

1 **Title:** Invader disruption of belowground plant mutualisms reduces carbon acquisition and alters  
2 allocation patterns in a native forest herb  
3

4 **Authors:** Alison N. Hale<sup>1\*</sup>, Line Lapointe<sup>2</sup>, and Susan Kalisz<sup>1</sup>  
5

6 **Affiliation:** <sup>1</sup>Department of Biological Sciences, University of Pittsburgh, Pittsburgh, PA 15260  
7 USA

8 <sup>2</sup>Departement de biologie, Université Laval, Québec, Québec G1V 0A6, Canada  
9

10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42

43 **Corresponding Author (\*):**

44 Email: [anm116@pitt.edu](mailto:anm116@pitt.edu)

45 Phone: 412-624-6164

46 Mailing Address: 4249 Fifth Avenue, Pittsburgh, PA 15260

47 **Summary (200 word limit)**

- 48 • Invasive plants impose novel selection pressures on naïve mutualistic interactions  
 49 between native plants and their partners. Since most plants critically rely on root fungal  
 50 symbionts (RFS) for soil resources, invaders that disrupt plant-RFS mutualisms can  
 51 significantly depress native plant fitness. Here, we investigate the consequences of RFS  
 52 mutualism disruption on native plant fitness in a greenhouse experiment with a forest  
 53 invader that produces known anti-fungal allelochemicals.
- 54 • Over five-months, we regularly applied 1) green leaves of the allelopathic invader  
 55 *Alliaria petiolata*, 2) a non-systemic fungicide to simulate *Alliaria*'s effects, or 3) green  
 56 leaves of non-allelopathic *Hesperis matronalis* (control) to pots containing the native  
 57 *Maianthemum racemosum* and its RFS. We repeatedly measured *Maianthemum*  
 58 physiology and harvested plants periodically to assess carbon allocation.
- 59 • *Alliaria* and fungicide treatment effects are indistinguishable: we observe significant  
 60 inhibition of the RFS soil hyphal network and significant reductions in *Maianthemum*  
 61 physiology (photosynthesis, transpiration, conductance) and allocation (carbon storage,  
 62 root biomass, asexual reproduction) in both treatments relative to the control.
- 63 • Our findings suggest a general mechanistic hypothesis for local extinction of native  
 64 species in ecosystems challenged by allelopathic invaders: RFS mutualism disruption  
 65 drives carbon stress, subsequent declines in native plant vigor, and if chronic, declines in  
 66 RFS-dependent species abundance.

67 **Key Words:** mutualism disruption, invasion, *Alliaria petiolata*, allelopathy, root fungal  
 68 symbionts, physiology, carbon allocation

69 **Introduction**

70 Mutualisms intimately link the success of partner species together. Consequently,  
 71 declines in one partner can profoundly affect the fitness of the other. A critical mutualism for  
 72 most terrestrial plants involves root fungal symbionts (RFS), such as mycorrhizae and dark  
 73 septate endophytes. Plants provide their symbionts with carbon, while RFS enhance the plant  
 74 resource status (Parniske, 2008; Jumpponen, 2001; Newsham, 2011). For plants, the resources  
 75 received are nontrivial – arbuscular mycorrhizal fungi (AMF) can alleviate plant water stress  
 76 (Augé, 2001) and provide up to 80% of plant phosphorus and 25% of plant nitrogen (Marschner  
 77 & Dell, 1994). Similarly, dark septate endophytes can increase shoot phosphorus and nitrogen up

78 to 100% relative to controls (Newsham, 2011). Because plant physiological activity is intimately  
79 tied to resource status (Wright *et al.*, 2004), even short-term disruption of RFS function can have  
80 dramatic impacts on plant growth (e.g. Stinson *et al.*, 2006) and photosynthesis (e.g. Hale *et al.*,  
81 2011).

82 The rapid pace of global change (Millenium Ecosystem Assessment, 2005) creates  
83 growing concern for the stability of plant mutualisms (Staddon *et al.*, 2003; Memmot *et al.*,  
84 2007; Tylianakis *et al.*, 2008; Kiers *et al.*, 2010; Aslan *et al.*, 2013). Specifically, exotic plant  
85 invasion is an important disrupter of pollination, seed dispersal, and microbial mutualisms  
86 (reviewed by Traveset & Richardson, 2014). Several invasive plants produce novel biochemical  
87 weapons (i.e. allelopathic compounds; Callaway & Ridenour, 2004) that are toxic to RFS (see  
88 Cipollini *et al.*, 2012; e.g. *Tamarix* sp., Meinhardt & Gehring, 2012; *Sisymbrium loeselii*,  
89 Bainard *et al.*, 2009; *Alliaria petiolata*, Roberts & Anderson, 2001; Stinson *et al.*, 2006;  
90 Callaway *et al.*, 2008; Cantor *et al.*, 2011). To explore RFS mutualism disruption, we focus on  
91 one of these allelopathic invaders, *Alliaria petiolata* (Brassicaceae, garlic mustard), a prominent  
92 invader of forest understories throughout northeastern North America (reviewed in Rodgers *et*  
93 *al.*, 2008). *Alliaria* produces glucosinolate-derived compounds that are bioactive and exhibit  
94 anti-fungal properties (e.g. allyl isothiocyanate derived from sinigrin; Vaughn & Berhow, 1999).

95 Prior work has established the potency of *Alliaria*'s allelochemicals: Field concentrations  
96 inhibit mycorrhizal spore germination and the growth of fungal hyphae in soil (Cantor *et al.*,  
97 2011) while single applications of fresh *Alliaria* tissue depressed soil respiration rates and the  
98 physiology of an RFS-dependent native herb in the field (Hale *et al.*, 2011). Declines in native  
99 perennial herb (Rodgers, 2008) and tree seedling abundance (Stinson *et al.*, 2007) have been  
100 documented in *Alliaria* invaded forests. Further, at the community level, native plant biodiversity  
101 is negatively correlated with *Alliaria* density (Stinson *et al.*, 2007). These results are consistent  
102 with mutualism disruption as the majority of forest trees, shrubs, and herbs form associations  
103 with RFS (Brundrett & Kendrick, 1988; Whigham, 2004). Thus, the success of this invader has  
104 in large part been attributed to its allelopathic effects on RFS (Roberts & Anderson, 2001;  
105 Stinson *et al.*, 2006; Callaway *et al.*, 2008; Lankau *et al.*, 2009, Lankau, 2010, Cantor *et al.*,  
106 2011). Data from these studies assume a causal chain between *Alliaria* mutualism disruption and  
107 native plant population declines. However, our mechanistic understanding of the intermediate  
108 links in this causal chain - if and how mutualism disruption affects native plant physiological

109 processes and carbon balance over the long-term and the ultimate outcome of sustained  
110 mutualism disruption for individual plants or populations - remains rudimentary.

111 We developed a cost:benefit model for native plants associating with RFS in the presence  
112 or absence of an allelopathic invader (Fig. 1). We hypothesize that if allelochemicals inhibit the  
113 RFS soil hyphal network, plants will receive diminished resource benefits and experience  
114 physiological declines but pay the continuing costs of maintaining internal RFS structures. Root  
115 colonization rates of forest herbs are known to be unaffected by *Alliaria* invasion (Burke, 2008).  
116 Thus, high maintenance costs are expected for these plant species because their internal RFS  
117 structures have low turnover rates and persist for several months in their long-lived roots  
118 (Brundrett & Kendrick, 1990). Re-establishment of the hyphal network is expected to be a high  
119 priority in species obligately-dependent on RFS. Repeated carbon allocation to fungi for re-  
120 growth of the external hyphal network is anticipated to further raise carbon costs. The high cost  
121 of supplying already limited carbon to a malfunctioning mutualist is expected to negate the  
122 benefits received by a plant (*sensu* Johnson *et al.*, 2015) and could profoundly influence its  
123 carbon budget and allocation strategy. This model yields three nested predictions: 1. *Alliaria*'s  
124 allelochemicals inhibit the *external* hyphal network. 2. Loss of this network will translate into  
125 diminished plant physiological activity. 3. Long-term declines in physiological activity and high  
126 carbon demands to re-establish RFS function will reduce plant carbon allocation to fitness-  
127 related traits.

128 Here, we report novel results from a long-term mutualism disruption experiment using  
129 the native forest herb *Maianthemum racemosum* (hereafter *Maianthemum*). We exposed  
130 *Maianthemum* to one of three treatments: 1) *Alliaria* allelochemicals delivered through  
131 applications of fresh leaves onto the soil of potted plants; 2) application of a non-systemic  
132 fungicide with no known phytotoxic effects (Petit *et al.*, 2012) to simulate *Alliaria*'s chemical  
133 effect (positive control); 3) applications of fresh *Hesperis matronalis* (Brassicaceae) leaves  
134 (negative control). We selected *Hesperis* for our negative control because it is an invasive  
135 mustard and RFS have been detected in its roots (DeMars & Boerner, 1995) indicating a  
136 negligible allelopathic effect. We quantified soil RFS hyphal density, plant physiological  
137 activities, and plant carbon allocation in the three treatment groups.

138

139 **Materials and Methods**

140 ***Native Species***

141 *Maianthemum* is a common perennial in *Alliaria*-invaded forests (e.g. Burke, 2008; Rodgers,  
142 2008) with long-lived coarse roots (Brundrett & Kendrick, 1988) and high RFS dependence  
143 (colonization rates by AMF 76-94%; Brundrett & Kendrick, 1988; Burke, 2008; and dark septate  
144 endophytes; Hough, 2008). *Maianthemum* also exhibits no change in internal RFS colonization  
145 rates (Burke, 2008) but declines in flowering frequency in the presence of *Alliaria* (Brouwer *et*  
146 *al.*, 2015).

147 ***Experimental Design***

148 Sixty-three dormant *Maianthemum* were procured (Source: Prairie Moon Nursery, Winona, MN;  
149 grown in raised beds with no pesticides), weighed and potted in a 3:1 mixture of autoclaved  
150 Fafard potting soil (Conrad Fafard Inc. Agawam, MA), Turface (Profile Products LLC, Buffalo  
151 Grove, IL) and 150 g of forest soil to ensure inoculation with RFS (pot volume:  $3.5 \times 10^{-3} \text{m}^3$ ).  
152 Plants were placed in a greenhouse with 2 layers of 65% shade cloth, where they experienced  
153 light levels approaching saturation for *Maianthemum* ( $\sim 115\text{-}117 \mu\text{mol}\cdot\text{m}^{-2}\cdot\text{s}^{-1}$  at midday; Hale *et*  
154 *al.*, 2011) and were watered every 2-4 weeks, as needed (pH of water: 6.6).

155         Plants were randomly assigned to *Alliaria*, *Hesperis*, or non-systemic fungicide  
156 treatments. For the *Alliaria* and *Hesperis* treatments, we placed 25 g of fresh leaves from either  
157 species onto the soil of each pot on 11 June 2010. Leaves were re-applied every 2 weeks through  
158 20 August. The mechanism of *Alliaria*'s allelochemical release into the soil is uncertain, but both  
159 root exudation and aboveground decomposition have been suggested (Stinson *et al.*, 2006). Our  
160 previous work suggests that plant decomposition may be a particularly important route for  
161 allelochemical delivery as our repeated assays for allelochemical in field soils across the growing  
162 season were only positive at the time of *Alliaria* leaf senescence (Cantor *et al.*, 2011). Therefore,  
163 our greenhouse treatment mimics a relevant biological process, with the ultimate goal of  
164 delivering biologically relevant concentrations of the allelochemicals into the soil of the pots. We  
165 verified that the *Alliaria* treatment effectively delivered allelochemicals by testing for the  
166 presence of sinigrin (the pre-cursor of *Alliaria*'s anti-fungal allelochemical allyl isothiocyanate,  
167 Vaughn & Berhow, 1999) using HPLC (see Supporting Information Methods). We found that  
168 the *Alliaria* treatment delivered sinigrin (Fig. S1) at concentrations consistent with those detected  
169 in the field (Cantor *et al.*, 2011). Fungicide-treated plant received monthly applications of either

170 Chipco 26019 (~50% (w/w) iprodione) or OHP 26 GT-0 (~23.3% (w/w) iprodione) at a rate of  
171 ~0.1 g active ingredient/plant.

## 172 ***Physiology***

173 Baseline physiological rates were measured prior to treatments and used as covariates in all  
174 analyses. Post-treatment assessments were made one week after the first treatment and repeated  
175 weekly for five weeks. This timeframe is appropriate because the bulk of annual carbon  
176 acquisition of forest herbs is completed within  $\leq 6$  weeks (Neufeld & Young, 2014).

177 Single leaf gas exchange measures were made using a LI-COR 6400 infrared gas  
178 analyzer (IRGA; LI-COR Inc., Lincoln, NE, USA) at 25°C, 40-50% relative humidity, 400  $\mu\text{mol}$   
179  $\text{CO}_2 \text{ mol}^{-1}$  air. Each plant was measured in random order between 0900 and 1600 h at a  
180 saturating irradiance of 600  $\mu\text{mol photons}\cdot\text{m}^{-2}\cdot\text{s}^{-1}$  (Hale *et al.*, 2011). We recorded light saturated  
181 photosynthetic rate ( $A_{\text{sat}}$ ), transpiration rate (E) and stomatal conductance to water vapor ( $g_{\text{sw}}$ )  
182 every 15 seconds for one minute.

183 To test for treatment effects on physiological traits, we used repeated measures linear  
184 mixed effects models (main effects: treatment, week of measurement (time), treatment $\times$ time  
185 interaction; covariates: air temperature, relative humidity, baseline physiological rates, initial  
186 plant wet mass). Data were transformed when necessary to meet the assumptions of normality.  
187 We selected the best covariance structure based on AIC.

## 188 ***Nutrient/carbohydrate concentration and allocation***

189 Non-experimental plants (N = 4) were harvested prior to treatments to assess early season  
190 nutrient content and allocation patterns. Experimental plants were then harvested at three post-  
191 treatment time points (N = 6-8 plants/treatment/harvest): 9 July, 6 August, and at senescence (i.e.  
192 when >40% of their leaf tissue yellowed and light saturated photosynthetic rate  $A_{\text{sat}}$  was <1.0  
193  $\mu\text{mol}\cdot\text{m}^{-2}\cdot\text{s}^{-1}$ ). We recorded the wet mass of roots at each harvest and a subset of roots from 10  
194 plants (3 from the fungicide treatment, 3 from *Alliaria*, and 4 from *Hesperis*) were stained  
195 (Brundrett *et al.*, 1984) and examined under 200 $\times$  magnification to confirm the presence of intact  
196 internal RFS structures. Leaf and the remaining root samples were dried at 70°C for 5-6 days.  
197 Rhizomes were flash-frozen in liquid nitrogen and lyophilized to a constant weight. Dried leaf,  
198 root, and rhizome samples were weighed and ground in a Wiley mill. Since bud formation in  
199 *Maianthemum* is not complete until the end of the growing season (LaFrankie, 1985), buds were  
200 counted only at the last harvest. For nitrogen and phosphorus analysis, leaf and rhizome samples

201 were sent to the Penn State Agricultural Analytical Services Laboratory (University Park, PA,  
202 USA). For carbohydrate analysis, rhizome samples were analyzed using HPLC to determine their  
203 fructan and soluble sugar content (Zuleta & Sambucetti, 2001). We report inulin and sucrose  
204 concentrations as they were the most abundant forms of carbohydrates in the rhizome throughout  
205 the season. All nutrient and carbon concentrations are expressed as a % of the sample dry mass.

206 We tested for significant treatment effects on leaf and rhizome nutrient concentration  
207 using two-way analysis of covariance (ANCOVA; model: trait = treatment + harvest date +  
208 treatment×harvest date + initial plant wet mass). For allocation to asexual reproduction, we used  
209 a one-way ANCOVA (model: bud number = treatment + initial plant wet mass).

210 Because inulin and sucrose concentrations ( $r = -0.55$ ;  $P < 0.0001$ ), rhizome and root mass  
211 ( $r_s = 0.47$ ,  $P = 0.0001$ ), and shoot mass and leaf area ( $r = 0.81$ ,  $P < 0.0001$ ) are highly correlated,  
212 we first used a multivariate analysis of covariance (MANCOVA) to test for an overall treatment  
213 effect (model: trait1, trait2 = treatment + harvest date + treatment×harvest date + initial plant wet  
214 mass). For each MANCOVA, we report F-statistics for Roy's greatest root (Scheiner, 2001). If a  
215 significant treatment effect was found by MANCOVA, we used two-way ANCOVA – and if  
216 significant, pairwise comparisons – to determine treatment effect on individual traits.

### 217 *Soil colonization by fungal hyphae*

218 Soil hyphal abundance was quantified following Baláz and Vosátka (2001). Mixed cellulose  
219 ester membranes (Millipore; pore size = 45 $\mu$ m) were inserted into each pot two weeks after the  
220 first treatment applications. We collected membranes at the first two post-treatment harvests  
221 only. [Note: Because plants senesced on different dates, membranes at final harvest are not  
222 comparable.] Membranes were stained and hyphal length determined following Cantor *et al.*  
223 (2011). We summed the hyphal length from all images/membrane to estimate hyphal  
224 length/membrane (mm), our proxy for soil hyphal abundance.

225 Robust regression provides estimates of means and standard errors for small sample sizes  
226 that are less affected by heterogeneous variance, non-normal residuals, and potential outliers  
227 (Wilcox, 1998). To test for a significant treatment effect on soil hyphal colonization, we used a  
228 robust regression (model: hyphal abundance = treatment + harvest date + treatment×harvest date)  
229 with bi-square MM-estimator using the R package *robustbase* (Rousseeuw *et al.*, 2015). Planned  
230 contrasts were performed between *Hesperis* and the two other treatments by pooling data across  
231 harvest dates (June and July) for each treatment.

232

233 **Results**234 ***Evidence for mutualism disruption***

235 We find that 100% of the *Maianthemum* plants examined are colonized by RFS (i.e. contain  
 236 intact, diagnostic fungal structures inside of the roots; Fig. S2). However, the abundance of RFS  
 237 hyphae outside the roots differs significantly across treatments ( $\chi^2_{2,35} = 7.54$ ,  $P = 0.023$ ). Linear  
 238 contrasts reveal that membrane hyphal lengths in the *Hesperis* treatment are significantly greater  
 239 than in the *Alliaria* and fungicide treatments ( $z = 2.16$ ,  $P = 0.031$ ). *Alliaria* and fungicide  
 240 treatments are not different ( $P = 0.56$ , see Table 1 for raw means).

241 ***Physiology***

242 Overall, physiological activities of *Maianthemum* from both the *Alliaria* and fungicide  
 243 treatments are significantly and similarly reduced relative to the *Hesperis* treatment (Fig. 2).  
 244 Compared to the *Hesperis* treated plants, plants in the *Alliaria* and fungicide treatments exhibit  
 245 significantly lower photosynthetic rates ( $A_{\text{sat}}$  declines by 15% and 17%, respectively;  $P = 0.031$ ,  
 246  $P = 0.019$ ; Fig. 2a), stomatal conductance ( $g_{\text{sw}}$  declines by 23% and 28%, respectively;  $P =$   
 247  $0.009$ ,  $P = 0.001$ ; Fig. 2b) and maximum transpiration rates (E declines by 18% and 21%,  
 248 respectively;  $P = 0.020$ ,  $P = 0.008$ ; Fig. 2c). A significant treatment $\times$ week effect is also evident  
 249 for each of the physiological traits (Table S1). This can be attributed to declines in physiological  
 250 function in weeks 4 and 5 for plants in the fungicide treatment. While the *Alliaria* treated plants  
 251 do not exhibit the same decline in those weeks, over time their physiological functions was  
 252 consistently lower than that of plants in the *H. matronalis* control. See Table S1 for a full  
 253 summary of results from mixed model analyses of physiological traits.

254 ***Foliar and rhizome nutrient concentrations***

255 Across all treatments, there is no significant difference in foliar or rhizome nitrogen (Fig. S3;  $P =$   
 256  $0.11$  and  $P = 0.37$ , respectively) or foliar or rhizome phosphorus concentrations ( $P = 0.23$  and  $P$   
 257  $= 0.96$ , respectively). Because all of *Maianthemum*'s leaves are pre-formed eight to twelve  
 258 months prior to emergence (LaFrankie, 1985), we did not expect treatment effects on foliar  
 259 nutrient concentrations. Furthermore, our ability to detect treatment effects on rhizome nutrient  
 260 concentration may be obscured by the ability of *Maianthemum* to resorb nitrogen and  
 261 phosphorus from its leaves at the end of the growing season (DeMars & Boerner, 1997). Indeed,  
 262 we find a significant effect of harvest date on all resource traits (foliar nitrogen  $P < 0.0001$ ; foliar



263 phosphorous  $P < 0.0001$ ; rhizome nitrogen  $P < 0.0001$ ; rhizome phosphorous  $P = 0.003$ ). Foliar  
 264 nutrient concentrations slowly decline across the growing season and rhizome nutrient  
 265 concentrations increase, indicating resorption.

### 266 **Carbon allocation**

267 Starch is not detectable in *Maianthemum* rhizomes. Instead, two major carbohydrates  
 268 predominate: inulin, a storage carbohydrate, and sucrose, a mobile sugar. Rhizome inulin and  
 269 sucrose concentrations are negatively correlated ( $r = -0.55$ ;  $P < 0.0001$ ). Treatment significantly  
 270 affects total non-structural carbon concentrations in the rhizome (Roy's greatest root = 0.32,  $P =$   
 271 0.001). Specifically, our treatments influence both inulin ( $F_{2,55} = 5.42$ ,  $P = 0.007$ ) and sucrose  
 272 ( $F_{2,55} = 7.52$ ,  $P = 0.001$ ) concentrations. *Maianthemum* in the *Alliaria* treatment store 17% less  
 273 inulin relative to the *Hesperis* control (Fig. 3a), but have significantly greater sucrose  
 274 concentration (Fig. 3b). Fungicide treated plants do not show reduced rhizome inulin (Fig. 3a)  
 275 and their sucrose content is intermediate between the *Hesperis* and *Alliaria* treatments (Fig. 3b).

276 Treatment also significantly alters allocation to belowground structures (Roy's greatest  
 277 root = 0.43,  $P = 0.0001$ ), which is driven by changes in root mass ( $F_{2,54} = 9.40$ ,  $P = 0.001$ ), not  
 278 rhizome mass ( $F_{2,54} = 0.22$ ,  $P = 0.80$ ). Plants from both the *Alliaria* and fungicide treatments  
 279 produce ~25% less root mass than the *Hesperis* controls (Fig. 3c). Our analyses show no  
 280 significant treatment effect on shoot mass ( $F_{2,54} = 0.66$ ,  $P = 0.52$ ) or leaf area ( $F_{2,52} = 0.04$ ,  $P =$   
 281 0.96). Due to leaf preformation in *Maianthemum* (LaFrankie, 1985), we did not expect responses  
 282 in aboveground tissues to our experimental treatments.

283 Treatment has a marginally significant effect on asexual reproduction ( $F_{2,15} = 3.01$ ,  $P =$   
 284 0.079). On average, *Maianthemum* in the *Hesperis* control produce two more asexual buds  
 285 compared to the *Alliaria* treatment; fungicide treated plants produce intermediate bud numbers  
 286 (Fig. 3d).

287

### 288 **Discussion**

289 Our results mechanistically link allelopathic mutualism disruption of root fungal symbionts  
 290 (RFS) to plant fitness declines and provide clear support for the predictions of our model (Fig.  
 291 1).

292 Prediction 1. Soil hyphal length was reduced in the *Alliaria* and fungicide treatments  
 293 relative to the *Hesperis* control (Table 1). These results corroborate the findings of Hale *et al.*

294 (2011) and Cantor *et al.* (2011), which showed that a similar *Alliaria* treatment reduced soil  
295 respiration rates around *Maianthemum* and that soil hyphal abundance was lower in *Alliaria*  
296 invaded field sites, respectively.

297 Prediction 2. *Alliaria* treatment significantly diminishes plant physiological activities  
298 (Fig. 2). In fact, *Alliaria* treated plants displayed nearly identical patterns of physiological  
299 suppression to those observed in the non-systemic fungicide treatment. The observed declines in  
300 *Maianthemum* physiological activity following loss of the soil RFS network could be the result  
301 of declines in RFS-delivered benefits, including plant nutrition, water availability, and/or  
302 changes in sink strength (Hale *et al.*, 2011). However, foliar and rhizome nitrogen and  
303 phosphorus concentrations did not differ among treatments (Fig. S3). Thus, nutrient limitation is  
304 not responsible for the reductions in *Maianthemum* physiological activities, but could become  
305 important if the experiment continued for multiple growing seasons. The significant  
306 photosynthetic and stomatal conductance declines in *Maianthemum* seen here point to reductions  
307 in water availability (Augé, 2001) and are consistent with our previous field study (Hale *et al.*,  
308 2011).

309 Prediction 3. RFS mutualism disruption manifested as changes in carbon allocation in  
310 *Maianthemum* plants. Significantly lower inulin concentrations, root mass and asexual  
311 reproduction in the *Alliaria* treatment are consistent signatures of carbon limitation (Fig. 3a,c,d).  
312 Reduced root mass with disrupted RFS is a somewhat surprising result, as other forest herbs have  
313 shown the opposite outcome (i.e. increased root growth) when RFS are disrupted (e.g. Lapointe  
314 & Molard, 1997). However, fungicides have been shown to reduce root length in some  
315 mycorrhizal plants (e.g. Sukarno *et al.*, 1993). The reduced root mass seen in our study further  
316 supports the idea that *Alliaria* and fungicide treatments cause carbon limitation in *Maianthemum*.

317 Comparison of leaf longevity and carbon storage across treatments provides insight into  
318 the carbon stress observed here. Plants in the *Alliaria* treatment senesced, on average, seven days  
319 later than those in the *Hesperis* treatment (data not shown). This observation is consistent with  
320 other studies of forest herbs that show increased leaf longevity in response to photosynthetic rate  
321 suppression and carbon stress (Muraoka *et al.*, 1997; Tomimatsu & Yoshimichi, 2008). Further,  
322 in just a single growing season, inulin concentrations in rhizomes of *Alliaria* treated plants were  
323 17% lower (Fig. 3a) relative to the *Hesperis* controls. This large magnitude decline mirrors the

324 observed decline in physiological traits (Fig. 2) and should affect plant fitness in the long run if  
325 RFS disruption is chronic as in heavily invaded forests.

326 Interestingly, in addition to reducing storage carbohydrates, the *Alliaria* treatment  
327 significantly increased mobile sucrose concentrations (Fig. 3b). We offer two potential, non-  
328 exclusive mechanisms for this contrasting result. First, sucrose may be in transit to the internal  
329 root symbionts for re-establishment of the external soil hyphal network. Sucrose is the precursor  
330 to the hexose sugars that are exchanged with AMF arbuscules in roots (Parniske, 2008).  
331 Repeated inhibition of the external hyphal network as it is regrown and disrupted by *Alliaria*  
332 could create an open-ended carbon sink further reducing carbon storage and increasing carbon  
333 stress. While such an open-ended sink is expected to up-regulate photosynthesis and  
334 carbohydrate production, we observe a decrease in photosynthetic rates following disruption of  
335 the soil hyphal network in the *Alliaria* treatment. Our results point to limited water availability  
336 constraining *Maianthemum*'s photosynthetic capacity and suggest an inability of *Alliaria* treated  
337 plants to respond to the increased sink strength. Second, the observed sucrose concentration  
338 could result from changes in gene expression. Sugar produced via photosynthesis acts as both a  
339 substrate for construction and as a signal that modulates gene expression in plants (Smeekens,  
340 2000; Rolland *et al.*, 2006). In general, when sugars from photosynthesis are scarce, genes  
341 controlling storage and growth are repressed, while genes controlling photosynthesis, nutrient  
342 mobilization, and export are enhanced (Rolland *et al.*, 2006). It is possible that the low carbon  
343 fixation rates we observed in *Alliaria* and fungicide-treated *Maianthemum* could repress genes  
344 directing carbon storage and root growth. Similarly, increased sucrose availability in the rhizome  
345 (Fig. 3b) could be explained by enhanced expression of export genes. The possibility for  
346 invaders to alter gene expression in native plants by mutualism disruption is an intriguing  
347 possibility that, to our knowledge, is completely unexplored. This avenue of research warrants  
348 future study, as it could lead to a deeper mechanistic understanding of the processes underlying  
349 the impacts of mutualism disruption.

350 Finally, we note that the results of our study could be due to direct, phytotoxic effects of  
351 *Alliaria* on *Maianthemum*. However, *Alliaria*'s effects were significantly greater than the  
352 fungicide treatment for only a single trait (inulin, Fig. 3a) and slightly, but not significantly,  
353 greater than the fungicide treatment for only two traits measured (sucrose, Fig. 3b, and asexual  
354 reproduction, Fig. 3d). In all other comparisons, *Alliaria* and the non-systemic fungicide

355 treatments were statistically indistinguishable and responded in a similar fashion. While direct  
356 allelopathic effects may explain slight differences between these treatments, the tandem declines  
357 in soil hyphal abundance in both *Alliaria* and fungicide treatments point to loss of the mutualism  
358 function as the key effect of the allelochemicals.

359 Many herbaceous species in northeastern North American temperate hardwood forests  
360 possess similar life histories (Bierzychudek, 1982; Whigham, 2004; Jolls and Whigham, 2014),  
361 growth patterns, and root traits (i.e. lack fine roots, exhibit 2–4 orders of root branching, and  
362 sustain long-lived fungal structures inside their roots (Fig. S2; Brundrett & Kendrick, 1988)). We  
363 suggest that many species will be highly susceptible to RFS disruption following *Alliaria*  
364 invasion (e.g. species in the genera *Maianthemum*, *Polygonatum*, *Trillium*, *Sanguinaria*,  
365 *Arisaema*, *Erythronium*, *Asarum*, *Allium*, and others). If the allocation changes observed by  
366 *Maianthemum* in the *Alliaria* treatment are multiplied over many growing seasons in the field,  
367 the ability of native plants to compete and tolerate additional environmental stresses could be  
368 compromised. Declines in hyphal length, as observed here and under field conditions (Cantor *et*  
369 *al.*, 2011), and our observed declines in root mass (Fig. 3c) indicate that on *Alliaria* invaded  
370 sites, plants that rely on belowground mutualisms will be less able to forage and compete for soil  
371 resources and water. Additionally, the observed declines in inulin storage (Fig. 3a) suggest that  
372 plants on invaded sites may be unable to store sufficient quantities of carbohydrates and tolerate  
373 additional stresses, such as high herbivore pressure. For example, white-tailed deer (*Odocoileus*  
374 *virginianus*) can consume up to 100% of flowering plants in *Maianthemum* populations annually  
375 (Ruhren & Handel, 2003; Kraft *et al.*, 2004). High concentrations of storage carbohydrates are  
376 critical for herbaceous perennials to survive such repeated herbivory episodes (Lapointe *et al.*,  
377 2010). Lastly, clonal reproduction, which is linked to bud number, is likely important in  
378 maintaining population growth in many forest herb species (Honnay *et al.* 2005), which exhibit  
379 slow growth and low germination rates (Bierzychudek, 1982; Whigham, 2004; Jolls and  
380 Whigham, 2014). Overall, mutualism disruption by *Alliaria* could lead to population decline and  
381 may generally explain the reduced abundance of native plants on invaded sites (Rodgers, 2008;  
382 Stinson *et al.*, 2007).

383 In summary, we have shown that allelochemicals from a widespread invader reduce  
384 fungal hyphal growth in the soil and may alter the carbon fixation capacity and allocation  
385 patterns in native plants. Reduced water absorption leads to persistent declines in plant

386 photosynthetic rates, and this source limitation coupled with an increased RFS sink demand  
387 culminates in reduced allocation to functionally important traits like storage, growth, and  
388 reproduction. Ultimately, these impacts could facilitate further invasion of the ecosystem by  
389 reducing the competitive ability of mutualism dependent native plants and compromise their  
390 ability to respond to other environmental stressors. Many exotic and invasive species produce  
391 novel chemicals that are toxic to root fungal symbionts (e.g. Bainard *et al.*, 2009; Meinhardt &  
392 Gehring, 2012), yet the majority remain untested for allelopathic potential. Allelopathic  
393 mutualism disruption may be an important, but under-recognized, mechanism contributing to  
394 ecosystem invasion. As the number of invasive species and their corresponding impacts continue  
395 to increase, studies of invasion employing the powerful tools of eco-physiology may be critical  
396 in revealing invasion mechanisms and informing management decisions.

### 397 **Acknowledgements**

398 We thank NSF DEB 0958676, Phipps Conservatory Botany-in-Action, and U. Pittsburgh's  
399 Andrew K. Mellon pre-doctoral fellowship for funding. We thank M. Khalf for rhizome inulin  
400 content measurement, M.B. Traw for help with sinigrin isolation, N.L. Brouwer for help with  
401 data analysis, and E. York and the Kalisz lab for assistance in the greenhouse. We also thank  
402 N.L. Brouwer and M. Heberling for helpful comments on drafts of this manuscript.

### 403 **References**

- 404 **Aslan CE, Zavaleta ES, Tershy B, Croll D. 2013.** Mutualism disruption threatens global plant  
405 biodiversity: a systematic review. *PLoS ONE* 8: e66993.
- 406 **Augé RM. 2001.** Water relations, drought, and vesicular-arbuscular mycorrhizal symbiosis.  
407 *Mycorrhiza* 11: 3-42.
- 408 **Bainard LD, Brown PD, Upadhyaya MK. 2009.** Inhibitory effect of tall hedge mustard  
409 (*Sisymbrium loeselii*) allelochemicals on rangeland plants and arbuscular mycorrhizal  
410 fungi. *Weed Science* 57: 386-393.
- 411 **Baláz M, Vosátka M. 2001.** A novel inserted membrane technique for studies of mycorrhizal  
412 extraradical mycelium. *Mycorrhiza* 11: 291-296.
- 413 **Bierzychudek P. 1982.** Life histories and demography of shade-tolerant temperate forest herbs:  
414 A review. *New Phytologist* 90: 757-776.

- 415 **Brouwer NL, Hale AN, Kalisz S. 2015.** Mutualism-disrupting allelopathic invader drives  
416 carbon stress and vital rate decline in a forest perennial herb. *AoB Plants*. doi:  
417 10.1093/aobpla/plv014.
- 418 **Brundrett MC, Piché Y, Peterson RL. 1984.** A new method for observing the morphology of  
419 vesicular-arbuscular mycorrhizae. *Canadian Journal of Botany* 62: 2128-2134.
- 420 **Brundrett M, Kendrick B. 1988.** The mycorrhizal status, root anatomy, and phenology of  
421 plants in a sugar maple forest. *Canadian Journal of Botany* 66: 1153-1173.
- 422 **Brundrett M, Kendrick B. 1990.** The roots and mycorrhizas of herbaceous woodland plants. I.  
423 Quantitative aspects of morphology. *New Phytologist* 114: 457-468.
- 424 **Burke DJ. 2008.** Effects of *Alliaria petiolata* (garlic mustard; Brassicaceae) on mycorrhizal  
425 colonization and community structure in three herbaceous plants in a mixed deciduous  
426 forest. *American Journal of Botany* 95: 1416-1425.
- 427 **Callaway RM, Ridenour WM. 2004.** Novel weapons: Invasive success and the evolution of  
428 increased competitive ability. *Frontiers in Ecology and the Environment* 2: 436-443.
- 429 **Callaway RM, Cipollini D, Barto K, Thelen GC, Hallett SG, Prati D, Stinson KA,**  
430 **Klironomos J. 2008.** Novel weapons: Invasive plant suppresses fungal mutualists in  
431 America but not in its native Europe. *Ecology* 89: 1043-1055.
- 432 **Cantor A, Hale AN, Aaron J, Traw MB, Kalisz S. 2011.** Low allelochemical concentrations  
433 detected in garlic mustard-invaded forest soils inhibit fungal growth and AMF spore  
434 germination. *Biological Invasions* 13: 3015-3025.
- 435 **Cipollini D, Rigsby CM, Barto EK. 2012.** Microbes as targets and mediators of allelopathy in  
436 plants. *Journal of Chemical Ecology* 38: 714-727.
- 437 **DeMars BG, Boerner REJ. 1995.** Arbuscular mycorrhizal development in three crucifers.  
438 *Mycorrhiza* 5: 405-408.
- 439 **DeMars BG, Boerner REJ. 1997.** Foliar phosphorus and nitrogen resorption in three woodland  
440 herbs of contrasting phenology. *Castanea* 62: 43-54.
- 441 **Hale AN, Tonsor SJ, Kalisz S. 2011.** Testing the mutualism disruption hypothesis:  
442 physiological mechanisms for invasion of intact perennial plant communities. *Ecosphere*  
443 2: 110.

- 444 **Honnay O, Jacquemyn H, Bossuyt B, Hermy M.** 2005. Forest fragmentation effects on patch  
445 occupancy and population viability of herbaceous plant species. *New Phytologist* 166:  
446 723-736.
- 447 **Hough M.** 2008. *Possible limiting agents to the early establishment and growth of understory*  
448 *herbs in post-agricultural forests in central New York.* Master thesis, State University of  
449 New York, Syracuse, NY, USA.
- 450 **Johnson NC, Wilson GWT, Wilson JA, Miller RM, Bowker MA.** 2015. Mycorrhizal  
451 phenotypes and the law of the minimum. *New Phytologist* 205: 1473-1484.
- 452 **Jolls CL, Whigham D.** 2014. Populations of and threats to rare plants of the herb layer: still  
453 more challenges and opportunities for conservation biologists. In: Gilliam FS, ed. *The*  
454 *herbaceous layer in forests of eastern North America.* New York, NY, USA: Oxford  
455 University Press, 134-164.
- 456 **Jumpponen A.** 2001. Dark septate endophytes – are they mycorrhizal? *Mycorrhiza* 11: 207-211.
- 457 **Kiers ET, Palmer TM, Ives AR, Bruno JF, Bronstein JL.** 2010. Mutualisms in a changing  
458 world: an evolutionary perspective. *Ecology Letters* 13, 1459-1474.
- 459 **Kraft SL, Crow TR, Buckley DS, Nauertz EA, Zasada JC.** 2004. Effects of harvesting and  
460 deer browsing on attributes of understory plants in northern hardwood forests, Upper  
461 Michigan, USA. *Forest Ecology and Management* 199: 219-230.
- 462 **LaFrankie Jr. JV.** 1985. Morphology, growth, and vasculature of the sympodial rhizome of  
463 *Smilacina racemosa* (Liliaceae). *Botanical Gazette* 146: 534-544.
- 464 **Lankau RA, Nuzzo V, Spyreas G, Davis AS.** 2009. Evolutionary limits ameliorate the negative  
465 impact of an invasive plant. *Proceedings of the National Academy of Sciences USA* 106:  
466 15362-15367.
- 467 **Lankau RA.** 2010. Resistance and recovery of soil microbial communities in the face of *Alliaria*  
468 *petiolata* invasions. *New Phytologist* 189: 536-548.
- 469 **Lapointe L, Molard J.** 1997. Costs and benefits of mycorrhizal infection in a spring ephemeral,  
470 *Erythronium americanum.* *New Phytologist* 135: 491–500.
- 471 **Lapointe L, Bussièrès J, Crête M, Ouellet J-P.** 2010. Impact of growth form and carbohydrate  
472 reserves on tolerance to simulated deer herbivory and subsequent recovery in Liliaceae.  
473 *American Journal of Botany* 97: 913-924.
- 474 **Marschner H, Dell B.** 1994. Nutrient uptake in mycorrhizal symbiosis. *Plant Soil* 159: 89-102.

- 475 **Meinhardt KA, Gehring CA. 2012.** Disrupting mycorrhizal mutualisms: a potential mechanism  
476 by which exotic tamarisk outcompetes native cottonwoods. *Ecology* 22: 532-549.
- 477 **Memmot J, Craze PG, Waser NM, Price MV. 2007.** Global warming and the disruption of  
478 plant-pollinator interactions. *Ecology Letters* 10: 710-717.
- 479 **Millennium Ecosystem Assessment. 2005.** Corvalan C, Hales S, McMichael A, core writing  
480 team. *Ecosystems and Human Well-Being: Synthesis*. Washington D.C., USA: Island  
481 Press.
- 482 **Muraoka H, Tang Y, Koisumi H, Washitani I. 1997.** Combined effects of light and water  
483 availability on photosynthesis and growth of *Arisaema heterophyllum* in the forest  
484 understory and an open site. *Oecologia* 112: 26-34.
- 485 **Neufeld HS, Young DR. 2014.** Ecophysiology of the herbaceous layer in temperate deciduous  
486 forests. In: Gilliam FS, ed. *The herbaceous layer in forests of eastern North America*.  
487 New York, NY, USA: Oxford University Press, 35-92.
- 488 **Newsham KK. 2011.** A meta-analysis of plant responses to dark septate root endophytes. *New*  
489 *Phytologist* 190: 783-793.
- 490 **Parniske M. 2008.** Arbuscular mycorrhiza: the mother of plant root endosymbiosis. *Nature*  
491 *Reviews Microbiology* 6:763-775.
- 492 **Petit AN, Fontaine F, Vatsa P, Clément C, Vaillant-Gaveau N. 2012.** Fungicide impacts on  
493 photosynthesis in crop plants. *Photosynthesis Research* 111: 315-326.
- 494 **Rousseeuw P, Croux C, Todorov V, Ruckstuhl A, Salibian-Barrera M, Verbeke T, Koller**  
495 **M, Maechler M. 2015.** *robustbase: Basic Robust Statistics*. [R package version 0.92-4.]  
496 URL <http://CRAN.R-project.org/package=robustbase>.
- 497 **Roberts KJ, Anderson RC. 2001.** Effect of garlic mustard [*Alliaria petiolata* (Beib. Cavara and  
498 Grande)] extracts on plants and arbuscular mycorrhizal (AM) fungi. *American Midland*  
499 *Naturalist* 146: 146-152.
- 500 **Rodgers V, Stinson KA, Finzi AC. 2008.** Ready or not, garlic mustard is moving in: *Alliaria*  
501 *petiolata* as a member of eastern North American forests. *BioScience* 58: 426-436.
- 502 **Rodgers VL. 2008.** *Impacts of Alliaria petiolata (garlic mustard) invasion on plant diversity*  
503 *and soil nutrient cycling in northern hardwood-conifer forests*. PhD Thesis, Boston  
504 University, Boston, MA, USA.



- 505 **Rolland F, Baena-Gonzalez E, Sheen J. 2006.** Sugar sensing and signaling in plants:  
506 Conserved and novel mechanisms. *Annual Review of Plant Biology* 57: 675-709.
- 507 **Ruhren S, Handel SN. 2003.** Herbivory constrains survival, reproduction and mutualisms when  
508 restoring nine temperate forest herbs. *Journal of the Torrey Botanical Society* 130: 34-42.
- 509 **Scheiner SM. 2001.** MANOVA: multiple response variables and multispecies interactions. In:  
510 Scheiner SM, Gurevitch J, eds. *Design and analysis of ecological experiments*. New  
511 York, NY, USA: Oxford University Press, 99–115.
- 512 **Smeekens S. 2000.** Sugar-induced signal transduction in plants. *Annual Review of Plant*  
513 *Physiology and Plant Molecular Biology* 51: 49-81.
- 514 **Staddon PL, Thompson K, Jakobsen I, Grime P, Askew AP, Fitter AH. 2003.** Mycorrhizal  
515 fungal abundance is affected by long-term climatic manipulations in the field. *Global*  
516 *Change Biology* 9: 186-194.
- 517 **Stinson KA, Campbell SA, Powell JR, Wolfe BE, Callaway RM, Thelen GC, Hallett SG,**  
518 **Prati D, Klironomos J. 2006.** Invasive plant suppresses the growth of native tree seedlings  
519 by disrupting belowground mutualisms. *PLOS Biology* 4: 727-731.
- 520 **Stinson KA, Kaufman S, Durbin L, Lowenstein F. 2007.** Impacts of *Alliaria* invasion on a  
521 forest understory community. *Northeastern Naturalist* 14: 73-88.
- 522 **Sukarno N, Smith SE, Scott ES. 1993.** The effect of fungicides on vesicular-arbuscular  
523 mycorrhizal fungi and plant growth. *New Phytologist* 25: 139-147.
- 524 **Tomimatsu H, Yoshimichi H. 2008.** Effect of soil moisture on leaf ecophysiology of  
525 *Parasenecio yatabei*, a summer-green herb in a cool–temperate forest understory in  
526 Japan. *Journal of Plant Research* 121: 43-53.
- 527 **Traveset A, Richardson DM. 2014.** Mutualistic interactions and biological invasions. *Annual*  
528 *Review of Ecology, Evolution, and Systematics* 45: 89-113.
- 529 **Tylianakis JM, Didham RK, Bascompte J, Wardle DA. 2008.** Global change and species  
530 interactions in terrestrial ecosystems. *Ecology Letters* 11: 1351-1363.
- 531 **Vaughn SF, Berhow MA. 1999.** Allelochemicals isolated from tissues of the invasive weed  
532 garlic mustard (*Alliaria petiolata*). *Journal of Chemical Ecology* 25: 2495-2504.
- 533 **Whigham DF. 2004.** Ecology of woodland herbs in temperate deciduous forests. *Annual Review*  
534 *of Ecology, Evolution, and Systematics*. 35: 583-621.

535 **Wilcox RR. 1998.** How many discoveries have been lost by ignoring modern statistical methods.  
 536 *American Psychologist* 53: 300-314.

537 **Wright IJ, Reich PB, Westoby M, Ackerly DD, Baruch Z, Bongers F, Cavender-Bares J,**  
 538 **Chapin T, Cornelissen JHC, Diemer M, et al. 2004.** The worldwide leaf economics  
 539 spectrum. *Nature* 428: 821-827.

540 **Zuleta A, Sambucetti ME. 2001.** Inulin determination for food labeling. *Journal of Agricultural*  
 541 *and Food Chemistry* 49: 4570-4572.

542

543 **Table 1.** Raw means and log-transformed means from robust regression for soil fungal hyphal  
 544 length/membrane (mm) in each treatment at June and July harvests. Each treatment has 6-7  
 545 replicates comprised of pots containing one *Maianthemum racemosum* individual and one  
 546 membrane. Because plants were harvested on different dates for the last harvest in August,  
 547 hyphal data were not comparable.

548

Harvest	Treatment	N	Raw data		Log-transformed data from robust regression	
			Mean	SE	Mean	SE
June	<i>Hesperis</i>	7	2.43	0.60	0.70	0.25
	Fungicide	7	3.59	1.23	1.06	0.51
	<i>Alliaria</i>	7	1.71	0.83	0.01	0.26
July	<i>Hesperis</i>	6	19.51	4.46	2.82	0.31
	Fungicide	7	6.38	1.37	1.70	0.26
	<i>Alliaria</i>	7	12.11	2.72	2.34	0.30

549

550

551

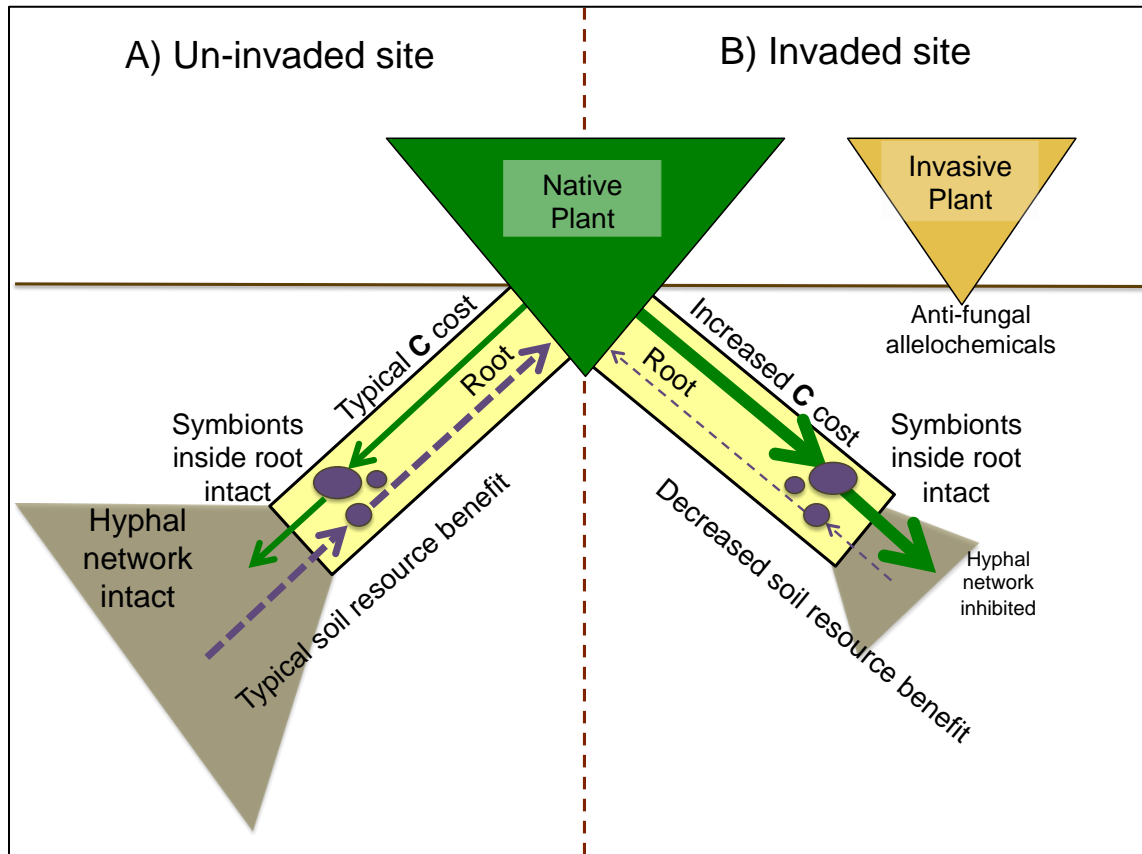
552

553

554

555

556 **Figure 1.** Hypothetical model depicting the costs and benefits to the plant of supporting root  
 557 fungal symbionts in A) the absence of *Alliaria* and B) in the presence of *Alliaria*. Solid arrows  
 558 represent carbon costs; dashed arrows represent resource benefits. Thickness of arrows  
 559 represents the magnitude of the costs and benefits.

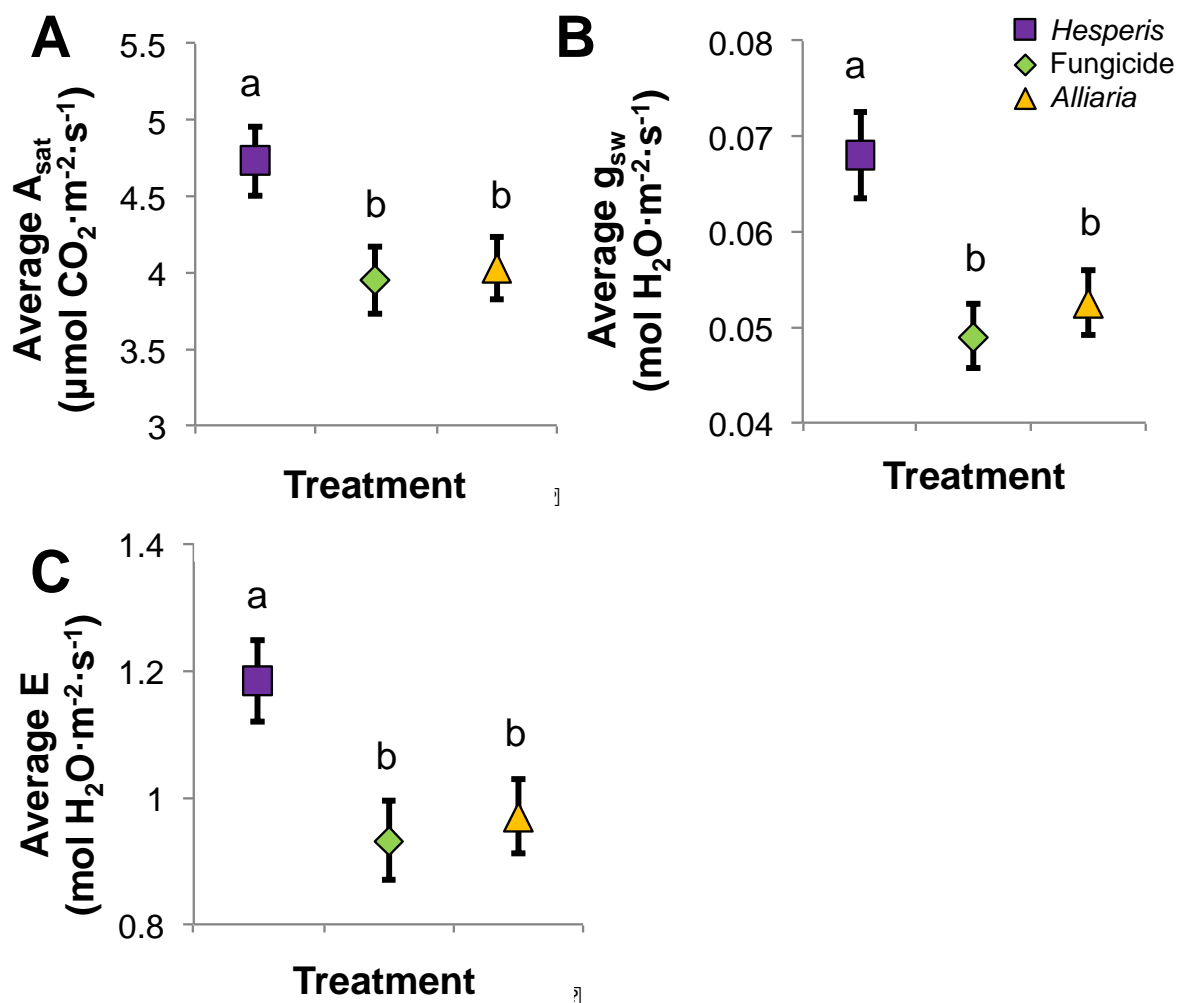


560  
 561  
 562  
 563  
 564  
 565  
 566  
 567  
 568  
 569  
 570

571 **Figure 2.** Five-week average physiological performance of *Maianthemum* plants exposed to  
572 *Hesperis*, fungicide or *Alliaria* treatments. A) Light saturated photosynthetic rate ( $A_{\text{sat}}$ ), B)  
573 stomatal conductance ( $g_{\text{sw}}$ ), and C) transpiration rate (E). Different letters are significantly  
574 different ( $P < 0.05$ ; values are least squares means  $\pm 1$  standard error).

575

576

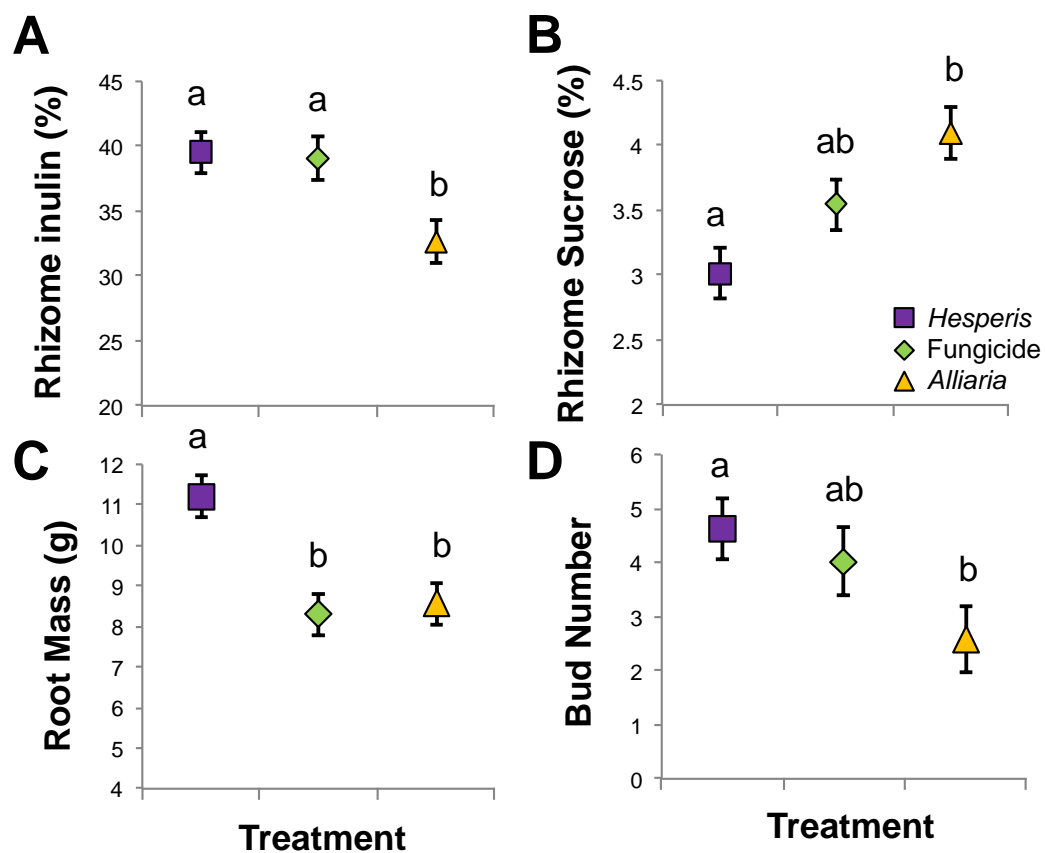


577

578

579

580 **Figure 3.** Average carbon allocation in *Maianthemum* plants exposed to *Hesperis*, fungicide or  
581 *Alliaria* treatments. Carbon can be stored as inulin (A), maintained in a mobile form as sucrose  
582 (B), allocated to root mass (C), or allocated to asexual reproduction (D). Asexual reproduction  
583 was measured as new bud production along the rhizome at the last harvest. As determined by  
584 pairwise comparisons, treatments with different lowercase letters are significantly different from  
585 each other ( $P < 0.05$ ; values are least squares means  $\pm 1$  standard error).



586

587