

Citation for published version:

Plosz, B, Climent, J, Griffin, C, Chiva, S, Mukherjee, R, Penkarski-Rodon, E, Clarke, M & Valverde-Perez, BO 2020, 'Hindered and compression solid settling functions – sensor data collection, practical model identification and validation', *Water Research*, vol. 184, 116129. <https://doi.org/10.1016/j.watres.2020.116129>

DOI:

[10.1016/j.watres.2020.116129](https://doi.org/10.1016/j.watres.2020.116129)

Publication date:

2020

Document Version

Peer reviewed version

[Link to publication](#)

Publisher Rights

CC BY-NC-ND

University of Bath

Alternative formats

If you require this document in an alternative format, please contact:
openaccess@bath.ac.uk

General rights

Copyright and moral rights for the publications made accessible in the public portal are retained by the authors and/or other copyright owners and it is a condition of accessing publications that users recognise and abide by the legal requirements associated with these rights.

Take down policy

If you believe that this document breaches copyright please contact us providing details, and we will remove access to the work immediately and investigate your claim.

Hindered and compression solid settling functions – sensor data collection, practical model identification and validation

Benedek G. Plósz^{1,3,*}, Javier Climent², Christopher T. Griffin¹, Sergio Chiva², Rani Mukherjee¹, Elena Penkarski-Rodon³, Matthew Clarke¹, and Borja Valverde-Pérez³

¹Department of Chemical Engineering, University of Bath, Claverton Down, Bath BA2 7AY, UK (Email: b.g.plosz@bath.ac.uk; chris.t.griffin@bath.edu; rm2100@bath.ac.uk; mjc62@bath.ac.uk)

²Universitat Jaume I, Department of Mechanical Engineering and Construction, Av. Vicent Sos Baynat, s/n 12071 Castellón (Spain), (Email: jcliment@uji.es, schiva@uji.es)

³Dept. of Environmental Engineering, Technical University of Denmark, Bygningstorvet, Building 115, 2800 Kgs. Lyngby, Denmark (Email: bvape@env.dtu.dk).

* *Corresponding author*

Abstract

Secondary settling tanks (SSTs) are the most hydraulically sensitive unit operations in activated sludge water resource recovery facilities (WRRF). Mathematical models for predicting activated sludge solids settling velocity include parameters that show irreducible epistemic uncertainty. Therefore, reliable and periodic calibration of the settling velocity model is key for predicting activated sludge process capacity, thus averting possible failures under wet-weather flow- and filamentous bulking conditions. The two main knowledge gaps addressed here are: (1) Do constitutive functions for hindered and compression settling exist, for which all velocity parameters can be uniquely estimated? (2) What is the optimum sensor data requirement of developing reliable settling velocity functions? Innovative settling column sensor and full-scale data were used to identify and validate amended Vesilind function for hindered settling and a new exponential function for compression settling velocity using one-dimensional and computational fluid dynamics

30 simulations. Results indicate practical model identifiability under well-settling and filamentous
31 bulking conditions.

32

33 **Keywords**

34 Hindered and compression solid settling velocity; Compression solid concentration and effective
35 solid stress; settling column sensor; one-dimensional model; computational fluid dynamics (CFD);
36 practical model identification.

37

38

39 **1. INTRODUCTION**

40 The impact of climate change on influent flow conditions, arising from snow melting and storms
41 events, will require adaptation on the part of water resource recovery facilities (WRRFs) to maintain
42 stringent water quality standards, in the future. The increasing frequency of hydraulic shock events –
43 as a result of climate change – necessitates more effective operation and control of secondary settling
44 tanks (SSTs) in WRRFs (Ramin et al., 2014a). Theoretically, the maximum permissible SST loading
45 capacity determines the maximum permissible hydraulic WRRF load. However, the SST capacity
46 varies with activated sludge settleability. As such, stable operation and control necessitates effective
47 sensor technology and identifiable simulation models (Jeppsson et al., 2013; Plósz et al., 2009).
48 Settling sensors should ideally provide experimental data for estimating settling velocity parameters;
49 yet, up to date, no simple and robust methods exist to calibrate hindered and compression settling
50 parameters. Parameter identifiability of activated sludge settling velocity models therefore remains a
51 challenge. Experimental data typically collected during offline sludge settleability monitoring (e.g.,
52 sludge volume index) are unreliable (e.g., Wágner et al., 2015) and insufficient means, considering
53 the complexity of settling velocity models. In contrast, De Clercq et al. (2005) present the hitherto
54 most complex observations on solids concentration profiles in batch settling of activated sludge using
55 solids radiotracer and gamma cameras – a technique, whilst capable of revealing hindered and

56 compression settling behaviours in high resolution, deemed too expensive to be implemented in
57 WRRFs. To improve data collection, Vanrolleghem *et al.* (1996) propose recording batch settling
58 curves using a scanner to measure the sedimentation of the sludge blanket over time (SettloMeter).
59 Furthermore, Derlon *et al.* (2017) present a cost-effective camera-based method to monitor sludge
60 blanket height (*SBH*). Ramin *et al.* (2014b) propose a sensor setup with a TSS sensor installed in the
61 bottom of a settling column, thus inferring *SBH* and the TSS concentration (X_{bottom}) time-series.
62 Valverde-Pérez *et al.* (2017) demonstrate, however, that *SBH* and X_{bottom} time-series do not provide
63 sufficient information for reliable identification of the settling velocity model by Ramin *et al.* (2014b),
64 and propose a novel multi-probe sensor setup, monitoring TSS concentration at multiple heights at
65 the side of the column sensor (X_{side}), besides *SBH*. Despite the extensive experimental data measured,
66 it is still not guaranteed that a unique set of model parameters can be reliably estimated for constitutive
67 functions for hindered and compression settling velocity. Results obtained using state-of-the-art
68 settling velocity models (Li and Stenstrom, 2016; Ramin *et al.*, 2014b) still suggest limitations in
69 terms of practical identifiability of compression settling velocity model parameters. Consequently,
70 besides experimental planning that proposes additional measurements to reduce parameter
71 uncertainties, model reduction is also necessary to adjust model complexity to the information
72 provided in the experimental data. To this end, Guyonvarch *et al.* (2015) assess the setting of the
73 variable compressive threshold concentration (X_C) parameter using state-of-the-art models (Bürger *et*
74 *al.*, 2013; Ramin *et al.*, 2014; De Clercq, 2006, De Clercq *et al.*, 2008). Setting X_C as a function of the
75 initial solid concentration and the SST feed solid concentration for simulating batch tests and SST,
76 respectively, is found superior over other methods (see also Supporting Information).

77 In the Bayesian framework, the parameter θ is treated as a random quantity with a specific prior
78 distribution $p(\theta)$, from which we can obtain the posterior distribution $p(\theta | x)$ via Bayes theorem, with
79 x denoting the input data. If the $p(\theta | x)$ are in the same probability distribution family as that for the

80 $p(\theta)$, the prior and posterior are then called conjugate distributions (Raiffa and Schlaifer, 1961;
81 Bernardo, 2000). Latin-Hypercube-Sampled priors for Simplex (LHSS) is a global approximate
82 Bayesian optimisation method, whereby uniform probability distribution of priors is used (Wágner et
83 al., 2016). One of the criteria for practical model identifiability in LHSS is that $p(\theta | x)$ is of
84 normal/Gaussian distribution. The question arises whether, once practical identifiability is established
85 through LHSS, results obtained in terms of uniqueness of normal posterior parameter distribution and
86 mean parameter values for each model structure with the available data series, estimates could be
87 improved by considering normally distributed (conjugate) priors in a subsequent run of parameter
88 estimation – a focal area chosen for this study.

89 As for the sources of uncertainty associated with settling velocity model identification, the design
90 of settling column setups can significantly influence measured data and thus the parameter estimates
91 (Vanrolleghem et al., 1996; Ekama et al., 1997). More research is still needed to understand how the
92 impact of column size affects model parameter estimates. Therefore, this study also addresses this
93 uncertainty source represented by the approach of using 1-D simulation models for estimating model
94 parameters, which are then used to calibrate CFD simulation models with higher complexity. As for
95 model validation, triangulation is the strategic use of multiple inquiries to address the same question,
96 each depending on different set of assumptions with their strengths and weaknesses (Lawlor et al.,
97 2016). Results agreeing across different inquiries are more likely to be replicated reliably.

98 The aims set in this study are (1) identifying constitutive functions for hindered-compression
99 settling velocity for which all parameters can be estimated using the sensor data with both good
100 settling and moderate filamentous bulking; (2) evaluating the feasibility of the sensor setup as a means
101 to infer experimental data on the effective compressive solid stress; (3) assessing uncertainty sources
102 associated with the model identification method and the settling column design; and (4) evaluating
103 and validating the new settling velocity constitutive functions using the triangulation approach.

104

105

106 2. MATERIALS AND METHODS

107

108 2.1. Sampling and settling sensor setup

109 Activated sludge samples were collected in two WRRFs in Denmark (Fredericia and Avedøre
110 WRRFs) and one in Sweden (Ellinge WRRF) with well-settling characteristics (Fredericia and
111 Ellinge with $SVI_{3,5} \leq 90$ ml/g) and moderate filamentous bulking (Avedøre, $SVI_{3,5} \sim 200$ ml/g). The
112 three activated sludge processes differed in terms of operating conditions. Secondary biological
113 treatment in Avedøre WRRF (320 000 PE – mostly municipal sewage) and Fredericia WRRF (350
114 000 PE – municipal sewage with significant industrial contribution) were operated at solids retention
115 time, $SRT=10-15$ days, and used polymers and chlorination for bulking control, respectively. Ellinge
116 WRRF (330 000 PE – mostly food industrial wastewater) was operated as a high-rate system, $SRT \sim 2$
117 days, without any bulking control measure taken. Settling tests were carried out using the multi-probe
118 sensor prototype (Valverde-Pérez *et al.*, 2017), consisting of a column (height: 0.7 m; bottom
119 diameter: 0.2 m) equipped with TSS SOLITAX (Hach, USA) infrared sensors installed at 0.21m
120 height in the side-wall (X_{Side}) and in the bottom of the column (X_{Bottom}). Measurements in a full-scale
121 SST were carried in the OBVA WRRF, Vila-Real, Spain (46 773 PE – mostly municipal sewage)
122 was operated at $SRT \sim 10$ days under moderate filamentous bulking conditions ($SVI_{3,5} \sim 270$ ml/g) and
123 without any bulking control measure taken. A SOLITAX and a SONATAX (Hach, USA) probes
124 were used to measure the SBH and the TSS in the bottom of the SST, respectively. For measuring
125 the SBH, the threshold TSS concentration was set to 0.3 kg m^{-3} . Radial velocity was measured in the

126 SST using a Vectrino (Nortek, USA) high-resolution acoustic velocimeter. More information on the
127 sensor positions in the SST is shown in the SI.

128

129 **2.2. Regression analysis to estimate X_{Infi} parameter**

130 Values of the maximum solids concentration parameter (X_{Infi} , kg m⁻³) are estimated with the X_{Bottom}
131 data series (e.g., Fig. 1b) obtained for each settling experiment using the regression equation

$$132 X_{Bottom}(t) = f_X + (X_{Infi} - f_X) \cdot (1 - e^{-k_X t}), \quad (1)$$

133 in the software, SigmaPlot 13 with k_X and f_X , denoting additional regression parameters.

134

135 **2.3. 1-D simulation model**

136 Solids settling in the column sensor is described as a PDE of second-order parabolic type as

$$137 \frac{\partial X}{\partial t} = -\frac{\partial(v_H(X) \cdot X)}{\partial z} + \frac{\partial}{\partial z} \left(D_{Comp}(X) \frac{\partial X}{\partial z} \right), \quad (2)$$

138 where X is the solid concentration, z is the vertical direction variable, v_H is the hindered solid
139 settling velocity, D_{Comp} denotes the compression settling (Eq. 5). The numerical scheme applied in
140 the simulation model implementation – in MATLAB® (Mathwork, Natick USA) – is according to
141 Guyonvarch *et al.* (2015). Briefly, the second-order PDE is discretized using 60 layers. The
142 Godunov approach is used to comply with the minimum settling flux conditions (e.g., Plósz *et al.*,
143 2007). The height of the sludge blanket is calculated as the distance between the bottom layer and
144 the first layer where the concentration is reduced below 0.9 g l⁻¹. Constitutive functions for v_H and
145 D_{Comp} are those shown in Eq. 4 and Eq. 6, respectively.

146

2.4. Practical model identifiability analysis

147
148 A three-level practical identification process (Table 1) was employed in this study, including the
149 (Level 1) assessing the normality (Gaussian) of posterior probability distribution of parameters by
150 employing uniform *a priori* probability distributions; and (Level 2) re-estimation of posterior
151 parameter distributions using normally distributed (conjugate) priors. Level 3 is carried out to assess
152 the goodness of experimental design and data inferred to achieve practical identifiability.

153 Level 1 was carried out using the Latin-Hypercube-Sampled priors (250 samples were found
154 sufficient to reach convergence) with uniform *a priori* probability distributions (LHSS) global
155 method (Wágner et al., 2015). The minimization of the sum of square of relative errors (SSRE),
156 obtained between 1-D model predictions and the experimental results, is carried out using the
157 MATLAB® function *fminsearch*, employing the Nelder-Mead algorithm – also known as the
158 Simplex method (Nelder and Mead, 1965). For selected constitutive functions with posterior
159 parameter values obtained in narrow, normally-distributed intervals (i.e. Eq. 4 and Eq. 6), posterior
160 parameter probability distributions are re-estimated using LHSS employing normally-distributed
161 priors. Therefore, in Level 2 (Table 1), parameter estimation is carried out by sampling from Gaussian
162 conjugate parameter probability ranges. In the LHSS, the Janus coefficient (J) is used to assess the
163 impact of parameter variability by considering a collinearity threshold for identifiability to be 0.7
164 (Ramin et al., 2017). This is done by comparing the relative predictive accuracy – calculated as the
165 sum of the root mean square of relative error for SBH , X_{bottom} and X_{side} – obtained using the upper and
166 lower parameter boundaries, calculated as the posterior mean parameter values +/- the 95%
167 confidence intervals obtained. For $J \sim 1$, it is concluded that parameters are uniquely identifiable.

168 Additionally, we report the Akaike's and Bayesian information criteria, AIC & BIC (Bozdogan,
169 1987) calculated for the new constitutive functions for hindered and compression settling velocity.

170 Dynamic global sensitivity (GSA) and uncertainty analysis were carried out using linear regression

171 of Monte Carlo simulation results obtained in Level 3 (Saltelli et al., 2008; Sin et al., 2011). In the
172 dynamic GSA, values of the standardised regression coefficient ($SRC_{j,p}$) is computed for each θ_j and
173 for each output (Y_k) at each time-step using the multivariate linear regression between the p^{th} LHS
174 sampled parameter value $\theta_{j,p}(t_i)$ and the k^{th} simulation output $Y_k(t_i)$ obtained using Monte Carlo
175 simulations, according to

$$176 \frac{Y_k(t_i) - \mu_{Yk}}{\sigma_{Yk}} = SRC_{j,p} \frac{\theta_{j,p}(t_i) - \mu_{\theta_j}}{\sigma_{\theta_j}} + \epsilon_k, \quad (3)$$

177 including the mean (μ_{Yk} and μ_{θ_j}) and standard deviation (σ_{Yk} and σ_{θ_j}) values of the simulation outputs
178 and parameters, respectively. Furthermore, ϵ_k is the error vector of the regression model (intercept).
179 The coefficient of determination, R^2 indicates the proportion of the total uncertainty of the model
180 output explained by the linear model. The SRC values are reliable to be used as sensitivity measures
181 when $R^2 > 0.6$. Also, only parameters with $SRC > 0.1$ are considered to be influential in predicting a
182 given output. GSA and uncertainty analysis can be used to inform and improve model calibration
183 exercises. Dynamic $SRC_{j,p}$ results were used to assess the sensitivity of predicting different
184 experimental data sets to model parameters and to locate specific experimental periods more
185 conducive to practical identifiability. For more on the calculation of SRC and R^2 , the reader is referred
186 to Saltelli et al. (2008). In addition to the settling function Eq. 4 and Eq. 6, the 3-parameter (3P)
187 logistic function for hindered settling by Diehl (Diehl, 2015; Torfs et al., 2017) in combination with
188 the compression settling function by De Clercq (2006) – i.e. the so-called Diehl-DeClercq model –
189 was tested through Levels 1-3. Furthermore, the hindered-transient-compression (HTC) settling
190 function by Ramin et al. (2014b) was assessed. The example of the Fredericia sludge (Initial
191 concentration: 3.44 g l⁻¹; Supporting Information) was chosen as experimental data for benchmarking
192 the different settling functions.

193 For the implementation of the 1-D simulation models and executing simulations the software Matlab
194 (The MathWorks, Inc., <http://www.mathworks.com/>) was used. Calculations of SRC were carried out

195 by transferring to and processing simulation results using Python (Python Software Foundation,
196 <https://www.python.org/>).

197

198 **2.5. CFD simulations**

199 The software ANSYS-CFX[®] (Academic Res. Release 17.2) was used to implement the solver,
200 according to Ramin *et al.* (2014b). Briefly, the solver employs an average Eulerian 2-phase flow
201 model. Turbulence is modelled using the k- ϵ model. Molecular viscosity of sludge is predicted using
202 the Herschel-Bulkley model (more information on model calibration is shown in the Supporting
203 Information, SI). For the full-scale SST simulations, the solver implementation development included
204 two scenarios, i.e. (1) the novel hindered-compression settling velocity function, (2) simple Vesilind
205 hindered settling function (the model calibration is described in the SI and Fig. S1). The initialization
206 of the 2-day transient state was carried out by converging a previous steady-state case with a constant
207 influent flow. For predicting activated sludge settling in the column sensor, the wall-with-no slip and
208 smooth roughness boundary conditions were used (more information on CFD simulations shown in
209 the SI).

210

211 **2.6. Model validation by triangulation (MVT)**

212 The MVT addresses the question of reliable prediction of hindered and compression solid settling
213 using the constitutive functions developed. MVT comprises two independent approaches, i.e. (A)
214 practical model identification using two independent sets of laboratory-scale measurements (samples
215 from Ellinge and Avedøre WRRFs) carried out with the new settling sensor setup (Fig. 1) and using
216 the new constitutive functions for hindered-compression settling; and (B) transient-to-steady-state
217 simulations using independent sets of dynamic full-scale measurement data (*SBH* and *TSS_{RAS}*) using
218 a CFD simulation model developed. Key sources of bias for approaches *A* and *B* are the highly

219 degenerated simulation model structure in 1-D and the lack of estimation of parameter values other
220 than settling velocity parameters through the calibration of the CFD simulation model, respectively.
221 No specific direction of bias of these sources can be explicitly identified. Results from these two
222 approaches are then compared through the 3-D CFD simulation of column tests for well-settling and
223 filamentous sludges in terms of SBH , X_{bottom} and X_{side} .

224

225 **2.7. Assessment of two sources of uncertainty**

226 One of the sources of uncertainty assessed using CFD simulations, involved the design boundary
227 conditions of the settling column setup, characterised with a design factor (F). The impact of the
228 column sensor design on the model parameter estimates was tested via CFD simulations, whereby
229 the CFD solver was calibrated with model parameters obtained for the Fredericia sludge at 3.44 g l^{-1}
230 as initial concentration and for the Avedøre sludge at 3.86 g l^{-1} as initial concentration. The base case
231 scenario (F=1) was that of the real setup (Fig. 1a), and factors (e.g., F=0.7 means 70%) were applied
232 to resize the height and diameter of the column – maintaining the original proportions – using five
233 scenarios. Additionally, the approach of using 1-D simulation models for estimating parameters using
234 batch experimental data – that are then used to calibrate CFD simulation models – was identified and
235 assessed as an additional uncertainty source. The uncertainties in parameter values introduced as a
236 result of using different settling column designs were quantified using CFD simulations by
237 considering the settling characteristics of the Fredericia and Avedøre sludge samples at $X_{ini}=3.44$ and
238 3.9 g/l as initial concentrations, respectively. 3-D profiles obtained were converted into 1-D
239 concentration profiles that were then used to re-estimate the posterior model parameter values. The
240 posterior parameter estimates were then benchmarked against that obtained using the 1-D simulation
241 model and the column sensor measurement data.

242

243

244 **3. RESULTS AND DISCUSSIONS**

245

246 **3.1. Practical model identification**247 *3.1.1. Hindered-compression settling velocity functions*

248 An iterative approach was taken to test the practical identifiability of parameters in a plethora of
 249 potential rate equations (more information shown in SI). Using practical identifiability as a selection
 250 criterion, a four-parameter model was identified to describe hindered-compression settling as

$$251 \quad v_S = \begin{cases} v_H(X) = v_0 e^{-r_H \cdot X} & 0 \leq X \leq X_C \\ v_H(X) \cdot \left(1 - \frac{\rho_S}{(\rho_S - \rho_f)gX} \frac{\partial \tau}{\partial X} \frac{\partial X}{\partial z}\right) & X > X_C, \end{cases} \quad (4)$$

252 where

$$253 \quad D_{Comp} = \begin{cases} 0 & 0 \leq X \leq X_C \\ v_H(X) \frac{\rho_S}{(\rho_S - \rho_f)gX} \frac{\partial \tau}{\partial X} & X > X_C, \end{cases} \quad (5)$$

254 with

$$255 \quad \frac{\partial \tau}{\partial X} = v_C e^{\left[r_C \frac{X}{X_{Infi}}\right]}, \quad (6)$$

256 where the effective solids stress (τ) derivative is formulated using an exponential term with v_C (m^2
 257 s^{-2}) and r_C (-) parameters. The maximum solids concentration (X_{Infi} , $kg\ m^{-3}$) is used to normalise
 258 local biomass concentration values, X . For hindered settling velocity (v_H , $m\ s^{-1}$), the model includes
 259 a pseudo 2-parameter exponential constitutive function with v_0 ($m\ s^{-1}$) and r_H ($m^3\ kg^{-1}$), denoting the
 260 hindered settling velocity parameters. In Eq. 4, ρ_S and ρ_f are the sludge and water density,
 261 respectively; g denotes the gravitational acceleration constant; z is the depth in the settling column.
 262 The setting of the compressive threshold concentration (X_C) is according to Guyonvarch *et al.*
 263 (2015). That is, for simulating batch column tests, 1-D advection-dispersion and 2-D CFD
 264 modelling of SSTs, $X_C = X_{ini}$; $X_C = X_{feed+1}$ and $X_C = X_{feed}$, respectively; where X_{feed+1} is the solid

265 concentration located just below the dynamic feed-layer in 1-D simulation model, and X_{feed} is the
266 solid concentration in the SST influent.

267 Compared to previous models, including the exponential term (Vesilind, 1968; Takács *et al.*, 1991),
268 instead of letting parameters independently vary, the ratio of v_0/r_H is estimated with v_0 set as
269 constant at $v_0=0.0025$ (m s⁻¹). The v_0/r_H ratio (kg m⁻² s⁻¹) – same unit as solid flux – can be linked
270 to the degree of sludge bulking (Ekama *et al.*, 1997; Wágner *et al.*, 2015), making it a suitable
271 controlled variable at WRRFs. Previous experimental studies (Daigger, 1995; Weiss *et al.*, 2007)
272 also indicate a constant v_0 . In fact, Daigger’s database show a value (0.0022, m s⁻¹) that is in very
273 close agreement with our observations obtained through practical model identification, thereby
274 providing experimental evidence to consider v_0 as constant.

275 The new settling velocity functions require the estimation of three parameters (v_0/r_H , v_C , r_C) using
276 global optimisation and the parameter, X_{Infi} is directly deducible from experimental data using
277 regression analysis. Uncertainty plots (Fig. 2a) show 95% confidence intervals effectively covering
278 most of the measured data. Values of relative dynamic sensitivity (Fig. 2c) and coefficient of
279 determination (Fig. 2d) suggest potential benefits of the set of experimental data for parameter
280 identifiability: high sensitivity of predicting SBH and TSS side sensor concentration data to v_0/r_H
281 parameter. Moreover, the prediction of TSS bottom and side concentration data show relatively high
282 sensitivity to the compression settling parameters, r_C and v_C . Based on the trajectories of SRC, an
283 extended experimental time beyond one hour does not seem to add any significant benefit to
284 parameter estimation. That is SRC does not increase significantly beyond the experimental time set
285 in here. Posterior parameter distributions (Fig. 3) show comparably narrow confidence intervals (CI),
286 i.e. CI-to-mean parameter ratios are <50% (Supporting information). Although, the covariance
287 matrices show values up to ~0.83 for compression parameters, parameter variability does not
288 significantly influence simulation outputs, i.e. $J \sim 1$ obtained, thereby indicating practically

289 identifiability of model parameters using the experimental data. Additionally, increasing the initial
290 solids concentration increases the SRC values only for v_0/r_H parameter (Fig. S2-S5) – whilst no
291 significant benefits in terms of identifiability of compression parameters can be drawn. Part of the
292 reason for this observation may stem from the position of the side sensor – perpendicular to the
293 direction of the settling solids in the column – that can possibly lead to deteriorating quality of sensor
294 data, in particular, at higher solid concentrations (e.g., Fig. S5). This shortcoming of the sensor;
295 however, does not seem to influence the overall quality of unique parameter sets obtained, i.e. mean
296 values of v_0/r_H , v_C , r_C are obtained with relatively narrow ranges of mean parameter values and with
297 overlapping 95% confidence intervals (Fig. 3) – except for the lowest r_C value obtained with
298 filamentous bulking sludge. Therefore, to improve parameter identifiability, future research should,
299 focus on optimising the position of the side sensor. In addition to Eq. 4 and Eq. 6 functions, the Diehl-
300 DeClercq and the HTC models were assessed, and were found practically non-identifiable based on
301 Step 6a & 6b@Level 1 (Table 1, Supporting Information). That is, for the Diehl-DeClercq functions,
302 histograms obtained (Step 6a@Level 1) indicate non-identifiability for v_0 , q , α (Supporting
303 Information). Comparably high SRC values were obtained only for v_0 , \bar{X} and C_g – an outcome that
304 reasonable agrees with the parameter identifiability results (Level 1), except for v_0 for which high
305 SRC values do not result in identifiability. Relatively high parameter correlation is found for v_0 - \bar{X}
306 and \bar{X} - C_g parameter pairs (Janus test not done). For the HTC functions, histograms obtained (Step
307 6a@Level 1) indicate non-identifiability for r_H , C_1 and C_2 (Supporting Information). Comparably
308 high SRC values were not obtained for any of the parameters, thus indicating major limitations of
309 practical identifiability. Relatively high parameter correlation is found for v_0 - \bar{X} and \bar{X} - X_C parameter
310 pairs (Janus test not done). Despite failing the identifiability criteria, to provide a comparison between
311 different settling functions, AIC and BIC values (Table 2) were also computed for the Diehl-DeClercq
312 and the HTC functions. Based on the information criteria, the strength of the evidence (e.g., ΔBIC)

313 against the functions with the higher BIC value is reasonably strong, i.e. $\Delta\text{BIC}>6-10$ (Kass and
314 Raftery, 1995).

315 *3.1.2. Modelling compression settling*

316 The present paper proposes a compression settling function to predict any effects of solid stress
317 propagating through the sludge blanket (i) by setting $X_C = X_{ini}$ and (ii) by formulating the solids stress,
318 in contrast to previous approaches, independently from the X_C value and from the relative
319 concentration ($X - X_C$). DeClercq (2006) suggest modelling sedimentation transport by considering
320 hindered and compression settling, and by employing a time-dependent onset of compression through
321 X_C . Partly because DeClercq (2006)'s model overestimate the transient settling velocity (in the falling
322 hindered settling rate region characterised with straight isoconcentration lines) and due to challenges
323 in implementing the proposed X_C models in SST simulation models, Ramin et al. (2014) propose a
324 model that additionally (A) includes a first-order transient settling function, formulated analogously
325 to hindered settling, and (B) employes two threshold concentrations for the onset of transient (X_T)
326 and for compression settling, X_C .

327 In Fig. 4a, at relatively low X values, upward propagating straight isoconcentration lines are shown,
328 above which, tangential isoconcentration lines propagate from the sediment-suspension interface.
329 These simulation results demonstrate the consistency of the simulation model with the actual physical
330 phenomena, i.e. with theory (Diplas and Papanicolaou, 1997, Kinnear, 2002) and experimental
331 observations (DeClercq et al., 2008 – the Destelbergen sludge). There is also a close agreement
332 between the evolution of the effective solid stress derivative obtained with the Destelbergen sludge
333 (DeClercq et al., 2008) – i.e. $0.25 - 1.95 \text{ m}^2 \text{ s}^{-2}$ for $X_{ini}=2.4, 3.2 \text{ g l}^{-1}$ – and that shown in this study
334 (Fig. 10). With well-settling solids (Fig. 4b,c), isoconcentration lines indicate a comparably fast
335 compressive solid consolidation behaviour compared to that obtained with filamentous sludge (Fig.
336 4a).

337 Additionally, the physical justification for setting X_C at the feed solid concentration for SST
338 modelling (Guyonvarch *et al.*, 2015) is that the density current of the feed slurry tends towards zero
339 buoyancy, thus propagating through volumes, under which, the descending particles shall increase
340 the local concentration only if they exhibit compressive solids settling. For more information, the
341 reader is referred to the Supporting Information.

342

343 **3.2. Model validation**

344 *3.2.1. Independent batch settling experiments*

345 Independent experimental settling data – obtained using solids with well-settling (Ellinge WRRF)
346 and filamentous bulking (Avedøre WRRF) characteristics – were used to test the practical
347 identifiability and the validity of the simulation model, including the new hindered-compression
348 settling functions. As for the Ellinge data (Fig. S6-S8), results obtained show close agreement with
349 the outcomes in the Fredericia case (Fig. 2) in terms of predictive accuracy for SBH, TSS_{side} , TSS_{bottom}
350 sensor data outputs and of parameter covariance. Parameter covariance obtained with the comparably
351 narrow probability density ranges (Fig. S6-S8; covariance indices up to ~ 0.8 for v_C and r_C) does not
352 significantly influence simulation outputs, i.e. $J \sim 1$. However, the uniqueness of v_C parameter
353 estimates is not significantly impacted (Fig. 3). Values of $SRC > 0.8$ values for both v_0/r_H and r_C
354 indicate high sensitivity of model predictions to both parameters that can potentially benefit practical
355 model identifiability (Fig. S6-S8). In contrast to the Ellinge dataset, prediction of settling of solids
356 collected in Avedøre WRRF extends the validity of the simulation model to filamentous bulking
357 conditions (Fig. 5; Fig. S9-S12). Again, the outcomes of the identifiability test closely agree with the
358 Fredericia and Ellinge cases. Disparities to the Fredericia and Ellinge cases, include the improved
359 prediction of the SBH and TSS_{bottom} with bulking sludge, whereas the accuracy of TSS_{side} prediction
360 is compromised at both low and high initial solids concentrations (Fig. 5a,d). Finally, according to

361 the increasing values of SRC for compression settling parameters for TSS_{side} , under filamentous
362 bulking conditions, practical identifiability tests may benefit from extending the length of experiment
363 over an hour at low initial solids concentration, i.e. $<3.3 \text{ g L}^{-1}$ (Fig. S11). In terms of uniqueness of
364 posterior parameter estimates (Fig. 3), the identifiability of compression settling parameters seems to
365 be compromised at higher initial solids concentrations – possibly, as a result of wall-effects – though,
366 this effect is not significant, based on the confidence intervals obtained (Fig. 3).
367 Taken together, the independent results obtained with Ellinge and Avedøre solids suggest the validity
368 of the identifiability approach and the simulation model structure for well-settling and filamentous
369 bulking conditions – an important aspect for future development of model-based control design
370 structures for WRRFs.

371

372 3.2.2. *CFD simulations of a full-scale SST*

373 As part of the MVT approach, forward CFD simulations were carried out to validate the hindered-
374 compression settling model. Simulations of the SBH, X_{RAS} reasonably agree with the measured full-
375 scale SST data collected during more than 40 hours (Fig. 6a and 6b). Furthermore, radial velocity
376 measurements (see also SI) indicate close agreement with that predicted using the new settling
377 velocity model. We note that CFD model prediction is somewhat compromised when predicting the
378 transient SST behaviour – in terms of SBH and TSS_{RAS} between maximum and minimum loading
379 conditions (between 20 and 30 hours). It is assumed that the lack of undertaking a more in-depth
380 optimisation of the e.g., SBH sensor calibration, CFD simulation model can explain the prediction
381 inefficiency observed.

382 Taken together, both approaches involved in the MVT support the hypothesis that the novel
383 constitutive functions for hindered and compression settling velocity combined with the X_C setting
384 method can reliably predict the real physical phenomena.

385

386 **3.3. Posterior parameter estimation and parameter intervals**

387 Fig. 3 summarises all parameter values with confidence intervals obtained with the sludge samples
388 taken from the three WRRFs. As for Fig. 3a., fixing v_0 was found to allow the estimation of v_0/r_H
389 values in a narrow range (Fig. 3) for the different initial concentrations and independently from the
390 compression parameters. This was otherwise impossible to achieve with any of the hindered and
391 compression functions and their combinations thereof tested (see more on this in the SI). Fig. 3a also
392 supports the hypothesis of v_0/r_H effectively gauging sludge settling properties (Wágner *et al.*, 2015).
393 Furthermore, an arbitrary threshold value $v_0/r_H \sim 0.005$ is proposed to distinguish between well-
394 settling and filamentous bulking solids. Despite the considerable difference between the three
395 WRRFs in terms of operating conditions – notably, SRT and bulking control measures – all four
396 settling velocity parameters obtained show consistent and comparable trends. That is, hindered and
397 compression parameters are independent of initial solids concentration and the parameter values
398 obtained for well-settling solids versus filamentous bulking sludge vary significantly at low initial
399 solid concentrations. In the practical identification analysis (Level 1), posterior values of mean \pm
400 confidence intervals obtained (Table 1; Fig. 7) suggest the dependence of most compression
401 parameter sets on the initial solid concentration – a possible physical phenomenon, agreeing with that
402 reported in literature (Ramin *et al.*, 2014b; DeClercq *et al.*, 2008). Furthermore, once practical
403 identifiability of model parameters is established through Level 1 (Table 1), Gaussian conjugate
404 priors are used in Level 2 (Table 1; Fig. 3). Posterior parameter estimates obtained in Level 2 indicate
405 the independence of hindered (v_0/r_H) and compression settling parameters (v_C and r_C) of the initial
406 solids concentration, thereby suggesting the effect, being a result of error propagation rather than of
407 real physical phenomena.

408

409 **3.4. Practical guidance on sensor application for parameter** 410 **estimation**

411 The settling column sensor application presented here has clear benefits and shortcomings. On the
412 one hand it can be used to infer data for practical identification and parameter estimation of the
413 constitutive function (Eq. 4 and Eq. 6). The results (Fig. 2, Fig. S2-S12) indicate the benefits of using
414 the bottom- and the side-TSS sensors to decrease the uncertainty of r_C and v_C , compression
415 parameters. This however, is only true at comparably low initial solid concentrations. On the other
416 hand, low parameter variability is obtained for all settling velocity parameters in the initial solid
417 concentration range examined (Fig. 3). Consequently, from a practical point of view, employing the
418 column sensor setup in 1 or 2 batch experiments – at initial solid concentrations $\leq 3 \text{ g L}^{-2}$ – to infer
419 data for parameter estimation seems reasonable and a reliable way of settling parameter estimation.

420

421 **3.5. Assessing sources of uncertainty using CFD simulations**

422 For selected initial solids concentrations with well-settling and bulking sludge, the close agreement
423 of measured and CFD simulation results (Fig. 8 at design factor $F=1$) indicate negligible uncertainties
424 introduced by the 1-D parameter estimation approach, and thus suggest the reliability of the parameter
425 estimation approach. Interestingly, compared to the 1-D case (compare Fig. 8a-c to Fig. 2), closer
426 agreement between the simulated and measured SBH data is achieved using the CFD simulation
427 model. This result suggests that the overestimation of the compressive SBH tail by the 1-D simulation
428 model may be a bias caused by the degenerated 1-D simulation model structure rather than the settling
429 velocity model structure – an impact that will be investigated in the future. The latter was the same
430 in both the 1-D and the 3-D CFD model. Torfs *et al.*, (2017) assessed the effect of overestimation of
431 the compressive SBH tail in more depth, suggesting the 1-D simulation model structure - in terms of

432 hindered settling velocity formulation – as the potential cause of this bias. Furthermore, to assess the
 433 variability of parameter values as a result of settling column design, CFD simulations, carried out
 434 within a wide range of column design boundary conditions (Fig. 8), were used to re-estimate the
 435 posterior model parameter values (Fig. 9). Results obtained suggest that, assuming negligible wall
 436 effects in the column sensor, the estimated r_C values can be expected to vary significantly in the
 437 design boundary range studied for both well-settling and filamentous sludge settling. However, it is
 438 not expected to introduce significant bias in the predictive accuracy based on $J \sim 0.91$ obtained using
 439 the 1-D simulation model calibrated with the lowest and highest values obtained for r_C (Fig. 9).

440

441 **3.6. Quantifying the effective solid stress derivative using sensor**

442 **data**

443 This study also addressed the question whether the multi-probe sensor setup could be used to quantify
 444 the τ derivative – a variable approximated using the sensor data according to

$$445 \frac{\partial \tau}{\partial X} = \frac{(\rho_S - \rho_L) \cdot g \cdot h_{side\ sensor}}{X_{Bottom}(t) - X_{Side}(t)}, \quad (7)$$

446 where the density difference between water and sludge ($\rho_S - \rho_L$) was assumed constant. Eq. 7 was
 447 formulated based on force balance analysis – assuming only the gravitational, buoyancy and solids
 448 pressure acting on particles – and by assuming quasi steady-state conditions (Xu et al., 2017).
 449 Simulation results obtained (Fig. 10) reasonably agree with the sensor τ -derivative values for sludge
 450 with well-settling and filamentous bulking characteristics, thereby indicating the feasibility of the
 451 sensor approach to quantify the solid stress derivative. This outcome is extremely important as it
 452 indicates that the sensor setup presented herein can be used by practitioners to estimate compression
 453 settling parameters directly from the measurements using simple regression analysis – similar to that
 454 conventionally used for hindered settling parameter estimation. To test the approach v_C and r_C values
 455 were estimated using exponential curve fitting of the τ -derivative – X/X_{infi} (Fig. 10). Discrepancies

456 between v_C and r_C values obtained and those estimated through global optimisation (Fig. 3) are
457 between 4 – 260%. The fitness of settling model calibration by means of simple regression analysis
458 should be evaluated, in the future, thereby also assessing the error introduced by assuming quasi
459 steady-state in approximating compressive solids stress using the sensor data.

460 De Clercq et al. (2008) present a study of calculating solids stress in settling column tests carried out
461 with the aid of radiotracers. These tests allow the spatial-temporal quantification of complete
462 concentration profiles. In contrast, our setup offers a simpler cost-effective technical solution for
463 practitioners with temporal TSS measurements only at two spatial points. In our study, in contrast to
464 De Clercq (2006), estimating the solids stress gradient and employing X_C fixed at the initial
465 concentration allowed the implementation of constitutive functions with reduced complexity and
466 number of parameters.

467

468

469 **4. OUTLOOK AND PERSPECTIVES**

470 Practical model identifiability of the constitutive functions, Eq. 4 and Eq. 6, describing hindered and
471 compression settling velocity is important because, it can, for the first time, allow for the estimation
472 of reliable, unique, posterior parameter values using sensor data. Bürger et al. (2011) propose a
473 consistent modelling methodology (CMM) for SST that requires the use of consistent and reliable
474 numerical method (solver) to satisfy the entropy condition – an admissibility criterion, ensuring
475 physically relevant (stable) discontinuities appear in the numerical approximate solution. Practical
476 identifiability as a criterion, extending the CMM, can alleviate ill-conditioned and ill-posed
477 calibration problems, thereby reducing uncertainties, propagating to the simulation results. Compared
478 to former constitutive assumptions (e.g., Takács et al., 1991; Torfs et al., 2017), the first study
479 assessing practical identifiability, i.e., Ramin et al. (2014b), indicate partial compliance with

480 identifiability criteria, and suggest further research on methods to infer experimental data and on
481 mathematical functions. These recommendations formed the main aims of the present study.
482 Regarding the column sensor setup presented here, the side TSS sensor indicate considerable
483 shortcomings at comparably high initial solid concentration. This inefficiency, once overcome, can
484 potentially decrease the uncertainties in compression settling model calibration by means of global
485 optimisation and simple regression analysis. The constitutive functions for hindered-compression
486 settling combined with the probe settling column setup proposed here can, for the first time, allow
487 practitioners to develop reliable and updatable model-based decision support and process control
488 structures to mitigate the impacts of hydraulic shocks on WRRFs. An example for such a decision
489 support system includes aeration tank settling (Thornberg *et al.*, 1998; Gernaey *et al.*, 2004) – an
490 effective means reducing the by-passed untreated sewage under wet-weather flow conditions.

491

492

493 **5. CONCLUSIONS**

494 The concluding remarks drawn in the study include:

- 495 • A pseudo two-parameter and a three-parameter exponential term were identified to describe
496 hindered settling velocity and the effective compressive solids stress gradient, respectively.
497 The ratio of v_0/r_H was estimated with v_0 set as constant in the hindered settling function.
498 Solids concentration is normalised using the X_{Infi} parameter easily obtainable in regression
499 analysis. The three parameters required to estimate using the global optimisation method are
500 v_0/r_H , v_C and r_C – all practically identifiable using the data obtained using the innovative
501 multi-probe column sensor setup.
- 502 • It is demonstrated that uncertainties, propagating from Bayesian prior settings to posterior
503 parameter estimates can cause significant bias; and that the three-level parameter estimation

504 method is effective in reducing this uncertainty propagation, and thus resulting in the
505 uniqueness of posterior parameter estimate solutions by employing the sequential uniform-
506 Gaussian Bayesian priors method;

- 507 • The novel constitutive functions for hindered-compression settling developed are validated
508 using independent batch column sensor data obtained with well-settling and filamentous
509 bulking solids. Additionally, model validation was carried out using independent full-scale
510 measurement and CFD simulation results.
- 511 • It is demonstrated that negligible uncertainties are introduced into CFD simulations by the
512 1-D parameter estimation approach using the column sensor data. Additionally, the multi-
513 probe settling sensor setup developed can be used to quantify the τ -gradient, and future
514 research should assess the benefits of using τ -gradient sensor data for settling model
515 calibration.

516

517 **ACKNOWLEDGEMENTS**

518 The co-authors wish to express their gratitude to University of Bath and the Technical University of
519 Denmark (DTU) for supporting this research with internal funding. All co-authors express their
520 deep appreciation and gratitude to the Reviewers of the manuscript, in particular, to Reviewer 1;
521 their comments and suggestions greatly improved the content of the manuscript. J. Climent and S.
522 Chiva gratefully acknowledge the support provided by Entitat de Sanejament d'Aigües de la
523 Comunitat Valenciana (EPSAR). B. Valverde-Pérez and B.G. Plósz acknowledge the financial
524 support obtained from the WISE project (Water Innovation SmE) funded by European Regional
525 Development Fund (RFH-15-011) and the Capital Region Denmark (15006962). B.G. Plósz wishes
526 to thank Prof Peter Steen Mikkelsen at DTU for his friendship, support and for the exciting

527 discussions on climate change mitigation in the Storm- and Wastewater Informatics (SWI) project
528 that inspired some of our ideas in the present research.

529

530

531 REFERENCES

532 Bernardo, J.M., 2000. *Bayesian theory*. Chichester: Wiley.

533 Bozdogan, H., 1987. Model selection and Akaike information criterion AIC - the general-theory and its analytical
534 extensions. *Psychometrika* **52**,345-370.

535 Bürger, R., Diehl, S., Farás, S., Nopens, I., Torfs, E., 2013. A consistent modelling methodology for secondary settling
536 tanks: a reliable numerical method. *Water Science and Technology* **68**, 192–208.

537 Bürger, R., Diehl, S., Nopens, I., 2011. A consistent modelling methodology for secondary settling tanks in wastewater
538 treatment. *Water Res.* **45**, 2247-2260.

539 Daigger, G.T., 1995. Development of refined clarifier operating diagrams using an updated settling characteristics
540 database, *Water Environ. Res.* **67**, 95–100.

541 De Clercq, J., Nopens, I., Defrancq, J., Vanrolleghem, P.A., 2008. Extending and calibrating a mechanistic hindered
542 and compression settling model for activated sludge using in-depth batch experiments. *Water Research* **42**, 781–
543 791.

544 De Clercq, J., 2006. *Batch and continuous settling of activated sludge: in-depth monitoring and 1-D compression*
545 *modelling*. Ph.D. Thesis, University of Gent, Belgium.

546 De Clercq, J., Jacobs, J., Kinnear, D.J., Nopens, I., Dierckx, R.A., Defrancq, J., Vanrolleghem, P.A., 2005. Detailed
547 spatio-temporal solids concentration profiling during batch settling of activated sludge using a radiotracer. *Water*
548 *Research* **39**, 2125–2135.

549 Derlon, N., Thürlimann, C., Dürrenmatt, D., Villez, K. 2017. Batch settling curve registration via image data modelling.
550 *Water Research* **114**, 327-337.

551 Diplas, P., Papanicolaou, A.N., 1997. Batch analysis of slurries in zone settling regime. *J. Environ. Eng.* **123**, 659–667.

552 Diehl, S., 2015. Numerical identification of constitutive functions in scalar nonlinear convection - diffusion equations
553 with application to batch sedimentation. *Applied Numerical Mathematics* **95**, 154-172.

554 Ekama, G.A., Barnard, J.L., Günthert, F.W., Krebs, P., McConcordale, J.A., Parker, D.S., Wahlberg, E.J., 1997.

- 555 Secondary settling tank: theory, modelling, design and operation. Scientific and Technical Report No. 6. IAWQ,
556 London.
- 557 Gernaey, K.V., Nielsen, M.K., Thornberg, D., Hook, B., Munk-Nielsen, T., Ingildsen, P., Jorgensen, S.B., 2004.
558 Conservation principles suspended solids distribution modelling to support ATS introduction on a recirculating
559 WWTP. *Water Science and Technology* **50**, 179-188.
- 560 Guyonvarch, E., Ramin, E., Kulahci, M., Plósz, B.G. 2015. iCFD: Interpreted computational fluid dynamics -
561 Degeneration of CFD to one-dimensional advection-dispersion models using statistical experimental design -
562 The secondary clarifier. *Water Research* **83**, 396-411.
- 563 Jeppsson, U., Alex, J., Batstone, D.J., Benedetti, L., Comas, J., Copp, J.B., Corominas, L., Flores-Alsina, X., Gernaey,
564 K.V., Nopens, I., Pons, M.N., Rodriguez-Roda, I., Rosen, C., Steyer, J.P., Vanrolleghem, P.A., Volcke, E. I. P.,
565 Vrecko, D., 2013. Benchmark simulation models, quo vadis? *Water Science and Technology* **68**, 1, 1-15.
- 566 Kass, R.E., Raftery, A.E., 1995. Bayes Factors. *Journal of the American Statistical Association* **90**, 773–795.
- 567 Kinnear, D.J., 2002. *Biological solids sedimentation: a model incorporating fundamental settling parameters*. Ph.D.
568 Thesis, Department of Civil and Environmental Engineering, University of Utah.
- 569 Lawlor, D.A., Tilling, K., Smith, G.D., 2016. Triangulation in aetiological epidemiology. *Int. J. Epidemiology*, **45**, 6,
570 1866-1886.
- 571 Li, B., Stenstrom, M.K., 2016. Practical identifiability and uncertainty analysis of the one-dimensional hindered-
572 compression continuous settling model. *Water Research* **90**, 235-246.
- 573 Nelder, J.A., Mead, R., 1965. A simplex method for function minimization. *Computational Journal*. **7**, 308-313.
- 574 Plósz, B.G., Weiss, M., Printemps, C., Essemiani, K., Meinhold, J. (2007) One-dimensional modelling of the secondary
575 clarifier - factors affecting simulation in the clarification zone and the assessment of the thickening flow
576 dependence. *Water Research* **41**, 3359–3371.
- 577 Plósz, B.G., Liltved, H., Ratnaweera, H., 2009. Climate change impacts on activated sludge wastewater treatment: A case
578 study from Norway. *Water Science and Technology* **60**, 2, 533–541.
- 579 Raiffa, H. and Schlaifer, R., 1961. *Applied statistical decision theory*. Clinton Press, Inc., Boston, USA.
- 580 Ramin, E., Sin, G., Mikkelsen, P.S., and Plósz, B.G., 2014a. Significance of settling model structures and parameter
581 subsets in modelling WRRFs under wet-weather flow and filamentous bulking conditions. *Water Research* **63**,
582 209-221.
- 583 Ramin, E., Wágner, D.S., Yde, L., Binning, P.J., Rasmussen, M.R., Mikkelsen, P.S., Plósz, B.G. 2014b. A new settling

- 584 velocity model to describe secondary sedimentation. *Water Research* **66**, 447-458.
- 585 Ramin, P., Valverde-Pérez, B., Polesel, F., Locatelli, L., Plósz, B.G. 2017. A systematic model identification method for
586 chemical and biochemical transformation pathways – the case of heroin biomarkers in wastewater. *Scientific*
587 *Report*, 7:9390, 1-11.
- 588 Saltelli, A., Ratto, M., Andres, T., Campolongo, F., 2008. *Global sensitivity analysis: the primer*. John Wiley & Sons,
589 West Sussex, England.
- 590 Sin, G., Gernaey, K.V., Neumann, M.B., van Loosdrecht, M.C.M., Gujer, W., 2011. Global sensitivity analysis in
591 wastewater treatment plant model applications: prioritizing sources of uncertainty. *Water Research* **45**, 639–651.
- 592 Takács, I., Patry, G.G., Nolasco, D., 1991. A dynamic model of the clarification-thickening process. *Water Research* **25**,
593 1263-1271.
- 594 Thornberg, D.E., Nielsen, M., Eriksson, J., 1998. Upgrading of Boras wastewater treatment plant based on intelligent
595 process and operation control. *Water Science and Technology* **37**, 57-63.
- 596 Torfs, E., Balemans, S., Locatelli, F., Diehl, S., Bürger, R., Laurent, J., Francois, P., Nopens, I., 2017. On constitutive
597 functions for hindered settling velocity in 1-D settler models: selection of appropriate model structure. *Water*
598 *Research* **110**, 38-47.
- 599 Valverde Pérez, B., Penkarski-Rodon, E., Zhang, X., Wágner, D.S., Haecky, P., Blackburn, N.D., Plósz, B.G. 2017.
600 Development and validation of a novel monitoring system for batch flocculant solids settling process. *IWA-*
601 *ICA2017–12th IWA Specialist Conf. on Instrumentation, Automation and Control*. Quebec City, Canada.
- 602 Vanrolleghem, P.A., Van der Schueren, D., Krikilion, G., Grijspeerd, K., Willems, P., Verstraete, W., 1996. On-line
603 quantification of settling properties with in-sensor-experiments in an automated settlometer. *Water Science and*
604 *Technology* **33**, 37–51.
- 605 Vesilind, P.A., 1968. Design of prototype thickeners from batch settling tests. *Water Sewage Works* **115**, 302-307.
- 606 Wágner, D.S., Valverde-Pérez, B., Sæbø, M., Bregua de la Sotilla, M., Van Wagenen, J., Smets, B.F., Plósz, B.G., 2016.
607 Towards a consensus-based biokinetic model for green microalgae—the ASM-A. *Water Research* **103**, 485-499.
- 608 Wágner, D.S., Ramin, E., Szabo, P., Dechesne, A., Plósz, B.G. 2015. *Microthrix parvicella* abundance associates with
609 activated sludge settling velocity and rheology—Quantifying and modelling filamentous bulking. *Water Research*
610 **78**, 121-132.
- 611 Weiss, M., Plósz, B.G., Essemiani, K., Meinhold, J. 2007. Suction-Lift Sludge Removal and Non-Newtonian Flow
612 Behaviour in Circular Secondary Clarifiers: Numerical Modelling and Measurements. *Chemical Engineering*

613 *Journal* **132**, 241–255.

614 Xu, G., Yin, F., Xu, Y., Yu, H.Q., 2017. A force-based mechanistic model for describing activated sludge settling process.

615 *Water Research* **127**, 118-126.

616

617

FIGURES

618

Hindered and compression solid settling functions – sensor data

619

collection, practical model identification and validation

620

621

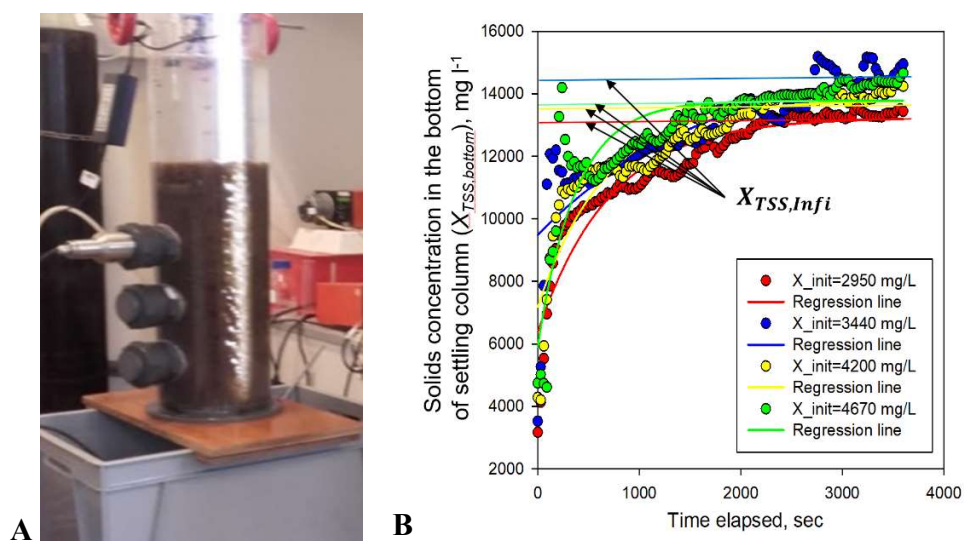
622 Benedek G. Plósz^{1,3,*}, Javier Climent², Christopher T. Griffin¹, Sergio Chiva², Rani Mukherjee¹, Elena

623

Penkarski-Rodon³, Matthew Clarke¹, and Borja Valverde-Pérez³

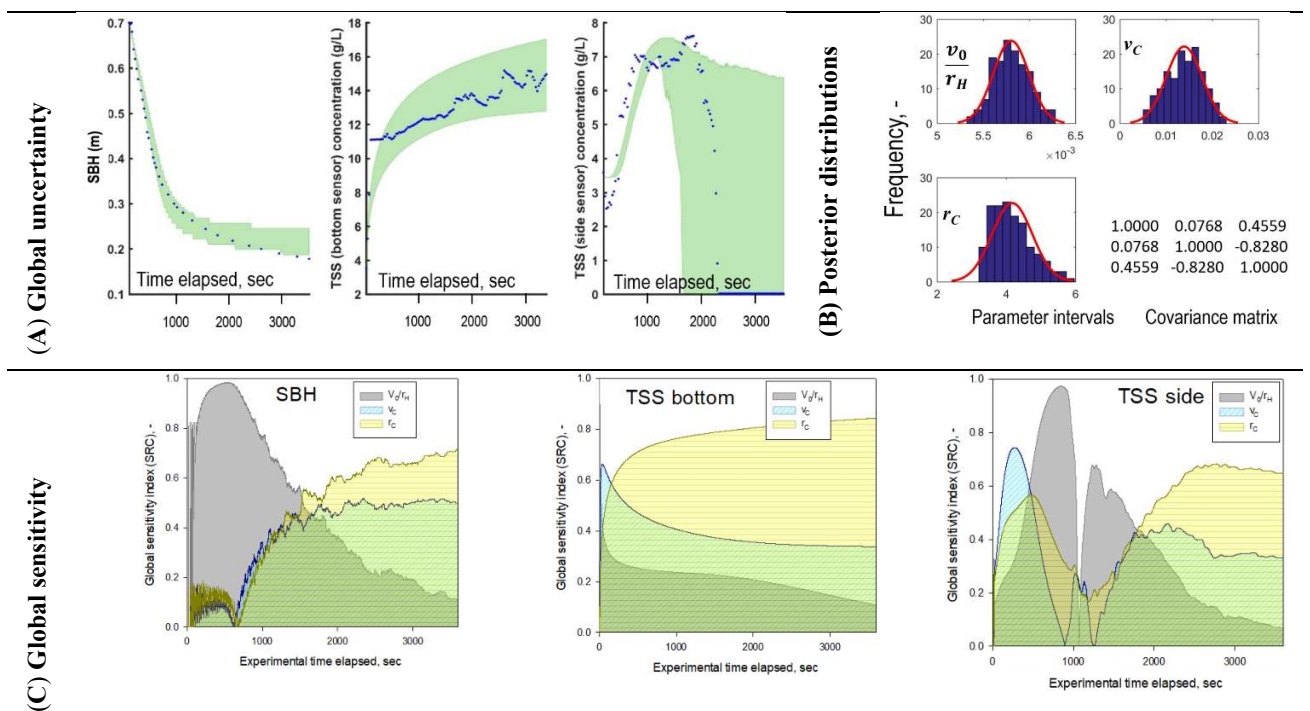
624

625



626

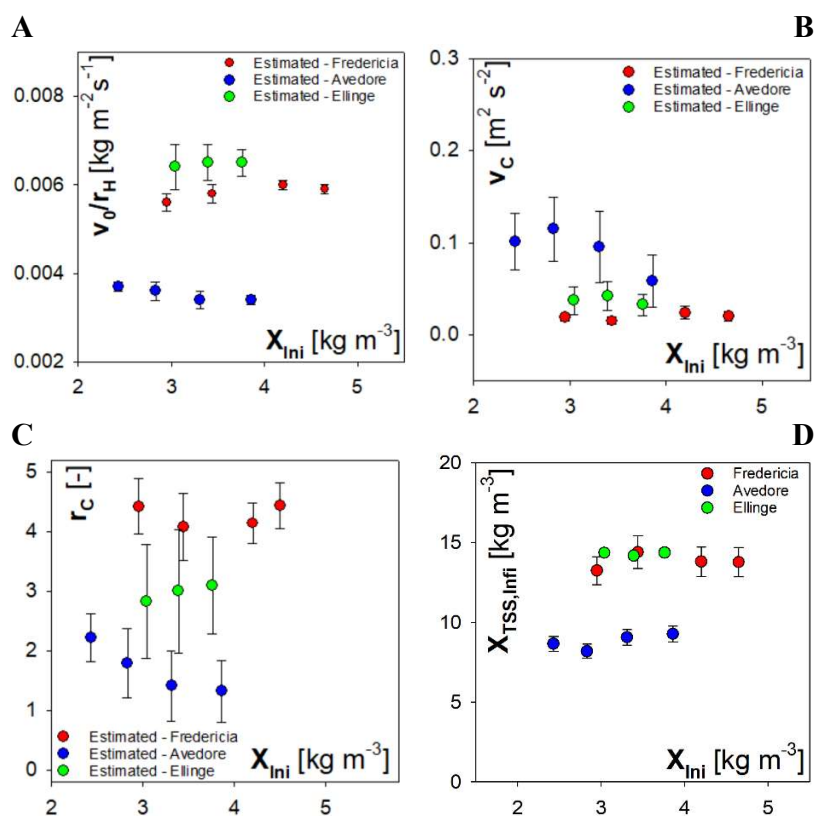
627 **Figure 1.** The multi-probe column sensor prototype – (A) equipped with two SOLITAX TSS
628 sensors installed in the bottom (not visible in the photo) and the sidewall of the settling column; and
629 (B) TSS values measured at the bottom of the settling column ($X_{TSS, Bottom}$) versus experimental time
630 and regression lines (Eq. 1) used to estimate $X_{TSS, Infi}$ values (Fig. 3d) - example shown here include
631 the settling experiments with Fredericia WRRF sludge.
632



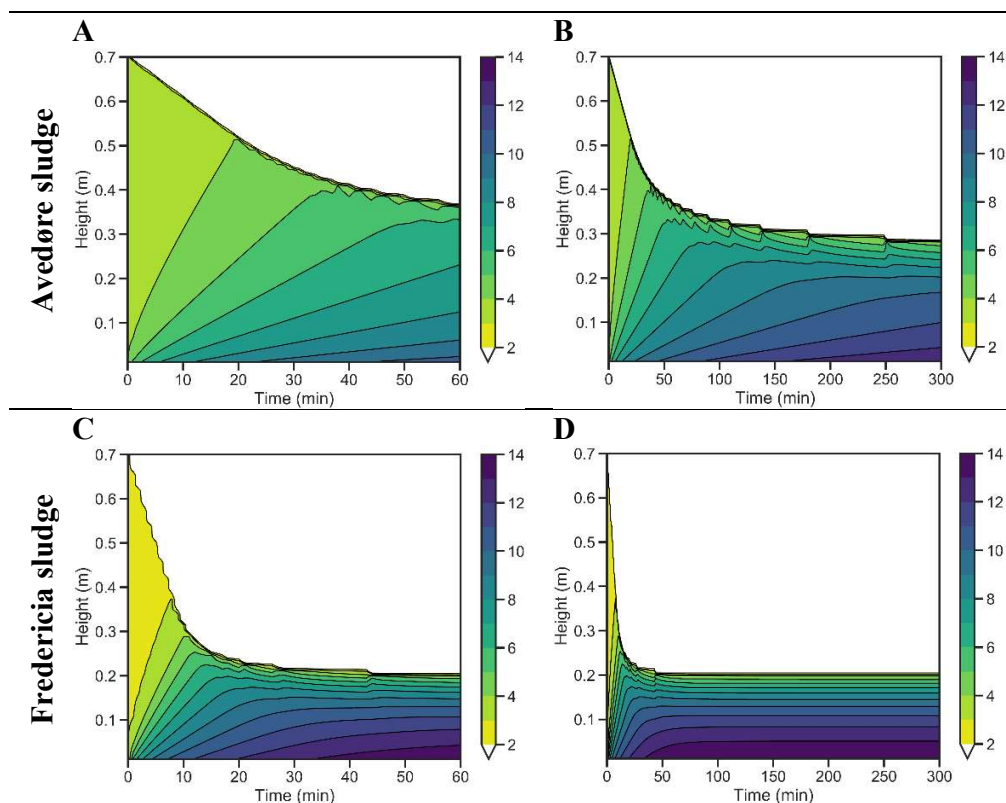
633 **Figure 2.** Practical model identification using data obtained with well-settling solids – measured
 634 and simulated data for activated sludge collected in Fredericia WRRF using the new hindered-
 635 compression settling velocity functions; Initial solid concentration: 3.44 g L^{-1} ; Proposed/*a priori*
 636 probability ranges: $v_0/r_H=[0.0052 \ 0.0063]$; $v_C=[0.005 \ 0.025]$; $r_C=[2.5 \ 5.5]$; (A) Global uncertainty
 637 plots with 95% confidence intervals, (B) posterior parameter density distributions with parameter
 638 intervals showing the *a priori* probability ranges, (C) values of dynamic sensitivity (SRC) and (D)
 639 coefficient of determination (R^2) computed for SBH, $X_{TSS,bottom}$ and $X_{TSS,side}$ concentration state-
 640 variables. Further results obtained with different dilutions of the activated sludge collected in
 641 Fredericia WRRF are shown in Fig. S2-S5.

642

643



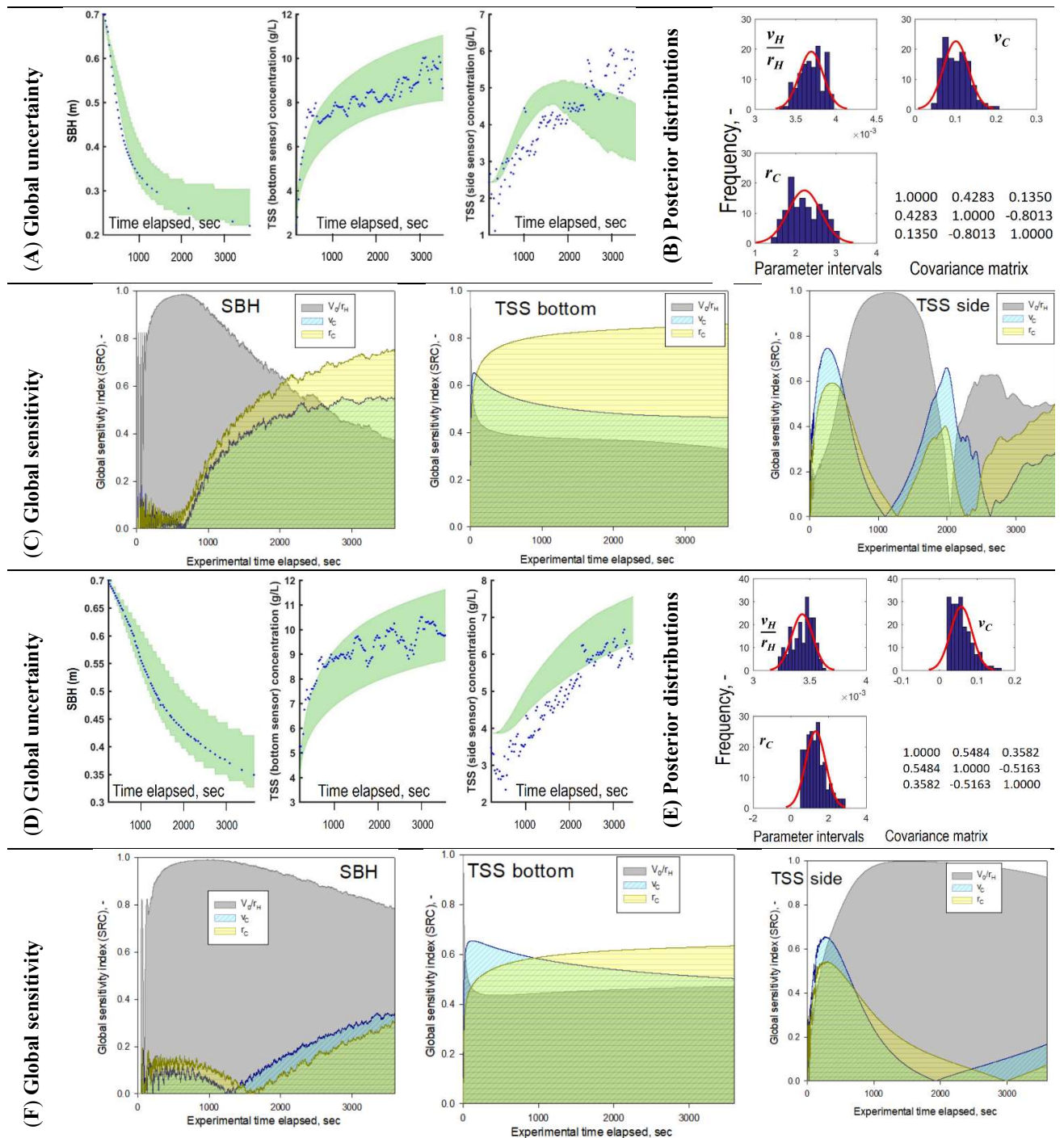
644 **Figure 3.** Posterior mean v_0/r_H (A), v_c (B), r_c (C) parameter values with 95% confidence interval
 645 denoted with error bars obtained with well-settling solids (Fredericia and Ellinge WRRFs) and solids
 646 with filamentous bulking (Avedøre WRRFs); and (D) $X_{TSS,Infi}$ values obtained using regression
 647 analysis for Fredericia, Ellinge and Avedøre WRRFs.
 648



649 **Figure 4.** Iso-concentration contour plots (X shown in colour fill legend) predicted in the batch
650 settling tests of (A-B) Avedøre WRRF – filamentous bulking sludge ($X_{Ini}=3.9 \text{ g l}^{-1}$) with experimental
651 times of 60 and 300 minutes; (C-D) Fredericia WRRF – well-settling sludge ($X_{Ini}=2.95 \text{ g l}^{-1}$) with
652 experimental times of 60 and 300 minutes.

653

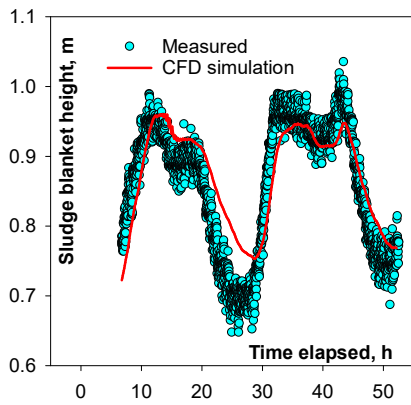
654



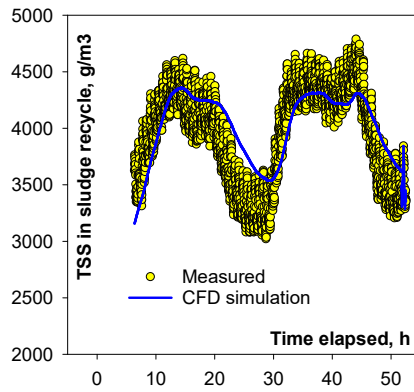
655 **Figure 5.** Model validation using data obtained with filamentous bulking solids – measured and
 656 simulated data for activated sludge collected in Avedøre WRRF using the new hindered-compression
 657 process model; Initial solid concentration: 2.4 g L⁻¹ (A, B, C); Proposed probability ranges:
 658 $v_0/r_H=[0.003\ 0.004]$; $v_C=[0.02\ 0.2]$; $r_C=[0.5\ 3]$; Initial solid concentration: 3.9 g L⁻¹ (D, E, F);
 659 Proposed probability ranges: $v_0/r_H=[0.003\ 0.004]$; $v_C=[0.01\ 0.2]$; $r_C=[0.1\ 3]$; Global uncertainty plots
 660 with 95% confidence intervals, posterior parameter density distributions; SRC and R² values shown
 661 in Fig. S9.
 662

663

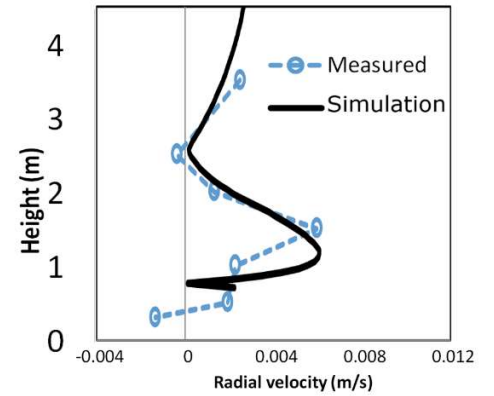
a



b



c

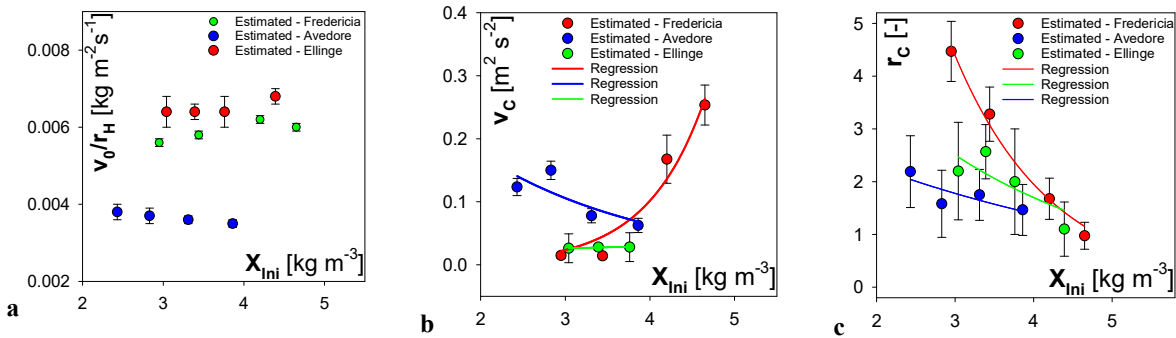


664 **Figure 6.** Measured and simulated (a) sludge blanket height, SBH, (b) TSS_{RAS} concentration, and (c)
665 vertical radial velocity profile (more information on the velocity metering shown in SI) in the full-
666 scale SST in OBVA WRRF, Vila-Real, Spain.

667

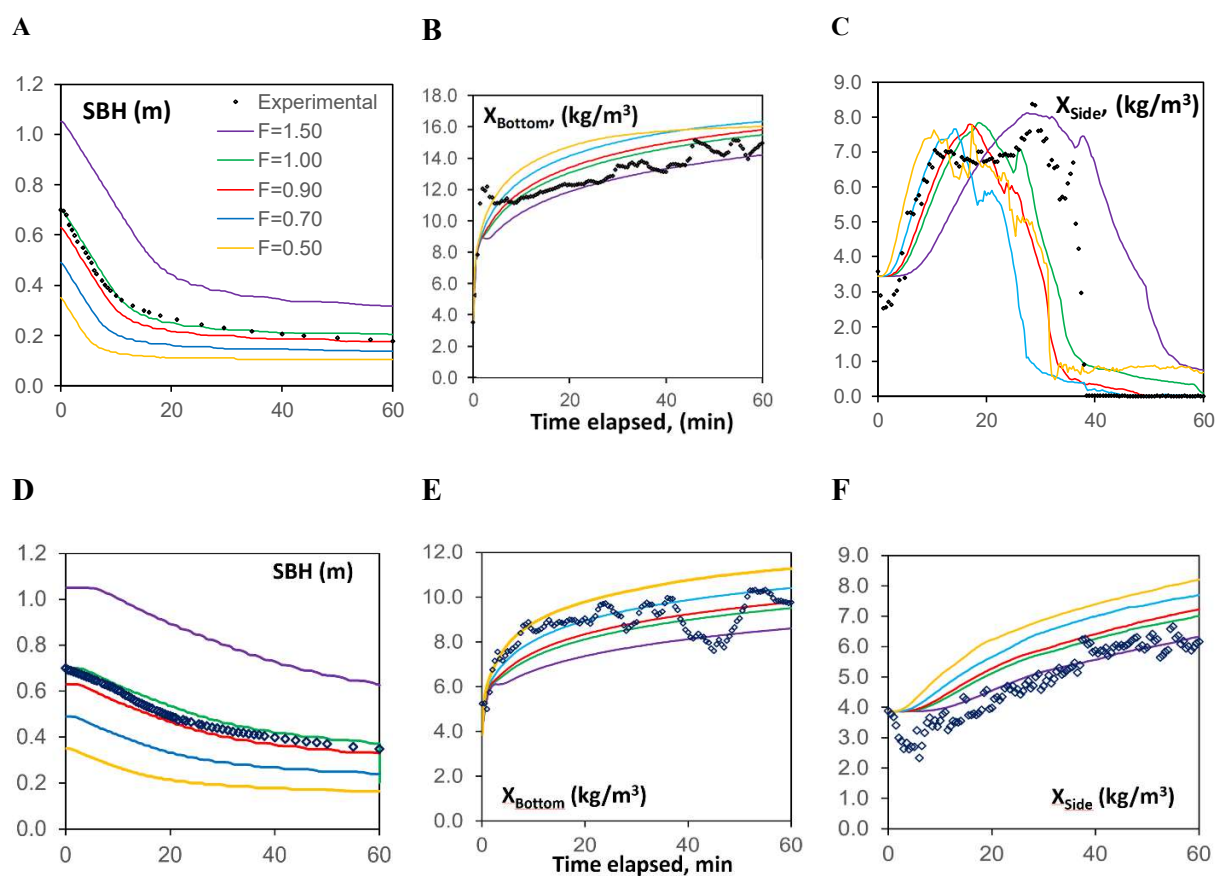
668

669



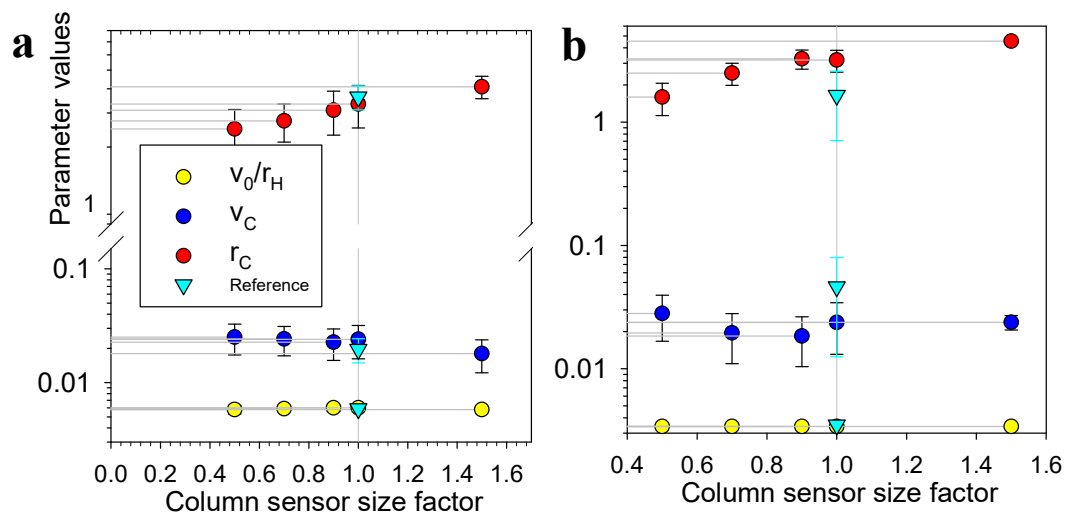
670 **Figure 7.** Posterior mean settling velocity parameter values with confidence interval (denoted with
671 error bars) obtained for Fredericia, Ellinge and Avedøre datasets in Level 1 (Table 1) using LHSS
672 with uniform *a priori*/proposed parameter probability density distribution; Parameter estimates
673 obtained in Level 2 (Table 1) using Gaussian *a priori* (conjugate) parameter probability distribution
674 is shown in Fig. 3.
675

676



677 **Figure 8.** Measured and CFD simulation results obtained for the settling column sensors, sized
 678 according to different design similarity factors (F) and compared to the real setup ($F=1$; Fig. 1a), in
 679 terms of SBH , X_{Bottom} and X_{Side} using solver calibrated according to parameter values obtained with
 680 (A, B, C) Fredericia WRRF sludge at $X_{ini}=3.44$ g/l (Fig. 2); and (D, E, F) Avedøre WRRF sludge at
 681 $X_{ini}=3.9$ g/l (Fig. S12).
 682

683

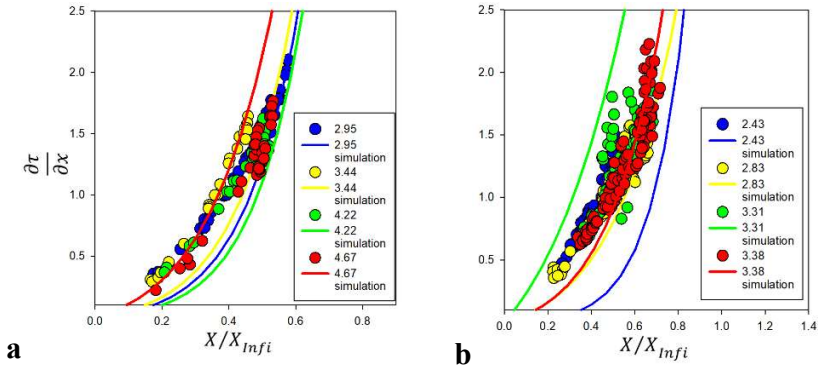


684

685 **Figure 9.** Settling model parameters estimated with the different column designs (see Fig. 5) using
686 CFD simulation output data obtained using calibration parameter sets for (A) well-settling sludge
687 from Fredericia WRRF and (B) sludge with filamentous bulking collected in Avedøre WRRF.
688 Reference parameter values shown were obtained using measured settling data (Fig. 2 and Fig. S12).
689

690

691



692 **Figure 10.** Calculated based on measured data (Eq. 4) and simulated compressive solid stress
693 derivative values plotted as a function of the X_{Side}/X_{Infi} values using (a) well-settling sludge from
694 Fredericia WRRF and (b) sludge with filamentous bulking collected in Avedøre WRRF.
695

696

TABLES

697

Hindered and compression solid settling functions – sensor data

698

collection, practical model identification and validation

699

700

701

Benedek G. Plósz^{1,3,*}, Javier Climent², Christopher T. Griffin¹, Sergio Chiva², Rani Mukherjee¹, Elena

702

Penkarski-Rodon³, Matthew Clarke¹, and Borja Valverde-Pérez³

703

704

705

706 **Table 1.** The two-level practical model identification and parameter estimation method

Algorithm – Level 1 : Assessing practical model identifiability using LHSS (discrimination of functions)	
1.	Definition of a priori parameter ranges, $p(\theta)$, where $\theta = \{\theta_1, \dots, \theta_j\}$ denotes a j-vector of model parameters using uniform probability distribution;
2.	Latin hypercube sampling, LHS from $p(\theta)$;
3.	<i>A priori</i> LHS parameter sets are used as initial values for the local optimisation MATLAB® function, <i>fminsearch</i> (Nelder-Mead algorithm, Simplex), employing the sum of square of relative errors (SSRE) as objective function to estimate posterior parameter values;
4.	Visualisation of posterior parameter probability density distribution, $p(\theta x)$ – with $x = \{x_1, \dots, x_n\}$ an n -vector of measurements - using histograms, excluding any parameter values with SSRE values higher than a selected threshold (10% of the minimum SSRE) – considered as local minima;
5.	Average parameter values, standard deviations and correlation matrix are computed;
6.	Practical model identifiability a. $(p(\theta x))$: histograms are interpreted in terms of (i) probability distribution: Gaussian distribution indicate parameter identifiability vs. uniform distribution indicates non-identifiability; (ii) narrow histograms indicate parameter identifiability; b. Correlation matrix, $cov(\hat{\theta})$: correlation of parameters are assessed by considering a collinearity threshold (CT) for identifiability to be 0.7; If for any pairs of parameters, $\theta_j\theta_j$, $CT > 0.7$, then these parameters are considered identifiable; If $J^2 = \frac{\frac{1}{n} \sum_{i=1}^n (y_{exp,p,i} - y_{sim,i}(t_i, \hat{\theta}_j + \sigma_j))^2}{\frac{1}{n} \sum_{i=1}^n (y_{exp,p,i} - y_{sim,i}(t_i, \hat{\theta}_j - \sigma_j))^2} = \frac{\sum_{i=1}^n (y_{exp,p,i} - y_{sim,i}(t_i, \hat{\theta}_j + \sigma_j))^2}{\sum_{i=1}^n (y_{exp,p,i} - y_{sim,i}(t_i, \hat{\theta}_j - \sigma_j))^2} \sim 1$ where, J^2 is the Janus coefficient, $y_{exp,p,i}$ is the i^{th} experimental data of the p^{th} variable, y , and $y_{sim,i}(t_i, \hat{\theta}_j + \sigma_j)$ is the simulation model output at the i^{th} time point, $\hat{\theta}_j$ is the mean optimal parameter value, σ_j is the corresponding standard deviation of θ_j , and n is the number of experimental data used; else the parameters are considered non-identifiable;
7.	Selection between candidate model structures. Calculate the Akaike Information Criterion (AIC) and Bayesian Information Criterion (BIC) values for the practically identifiable models (Supporting Information), and use AIC and BIC as selection criteria to compare alternative model structures.
Algorithm – Level 2 : Re-estimation of $p(\theta x)$ by considering Gaussian conjugate priors	
8.	Definition of <i>a priori</i> parameter ranges $p(\theta)$ using normal, Gaussian probability distribution;
9-12. These steps follow Steps 2-5 in Level 1 .	
Algorithm – Level 3 : Assessing experimental design conducive to practical identifiability	
13.	The standardised regression coefficient (SRC) is calculated; experimental data with high SRC indicate high probability of parameter identifiability.

707

708

709

710 **Table 2.** Akaike Information Criterion (AIC) and Bayesian Information Criterion (BIC) values
 711 calculated using the objective functions (OF) calculated for $\hat{\theta}_j \pm \sigma_j$. Lower AIC/BIC values indicate
 712 a superior set of functions. See also (Step 7@Level 1, Table 1).

	<i>New functions</i>	<i>Diehl-DeClercq functions</i>	<i>HTC functions</i>
AIC (OF ₁ , $\hat{\theta}_j + \sigma_j$)	10.7	16.4	18.4
AIC (OF ₁ , $\hat{\theta}_j - \sigma_j$)	10.6	16.8	18.5
AIC (OF ₂ , $\hat{\theta}_j + \sigma_j$)	12.5	17.8	19.9
AIC (OF ₂ , $\hat{\theta}_j - \sigma_j$)	12.4	18.8	20.1
BIC (OF ₁ , $\hat{\theta}_j + \sigma_j$)	21.6	38.1	43.7
BIC (OF ₁ , $\hat{\theta}_j - \sigma_j$)	21.5	38.5	43.8
BIC (OF ₂ , $\hat{\theta}_j + \sigma_j$)	23.4	39.5	45.2
BIC (OF ₂ , $\hat{\theta}_j - \sigma_j$)	23.2	40.5	45.4

713

714

715

716

717

SUPPORTING INFORMATION

718

719 **Hindered and compression solid settling functions – sensor data**
720 **collection, practical model identification and validation**

721

722

723 Benedek G. Plósz^{1,3,*}, Javier Climent², Christopher T. Griffin¹, Sergio Chiva², Rani Mukherjee¹, Elena

724 Penkarski-Rodon³, Matthew Clarke¹, and Borja Valverde-Pérez³

725

726

727 ¹Department of Chemical Engineering, University of Bath, Claverton Down, Bath BA2 7AY, UK (Email:

728 *b.g.plosz@bath.ac.uk; ctg24@bath.ac.uk; rm2100@bath.ac.uk; mjc62@bath.ac.uk*)

729 ²Universitat Jaume I, Department of Mechanical Engineering and Construction, Av. Vicent Sos Baynat, s/n 12071

730 Castellón (Spain), (Email: *jcliment@uji.es, schiva@uji.es*)

731 ³Dept. of Environmental Engineering, Technical University of Denmark, Bygningstorvet, Building 115, 2800 Kgs.

732 Lyngby, Denmark (Email: *bvape@env.dtu.dk*).

733 * *Corresponding author*

1. Discrimination of the hindered – compression settling functions based on practical identifiability criteria

With the exception of the the process model by Ramin *et al.* (2014), the hindered – compression settling functions tested in this study – other than that presented in the main body of text – can be described with the equation

$$v_s = \begin{cases} v_H & X \leq X_{TSS,C} \\ v_H \left(1 - \frac{\rho_s}{(\rho_s - \rho_f)gX} \frac{\partial \sigma}{\partial X} \frac{\partial X}{\partial z} \right) & X > X_{TSS,C} \end{cases}$$

, where v_H is the hindered settling velocity; X is the sludge concentration; X_{TC} is the threshold concentration for the onset of compression settling, $\partial \sigma / \partial X$ is the effective solid stress gradient, z is the vertical direction variable, and g is the gravity acceleration constant. D_i denotes generic model parameters with a subscript representing the number of parameters in each constitutive function. We note that the process model by Ramin *et al.* (2014) also includes a transient settling velocity function; and for the formulation of the process model, the reader is referred to the original publication.

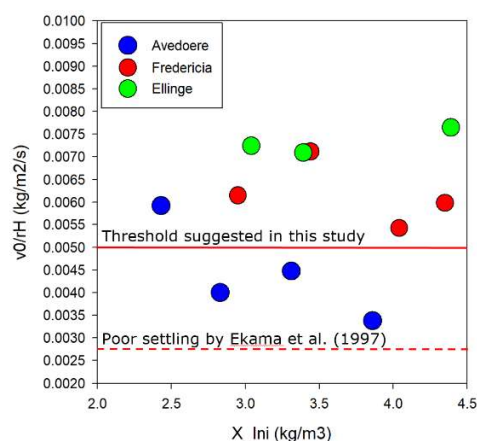
Briefly, the HTC model and the Diehl hindered settling function were tested, and results are presented in this chapter. In summary, the practical identifiability of the HTC model was assessed for all three sets of experiments, and related shortcomings – further detailed below - with the functions formed the main motivation for the present study. The practical identifiability of the compression function by De Clercq *et al.* (2008) was assessed by Ramin *et al.* (2014b), and shortcomings with the function – reported in the same paper - then formed the main motivation for developing the compression settling model presented as the HTC model by Ramin *et al.* (2014b). Noteworthy is that these shortcomings of the De Clercq's compression settling function were also shown by Li and Stenstrom (2016).

For discrimination of hindered – compression settling functions, the Diehl-DeClercq and the HTC models, both including hindered and compression settling functions were assessed in-depth using practical identifiability criteria.

765 **Table SI-i1.** Alternative constitutive functions for hindered-transient-compression settling assessed
 766 in this study.

Name of function, references	Constitutive functions
Hindered settling (v_H)	
Diehl (2008); Torfs et al. (2017) <i>Function in the Diehl-DeClercq model</i>	$v_0/(1+(X(i)/\bar{X})^q)$
Vesilind (1968) <i>Function in the HTC model</i>	$v_0*\exp(-r_H*X(i))$
Transient settling	
Ramin et al. (2014) <i>Function in the HTC model</i>	$v_t*\exp(-r_t*X(i))$
Compressive/effective solid stress	
Ramin et al. (2014) <i>Function in the HTC model</i>	$((X(i) - X_{C,limit})/C_1)^{C_2}$, where C_2 is found dependent (exponential function) on the initial solids concentration. The exponential function includes two regression parameters.
De Clercq et al. (2008) <i>Function in the Diehl-DeClercq model</i>	$\alpha*\ln((X(i) - C_g + \beta)/\beta)$ where D_2 is found dependent (power function) on the initial solids concentration. The power function includes two regression parameters.
Note: other functions were also tested; however, are not shown herein.	

767



768 **Figure. SI-i1** Examples drawn from the process model development study using the HTC model.
 769 (top chart) Posterior mean v_0/r_H values estimated with the HTC model (Ramin et al., 2014),
 770 showing high variability in parameter estimates as a function of initial sludge concentration. For the
 771 Avedoere WRRF, v_0/r_H values show both reasonably good and poor settling for the same sludge
 772 sample, which is one of the main drawbacks (other than those described in Ramin et al., 2014) that
 773 prompted the present study to be undertaken.
 774

775

2. Calculation of the AIC and BIC values (Table 1)

776
777
778 The two-level practical model identification and parameter estimation method is carried out
779 according to Table 1. In Step 7@Level 1, the selection between candidate model structures is
780 carried out Calculate the Akaike Information Criterion (AIC) and Bayesian Information Criterion
781 (BIC) values, including the objective functions (OF_j) of mean sum of the squared errors (SSE)
782 calculated for $\hat{\theta}_j \pm \sigma_j$

$$783 \quad OF_1 = \frac{SSE}{n} = \frac{\sum_{i=1}^n (y_{exp,p,i} - y_{sim,i}(t_i, \hat{\theta}_j \pm \sigma_j))^2}{n}$$

784 and and the root mean squared errors (RMSE):

$$785 \quad OF_2 = RMSE = \sqrt{\frac{SSE}{n}}$$

786 in the AIC as

$$787 \quad AIC = 2 \cdot (\log(OF) + j)$$

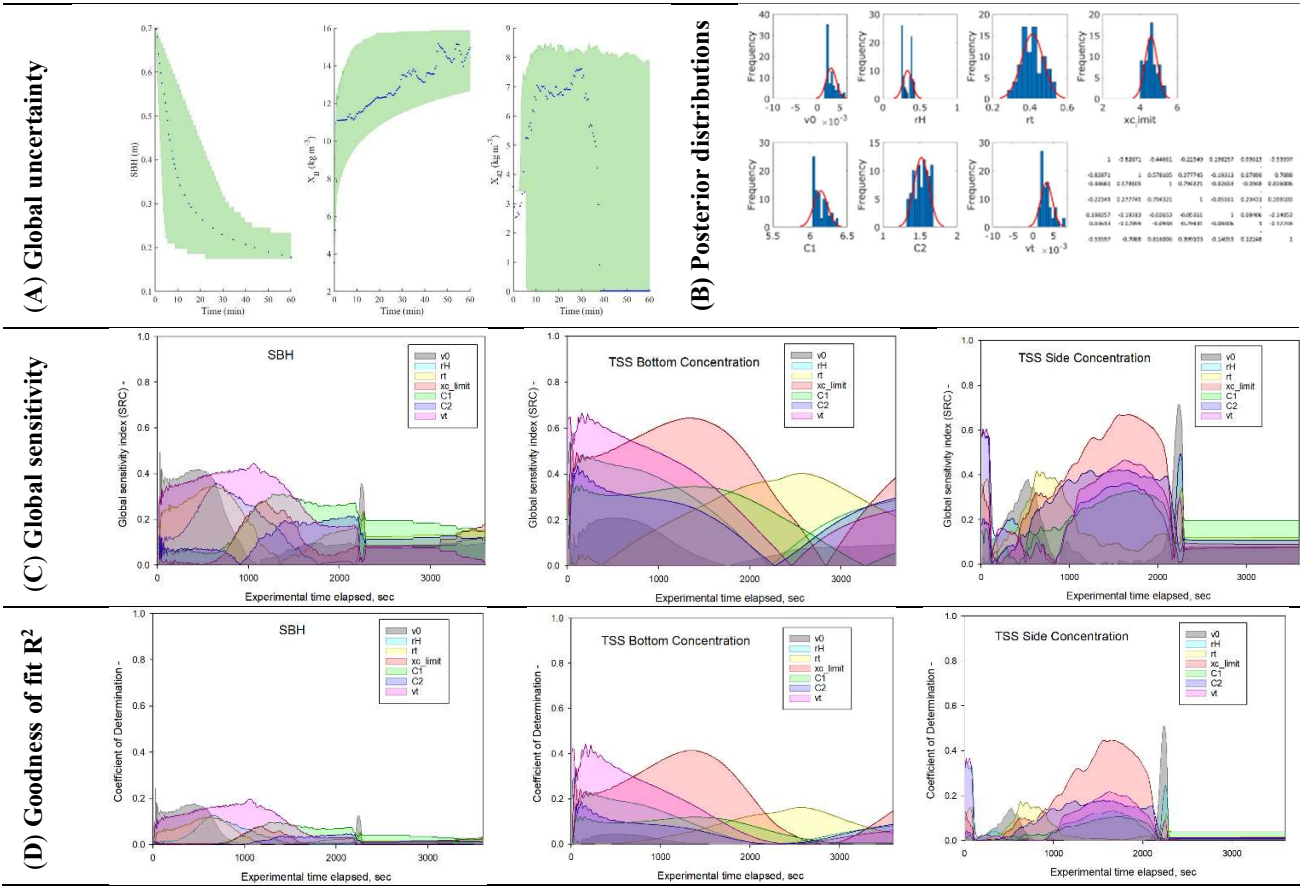
788 and use AIC as a selection criterion to compare model structures; alternatively, one can also use
789 e.g., the BIC, given as

$$790 \quad BIC = 2 \cdot \log(OF) + j \cdot \log(n)$$

791 where, $y_{exp,p,i}$ is the i^{th} experimental data of the p^{th} variable, y , and $y_{sim,i}(t_i, \hat{\theta}_j \pm \sigma_j)$ is the simulation
792 model output at the i^{th} time point, $\hat{\theta}_j$ is the mean optimal parameter value, σ_j is the corresponding
793 standard deviation of θ_j , and n is the number of experimental data used.

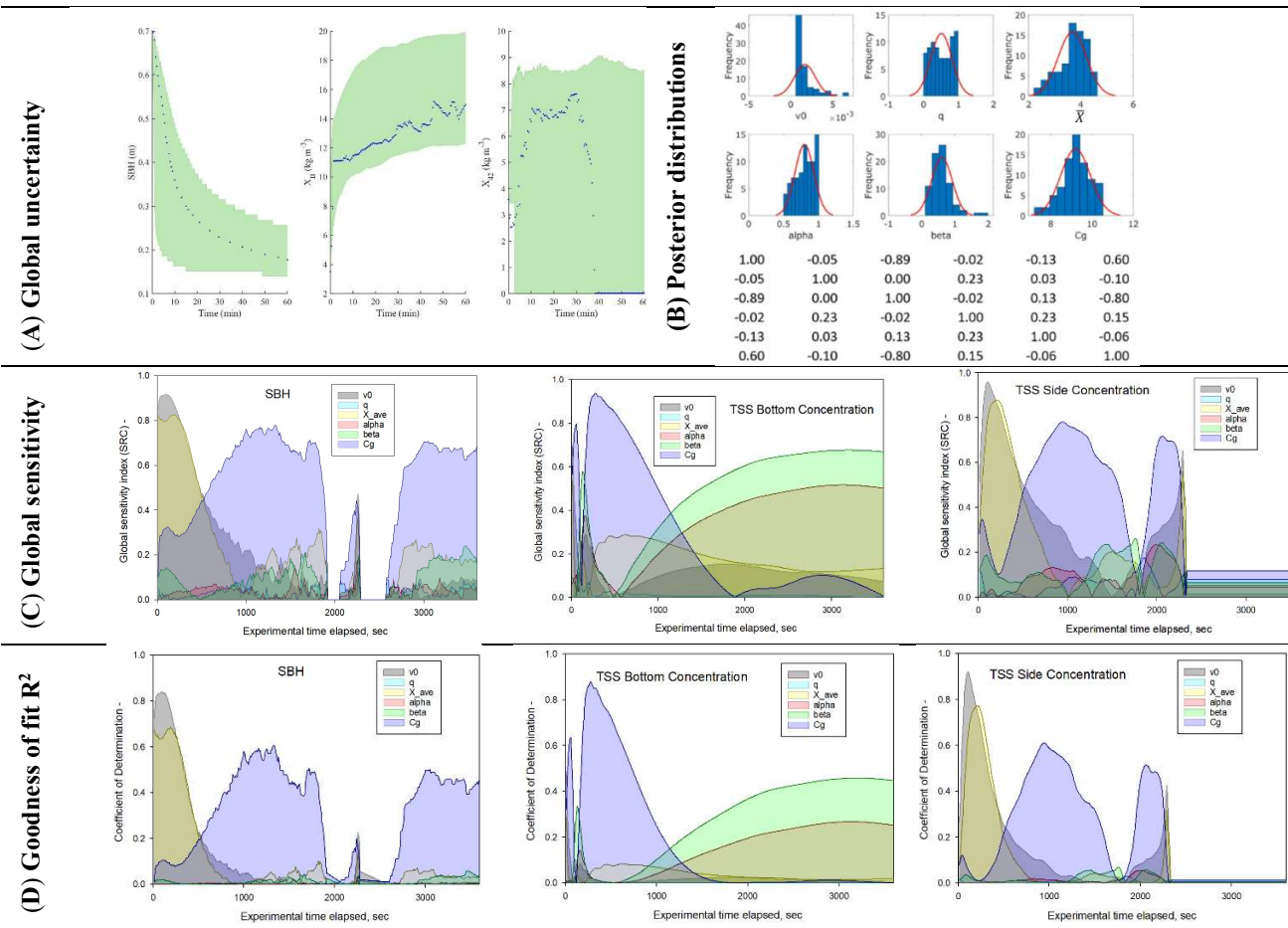
794

795



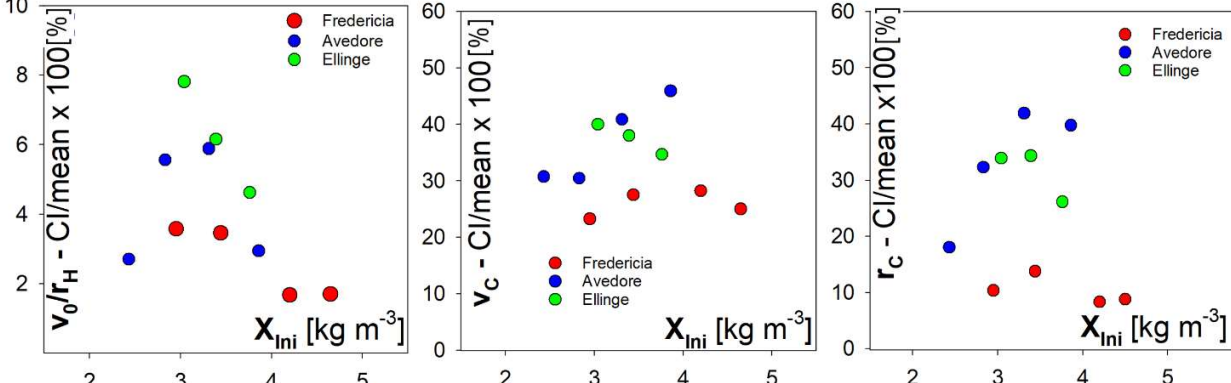
796
797
798
799
800

Figure. SI-ii2 Practical model identification using the HTC model; Posterior histograms obtained in Step 6@Level 1 (Table 1) as well as values of SRC and Goodness-of-fit for linear regression (R²).



801
 802 **Figure. SI-ii3** Practical identifiability test of the Diehl-DeClercq model. Posterior histograms
 803 obtained in Step 6@Level 1 (Table 1) as well as values of SRC and Goodness-of-fit for linear
 804 regression (R^2).
 805
 806
 807

808 **3. Confidence interval as parameter estimator for the new settling**
809 **velocity function**
810



811 **Figure. SI-iii1** Width of confidence intervals, CI [%] – Values of confidence interval normalised to
812 the mean posterior parameter values estimated times 100.
813
814

815

816 **4. CFD simulations**

817

818 **4.1. Full-scale secondary settling tank, OBVA WRRF, Vila-Real, Spain**

819

820 A 3-D axi-symmetrical domain was developed for the SST (Circular; diameter: 30 m; Q=22486
821 m³/day). Turbulence was predicted in the simulation model using the Shear Stress Transport model.822 Transient simulations Herschel-Bulkley model was implemented to predict the rheological
823 behaviour of sludge

$$\tau = \tau_0 + K\gamma^n$$

$$\eta = \frac{\tau}{\dot{\gamma}}$$

824

825 where τ_0 is the yield stress; K is the consistency index; n is the flow behaviour index; and h is the
826 apparent viscosity of sludge. These three variable parameters were calculated using the regression
827 equations

828
$$\tau_0 = AX^B.$$

829
$$K = \mu_w \exp(C.X)$$

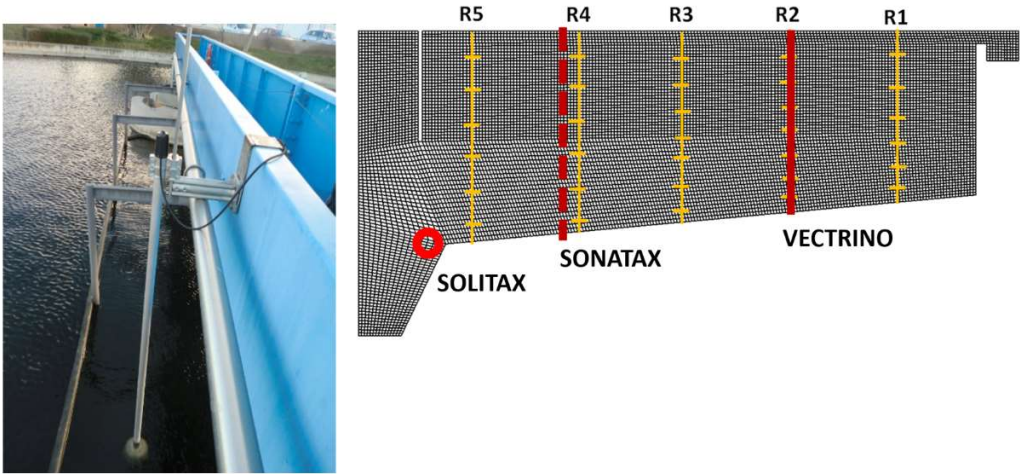
$$n = \frac{1}{1 + D.X^E}$$

830

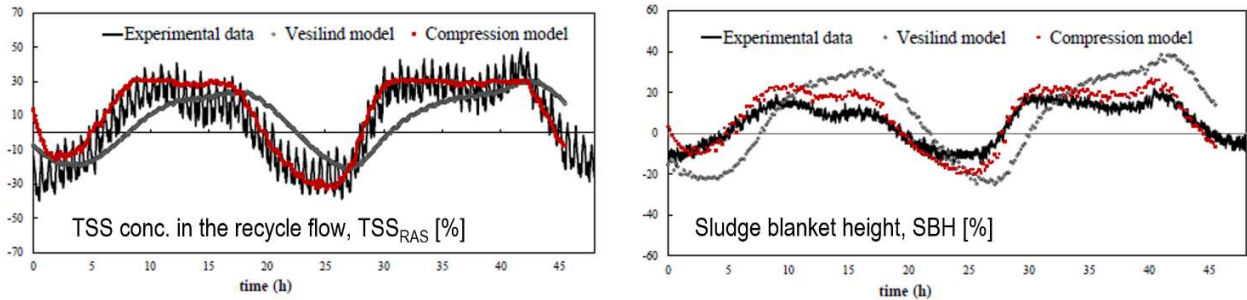
831 where $A = 0.00066$ [$\text{kg}^{1-B}\text{m}^{3-B-1}\text{s}^{-2}$], $B = 2.18$ [-], $C = 0.28$ [m^3kg^{-1}], $D = 0.00083$ [$\text{m}^{3-E}\text{kg}^{-E}$], $E = 2.57$ [-]
832 according to Ramin *et al.* (2014).833 For calibrating the hindered settling function, a sequence of six batch settling column tests were
834 carried out onsite (Fig. S1), and the v_0/r_H parameter value ($0.0024 \text{ kg m}^{-2} \text{ s}^{-1}$) was estimated by
835 considering $v_0 = 0.0025$ (m s^{-1}) and using the exponential regression function, $f = 0.0025 * \exp(-$
836 $r_H * X(i))$ in SigmaPlot 13. The r_H value obtained is 0.98 ($\text{m}^3 \text{kg}^{-1}$). In the absence of column sensor
837 data in terms of solids concentration in the bottom and side-wall – as proposed in this study – the
838 compressive solid stress was calibrated in the CFD model using the average v_C and r_C parameter
839 values (0.1 [$\text{m}^2 \text{ s}^{-2}$] and 1.5 [-]) obtained for the Avedøre WRRF. This was done based on the v_0/r_H
840 values and due to the similarity in terms of SVI values obtained in the two WRRFs.

841

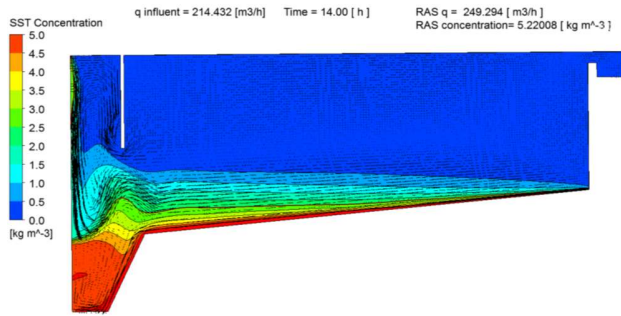
842 In the on-site measurements, radial velocity was measured using a Vectrino (Nortek, USA) high-
843 resolution acoustic veloci-meter at five equidistant positions.



844
 845 **Figure. SI-iv1** The location of velocimeter measurements done in one of the SSTs in the OBVA
 846 WRRF, Vila-Real, Spain. Red solid line indicates the position where the velocity meter profile
 847 (VECTRINO), shown in Fig. 4, were recorded at. The positions of other sensors including the
 848 SONATA and SOLITAX sensors are indicated with dashed line and circle, respectively.
 849



850
 851 **Figure. SI-iv2** Experimental and simulation results obtained in the measurement campaign in the
 852 Vila Real WWTP, Spain. % values were calculated by dividing each value with the mean measured
 853 value times 100. Simulation results were obtained using only Vesilind and the new settling velocity
 854 function, including the hindered-compression constitutive functions.
 855
 856



857
 858 **Figure. SI-iv3** TSS performance of the clarifier at t=14 hours of the 48-hour long measurement
 859 campaign.
 860

861

862 **4.2. Development of a CFD simulation model of the experimental sludge settling column**

863 A Computational Fluid Dynamics (CFD) model of the sludge settling column was developed in the
 864 commercial software package: ANSYS® CFX. The fundamental two-phase modelling approach
 865 used was Eulerian. More specifically, the single-phase Eulerian Drift Flux Model (DFM) was used
 866 to describe the behaviour of the dispersed phase (biological flocs) relative to the constrained
 867 continuous phase (pure water). The DFM solves a single set of continuity (Eqn. 1) and momentum
 868 (Eqn. 2) equations for a fictitious variable composition mixture, with an additional ‘drift’ equation
 869 (Eqn. 3) describing the relative motion of the dispersed to the continuous material. This modelling
 870 approach is commonly applied to activated sludge modelling (Brennan, 2001; Ramin *et al.*, 2014).

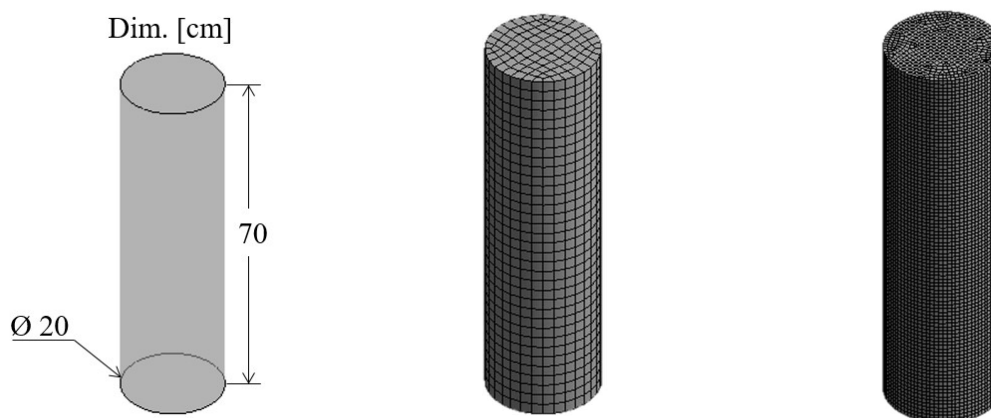
$$\frac{\partial \rho_m}{\partial t} + \nabla \cdot (\rho_m v_m) = 0 \quad (1)$$

$$\frac{\partial \rho_m v_m}{\partial t} + \nabla \cdot (\rho_m v_m v_m) \quad (2)$$

$$= -\nabla \cdot P_m + \nabla \cdot [\tau + \tau^t] - \nabla \cdot \left(\frac{\alpha_d}{1 - \alpha_d} \frac{\rho_c \rho_d}{\rho_m} v_{dj} v_{dj} \right) + \rho_m g + M_m$$

$$\frac{\partial \alpha_d}{\partial t} + \nabla \cdot (\alpha_d v_m) = -\nabla \cdot \left(\frac{\alpha_d \rho_c}{\rho_m} v_{dj} \right) + \nabla \cdot \Gamma \nabla \alpha_d \quad (3)$$

871 Implementing the above modelling approach and the Hindered-Compression settling velocity model
 872 developed in this study, a prototype CFD model of the experimental settling column was produced
 873 in ANSYS® CFX. This prototype model used a set of model parameters and activated sludge
 874 physical data and could hence be benchmarked against their experimental settling data during the
 875 subsequent column meshing studies. These meshing studies considered a column geometry
 876 identical to that of the experimental sludge settling column and tested mesh sizes ranging from
 877 coarse (3,360 elements) to very-fine (57,288). It was concluded that the medium mesh size,
 878 comprising 7,708 elements, resulted in an accurate reproduction of experimental data while
 879 minimising computational effort and avoiding issues associated with numerical instability and poor
 880 convergence in near-wall regions experienced when using a very-fine mesh.



881 **Figure. SI-iv4** The basic column geometry considered for the CFD model, **(3.1)** the coarsest
 882 (‘coarse’) column mesh tested (3,360 elements), **(3.2)** the finest (‘very-fine’) column mesh tested
 883 (57,288 elements)

884

885 In further developing the sludge settling column CFD model, it was necessary to consider both the
 886 sludge rheology and turbulence modelling approaches taken. This was done via scenario simulation,
 887 once again using the experimental data as a benchmark for sludge blanket height (SBH) and sludge

888 concentration at the bottom (X_B) and side (X_{side}) of the column over the course of the 60 minute
889 settling experiment. It was determined, by the comparison of a simple Newtonian viscosity model
890 and the non-Newtonian Ostwald de Waele or Power Law model, that the treatment of sludge
891 rheology in the CFD model had little to no impact on the accuracy with which experimental data
892 sets could be replicated. For this reason, the simpler Newtonian model was adopted to minimise the
893 complexity of the simulation. Alternatively, the chosen turbulence modelling approach was found
894 to greatly impact the replication of experimental SBH, X_B and X_{side} profiles, particularly within the
895 so-called ‘lag-phase’ at the commencement of the settling experiment. The Re-Normalisation
896 Group (Yakhot et al., 1992) and Shear Stress Transport (Menter, 1994) k- ϵ derivatives were
897 compared via scenario simulation and the latter found far superior with regard to the present
898 application, particularly in the prediction of X_{side} .
899
900

5. Compression settling and compression solids concentration

For modelling of the onset of compression settling, Guyonvarch et al. (2015) assess the setting of the variable compressive threshold concentration (X_C) using state-of-the-art models (Bürger et al., 2013; Ramin et al., 2014; De Clercq, 2006, De Clercq et al., 2008). Based on the relative predictive error, computational time and a separate model validation test, the approach of setting X_C as a function of the initial solid concentration and the SST feed concentration for simulating batch tests and SST, respectively, is found superior over other methods.

Notably, the two models of time-dependent X_C by DeClercq (2006) are of interest as they are identified based on in-depth radiotracer experimental data, i.e. (a) $X_C = X_{SBH+5}$ where X_{SBH+5} is the concentration of the layer located 5 layers below the top of the sludge blanket (De Clercq, 2006); and (b) the concentration of the highest layer within the sludge blanket where the concentration gradient falls below 200g/L/m (De Clercq et al., 2008). It is noteworthy that both of these models failed the discrimination tests (Table SI-v.1), which partly prompted the present research as well.

Table SI-v.1 Comparison of different second-order 1-D-model structures in terms of feed-layer location and compression threshold concentration – a model discrimination study carried out by Guyonvarch et al. (2015). For the discretization of the Model 0, 90 layers are used and pseudo-dispersion D_f is considered only around the feed layer, at a distance $\min(H_{in}; SWD - H_{in})/2$ (Bürger et al., 2013). For Model 1 to 8, the number of layers used is 60 layers and a single pseudo-dispersion D_0 constant along the tank (Plósz et al., 2007). For more details on the validation test, see Chapter 2.3.3 in Guyonvarch et al. (2015). ^a - Computational time evaluation is expressed as *Low* (few seconds), *High* (hours) and *Too high* (up to several days). An *Acceptable* computational time is considered from several seconds to few minutes. ^b - The validation test is considered as failed if the mean SSRE between 1-D model predictions and CFD outputs is significantly higher than 1.

Model #	Feed layer location	Compression threshold X_C	Selection criteria		
			Averaged SSRE for the LHS experiments after calibration (except experiment #35)	Computational time ^a	Validation test ^b
Model 0	Fixed feed layer at the middle of the tank (Bürger et al., 2011, 2005)	$X_C = 6g/L$ (Bürger et al., 2013, 2011)	25.2	High	
Model 1	<ul style="list-style-type: none"> Highest layer where $X > X_{in}$ (Fig. 5) Maximum height of the feed layer at 53% of the clarifier depth (Plósz et al., 2007) 	$X_C = 1.1 * X_{in}$ (Ramin et al., 2014)	139	Low	Failure
Model 2	<ul style="list-style-type: none"> Highest layer where $X > X_{in}$ Maximum height of the feed layer at 53% of the clarifier depth (Plósz et al., 2007) 	$X_C = 1.1 * X_{feed_layer}$	1.28	Low	Failure
Model 3	Highest layer where $X > X_{in}$	$X_C = X_{SBH+5}$ where X_{SBH+5} is the concentration of the layer located 5 layers below the top of the sludge blanket (De Clercq, 2006)	Higher SSRE than with Model 2 for the cases with low sludge blanket (most of the cases)	Acceptable	
Model 4	Layer where $X = X_C$ (De Clercq, 2006)	Concentration of the highest layer within the sludge blanket where the concentration gradient falls below 200g/L/m (De Clercq et al., 2008)	Higher SSRE than with Model 2 for the cases with low sludge blanket (most of the cases)	Too high	
Model 5	Feed layer located at the top of the sludge blanket ($X > 0.8g/L$) (based on Anderson, 1945 and Fig.5)	$X_C = 1.1 * X_{feed_layer}$	0.214	High	Failure
Model 6	Highest layer where $X > X_{in}$	$X_C = 1.1 * X_{feed_layer}$	0.247	Acceptable	
Model 7 SELECTED MODEL	Highest layer where $X > X_{in}$	$X_C = X_{feed_layer+1}$ where $X_{feed_layer+1}$ is the concentration located just below the feed layer	0.206	Acceptable	Success
Model 8	Fixed feed layer at $z = H_{in}$ (Bürger et al., 2012)	$X_C = X_{feed_layer+1}$ where $X_{feed_layer+1}$ is the concentration located just below the feed layer	34.4	Low	

Theoretically, the solid concentration in the volume of slurry settling in the hindered regime in a settling column test corresponds to that of the initial concentration. Any increase of the local concentration in the volume of slurry settling in the **hindered settling** regime is caused by some

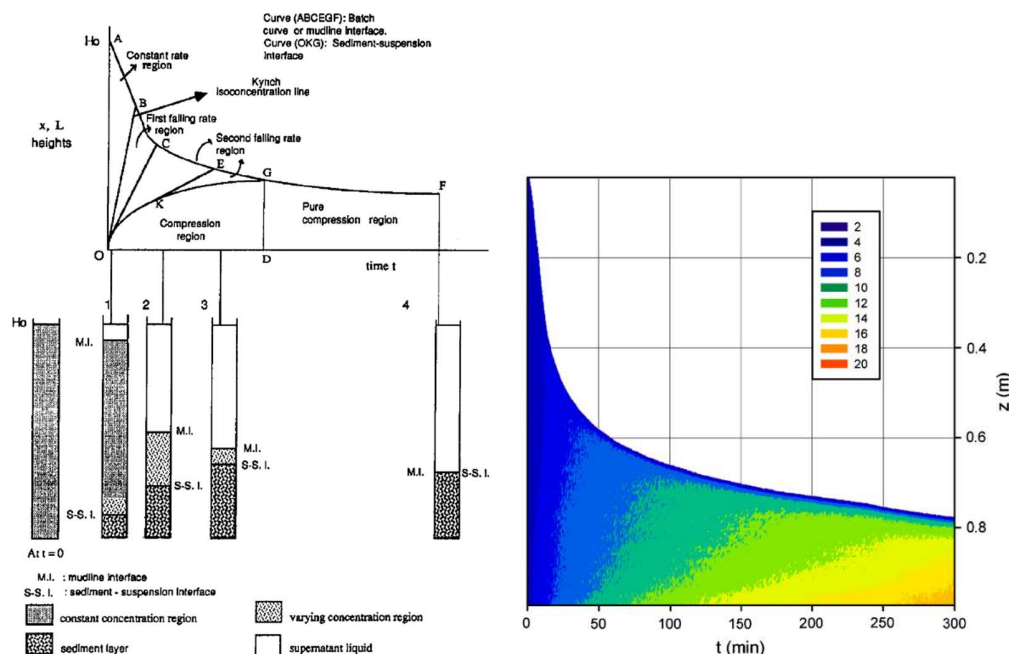
931 degree of solid stress through interactions with the increasing cake in the bottom of batch column
 932 (Fig. SI.v.2). According to Kynch's theory, in the falling hindered settling rate region, also referred
 933 to as **transient settling**, straight isoconcentration lines propagate from the bottom of the cylinder.
 934 Furthermore, tangential isoconcentration lines propagate from the sediment-suspension interface
 935 during **compression settling**. Straight isoconcentration lines suggest a first-order process, same as
 936 hindered settling, whereas the curved isoconcentration lines indicate a second-order phenomenon,
 937 i.e. compression settling.

938 DeClercq (2006) suggest modelling sedimentation transport by considering only hindered and
 939 compression settling, and by employing a time-dependent onset of compression through the
 940 aforementioned X_C models.

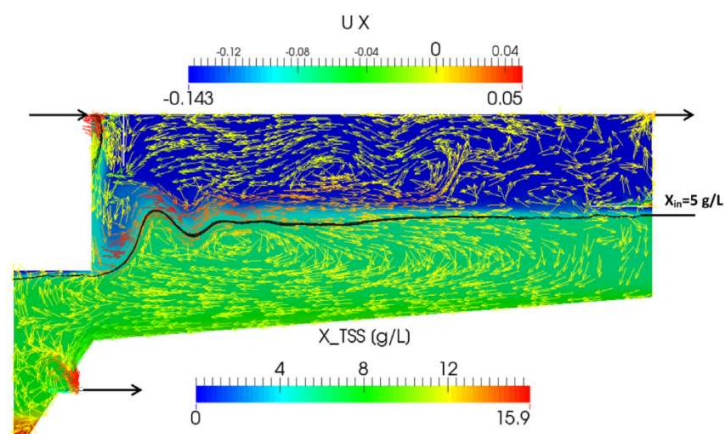
941 Partly because DeClercq (2006)'s models seem to overestimate the transient settling velocity and
 942 due to challenges in implementing the proposed X_C models in SST simulation models, Ramin *et al.*
 943 (2014) propose a model that additionally includes a first-order transient settling function,
 944 formulated analogously to hindered settling (straight isoconcentration lines), and by employing two
 945 threshold concentrations for the onset of transient (X_T) and then for compression settling, X_C .

946 In contrast to these previous approaches, besides the hindered settling, the present paper proposes a
 947 compression settling function to predict any effects of solid stress propagating through the sludge
 948 blanket (i) by setting $X_C = X_{Ini}$ and (ii) by formulating the solids stress independently from the X_C
 949 value and from the relative concentration ($X - X_C$) that can allow first-order solid settling behaviour
 950 to occur at relatively low X .

951 Additionally, the physical justification for setting X_C at the feed solid concentration for SST
 952 modelling is that the density current of the feed slurry tends towards zero buoyancy, and propagates
 953 through volumes, under which, the descending particles shall increase the local concentration only
 954 if they exhibit compressive solids settling (Fig. SI.v.3; Guyonvarch *et al.*, 2015).
 955



956
 957 **Figure SI.v.2.** Solid sedimentation – (left) Theory according to (Diplas and Papanicolaou, 1997)
 958 and (right) measurements according to DeClercq *et al.* (2008) - Iso-concentration contour plot
 959 during batch settling of Destelbergen Sludge.
 960



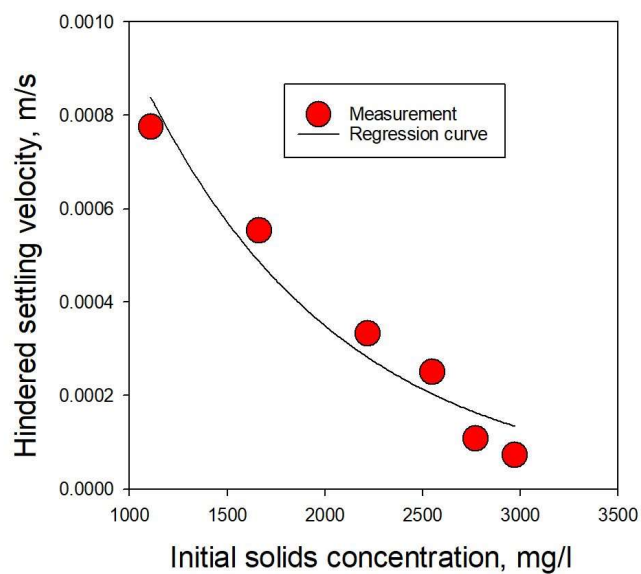
961 **Figure SI.v.3.** The SST influent density current – the rationale for the X_C modeling approach in
 962 CFD as well as feed-layer selection in 1-D SST modelling (Guyonvarch *et al.*, 2015). The velocity
 963 vectors (arrows) are coloured according to the magnitude of the horizontal velocity component (UX
 964 in m/s) – not scaled according to the velocity magnitude. X (g/L) is represented across the tank, and
 965 the iso-contour corresponds to the inlet/feed concentration X_{Feed} (solid black line).
 966
 967
 968

969 References

- 970
 971 Brennan, D., 2001. The numerical simulation of two phase flows in settling tanks. Imperial College
 972 London (University of London).
 973 De Clercq, J., 2006. *Batch and continuous settling of activated sludge: in-depth monitoring and 1-D*
 974 *compression modelling*. Ph.D. Thesis, University of Gent, Belgium.
 975 De Clercq, J., Nopens, I., Defrancq, J., Vanrolleghem, P.A., (2008). Extending and calibrating a
 976 mechanistic hindered and compression settling model for activated sludge using in-depth batch
 977 experiments. *Water Res.* 42 (3), 781-791.
 978 Diplas, P., Papanicolaou, A.N., 1997. Batch analysis of slurries in zone settling regime. *J. Environ.*
 979 *Eng.* 123, 659–667.
 980 Diehl, S., (2008). The solids-flux theory – confirmation and extension by using partial differential
 981 equations. *Water Res.* 42, 4976–4988.
 982 Guyonvarch, E., Ramin, E., Kulahci, M., Plósz, B.G. 2015. iCFD: Interpreted computational fluid
 983 dynamics - Degeneration of CFD to one-dimensional advection-dispersion models using
 984 statistical experimental design - The secondary clarifier. *Water Research* 83, 396-411.
 985 Menter, F.R., 1994. 2-Equation Eddy-Viscosity Turbulence Models for Engineering Applications.
 986 *Aiaa Journal*, 32(8), pp. 1598-1605.
 987 Ramin, E., Wágner, D.S., Yde, L., Binning, P.J., Rasmussen, M.R., Mikkelsen, P.S., Plósz, B.G.
 988 (2014b). A new settling velocity model to describe secondary sedimentation. *Water Research*
 989 66, 447-458.
 990 Torfs, E., Balemans, S., Locatelli, F., Diehl, S., Bürger, R., Laurent, J., Francois, P., Nopens, I.,
 991 (2017). On constitutive functions for hindered settling velocity in 1-D settler models: selection
 992 of appropriate model structure. *Water Research* 110, 38-47.
 993 Valverde-Perez, B., Penkarski-Rodon, E., Wagner, D. and Plosz, B., 2016. Secondary settling sensor
 994 setup development – testing prototypes and compression models via practical model parameter
 995 identifiability assessment IWA Specialist Conference: Advances in particle science and
 996 separation, Oslo, Norway.
 997 Vesilind, P.A., (1968). Design of prototype thickeners from batch settling tests. *Water Sewage Works*

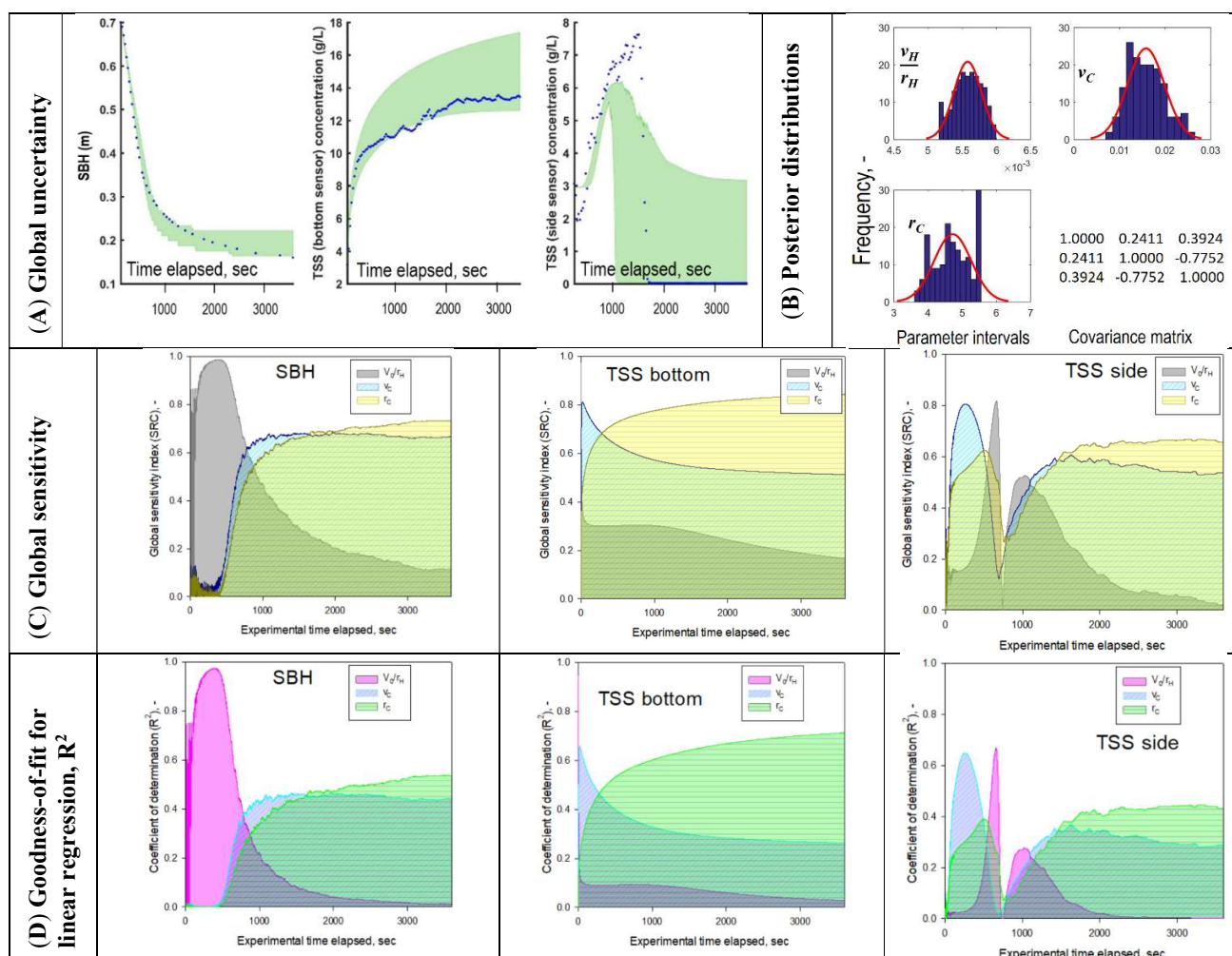
- 998 115, 302-307.
999 Yakhot, V., Orszag, S.A., Thangam, S., Gatski, T.B. and Speziale, C.G., 1992. Development of
1000 Turbulence Models for Shear Flows by a Double Expansion Technique. *Physics of Fluids a-*
1001 *Fluid Dynamics*, 4(7), pp. 1510-1520.
1002
1003
1004
1005

1006 **6. Figures specifically referred to in the main body of text**
1007

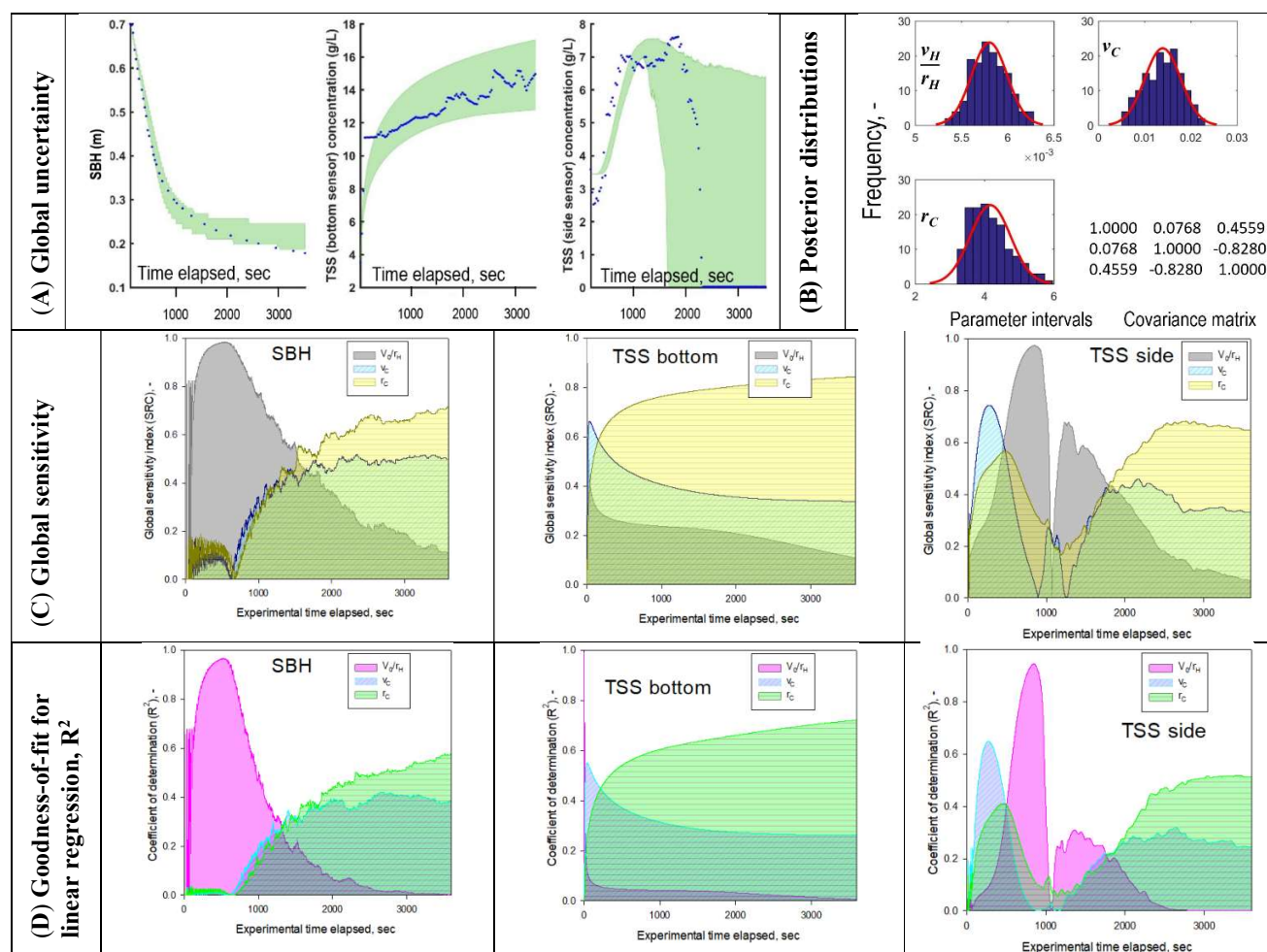


1008
1009 **Figure S1.** Hindered settling velocity model calibration in the CFD simulation model using 6
1010 dilutions of activated sludge samples in the full-scale secondary settling tank, OBVA WRRF, Vila-
1011 Real, Spain.

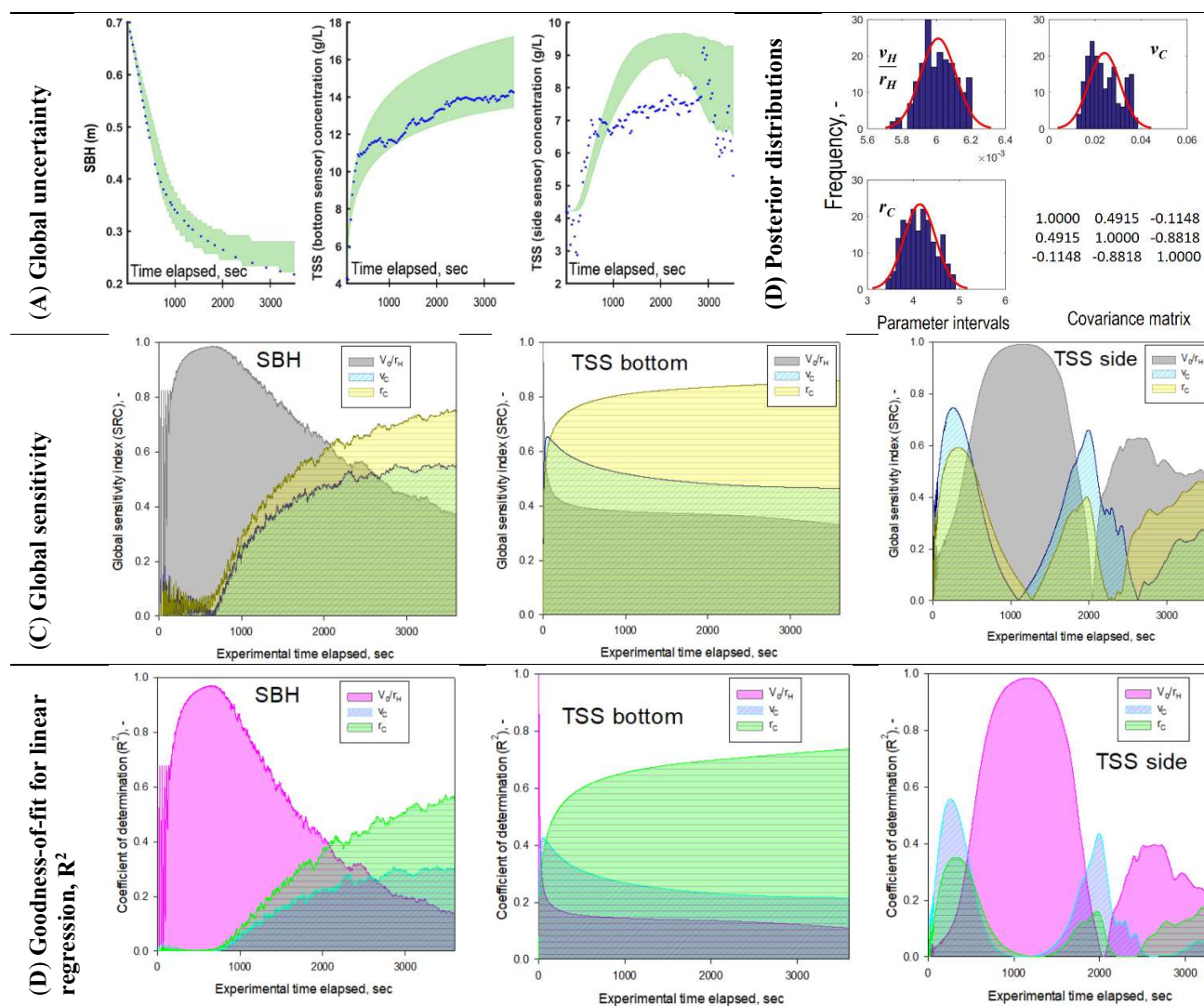
1012
1013
1014



1015 **Figure S2.** Practical model identification using data obtained with well-settling solids – measured
 1016 and simulated data for activated sludge collected in Fredericia WRRF using the new hindered-
 1017 compression process model; Initial solid concentration: 2.95 g L^{-1} ; Proposed/a priori probability
 1018 ranges: $v_0/r_H=[0.005 \ 0.0062]$; $v_C=[0.005 \ 0.027]$; $r_C=[2.5 \ 5.5]$; Global uncertainty plots with 95%
 1019 confidence intervals, posterior parameter density distributions, values of dynamic global sensitivity
 1020 (SRC) and Goodness-of-fit for linear regression (R^2) computed for SBH, TSS bottom and side
 1021 concentration state-variables.
 1022

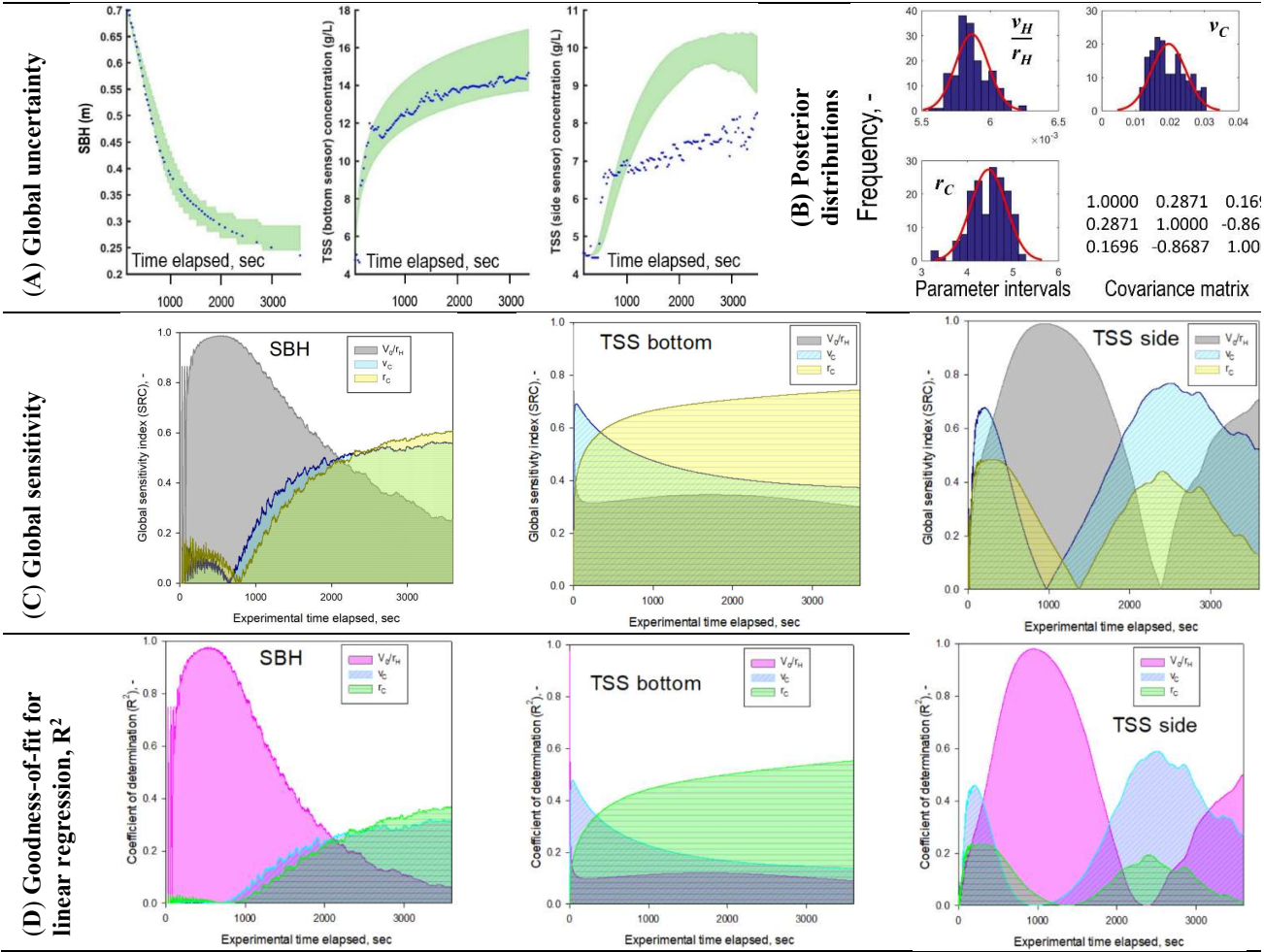


1023 **Figure S3** Practical model identification using data obtained with well-settling solids – measured
 1024 and simulated data for activated sludge collected in Fredericia WRRF using the new hindered-
 1025 compression process model; Initial solid concentration: 3.44 g L^{-1} ; Proposed/a priori probability
 1026 ranges: $v_0/r_H=[0.0052 \ 0.0063]$; $v_C=[0.005 \ 0.025]$; $r_C=[2.5 \ 5.5]$; Global uncertainty plots with 95%
 1027 confidence intervals, posterior parameter density distributions, values of dynamic global sensitivity
 1028 (SRC) and Goodness-of-fit for linear regression (R^2) computed for SBH, TSS bottom and side
 1029 concentration state-variables.
 1030
 1031
 1032
 1033
 1034



1035
 1036 **Figure S4.** Practical model identification using data obtained with well-settling solids – measured
 1037 and simulated data for activated sludge collected in Fredericia WRRF using the new hindered-
 1038 compression process model; Initial solid concentration: 4.2 g L^{-1} ; Proposed/*a priori* probability
 1039 ranges: $v_0/r_H=[0.0052 \ 0.0063]$; $v_C=[0.005 \ 0.025]$; $r_C=[2.5 \ 5.5]$; Global uncertainty plots with 95%
 1040 confidence intervals, posterior parameter density distributions, values of dynamic global sensitivity
 1041 (SRC) and Goodness-of-fit for linear regression (R^2) computed for SBH, TSS bottom and side
 1042 concentration state-variables.
 1043
 1044

1045



1046

1047

1048

1049

1050

1051

1052

1053

1054

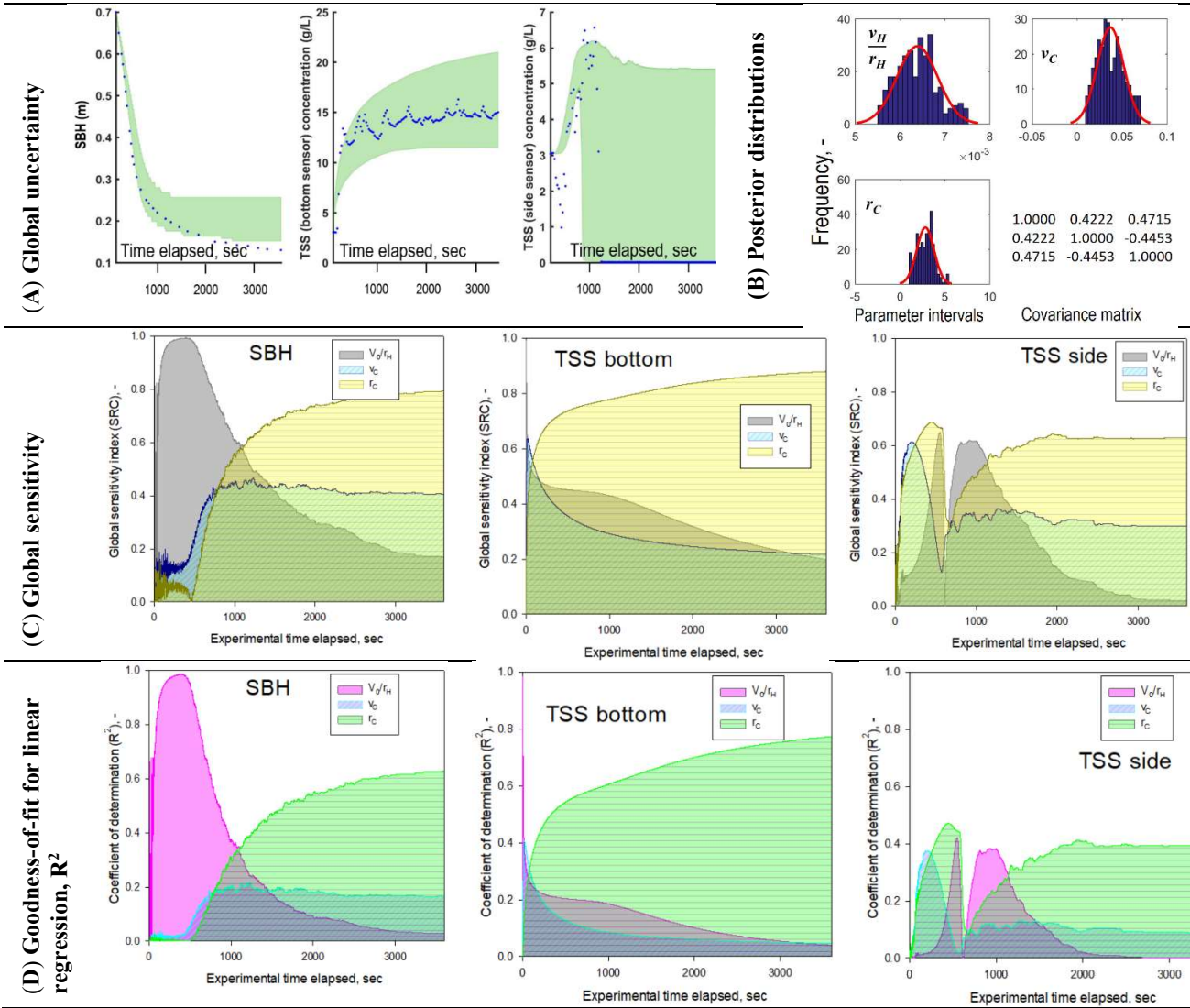
1055

1056

1057

Figure S5. Practical model identification using data obtained with well-settling solids – measured and simulated data for activated sludge collected in Fredericia WRRF using the new hindered-compression process model; Initial solid concentration: 4.5 g L^{-1} ; Proposed/*a priori* probability ranges: $v_0/r_H=[0.0052 \ 0.0063]$; $v_C=[0.005 \ 0.025]$; $r_C=[2.5 \ 5.5]$; Global uncertainty plots with 95% confidence intervals, posterior parameter density distributions, values of dynamic global sensitivity (SRC) and Goodness-of-fit for linear regression (R^2) computed for SBH, TSS bottom and side concentration state-variables.

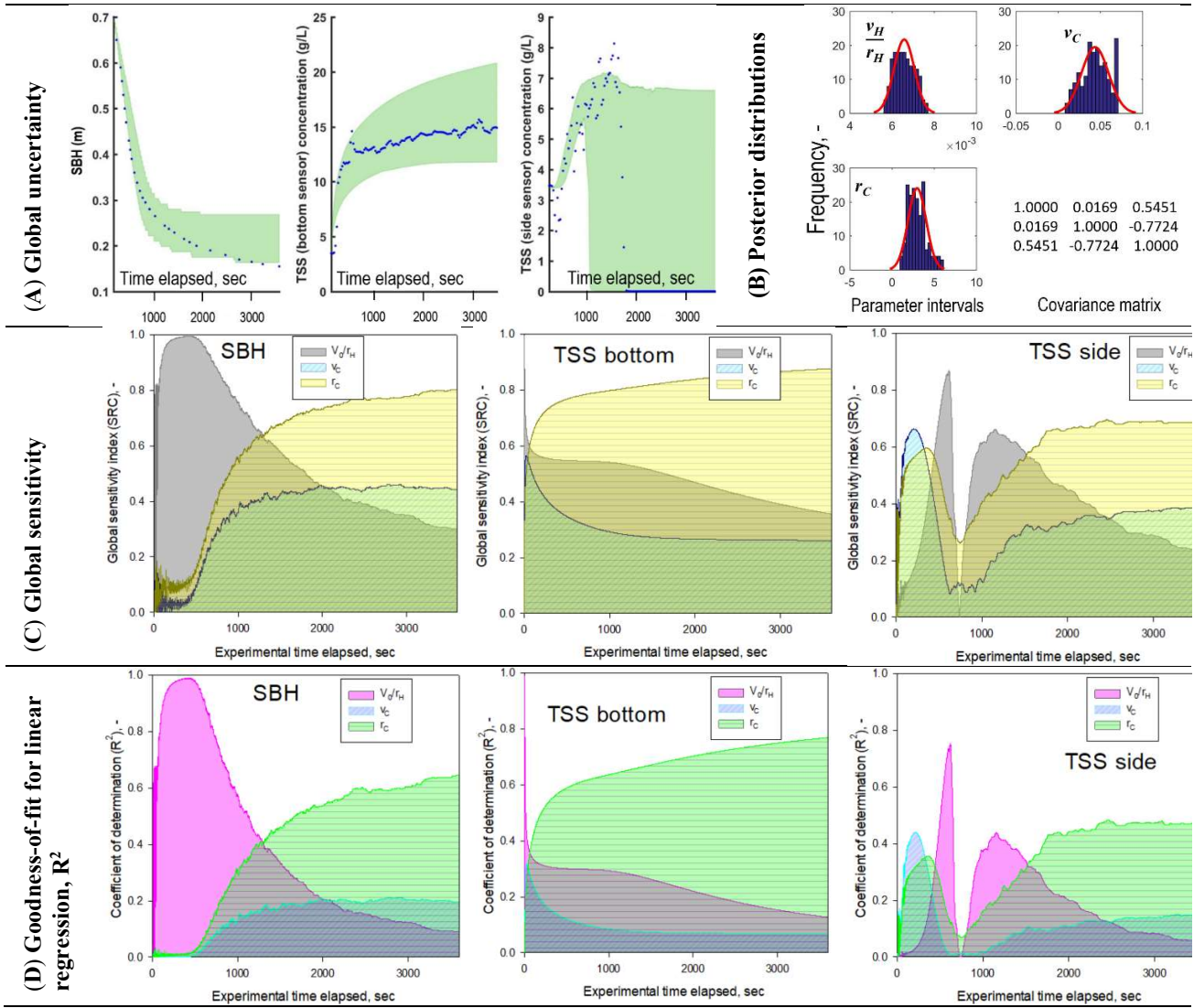
1058



1059
1060
1061
1062
1063
1064
1065
1066
1067
1068

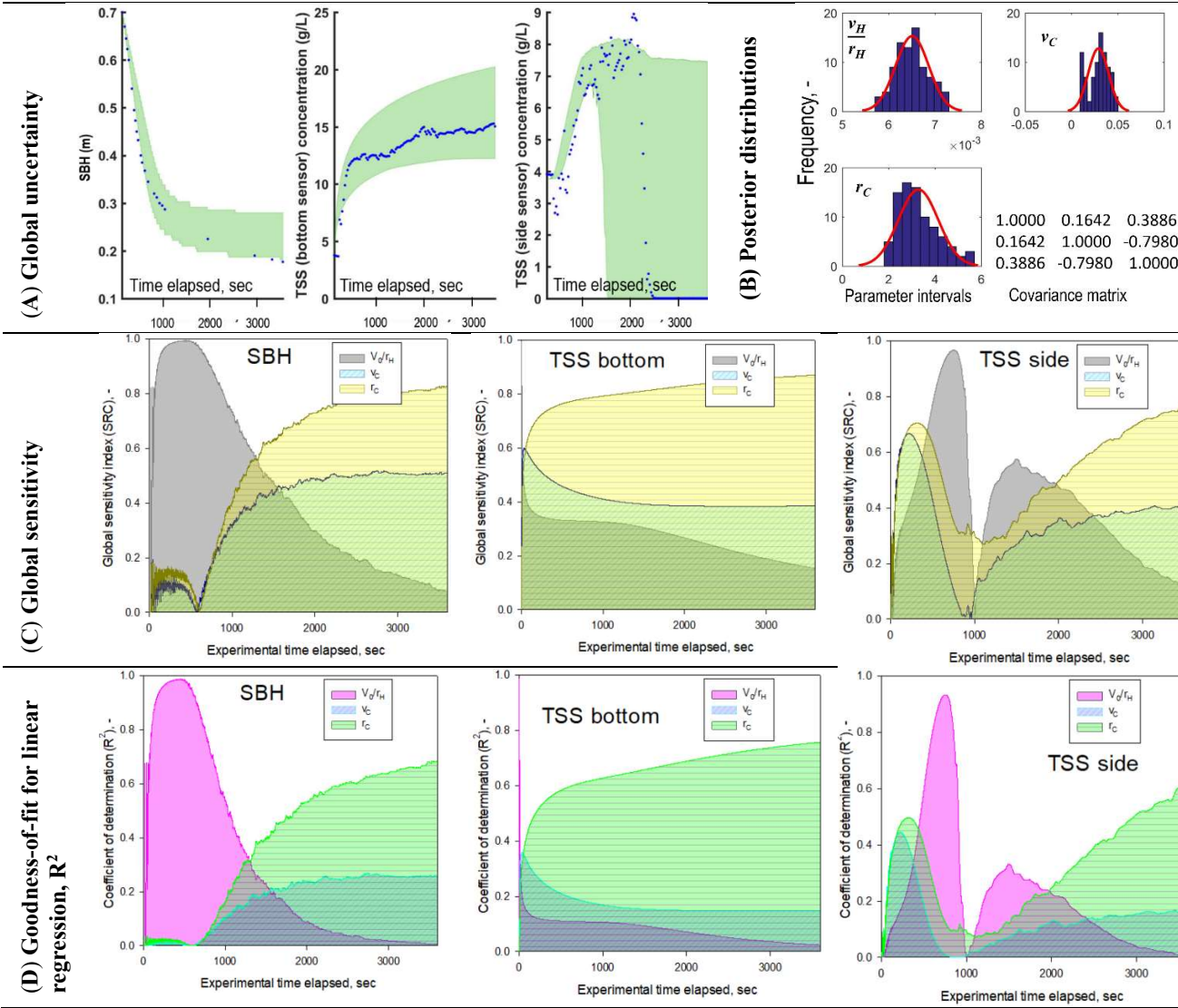
Figure S6. Model validation using data obtained with well-settling solids – measured and simulated data for activated sludge collected in Ellinge WRRF using the new hindered-compression process model; Initial solid concentration: 3.0 g L^{-1} ; Proposed/*a priori* probability ranges: $v_0/r_H=[0.0051 \text{ } 0.0076]$; $v_C=[0.01 \text{ } 0.07]$; $r_C=[0.5 \text{ } 6]$; Global uncertainty plots with 95% confidence intervals, posterior parameter density distributions, values of dynamic global sensitivity (SRC) and Goodness-of-fit for linear regression (R^2) computed for SBH, TSS bottom and side concentration state-variables.

1069



1070
 1071 **Figure S7.** Model validation using data obtained with well-settling solids – measured and simulated
 1072 data for activated sludge collected in Ellinge WRRF using the new hindered-compression process
 1073 model; Initial solid concentration: 3.4 g L^{-1} ; Proposed/a priori probability ranges: $v_0/r_H=[0.005$
 1074 $0.0077]$; $v_C=[0.01 \ 0.07]$; $r_C=[0.5 \ 6]$; Global uncertainty plots with 95% confidence intervals,
 1075 posterior parameter density distributions, values of dynamic global sensitivity (SRC) and
 1076 Goodness-of-fit for linear regression (R^2) computed for SBH, TSS bottom and side
 1077 state-variables.
 1078

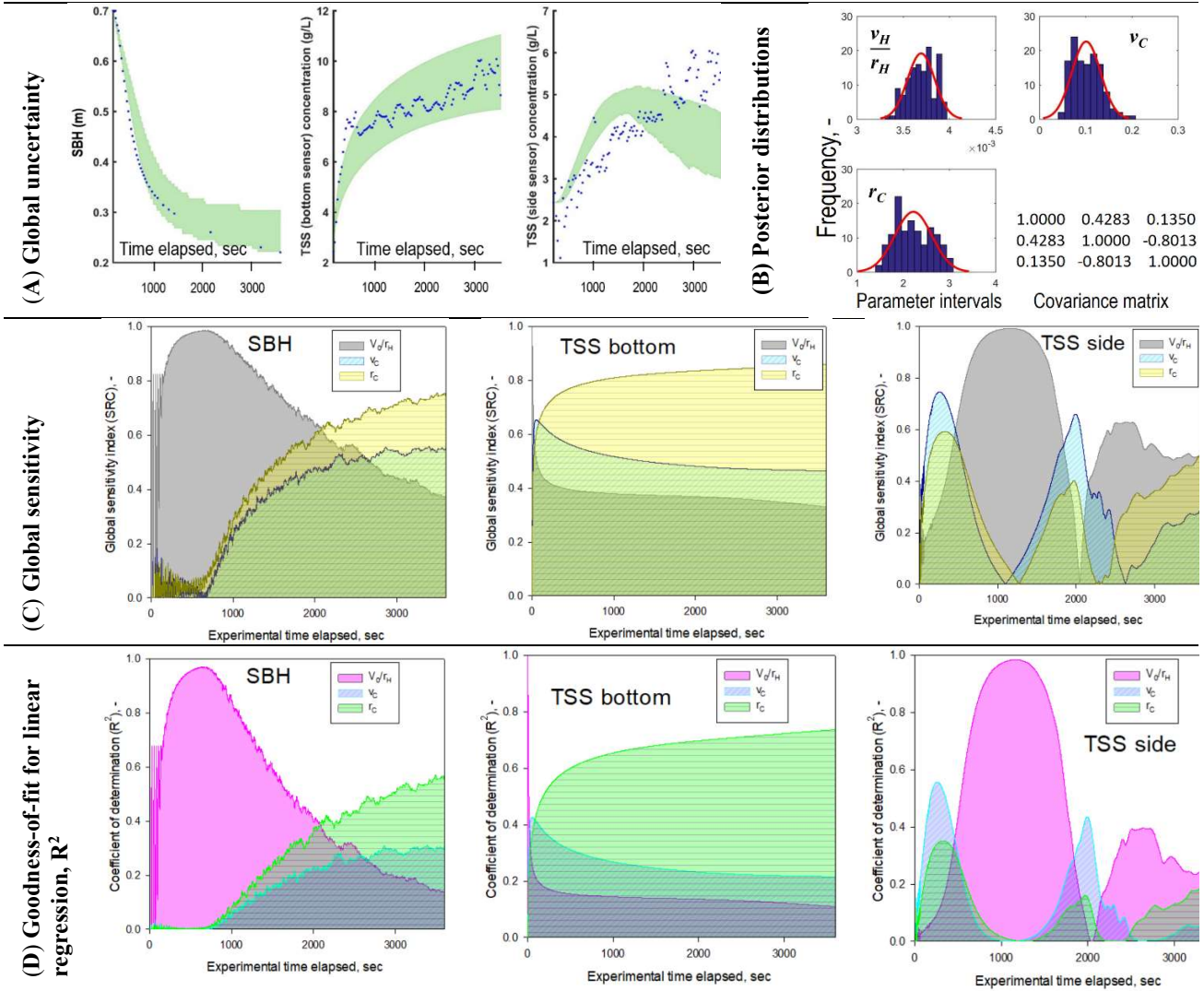
1079



1080
1081
1082
1083
1084
1085
1086
1087
1088
1089

Figure S8. Model validation using data obtained with well-settling solids – measured and simulated data for activated sludge collected in Ellinge WRRF using the new hindered-compression process model; Initial solid concentration: 3.8 g L^{-1} ; Proposed/*a priori* probability ranges: $v_0/r_H=[0.0051 \text{ } 0.0075]$; $v_C=[0.01 \text{ } 0.07]$; $r_C=[0.8 \text{ } 6]$; Global uncertainty plots with 95% confidence intervals, posterior parameter density distributions, values of dynamic global sensitivity (SRC) and Goodness-of-fit for linear regression (R^2) computed for SBH, TSS bottom and side concentration state-variables.

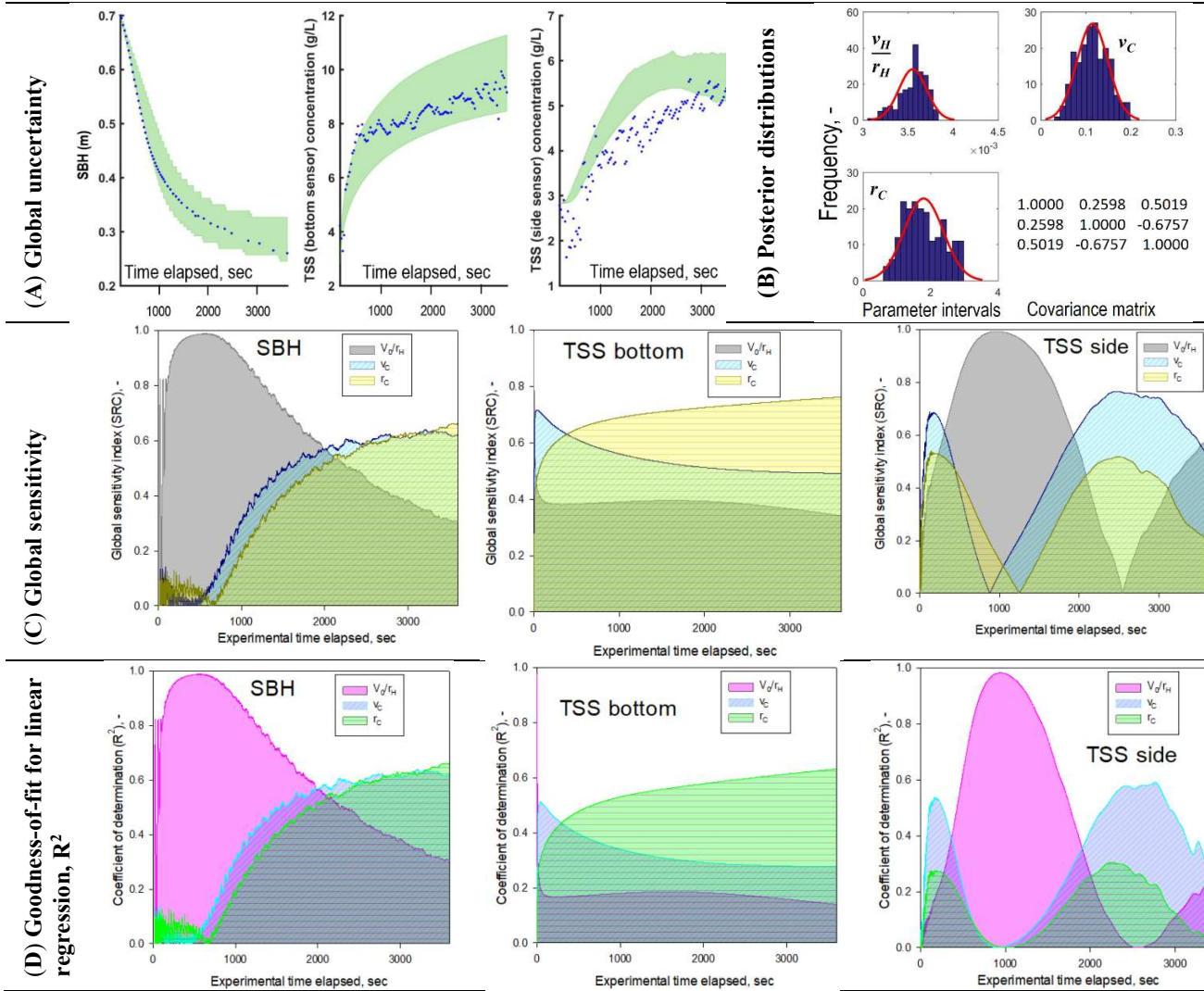
1090



1091
1092
1093
1094
1095
1096
1097
1098
1099
1100

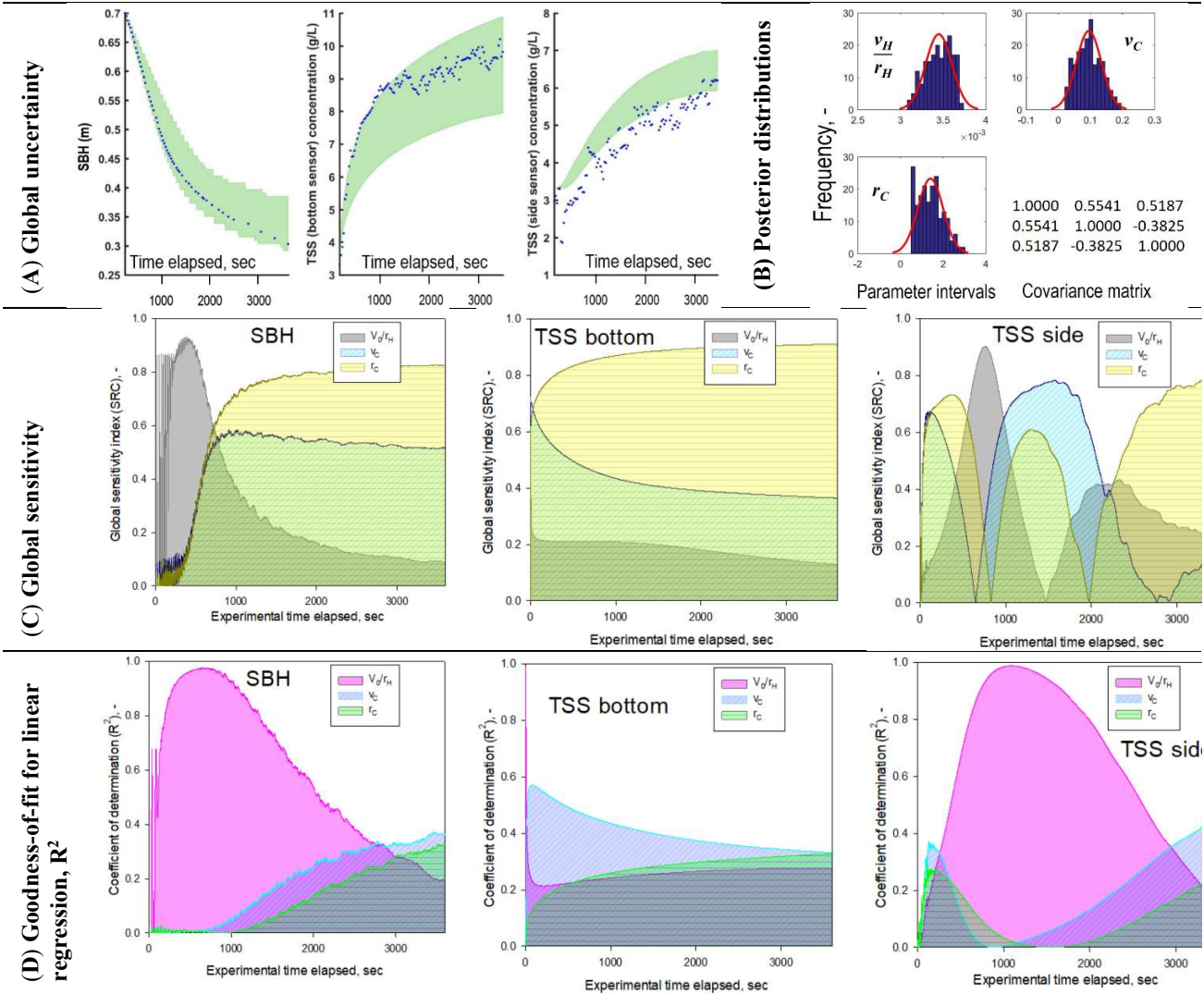
Figure S9. Model validation using data obtained with filamentous bulking solids – measured and simulated data for activated sludge collected in Avedøre WRRF using the new hindered-compression process model; Initial solid concentration: 2.4 g L^{-1} ; Proposed *a priori* probability ranges: $v_0/r_H=[0.003 \ 0.004]$; $v_C=[0.02 \ 0.2]$; $r_C=[0.5 \ 3]$; Global uncertainty plots with 95% confidence intervals, posterior parameter density distributions, values of dynamic global sensitivity (SRC) and Goodness-of-fit for linear regression (R^2) computed for SBH, TSS bottom and side concentration state-variables.

1101



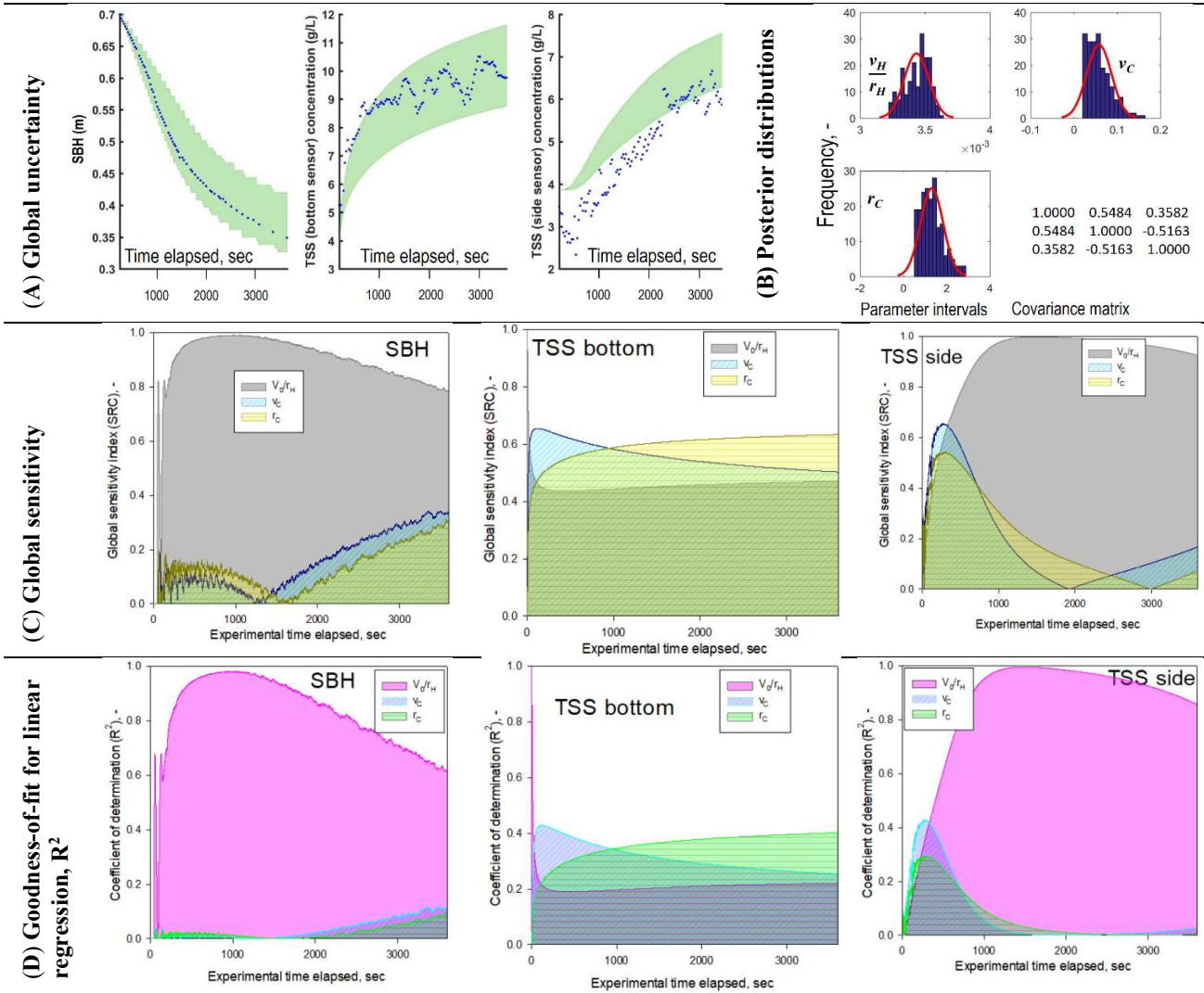
1102
 1103 **Figure S10.** Model validation using data obtained with filamentous bulking solids – measured and
 1104 simulated data for activated sludge collected in Avedøre WRRF using the new hindered-
 1105 compression process model; Initial solid concentration: 2.8 g L⁻¹; Proposed/a priori probability
 1106 ranges: $v_0/r_H=[0.003\ 0.004]$; $v_C=[0.02\ 0.2]$; $r_C=[0.5\ 4]$; Global uncertainty plots with 95%
 1107 confidence intervals, posterior parameter density distributions, values of dynamic global sensitivity
 1108 (SRC) and Goodness-of-fit for linear regression (R^2) computed for SBH, TSS bottom and side
 1109 concentration state-variables.
 1110

1111



1112 **Figure S11.** Model validation using data obtained with filamentous bulking solids – measured and
 1113 simulated data for activated sludge collected in Avedøre WRRF using the new hindered-
 1114 compression process model; Initial solid concentration: 3.3 g L⁻¹; Proposed/*a priori* probability
 1115 ranges: $v_0/r_H=[0.003\ 0.004]$; $v_C=[0.02\ 0.2]$; $r_C=[0.8\ 4]$; Global uncertainty plots with 95%
 1116 confidence intervals, posterior parameter density distributions, values of dynamic global sensitivity
 1117 (SRC) and Goodness-of-fit for linear regression (R^2) computed for SBH, TSS bottom and side
 1118 concentration state-variables.
 1119
 1120
 1121
 1122
 1123
 1124

1125



1126
 1127 **Figure S12.** Model validation using data obtained with filamentous bulking solids – measured and
 1128 simulated data for activated sludge collected in Avedøre WRRF using the new hindered-
 1129 compression process model; Initial solid concentration: 3.9 g L⁻¹; Proposed/a priori probability
 1130 ranges: $v_0/r_H=[0.003\ 0.004]$; $v_C=[0.01\ 0.2]$; $r_C=[0.1\ 3]$; Global uncertainty plots with 95%
 1131 confidence intervals, posterior parameter density distributions, values of dynamic global sensitivity
 1132 (SRC) and Goodness-of-fit for linear regression (R²) computed for SBH, TSS bottom and side
 1133 concentration state-variables.
 1134
 1135
 1136
 1137

1138