

Structure of mixed ombrophylous forests with *Araucaria angustifolia* (Araucariaceae) under external stress in Southern Brazil

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Abstract: This study is part of the Floristic and Forest Inventory of Santa Catarina, conceived to evaluate forest resources, species composition and structure of forest remnants, providing information to update forest conservation and land use policy in Southern Brazilian State of Santa Catarina (95 000km²). In accordance to the Brazilian National Forest Inventory (IFN-BR), the inventory applies systematic sampling, with 440 clusters containing four crosswise 1 000m² plots (20x50m) each, located on a 10x10km grid overlaid to land use map based on classification of SPOT-4 images from 2005. Within the sample units, all woody individuals of the main *stratum* (DBH≥10cm) are measured and collected (fertile and sterile), if not undoubtedly identified in field. Regeneration *stratum* (height>1.50m; DBH<10cm) is registered in 100m² in each sample unit. Floristic sampling includes collection of all fertile trees, shrubs and herbs within the sample unit and in its surroundings. This study performs analysis based on 92 clusters measured in 2008 within an area of 32320km² of mixed ombrophylous forests with *Araucaria angustifolia* located at the state's high plateau (500m to 1 560m above sea level at 26°00'-28°30' S and 49°13'-51°23' W). Mean density (DBH≥10cm) is 578 individuals/ha (ranging from 85/ha to 1 310/ha), mean species richness in measured remnants is 35 (8 to 62), Shannon and Wiener diversity index (H') varies between 1.05 and 3.48. Despite high total species diversity (364 Magnoliophyta, five Coniferophyta and one tree fern) and relatively high mean basal area (25.75m²/ha, varying from 3.87 to 68.85m²/ha), the overwhelming majority of forest fragments are considered highly impacted and impoverished, mostly by logging, burning and extensive cattle farming, turning necessary more efficient protection measures. Basal area was considered an appropriate indicator for stand quality and conservation status. Rev. Biol. Trop. 59 (3): 1371-1387. Epub 2011 September 01.

Key words: regional forest inventory, *Araucaria angustifolia*, forest structure, human impacts on forests.

Subtropical native forests have suffered drastic reduction in Southern Brazil during the last hundred years and continue under serious pressure (Metzger 2009). Within this scenario, the Santa Catarina State with a total area of 95000km² shows a relatively high native forest cover of about 22.4% (SOS Mata Atlântica 2010), besides agriculture and extensive cattle farming and approximately 500 000ha of forest

plantations, mostly of *Pinus* spp., equivalent to 5% of its territory. Native forests are not homogeneously distributed in the state's three main phytogeographic regions (Klein 1978, 1979, Veloso *et al.* 1991), but remnants are concentrated in the region of the coastal range covered by dense Atlantic rain forests. Mixed ombrophylous forests (MOF) with *Araucaria angustifolia* (Bertol.) Kuntze are one of the

most threatened forest formations in Brazil, resting only 12.6% of its original extension (Ribeiro *et al.* 2009) and only 7% within Santa Catarina State. The term mixed ombrophylous forest is used to describe a type of forest vegetation characterized by elements of both the tropical (afro-brazilian) and temperate (austral-anthartic- andinean) flora (Veloso *et al.* 1991, Leite 2002), under a humid subtropical climate without a dry season. This forest type is also known as *Araucaria* forest, due to the physiognomic dominance of *A. angustifolia*, occurring at elevations above 800m.a.s.l. and from latitudes 20°S to 29°S. Although *A. angustifolia* is the physiognomic dominant species, it is in general accompanied, by species as *Podocarpus lambertii* Klotzsch ex Endl., *Drymis brasiliensis* Miers, *Ocotea porosa* (Nees & Mart.) Barroso, *Schinus terebinthifolius* Raddi, *Ocotea pulchella* (Nees & Mart.) Mez and *Ilex paraguariensis* A.St.-Hil. (Veloso *et al.* 1991). In open areas are frequently found light-demanding tree species like *Piptocarpha angustifolia* Dusén ex Malme, *Mimosa scabrella* Benth., *Vernonanthura discolor* (Spreng.) H.Rob. and *Ocotea puberula* (Rich.) Nees, according to Klein & Hatschbach (1970/71). Despite of numerous studies about MOF in the Southern states of Paraná and Rio Grande do Sul (reviewed by Figueiredo Filho *et al.* 2006), knowledge about these forests in Santa Catarina is based on few works (Klein 1960, 1990, Negrelle & Silva 1992, Formento *et al.* 2004, Lingner *et al.* 2007).

The Floristic and Forest Inventory of Santa Catarina (IFFSC) is an initiative of the local government, realized between 2007 and 2010 in order to evaluate the conservation *status* of forest cover and to support the formulation of forest conservation and land use policy (Vibrans *et al.* 2008). Focused on endangered tree species in a highly fragmented forest cover under permanent pressure by agriculture and forest plantations, the inventory is composed by five components: (1) Integration of the state's four herbaria collections, (2) field inventory *strictu sensu*, (3) assessment of genetic structure of endangered tree populations, (4)

analysis of socio-economic importance of forest resources and (5) on-line geo-referenced database to be accessed by decision makers and the public. The Santa Catarina inventory applies a systematic sampling with 440 clusters located on a 10x10km grid, in accordance to the Brazilian National Forest Inventory (IFN-BR) to be started in 2010.

The purpose of this study is to analyze the structure of MOF in the state of Santa Catarina, in Southern Brazil, relating it to human impacts and threats to which it is exposed.

MATERIALS AND METHODS

The present work consists of a preliminary evaluation of field data collected between 2007 and 2008 in the first stage of IFFSC, realized in mixed ombrophylous forests with *A. angustifolia*. The study performed analysis of woody species based on 92 clusters containing four crosswise 1000m² plots (20x50m), within an area of approximately 32 320km² at the state's high plateau at 500m to 1560m above sea level, located at 26°00' -28°30' S - 49°13' -51°23' W. The 10x10km sampling grid was overlaid to land use map based on unsupervised classification of multispectral SPOT-4 images from 2005 with forest and non-forest classes to determine cluster location at forest areas. No isolation distance was established from the forest border, since the grid location coordinates were strictly obeyed. While 87 of the analyzed clusters are located on the regular grid, five clusters (n° 2001 to 4000) were located arbitrarily, within the studied area, in areas of public domain that suffered no logging activities for at least 40 years, in order to obtain data from what are supposed to be forests under better conservation conditions.

Within the sample units (Fig. 1) all woody individuals of the main *stratum* with diameter at breast height (DBH≥10cm) were measured and collected (fertile and sterile), if not undoubtedly identified in field (Vibrans *et al.* 2010). Regeneration *stratum* (height>1.50m; DBH<10cm) was registered in 100m² in each sample unit, including regrowth of tree species

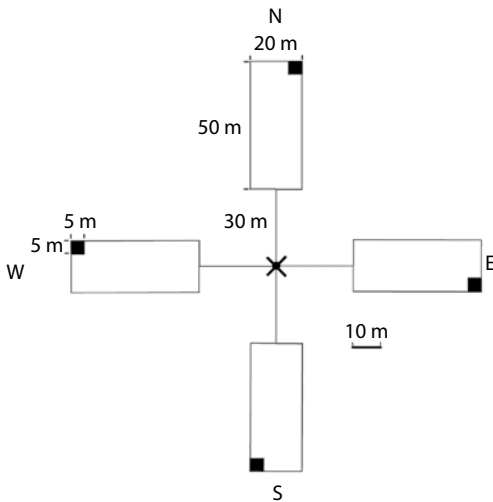


Fig. 1. Sample unit of IFFSC, composed by four crosswise 1 000m² plots (20x50m), at 30m distance from centre point.

and understory shrubs. Variables measured in main *stratum* were DBH, height, stem quality, tree sanity and crown length, while in regeneration *stratum* only density and height of plants. Floristic sampling included collection of fertile trees, shrubs, herbs and epiphytes within the sample unit and in its surroundings. All measured sites were submitted to a detailed physiognomic description by the field crew, including description of disturbing factors and any type of human impact within and in the sampled forest surroundings. A number of 1 809 voucher specimen collected during execution of this study were deposited at herbarium FURB under reference number 6 828 to 20 684 (not consecutive).

Floristic and structural indices were calculated according to Mueller-Dombois & Ellenberg (1974). Sampled forest fragments, that we also call remnants, were classified into three groups, using basal area, species composition, successional stage and conservation status as indicators, described as follows:

- **Group 1:** Vegetation characterized by highly degraded, mostly open arboreous

vegetation in early successional stage, due to intensive exploitation or clear cutting, invasion of pioneer tree and shrub species; sometimes with sparse tree cover, presence of silvopastoral systems, canopy cover less than 50%.

- **Group 2:** Vegetation in intermediate successional stage, after intensive logging and exploitation of *Ilex paraguariensis* and extensive cattle farming; vegetation predominantly arboreous, in variable conservation status, creating mosaic of vegetation types, evidence of recent and historic timber exploitation, frequent presence of pasture.
- **Group 3:** Vegetation in advanced successional stage, with species composition near to original, but diameter structure and structural parameters yet far from primary forests; vegetation with low human influence, absence of pasture and recent logging, wide diameter amplitude.

RESULTS

The main floristic and structure variables including successional group are shown for each cluster in Table 1, separately for the main *stratum* (DBH \geq 10cm) and regeneration *stratum* (all plants with total height >1.50m and DBH<10cm).

The distribution of stand variables within the set of 92 examined remnants is shown in Fig. 2. While species richness, diversity, density and dominant height were nearly normally distributed, basal area and richness of regeneration species distributions seemed to tend leftwards. Shannon and Wiener Index and density showed intermediate values in most of the remnants, although it is not possible to establish direct relationship between conservation status and diversity or density, for high values of both can occur in unperturbed as well as in perturbed forests.

Considering vegetation physiognomy, species composition and conservation status, 46 (50% of total) remnants were classified as belonging to Group 1, 41 (45%) belonged

TABLE 1
Main floristic and structure variables of 92 clusters measured by IFFSC between 2007 and 2008

Cluster	Main stratum DBH \geq 10cm									Regeneration		
	Alt. ¹	DE ²	S ³	H ⁴	J ⁵	D ⁶	BA ⁷	DBH ⁸	DH ⁹	D _{reg} ¹⁰	S _{reg} ¹¹	SG ¹²
69	1 091	82	19	1.67	0.57	515	24.47	17.9	12.3	2 900	8	2
70	1 089	150	22	2.09	0.68	418	12.07	16.1	6.7	1 000	6	1
85	990	313	30	2.23	0.66	730	43.54	24	11.2	1 300	8	1
87	867	29	39	3.15	0.86	503	22.8	20.7	10.9	1 500	5	2
89	1 255	111	24	2.52	0.79	733	33.77	19.8	13.6	8 200	22	2
104	825	627	51	3.2	0.81	780	27.84	19	18.3	9 700	26	2
113	1 430	201	18	2.08	0.72	538	25.28	18.8	9.6	3 300	12	1
114	1 449	103	12	1.5	0.6	345	29.31	28.6	13	4 500	17	2
134	1 068	107	31	2.76	0.8	685	32.71	20.4	12.2	1 000	6	2
137	1 359	82	29	2.8	0.83	770	32.29	19.8	11.4	4 600	11	2
139	1 475	146	8	1.33	0.64	190	8.62	20.3	9.2	900	3	1
140	1 560	146	10	1.64	0.71	303	20.2	22.6	8.1	2 300	5	1
177	826	120	37	2.85	0.79	643	23.11	18.2	15.7	13 000	28	1
188	1 231	294	34	1.67	0.47	1230	54.86	21.3	7.2	2 600	10	1
206	998	38	34	3.05	0.86	497	17.29	18.6	17.7	3 000	10	1
208	870	56	38	3.04	0.84	683	16.1	15.3	12.9	7 700	26	2
211	1 129	147	37	2.84	0.79	838	27.94	17.9	17.1	6 100	20	1
217	1 283	130	35	1.97	0.55	1310	50.75	20.1	10.6	2 400	15	2
242	1 110	250	16	2.29	0.83	268	7.55	16.8	11.2	1 600	8	1
246	1 015	372	34	2.93	0.83	375	22.11	24.3	15.8	0	0	2
297	1 051	143	41	2.84	0.77	970	28.15	17.3	16.4	7 600	25	2
301	891	51	14	2.38	0.9	85	3.87	21.6	13.2	3 600	13	1
304	863	299	26	2.71	0.83	289	21.26	26.7	12.9	2 700	9	1
328	850	101	50	3.47	0.89	858	30.08	18.9	17.6	900	4	1
336	840	100	31	3.06	0.89	406	10.27	15.9	11.7	8 800	20	1
337	888	167	40	2.8	0.76	595	19.02	16.7	12.3	5 200	14	2
365	941	192	37	3.02	0.84	378	16.44	20.8	17.8	6 500	14	2
367	846	148	43	3.28	0.87	915	25.58	17.3	16.5	11 000	28	1
369	811	100	39	3.01	0.82	573	19.02	18	16.7	8 100	18	1
409	962	115	54	3.48	0.87	846	29.02	19.1	16.4	4 700	17	1
413	1 150	106	50	3.29	0.84	553	36.76	24	15.2	800	5	3
415	849	176	48	3.5	0.9	535	27.59	22	11	2 900	14	1
450	908	185	49	3.31	0.85	550	25.27	20	17.2	8 100	13	1
453	993	100	50	3.21	0.82	733	27.92	18.7	15.7	3 700	20	1
455	1 070	213	35	1.98	0.56	1228	43.21	19.4	12	400	3	2
456	989	320	41	2.01	0.54	740	51.36	25.6	14.8	300	3	2
495	788	112	28	2.48	0.74	378	12.74	18.1	16.3	4 200	12	2
555	943	142	52	3.36	0.85	635	21.8	18.3	15.9	5 800	28	2
561	869	100	38	2.95	0.81	620	39.23	21.8	16.1	2 700	9	2
562	985	92	43	3.11	0.83	555	22.87	20.3	15.5	2 600	12	1
565	1 055	413	50	2.83	0.72	615	44.63	26.3	17.2	1 300	8	3
566	802	325	45	3.44	0.9	318	15.5	21.1	18.8	800	4	2
615	954	118	28	2.98	0.89	260	12.47	22.6	17.1	5 700	10	1
619	954	458	32	2.95	0.85	590	20.47	19.1	14.9	5 400	17	1
623	1 190	295	25	1.43	0.44	1228	64.1	23.4	10.4	0	0	2
668	1 089	45	44	3.41	0.9	533	22.97	20.1	16.2	9 700	15	1
669	1 012	63	40	3.08	0.83	553	12.95	13.7	12.5	7 000	21	1
672	970	117	30	2.65	0.78	663	20.31	17.2	16.3	3 000	13	1
673	1 025	107	34	2.87	0.81	473	30.08	25.4	19.1	300	3	2

TABLE 1 (Continued)
Main floristic and structure variables of 92 clusters measured by IFFSC between 2007 and 2008

Cluster	Main stratum DBH \geq 10cm									Regeneration		
	Alt. ¹	DE ²	S ³	H ⁴	J ⁵	D ⁶	BA ⁷	DBH ⁸	DH ⁹	D _{reg} ¹⁰	S _{reg} ¹¹	SG ¹²
732	1 175	123	44	3.19	0.84	743	24.7	18.2	15.2	1 400	10	2
736	1 287	187	25	2.33	0.72	973	32.94	17.8	10.8	1 300	7	2
739	572	82	46	3.05	0.8	718	20.8	17.3	13.6	4 100	21	1
784	1 348	92	38	2.69	0.74	778	46.9	24.4	16	1 400	5	2
789	1 026	142	28	2.6	0.78	483	13.6	16.5	11.9	5 500	16	1
793	1 110	202	37	3.07	0.85	510	12.01	15.7	12.5	300	3	1
794	682	146	62	3.46	0.84	515	24.47	21.6	16.8	1 600	8	2
843	1 269	162	24	1.05	0.33	768	42.88	24.1	12.3	0	0	2
845	1 060	250	30	2.75	0.81	469	28.28	21.8	15.3	300	3	1
850	560	100	42	3.25	0.87	353	14.81	19.7	16.2	800	4	1
887	1 047	565	30	2.53	0.74	670	32.34	20.4	14.4	200	2	2
894	867	80	42	3.02	0.81	470	12.26	14.8	15.4	300	2	2
895	861	171	44	3.32	0.88	713	20.87	16.2	18.4	3 100	17	2
901	943	489	56	3.43	0.85	770	42.53	20.6	15.5	4 400	18	3
902	979	690	35	3.13	0.88	238	27.4	29.4	16.8	6 500	10	2
933	1 098	250	46	2.97	0.78	698	20.18	17.1	12.6	800	6	1
934	1 143	284	53	2.82	0.71	710	30.09	21.3	14.9	2 800	19	1
939	866	327	54	3.51	0.88	393	24.41	23.8	17.3	1 100	5	2
945	865	96	37	2.95	0.82	430	14.6	16	17	5 500	21	2
946	932	187	35	2.6	0.73	583	17.47	17.1	13.4	400	4	1
949	865	104	28	2.64	0.79	438	19.35	21.8	18.8	800	5	2
974	1 011	81	46	3.43	0.9	373	19.46	21.5	15.7	1 700	10	1
976	820	250	49	3.25	0.84	575	22.16	17.4	17.8	2 400	16	2
978	1 226	953	39	2.39	0.65	543	33.55	25	18.7	0	0	3
979	800	134	37	2.95	0.82	510	16.01	17.6	15.1	2 300	19	1
982	845	122	29	2.71	0.8	245	14.23	22.2	20.6	800	7	1
984	740	401	19	2	0.68	390	22.83	23.1	19.8	500	2	1
1 010	807	95	16	1.98	0.71	273	16.69	20.1	11.2	800	2	1
1 013	813	369	15	2.22	0.82	135	15.25	34.1	23.1	7 900	9	1
1 016	896	152	26	2.67	0.82	308	24.84	26.6	17.6	1 700	7	1
1 019	903	149	38	3.34	0.92	283	17.57	25.4	17.6	5 000	9	2
1 024	1 005	111	47	3.3	0.86	713	17.41	16.4	12.3	5 100	25	2
1 034	835	134	17	2.23	0.79	108	7.16	25	13	0	0	1
1 042	846	114	26	2.87	0.88	295	18.51	23.5	16.8	2 300	15	1
1 055	796	144	21	2.15	0.71	408	19.32	21.7	21.4	800	6	1
1 059	792	83	47	3.34	0.87	575	26.21	19.8	22.8	2 700	14	2
1 061	879	110	40	2.81	0.76	710	29.89	19.9	23	4 700	19	1
1 063	861	153	47	2.77	0.72	825	36.19	19.9	17.1	4 300	21	1
2 001	824	250	30	2.36	0.69	598	33.6	23.2	19.6	1 400	11	2
2 002	806	52	31	2.75	0.8	710	39.68	22	20	1 500	8	2
3 001	842	496	42	3.13	0.84	635	26.51	20.3	20.5	4 000	9	2
3 002	821	929	21	2.27	0.75	470	26.51	21.9	18.3	300	2	2
4 000	995	795	31	1.73	0.5	1083	68.85	23.1	20.1	200	2	3
Average	980.2	209	35.3	2.73	0.78	578.4	25.75	20.6	15.3	3 200	11.1	

¹Altitude above sea level (m); ²Distance of clustercenter from the forest edge (m); ³Species richness within plot; ⁴Shannon and Wiener Index; ⁵Pielou evenness Index; ⁶Density (n/ha); ⁷Basal Area (m²/ha); ⁸Mean DBH (cm); ⁹Dominant height - mean of trees with height $> \bar{x} + sd$ (m); ¹⁰Density of regeneration (n/ha); ¹¹Species richness of regeneration within cluster; ¹²Successional group.

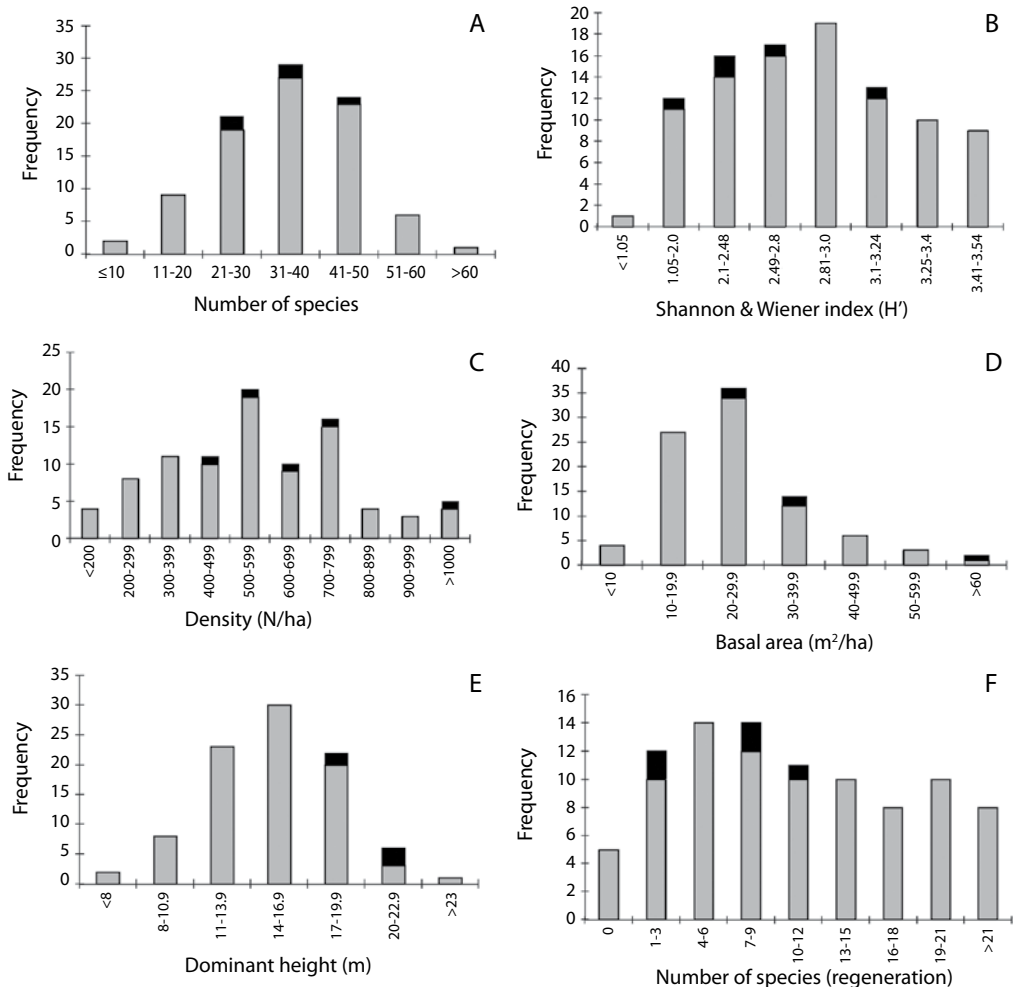


Fig. 2. Number of species (A), Shannon & Wiener Index (B), density (C), basal area (D), dominant height (E) and regeneration species number (F), from 92 sample plots (main stratum, DBH \geq 10cm); plots within conservation units in black.

to Group 2 and only five remnants (5%) to Group 3.

Classification of sampled remnants into the three successional groups (Table 2) showed that basal area (ANOVA, $\alpha < 0.01$) and density (ANOVA, $\alpha < 0.05$) were significantly different between groups. Furthermore analyzing basal area, Group 1 and 2 differed more significantly (Student, $\alpha < 0.01$) than Group 2 and 3 (Student, $\alpha < 0.05$). Considering density, only Group 1 and 2 showed significant difference (Student,

$\alpha < 0.05$). Yet neither DBH, nor dominant height or species richness, showed significant difference between successional groups. This indicates that basal area can be considered the most appropriate indicator of stand quality and conservation status in the observed universe within MOF.

Remnants located in public areas, supposed to be forests under better conservation conditions, did not have highest values of floristic and structure variables; they showed

TABLE 2
Mean values and standard deviation (sd) of vegetation variables by successional group

Group	BA ¹	sd	DBH ²	sd	DH ³	sd	S ⁴	sd	H ⁵	Sd	D ⁶	sd
1	21.1	9.4	20.2	3.5	14.8	3.7	33.1	12.3	2.85	2.06	515	238
2	28.5	11.5	20.7	3.5	15.6	3.1	36.5	10.2	2.88	2.12	633	248
3	45.3	13.0	23.8	2.1	17.3	2.1	45.2	10	2.90	2.36	713	226

¹Basal Area (m²/ha), ²DBH (cm), ³Dominant height (m), ⁴Species richness (S), ⁵Shannon and Wiener Index, ⁶Density (n/ha). Data from 92 analyzed remnants by IFFSC between 2007 and 2008.

higher values of basal area and dominant height than the average, but surprisingly lower species richness (not diversity) in the main *stratum* and in the regeneration *stratum* (Fig. 2), possibly due to ancient exploitation and degradation. Even in the site considered most closely to primary conditions (cluster n° 4 000), richness and diversity were quite lower than expected; this can mean that species richness is not automatically correlated to conservation status, besides the possible restriction that the applied sampling method was not appropriate and could have failed recording the existing plant diversity. On the other hand, due to the hegemony of *Dicksonia sellowiana* Hook (61.4% of relative density) and *A. angustifolia* (52% of relative dominance), this particular community (cluster n° 4 000) could represent a monodominant forest, due to special site conditions. In this case ideal site conditions for one or two species can be adverse or even exclusive for a lot of other species and reduce richness and diversity values.

Species richness and diversity: A total of 342 tree and shrub species and 28 morpho-species were detected in the sampled areas (five Coniferophyta, one tree fern, 364 Magnoliophyta), belonging to 66 families. There were 310 species in the main *stratum* (113 of them exclusives in this *stratum*) and 260 in the regeneration *stratum* (59 exclusives); the families with major richness were Myrtaceae, Asteraceae, Lauraceae, Fabaceae, Solanaceae, Rubiaceae, Euphorbiaceae and Melastomataceae. The species richness reflected the great

range of forest sites and environmental conditions within MOF in Santa Catarina.

The mean species richness in measured remnants was 35 (ranging from eight up to 62 tree species). Shannon and Wiener diversity index (H') varied between 1.05 and 3.48. Although beta and gamma diversity was high, the results showed that the individual fragments have low species richness (alfa diversity).

Only 13 species were present in more than 50% of sampled clusters (main *stratum*, in decrescent order of frequency): *A. angustifolia*, *Prunus myrtifolia* (L.) Urb, *Clethra scabra* Pers., *Vernonanthura discolor*, *D. sellowiana*, *Myrsine coriacea* (Sw.) R. Br. ex Roem. & Schult., *Matayba elaeagnoides* Radlk., *Ocotea puberula*, *Ilex paraguariensis*, *Casearia decandra* Jacq., *Sapium glandulosum* (L.) Morong, *Zanthoxylum rhoifolium* Lam., *Ocotea pulchella*; all them can be considered generalists and secondary species with wide range of environmental conditions. In contrast, a high number of species (178 equivalent to 57% of total species number) were present in less than six clusters.

The high number of rare species corroborates this assumption, showing 74 species occurring in only one cluster, and 120 species occurring in up to two clusters. Among these species, 44 occurred with only one specimen and 72 species were sampled with only two specimens.

Among them were found naturally rare species, like *Myrcia rupicola* D. Legrand, *Myrceugenia alpigena* (DC.) Landrum; yet also species highly impacted by exploitation like

Ocotea odorifera (Vell.) Rohwer and *Ocotea porosa* and indirectly impacted by pasture burning close to forest remnants like *Myrcogenia venosa* D. Legrand, certainly contributed to increase the number of rare species.

The floristic survey within and around the clusters resulted in an important increase of registered plant diversity. There were found another (exclusive) 378 species belonging to 84 families. The Lycophyta and Monilophyta groups were represented by 99 species (terricole and epiphyte) and the Magnoliophyta group by 279 species, being 65 trees and shrubs species undetected by the inventory, 33 lianas, 35 epiphytes, 145 terricole herbs and one parasite.

Species richness within the three successional groups showed increasing pattern from group one to group three in the main stratum and decrease in the regeneration stratum (Fig. 3).

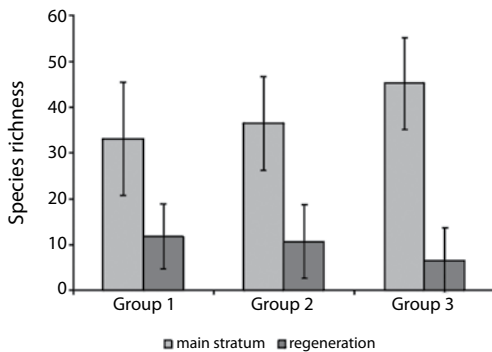


Fig. 3. Mean species richness within three successional groups and standard deviation (main stratum: DBH \geq 10cm; regeneration stratum: total height $>$ 1.50m, DBH $<$ 10cm).

Species composition and forest structure: Mean density of trees with DBH \geq 10cm was 578 individuals/ha (ranging from 85/ha to 1 310/ha, sd=248); medium basal area was 25.75m²/ha (3.87 to 68.85m²/ha, sd=12.06), dominant height 15.3m (6.7m to 23.1m, sd=3.44).

Considering Importance Values (IV), calculated for the whole set of 92 clusters, there

were 25 species with IV \geq 1% (8.0% of total number of species). This relatively restricted group of species represents 67.8% of all trees, 73.5% of total basal area and 60.6% of the total importance value. Collected data showed a relatively small group of species of high importance within horizontal structure in the MOF.

Within these species occurred, in decrescent order, *D. sellowiana*, *A. angustifolia*, *C. scabra*, *Lithraea brasiliensis* Marchand, *O. porosa*, *M. elaeagnoides*, *P. myrtifolia*, *O. puberula*, *O. pulchella* and *I. paraguariensis*. Two of them, *D. sellowiana* and *A. angustifolia*, are endangered and red list species (MMA, 2008).

Structure analysis by successional group showed that:

- **In group one dominated:** (highest IV, in decrescent order) *D. sellowiana*, *A. angustifolia*, *O. porosa*, *I. paraguariensis*, *Nectandra megapotamica* (Spreng.) Mez, *C. decandra*, *M. elaeagnoides*, *Allophylus guaraniticus* (A. St.-Hil.) Radlk., *Myrcia hatschbachii* D. Legrand;
- **In group two dominated:** *D. sellowiana*, *A. angustifolia*, *L. brasiliensis*, *P. lambertii*, *O. pulchella*, *M. elaeagnoides*, *C. scabra*, *O. porosa*, *I. paraguariensis*;
- **In group three dominated:** *D. sellowiana*, *A. angustifolia*, *L. brasiliensis*, *P. lambertii*, *O. pulchella*, *M. elaeagnoides*, *C. scabra*, *Sebastiania commersoniana* (Baill.) L. B. Sm. & Downs, *O. porosa*, *I. paraguariensis*.

Density showed increasing mean values from group one to three (515/633/713 ind./ha), the same tendency is observed for basal area (21.1/28.45/45.26m²/ha) and mean species richness (33/36/45 species/cluster), while regeneration richness (12/11/7 species/cluster) and density (3 528/2 884/1 340 ind./ha) were decreasing. In this context it has to be mentioned that edge effect could have influenced species composition and structure of analyzed forests, overlapping or masking the described succession processes. Than basal area ($\alpha<$ 0.01)

and dominant height ($\alpha < 0.05$) were positively correlated (Pearson) to edge distance (Table 1), while richness of regeneration *stratum* ($\alpha < 0.05$) was inversely correlated and the other structural variables did not show any relationship to the edge distance.

Diameter distribution: Diameter distribution ranged from 10cm to 165cm, with major concentration of trees (84.21% of total number) in lower diameter classes (Fig. 4). Besides tree size distribution in tropical forests is to be expected like reversed J shape, small number of big trees was quite astonishing. In the classes ≤ 30 cm were found 84.4% of total number of trees, in classes ≤ 40 cm, 94.2% and ≤ 50 cm, 97.7%. Therefore the overwhelming majority of sampled trees belonged to regrowth and to lower or intermediate diameter classes, probably due to intensive exploitation and successional stage of forest remnants.

Between all the tree species, only 10 (2.7%) showed diameters greater than 100cm: *A. angustifolia*, *Cedrela fissilis* Vell., *Cinnamomum amoenum* (Mez) Kosterm., *Ficus cestrifolia* Schott ex Spreng., *Ilex brevicuspis* Reissek, *Luehea divaricata* Mart., *Nectandra lanceolata* Nees, *Ocotea porosa*, *Ocotea pulchella* and *Sloanea monosperma* Vell. All of them, despite *F. cestrifolia*, were found with decreasing diameter distribution (reversed J shape), showing stability of these populations within MOF (Fig. 5).

On the other hand, 67.9 % of all species only occurred in lower diameter classes (< 40 cm). Within these species, the following ones had major density values (> 50 specimen/ha) and can be considered typical components of early and intermediate successional stages of MOF in Santa Catarina: *A. guaraniticus*, *Cupania vernalis* Cambess., *Ilex dumosa* Reissek, *Jacaranda puberula* Cham., *S. terebinthifolius*, *Symplocos uniflora* (Pohl) Benth., *Z. rhoifolium*, *Myrcia splendens* (Sw.) DC., *Inga lentiscifolia* Benth., *Lonchocarpus campestris* Mart. ex Benth., *Inga vera* Willd., *Schinus lentiscifolius* Marchand and *Drimys angustifolia* Miers.

The number of species present only in the lower diameter classes are shown in Fig. 6. Their number was relatively stable in the three successional groups, meanwhile number of late successional or climax species with decreasing diameter distribution was increasing accordingly to the successional stage of vegetation. This means that an increasing number of species reached stable population levels in later successional stages of the analyzed forest vegetation.

Regeneration: In the regeneration were found an average of 3 200 individuals/ha ($sd = 2\ 890$), belonging to 260 species and morphospecies, 59 of them were exclusive of this *stratum* and were not sampled in the main *stratum*. Absolute density of species was ranging from one to 120 individuals/ha, relative density varied from 0.03 to 3.5%, considering

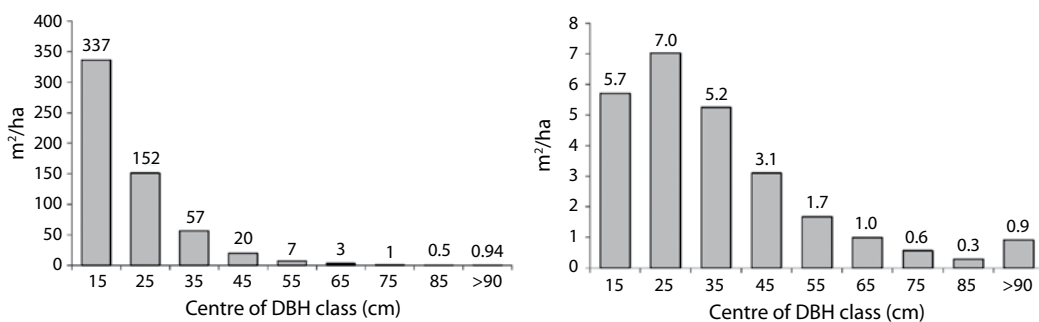


Fig. 4. Total diameter distribution for density ($N \cdot ha^{-1}$, left) and basal area ($m^2 \cdot ha^{-1}$, right) of main *stratum* ($DBH \geq 10$ cm) from 92 sample plots in MOF.

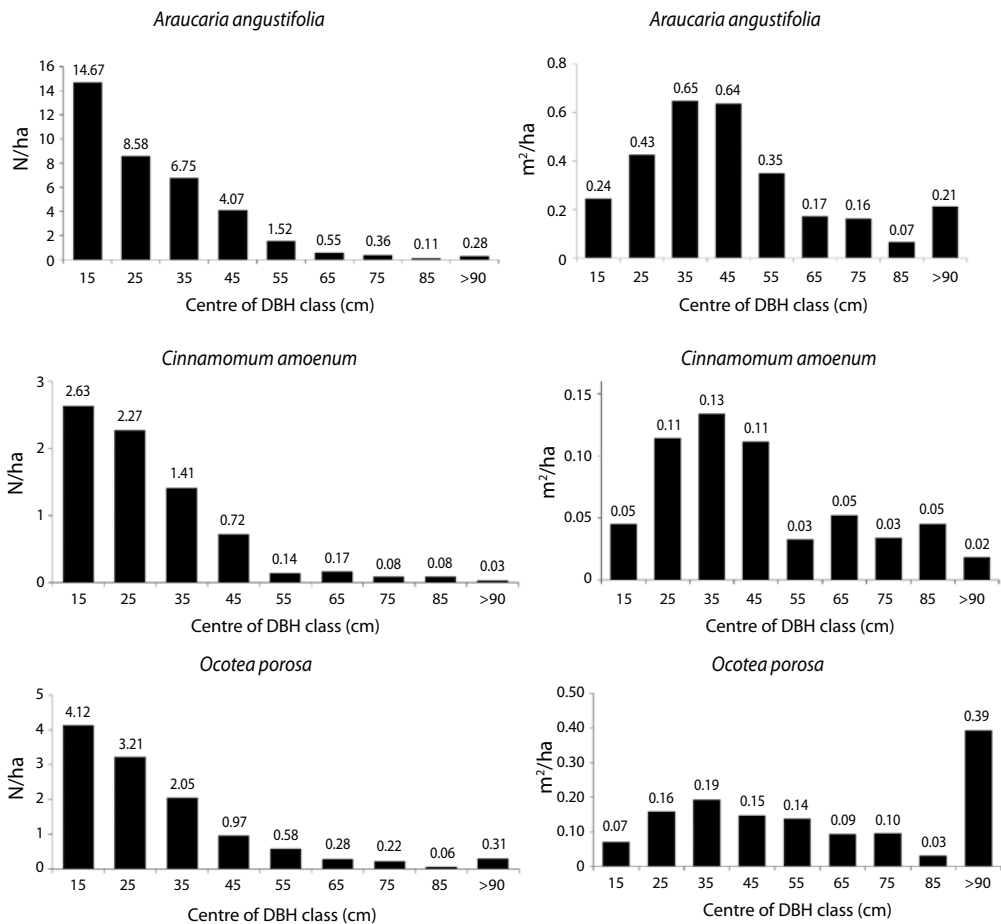


Fig. 5. Total diameter distribution for density ($N \cdot ha^{-1}$) and basal area ($m^2 \cdot ha^{-1}$) of *Araucaria angustifolia*, *Cinnamomum amoenum* and *Ocotea porosa* in MOF.

the whole data set of 92 clusters. Only 14 species (7.7%) were found with density major than 50 individuals/ha (in decrescent order): *Myrcia oblongata* DC, *Sebastiania brasiliensis* Spreng., *Solanum sanctaecatharinae* Dunal, *C. vernalis*, *C. decandra*, *Acca sellowiana* (O. Berg) Burret, *Solanum variabile* Mart., *Grazielia serrata* (Spreng.) R.M. King & H. Rob., *Gymnanthes concolor* (Spreng.) Müll. Arg., *M. elaeagnoides*, *S. commersoniana*, *A. angustifolia*, *V. discolor* and *I. paraguariensis*. The cited species summarized 30.8% of total density.

Considering species frequency, maximum absolute frequency was 20 (23%) and only a

small group of 28 (10.8%) species were present in at least 10% of the sampled clusters, meanwhile 183 species (or 70%) had been detected in less than 5% of the sampled forests; 97 species were found in a unique cluster. Within the most frequent and representative species were *C. decandra*, *A. angustifolia*, *C. vernalis*, *I. paraguariensis*, *M. elaeagnoides*, *S. commersoniana*, *Styrax leprosus* Hook. & Arn. and *D. brasiliensis*, but only some of these were also frequently founded in the main *stratum*. This means that MOF has diverse species compositions due to great variability of site conditions within the total area.

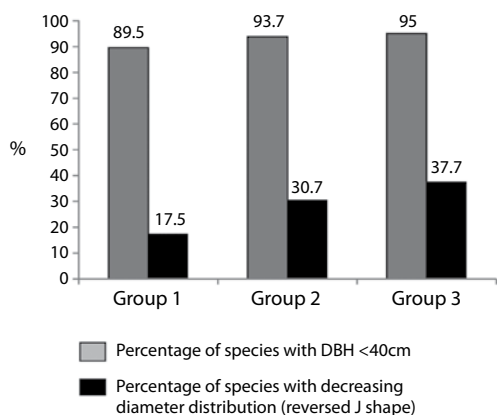


Fig. 6. Species with DBH<40cm and species with decreasing (reversed J shape) diameter distribution in % at MOF.

Data analysis of successional groups resulted in the following species compositions (most frequent species):

- **Group 1:** *M. oblongata*, *G. serrata*, *C. decandra*, *Myrciaria tenella* (DC.) O. Berg, *S. commersoniana*, *M. elaeagnoides*, *A. sellowiana*, *I. paraguariensis* and *Annona sylvatica* (A. St.-Hil.) Mart.
- **Group 2:** *Eugenia verticillata* (Vell.) Angelly, *M. hatschbachii*, *Sloanea guianensis* (Aubl.) Benth., *Cordia concolor* (Cham.) Kuntze, *Strychnos brasiliensis* (Spreng.) Mart., *Coussarea contracta* (Walp.) Müll. Arg., *Mollinedia clavifera* Tul., *Sorocea bonplandii* (Baill.) W. C. Burger, Lanj. & Wess. Boer, *Banara parviflora* (A. Gray) Benth. and *C. vernalis*;
- **Group 3:** *Myrsine gardneriana* A. DC., *C. scabra*, *Cinnamodendron dinitzii* Schwacke, *Myrceugenia myrcioides* (Cambess.) O. Berg, *A. guaraniticus*, *Rudgea jasminoides* (Cham.) Müll. Arg., *C. vernalis* and *D. sellowiana*.

The results showed that the most frequent species in the regeneration were different from the most important of the main *stratum*. For group 1 and 2 as initial and intermediate

stage vegetation types, these differences can be explained by the process of natural succession dynamic, which led to subsequent alteration of the dominant species pool and permanent substitution of early succession species by late succession components. On the other hand, in group 3 was expected that leading species of main *stratum* would be well represented also in regeneration *stratum*; their absence in regeneration can be explained either by deficit of fruit producing by or by weak conditions for late succession species in the understory, mostly due to manual clearing of understory vegetation and extensive cattle pasture, which is still a local practice to provide survival of cattle in wintertime when grasslands outside of forests does not offer sustain for the animals. It also had to be considered that part of the species of the regeneration *stratum* are understory specialists and absent in the upper stratum and therefore are not “regenerating” species.

Conservation status and human impacts:

Detailed site description revealed evidences of impacting and stress factors inside forests and in the surroundings (Table 3). Main pressure factors identified inside forests were selective logging, of more or less intensity, mostly not recent, but always without any silvicultural criteria and extensive cattle pasture (present in 51% of all fragments), damaging the understory and regeneration through grazing and trampling; these activities were often driven by vectors like (desirable and

TABLE 3
Impacting and degrading human factors observed in 92 sampled forest remnants

Factor	N° of remnants	%
Recent or historic selective logging	72	78.3
Extensive cattle pasture	50	54.3
Logging roads	25	27.2
Understory clearing	18	19.6
Exploitation of <i>Ilex paraguariensis</i>	11	12.0

More than one factor possible in one fragment.

sustainable) non-destructive exploitation of leaves and branches of *I. paraguariensis* (mate tea). Invasion of herbs, climbers and bamboo species (*Merostachys* spp. and *Chusquea* spp.) were consequences of excessively opened forest ecosystems.

In the forest surroundings most frequently detected land uses were cattle farming, afforestation with fast growing exotic species and agriculture (soybean, corn and wheat crops), as seen by the attractiveness of alternative land-use systems that increases pressure on forest lands. Native vegetation in good conservation status was rarely found. In the studied forest sites, grazing and cattle trampling may have caused differential mortality of herbs and woody seedling. The occasional penetrations of fire also may have caused significant changes in vegetation structure and composition at the edge affected area that penetrates, at least, up to 50m inside the forest. In the public areas were detected vestiges from hunters and exploitation of *I. paraguariensis*, but they did not suffer directly from logging, cattle pasture, understory clearing or burning, but indirectly from these impacts present in the surrounding areas. The ensemble of the main factors pressuring on forest remnants in Santa Catarina's MOF are depicted in Fig. 7.

DISCUSSION

Despite of high total number of species founded, species richness found by this study in most of the sampled forest remnants was considered low, comparing to data collected by Longhi *et al.* (2006), who found 75, 96 and 110 tree species in three well conserved MOF sites at São Francisco de Paula, in the state of Rio Grande do Sul. Considering the same sampling criteria (DBH \geq 10cm), Formento *et al.* (2004) registered 70 tree species and Figueiredo Filho *et al.* (2006) 108 tree species in primary remnants of MOF in Santa Catarina and Paraná state. The overwhelming majority of forest fragments of this study were highly impacted and impoverished, in process of progressive succession (regeneration) after exploitation near to clear cutting or in regressive succession due to repeated and constant, historic or recent, intensive human impact. Shannon and Wiener Index (H') showed similar average values in all the three groups (with huge standard deviation within the groups), despite of differences of species composition and ecological category of species in these groups. Murphy & Lugo (1986) observed that, besides the influences of pedoclimatic conditions, diversity of secondary forests depends on the type of previous use. Thus intense cropping or pasture, with short

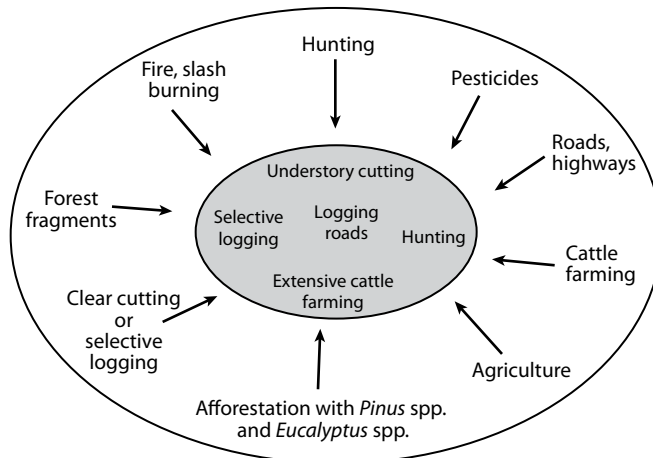


Fig. 7. Main environmental stress factors impacting inside (grey) and from outside on remnants of Santa Catarina's Mixed Ombrophyllous Forests.

regrowth periods and frequent use of fire, as mentioned by Lugo (1990) and Lamprecht (1989), leads to degradation of the original ecosystem, lagging or irreversibly obstructing the successional process. In the case of MOF, also intense and repeated selective logging without clearcutting apparently led to heavily degraded forest sites.

The high number of rare species (120 species occurring in only two clusters, 72 species with only two specimens) is not unusual in tropical forests, since their high α -diversity is directly related to the presence of rare species (Connell *et al.* 1984, Lepsch-Cunha, 1996). As pointed out by Cavallari Neto (2004), these species have evolved to be rare because their genetic flows are compatible with their rarity. Most of them colonize small gaps, have temporary seedlings, and show characteristics that are related with Janzen hypothesis. Some of them may be dependent of resource pulses linked to stochastic disturbance, as has been discussed in Tilman & Pacala (1993). In the studied area it seemed evident that besides the naturally rare species like *Myrcia rupicola*, *Myrceugenia alpigena* (DC.) Landrum (Legrand & Klein 1969, 1970, Klein 1990), many others like *Ocotea odorifera* and *O. porosa* have turned rare and threatened by anthropogenic stress factors as. On the other hand, also the wide spread of the sampled territory and his environmental gradients could have influenced the results (Longhi 1997).

It is notable that *D. sellowiana* was the predominant species in the understory, and can be considered the most important component of MOF in Santa Catarina, due to his elevated density, showing that this species seems to be in regeneration process within the remnants of MOF. It's high Importance Value (IV) in all successional groups, may be related to the fact that in eight clusters the communities were dominated by this species, forming monodominant forests, *sensu* Connell & Lowman (1989), with more than 50% of all upper story individuals and basal area. The overwhelming presence of this species generally occurred in areas with wet or even soggy soils in valleys

and plains, but also on slopes. These soil conditions can lead to monodominant tree species compositions, as observed by Peters *et al.* (1989), Steege (1994) and Arieira & Cunha (2006). Due to their high density many of these species in monodominant forests constitute important economic resources (Peters *et al.* 1989, Marimon & Felfili 2000), like *D. sellowiana* in Santa Catarina, which was intensively exploited during the twentieth century (Reitz *et al.* 1978, Windisch 2002).

At the same time, *A. angustifolia* as the second most important species, seems to be benefited by many degraded and open forest stands, that permitted her regrowth as a long living pioneer species.

Increase of species richness within the three successional groups showed from group 1 to group 3 in the main *stratum* (Fig. 3) is a tendency expected and consistent with current theory of ecosystem development and plant succession (Odum 1983), which supposes increasing richness, diversity and complexity of plant communities along the succession process. On the other hand, decreasing species richness in the regeneration *stratum*, also depicted in Fig. 3, has been explained by Lamprecht (1989); according to this author, the decrease of understory species richness during the succession process, caused by disappearance of light-demanding pioneers and also many ephemeral gap opportunist species, can be larger than the migration of new shade-tolerant species in areas with more homogenized light conditions in later successional stage.

Considering tree size distribution, most of the sampled trees belonged to lower or intermediate diameter classes and big trees are very rare. Species present in higher diameter classes were found with decreasing diameter distribution (reversed J shape). This diameter frequency distribution has been observed frequently in tropical forests, also in brazilian stands (Felfilli & Silva-Junior 1988, Oliveira-Filho *et al.* 1994, Felfilli 1997); yet diameter distribution of species, varies as a function of their environmental and successional context and of their populations' spatial pattern (Felfilli

1997, Marimon & Felfilli 2000). Some authors like Lieberman & Lieberman (1987) and Condit *et al.* (1993) singled out the great range of increase rates between tropical tree species, specially between understory species, due to the heterogeneity of light conditions.

Comparison between main *stratum* and regeneration *stratum* turned evident that pools of dominant species are different, in part due to succession processes in course after and during the action of degradation factors. While it can be expected that the species composition of both strata would be more or less approximate, it is worth noting what was reported by Guariguata & Ostertag (2001), who argue that the succession processes are strongly dependent on site characteristics, as well as on earlier land use. The same authors comment that the more intensive human impacts become, the potential for forest regeneration from seed bank decreases, making regeneration dependent on the inflow of new seeds, both from autochthonous and allochthonous sources. However, not all species will have life story characteristics that will favor their establishment, and it is likely that, in the case of this study, the autochthonous species do not possess characteristics that allow their easy regeneration. This means that although the characteristics of plant communities, such as the density and basal areas, are relatively predictable, the floristic composition along a chronological sequence is practically impossible to be predicted (Guariguata & Ostertag 2001, Chazdon *et al.* 2007, Letcher & Chazdon 2009) and therefore the described differences in the composition, could be expected.

Regeneration of upper story species is yet affected by clearing of understory vegetation through pasture and burning practices in agriculture land around the forests. The action of livestock in MOF in Santa Catarina has been described by Sampaio & Guarino (2007), who appointed to the processes of simplifying forest ecosystems and loss of biodiversity, benefitting pioneer species. Also apparently well conserved forests revealed in many cases extremely simplified structure with only young, small and not reproducing trees,

lack of regeneration, impoverished understory and forest floor vegetation, inducing regressive succession and leading to “empty forests”, as cited by Redford (1992) and Richards (1996). Besides their extreme impoverishment, ‘empty forests’ are frequently considered ‘closed forests’ by remote sensing monitoring, error caused by lack of detailed knowledge of relations between biophysical vegetation variables from field survey and data collected by remote sensing techniques.

On the other hand, in tropical forests some species show intense regeneration while others are rarely fruiting or their seedlings hardly succeed to establish, as observed by Whitmore (1998). The simultaneous occurrence of species typical to different successional stages has been frequently observed in tropical forests (Budowski 1965), yet the more frequent and prevailing species can serve to define their successional stage. Meanwhile the total absence of seedlings of red list species (MMA, 2008) *O. porosa* (Lauraceae) drew attention, once it was the main commercial timber species in Santa Catarina and composes the group of formerly frequent and dominant species that were actually turned rare by excessive and inordinate exploitation.

Considering results of species composition, forest structure data and field evidences, only five percent of fragments showed characteristics similar to primary forests, while 49% could be considered under medium impact and 47% under high external impact, mostly in secondary succession. Only six percent of sampled forests have got well conserved vicinity with favorable conditions towards conservation. Therefore Santa Catarina’s MOF must be considered threatened by isolation and internal impoverishment and structural simplification, due to over-logging and due to the presence of livestock inside the understory; these types of threatens seams to be performed mostly by small land owners. Lack of seed trees and seedlings of late succession and commercial tree species like *O. porosa* completes the ‘tableau’ of empty forest remnants, often identified as closed forests by remote sensing surveys. External pressures exercised on forests by

other land-uses like agriculture, cattle farming and afforestation with rapid growing species appear to be related to major land owners and private firms. For rare and endangered species, reduced area and isolation of forests stand for additional threaten leading to loss of genetic diversity by bottle-neck effects (Tarazi *et al.* 2010). Floristic surveys should be included in large scale inventories in order to increase databases and to recognize patterns of spatial distribution of rarer species.

Effective measures have to be submitted to decision makers, including financial compensation of forest owners, mostly small farmers, for their conservations efforts (maintaining forests and keeping livestock out of woodlands) and for providing environmental services like carbon stocking in vegetation and soil, so as water and soil protection. Besides educational measures and stimulatory actions towards conservation and sustainable use of forest resources, more efficient direct surveillance and protection has to be realized in order to combat or at least reduce illegal clear-cutting and timber logging. Finally, research has to focus on developing conservation indices and ecological indicators in order to define successful conservation strategies and priority areas within the mixed ombrophylous forests.

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RESUMEN

Este estudio es parte del Inventario Florístico Forestal de Santa Catarina, realizado para evaluar los recursos forestales, la composición de especies y la estructura de remanentes de bosque, y proporciona información para actualizar la conservación de los bosques y políticas de uso de la tierra en el estado brasileño de Santa Catarina (95 000km²). El inventario se aplica al muestreo sistemático, de 440 conglomerados en cuatro parcelas de 1 000m²

cada una, situados en una red de 10x10km. Dentro de las parcelas, todos los individuos leñosos (DAP \geq 10cm) fueron medidos. El estrato de regeneración (altura $>$ 1.50m, DAP $<$ 10cm) se registra en 100m² en cada conglomerado. Este estudio realiza un análisis de 92 conglomerados medidos en 2008 dentro de un área de 32 320km² de bosques ombrófilos mixtos con *Araucaria angustifolia* ubicados en el altiplano del estado. La densidad media (DAP \geq 10cm) es de 578 individuos/ha (desde 85/ha hasta 1 310/ha), la media de la riqueza de especies en los remanentes es de 35 (8-62), la diversidad (H') de Shannon y Wiener varía entre 1.05 y 3.48. A pesar de la alta diversidad total de especies (364 Magnoliophyta, cinco Coniferophyta y un helecho arborescente) y el alto promedio del área basal (25.75m²/ha, variando de 3.87 a 68.85m²/ha), la mayoría de los fragmentos de bosque se consideran altamente impactados por la tala, quema y ganadería extensiva, por lo tanto es necesario más medidas eficaces de protección.

Palabras clave: inventario forestal regional, *Araucaria angustifolia*, estructura forestal, impactos antrópicos en forestas.

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