because of the fact that the water level gradient represents a complex gradients (correlation with the other environmental factors, e.g. soil 235 236 nutrient, Lou et al., 2013) and the effects of the other substrates on species abundance may be 237 integrated into that of water level.

238 The univariate negative correlation between species abundance change and optimum water239 level (Table 1) indicates that the species with high optimum water level increase their

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240 dominance, while the species with low optimum water level decrease their dominance as a 241 consequence of habitats drying. This is consistent with the niche position (or habitat 242 availability) hypothesis and with other studies which assumed that the closer the niche is to 243 the margins of the environmental gradient, the more sensitive the species are to habitat change 244 (Heino & Mendoza, 2016; Williams et al., 2007; Broennimann et al., 2006). This suggests 245 that the potential sensitivity of species to habitat change can, at least partly, be estimated a priori from their niche characteristics along key environmental gradients, and this was 246 247 confirmed in another study as well (Thuiller et al., 2005).

248 In the present study, biological traits seemed to play a relatively minor role in the prediction 249 of marsh species abundance changes. Out of the investigated traits, there was a relatively big 250 difference only in leaf N and SLA between the decreasing and increasing species. Leaf N of 251 the increasing species was higher than that of the decreasing species. On the one hand, leaf N 252 is related to leaf growth, and species with high leaf N have generally a rapid growth rate 253 (Reich et al., 2008). Laliberte et al (2012) found that species associated with a rapid growth 254 rate become dominant under soil resource addition, and this supports our result based on the 255 fact that the N deposition is serious in study area (Lou et al., 2015). On the other hand, this suggests that the minor importance of leaf N may be a result of its correlation with the 256 257 optimum water level which was stronger than that of all the other tested traits (supporting 258 information II), i.e., its contribution may be a result of its covariance with the optimum water 259 level. Similarly, SLA of the increasing species was larger than that of the decreasing species, 260 and this is also consistent with the result of Laliberte et al. (2012). Moreover, SLA has been

proved to impact abundance dynamics in the studies by Soudzilovskaia *et al.* (2013) and Estrada *et al.* (2015). However, this trait had very low explanatory power for abundance changes in our study, most likely because the capacity for rapid growth and competition conferred by SLA is possibly not useful to new ramets regeneration, which is the main way to increase species dominance in wetland communities dominated by colonization.

266 The exclusion of a variety of reproductive traits along with population dispersal may also 267 impact the relationships between biological traits and species abundance changes. The role of reproductive traits in species persistence and colonization has been highlighted by many 268 269 researchers (Thuiller et al., 2012; Estrada et al., 2015; Pacifici et al., 2017; MacLean & 270 Beissinger, 2017). Most wetland vascular plant species feature clonal reproduction and their 271 clonal growth traits (e.g., tiller number, rhizome biomass, root depth, etc.) were shown to be 272 important for species distribution dynamics at small scales (Purcell et al., 2019; Moor et al., 273 2015; 2017). For example, species with deeper roots are more able to resist habitat changes 274 than shallower-rooted species (Willis, 2017). In the present study, the increased abundance of 275 Menyanthes trifoliata may be related to its deep roots. In addition, our approach did not 276 consider intra-specific plasticity or variation of traits, which can be substantial since species traits depend on hydrological heterogeneity (e.g., Kostikova et al., 2013) and are well 277 documented as key mechanisms affecting species resistance to habitat change (Valladares et 278 279 al., 2014; Liancourt et al., 2015).

The correlation between niche properties and functional traits (especially life history traits)
may interfere with our results. The study carried out by Thuiller *et al.* (2004) confirmed that

282 niche position was related to leaf and flowering traits. In the present study, niche position 283 water level may have been determined by the traits related with flooding conditions, by other 284 adaptations to anoxia, and perhaps by the presence of mycorrhiza at the dry side of the 285 gradient. For example, Purcell et al. (2019) demonstrated that fine root traits (such as 286 aerenchyma and dry matter content of fine root) are correlated with flooding duration in 287 ephemeral wetlands. Such traits were not included in this study because of the difficulty in 288 collecting data on these traits. In the follow-up studies, the correlation analysis of niche 289 properties and life history traits related to flooding conditions needs to be carried out.

290 Finally, water level changes was not showed as the variable response for species dominance change in this study, but its role was third next to that of leaf N and optimum water level 291 292 (Table 2). This demonstrates that water level change is still an important driving factor of 293 vegetation change, and should be strengthened in the future research. As for the driving 294 factors of water level decrease, besides climate warming and drainage mentioned in our 295 previous study (Lou et al., 2015), the impact of land use change needs to be reemphasized 296 here. The area of paddy field in this region increased from 3,200 ha in 1990 to 91,300 ha in 2005, and much more groundwater was extracted for irrigation, with the result that the buried 297 298 depth of groundwater level in this area decreased by 3 m on average (Li et al., 2007).

299 **Conclusions**

300 This study highlighted that the response of species abundance to habitat change depends more301 on their niche position water level than on their functional traits. It is well known that local

302 abundances of species are related to the extent of their geographical distribution (Brown, 303 1984); thus, niche position water level, and possibly leaf N, should be considered for 304 integration into species distribution models for predicting regional distribution of wetland 305 plants. This study suggested at least three perspectives for further research, concerning: a) 306 disentangling the relative contribution of soil nutrient availability in the complex gradient 307 represented by water level; b) including reproductive and life history traits related with 308 flooding conditions into the driving mechanism of abundance changes of marsh plant species, 309 and c) acquiring a better understanding of site hydrology as it is an important controlling 310 factor of vegetation changes in marsh ecosystems.

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- 465 Appendix A. Supplementary data
- 466
- 467
- 468

Species	Number of plots		Mean water level (cm) (mean ± SE)			
	1973	2012	1973	2012	р	
Carex lasiocarpa	109	101	11.6 ± 0.83	6.6 ± 0.53	0.000	
Glyceria spiculosa	84	50	13.2 ± 1.17	6.1 ± 0.64	0.000	
Carex pseudocuraica	58	93	10.8 ± 1.25	6.3 ± 0.63	0.003	
Lysimachia thyrsiflora	57	70	9.6 ± 1.02	$\textbf{5.4} \pm \textbf{0.48}$	0.000	
Caltha palustris	56	80	$\textbf{8.6} \pm \textbf{0.83}$	$\textbf{3.8} \pm \textbf{0.53}$	0.000	
Equisetum fluviatile	52	57	10.3 ± 1.11	5.6 ± 0.61	0.001	
Menyanthes trifoliata	50	48	10.4 ± 1.06	6.4 ± 0.55	0.005	
Iris laevigata	47	38	7.6 ± 0.86	5.9 ± 0.54	0.066	
Salix myrtilloides	47	47	10.3 ± 1.06	$\textbf{5.2} \pm \textbf{0.72}$	0.000	
Comarum palustre	44	80	8.4 ± 0.93	7.2 ± 0.67	0.442	
Carex meyeriana	40	44	8.2 ± 1.0	5.7 ± 0.86	0.075	
Carex appendiculata	38	43	6.9 ± 0.96	$\textbf{1.4} \pm \textbf{0.27}$	0.000	
Calamagrostis angustifolia	31	46	5.5 ± 0.90	$\textbf{3.1} \pm \textbf{0.42}$	0.025	
Sanguisorba tenuifolia	28	22	4.9 ± 0.77	3.6 ± 0.59	0.186	
Lycopus lucidus	21	51	4.1 ± 0.95	1.7 ± 0.48	0.019	
Stachys baicalensis	9	19	2.9 ± 1.01	4.3 ± 0.64	0.158	
Scutellaria baicalensis	4	20	3.3 ± 0.88	4.7 ± 1.00	0.608	

Table 1. Mean water level for 17 dominant species in Sanjiang Plain marsh in the two sampling years. *p*-value was obtained from paired t-test. Species are sorted by their presence in the plots 1973.

470 **Table 2.** Results of the regression analyses of abundance changes and plant functional traits. The 471 ecological trait was the optimum water level, and the biological traits were individual height, leaf N, P, 472 and N:P, specific leaf area (SLA) and leaf area (LA). Only the variables strongly related to change in 473 species abundance are shown. OWL, optimum water level; WLC, water level change; N, number of 474 species; B, regression coefficient.

Response variable	Model	Significant predictors	В	AIC	F	R^2_{adj}	р
	multiple stepwise – regression _ (n=17)	OWL + Leaf N		87.160	17.82	0.68	< 0.001
		OWL	-3.54				< 0.001
		Leaf N	1.02				0.200
Cover	individual — predictor (n=17 for each predictor) —	OWL	-3.98	87.222	32.12	0.66	< 0.001
change		Leaf N	2.55	101.111	5.814	0.23	0.029
6		WLC	1.50	104.951	1.606	0.036	0.224
		Leaf N:P ratio	0.91	106.061	0.556	< 0	0.468
		SLA	0.80	106.205	0.43	< 0	0.524
		Leaf P	0.58	106.436	0.216	< 0	0.648
	-	Leaf area	0.40	106.56	0.11	< 0	0.749
		Height	-0.11	106.67	0.01	< 0	0.93

476 Figure legends

- 477 Fig. 1 Distribution of the investigated sampling plots in the wetlands of the Sanjiang Plain,
 478 Northeast China.
- 479 Fig. 2 The comparison of monthly mean air temperature (a) and monthly precipitation (b) of
- 480 five national meteorological stations (Fujin, Jiamusi, Baoqing, Hulin and Hegang Station) in
- 481 study area for the sampling year 1973 and 2012.
- 482 **Fig. 3** Abundance changes of the 17 most common wetland species in Sanjiang Plain between
- 483 the sampling year 1973 and 2012.
- 484 Fig. 4 Comparison of niche property and functional traits between the decreasing and
- 485 increasing species. DS: decreasing species; IS: increasing species.
- 486 **Fig. 5** The relationship between optimum water level, leaf N, and species abundance changes.



Fig. 1 Distribution of the investigated sampling plots in the wetlands



Fig. 2 Monthly mean air temperature (a) and accumulated rainfall in January-December (b) of five national meteorological stations (Fujin, Jiamusi, Baoqing, Hulin and Hegang Station) in study area for the sampling year 1973 and 2012.



Fig. 3 Abundance changes of the 17 most common wetland species in Sanjiang Plain between the sampling year 1973 and 2012. * indicates significant abundance change based on the paired-t test.



Fig. 4 Comparison of niche property and functional traits between the decreasing and increasing species. DS: decreasing species, n=9; IS: increasing species, n=8. The core boxes of boxplots indicate the interquartile range of data whereas the whisker lines symbolize the centiles 5 and 95, middle black line is the median.



Fig. 5 The relationship between optimum water level, leaf N and species abundance changes (n=17 species).

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Conflict of interest statement

We declare that we have **no** financial and personal relationships with other people or organizations that can inappropriately influence our work, there is **no** professional or other personal interest of any nature or kind in any product, service and/or company that could be construed as influencing the position presented in, or the review of, the manuscript entitled " Abundance changes of marsh plants over 40 years are better explained by niche position than functional traits". Credit author statement

Yanjing Lou: investigation, data curation, writing, original draft preparation; Jutta
Kapfer: methodology, reviewing and editing; Pete Smith: reviewing and editing;
Ying Liu: investigation, data analysis, figures. Ming Jiang and Xianguo Lu: data
collecting and reviewing; Michael Ashcroft: reviewing and editing.