

1	
2	
3	
4	
5	
6	
7	
8	
9	
10	Effect of oxy-combustion flue gas on mercury
10	
11	oxidation
12	
13	
14	
15	
16	
17	
18	Nuria Fernández-Miranda, M. Antonia Lopez-Anton*, Mercedes Díaz-Somoano, M.
19	Rosa Martínez-Tarazona
20	
21	
22	Instituto Nacional del Carbón (CSIC), Francisco Pintado Fe, 26, 33011, Oviedo, Spain
23	
24	
25	
26	
27	
28	*Corresponding author:
29	$\mathbf{P}_{\mathbf{r}} = -24.095.110000$
30 21	Phone: +34 985 119090
21	Fax: +34 985 29/002
32 22	Email. mailan@incar.csic.es
22 24	
34	
36	
37	
38	
39	
40	
10	

41 Abstract.

42 This study evaluates the effect of the gases present in a typical oxy-coal combustion 43 atmosphere on mercury speciation and compares it with the mercury speciation produced in conventional air combustion atmospheres. The work was performed at 44 laboratory scale at 150°C. It was found that the minor constituents (SO<sub>2</sub>, NOx and HCl) 45 significantly modify the percentages of  $Hg^{2+}$  in the gas. The influence of these species 46 47 on mercury oxidation was demostrated when they were tested individually and also 48 when they were blended in different gas compositions, although the effect was different 49 to the sum of their individual effects. Of the minor constituents, NOx were the main 50 species involved in oxidation of mercury. Moreover, it was found that a large 51 concentration of H<sub>2</sub>O vapor also plays an important role in mercury oxidation. Around 52 50% of the total mercury was oxidized in atmospheres with H<sub>2</sub>O vapor concentrations 53 typical of oxy-combustion conditions. When the atmospheres have similar concentrations of SO<sub>2</sub>, NO, NO<sub>2</sub>, HCl and H<sub>2</sub>O, the proportion of Hg<sup>0</sup>/Hg<sup>2+</sup> is similar 54 regardless of whether CO<sub>2</sub> (oxy-fuel combustion) or N<sub>2</sub> (air combustion) are the main 55 56 components of the gas.



- 59 Introduction.
- 60 61

Over the past three decades ever since the incident of Minamata Bay the effect of mercury on human health and ecosystems has continued to attract great interest. Mercury is a toxic element and, although mercury species are found in coal in trace concentrations, coal-fired power plants for energy production is one of the greatest anthropogenic sources of mercury emissions to the environment.<sup>1,2</sup> It is therefore imperative to take measures to ensure its control.<sup>3-6</sup>

68 Nevertheless, the main cause of alarm in coal fired power plants are the emissions of 69 CO<sub>2</sub> and several technologies are being developed for its capture and sequestration. 70 Among these, oxy-combustion is the most promising but, as with any relatively new 71 technology, many problems still remain unsolved. This is the case of the effects of 72 mercury species, whose presence in the oxy-combustion flue gases is not only an environmental problem, but also a technological problem. The elemental mercury (Hg<sup>0</sup>) 73 74 present in the flue gas can accumulate in the CO<sub>2</sub> compression units and damage the aluminium heat exchangers.<sup>7,8</sup> For these reasons, special attencion needs to be paid to 75 76 mercury oxidation in oxy-combustion processes, and new control measures need to be 77 investigated.

Because mercury behavior is influenced by its mode of occurrence, the efficiency of mercury control systems will depend on mercury speciation. The main mercury species present in the flue gas steam are in gas and particulate phases. The particulate mercury is mainly retained in the particulate control devices, whereas the capture of species in gas phase will depend on their reaction and adsorption behavior. Therefore, as a first step towards solving the problem it is necessary to determine in what way the composition of the flue gases produced in the oxy-fuel process affect mercury speciation and the differences that can be expected compared to air combustion
systems.<sup>9</sup>

It has been demonstrated that NO<sub>x</sub>, SO<sub>x</sub> and halogens play an important role in 87 mercury reactions in flue gases from coal combustion in air.<sup>10-13</sup> Therefore they can be 88 89 expected to influence mercury behavior in oxy-fuel facilities. Although the amount of  $SO_x$  is probably of the same order both in air combustion and in oxy-combustion,<sup>14</sup> 90 there is a wide variation in the SO<sub>3</sub>/SO<sub>2</sub> conversion rates.<sup>15,16</sup> Moreover, in oxy-91 92 combustion, as the flue gas is recycled, the concentration of SO<sub>2</sub> will be higher than in conventional air-coal combustion.<sup>17,18</sup> This could lead to a higher SO<sub>3</sub> concentrations in 93 the system which might inhibit mercury oxidation.<sup>9</sup> Consequently, a higher proportion 94 of Hg<sup>0</sup> can be expected under oxy-combustion than under air combustion. A higher 95 proportion of Hg<sup>0</sup> would then entail a lower retention in the Flue Gas Desulfuration 96 97 Systems (FGD) (if any) and increase the probability of amalgamation with aluminum which might cause damage to the CO<sub>2</sub> compression unit. The concentration of NOx 98 99 may also vary in different combustibles and plants. However, in oxy-fuel combustion, the levels of NOx will be lower than in air because the low  $N_2$  content in the boiler will 100 lead to the formation of less thermal NOx.<sup>16</sup> Several studies<sup>9,19</sup> have established that 101 102 mercury oxidation depends on the concentration of NO. For instance, it was found that 103 mercury oxidation increased when the NO concentration rose from 80 to 260 ppm, remaining constant above 550 ppm.<sup>9</sup> In addition, the H<sub>2</sub>O vapor concentration is 104 105 significantly higher in flue gas from oxy-combustion. In fact, the concentration of H<sub>2</sub>O 106 through recycling flue gases to the boiler can, in some cases, reach values of 30%. The role of water in the behavior of mercury has so far received little attention.<sup>8</sup> However, it 107 108 has been observed that the effect of H<sub>2</sub>O on mercury oxidation depends on its 109 concentration. It was found that a small amount of moisture can enhance oxidation

110 (<0.74%) whereas a large amount may have the opposite effect.<sup>20</sup> This could be due in 111 part to the interactions with other gases such as SO<sub>2</sub> and NO, which in some cases 112 favour oxidation.<sup>19,21-23</sup>

On the basis of knowledge acquired so far, the present work evaluates the effect of the components of a typical oxy-coal combustion atmosphere on mercury speciation and compares this behaviour with that of a typical air coal combustion atmosphere, focusing on the impact of different proportions of  $CO_2$  combined with  $O_2$ ,  $NO_x$ ,  $SO_2$ ,  $H_2O$  and HCl upon mercury speciation. The results provide a basis for the development of systems for controling mercury in oxy-combustion processes.

119

# 120 **2. Experimental.**

121 A schematic diagram of the experimental laboratory scale device employed in this122 study is shown in Figure 1. The experimental device consists of

- 123 i) a system to generate a stable concentration of  $Hg^0$  in the gas phase. A  $Hg^0$ 124 concentration of 100  $\mu g \cdot m^{-3}$  was produced by means of a VICI Metronic 125 calibrated permeation tube placed inside a glass "U" tube which was immersed 126 in a water bath at 40°C.
- 127 ii) a gas blending station where different gas compositions were prepared to 128 evaluate the effect of the gas components and their concentrations. Bronkhorst 129 mass flow controllers were used to control the gas flow into the system. 130 Atmospheres consisting of blends of two or more gases, -O<sub>2</sub>, CO<sub>2</sub> and H<sub>2</sub>O vapour in major concentrations and  $N_2,\,SO_2,\,HCl$  and  $NO_x$  in minor and trace 131 concentrations-, were incorporated to the Hg<sup>0</sup> gas stream to assess the influence 132 of each gas upon mercury speciation. The transfer lines for mercury are made of 133 134 PTFE (polytetrafluoroethylene) and for the rest of the gases of 316 stainless steel

136

with a diameter of  $\frac{1}{4}$  in. The temperature of the gases was maintained at 150°C. The residence time was 3.2 s<sup>-1</sup>.

a system for analyzing the mercury species. A continuous Hg gas analyzer (VM 137 iii) 3000) was used to monitor the  $Hg^0$ . The  $Hg^{2+}$  was determined by capturing it in 138 139 an ion exchanger resin (Dowex® 1x8), suitable for the selective extraction of  $Hg^{2+}$  species.<sup>24</sup> The resin was conditioned with a solution of HCl:H<sub>2</sub>O (1:1) at 140 90°C for 30 minutes and then filtered and dried. The resin was placed prior to the 141  $Hg^0$  continuous analyzer ((1) in Figure 1) in such a way that the total mercury 142 concentration was balanced. The Hg<sup>2+</sup> in the resin at the end of the experiments 143 was determined by means of an automatic mercury analyzer (AMA 254). The 144 Hg<sup>2+</sup> results obtained with the resin were compared with those obtained by the 145 standard Ontario Hydro (OH) method and a commercial speciation sorbent trap 146 147 (Ohio Lumex). The OH procedure is described in the standard ASTM D6784 test method. In accordance with this method, Hg<sup>2+</sup> is collected in impingers 148 containing a chilled aqueous KCl solution. Hg<sup>0</sup> is collected in subsequent 149 impingers: one impinger containing a chilled aqueous solution of HNO<sub>3</sub>/H<sub>2</sub>O<sub>2</sub> 150 151 and the other three containing chilled aqueous acidic solutions of  $KMnO_4$  ((2) in 152 Figure 1). The solutions are then analyzed by AMA 254. The Ohio Lumex sorbent trap allows  $Hg^{2+}$  and  $Hg^{0}$  to be captured separately. Moisture resistance 153 154 and acid gas scrubbers come standard with these traps in order to prevent interference with  $Hg^{2+}$  ((3) in Figure 1). 155

In order to identify possible reactions between gases other than mercury, the contents
of O<sub>2</sub>, CO<sub>2</sub>, SO<sub>2</sub>, NO and NO<sub>2</sub> in the gases were determined using a TESTO 350 Gas
Analyzer. Different gas compositions at different temperatures were evaluated.

The water vapour in the outlet of the experimental device was condensed and collected in two flasks prior to the VM analyzer. The sulphate, nitrite, nitrate and chloride retained in the condensate were determined by ionic chromatography.

162

# 163 **3. Results and discussion.**

164 3.1 Mercury speciation in gas phase.

165 A comparative study of the three methods employed to determine the mercury 166 speciation was carried out. The atmosphere was prepared using all the gases that might be produced during oxy-combustion in the following concentrations: 1000 ppm of SO<sub>2</sub>, 167 168 1000 ppm of NO, 100 ppm of NO<sub>2</sub>, 25 ppm of HCl, 4% of O<sub>2</sub>, 64% of CO<sub>2</sub>, 12% of H<sub>2</sub>O and balanced with N<sub>2</sub>. Figure 2 compares the percentage of Hg<sup>2+</sup> determined with 169 170 the resin, OH and trap methods. Similar results were obtained with the resin and OH 171 method. However, the result was slightly lower in the case of the trap sorbent. This may 172 be due to the fact that the experimental conditions of this work differ from the optimal 173 conditions for which the traps were designed. They were manufactured to be kept at 174 100°C and in atmospheres with low H<sub>2</sub>O vapor content. On the other hand, the flow rate 175 through these traps must be between 200-300 ml/min to avoid breakthrough on the KCl 176 bed whereas the flow rate used in this study was 500 ml/min. Moreover, the OH method 177 is tedious and employs solutions that can become unstable with time. Since resin is 178 easily conditioned, economically viable and produces accurate results in samples with a large amount of moisture,<sup>25-27</sup> the resin method was employed in this study to evaluate 179 mercury oxidation. 180

181 3.2. Effect of SO<sub>2</sub>.

182 The main difference between a typical coal combustion atmosphere and an oxy-183 combustion atmosphere is the higher CO<sub>2</sub> and lower N<sub>2</sub> concentration in the oxy-fuel 184 flue gas. In this study, the CO<sub>2</sub> content was 16% in the air combustion atmosphere and 185 approximately 70% in oxy-combustion. The atmospheres were balanced with the 186 reactive gases and N<sub>2</sub>. When the influence of SO<sub>2</sub> on mercury speciation was evaluated 187 in the two atmospheres: a) 1000 ppm SO<sub>2</sub>, 70% CO<sub>2</sub>, 4% O<sub>2</sub> and 26% N<sub>2</sub> and b) 1000 ppm SO<sub>2</sub>, 16% CO<sub>2</sub>, 4% O<sub>2</sub> and 80%  $N_{2}$  it was found that approximately 3% of  $\mathrm{Hg}^{2+}$ 188 189 was formed in both atmospheres. This percentage increased slightly to 4-5% when the SO<sub>2</sub> concentration was increased to 2000 ppm. As postulated in previous works<sup>13,21,23</sup> in 190 191 these simple atmospheres, mercury oxidation may have occurred as a consequence of 192 reaction (I).

193 
$$2Hg^{0}(g) + 2SO_{2}(g) + 2O_{2}(g) \rightarrow 2HgO(g) + 2SO_{3}(g)$$
 fast (I)

194 Because it is not possible to identify of HgO(g) to prove that this reaction occurs, the change in the SO<sub>2</sub>/SO<sub>3</sub> ratio was evaluated. To this end, the N<sub>2</sub>, CO<sub>2</sub>, O<sub>2</sub> and SO<sub>2</sub> 195 196 contents in the outlet gases were compared with the composition in the inlet gas, and it 197 was found that the SO<sub>2</sub> content decreased. Figure 3 shows the concentration of SO<sub>2</sub> in 198 the outlet stream when the inlet gas had 1000 ppm (Figure 3a) and 2000 ppm (Figure 199 3b) of SO<sub>2</sub> blended with O<sub>2</sub> and CO<sub>2</sub>. This decrease can be attributed to the oxidation of  $SO_2$  to  $SO_3$  according to reaction I. The mercury oxidation was around 3% Hg<sup>2+</sup> in both 200 201 atmospheres (16 and 70% CO<sub>2</sub>). In both cases the concentration of SO<sub>2</sub> was sufficient to 202 produce the oxidation of mercury through Reaction I.

203 3.3. Effect of NOx

Figure 4 shows the percentage of  $Hg^0$  and  $Hg^{2+}$  produced in an atmosphere with 1000 ppm of NO in different simple gas compositions containing N<sub>2</sub>, CO<sub>2</sub> and O<sub>2</sub> in different proportions. As can be observed,  $Hg^0$  was oxidized by NO in an atmosphere of N<sub>2</sub> in a proportion of ~15% which can be explained by reaction II. The same percentage of  $Hg^{2+}$  was produced in the atmospheres having 4% O<sub>2</sub> + N<sub>2</sub> and 16% CO<sub>2</sub> + N<sub>2</sub>. However, the oxidation was higher (~32%) when both CO<sub>2</sub> and O<sub>2</sub> were present in the gas blend, independently of CO<sub>2</sub> concentration.

211 
$$2 \operatorname{Hg}^{0}(g) + 2 \operatorname{NO}(g) \rightarrow 2 \operatorname{HgO}(g) + N_{2}(g)$$
 very fast (II)

The analysis of the outlet gas composition (Figure 5) revealed that some of NO was converted into NO<sub>2</sub> (Reaction III) and this transformation was higher in the presence of CO<sub>2</sub> (Figure 5b). For example, 1000 ppm of NO resulted in 50 ppm of NO<sub>2</sub> in an atmosphere with 70% CO<sub>2</sub> whereas only 5 ppm of NO<sub>2</sub> was formed in an atmosphere free of CO<sub>2</sub>. Therefore, the higher oxidation of mercury observed in the presence of NO, CO<sub>2</sub> and O<sub>2</sub> could have been due to the formation of a high proportion of NO<sub>2</sub> which would react with  $Hg^0$  according to reactions IV and/or V.

219 
$$2NO(g) + O_2 \rightarrow 2NO_2(g)$$
 fast (III)

220 
$$\operatorname{Hg}^{0}(g) + \operatorname{NO}_{2}(g) \rightarrow \operatorname{HgO}(g) + \operatorname{NO}(g)$$
 slow (IV)

221 
$$\operatorname{Hg}^{0}(g) + 2\operatorname{NO}_{2}(g) \rightarrow \operatorname{Hg}(\operatorname{NO}_{2})_{2}(g)$$
 slow (V)

Although in a combustion flue gas  $NO_2$  is usually present in lower concentrations than NO, the  $NO_2$  content could increase via reaction III in which case reactions IV and V would occur. To explore this possibility, the  $Hg^{2+}$  content was determined in an atmosphere with 100 ppm of  $NO_2$  blended with  $CO_2$ ,  $O_2$  and  $N_2$ . The results (Figure 6)

demonstrated the presence of  $Hg^{2+}$  in all cases. It is worth noting that i) oxidation was 226 higher than that produced by NO in an atmosphere with only N<sub>2</sub> (Figure 4) and that ii) 227 228 the oxidation decreased when CO<sub>2</sub> was present. The lower oxidation of mercury 229 produced in the presence of CO<sub>2</sub> could have been due to the conversion of NO<sub>2</sub> to NO or N<sub>2</sub>O<sub>4</sub> through reactions VI and/or VII<sup>28,29</sup>. Analysis of the gases at different 230 temperatures and concentrations showed that the decomposition of NO<sub>2</sub> to NO through 231 Reaction VI hardly occurred (Figure 7) whereas reaction VII may have produced. As 232 can be observed in Figure 8, the NO2 signal decreased when CO2 was present in the 233 234 atmosphere.

235 
$$2NO_2(g) \rightarrow 2NO(g) + O_2(g)$$
 slow (VI)

$$236 \qquad 2NO_2(g) \rightarrow 2N_2O_4(g) \qquad \qquad \text{slow} \qquad (VII)$$

According to the data described above, NO and NO<sub>2</sub> individually are able to oxidize mercury in different proportions (13-52%) depending on whether  $CO_2$  is present (Figures 4 and 6). However, when both gases were present in an atmosphere composed of 1000 ppm NO, 100 ppm NO<sub>2</sub>, 4%O<sub>2</sub>, 70%CO<sub>2</sub> and 26%N<sub>2</sub>, mercury oxidation decreased to 6±1%. This could have been due to:

i) a new reaction  $(Reaction VIII)^{15,29}$  yielding N<sub>2</sub>O<sub>3</sub> which is not able to oxidize mercury. This would explain why the analysis of the total NO<sub>x</sub> content was lower than the sum of NO and NO<sub>2</sub> (Figure 9)

245 
$$NO(g) + NO_2(g) \rightarrow N_2O_3(g)$$
 slow (VIII)

ii) a higher concentration of NO (1000 ppm) than NO<sub>2</sub> (100 ppm) which would tend
to drive reaction IV to the left making mercury oxidation less likely.

248 3.4. Effect of HCl.

Mercury oxidation does not take place in atmospheres with 25 ppm of HCl + 4%  $O_2$ in different proportions of CO<sub>2</sub>. Although theoretical models based on thermodynamic equilibrium data (HSC Chemistry) predict the formation of HgCl<sub>2</sub>(g) and HgCl(g), these reactions did not occur in the experimental conditions of this work due to kinetic limitations. These results agree with previous results obtained by the authors<sup>13,21,23</sup> and confirm that HCl does not influence homogeneous mercury oxidation.

$$3.5.$$
 Effect of the mixture of SO<sub>2</sub>, NO<sub>x</sub> and HCl.

257 In an atmosphere composed of 1000 ppm SO<sub>2</sub>, 1000 ppm NO, 100 ppm NO<sub>2</sub>, 25 ppm HCl,  $4\%O_2$ ,  $70\%CO_2$  and  $26\%N_2$  the percentage of Hg<sup>2+</sup> was 9±1%. Similar 258 259 percentages of oxidation were obtained when CO<sub>2</sub> was subsitute for N<sub>2</sub>. From a quantitative point of view, the results suggest that NOx and SO<sub>2</sub> have an additive effect 260 261 on mercury oxidation. It must be remembered that the oxidation of mercury by SO<sub>2</sub> and 262 NOx was approximately 3 and 6%, respectively, as a consequence of reactions I, II, IV 263 and V. It is remarkable that when HCl was removed from the atmosphere, the percentage of  $Hg^{2+}$  increased to 13±1% in both the atmosphere rich in N<sub>2</sub> (air 264 265 combustion) and the atmosphere rich in  $CO_2$  (oxy-combustion). Two reactions may 266 explain this behavior:

267 
$$\operatorname{NO}_2(g) + \operatorname{SO}_2(g) \to \operatorname{NO}(g) + \operatorname{SO}_3(g)$$
 fast (IX)

268 
$$4HCl(g) + 2NO(g) + 0.5O_2(g) \rightarrow 2Cl_2(g) + N_2O(g) + 2H_2O(g)$$
 very fast (X)

According to these reactions, the absence of HCl could be favouring reaction IX, and hence, an increase in mercury oxidation through the formation of NO. On the other hand the presence of HCl could give rise to reaction X which would consume NO in detriment to the production of  $Hg^{2+}$ . In both cases the absence of HCl would benefit mercury oxidation in the presence of NOx, SO<sub>2</sub>, O<sub>2</sub>, CO<sub>2</sub> and N<sub>2</sub>.

 $274 \qquad 3.6. \text{ Effect of } H_2O.$ 

275 As already mentioned, one of the most important differences between the 276 composition of a flue gas from coal combustion in air and from oxy-combustion is that 277 in the latter the concentration of H<sub>2</sub>O is much higher. The effect of different H<sub>2</sub>O vapor concentrations (7, 12 and 20%) in the atmosphere formed by 1000 ppm SO<sub>2</sub>, 1000 ppm 278 NO, 100 ppm NO<sub>2</sub>, 25 ppm HCl, 4%O<sub>2</sub>, 56-70%CO<sub>2</sub> and N<sub>2</sub> (simulated flue gas, SFG) 279 was evaluated. The  $Hg^{2+}$  percentages are shown in Figure 10. Mercury oxidation was 280 281 found to be higher (25 to 45%) in the atmospheres containing  $H_2O$  vapor than in the dry atmosphere (9%). In an atmosphere with smaller CO<sub>2</sub> and H<sub>2</sub>O concentrations (1000 282 283 ppm SO<sub>2</sub>, 1000 ppm NO, 100 ppm NO<sub>2</sub>, 25 ppm HCl, 4%O<sub>2</sub>, 16%CO<sub>2</sub>, 6%H<sub>2</sub>O and N<sub>2</sub>) as in the case of coal combustion in air, the  $Hg^{2+}$  obtained was  $30\pm9\%$ . Therefore, in all 284 285 cases mercury oxidation is favored by the presence of H<sub>2</sub>O.

Apart from homogeneous oxidation, which is the subject of this study, reactions in aqueous phase can also take place giving rise to the formation of acid gases.<sup>30</sup> These reactions may occur in any condensate produced at the cold points during the combustion process and also need to be taken into consideration. The most important reactions of this type are as follows:

291 
$$SO_3(aq,g) + H_2O(l,g) \rightarrow H_2SO_4(aq)$$
 fast (XI)

292 
$$2NO_2(g) + H_2O(l,g) \rightarrow HNO_2(aq) + HNO_3(aq)$$
 slow (XII)

293 
$$3HNO_2(aq) \rightarrow HNO_3(aq) + 2NO(g) + H_2O(l,g)$$
 fast (XIII)

294 
$$NO_2(g) + NO(g) + H_2O(l,g) \rightarrow 2HNO_2(aq)$$
 slow (XIV)

295 
$$N_2O_4(g) + H_2O(l,g) \rightarrow HNO_2(aq) + HNO_3(aq)$$
 fast (XV)

296 
$$N_2O_3(g) + H_2O(l,g) \rightarrow 2HNO_2(aq)$$
 fast (XVI)

To identify these possible reactions, the condensates obtained by cooling the gases collected prior to the VM 3000 in a flask (Figure 1) were analyzed by ion chromatography. These analysis corroborated the presence of sulphates, nitrites and nitrates. The higher oxidation of  $Hg^0$  in the presence of  $H_2O$  could have been a consequence of the above mentioned reactions in which, in addition to the formation of sulphates, nitrites and nitrates, NO may have produced (Reaction XIII) favouring the formation of HgO through reaction II.

As mentioned in section 3.5, in an atmosphere without HCl (1000 ppm SO<sub>2</sub>, 1000 304 ppm NO, 100 ppm NO<sub>2</sub>,  $4\%O_2$ ,  $12\%H_2O$ ,  $64\%CO_2$  and N<sub>2</sub>), the percentage of Hg<sup>2+</sup> is 305 306 higher (45%) than in the same atmosphere with HCl (37%) (Figure 10). This supports the view that, in addition to the reactions listed above, reactions IX and X would occur, 307 producing a higher amount of  $Hg^{2+}$ . In summary, mercury oxidation is promoted by the 308 presence of H<sub>2</sub>O, NO, NO<sub>2</sub> and SO<sub>2</sub> in the gas. The extent of this oxidation depends on 309 their relative proportions and concentrations and the presence of other gases, such as 310 HCl and CO<sub>2</sub>. This suggests that it is difficult to establish the exact percentage of  $Hg^{2+}$ 311 312 that might be produced in a combustion process. Nevertheless, from the results of this 313 work some generalizations can be made that will be helpful for modelling and 314 predicting the behaviour of mercury in each particular case. NO and NO<sub>2</sub> are involved 315 in mercury oxidation through several competing reactions which are affected by the 316 concentration of CO<sub>2</sub>. The presence of HCl together with the other gases decreases mercury oxidation. On the other hand, H<sub>2</sub>O favors the formation of Hg<sup>2+</sup> which can 317 318 reach values as high as 50%. In oxy-combustion where high amounts of H<sub>2</sub>O may be 319 present, oxidation may be higher than in conventional coal combustion. Independently 320 of the problems that the presence of a large amount of H<sub>2</sub>O originates in oxy-321 combustion, for tackling the problem of mercury, H<sub>2</sub>O would be an advantage. An increase in mercury oxidation implies a smaller amount of Hg<sup>0</sup> and therefore a lesser 322 risk of damage to the CO<sub>2</sub> compression units. 323

324

#### 325 Acknowledgments

The financial support for this work was provided by the project CTM2011-22921. The authors thank PCTI Asturias for awarding Nuria Fernandez-Miranda a pre-doctoral fellowship.

329

### 330 References

(1) Pacyna, E. G.; Pacyna, J. M.; Sundseth, K.; Munthe, J.; Kindbom, K.; Wilson,
S.; Steenhuisen, F.; Maxson, P. Global emission of mercury to the atmosphere from
anthropogenic sources in 2005 and projections to 2020. *Atmos. Environ.* 2010, *44*,
2487-2499.

(2) Rallo, M.; López-Antón, M. A.; Contreras, M. L.; Maroto-Valer, M. M.
Mercury policy and regulations for coal-fired power plants. *Environ. Sci. Pollut. R.* **2012**, *19*, 1084-1096.

(3) United States Environmental Protection Agency, National Emission Standards 338 339 for Hazardous Air Pollutants from Coal- and Oil-fired Electric Utility Steam Generating 340 Units and Standards of Performance for Fossil-Fuel-Fired Electric Utility, Industrial-341 Commercial-Institutional, and Small Industrial-Commercial-Institutional Steam 342 Generating Units. 2011; http://www.epa.gov/airquality/powerplanttoxics/pdfs 343 /proposal.pdf

344 (4) Council of the European Union, Council conclusions, Review of the
345 Community Strategy concerning Mercury, 3075th ENVIRONMENT Council meeting,
346 Brussels, 14 March 2011;

347 http://www.consilium.europa.eu/uedocs/cms\_data/docs/pressdata/en/envir/119867.pdf

348 (5) U.N.E.P. Global Mercury Assessment 2013: Sources, Emissions, Releases and
349 Environmental Transport, UNEP Chemicals Branch, Geneva, Switzerland, 2013.

350 (6) U.N.E.P. Mercury-Time to Act 2013, 2013.

351 (7) Santos, S. O. IEA Greenhouse. Gas R&D Programme Cheltenham, MEC7
352 Workshop DLCS, Strathclyde University, United Kingdom, 2010.

353 (8) Mitsui, Y.; Imada, N.; Kikkawa, H.; Katagawa, A. Study of Hg and SO3
354 behavior in flue gas of oxy-fuel combustion system. *Int. J. Greenh. Gas Con.* 2011, *55*,
355 5143-5150.

(9) Wu, H.; Liu, H.; Wang, Q.; Luo, G.; Yao, H.; Qiu, J. Experimental study of
homogeneous mercury oxidation under O2/CO2 atmosphere. *Proceedings of the Combustion Institute*, 2013, *34*, 2847-2857.

359 (10) Eswaran, S.; Stenger, H. G. Effect of halogens on mercury conversion in SCR
360 catalysts. *Fuel Process. Technol.* 2008, 89, 1153-1159.

- 361 (11) Van Otten, B.; Buitrago, P. A.; Senior, C. L.; Silcox, G. D. Gas-phase
  362 oxidation of mercury by bromine and chlorine in flue gas. *Energy Fuel.* 2011, *25*, 3530363 3536.
- 364 (12) Zhou, J.; Luo, Z.; Hu, C.; Cen, K. Factors impacting gaseous mercury
  365 speciation in postcombustion. *Energy Fuel.* 2007, *21*, 491-495.
- 366 (13) Fuente-Cuesta, A.; López-Antón, M. A.; Díaz-Somoano, M.; Martínez367 Tarazona, M. R. Retention of mercury by low-cost sorbents: Influence of flue gas
  368 composition and fly ash occurrence. *Chem. Eng. J.* 2012, *213*, 16-21.
- 369 (14) Anheden, M.; Yan, J.; De Smedt, G. Denitrogenation (or oxyfuel concepts).
  370 *Oil & Gas Science and Technology*-Rev. IFP, **2005**, *60*, 485-495.
- 371 (15) Stanger, R.; Wall, T. Sulphur impacts during pulverized coal combustion in
  372 oxy-fuel technology for carbon capture and storage. *Prog. Energ. Combust.* 2011, *37*,
  373 69-88.
- 374 (16) Koornneef, J.; Ramirez, A.; Van Harmelen, T.; Van Horssen, A.; Turkenburg,

W.; Faaij, A. The impact of CO2 capture in the power and heat sector on the emission

of SO2, NOx, particulate matter, volatile organic compounds and NH3 in the European

- 377 Union. Atmos. Environ. 2010, 44, 1369-1385.
- 378 (17) Croiset, E.; Thambimuthu, K. V.; Palmer, A. A coal combustion in O2/CO2
- 379 mixtures compared with air. Can. J. Chem. Eng. 2000, 78, 402-407.
- 380 (18) Croiset, E.; Thambimuthu, K. V. NOx and SO2 emission from O2/CO2
- 381 recycled coal combustion. *Fuel*, **2001**, *80*, 2117-2121.

- (19) Niksa, S.; Helble, J. J.; Fujiwara, N. Kinetic modelling of homogeneous
  mercury oxidation: the importance of NO and H2O in predicting oxidation in coalderived systems. *Environ. Sci. Technol.* 2001, *35*, 3701-3706.
- 385 (20) Wang, Z. H.; Jiang, S. D.; Zhu, Y. Q.; Zhou, J. S.; Zhou, J. H.; Li, Z.S.; Cen,
- 386 K.F. Investigation on elemental mercury oxidation mechanism by non-thermal plasma
- 387 treatment. *Fuel Process. Tehcnol.* **2010**, *91*, 1395-1400.
- 388 (21) Rodríguez-Pérez, J.; López-Antón, M. A.; Díaz-Somoano, M.; García, R.;
- Martínez-Tarazona, M. R. Regenerable sorbents for mercury capture in simulated coal
  combustion flue gas. *J. Hazar. Mater.* 2013, *260*, 869-877.
- 391 (22) Zhao, Y.; Mann, M. D.; Olson, E. S.; Pavlish, J. H.; Dunham, G. E. Effects of
  392 sulphur dioxide and nitric oxide on mercury oxidation and reduction under
  393 homogeneous conditions. *J. Air Waste Manage. Assoc.* 2006, *56*, 628-635.
- 394 (23) Abad-Valle, P.; López-Antón, M. A.; Díaz-Somoano, M.; Martínez-Tarazona,
- M. R. The role of unburned carbon concentrates from fly ashes in the oxidation and
  retention of mercury. *Chem. Eng. J.* 2011, *174*, 86-92.
- 397 (24) Fuente-Cuesta, A.; Diaz-Somoano, M.; Lopez-Anton, M. A.; Martinez398 Tarazona, M. R. Oxidised mercury determination from combustion gases using an ionic
  399 exchanger. *Fuel*, 2014, *122*, 218-222.
- 400 (25) Cañada Rudner, P.; Garcia de Torres, A.; Cano Pavon, J. M.; Sanchez Rojas,
  401 F. On-line preconcentration of mercury by sorption on an anion-exchange resin loaded
  402 with 1,5-bis[(2-pyridyl)-3-sulphophenyl methylene] thiocarbonohydrazide and
  403 determination by cold-vapour inductively coupled plasma atomic emission. *Talanta*,
  404 **1998**, *46*, 1095-1105.
- 405 (26) De Wuilloud, J. C. A.; Wuilloud, R. G.; Olsinaa, R. A.; Martinez, L. D.
  406 Separation and preconcentration of inorganic and organomercury species in water

407 samples using a selective reagent and an anion exchange resin and determination by
408 flow injection-cold vapor atomic absorption spectrometry. *J. Anal. At. Spectrom.* 2002,
409 *17*, 389–394.

- 410 (27) Metzger, M.; Braun, H. In-situ mercury speciation in flue gas by liquid and
  411 solid sorption systems. *Chemosphere* 1987, *16*, 821-832.
- 412 (28) Patnaik, P. *Handbook of Inorganic Chemicals*, McGraw-Hill Companies,
  413 United States of America, 2003.

414 (29) Ting, T.; Stanger, R.; Wall, T. Laboratory investigation of high pressure NO
415 oxidation to NO2 and capture with liquid and gaseous water under oxy-fuel CO2
416 compression conditions. *Int. J. Greenh. Gas Con.* 2013, *18*, 15-22.
417 (30) Liémans, I.; Alban, B.; Tranier, J.-P.; Thomas, D. SOx and NOx absorption

418 based removal into acidic conditions for the flue gas treatment in oxy-fuel combustion.

419 *Energy Procedia* **2011**, *4*, 2847-2854.

420

421

422

# 424 Figure captions

- 425 426
- 427 **Figure 1**. Schematic diagram of the experimental device.
- 428 Figure 2. Percentages of  $Hg^0$  and  $Hg^{2+}$  in the presence of 1000ppm SO<sub>2</sub>, 1000ppm NO,
- 429 100 ppm NO<sub>2</sub>, 25 ppm HCl, 4%O<sub>2</sub>, 12% H<sub>2</sub>O, 64%CO<sub>2</sub> and N<sub>2</sub> retained by the resin,
- 430 Ontario Hydro (OH) and trap method
- 431 **Figure 3.** Analysis of the SO<sub>2</sub> concentration in the presence of O<sub>2</sub> and CO<sub>2</sub>
- 432 Figure 4. Percentages of  $Hg^0$  and  $Hg^{2+}$  in the presence of NO, O<sub>2</sub>, CO<sub>2</sub> and N<sub>2</sub>
- 433 determined by VM analyser and resin, respectively.
- 434 **Figure 5.** Conversion of NO to NO<sub>2</sub> at 150°C in the presence of a) 4%O<sub>2</sub> and N<sub>2</sub> and b)
- 435 4%O<sub>2</sub> and 70% CO<sub>2</sub>
- 436 Figure 6. Percentages of  $Hg^0$  and  $Hg^{2+}$  in the presence of NO<sub>2</sub>, O<sub>2</sub>, CO<sub>2</sub> and N<sub>2</sub>
- 437 determined by VM analyser and resin, respectively.
- 438 **Figure 7.** Conversion of NO<sub>2</sub> to NO at different temperatures
- 439 Figure 8. Analysis of the  $NO_2$  concentration in the presence of  $O_2$ ,  $CO_2$  and  $N_2$
- 440 Figure 9. Analysis of the  $NO_x$  concentration in the presence of  $O_2$ ,  $CO_2$  and  $N_2$
- 441 Figure 10. Percentages of  $Hg^0$  and  $Hg^{2+}$  in the presence of 1000ppm SO<sub>2</sub>, 1000ppm
- 442 NO, 100 ppm NO<sub>2</sub>, 25 ppm HCl, 4%O<sub>2</sub>, 70%CO<sub>2</sub> and N<sub>2</sub> (SFG: simulated flue gas) with
- 443 different amounts of H<sub>2</sub>O determined by VM analyser and resin, respectively.
- 444
- 445















465 Figure 4



470 Figure 5













487 Figure 9



492 Figure 10