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Life Cycle Costs and Carbon Emissions of Onshore Wind Power

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Summary

There is a significant diversity of views on the life cycle levelised costs and carbon emissions of energy technologies, including onshore wind. ClimateXChange has commissioned a briefing paper to help Scottish policy makers and other interested parties better understand these perspectives, the uncertainties associated with them, and the differing underpinning assumptions. In particular, this paper:

- Identifies the varied academic and wider perspectives on the life cycle costs and emissions of onshore wind technologies and associated infrastructure;
- Synthesises the existing evidence and assumptions used to support these perspectives;
- Identifies variations in the evidence and assumptions;
- Identifies areas of consensus and any outliers.

1 Introduction

Energy policy is right at the top of the political agenda following concerns over the cost of living, recent price rises by the main utilities and the cause of these price rises. There is very vocal argument about the impacts of 'green obligations' and subsidies for renewable energy sources, particularly as wind energy production reaches record levels in Scotland and across the United Kingdom. Additionally, reports of low generation margins and risks to security of supply are adding to the mix.

Carbon emissions, affordable energy and security of supply are strands of the energy policy 'trilemma'. The three aspects are very heavily interdependent and, consequently, there is substantial scope for disagreement - particularly where one aspect is focussed on while others are ignored. This makes rational policymaking challenging.

Understanding the economics of wind energy is vitally important to ensure a rational discussion about the role of wind power within the energy mix. The challenge is that 'cost' means different things to different people, with often conflicting views apparently supported by 'evidence'. In part this is due to confusion about current and likely future costs of generation, what might be included or excluded and the characteristics of wind relative to other generation types. Additionally, there is conflation of 'costs', 'prices' within the power markets and 'subsidies'.

Another key issue is the debate over whether onshore wind farms actually achieve a net carbon emissions saving over their lifetime. The carbon emissions reduction of wind power cannot simply be estimated as equal to the

ClimateXChange is Scotland's Centre of Expertise on Climate Change, supporting the Scottish Government's policy development on climate change mitigation, adaptation and the transition to a low carbon economy. The centre delivers objective, independent, integrated and authoritative evidence in response to clearly specified policy questions.

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carbon emissions of conventional coal- or gas-fired generation: firstly, wind power generation is not zero carbon, as greenhouse gases are emitted during installation, maintenance and decommissioning; secondly, wind power will not replace all forms of conventional generation equally, so the true carbon emissions displacement will depend upon a combination of factors – including the types of power generation being replaced, any decrease in efficiency of conventional plant operating at part load, and the impact of any increase in frequency of start-up and shut-down of conventional plant. There may also be longer-term impacts associated with the installation of new conventional plant to back up an increase in installed wind capacity. Many of the existing publications examining the carbon emissions of onshore wind concentrate on either one or other of the above issues, with positive reports often focussing on the relatively small life cycle emissions of wind power in comparison to fossil-fuelled generation, and negative reports highlighting the uncertainty of calculating the true emissions displacement.

This briefing paper critically examines both of these issues in order to provide guidance on the most realistic estimates of life cycle costs and carbon emissions savings for onshore wind power generation in Scotland and the UK. The specific issues addressed in this paper are:

Life Cycle Costs – The cost of producing energy from onshore wind compared to conventional sources.

Life Cycle Carbon Emissions – The overall carbon emissions associated with onshore wind over the life cycle of the plant: examining the existing evidence of these life cycle emissions, and comparing them with other technologies.

System Costs and Emissions – The impact on life cycle cost and emissions of the technology required to complement onshore wind power as a mainstream energy source, including the costs and emissions associated with the installation and operation of conventional generation to cope with the variable output of wind power, and an understanding of the emissions displacement of onshore wind for realistic estimates of emissions savings.

1.1 Glossary

This list is intended as a quick reference to clarify specific terms used in this paper.

Carbon emissions	Greenhouse gas (GHG) emissions
Carbon footprint	Life cycle carbon/greenhouse gas emissions
Carbon payback period	The time for displaced emissions to equal the life cycle carbon emissions
Discount rate	A value which determines the future value of costs in present value terms
Displaced emissions	A measure of the greenhouse gases not emitted from conventional generators due to power from wind
Efficiency penalty	The decrease in efficiency of conventional generators when operating at part load
Emissions intensity	The greenhouse gas emissions per unit of output energy
External cost	An impact that has economic value but is not captured by traditional financial cost measures
Lifetime emissions savings	Net reduction in greenhouse gas emissions over the life time of the wind farm
Levelised cost of energy	Measure of life cycle costs expressed per unit of electricity generated

Marginal generation	The type of power generation operating on the margin
System costs	Costs associated with the operation and planning of the wider electricity system

1.2 Wind farm life cycle

Costs and carbon emissions arise during every stage of the life cycle of a wind farm, illustrated by Figure 1. The elements of each of these stages considered in this report are further explained below.

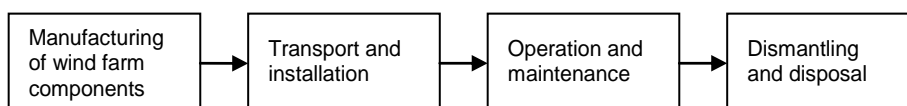


Figure 1 - Life cycle of a wind farm

Manufacturing of wind farm components

The first stage includes the extraction and production of raw materials, and the manufacture of wind farm components. Figure 2 illustrates which components are typically included within the system boundary for estimation of the costs and emissions of an onshore wind farm, with the principal components of the turbine itself illustrated in Figure 3, and described in greater detail in Appendix 2. The wind turbine assembly varies little for onshore or offshore installations, with the main difference being the tower height – typically 80 m offshore and 100 onshore (Vestas, 2006b) – instead the principal differences between onshore and offshore farms are in the design of the foundations, groundworks and transmission equipment.

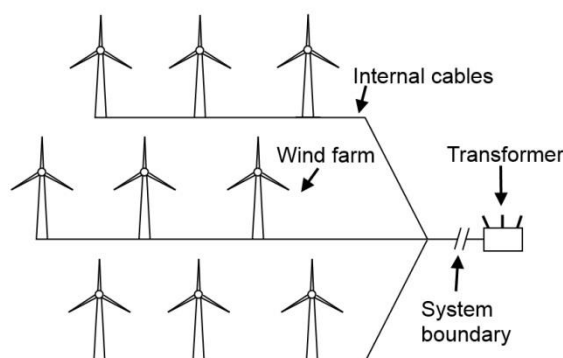


Figure 2 - System boundary for an onshore wind farm (after (Vestas, 2006b))

Wind turbine designs, however, do vary significantly from manufacturer to manufacturer, principally in size and choice of materials - this will affect the cost and emissions of each design; for example, the precise design of composite materials used in the nacelle and hub may vary, while cables, electrical equipment, hydraulic equipment and foundations also use different quantities of materials depending upon the location and design of the farm itself. Furthermore, different manufacturers in different locations may use different proportions of recycled raw materials, which will also affect both the costs and emissions at this stage.

Transport and installation

The second life cycle stage is transport and installation, which stage for an onshore wind farm includes the construction of access roads and foundations, and the transport, installation and commissioning of the wind turbines. Costs and emissions arise from all of these processes, and there may also be an impact on carbon

emissions due to change in land-use at the installation site. Land-use change is of particular consideration for onshore wind farms in Scotland, where a high proportion are built on peatlands; peat plays a significant role in the carbon cycle, absorbing and releasing carbon dioxide and methane according to its moisture content, and it is likely that the ground works associated with a wind farm will cause a net emission of carbon from the soil (Nayak et al., 2008).

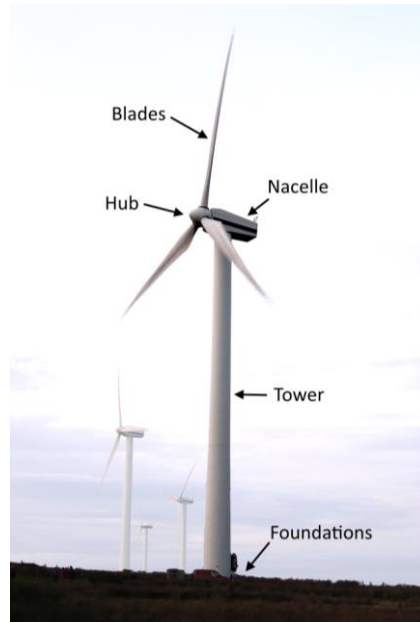


Figure 3 - Wind turbine

Operations and maintenance

Costs and carbon emissions that arise during the third life cycle stage of a wind farm are largely due to maintenance activities, such as inspection visits (including transport of equipment and people to and from the site), regular changes of oil and other lubricants, maintenance of paintwork and component renovation or replacement (including the impacts associated with the materials and manufacture of these components, and associated disposal of any operational waste) (Vattenfall, 2013; Vestas, 2006b). There are some costs and emissions associated with the operation of wind turbines, due to energy consumption to operate the yaw system, the brakes, and power up the generator (Guezuraga et al., 2012), but this is usually subtracted from the estimated energy production and is, therefore, included only as a loss of earnings or reduction in total output.

Dismantling and disposal

The final stage in the life cycle of a wind farm is decommissioning, which includes all dismantling, transport, disposal and recycling (Vestas, 2006a; Vestas, 2006b). There are costs and carbon emissions associated with all of these processes, although recycling can also result in a carbon saving or financial revenue.

Key Messages

- There is confusion about current and likely future costs of generation, what might be included or excluded in estimates and the characteristics of wind relative to other generation types.
- There is conflation of 'costs', 'prices' within the power markets and 'subsidies'.
- The carbon emissions reduction of wind power is complex, as life cycle emissions of wind are non-zero and true carbon emissions displacement will depend upon the operation of the whole grid
- Variations in cost and carbon emissions estimates are affected by assumptions made in the calculation itself and also differences in wind turbine designs, manufacturing and installation locations, maintenance and disposal.

2 Life Cycle Costs

2.1 Expenditure and levelised cost

There are a wide variety of costs associated with electricity generation technologies but these can be grouped into three main components:

- Capital costs (CAPEX): the fixed costs of construction including manufacturing, installation and transport;
- Operation and maintenance costs (OPEX): the annual fixed costs associated with running the generator (e.g. maintenance) as well as those that vary with production (e.g. fuel);
- Decommissioning: the cost of taking the plant out of commission, dismantling and remediation.

Technologies may be compared on the basis of any of these costs: the capital cost per unit of installed capacity is a common measure of how expensive a given technology is to build; operational costs tend to distinguish between technologies that have high operational costs (particularly those using fossil fuels), and those with low operational costs (which would include most renewable and nuclear technologies); some technologies have significant costs associated with decommissioning (e.g. nuclear) and others are fairly limited. While it is possible to compare technologies by looking at individual cost categories this tends to distort the picture as it is not automatically the case that a technology with high capital cost is the 'most expensive'.

A more holistic view of 'cost' can be gained by looking across the life cycle of the technology and considering their overall cost. Discussion of the economic merit of electricity generating technologies is, therefore, generally based on their levelised costs of energy (LCOE), which offer a measure of the overall costs of a technology over its life cycle per unit of electricity produced. It is expressed either as £/MWh or p/kWh, with £10/MWh being equivalent to 1 p/kWh. The results from such analyses give a cost or a range of costs for each technology, and are typically used to compare one technology with another.

It is important to note that LCOE, and cost in general, is not the only important factor in the economics of electricity generation; investors will also look at overall return on investment, which requires estimates of revenue to be determined. In a market setting this is a complex exercise, and the source of much uncertainty and risk. The extent to which this uncertainty can be mitigated is a large determinant of whether a particular generating technology can be regarded as an 'economic' investment. As such, LCOE alone is rarely used for actual investment decisions but it is regarded as a useful tool for policymaking, as long as the limitations are well understood (Royal Academy of Engineering, 2014).

2.2 Calculation Methodology

The levelised cost of energy is the sum of the discounted costs over the generator's lifetime, spread across the discounted units of energy produced over the lifetime. This is not simply 'adding up' the various costs, but requires future costs to be expressed in 'present value' terms by the process of discounting.

While there is no 'official' standard governing calculation of LCOE, there are several methodologies in use, including the 'IEA Method', the 'annuity method' and 'full cash flow' methods. The IEA method is the most common; for example, it has been used in studies by the International Energy Agency (IEA, 2010), and UKERC (Gross et al., 2007; Gross et al., 2013), as well as recent UK 'governmental' studies for, or by, the Department of Energy and Climate Change (DECC) and the Committee on Climate Change (CCC): Parsons Brinckerhoff (2010), Mott MacDonald (2010), Arup (2011), DECC (2012) and Poyry (2013).

The LCOE is given by:

$$\text{LCOE} = \frac{\sum_t^T \frac{C_t + O_t + F_t + D_t}{(1+r)^t}}{\sum_t^T \frac{E_t}{(1+r)^t}}$$

where C is the capital cost (£); O is operations and maintenance (O&M) cost (£); F is fuel cost (£); D is the decommissioning cost (£); E is the electricity produced (MWh); r is the discount rate (%); and t is the year in which a cost occurs during the project lifetime T . For a wind farm, no fuel is burned to generate power, so fuel cost is zero; however, indirect fuel use for transport is associated with many activities during the farm’s life.

Irrespective of which method is used, the calculation of LCOE requires a substantial number of factors to be determined, which can be split into those that determine cost and those that determine energy production. Figure 4 shows the main information that is required to estimate the costs and energy production of a typical wind farm. These reduce to three main factors: CAPEX, OPEX and energy production, which can then be considered along with the discount rate and other financial parameters.

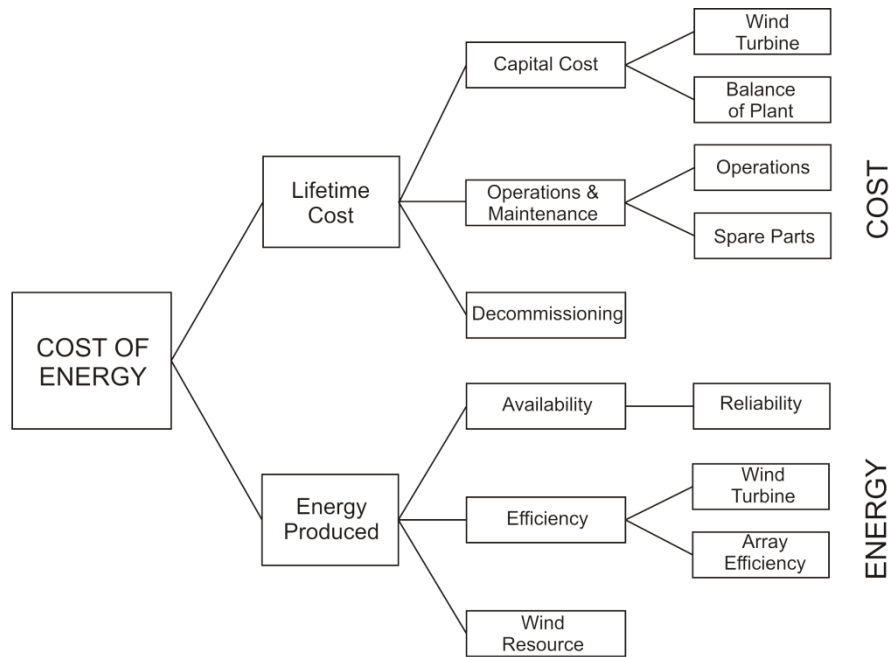


Figure 4 - Cost of energy for a wind farm

As LCOE is applied to many different generating technologies with a wide range of intended applications, there is substantial scope for variation introduced by different assumptions, methods and uncertainty. Figure 5 illustrates the areas where variation can be introduced in estimates of LCOE. These can be divided into four categories: variation in input data arising from the scenarios used, timing and locations, as well as uncertainty in the data itself; uncertainties introduced by the financial assumptions, again arising from location such as tax rates and treatment, prevailing financial treatments, whether pre- or post-tax rates are used, and adjustments for risk or inflation; variations in the physical and temporal boundaries analysed, and whether specific cost categories are included or not; and finally, differences in the methodology used, and intended scope.

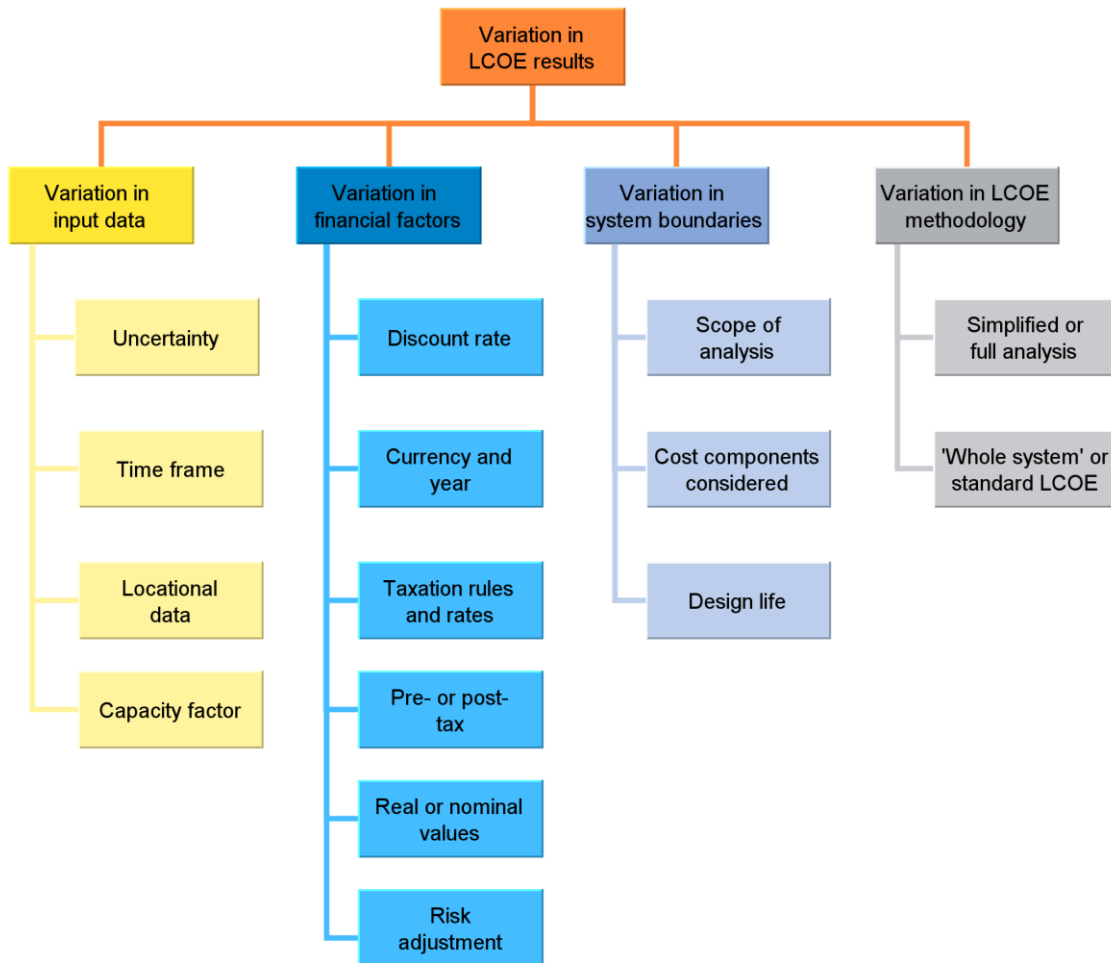


Figure 5 - Causes of variation in LCOE for a wind farm

A brief explanation of the terms used in Figure 5 follows:

- Cost uncertainty** – As UKERC (Gross et al., 2007) and the IEA (2010) point out, high-quality data is needed to produce reliable cost figures, but, as a result of privatisation and market liberalisation, there is often restricted access to commercially sensitive data on production costs. As a result, there is uncertainty around the figures. UKERC suggest that (engineering) consultants, through their role as advisors on projects, may have the best access to reliable and up-to-date cost data; information from consultants is the basis for much of the UK-specific analysis over recent years (CCC, 2011; Mott MacDonald, 2010; Parsons Brinckerhoff, 2011; Poyry, 2013). There is also a more fundamental issue regarding what is actually meant by ‘cost’, as it is possible to estimate costs in several ways: purely on the materials used; the actual cost of a component or system, including labour costs and overheads of the business; or the purchase price of a component or system, which includes profit for the seller. The latter is particularly important in market situations as, in time of scarcity, prices may rise; there is evidence of ‘congestion rent’ existing in the wind turbine and component market over recent years (Gross et al., 2013; Parsons Brinckerhoff, 2011).
- Time frame** – Wind farms built in different years will have different costs as designs, technical performance and practices change; these will be exacerbated by the impact of economic and financial factors including currency, inflation and financing terms.
- Locational data** – The costs associated with components and approaches will vary between locations, and the difficulties in comparisons are exacerbated by information from other countries. Most studies, particularly those for the UK, use ‘typical’ values for many aspects, providing a homogenised value, but some, for example Poyry (2013) apply location-specific costs.

- Capacity factor (or load factor)** – A measure of the energy production of a wind farm, defined as the proportion of energy generated over a period compared to maximum possible output. A great deal of emphasis has been placed on capacity factor as ‘evidence’ of wind farms being a poor choice; often this is as a result of it being mistaken for efficiency, or the amount of time that the wind farm operates for. The value depends on capacity and production, which means that although a large generator will produce more energy than a smaller one, it may not have a higher capacity factor. Capacity factor is, therefore, a major determinant of LCOE. It is the case that early turbines have poorer performance, while more modern devices have higher availability as experience and maintenance have improved, alongside much improved turbine siting and matching of turbine to conditions at the farm site. DECC (2011a) report that average onshore wind load factor in Scotland was 27.9% (27% for UK) over 2000 to 2012, with substantial inter-annual variation arising from wind patterns (2010: 21.7%; 2004: 33.6%). It is apparent that average capacity factors are above those suggested by some commentators, such as Gibson (2011). Recent concerns raised by Hughes (2012a) for the Renewable Energy Foundation suggested very rapid reductions in wind capacity factors with age; however, analysis by Staffell and Green (2014), that include a correction for wind speeds, demonstrate that actual turbine performance degradation is more limited.
- Discount rate** – Discounting is central to the LCOE calculation and describes the time value of money where the value of cash sums declines over time due to inflation, expectations of real returns and, critically, the risk that future costs may turn out to be different than expected. The discount rate is normally taken to be the weighted average cost of capital, combining higher expected rates of return to equity and lower debt rates. The discount rate reduces future costs whilst leaving capital costs largely unchanged; this is important when comparing technologies with very different cost profiles. Studies use discount rates expressed as pre-tax or post-tax as well as real or nominal rates; care must be taken in comparing studies, as post-tax rates will be lower than pre-tax, and nominal rates will be higher than real. In general, LCOE assessments use a single real pre-tax discount rate for all technologies, with recent UK and IEA LCOE studies using 10% as the real cost of capital for generation; however, other recent analyses (Oxera, 2011) have differentiated on the basis of risk.
- Risk adjustment** – Using the same discount rate across technologies, or for technologies across time, effectively ignores differences in risk (Awerbuch and Yang, 2008). Oxera (2011) currently estimate well-established dispatchable technologies (gas, hydro) to have a pre-tax real discount rate of 6 to 9%, onshore wind at 7 to 10% and offshore wind at 10 to 14%. While these adjustments are effective in differentiating project risk they do not tackle a more fundamental issue with most LCOE analyses: while trends in fossil fuel costs are captured, the risk arising from cost volatility is not considered (Awerbuch and Yang, 2008).
- Currency and year** – When and where studies relate to has a bearing on the values that are quoted. In particular, there are substantial swings in currency values relative to Sterling which can create changes in relative costs; this is a particularly important factor in the wind sector where the main suppliers are based outside the UK. This, along with changes in inflation and commodity prices (e.g. steel), can have a big impact on costs. Studies such as UKERC (Gross et al., 2013) and Bolinger and Wiser (2012), that take a longitudinal view, do account for these relative movements.
- Taxation rules and rates** – Most LCOE studies apply the IEA method in which such factors do not appear directly, although their impact arises indirectly in terms of expected pre-tax discount rates, which would be higher than post-tax rates. Studies using the full cash flow models explicitly account for these factors.
- Scope of analysis** – Different studies set different physical system boundaries for analysis: a single turbine, a farm including other infrastructure such as grid connection, or inclusion of ‘knock on effects’ elsewhere in the system – this is considered in detail in Section 3.
- Cost components considered** – Credible analysis of LCOE requires information on all cost components; as a minimum these need to include capital costs and operating costs. There are also costs associated with project development, which are detailed in most work, and decommissioning costs, which tend to be more uncertain so it has been practice to assume these to be equal to the scrap value of the assets (Royal Academy of Engineering, 2014). Nuclear differs with high decommissioning costs and uncertainty; although, discounting over many decades means that, at project evaluation, decommissioning costs are virtually negligible at realistic discount rates (IEA, 2010). The inclusion of ‘interest during construction’ (IDC) varies between studies, and particularly affects projects that have long construction periods where there are borrowings but no production. Effective assessment of this requires knowledge of the construction schedule and the financing; IEA (2010) and Gibson

(2011) both estimate the IDC. Most use an ‘overnight’ cost that includes pre-construction work, construction and contingency (Mott MacDonald, 2010). Onshore wind construction periods tend to be short so the impact will be limited.

- **Design life** – Typically a wind farm is considered to have a design life of 20 years, although there is variation in assumptions. A shorter design life will tend to raise LCOE and vice versa. The actual life time of the wind farm varies, normally determined by economic decisions around whether or not to ‘re-power’ the farm (where turbines are replaced with modern, larger turbines).
- **Full versus simplified analysis** – LCOE analyses using the IEA Method are simplified versions of assessments based on full cash flow models that explicitly consider a project from the investor (or equity) point of view and allow a more realistic evaluation of all costs applicable within specific jurisdictions. It is explicit about financing arrangements (debt/equity ratio and returns), loan periods, tax rates and depreciation. It takes the form (Schwabe, 2011):

$$\text{LCOE} = \frac{e \times C + \sum_{t=1}^T \frac{(1 - \text{Tax}) \times (O_t + F_t + D_t) - \text{Tax} \times (\text{Int}_t + \text{Dep}_t)}{(1 + r_e)^t}}{\sum_{t=1}^T \frac{E_t (1 - \text{Tax})}{(1 + r_e)^t}}$$

where e is the proportion of the project funded by equity; r_e is the return on equity; Tax is the tax rate; Int is the interest paid on the loan and Dep is depreciation. This makes explicit assumptions about accounting rules in different jurisdictions, which are complex and varied and which affect the timing and amounts of cash flows (Schwabe, 2011).

- **‘Whole system’ or standard LCOE** – There are many views on what the ‘true’ cost of wind power is and what additional costs can be attributed to it. Many of these additional costs are associated with the impacts of wind power on the operation and makeup of the electricity system, and include transmission upgrades, system balancing and provision of backup. Most studies do not consider these costs, but some do, including PB Power (2004), Gibson (2011) and Civitas (Lea, 2012). These issues are examined in more detail in Section 3. There are also other non-financial costs not captured by LCOE; these ‘external’ costs tend to be environmental and health impacts which, while challenging to quantify, show that fossil fuelled generation has relatively high external costs, while those for wind are very low. It is notable that carbon costs are also now being routinely included in LCOE analyses (CCC, 2011; IEA, 2010; Parsons Brinckerhoff, 2011), although remain absent in others (Gibson, 2011). IEA (2010) justifies their inclusion due to the existence of mature carbon policies such as the EU Emissions Trading Scheme, which associates real financial costs with carbon pricing.

2.3 Current Cost Estimates

In recent years there have been a series of studies providing estimates of costs for onshore wind farm and comparator technologies. These include UK-specific work for, and by, DECC, and the CCC (Arup, 2011; DECC, 2012; Mott MacDonald, 2010; Mott MacDonald, 2011; Poyry, 2013), as well as a range of work by international and overseas bodies. There has been only a modest amount of peer reviewed academic work published alongside this, some by pressure groups and individuals; of note are the longitudinal investigations into the variations in levelised costs over time including those by UKERC (Gross et al., 2013), the IEA Wind Task 26 (Lantz et al., 2012) and the Berkeley Laboratory (Wiser, 2012), which have been valuable in indicating the basis for wind cost variation over recent years. Given the reported variations in costs, only relatively recent studies have been included here.

The methods employed in estimating costs are varied: parametric cost models (Tegen, 2013); project development, survey or reverse engineered (Mott MacDonald, 2011); anonymised price reporting (Milborrow, 2013), or re-engineered from other sources (Gibson, 2013; Gibson, 2011). A summary of the LCOE studies analysed is shown in Table 2, where the LCOE is given in its original currency values along with costs corrected to 2011 pounds sterling (indicated by ‘£2011’).

Treatment of uncertainty varies between studies: none; simple percentage ranges; scenarios for specific parameters based around a central value with high and low values (CCC, 2011); reporting full ranges of parameter sensitivities (Tegen, 2013); ‘probabilistic’ estimates using subjective weighting for key parameters (Gibson, 2011); or location-specific parameter values allowing differentiation between capacity factor, costs and ultimately LCOE (Poyry, 2013).

Capital cost

For onshore wind, capital cost is the dominant determinant of LCOE. It typically accounts for 80 to 90% of overall life cycle costs and is either expressed in terms of cost per unit capacity of wind farm (£/kW), as a total cost, or as a component of the levelised cost (£/MWh). Most of the studies reviewed provided capital costs explicitly.

The capital cost is itself broken down by a series of major cost items relating to the development of the project, purchase of equipment, transportation, site preparation and installation. Figure 6 shows an example for a large UK onshore wind farm (Mott MacDonald, 2011). This, and most other studies, gives figures on a ‘farm’ basis, and includes the costs of connecting the farm to the grid but excludes interest during construction.

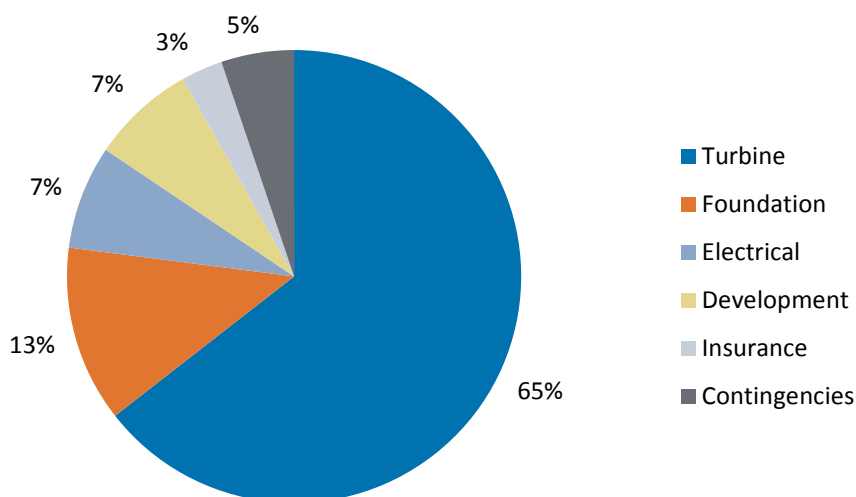


Figure 6 - Typical breakdown of capital cost for large onshore wind farm (Mott MacDonald, 2011)

The vast majority of the capital cost is represented by the turbine itself, accounting for 60 to 80% (Krohn, 2009), with the example in Figure 6 at the lower end of the scale. This variation arises not only from variation in the cost of the turbines, but also the attributes of the farm: country of installation, distance from suitable grid connection points, terrain and geology. In many cases, grid connection will be the most expensive item after the turbine (Krohn, 2009). The capital costs generally also include other costs, such as development, insurance and contingency.

Materials costs are a determinant of capital costs, and changes in commodity prices (particularly steel and copper) contribute to price variations. Perhaps surprisingly, materials costs for onshore wind farms contribute a modest 2.5%, while labour costs associated with manufacture, on-site or project management represent the largest part of the capital cost (Mott MacDonald, 2011). This is particularly true of the carbon- and glass-fibre manufacturing processes for the turbine rotors, which remains largely manual. Additionally, the extent of the component supply chain, competition and the impacts of scarcity and supplier’s contingency costs contribute to variations in capital costs.

The price of turbines has fluctuated substantially over the last ten years, as Bolinger and Wiser (2012), Lantz et al (2012) and UKERC (Gross et al., 2013) show. Among the studies that report separate capital costs for onshore wind turbines there is some variety: Mott MacDonald (2011) suggest a price of £870/kW, Arup (2011) around £1000/kW, IRENA (2012) report turbine prices of \$1400/kW (£950 in £2011), and Lantz et al (2012) suggest \$1,300/kW to \$2,150/kW (£800-1350/kW) with the upper end of the range being newer turbines designed for low wind speed conditions. There is a tendency for larger turbines to have lower prices per unit capacity than small ones.

Given the contribution of turbine costs, there is variation in the reported onshore wind farm capital costs. Figure 7 provides an overview of UK-specific costs, and a selection of studies reporting international (INT), European (EU) and country-specific costs (DK and DE). Even correcting for currency and inflation, there is variation arising from the year of study and location. The central points of these studies suggest a typical capital cost of around £1350/kW, adding credibility to the values suggested by UK-specific studies by CCC (2011), Mott MacDonald (2011) and Poyry (2013) and suggesting UK costs are about average, internationally. The range of central values is £1100 to 1579/kW. Most studies offer a range of costs, although, as there is limited consistency in terms of how uncertainty in capital costs is reported, interpretation requires care. In Figure 7, IEA (2010) reports costs from a range of OECD countries, with the high upper cost for Switzerland arising from complex terrain and low wind speeds.

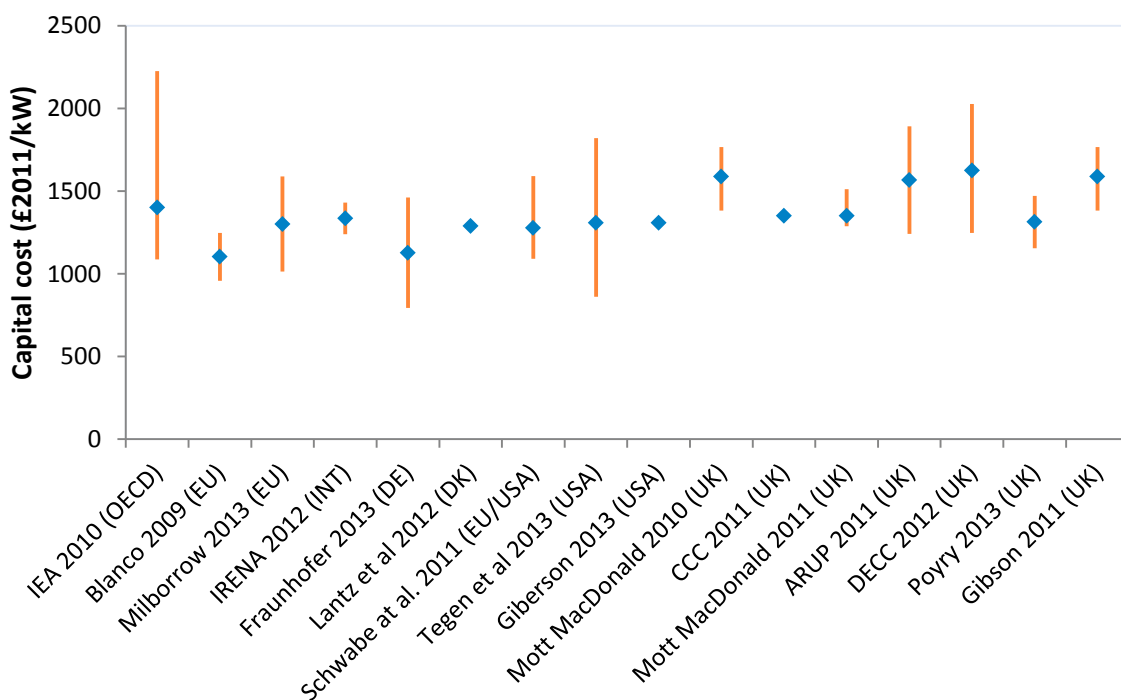


Figure 7 - Onshore wind farm capital costs (£2011). The vertical lines shows the reported range of costs within each study and the horizontal bar shows the reported median value or mean of the range where none is given.

Operation and maintenance costs

Although operating costs are less significant than capital costs they remain a key input to levelised cost calculations; as a proportion of LCOE O&M accounts for 15 to 24% of overall life time costs. They are expressed as fixed and/or variable components in a number of different ways: a fixed annual cost based on percentage of capital cost (%); a fixed annual cost per unit of capacity (£/kW/yr); or a variable cost or levelised cost per unit of production (£/MWh). The wide range of presentations makes direct comparison less straightforward – in general

O&M costs for onshore wind are low. Studies expressing O&M as a proportion of CAPEX suggest values around 1 to 1.5% which is in line with those expressing costs on other bases.

Decommissioning

Decommissioning costs are largely neglected in studies as, for the reasons outlined earlier, the discounted value is generally minimal. In studies that include such costs for wind power, they are included as a percentage of capital cost, e.g. 5% (IEA, 2010); or per kW, e.g. \$2010 0.6/kW (Schwabe, 2011).

Levelised Costs

The variations in capital and operating costs feed through into the overall levelised cost of energy estimates. Here they are joined by a series of other factors that lead to significant variation in LCOE. Figure 8 shows the range of LCOE estimates (in £2011) for the same studies shown in Figure 7, as well as values from Civitas (Lea, 2012) and the UKERC (Gross et al., 2013) review for comparison.

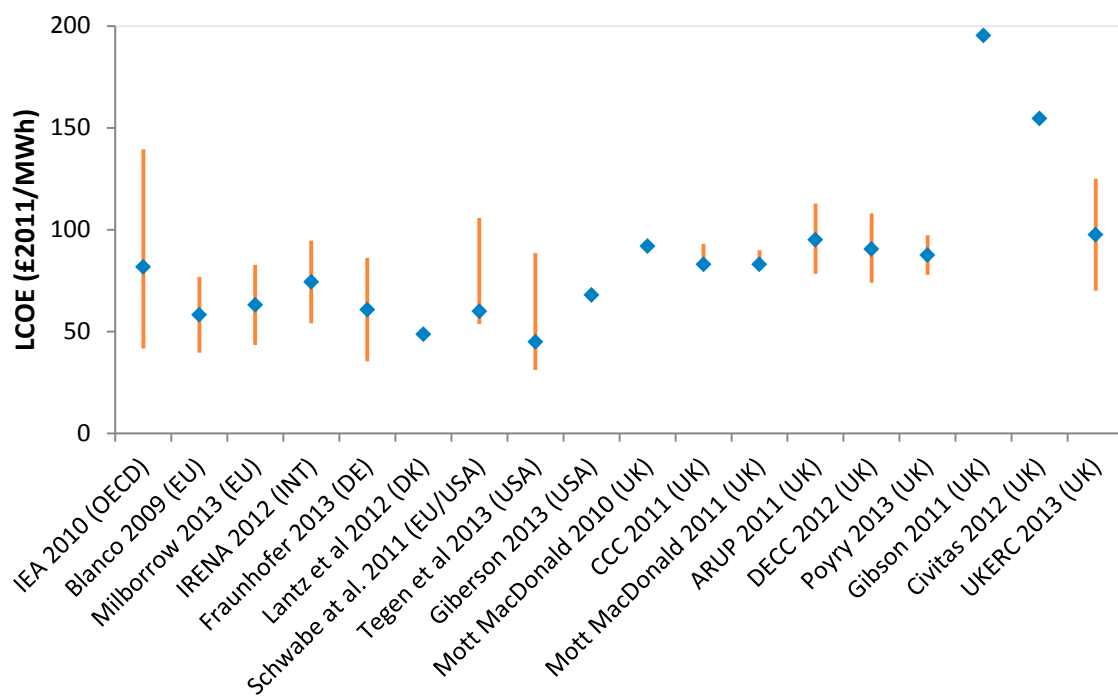


Figure 8 - LCOE of onshore wind farms (£2011). The bar shows the reported range of costs within each study and the bar shows the reported median value or where none given the mean of the range.

Several things are apparent:

- Higher values of CAPEX do not automatically translate into higher LCOE; for example, Tegen et al. (2013) has one of the higher CAPEX ranges but one of the lowest LCOE ranges;
- UK-specific studies and the UKERC ranges tend to show higher LCOE values than those for overseas;
- The spread of values is much greater overall with two studies in particular indicating substantially higher LCOE.

The UK studies have a mean capacity factor of 27%, whereas the international studies have higher capacity factors, on average, with a spread between Germany and the USA. It is notable that the lowest LCOE are from Fraunhofer (2013) for Germany and Tegen et al. (2013) for the USA. The UK studies almost universally apply the simplified LCOE method and have 10% pre-tax real discount rates or, even when risk-adjusted, just below. Other than the IEA (2010) and IRENA (2012) studies, which use a similar discount rate and method, the international studies tend to have substantially lower real discount rates: Fraunhofer use a very low 3.8%, while the 8% nominal discount rate

used by Tegen et al. (2013) is equivalent to 5.7% real. Gibson (2011) is unusual in applying a post-tax rate. The national variation in discount rates reflects expectations of cost of debt and equity as well as financing preferences. In addition, Oxera (2011) note that discount rates also reflect perceptions of a range of risks including those from policy.

The levelised cost of onshore wind is very sensitive to assumptions on capacity factor, lifetime, discount rate and financing structure (Schwabe, 2011). A good illustration of this is in comparing the central LCOE figure from Tegen et al. (2013) with Giberson (2013); Giberson uses the cost data in Tegen et al (2013) and makes a series of adjustments that, in their opinion, are 'reasonable', that raise LCOE from \$72/MWh to \$109/MWh (£45 and £68/MWh in £2011 prices). These adjustments are shown in Table 1.

	\$/MWh	\$/MWh
Tegen et al (2011) LCOE estimate		72
Adjustments:		
Reduced capacity factor from 38% to 33%	+8	
Use of a higher nominal discount rate of 10% rather than 8%	+10	
Use of a 20 year depreciation schedule rather than the 5-year write down allowed by US laws for renewable energy projects	+8	
Use of \$21/MWh rather than \$11/MWh operating expense	+10	
Total adjustments		+37
Giberson (2013) LCOE estimate		109

Table 1 - Summary of selected wind LCOE analyses and their key parameters

The final point is that that Gibson (2011) and Civitas (Lea, 2012) have much higher apparent LCOE as a result of adding 'system costs' to the baseline levelised costs. Civitas (Lea, 2012) combines the £88/MWh baseline LCOE from Mott MacDonald (2010) with £60/MWh of system costs based on Gibson's estimates of balancing, additional backup and transmission costs. Gibson's higher LCOE estimate is made up of a £75/MWh system cost and a baseline LCOE of £112/MWh, despite also using Mott Macdonald (2010) cost components. In part, both figures are higher as a result of a more conservative 25% capacity factor. More importantly, close inspection of Gibson's spreadsheets suggests a series of factors that serve to inflate the LCOE: a 'full' LCOE method is used that calculates IDC using a very high 12.5% post-tax equity rate of return, low gearing and a separate debt repayment charge applied at the overall discount rate. The latter item is effectively double-counting and it is notable that the financial treatment of on- and offshore wind differs from the other generation types examined. Both studies are clearly marked in Figure 10. The system costs are examined in more detail in Section 3.

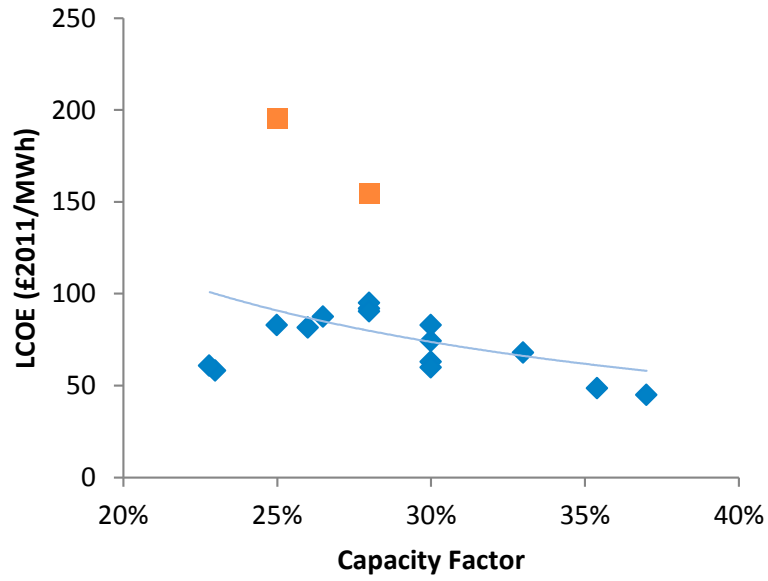


Figure 9 - LCOE variation with capacity factor. Studies that include system costs are clearly identified by square markers.

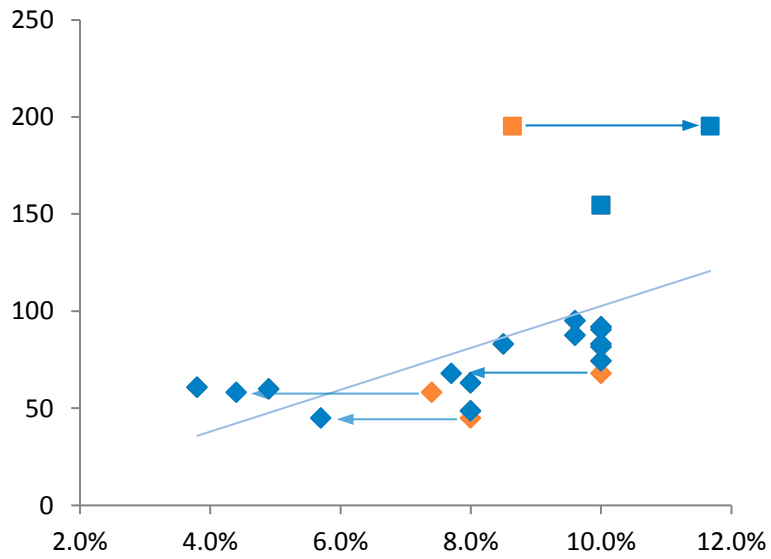


Figure 10 - LCOE variation with discount rate. Blue markers indicate headline discount rates with orange markers and arrows showing studies where discount rates have been restated on a pre-tax real basis. Studies that include system costs are identified by square markers.

Study	Location	Type of analysis	Currency unit	LCOE	LCOE	Capital costs	Capacity factor	Lifetime	Discount rate	System costs
				Currency/MWh	£2011/MWh	£2011/kW	%	Years	%	
IEA (2010)	OECD	Simplified	\$2008	137 [70, 234]	82 [42, 139]	1400 [1099, 2214]	26 [20, 41]	25	10.0%	No
Blanco (2009)	EU	Full	€2008	[45, 87]	58 [40, 77]	1102 [970, 1235]	23 [19, 35]	20	7.4% ^{PN}	No
Milborrow (2013)	EU	Simplified	€2012	[55, 105]	63 [43, 83]	1301 [1025, 1577]	30 [20, 40]	20	8.0%	No
IRENA (2012)	INT	Simplified	\$2010	[80, 140]	74 [54, 95]	1335 [1250, 1419]	30 [25, 35]	20	10.0%	No
Fraunhofer (2013)	DE	Simplified	€2013	[44, 107]	61 [35, 86]	1127 [805, 1449]	23 [15, 31]	20	3.8%	No
Lantz et al (2012)	DK	Full	\$2010	72	49	1289	35	20	8.0% ^{PN}	No
Schwabe et al. (2011)	EU/USA	Full	€2008	68 [61, 120]	60 [54, 106]	1278 [1102, 1578]	30	20	4.9%	No
Tegen et al (2013)	USA	Full	\$2011	72 [50, 142]	45 [31, 89]	1308 [873, 1808]	37	20	8.0% ^{PN}	No
Giberson (2013)	USA	Full	\$2011	109	68	1308	33	20	10.0% ^{PN}	No
Mott MacDonald (2010)	UK	Simplified	£2010	88	92	1588 [1394, 1755]	28 [25, 31]	24	10.0%	No
CCC (2011)	UK	Simplified	£2011	83 [81, 93]	83 [81, 93]	1350	25	20	10.0%	No
Mott MacDonald (2011)	UK	Simplified	£2011	83 [83, 90]	83	1350 [1300, 1500]	30	20	8.5%	No
ARUP (2011)	UK	Simplified	£2010	91 [75, 108]	95 [78, 113]	1567 [1253, 1880]	28	24	9.6%	No
DECC (2012)	UK	Simplified	£2012	93 [76, 111]	90 [74, 108]	1624 [1258, 2014]	28	20	10.0%	No
Poyry (2013)	UK	Simplified	£2012	[80, 100]	88 [78, 97]	1313 [1167, 1459]	27 [22, 31]	24	9.6%	No
Gibson (2011)	UK	Full	£2010	187 [160, 215]	195	1588 [1394, 1755]	25 [22, 28]	20	8.6% ^{TR}	Yes
Civitas (Lea, 2012)	UK	Simplified	£2010	148	155	1588 [1394, 1755]	28	24	10.0%	Yes
UKERC (Gross et al., 2013)	UK	Review	£2011		98 [70, 125]		-	-	-	No

Table 2 – Summary of selected onshore wind LCOE analyses and their key parameters. LCOE, CAPEX and Capacity factor values are in the form “Central [Low, High]”. Discount rates are real pre-tax rates (weighted average cost of capital) except ^{PN} pre-tax nominal, ^{TN} post-tax nominal and ^{TR} post-tax real rates. Low outliers are highlighted in blue, high outliers in orange.

Comparison with other generating technologies

Many studies reviewed and referred to in the cost analyses presented earlier offer comparisons between wind and other technologies. In the main the UK-specific analyses are representative, and the UKERC study (Gross et al., 2013) conveniently provides an analysis of current levelised costs, as summarised in Table 3. It is apparent that there are substantial uncertainties around all technologies: capital cost, capacity factor and discount rate are important for nuclear while fossil fuel and carbon costs are important factors for CCGT.

An aspect that often gets overlooked in comparisons is that the LCOE for thermal power plant generally assume operation as baseload with capacity factors that are at the upper end of the range (85-90%). In an electricity system with variable demand it is not possible that all thermal plant will operate as baseload, as marginal cost will dictate that some will operate less frequently so their capacity factor will decline and LCOE will increase; this effect is expected to be enhanced as more wind enters the system, squeezing operational opportunities for gas and coal generation.

Although it is evident that offshore wind is substantially more expensive at present, the overlapping of the ranges for nuclear, onshore wind and combined cycle gas turbines means there is no clear outcome in terms of which technology is currently 'cheapest' on the basis of levelised costs.

Generation technology	Range (£/MWh)
Nuclear	70 – 105
Gas (CCGT)	60 – 100
Onshore wind	70 – 125
Offshore wind	100 – 200

Table 3 - LCOE of a range of generating technologies: on and offshore wind, combined cycle gas turbines and nuclear generation (in £2011) based on sample of UK studies by UKERC (Gross et al., 2013)

2.4 Outlook for LCOE of Onshore Wind

For many new and established generating technologies there is an expectation that costs will come down and performance will increase with time; a wide range of literature on innovation supports this view. UKERC (Gross et al., 2013) summarises the mechanisms through which this occurs and compares the two main approaches used to project future costs:

1. Technical engineering assessment; and
2. Extrapolation using experience curves (or learning rates).

Engineering assessment breaks down a system into constituent parts, and parametric modelling is used to examine contributions to overall cost and scope for improvements (Mukora et al., 2009). Experience curves, on the other hand seek, mathematical relationships between historic costs and the cumulative production of a product; this can be extrapolated into the future to assess potential costs at specific levels of deployment. The key parameter in experience curve analysis is the 'learning rate' – with a higher value resulting in a faster decrease in costs with installed capacity. Such studies have been widely used, but UKERC (Gross et al., 2013) have identified a number of limitations, and conclude that engineering assessment may be the most appropriate method for assessment of emerging technologies, while learning rates then become more appropriate once a track record is established. Gross (2013) further note that cost gains due to learning may be overwhelmed by external factors, including fuel and commodity prices and supply chain issues, and that many of these factors are uncertain and volatile.

Although it does not explicitly identify cost projections from individual studies, UKERC's analysis of available literature suggests a generally downward cost trend for most technologies apart from gas, but identifies that a substantial range exists, as Table 4 shows. To illustrate the point several studies for onshore wind have been picked out for further analysis.

Generation technology	2020		2030	
	Central value	Range	Central value	Range
Nuclear	70	30 - 130	60	30 - 125
Gas (CCGT)	94	55 - 108	96	52 - 138
Onshore wind	83	47 - 112	88	71 - 104
Offshore wind	127	92 - 140	112	98 - 130

Table 4 - Forecast LCOE for generating technologies: on and offshore wind, combined cycle gas turbines and nuclear generation (in £2011) based on sample of UK studies by UKERC (Gross et al., 2013)

Mott MacDonald (2011) expect capital costs for onshore wind farms to fall in real terms by around 12% by 2020, with a more substantial 22% reduction projected by 2040; the turbine contributes most to this. Together with reductions in discount rate from 8.5% currently to 6.4% in 2040, this would equate to LCOE falling from £83-93/MWh to £63-72/MWh in 2020 and £51-61/MWh in 2040. A more substantial 20 to 30% drop in LCOE by 2030 is suggested as credible by Lantz et al. (2012); however, Arup (2011) anticipates modest (0.5%) cost reductions over the next 20 years as industry learning is partly offset by steel price increases. With no change in discount rate or capacity factor, Arup expects median LCOE to fall from £91/MWh to £86/MWh by 2020 and further decrease to £82/MWh by 2030.

Key Messages

- Capital costs for onshore wind are approximately £1350/kW.
- Two studies (Gibson, 2011; Lea, 2012) show life cycle costs that are notably above others arising from inclusion of very high estimates of system costs. Further Gibson (2011) uses high discount rates, low capacity factors and otherwise unusual financial treatments.
- Lantz et al (2012) and Tegen et al (2013) suggest exceptionally low cost of energy which is attributed to relatively low discount rates and capacity factors that are very high for the UK (but credible for USA).
- Discount rate assumptions are critical to the eventual levelised cost of onshore wind; post tax real discount rates of 10% are typical for the UK and higher than international comparators.
- Currently onshore wind is substantially cheaper than offshore wind and broadly comparable with nuclear and gas generation; there appears to be moderate scope to reduce costs by 2020.

3 Effect of wind power on system costs

The impact of wind on other generators, and the system as a whole, is generally excluded from levelised cost calculations, although some studies do include them; for example PB Power (2004), Gibson (2011), Civitas (Lea, 2012), and the American Tradition Institute (ATI) (Taylor and Tanton, 2012). Some of these studies that include 'system costs' use it as evidence that wind energy costs are "significantly understated [because] they failed to take its unusual indirect and infrastructure costs into account" (Taylor and Tanton, 2012). These studies suggest that inclusion of the system costs of onshore wind increases the apparent cost by 45% (PB Power, 2004), 67% (Gibson, 2011), 68% (Lea, 2012), or 50-90% (Taylor and Tanton, 2012).

In essence the 'system' costs that are referred to are:

- The costs of balancing the power system to cope with the variable output of wind farms;
- The costs of providing 'backup' or, more specifically, costs of ensuring there is sufficient generation capacity to meet demand;
- The cost of additional transmission that is required to connect wind plants, and the losses associated with it.

There have been several reviews of aspects of these costs – notably Costs and Impacts of Intermittency (Gross et al., 2006), as well as a wide range of relevant studies since then. The IEA (2010) make the point that "there is no disagreement between experts that such system costs for non-dispatchable renewables exist [but there is] little agreement (and, in fact, very little information) about their precise amount". Studies show that generation mix, network capacity and interconnection, as well as the availability of mechanisms for managing variability, are important in determining costs, which makes comparison challenging.

Additionally, while the operation of the power system (or national grid) operates on relatively simple concepts, the system itself is highly complex, requiring substantial engineering expertise to operate securely and efficiently. Furthermore, the engineering practices required to achieve this do not feature in the (classical) economic theories that explain market operation; as such, there is substantial scope for misunderstanding terms and outcomes.

3.1 Balancing

The variable nature of wind power, in contrast to conventional, dispatchable technologies, requires flexible 'reserves' to be on hand for times when the resource is not available (IEA, 2010); therefore, the cost of onshore wind is higher at system level than at farm level.

Reserves are used to handle unpredicted variations in demand or generation on a range of timescales from seconds to around four hours. They include 'frequency response' generation that automatically reacts to rapid changes such as the sudden loss of a large generator, and operating reserve, which deals with slower variations over time, such as changing generator availability or incorrect forecasts. Operating reserves are provided by power stations running at part load, standby generators that can be started quickly (hydro, diesel, open cycle gas turbines), as well as (some) contracted demand response. Reserve therefore creates costs in terms of operating power plants less efficiently, as well as the cost of contracts for ensuring standby generation is available. The amount of reserve is specified by National Grid on the basis of the largest generator than can be lost, and the level of error in forecasting demand and wind four hours ahead of delivery. Increases in wind capacity will therefore increase the amount of reserve that needs to be held, but this amount depends on overall expected errors, not simply that of wind. The four hour window is important as this is the standard lead time to start a thermal power plant to cover shortfalls. National Grid handles this through the Balancing Mechanism and several other schemes.

IEA (2010) compares several international studies that show balancing costs increase with wind penetration, although the rate of increase does level off: at penetration levels of up to around 20%, costs are around £0.60 to

4/MWh (\$1 to 6/MWh), or around 10% of wind cost. Katzenstein and Apt (2012) note that the costs of handling variability of wind power in Texas reduces as wind capacity factors increase, and as the number of plants increases. The Eastern Interconnection Wind Integration Study (EnerNex Corporation, 2011) shows that for large balancing areas and fully developed regional markets, the cost of integration is about \$5/MWh (US\$ 2009). Specific studies for the UK also suggest increases in the volume of reserve held: Strbac et al (2007) suggest an extra 4.6 to 6.3 GW of reserve will be necessary to integrate 25 GW wind, costing £3.4 to 6.3/MWh (corrected to £2011); National Grid (2010) estimate that the extra balancing costs for wind for a 40% wind penetration in 2020 are of the order of £500–1000 million per annum (£3.5–7.0/MWh of wind). The uncertainty in these estimates arises from the uncertainty of the future trajectory of the costs of balancing services, as they are dependent on fuel prices. For comparison, the cost of balancing the system in 2012/13 was £803 million (~1% of customer bills), of which £170 million was due to managing grid constraints and £7 million for constraining wind farms.

A concern that has arisen in recent years has been around the impact that ‘cycling’ of thermal power plants has on the fuel savings due to wind operation. While one of the less credible studies (le Pair, 2011b) is examined in detail in Appendix 3, there is a reasonable basis for concern. The issue arises from the need to operate thermal power plant flexibly to respond to wind power production, leading to part-loading, increased ramping, and additional shutdowns and start-ups. This potentially leads to costs associated with higher fuel consumption per MWh due to less efficient operation, as well as impacts on operations, maintenance and reliability. Denny and O’Malley (2009) suggest that fuel associated with on-off cycles represent a modest part of the costs, between 2 and 50% depending on the generator. The ATI (Taylor and Tanton, 2012) speculate that ‘additional gas consumption’ would cost \$4 to 8/MWh despite admitting that they were unaware of the true penalty. A more credible analysis by NREL (2013b) found that, for the Western Integration in the USA, the increase in O&M costs from cycling were \$0.14–\$0.67 per MWh (<40p/MWh) compared to around \$30/MWh of cost savings associated with avoided fossil fuel use.

Overall, the literature suggests that balancing costs are likely to be lower in larger markets, with a geographical spread of plants, and when wind is part of a complementary portfolio of other generation technologies (IEA, 2010). This is important in considering wind integration in Scotland as, while Scotland’s wind penetration will be locally very high, it is the penetration at GB level and the extent of transmission and external interconnections that will strongly govern balancing costs. While there are undoubtedly additional balancing costs arising from integrating variable wind, the IEA (2010) and other studies suggest they are not prohibitive. Additionally, the Committee on Climate Change suggest that, with the right investment in flexibility in the form of storage, demand side management and interconnection, costs can be managed (at around £10/MWh) at even relatively high levels of renewables penetration. Furthermore, these are generally one-off investment costs and “low compared to costs of deploying renewable generation” (Barrs, 2011).

3.2 Backup

Ensuring that there is sufficient generating capacity to provide secure electricity supply is a key issue, and concern is expressed about ensuring ‘backup’ is provided to cover days when there is little or no wind; however, in analysing this issue some studies make an explicit assumption that additional dedicated generating capacity must be built to ‘firm up’ wind, and that this entails high additional costs to cover capital and operating expenses. Gibson (2011) and Civitas (Lea, 2012) refer to this as ‘Planning Reserve’ but the same idea is used in studies by PB Power (2004), the ATI (Taylor and Tanton, 2012) and Hughes (2012b). In arriving at a cost of £16.7/MWh (~£20/MWh in £2011), PB Power (2004) assume that open cycle gas turbines are built to cover the equivalent of 65% of the wind capacity, with the capital cost, fixed O&M and the difference between the marginal fuel cost of OCGT and CCGT generation attributed to wind. Gibson (2011) assumes 92% of wind capacity is required as backup,

hence a higher £28/MWh charge (and £24/MWh for Civitas (Lea, 2012)) to cover capital costs. The ATI (Taylor and Tanton, 2012) assume 75% of the wind capacity would be backed up by CCGT, amounting to \$17/MWh of backup costs.

Milborrow (2009) and others make the point that this is not a realistic representation as, in reality, all fossil fuel and nuclear power stations in a given system provide backup to all others. As each has a statistical probability of experiencing an outage, more capacity is built than is required at peak demand levels to cover this eventuality. Wind is essentially no different, although its availability is governed by the weather rather than the mechanical reliability. As wind is added to the system it, in itself, adds to system reliability and this is referred to as its 'capacity credit'. Importantly, as wind is added, it is not automatically the case that other power plants are retired. Unlike in planned systems, there is no specific entity responsible in Great Britain for deciding when power plants should be connected or withdrawn; rather market participants decide on the basis of expected profitability, among other considerations. The theory is that price signals should ensure generation is built at appropriate times, although the recent Electricity Market Reform introduced a capacity payment to make capacity signals more explicit. Given this, there is a drastic over-estimate of the cost of backup power in analyses that assume that wind needs dedicated provision. Additionally, the very low (8%) capacity credit estimates used by Gibson (2011) and Civitas (Lea, 2012) serve to inflate the amount of backup; Milborrow (2009) notes that other estimates for capacity credit are in the region of 20%.

There have been a number of system planning studies undertaken in recent years that aim to optimally plan the GB system with high penetrations of renewables, such as Poyry (2011) for the CCC. The default assumption is that, in constructing a generation portfolio to meet variable demand, some peaking plant is required. Typically these are open cycle gas turbines with very low capital cost but high running costs. Analyses for the CCC show that use of flexibility introduced by storage, demand side response and interconnection, mean requirement for peaking plant such as OCGT for meeting shortfalls is low, but not eliminated: costs of £30 million/year at 40% GB renewable penetration equate to 'backup costs' of £0.2/MWh. The key point is that provision of full back up of renewable capacity is not necessary for secure supplies where flexibility is encouraged.

3.3 Transmission

The cost of investment in transmission lines, cables and associated infrastructure is also a key theme, with Gibson (2011) and Civitas (Lea, 2012) attributing very high transmission costs to wind. In determining a cost of transmission Gibson (2011) uses the cost of the contentious Beaulieu-Denny line, and extrapolates to a reinforcement cost of £31/MWh. While this is, in principle, a reasonable approach, the calculation uses a low capacity factor for wind and, in the financing calculations there appears to be double counting and the use of a cost of capital that is much higher than the 4.5% value for a regulated network utility suggested by Ofgem (2012).

Parsons Brinckerhoff (2012) offer lifetime transmission costs for a range of overhead line and cable installations that can be used to derive indicative costs per MWh for a wind farm connected to one end; for example, 40-year lifetime costs of £168 million (£2.2 million/km) are estimated for a 75km double circuit overhead line able to carry 3190 MVA (a substantial proportion of this cost is due to transmission losses). Assuming 1 to 3 GW of wind farms connect to the line, this suggests a range of costs of £1.6 to £5/MWh (with a capacity factor of 28%, variation of losses with circuit loading ignored and discounting wind output at 6.25%). The lower the use of the line the more expensive it becomes on a per MWh basis.

A difficulty in estimating cost of transmission on this basis is that transmission lines generally add to, or uprate, an existing interconnected system. The power flows are therefore more complex: lines are not loaded to maximum to ensure stability and post-fault security, and there are often a series of related upgrades. These factors make it

difficult to attribute costs to wind, although more realistic estimates can be gained from full transmission studies. Mills et al (2009) reviewed an extensive set of US transmission studies for wind connection, and found a median cost of \$300/kW of wind installed (~15% of then wind farm capex at the time) and median unit cost of \$15/MWh wind produced [(~£10/MWh). Similarly the Eastern Interconnection study (EnerNex Corporation, 2011) found transmission costs to be \$15/MWh (\$2009). The choice over how much transmission capacity to build can be determined either by system security considerations, which define the level of redundancy, or on a cost-benefit basis, where the cost of transmission is compared to the cost of constraints. Constraint costs have been a particular area of contention in recent years, particularly where payments have been to wind farms. National Grid estimate that the cost of constraints in 2011/12 was £324million, of which £31million was for wind constraints, while in 2012/13 constraint costs dropped to was £170million, with £7million for wind. They attribute the reason for reduced costs to investment in the network. In its analyses, the CCC (Barrs, 2011) suggest that transmission costs are likely to rise with renewables penetration: wind generation in the north will tend to increase need for capacity, but where it is closer to the south it may save on the cost of transmission to accommodate non-renewable plant elsewhere. The CCC estimates the cost of transmission requirements to be between £5 and £10/MWh (Barrs, 2011).

In considering the cost of transmission expansion, it is important to note that other generation sources will also require transmission expenditure, not just wind. It is reasonable to say that gas power plants have more choice over location than wind (Giberson, 2013) or, for that matter, coal and nuclear plants, which require ready access to cooling water. Many replacement plants choose to locate at existing sites precisely to avoid transmission expansion (although there may well be requirements to extend or reinforce the gas transmission network to accommodate gas power plants). Mills et al. (2009) emphasise that transmission expansion typically serves multiple purposes, and that assigning the full costs of expansion to additional (wind) generation capacity effectively ignores other benefits. Furthermore, a problem with some analyses is that they ignore the fact that the GB transmission system, as it exists in present form, was *deliberately* planned to strategically accommodate given resources, with costs socialised largely through the nationalised industry. It simply happens that those strategic assets were nuclear and coal power plants and large scale pumped storage, rather than wind.

3.4 Total 'System' Costs

While there are variations in the literature for each of the components of 'system cost', there is clear distinction between these and the overall costs suggested by the four LCOE studies that incorporate 'system costs':

- **Backup costs** – Overstated in all cases as a result of a partial understanding of the system;
- **Transmission costs** – Gibson (2011) and Civitas (Lea, 2012) overstate this, although the ATI (Taylor and Tanton, 2012) are in line with other US literature (the ATI do, however, add an extra \$10/MWh for transmission losses, despite transmission assessments including loss costs as a key cost);
- **Balancing costs** – The ATI (Taylor and Tanton, 2012) are broadly in line with other literature, while Gibson (2011) and Civitas (Lea, 2012) are high. This aspect has been specifically criticised, and it transpires that the £16/MWh is quoted from PB Power (2006), which in turn references Dale et al. (2004). That study estimates the difference in total costs of a system with 20% wind and that of a gas-only system, and includes overall generation, transmission and distribution costs alongside fuel and balancing costs. The £16/MWh figure is the additional cost per unit of wind produced which equates to £3/MWh per unit of electricity sold.

Taking the range of credible estimates from the literature for each component of system costs allows an estimate to of total systems costs to be made for penetrations of up to 40% wind shown in Table 5. A total of £7 to 18/MWh straddles the £10/MWh cost of intermittency suggested by the Committee on Climate Change (Barrs, 2011), and would represent around 5 to 15% of baseline levelised cost of onshore wind.

Cost component	Range (£/MWh)
Balancing costs	2 – 7
Backup costs	0.2 – 0.5
Transmission costs	5 – 10
Total 'system' costs	7 – 18

Table 5 - Range of 'system' costs associated with wind power (in £2011)

Ultimately, while the estimates of Gibson (2011), Civitas (Lea, 2012) and others for the system costs of wind are very overstated, it remains the case that system costs are real. The IEA (2010) suggest that “part of the cost of such system’s reserves should in principle, be added to the LCOE of intermittent renewables when compared to other baseload generation sources”. Substantial system costs exist, even in zero wind systems, precisely because the nature of electricity supply requires backup, balancing and transmission to allow individual, isolated, generators to contribute.

Key Messages

- System costs such as balancing, provision of backup and transmission arise in accommodating wind energy.
- These are relatively small compared to the overall cost of electricity supply or the levelised cost of wind; the review suggests values of £7 to 18/MWh are credible.
- Studies that included system costs in their levelised cost estimates (PBPower, 2004; Gibson, 2011; Lea, 2012; Taylor and Tanton, 2012) tended to systematically overstate them.
- There is particular overstatement of the cost of 'backup' where it is assumed that dedicated conventional generation must be constructed for periods when the 'wind is not blowing'; the reality is that with other generation on the system operating flexibly the requirement for backup is modest.
- There is substantial scope for energy storage, demand side response to provide flexibility.

4 Life Cycle Carbon Emissions

4.1 Carbon, greenhouse gases and emissions savings

Much of the confusion in the existing debate over the life cycle carbon emissions of wind power generation is thought to arise in a lack of comprehension of the different aspects of the calculation, along with uncertainty over the terminology. The term 'carbon emissions' itself is unclear: it may be a measure of the emissions of all gases containing carbon, including those that have no global warming potential; it may focus solely on carbon dioxide and/or methane; or it may include all greenhouse gases (GHGs) identified by the Intergovernmental Panel on Climate Change (IPCC, 2011). For the purposes of this paper, the term 'carbon' is taken to be interchangeable with 'greenhouse gas', with carbon emissions measured in grams of carbon dioxide equivalent (g CO₂ eq). While this should include emissions of all greenhouse gases, the implications of only considering one or two in an estimate of life cycle carbon emissions are discussed later.

Estimates of the life cycle carbon emissions of onshore wind farms are not, in themselves, particularly useful, and are only really of interest for comparison with other forms of low-carbon generation. Further interpretation is required to calculate other values that may be more meaningful – the first of these is the lifetime emissions savings of an onshore wind farm. This is the net reduction of greenhouse gas emissions taking into account both the life cycle carbon emissions and the lifetime emissions displacement. Estimates of the latter vary widely, as they are a measure of the displaced emissions resulting from wind power replacing other forms of generation.

Another useful metric is the carbon payback period. This is an estimate of the time for the carbon emissions of an onshore wind farm to be offset by the displaced emissions. Provided that the carbon payback period is significantly shorter than the design life, a net reduction in emissions will be achieved. This value is less sensitive than the lifetime emissions savings to assumptions about design life, annual energy production and any changes in emissions displacement due to long-term network changes over time.

4.2 Calculation Methodology

The life cycle carbon emissions of wind farms are conventionally calculated using partial process-based Life Cycle Assessment (LCA), defined by a number of national and international standards (BSI, 2011; ISO, 2006a; ISO, 2006b; ISO, 2013; The International EPD Cooperation, 2008; WRI and WBCSD, 2011). This involves systematically analysing the greenhouse gas emissions of each process in each stage of the life cycle of the wind farm, as illustrated by Figure 1, and described in further detail in Appendix 1.

As the LCA method is designed to be applied to a wide range of goods and services, there is considerable scope for variations to be introduced to the results by variations in assumptions, methodological choices and data uncertainty (ISO, 2006a; ISO, 2006b). Figure 11 illustrates the key areas where such variations might be introduced to estimates of the life cycle carbon emissions of onshore wind power. These can be divided into four categories (Adams et al., 2013): the first concerns variations in the input data stemming from variations in the wind farm scenarios considered in different studies - farms built at different times, in different locations, with different equipment - along with any uncertainty in this input data; the second includes uncertainties in the emissions factors extracted from life cycle datasets, variations in processes in different countries, and differences in the scope of GHG emissions included; the third focuses on differences in physical and temporal system boundaries; and the fourth includes all variations introduced by differences in LCA methodology.

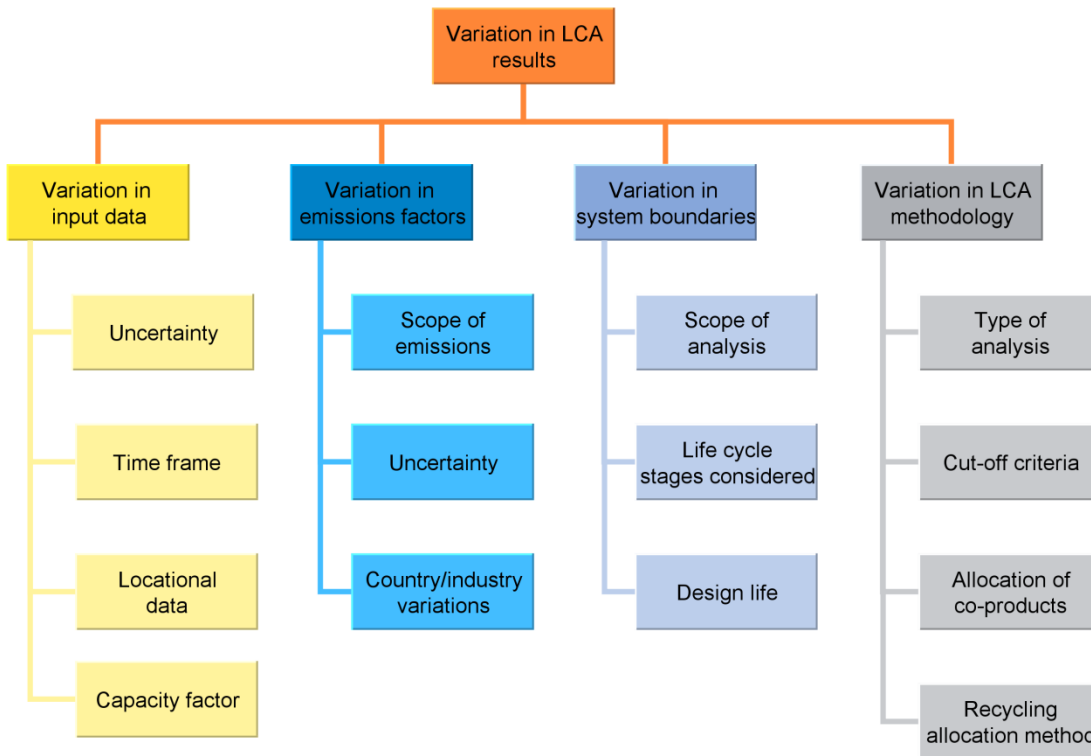


Figure 11 - Causes of variation in carbon emissions estimates of wind farms (after Adams (2013))

A brief explanation of the terms used in Figure 11 follows:

- **Uncertainty** – Often reported as a probability distribution, this includes potential measurement errors and uncertainties of assumptions; for example, at the construction stage, the design life of a wind farm is an assumption – the actual value might be different by a few years.
- **Time frame** – Wind farms built in different years will have different impacts, as designs, materials and processes evolve.
- **Location/country/industry** – Carbon emissions associated with different processes vary across industries and in different locations. Furthermore, life cycle carbon emissions are affected by transport distances, which are specific to the locations chosen in the given scenario.
- **Capacity factor** – As with LCOE, the capacity factor is an important value in estimating the carbon emissions. It is a measure of the power output of a wind farm, described as a proportion of the maximum possible output. This is typically assumed to be around 30% (see Table 6), but will vary according to the wind profile at the given installation location, and is currently estimated to average 27% for the UK, over 2000 to 2012, and 27.9% for Scotland (DECC, 2011a).
- **Scope of emissions** – Ideally an estimate of the life cycle carbon emissions will include emissions of all greenhouse gases identified by the IPCC (2007), but often only carbon dioxide, methane and nitrous oxide are included, with many studies focussing solely on CO₂ (Vestas, 2006a; Vestas, 2006b). Alternatively, some studies include the six gases identified by the Kyoto Protocol (The Carbon Trust, 2012).
- **Scope of analysis** – Different studies might set different system boundaries; it should be the point of connection with the grid, as illustrated in Figure 2 in Section 1.2, but some published analyses of onshore wind farms consider only a single turbine, rather than a complete farm (Guezuraga et al., 2012; Tremeac and Meunier, 2009). Furthermore, there is debate over whether the analysis should include the emissions associated with the life cycle of the capital goods, such as vehicles used to transport the turbines (Crawford, 2005; Goedkoop et al., 2008).

- **Life cycle stages** – Although all life cycle stages should be considered when estimating the carbon emissions of a wind farm, some studies consider stages such as maintenance or disposal more thoroughly than others.
- **Design life** – This is often taken to be the temporal system boundary of an analysis, and can vary across studies, as well as having an associated uncertainty.
- **Cut-off criteria** – The LCA standards allow cut-off criteria to be specified to exclude materials or processes with a low environmental impact from the analysis (ISO, 2006a; ISO, 2006b). Often it is time consuming to estimate the carbon emissions of small items that will not contribute significantly to the overall carbon emissions, but this practice does introduce scope for errors.
- **LCA methodology** – Life cycle carbon emissions are usually calculated with process-based LCA, but the setting of a system boundary and the application of cut-off criteria can introduce 'truncation errors' (Crawford, 2005; Crawford, 2009). Alternative hybrid methods have, therefore, been developed that combine national data from economic input-output tables (or supply-and-use tables) with detailed process information, to comprehensively consider all impacts at process level (Crawford, 2005); however, there is a suggestion that these methods might double-count the carbon emissions and result in overestimates (Davidsson et al., 2012; Wiedmann et al., 2011). The results of both process-based and hybrid analyses should, therefore, be considered.
- **Allocation** – This is the method by which emissions are divided between multiple co-products. In the case of recycling, the question is whether the recycling 'credit' should be allocated to the product that was recycled (closed-loop approximation method), or to the product made from this recycled material (recycled content method). Recycling allocation methods are described in greater detail in Hammond and Jones (2010) and Thomson (2014).

Thomson (2014) recently examined the impact of several of these potential causes of variation on the estimate of the carbon emissions of a wave energy converter, and the results are summarised in Figure 12. It can be seen that the uncertainty of emissions factor data introduces significant uncertainty to the results; however, this study also found that the variation between results using different datasets of emissions factors was fairly small, suggesting that this uncertainty can be ignored when comparing carbon emissions. In contrast, the choice of recycling allocation method may vary across studies, and Thomson found that, in the case of a wave energy converter largely constructed of highly recyclable steel, this choice could significantly affect the results: the estimate from the recycled content method was 32% higher than that calculated with the closed-loop approximation method (the original value calculated in this work included half of the recycled-content credit and half of the end-of-life-recycling credit). Unfortunately, recycling allocation methods are rarely reported in studies of the life cycle carbon emissions of wind farms, so it is difficult to assess whether this choice has a similar impact where devices include a much higher proportion of composite materials, which are not so readily recycled.

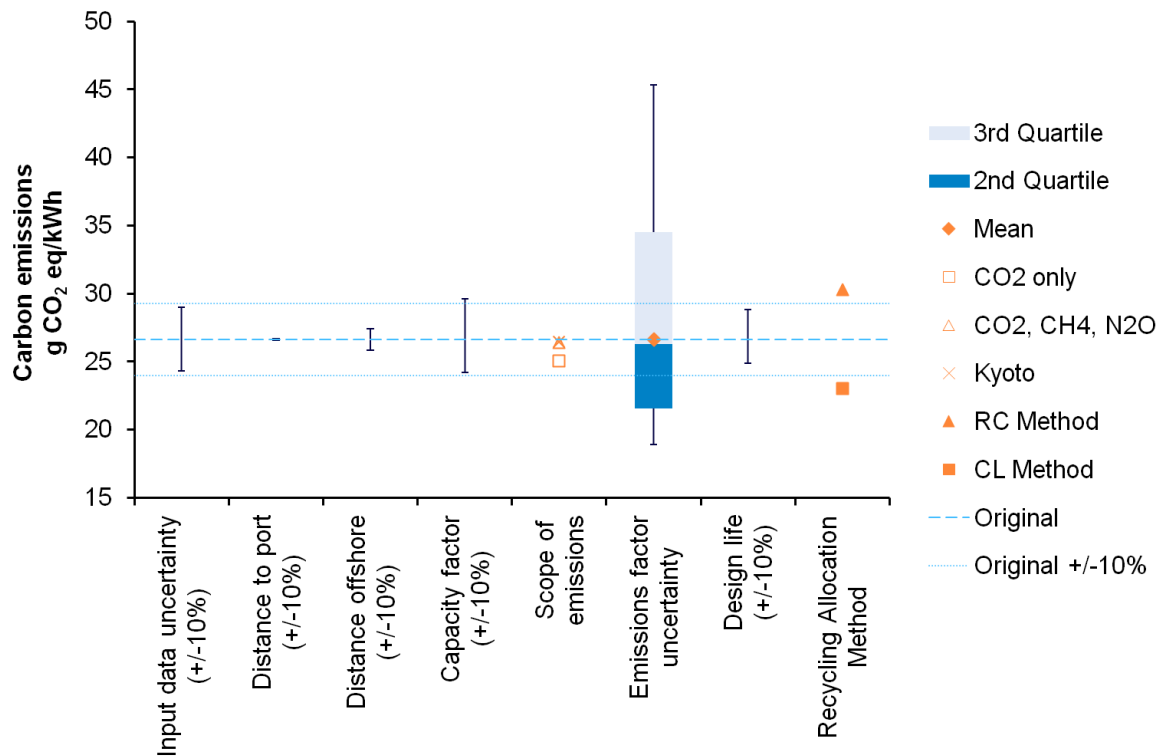


Figure 12 - Variations in carbon emissions estimates for the Pelamis Wave Energy Converter (after Thomson (2014))

Thomson also found variations in the scope of included greenhouse gases not to significantly affect the carbon emissions, with the value estimated by only considering carbon dioxide, methane and nitrous oxide being less than 1% lower than that calculated by including all greenhouse gases specified by the IPCC. Many studies of the carbon emissions of wind farms do include these three greenhouse gases as a minimum (Dolan and Heath, 2012; Guezuraga et al., 2012; Tremeac and Meunier, 2009; Vestas, 2006a; Vestas, 2006b).

The rest of this sensitivity analysis involved varying the input data (quantities of input material), distances, capacity factor and design life by $\pm 10\%$; this found that location was not a significant source of variation in the carbon emissions, but the others were. (Note that this analysis did not consider the impact of variations in the system boundary.) Recently the National Renewable Energy Laboratory (NREL) in the USA carried out a comprehensive review and harmonisation of life cycle carbon emission studies for onshore wind, and examined the impact of adjusting the published carbon emissions estimates to consistent system boundaries (in terms of major life cycle stages), emissions factors, capacity factors and system lifetimes (Dolan and Heath, 2012), concluding that variations in the capacity factor had the greatest impact on the results. It is important to note that capacity factor and system lifetime are a function of the specific design and location of each wind farm, and are therefore likely to vary between farms. Good quality studies should include a sensitivity and uncertainty analysis to test the robustness of the results and their sensitivity to variations in methodology and key assumptions (ISO, 2006a; ISO, 2006b); estimates of life cycle emissions should be presented with uncertainty ranges.

4.3 Life Cycle Impacts

Materials and Manufacture of Components

The manufacture and installation stages together account for over 90% of the total life cycle carbon emissions of an onshore wind farm (Ardente et al., 2008; Guezuraga et al., 2012; Tremeac and Meunier, 2009), with the vast majority of these emissions arising during the extraction of materials and manufacture of components (Figure 13).

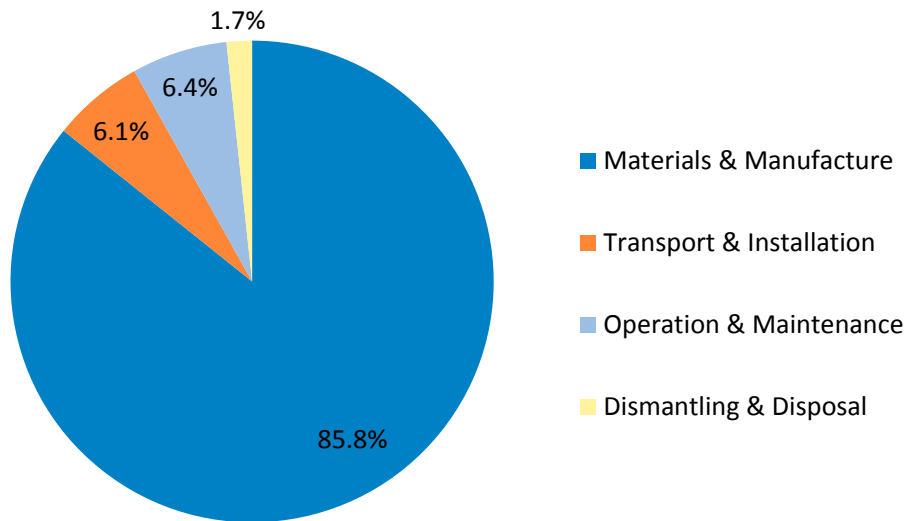


Figure 13 - Contribution of each life cycle stage to carbon emissions (from Ardente (2008))

As mentioned in Section 1.2, there is some variation in the design of wind turbines across different manufacturers, but a typical materials balance and corresponding emissions balance is shown in Figure 14 (Ardente et al., 2008). It can be seen that the materials used in the foundations and ground works dominate the mass balance, but contribute relatively little to the total carbon emissions of this stage; in this example, the turbine tower is made of steel plate, contributing over 50% to the total emissions of material extraction and manufacture.

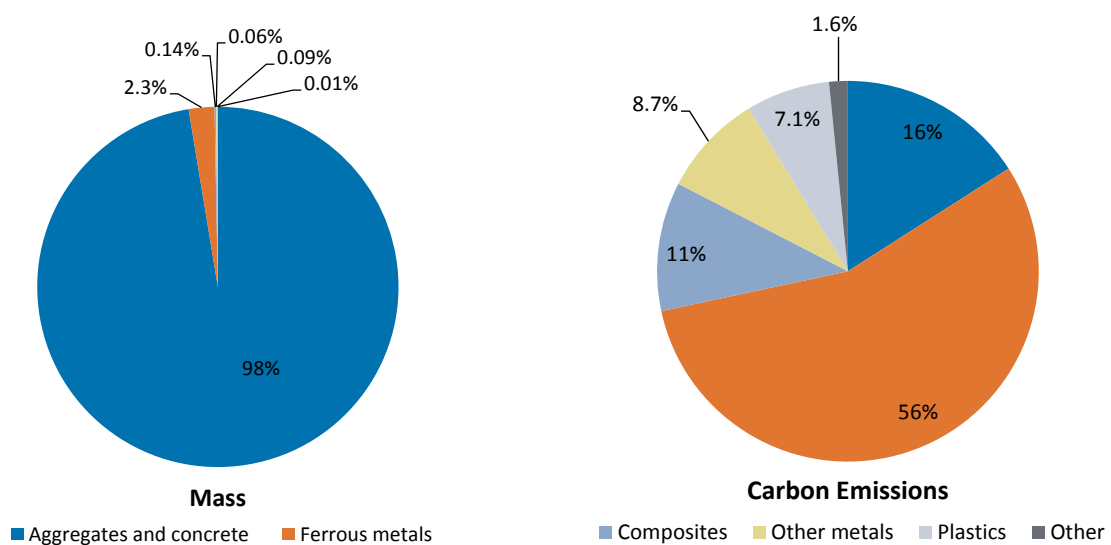


Figure 14 - Materials balance for a wind farm by mass and carbon emissions

Transport and Installation

The transport and installation stage typically contributes about 6% to the total life cycle carbon emissions of an onshore wind farm (Figure 13), although the carbon impacts of land-use change are rarely included. This is of particular consideration in Scotland where a high proportion of farms are built on peatlands, some forested, and many of the existing published studies either assume a typical installation location, so do not consider the wider implications of the effect on the land (Crawford, 2009; Tremeac and Meunier, 2009; Vestas, 2006a; Vestas, 2006b), or examine wind farms installed on clear land with other types of soil (Ardente et al., 2008; Guezuraga et al., 2012). The installation of foundations, cable ducts and access roads for a wind farm on peatlands will reduce the moisture content of the soil, and studies have shown that this will lead to a net decrease in the embodied carbon of the peat (Nayak et al., 2008) - such an impact should be allocated to the wind farm. Similarly, the impact of any forestry clearance should also be considered.

Several studies examining the impacts of installing wind farms on forested peatlands have been published (Mitchell et al., 2010; Nayak et al., 2008; Smith et al., 2014). Nayak et al, in particular, provides detailed explanations of the processes and carbon impacts of different types of construction. Analyses of wind farms on forested peatlands located in Scotland or the north of England suggest that the total carbon emissions of such farms is 4 to 7 times greater than that estimated when the impacts of peat and forestry disturbance are not considered (Mitchell et al., 2010; Nayak et al., 2008). In their comprehensive study, Nayak et al. estimated the impacts of peat and forestry disturbance to contribute 68 and 18g CO₂eq/kWh to the total life cycle emissions respectively (see Table 6). Suggestions for mitigating some of these impacts by peat and forestry restoration measures reduced the total life cycle carbon emissions by 41% to 62g CO₂eq/kWh, which is still significantly higher than other estimates, possibly limiting the viability of constructing wind farms on forested peatlands in the long term.

Operation and Maintenance

As discussed in Section 1.2, most of the carbon emissions that arise during the operational stage of a wind farm are attributable to maintenance activities - contributing around 6% to the total life cycle impacts of the wind farm (Figure 13). Maintenance activities considered in existing published studies typically include the renovation or replacement of the gearboxes, generators and transformers, although some studies also consider the replacement of blades during regular maintenance (Tremeac and Meunier, 2009).

As discussed in Section 4.2, the assumed operational capacity factor and lifetime of a wind farm can significantly influence the estimated life cycle carbon emissions. The capacity factor is the average annual power output described as a proportion of the maximum possible output if the farm is generating at its rated output for the whole year; in their review of existing carbon emissions estimates for onshore wind farms, (Dolan and Heath, 2012) found the mean assumed capacity factor to be 30%, which is slightly higher than the reported average figure for Scotland of 27.9% (DECC, 2011a), suggesting that the true carbon emissions of onshore wind in the UK might be 7% higher than the estimates provided by (Dolan and Heath, 2012).

The operational lifetime of a wind farm will vary, but most turbines have a design life of 20 years (Vestas, 2006a; Vestas, 2006b), and farms in the UK are often built on land with a fixed-term lease and planning permission of 20 years (Mathers, 2013), so this is considered a reasonable estimate.

Decommissioning

Typically, it is estimated that the decommissioning stage contributes 2% to the total life cycle carbon emissions of a wind turbine. The impacts or credits of recycling can, however, introduce some significant variation to carbon

emissions estimates at this stage, due to varying assumptions about the use of recycled materials and the recyclability of any waste.

As discussed in Section 4.2 and Appendix 1, there are different ways in which the impacts (or credits) of recycling are dealt with when assessing the life cycle emissions of a wind turbine, as recycled materials may be used in the initial manufacturing stage, and materials may also be recycled at the end-of-life. Both of these practices may affect the costs and emissions at both the manufacturing and decommissioning stages, but are open to double-counting. Furthermore, irrespective of which recycling method is used, assumptions about the end-of-life recyclability will affect the costs and emissions associated with disposal of the waste materials, as recycled material will not need to undergo waste treatment. The majority of studies do not explicitly state the emissions savings due to assumptions about recycling; however, in a one study of an onshore turbine this was examined, and it was found that the inclusion of recycling credit (using the recycled content method) decreased the overall carbon emissions by 44% (Guezuraga et al., 2012).

4.4 Summary

A range of stakeholders have examined the life cycle emissions of wind power, including (but not limited to) turbine manufacturers (Vestas, 2006a; Vestas, 2006b), wind farm operators (Vattenfall, 2013) and academics (Ardente et al., 2008; Guezuraga et al., 2012; Tremeac and Meunier, 2009). While many robust and reliable studies exist, the quality of those published does vary considerably - in a recent comprehensive review of the carbon emissions of wind power generation, researchers at the National Renewable Energy Laboratory (NREL) in the USA identified 175 studies, of which only 41% passed their basic quality screening criteria (Dolan and Heath, 2012). This summary, therefore, only examines a selection of the most robust and reliable carbon emissions estimates.

These estimates are summarised in Table 6, Figure 15 and Figure 16. It can be seen that there is considerable variation in these values; however, they are all significantly lower than estimates of the carbon emissions of gas- or coal-fired conventional generation (typically around 500 and 1000g CO₂eq/kWh respectively). The lowest estimates are published by Vestas and Guezuraga (Guezuraga et al., 2012; Vestas, 2006a; Vestas, 2006b). In the case of the Vestas analyses these low estimates can be attributed to a high assumed capacity factor, along with particular methodological choices in the analysis, such as: the exclusion of processes, like transport, that were thought to have negligible impact; the consideration of only carbon dioxide instead of all greenhouse gases; and the assumption that the emissions due to electricity consumption in manufacturing are particularly low due to the use of wind power at Vestas factories. The study by Guezuraga may have a particularly low estimate of carbon emissions because it has considered only a single wind turbine, without all of the associated ground works and transmission required for a full wind farm; however, the results of the sensitivity analysis (included in Table 6) show that these low impacts are mostly attributable to assumptions about the recycled content of the raw materials, taken to be the European average; where data for a specific country is applied the estimates are much higher. Finally, the findings of the hybrid study by Crawford are comparable to those of process-based studies (Crawford, 2009).

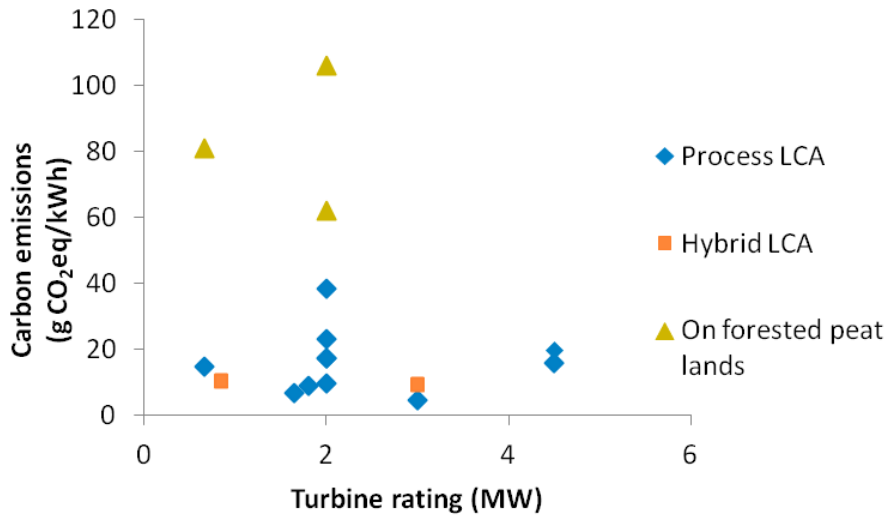


Figure 15 - Selection of carbon emissions estimates for different turbine ratings

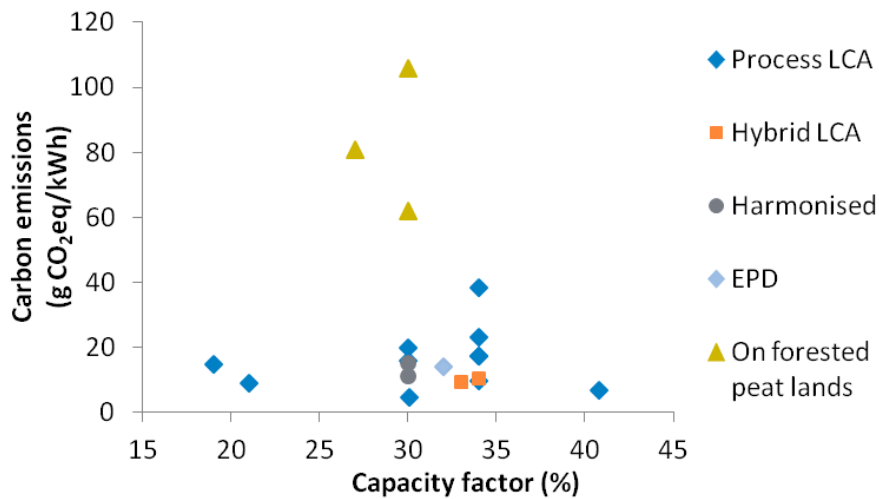


Figure 16 – Selection of carbon emissions estimates for different assumed capacity factors

Figure 15 and Figure 16 also demonstrate that the variation between studies does not show any strong relationship between carbon emissions and turbine rating or assumed capacity factor.

In order to develop a better understanding of these variations and provide a more robust estimate of the carbon emissions of onshore wind, NREL carried out a review and harmonisation of existing studies - the most comprehensive analysis of the existing body of knowledge on the carbon emissions of wind power to date (Dolan and Heath, 2012). This study included a systematic review of over 200 published estimates of the carbon emissions of wind power, to identify the most robust and reliable studies, and then carried out harmonisation to align methodological inconsistencies, with the aim of enabling comparison between studies while still maintaining variations introduced by each study's unique perspective; only key areas were harmonised, including capacity factors, system boundaries and functional units.

Life Cycle Costs and Carbon Emissions of Onshore Wind Power

Device	Rating (MW)	Capacity factor (%)	Design life (years)	Carbon emissions (g CO ₂ eq/kWh)	Type of Analysis	References
Nordic wind farms	Mixed	32 (19 – 44)	20	14	Environmental Product Declaration	(Vattenfall, 2013)
Vestas v90 farm	3.0	30.02	20	4.64	Process LCA	(Vestas, 2006b)
Vestas v82 farm	1.65	40.8	20	6.6	Process LCA	(Vestas, 2006a)
Gearless turbine - Europe	1.8	21	20	8.82	Process LCA – turbine only	(Guezuraga et al., 2012)
Geared turbine - Europe	2.0	34	20	9.73	Process LCA – turbine only	(Guezuraga et al., 2012)
Geared turbine – Europe no recycled material	2.0	34	20	17.35	Process LCA – turbine only	(Guezuraga et al., 2012)
Geared turbine – Germany	2.0	34	20	17.35	Process LCA – turbine only	(Guezuraga et al., 2012)
Geared turbine - Denmark	2.0	34	20	23.26	Process LCA – turbine only	(Guezuraga et al., 2012)
Geared turbine – China	2.0	34	20	38.33	Process LCA – turbine only	(Guezuraga et al., 2012)
Italian wind farm	0.66	19	20	14.8	Process LCA	(Ardente et al., 2008)
Turbine in France (with recycling)	4.5	30	20	15.8	Process LCA – turbine only	(Tremeac and Meunier, 2009)
Turbine in France (no recycling)	4.5	30	20	19.9	Process LCA – turbine only	(Tremeac and Meunier, 2009)
Scottish farm including peat/forestry impacts	2	30	25	106	Process LCA	(Nayak et al., 2008)
Scottish farm including peat/forestry impacts with improvement	2	30	25	62	Process LCA	(Nayak et al., 2008)
English farm including peat/forestry impacts	0.66	27	20	81	Process LCA	(Mitchell et al., 2010)
3.0MW turbine	3.0	33	20	9.3	Hybrid LCA	(Crawford, 2009)
850 kW turbine	0.85	34	20	10.3	Hybrid LCA	(Crawford, 2009)
Mean	Mixed	Mixed	Mixed	16	Review	(Dolan and Heath, 2012)
Harmonised mean	Mixed	30	20	15	Review and harmonisation	(Dolan and Heath, 2012)
Median	Mixed	Mixed	Mixed	12	Review	(Dolan and Heath, 2012)
Harmonised median	Mixed	30	20	11	Review and harmonisation	(Dolan and Heath, 2012)

Table 6 - Carbon emissions estimates for onshore wind power generation (low outliers are highlighted in blue, high outliers in orange)

The results of this analysis are shown in Figure 17. Note that these statistical results only show the distributions of existing life cycle carbon emissions estimates, and are not, therefore, indicative of their accuracy, just their comparability. It is of significance, however, that this study concluded that “the large number of previously published life cycle GHG emission estimates of wind power systems and their tight distribution suggest that new process-based LCAs of similar wind turbine technologies are unlikely to differ greatly” (Dolan and Heath, 2012).

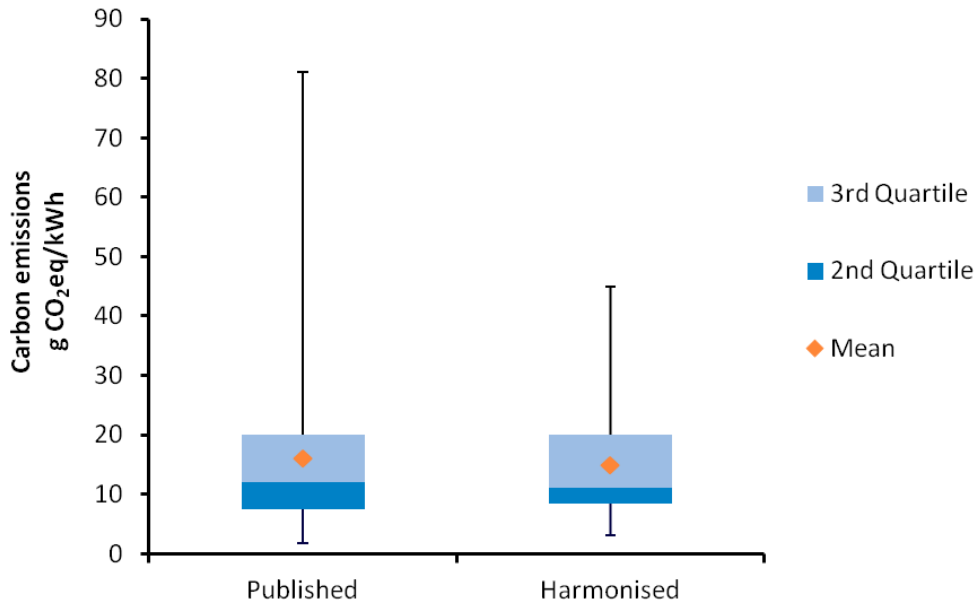


Figure 17 - Range of carbon emissions of onshore wind farms, before and after harmonisation (Dolan and Heath, 2012)

This work also found that harmonisation of the capacity factor had the most significant impact on the results, which will be important for real wind farms with their own, individual, capacity factors. NREL assumed a capacity factor of 30% for onshore wind farms, and adjusted the published estimates accordingly (Dolan and Heath, 2012). The following equation may be used to adjust the estimated carbon emissions for a known capacity factor:

$$E_{new} = \frac{E_{old} \times cf_{old}}{cf_{new}}$$

where E is the life cycle carbon emissions and cf is the capacity factor. The findings of Dolan and Heath (2012) can be adjusted for the Scottish average capacity factor of 27.9% (27% for the UK), giving mean and median estimates of 16 and 12g CO₂eq/kWh respectively. If the wind farm is built on forested peatlands this could rise significantly (see Section 4.3), and possibly be as much as 114g CO₂eq/kWh if no improvements are made following construction.

Despite these variations in estimated carbon emissions of wind power generation, it is significant to note that these are all significantly lower than for fossil fuelled generation. Figure 18 compares the values presented here with those gathered by NREL for other types of generation (NREL, 2013a; Warner and Heath, 2012; Whitaker et al., 2012), with the ranges showing the maximum range of published estimates. There is no overlap between onshore wind generation and any type of fossil fuelled generation. Furthermore, there is greater consensus on the carbon emissions of onshore wind than there is for other forms of low carbon generation, such as hydroelectricity and nuclear power. It is also worth noting that the land-use change impacts of building wind farms on forested peatlands are also applicable to other technologies, and results are rarely consistent across studies; this is

particularly relevant for hydro where the inundation of vegetation when a new reservoir is created can lead to significant greenhouse gas emissions (Vattenfall, 2011).

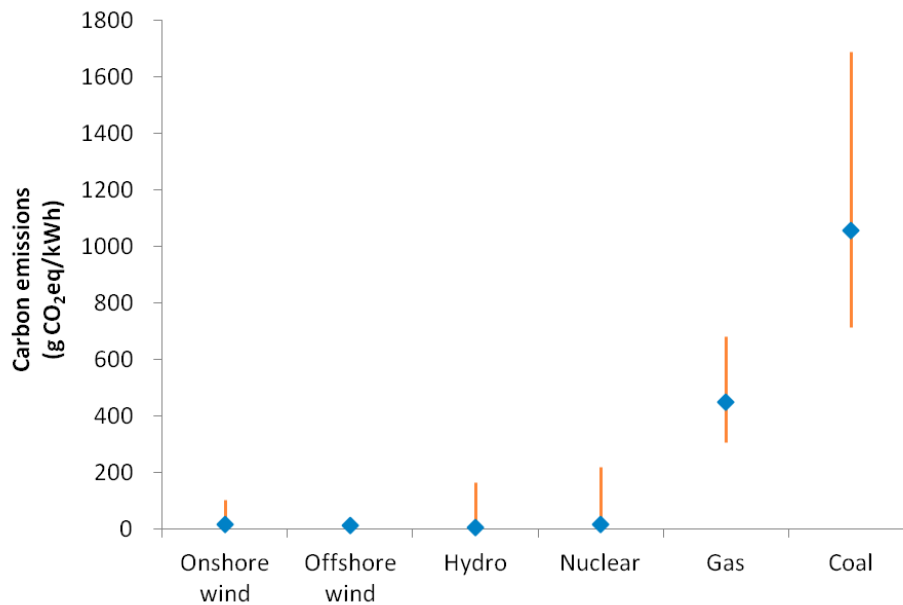


Figure 18 - Comparison of carbon emissions of onshore wind with other types of generation

Key Messages

- This report considers only a selection of the most reliable and robust published studies.
- Significant variations to carbon emissions estimates are introduced by uncertainties in raw data, assumed capacity factor and design life, and the way that recycling is dealt with, as well as the inclusion of land-use-change impacts, such as deforestation and peat disturbance.
- There is disagreement over whether process-based analyses or hybrid methods (which use input-output data) are the most reliable; however, this review has found that the results of both types of study are comparable.
- The lowest estimates are published by Vestas (2006a; 2006b) – attributable to a high assumed capacity factor, consideration of carbon dioxide only and low assumed impacts for electricity consumption.
- The highest published estimates are those that consider the impacts of land-use change on forested peat lands (Mitchell et al., 2010; Nayak et al., 2008; Smith et al., 2014).
- A comprehensive review and harmonisation of published carbon emissions estimates for onshore wind power published by Dolan and Heath (2012) found that credible estimates of the carbon emissions for onshore wind range from 3 to 45 g CO₂eq/kWh.
- The above review did not include studies that examined the impact on forested peatlands – a particular issue in Scotland. When such impacts are considered the life cycle carbon emissions for onshore wind will likely be in the range of 62 to 106g CO₂ eq/kWh.

5 System Carbon Emissions

In order to develop a more complete picture of the carbon emissions of onshore wind power, it is necessary to understand the wider implications that wind power has on the electricity network and other types of generation. Emissions arise due to the installation and/or operation of conventional generation required to complement wind power, and any carbon savings are made through wind displacing other types of generation. This section examines the wider effects of wind power on the electricity generation and transmission system over the short and long term in order to fully understand the true emissions impacts.

5.1 Carbon emissions displacement of wind power in Great Britain

As mentioned in Section 4.1, an understanding of the carbon emissions displacement of wind power is required in order to estimate both the lifetime emissions savings and carbon payback period – information that is used to support planning applications and inform government policy. Current practice in both scholarly research and policy implementation is to estimate this as the average emissions of the whole network using annual figures published by DECC/Defra, most recently 460g CO₂eq/kWh for 2012 (Ricardo-AEA, 2012; Siler-Evans et al., 2012). The use of this value is widely debated, however, as it does not reflect the fact that wind power only replaces certain types of generation (low-carbon nuclear, for example, does not respond to fluctuations in available wind energy). Also, as mentioned in Section 3, there are inefficiencies associated with operating conventional plant at lower output when the wind is blowing, which may increase the emissions intensity of energy from these generators (ASA, 2005; ASA, 2007a; ASA, 2007b).

Marginal emissions

The fluctuating output power of wind farms will displace generation from the plants operating on the margin – normally a mixture of coal- and gas-fired generation in Great Britain (confirmed by National Grid in ASA (2007b)). This marginal generating mix varies according to changes of load at different times of day and throughout the year, depending upon the relative prices of coal and gas, as dictated by the liberalised energy market. Figure 19 shows the generating mix for a windy winter's day in 2012. While it can be seen that the marginal generating mix is mostly coal and gas, it is not clear how much of the change in output from these plants is due to changes in demand, and how much is due to changes in wind power output. Furthermore, it can be seen that pumped storage hydro power is responding to a drop in wind power output in the early hours of the morning. One of the challenges of identifying a robust emissions displacement estimate for wind power is that it may not be equal to the marginal emissions of fluctuating demand; it is often assumed that the dispatchable generators on a network will respond to fluctuating supply from wind power as though these are negative fluctuations in demand (Farhat and Ugursal, 2010), but this assumption neglects the impacts of differences in forecasting accuracy for demand and wind power output, and any requirements for greater reserve capacity for fluctuating wind power generation (Thomson, 2014).

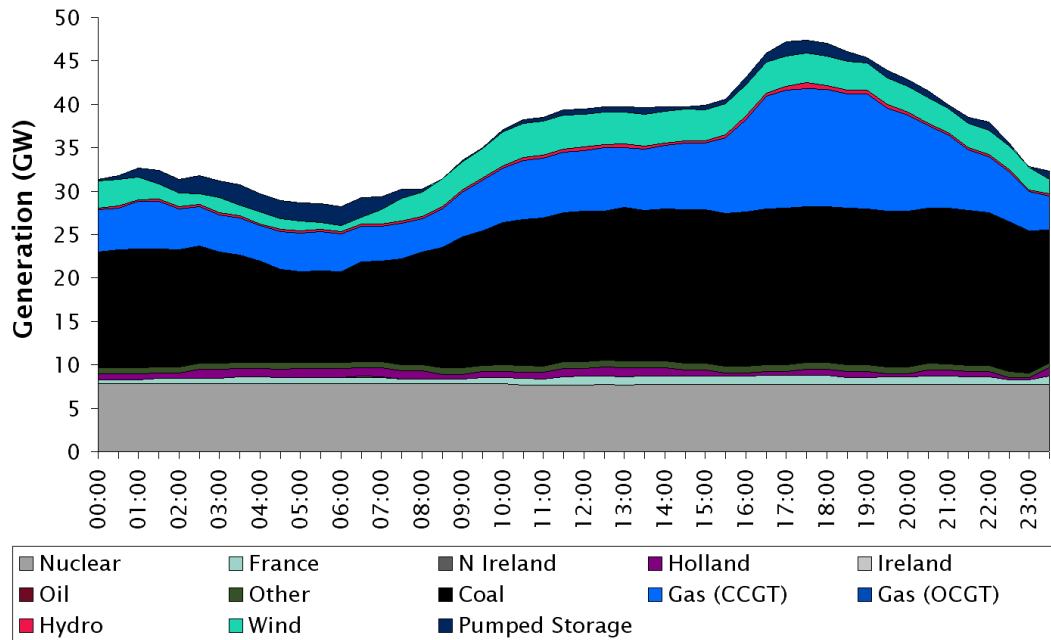


Figure 19 - Stacked demand curve for 9th December 2012 (Elxon, 2013a)

Efficiency penalties

Conventional dispatchable power stations will operate at part load in response to an increase in wind power availability, or to provide additional reserve capacity to cope with any sudden drops in wind speed. Another complexity of estimating the marginal emissions displacement of wind power is that there are significant efficiency penalties associated with operating these power stations at such reduced outputs, illustrated in Figure 20. The *emissions intensity* of these generators, therefore, *increases* due to the presence of wind generation on the network (Figure 21). This effect has led to some reports that wind power generation actually results in an increase in carbon emissions (Lea, 2012; Udo, 2011), but this is incorrect – while the carbon emissions per unit of energy output does increase, overall the total emissions still decrease with decreasing output (represented by the solid lines in Figure 22). The effect of the efficiency penalties is to decrease the magnitude of these emissions savings, and reduce the gradient of the total emissions curve; this is illustrated by the two dotted lines showing the carbon emissions if it is assumed that the power stations operate consistently at maximum efficiency. As conventional generators operate at a lower efficiency when they are part loaded, their carbon emissions aren't reduced by as much as might be expected, but *there is still an emissions saving*. Studies of the marginal emissions of networks around the world have demonstrated that the efficiency penalties do have an impact (Kaffine et al., 2011; Siler-Evans et al., 2012; Thomson, 2014; Voorspools and D'Haeseleer, 2000), and therefore must be considered in an accurate estimate of the carbon emissions displacement of wind power.

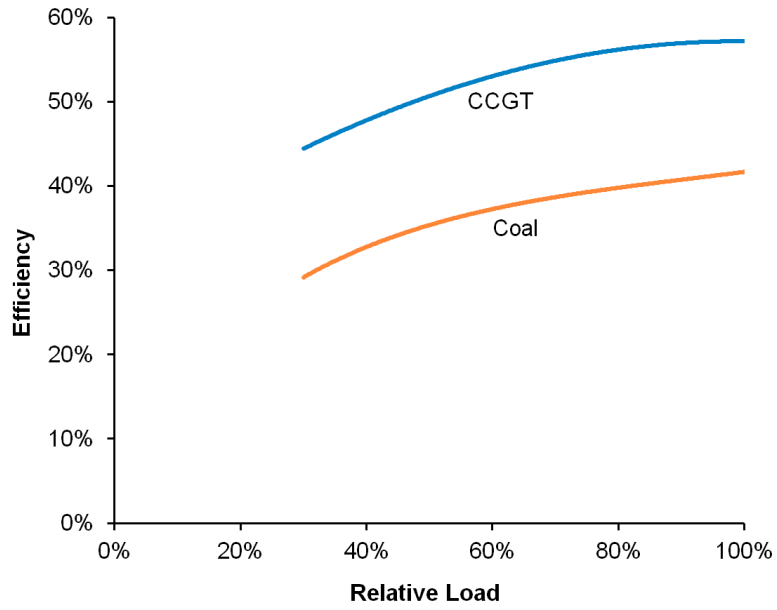


Figure 20 - Efficiency penalties of coal and CCGT power stations (Thomson, 2014)

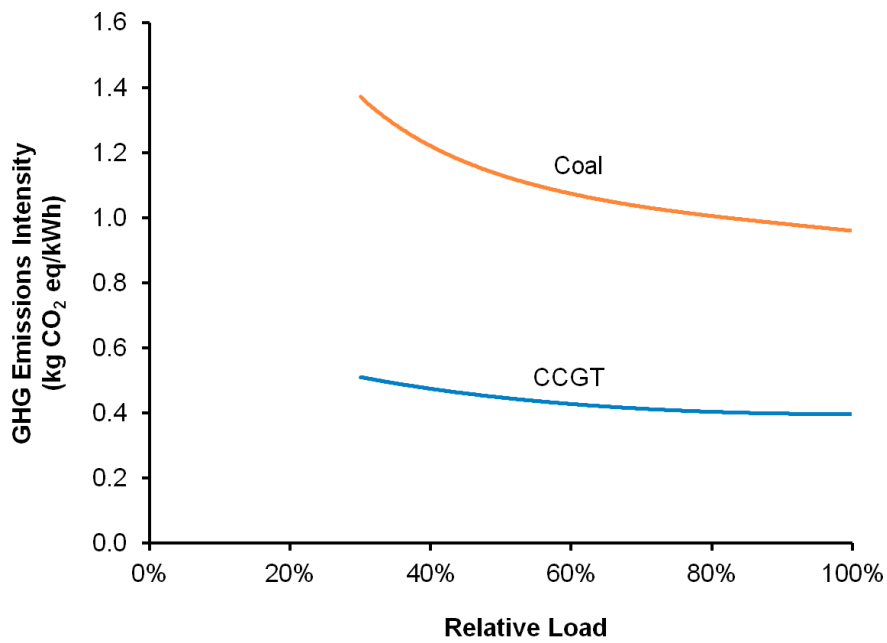


Figure 21 - Change in emissions intensities at part load (Thomson, 2014)

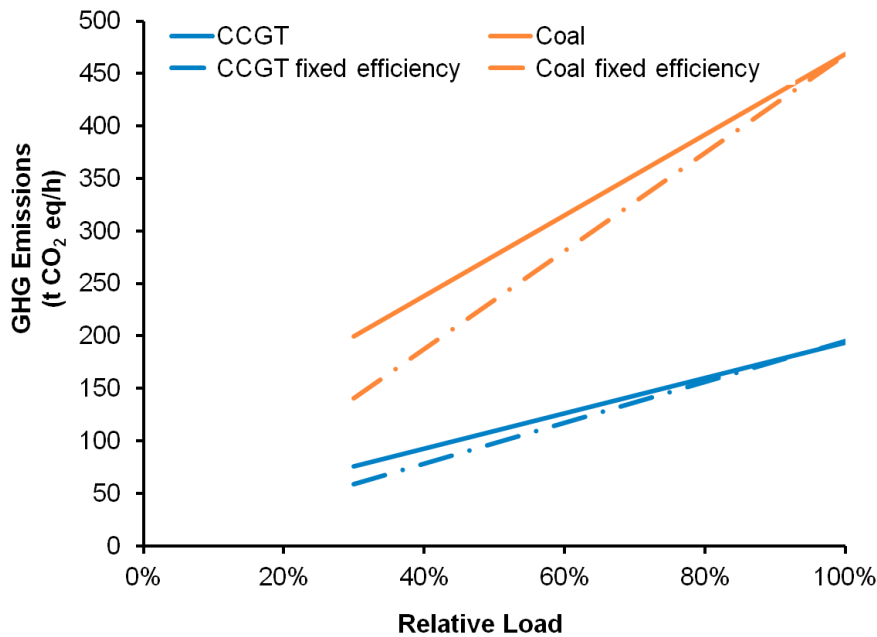


Figure 22 - Change in GHG emissions at part load for typical coal and gas-fired power plant, rated at 488 MW and 492 MW respectively

Summary of Current Research

Many studies have attempted to identify the marginal carbon emissions of networks around the world, and these have all confirmed that the marginal emissions are significantly different from the system average emissions of the corresponding networks (Bettle et al., 2006; Farhat and Ugursal, 2010; Gil and Joos, 2007; Hawkes, 2010; Marnay et al., 2002; Siler-Evans et al., 2012; Thomson, 2014). The actual values depend upon the types of generation available on the network, and the relative prices of different fuels; networks where the baseload generation is mostly low-carbon (which is the case in Great Britain, where nuclear power provides much of the baseload) generally have higher marginal emissions than their average emissions, while the marginal emissions of networks with high-carbon coal-fired plants as the baseload generators (such as the island of Ireland) may well be lower than the system-average emissions (Marnay et al., 2002; Siler-Evans et al., 2012; Wheatley, 2013). Furthermore, the marginal emissions are likely to reduce over time, as the most polluting power stations are decommissioned and replaced with lower carbon alternatives (Hawkes, 2010; Voorspools and D'Haeseleer, 2000).

No studies have been published that focus solely on the carbon emissions displacement of marginal changes of wind power in Scotland, but three studies have examined the network in Great Britain (Bettle et al., 2006; Hawkes, 2010; Thomson, 2014). Two of these focus solely on the emissions of marginal fluctuations in demand: the first (Bettle et al., 2006) is based on a theoretical order of generator dispatch, and is therefore unlikely to truly reflect the operation of the network following market liberalisation; the second (Hawkes, 2010) is based on near real-time market dispatch forecasts for each generator and is much more robust. This study concluded that the marginal carbon emissions of demand-side fluctuations were 690g CO₂/kWh, 35% higher than the corresponding system average emissions; however, it did not examine the marginal emissions displacement of fluctuating wind power output, nor did it consider the impact of efficiency penalties.

The most recent study has specifically examined the marginal emissions displacement of wind power in Great Britain, also taking efficiency penalties into account (Thomson, 2014). This was based on publicly-available

measured output data combined with some market dispatch forecasts, and found that efficiency penalties did reduce marginal emissions estimates, but that they remained higher than the corresponding system average emissions; the marginal emissions displacement of wind power was found to be **560g CO₂ eq/kWh** for November 2008 to June 2013, 9% higher than the corresponding system average emissions. Significantly, the carbon emissions associated with marginal changes in wind power output were found to be 7% lower than those for demand, demonstrating that the network does respond differently to fluctuations in supply or demand. This study also examined the annual carbon emissions, and found the marginal displacement of wind power to be **550g CO₂ eq/kWh for 2012**, a value that is 20% higher than the UK-average emissions for that year as reported by the government (460g CO₂ eq/kWh – (Ricardo-AEA, 2012)).

The findings of the study by Thomson (2014) are supported by analyses of similar networks in the USA, which found similar results using empirical emissions and power output data published by the Environmental Protection Agency (Kaffine et al., 2011; Siler-Evans et al., 2012); however, other studies have found very different findings. One such study, recently published by Wheatley for the network in Ireland, found the marginal displacement of wind power to be 280g CO₂/kWh, much lower than the system average emissions of 520g CO₂/kWh (Wheatley, 2013). Significantly, while the results of this study are of relevance when considering the marginal displacement of wind power in the whole of the UK, these findings do not conflict with those of Thomson, as the Irish network is very different to that in Great Britain and has a much higher penetration of wind power, no nuclear generators, and coal as the baseload fuel.

There are also several other reports and papers that have gained prominence in the UK media and conflict with the findings of the Thomson study. One of these is a report published by the think-tank Civitas, which suggests that wind power is not effective in cutting CO₂ emissions (Lea, 2012); firstly by highlighting that conventional generation may suffer from efficiency penalties (an issue that was addressed in the study by Thomson); and secondly by referring to a paper that found that fuel consumption of gas-fired plant increases when wind power is connected to a network (le Pair, 2011b). The latter paper, however, which was published online rather than in a peer-reviewed journal, has been widely criticised (Barnard, 2013; Carrington, 2012; Goggin, 2012; Hickman, 2012a; Hickman, 2012b; MarkR, 2012; UKERC, 2012), as it contains flawed assumptions and examines a very simplistic model that is not representative of any generation network around the world. The simplistic model used by le Pair assumes that wind power only displaces generation from combined cycle gas turbines (CCGTs) and open cycle gas turbines (OCGTs), but analyses of the real historical data from grids like that in Great Britain have found that wind power mostly displaces a combination of CCGT and coal-fired plant, with OCGT output rarely being affected, a conclusion supported by a statement from National Grid (ASA, 2007b; Kaffine et al., 2011; Thomson, 2014). Furthermore, a significant variable (demand) has been fixed, which is highly unrealistic.

Even given the limitations of this model, calculation errors in the treatment of the dynamic effects of changing wind power output on fuel consumption of the CCGT plant, and incorrect assumptions in calculating the impact of life cycle emissions on the net emissions savings, further invalidate the results. A detailed review of this paper is included in Appendix 3, and demonstrates that the correction of these errors results in a consistent **decrease** in fuel consumption and carbon emissions of the associated gas-fired plants when operating with wind power generation – completely opposite to le Pair's own conclusion (le Pair, 2011b).

The carbon emissions displacement of wind power generation in Great Britain can, therefore, be approximated by the UK-average annual emissions of the entire network, but this will underestimate the positive impacts of wind power on carbon emissions. Robust studies, such as that published by Thomson (2014), do provide insight into the true emissions displacement, but these are limited by their reliance on carbon emissions calculated from power output data, rather than metered emissions data.

5.2 Long-term impacts of infrastructure changes on carbon emissions and displacement

The review above considers only the carbon displacement in the short-term, on an hourly or minute-by-minute basis. Over the longer term, there may be additional carbon emissions from the development of other generation and infrastructure to complement an increased penetration of wind power on the network, as well as changes to the potential emissions displacement as a result of changes to the generation mix.

As discussed in Section 3, increased penetration of wind farms on the grid are likely to lead to significant upgrades to the existing network infrastructure, and their variable output will have an impact on the development and commissioning of new different types of generation required on the rest of the network (the latter is also briefly alluded to in Section 5.1 and Appendix 3, with regards to open-cycle or closed-cycle gas turbines). It is difficult to isolate which infrastructure changes are attributable to wind farm developments, and no reliable studies have been identified that consider this impact on carbon emissions. It is, however, worth noting that the impacts of transmission losses on emissions have been found to far outweigh the impacts of the construction and decommissioning of the network infrastructure itself (Harrison et al., 2010).

In contrast, there are many studies that examine the impact of long-term infrastructure changes and an increase in wind capacity on the system carbon emissions – the average emissions of generation on the whole system. Studies that examine wind power have focussed on identifying the “carbon abatement potential” of an increase in installed wind capacity; such studies model possible future scenarios (some including planned developments to other types of generation) in order to identify the marginal change in average emissions attributable to incremental increases in wind penetration, including a consideration of the efficiency penalties of cycling or increased start-up/shut-down of thermal generators (Delarue et al., 2009; Denny and O'Malley, 2006; Valentino et al., 2012). All of these studies find that average emissions decrease with an increase in wind penetration; however, the results of such studies cannot be used to estimate the future carbon displacement of wind power generation.

It is likely that large scale infrastructure changes will have a significant impact on the marginal and average emissions of the National Grid in Great Britain, and therefore on the emissions displacement of wind power. Thomson examined the trend in average emissions and marginal displacement of wind power from 2008 to 2013, and found that both of these were significantly affected by a step change in the relative prices of coal and gas towards the end of 2011, which resulted in a switch from coal to gas as the marginal generator, increasing the average emissions of the grid and **decreasing** the marginal displacement of wind power compared to previous years (see Figure 23) (Thomson, 2014). The marginal displacement estimate for 2012, however, did remain higher than the average factor published by DECC (Ricardo-AEA, 2012).

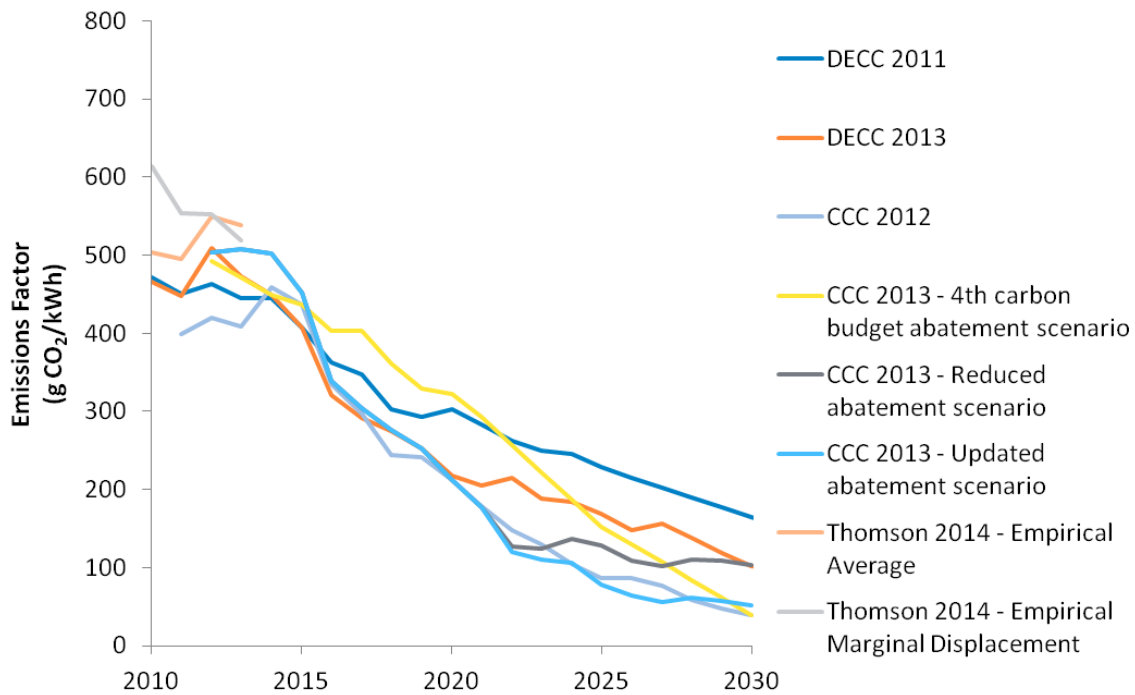


Figure 23 - Forecast trends in emission factor of the UK grid

Such fluctuations in relative prices are difficult to predict, and therefore studies attempting to identify the future average and marginal emissions of networks have concentrated on either examining historical trends (Marnay et al., 2002; Siler-Evans et al., 2012), or analysing different possible scenarios (Bettle et al., 2006; Farhat and Ugursal, 2010; Hawkes, 2010; Lund et al., 2010; Voorspools and D'Haeseleer, 2000). There have been several attempts to forecast the different scenarios towards decarbonisation of the National Grid, and the resulting average emissions of electricity generation are shown in Figure 23 (CCC, 2012; CCC, 2013; DECC, 2011b; DECC, 2013). These vary in both the assumed rapidity of decarbonisation, and in the final targets. Furthermore, they do not reflect the change in marginal emissions, although studies suggest that these will also decrease over time (Bettle et al., 2006; Hawkes, 2010); however, these studies have been focussed on marginal changes in demand, so do not take into account the possibility that networks might actively be managed so that wind power replaces the most carbon-intensive forms of generation, and that the marginal displacement of wind power may be higher.

It is likely that the planned long-term reduction in average carbon emissions of the National Grid will lead to a long-term reduction in carbon emissions displacement of wind power, in turn increasing the carbon payback period of wind power generation, and thus the pressure for wind farm construction, maintenance and decommissioning to have a low carbon footprint.

Key Messages

- Estimates of carbon emissions displacement are currently based on the average emissions of the whole network – 460g CO₂eq/kWh for 2012 (Ricardo-AEA, 2012).
- Wind power will not replace all forms of generation equally and on the GB grid marginal emissions displacement of wind is typically higher than average emissions.
- While wind power reduces the efficiency of conventional fossil-fuelled plant, it serves to reduce emissions savings a by a few percent and does *not* increase carbon emissions as some suggest (Lea, 2012; Udo, 2011).
- An influential report by Civitas (Lea, 2012) suggesting that wind power is not effective at reducing CO₂ emissions is based on flawed analysis by le Pair (2011b).
- The most reliable recent estimate for the emissions displacement of wind power in Great Britain is 550g CO₂eq/kWh for 2012 (Thomson, 2014), some 20% higher than ‘official’ estimates.

6 Payback periods and lifetime emissions savings

6.1 Carbon Payback Period

The carbon payback period is the time for the carbon emissions displaced by wind power to equal the life cycle carbon emissions of the wind farm. In order to achieve a net reduction in GHG emissions, the carbon payback period should be significantly shorter than the intended lifetime. The payback period can be calculated using the following equations (Thomson, 2014):

$$t_{C_{payback}} = \frac{E_{life}}{D_{yr}}$$

$$E_{life} = E \times W_{out}/yr \times L$$

$$D_{yr} = D \times W_{out}/yr$$

$$t_{C_{payback}} = \frac{E}{D} \times L$$

where $t_{C_{payback}}$ is the payback period, E_{life} is the total lifetime carbon emissions, D_{yr} is the annual emissions displacement, D is the emissions displacement per kWh, E is the life cycle carbon emissions per unit of energy output, W_{out} is the energy output, and L is the design life. Variations in either the calculated life cycle carbon emissions or the emissions displacement can introduce variations to the carbon payback period. Table 7 shows carbon payback estimates for a selection of studies, applying two different estimates of the emissions displacement of wind power - the UK-average emissions for 2012 (460g CO₂ eq/kWh (Ricardo-AEA, 2012)), and the marginal emissions displacement of wind power from November 2008 to June 2013 (560g CO₂ eq/kWh (Thomson, 2014)). The latter values are illustrated graphically in Figure 24 and Figure 25. It can be seen that carbon payback is achieved in all cases well within the typical design life, with all estimates falling within the first two years of operation

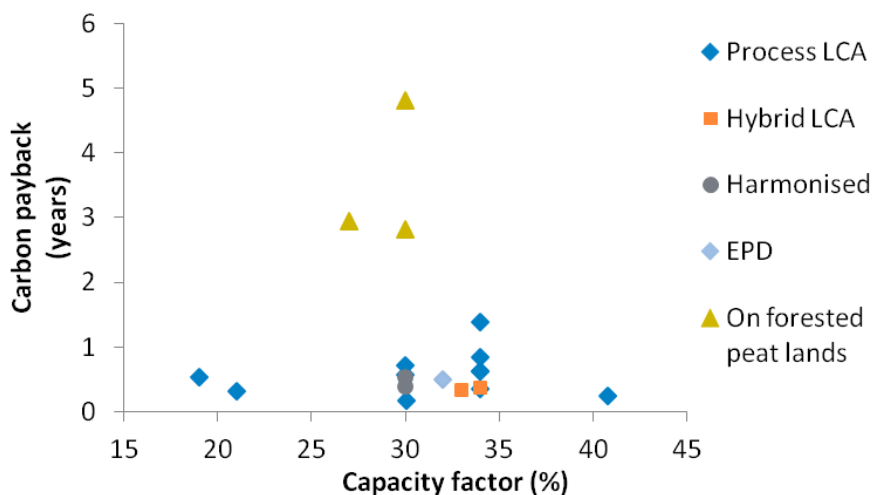


Figure 24 - Range of carbon payback estimates from selected studies

Device	Carbon Payback Period (years)		References
	Grid average for 2012	Marginal displacement for 2012	
Nordic wind farms	0.61	0.51	(Vattenfall, 2013)
Vestas v90 farm	0.20	0.17	(Vestas, 2006b)
Vestas v82 farm	0.29	0.24	(Vestas, 2006a)
Gearless turbine - Europe	0.38	0.32	(Guezuraga et al., 2012)
Geared turbine - Europe	0.42	0.35	(Guezuraga et al., 2012)
Geared turbine – Europe no recycled material	0.75	0.63	(Guezuraga et al., 2012)
Geared turbine - Germany	0.75	0.63	(Guezuraga et al., 2012)
Geared turbine - Denmark	1.01	0.85	(Guezuraga et al., 2012)
Geared turbine – China	1.67	1.39	(Guezuraga et al., 2012)
Italian wind farm	0.64	0.54	(Ardente et al., 2008)
Turbine in France (with recycling)	0.69	0.57	(Tremeac and Meunier, 2009)
Turbine in France (no recycling)	0.87	0.72	(Tremeac and Meunier, 2009)
Scottish farm including peat/forestry impacts	5.76	4.82	(Nayak et al., 2008)
Scottish farm including peat/forestry impacts with improvement	3.37	2.82	(Nayak et al., 2008)
English farm including peat/forestry impacts	3.52	2.95	(Mitchell et al., 2010)
3.0MW turbine	0.40	0.34	(Crawford, 2009)
850 kW turbine	0.45	0.37	(Crawford, 2009)
Harmonised mean	0.65	0.55	(Dolan and Heath, 2012)
Harmonised median	0.48	0.40	(Dolan and Heath, 2012)

Table 7 - Carbon payback estimates for onshore wind farms (low outliers are highlighted in blue, high outliers in orange)

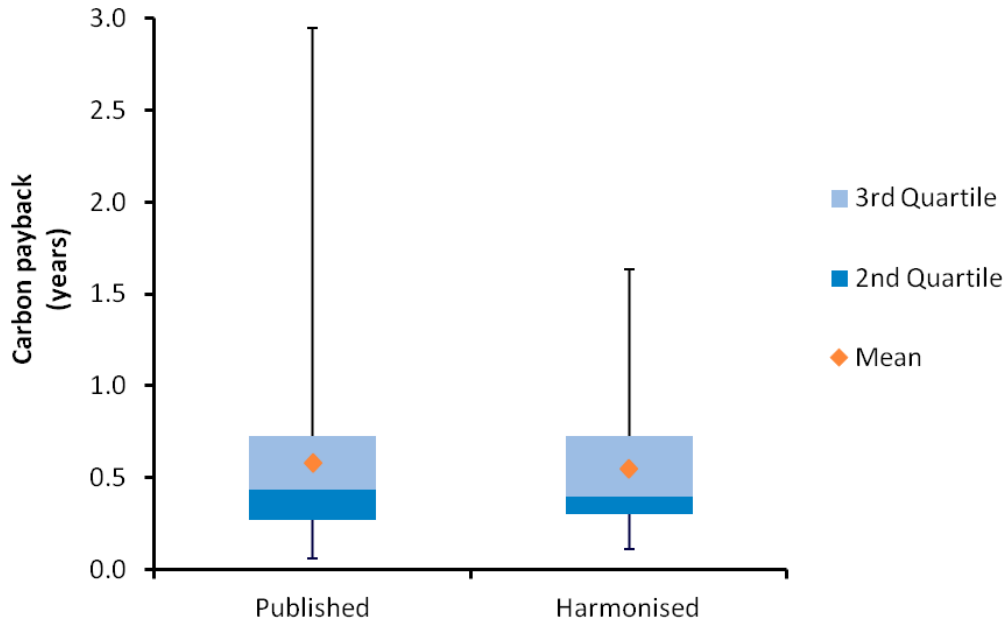


Figure 25 - Range of payback period estimates from LCA Harmonization Project (NREL, 2013a)

The longer the carbon payback period, the greater its uncertainty, as the emissions displacement of wind power on the grid is likely to decrease over time (see Section 5.2), dependent upon the design and operation of the system, and the relative prices of different fuels. (Smith et al., 2014) examined this problem when considering the payback period of onshore wind farms on forested peatlands, and identified that carbon payback will be achieved as long as the lifetime average emissions displacement of the wind farm is greater than its own carbon footprint. Figure 26 shows the lifetime average emissions factor for wind farms with a design life of 20 years, constructed between 2010 and 2050, based on the most recent DECC forecasts (DECC, 2013). It can be seen that the majority of the estimates of carbon emissions included in Table 6 fall below this line and will achieve carbon payback; however, the highest three, which correspond to wind farms constructed on forested peatlands, will not achieve carbon payback if they are constructed after 2022-2028. This is of consideration for wind farms currently being planned, and highlights the importance of ensuring that peatlands are disturbed as little as possible.

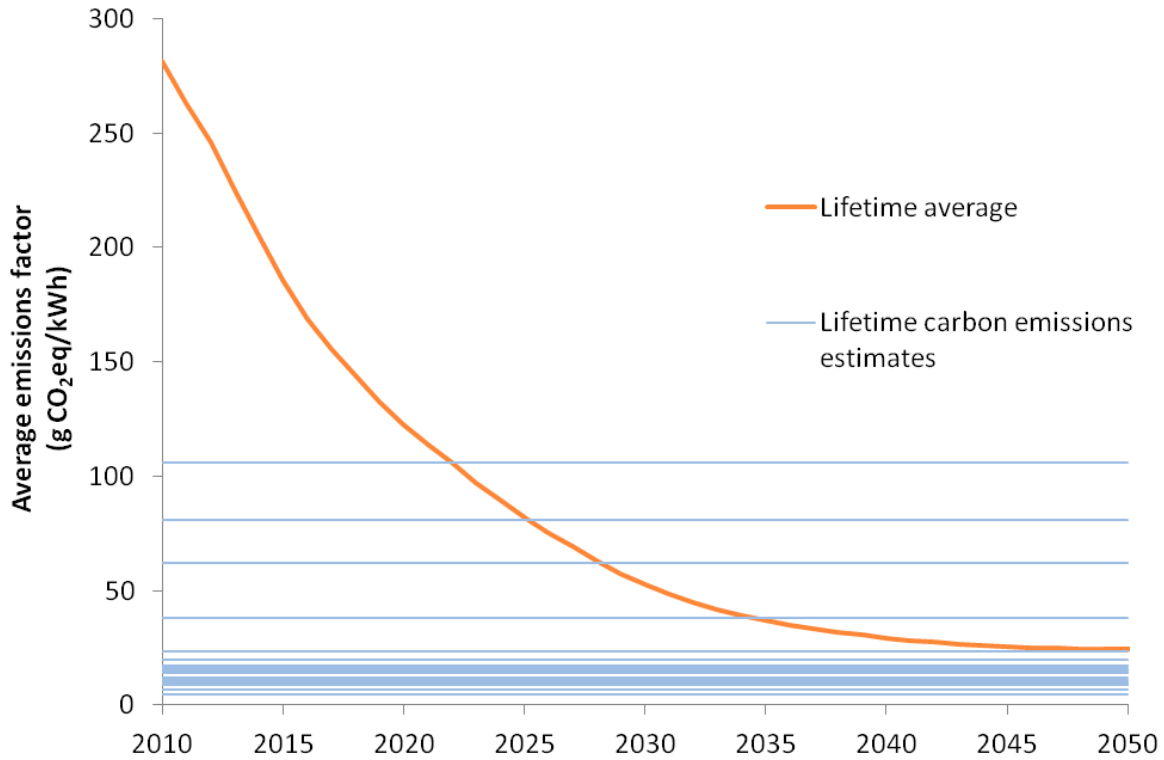


Figure 26 - Carbon payback thresholds for wind farms constructed in the future

6.2 Net Lifetime Emissions Savings

If a wind farm achieves carbon payback, it will achieve a net saving in carbon emissions over its lifetime. One of the advantages of calculating the carbon payback period instead of the lifetime emissions saving is that it is less susceptible to long-term changes in emissions displacement, particularly if it is short, as well as being less sensitive to assumptions about design life and annual energy production. The lifetime emissions savings, however, are of interest for headlines on the benefits of wind power generation. They can be calculated as follows:

$$S_{life} = \overline{D}_{yr} \times L - E_{life}$$

$$\overline{D}_{yr} = \overline{D} \times W_{out}/yr$$

$$E_{life} = E \times W_{out}/yr \times L$$

$$S_{life} = \overline{D} \times W_{out}/yr \times L - E \times W_{out}/yr \times L = (\overline{D} - E) \times W_{out}/yr \times L$$

where S_{life} is the lifetime emissions savings, \overline{D}_{yr} is the average annual emissions displacement, L is the design life, E_{life} is the total lifetime carbon emissions, \overline{D} is the average emissions displacement per kWh, W_{out} is the energy output, and E is the life cycle carbon emissions per unit of energy output.

It can be seen that the estimate of average emissions displacement will have a significant impact on the lifetime savings. Currently, it is common practice to estimate the net emissions reduction by assuming that the displaced emissions will be equal to the most recently published figures by the UK government (Ricardo-AEA, 2013), as illustrated by the values in the second column of Table 8, although this may be an underestimate because wind power will actually displace the marginal generation – as represented by the third column (Thomson, 2014). In

reality, both of these are likely to be significant overestimates. As discussed in Section 5.2, the average emissions of the National Grid are expected to decrease, and therefore it is likely that the marginal displacement of wind power will also decrease. Assuming a lifetime of 20 years, and calculating the lifetime average emissions based on the latest DECC forecasts (DECC, 2013), the net emissions savings estimates for wind farms constructed in 2015, 2020 and 2030 are shown in columns 3, 4 and 5 respectively. It can be seen that these are considerably lower than might be expected from current figures, although a net emissions saving is still achieved (except in the case of wind farms constructed on forested peatlands in 2030). The considerable variation in emissions savings from one carbon footprinting study to the next is largely due to the different sizes of wind turbine, as illustrated in Figure 27.

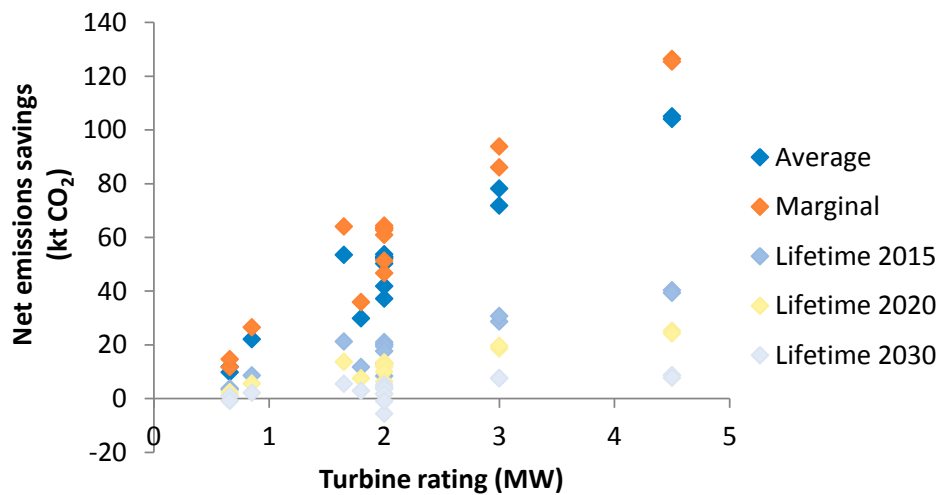


Figure 27 - Net emissions savings as a function of turbine rating

Although Figure 27 highlights the uncertainty of estimating the net emissions savings over the entire lifetime of a wind farm, it also demonstrates that wind power does reduce carbon emissions, and new farms will continue to do so. An analysis has been carried out on the existing installed wind capacity in Great Britain, using metered output data, and has shown that the emissions displacement between November 2008 and June 2013 offset all of the lifecycle carbon emissions of the existing farms, and was responsible for a further emissions reduction of 18 – 20 Mt CO₂eq (Thomson, 2014). The only restriction on this is that efforts must be made to minimise the carbon impacts of construction of wind farms on forested peatlands (the latter impacts were not included in the study by Thomson).

Key Messages

- In order to achieve a net reduction in carbon emissions, the carbon payback period should be significantly shorter than the intended wind farm lifetime (typically 20 years).
- Harmonised estimates for onshore wind range from 6 months to 2 years. When the impacts of construction on forested peatlands is considered, this can be as high as 5 years 9 months, based on 2012 values.
- When expected decrease in grid-average emissions is taken into account, most current lifecycle emissions estimates indicate payback will be achieved within the farm lifetime up to construction in 2050. The exceptions include a particularly carbon intensive wind turbine constructed in China (Guezuraga et al., 2012), and wind farms constructed on forested peatlands.
- Efforts must be made to minimise the carbon impacts of construction of wind farms on forested peatlands.

Device	Net lifetime emissions savings per turbine (kt CO ₂)					References
	Displacement factor:					
	UK average 2012	Marginal Displacement 2012	Lifetime average 2015	Lifetime average 2020	Lifetime average 2030	
Vestas v90 farm	72	86	28.6	18.6	7.6	(Vestas, 2006b)
Vestas v82 farm	53	64	21.2	13.7	5.4	(Vestas, 2006a)
Gearless turbine - Europe	30	36	11.7	7.5	2.9	(Guezuraga et al., 2012)
Geared turbine - Europe	54	64	21.0	13.4	5.1	(Guezuraga et al., 2012)
Geared turbine – Europe no recycled material	53	63	20.1	12.5	4.2	(Guezuraga et al., 2012)
Geared turbine – Germany	53	63	20.1	12.5	4.2	(Guezuraga et al., 2012)
Geared turbine - Denmark	52	63	19.4	11.8	3.5	(Guezuraga et al., 2012)
Geared turbine – China	50	61	17.6	10.0	1.7	(Guezuraga et al., 2012)
Italian wind farm	10	12	3.8	2.4	0.8	(Ardente et al., 2008)
Turbine in France (with recycling)	105	126	40.3	25.2	8.7	(Tremeac and Meunier, 2009)
Turbine in France (no recycling)	104	125	39.3	24.3	7.7	(Tremeac and Meunier, 2009)
Scottish farm including peat/forestry impacts	37	47	8.4	1.7	-5.6	(Nayak et al., 2008)
Scottish farm including peat/forestry impacts with improvement	42	51	13.0	6.4	-1.0	(Nayak et al., 2008)
English farm including peat/forestry impacts	12	15	3.3	1.3	-0.9	(Mitchell et al., 2010)
3.0MW turbine	78	94	30.7	19.6	7.5	(Crawford, 2009)
850 kW turbine	22	27	8.6	5.5	2.1	(Crawford, 2009)

Table 8 - Estimates of lifetime emissions reduction of some typical onshore turbines (low outliers are highlighted in blue, high outliers in orange)

7 Conclusions

While levelised cost and life cycle carbon analysis is well established with clear methodologies, there is scope within these to create quite large variations in headline figures for onshore wind power.

For levelised cost of energy estimates, the most important factors are capital cost of turbines, capacity factor of onshore wind, and financing assumptions – specifically the discount rate. There are a substantial number of different ways of expressing these key factors, which makes comparison challenging. Although a full harmonisation of factors was not carried out in this review, cost estimates were corrected for exchange rate and inflation changes to allow comparison. There was a modest range of estimates of capital cost, although this review found that recent studies for DECC and the CCC are broadly comparable, and that UK estimates are average internationally. In levelised cost estimates there was again a spread, but with UK-focused studies showing costs that were higher, on average, than international work. This can largely be attributed to the use of simplified LCOE methodologies and above average discount rates. Several cost estimates were much higher than others, and these were found to be largely down to the inclusion of ‘system costs’.

It is customary for levelised cost of energy calculations to not include ‘system effects’. The review found that system costs arising from accommodating wind (balancing, provision of backup and transmission) do exist, but at relatively modest levels that are not prohibitive when compared to overall costs of delivering electricity supplies, or the levelised cost of wind power itself. The review found that studies that did include system costs in the levelised cost analysis tended to systematically overestimate them, and suffered from a number of methodological flaws. There is no issue methodologically in including system costs, but credible approaches must be applied and other factors that are also not generally included in LCOE, such as external costs, should also be considered.

At present, the range of levelised costs for onshore wind is broadly comparable to nuclear and combined cycle gas turbines, and substantially cheaper than offshore wind. Future prospects for cost reduction exist for onshore wind, although there is some uncertainty as to the extent of reductions that are actually possible.

For life cycle carbon emissions the most important aspects are the materials used in manufacturing, the wind farm capacity factor, the approach to recycling credits, and the uncertainty in emissions factors, as well as emissions associated with constructing wind farms on peatlands. Extensive use was made of a recent harmonisation project which allowed different studies to be compared. While there is uncertainty associated with estimates of lifecycle emissions, onshore wind is substantially lower than unabated gas and coal generation, and there are fewer inherent uncertainties than nuclear.

Lifecycle carbon emissions also generally exclude ‘system effects’ but the literature showed that, although efficiency penalties associated with operating thermal generation at less than full load exist, the effect is modest. Wind generation is, therefore, effective at displacing fossil fuelled generation and reducing emissions; carbon payback periods are typically less than a year, although construction on peatland can extend this to several years. Long term, the expectation is that wind will remain effective at reducing emissions even within an electricity system undergoing major decarbonisation. Furthermore, while wind farms constructed on peatlands may soon reach the point that they will not be effective in reducing emissions, careful management of the land and mitigation of construction impacts can push this point some distance into the future.

Appendix 1 Life Cycle Assessment

The life cycle assessment process, as defined by ISO 14040 (2006a) and 14044 (2006b), is illustrated in Figure 28. The complete methodology is designed to consider a wide range of environmental impacts, with carbon emissions described under the single impact category of 'global warming potential'.

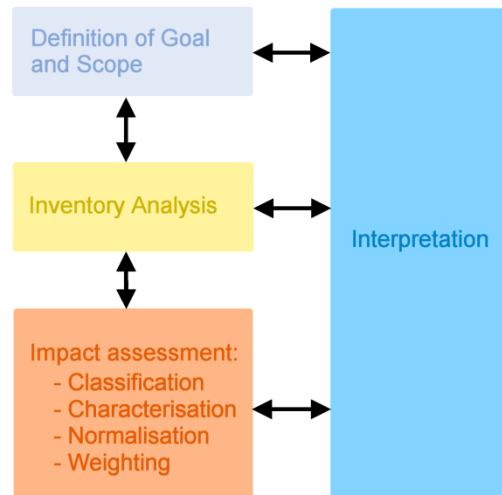


Figure 28 - Life cycle assessment framework

The clear definition of a goal and scope is an integral part of any LCA, and allows the context and purpose of the study to be defined, and the system boundary and functional unit to be identified. For analyses of the carbon emissions of wind power the functional unit is 1kWh of output energy, but the system boundary may vary between studies. A clear definition of the system boundary is important, as it defines which processes will be included in the analysis and the level of detail to which they will be studied (ISO, 2006b).

The system boundary for an onshore wind farm is illustrated in Figure 2 in Section 1.2; it can be seen that this includes the turbines, foundations, equipment to collect, transform and export the power to shore and onshore cables to the grid connection point. As well as defining the physical boundary of the analysis, the system boundary definition should also clearly state any geographical and temporal limits; such as the location of the wind farm, its age, and the year of the study.

Two calculation stages follow the definition of the goal and scope; firstly, a Life Cycle Inventory (LCI) is created detailing all relevant resource consumption and pollutant emissions over every stage in the life cycle; secondly, a Life Cycle Impact Assessment (LCIA) is carried out to make these results more understandable and environmentally relevant. When considering carbon emissions, the LCI is a detailed list of all greenhouse gas emissions, such as carbon dioxide, methane, sulphur hexafluoride and perfluorocarbons; in itself, such a list is not very informative, although it may be used to demonstrate the emissions of the six greenhouse gases defined by the Kyoto Protocol (The Carbon Trust, 2012). The LCIA allows the potential impacts of the emissions listed in the LCI to be calculated: characterisation factors are applied so that all relevant emissions can be reported in terms of their global warming potential in kilograms of carbon dioxide equivalent. The underlying physical mechanisms of global warming have been studied in detail, and there is, therefore, general agreement that the characterisation factors (or global warming potency) of the different greenhouse gases are those reported by the IPCC (2007). (There will be some variation in studies carried out over time, as the IPCC continues to review and update reported values.)

Discrepancies can be introduced to the results of an LCA through variations in assumptions and detailed methodology. One issue of particular contention is the choice of process-based, economic input-output or hybrid

methodologies. Process-based LCA is the methodology most clearly described by ISO 14040 and 14044, and involves systematically analysing the emissions associated with each process involved in the product life cycle (ISO, 2006a; ISO, 2006b). This process is limited by the application of a system boundary, however, and therefore wider impacts associated with the transport of staff to and from their workplace, or the impacts of ancillary services such as human resources, legal departments and insurance, are ignored (Baumann and Tillman, 2004). An alternative option for such analyses is to use economic input-output methodologies, which are usually based on input-output tables developed from national average statistics that model the financial flows between sectors of the economy (Crawford, 2005). The challenge of using such data for individual processes, such as the life cycle of a wind farm, is that they are generally at industry level, so hybrid methods have been developed that combine the detail of process-based methodologies with data derived from input-output tables. This should encompass all of the impacts associated with a given process, and thus eliminate any cut-off imposed by the system boundary (Crawford, 2005; Lenzen and Munksgaard, 2002; Wiedmann et al., 2011); however, there is some suggestion that double-counting of impacts is a problem, and that the results of hybrid studies may be significant over-estimates (Davidsson et al., 2012).

Another significant variation in LCA methodology, particularly for process-based analyses, is the choice of recycling allocation method. The carbon emissions associated with the manufacture of a product, such as a wind turbine, may be reduced by both using recycled materials in the manufacturing stage, and recycling any waste materials at the end-of-life. Allocating an emissions reduction, or credit, to the wind turbine for both of these activities may result in double-counting (waste steel from one wind turbine may be recycled and used in the manufacture of another wind turbine) (Hammond and Jones, 2010; Thomson, 2014). There are, therefore, two principal recycling allocation methods recognised by the literature: the recycled content method and the closed-loop approximation method (ISO, 2013; WRI and WBCSD, 2011). The former allows for recycling credit only to be applied at the manufacturing stage for the use of recycled materials; at the end-of-life stage the recycling of waste material only results in avoiding the emissions associated with long-term disposal of this waste. The latter is a method for considering only the emissions reductions associated with end-of-life recycling, and requires that no credit is given for recycled material is used in the manufacture of the turbine. (A third method, the 50:50 method, is sometimes used. This was developed to encourage both the use of recycled material and the practice of designing for greatest end-of-life recyclability, and simply takes 50% of the credit from the recycled content method, and 50% of the credit from the closed-loop approximation method (Hammond and Jones, 2010). It was applied in the study by Thomson (2014), referenced in Section 4.2 .)

Appendix 2 Design of Wind Turbines and Farms

The principal components of a wind turbine are illustrated in Figure 3 in Section 1.2. Energy is extracted from the wind by rotor blades, attached to the hub - this assembly is referred to as the turbine or rotor. The hub is mounted on the front of the nacelle, with the shaft of the turbine connected to a generator inside the nacelle, normally via a gearbox that allows the turbine's speed of 6-30 rpm to be converted to 1500 rpm for the generator. The generator produces electricity at 400 - 1000V, which is transformed to the local transmission voltage before being distributed. The nacelle also contains the yaw system, hydraulic systems and cables. The main supporting structure is the tower, and a gearing system between the nacelle and tower allows the turbine to be automatically turned towards the prevailing wind (Vattenfall, 2013). Depending upon the precise design of the turbine, electrical switchgear and a transformer may be located in the nacelle, the tower or in a nearby building (Ardente et al., 2008; Vestas, 2006a; Vestas, 2006b).

Onshore turbine towers are mounted on heavy reinforced-concrete foundations, which are normally at least 3 times the mass of the above-ground structure (Vattenfall, 2013; Vestas, 2006a; Vestas, 2006b). In addition to the foundations, access roads, working areas and turning areas are required. The turbines are all connected to the distribution grid at a substation, with cables typically laid in the ground. Some studies include the distribution transformer in the system boundary of the wind farm, while others consider it part of the existing network (Ardente et al., 2008; Vestas, 2006b).

Appendix 3 Review of “Electricity in the Netherlands – Wind turbines increase fossil fuel consumption & CO₂ emission”, C. le Pair, 2011

This paper (le Pair, 2011b), published online rather than in a peer-reviewed journal, has had a significant impact on the wind farm debate in the UK, particularly due to its citation in a report by the think tank Civitas (Lea, 2012). The analysis is, however, flawed, and is not representative of the situation in Great Britain. A detailed review of this paper, with reference to the National Grid in Great Britain, is provided here. (Note that the situation in Northern Ireland is different – as this network has a very different topology.)

A 3.1 Fuel Mix

The first assumption that is used as a basis for much of the analysis within this paper is that, when examining the generation displaced by wind, “coal and nuclear plants are almost irrelevant ... as they cannot be ramped up and down sufficiently fast to follow wind variations” (le Pair, 2011b). This statement is incorrect for the network in Great Britain; while coal and nuclear plants cannot be ramped quickly to follow very short-term fluctuations in wind power output, they are able to follow forecast changes. On the National Grid, it has been observed that it is not cost-effective to have nuclear power, which has particularly slow ramp rates, from following fluctuations in demand or wind power output, but coal-fired plants do decrease their output in response to increases in wind power output. This is supported by the findings of analyses of measured historical data for the British grid and similar American networks by Thomson (2014) and Kaffine (2011), and also by a statement by National Grid (ASA, 2007b).

A further limitation with comparing this modelled system with wind power in Great Britain is the assumption that rapid response to wind power fluctuations is provided by OCGT plants. An observation of real metered operational data from the National Grid has shown that OCGTs are rarely used, typically only over very short periods (observed to be around 1 hour every 2 or 3 days) to make up shortfalls between demand and supply during the morning pickup and evening peak (Elexon, 2013a). The annual output of OCGT plants has remained roughly constant from 2009 to 2012, despite a significant increase in installed capacity of wind power, with OCGT meeting only 0.008% of demand in 2012, compared to 4% from wind power (OCGT generation was 0.20% of the total output from wind power) (Elexon, 2013a). The assumptions made by le Pair are summarised in Table 9, and it can be seen that these assume a much greater penetration of OCGT on the network, with OCGT output corresponding to over 70% of that from the wind farm.

Wind Capacity (MW)	100	200	300
Wind Production (MWh/day) (le Pair, 2011a)	448.7	897.4	1346.1
Wind Penetration	3.7%	7.5%	11.2%
Proportion of gas generation that is OCGT (le Pair, 2011a)	3%	6%	10%
OCGT Penetration	2.9%	5.6%	8.9%
OCGT as proportion of wind penetration	77.2%	74.2%	79.1%

Table 9 - Assumed Penetration of OCGT Plant in (le Pair, 2011b)

A 3.2 Modelled System

The system examined by le Pair consists of a 500MW combined-cycle gas turbine (CCGT) plant, connected to some open-cycle gas turbines (OCGTs), and a wind farm. Three different sizes of wind farm are modelled, with installed capacities of 100MW, 200MW and 300MW. It is assumed that the demand is **constant** at 500MW (le Pair, 2011b).

This model is extremely simple, and also quite unrealistic. Firstly, on a system such as the National Grid, it is a requirement that enough backup capacity is provided to cover failure of the largest plant. In the case of this model, another 500MW of backup generation is required to be able to cope with failure of the CCGT plant, and to failover to this plant almost instantly. In the UK, pumped storage hydro plant and other fast responding generators are kept in reserve to back up failure of the largest connected generator, which is currently larger than any wind farm (Elexon, 2013b).

The second over-simplification is that only one conventional plant responds to the fluctuations in wind power output from the connected wind farm. National Grid states that “it is a property of the interconnected transmission system that individual and local independent fluctuations in output are diversified and averaged out across the system” (National Grid, 2011). Therefore, in a real network like the National Grid, the fluctuations of wind power output are compensated for by using the most cost-effective generation or demand-side response at that time; this is likely to be a combination of different plants, particularly for larger wind farms – a conclusion that is supported by the findings of Thomson (2014) and Kaffine (2011).

Thirdly this analysis assumes that the output of the CCGT plant is constant when there is no wind power on the network, by assuming a fixed demand. This is, again, a simplification that makes the model quite unrealistic, as real demand is constantly fluctuating, and the generators that respond to fluctuating wind power output will be the same ones that usually respond to fluctuating demand.

This model is, therefore, very simplistic and not at all representative of the network in Great Britain. Dr Robert Gross of the UK Energy Research Centre highlights that “Extreme estimates usually result from flawed or overly simplistic methodologies, [and] unrealistic assumptions...” (Hickman, 2012a).

A 3.3 Cycling and Hysteresis

The analysis by le Pair applies a “quasi-stationary” model to account for the impact on fuel consumption of the reduced efficiency of CCGT plant at part load, similar to that applied by Thomson (le Pair, 2011b; Thomson, 2014). This applies efficiency curves calculated from heat-rate curves to determine the total fuel consumption; however, the le Pair model assumes that the plant will always ramp up or down at its maximum ramp rate, which is unlikely to be the case. With an assumed maximum ramp rate of 12 MW/min for the CCGT generator, the calculation by le

Pair assumes that a change from 500MW to 400MW in a half hour period would result in the ramp occurring over 8min 20sec, with the output then remaining constant at 400MW for 21min 40sec. In reality, if a generator is required to change its output it will do so at any ramp rate up to the maximum possible – so may take the full half hour and ramp at 1.7MW/min. The assumption by le Pair may actually underestimate the true fuel consumption when ramping down, and, conversely, overestimate the fuel consumption when ramping up, as illustrated by Figure 29.

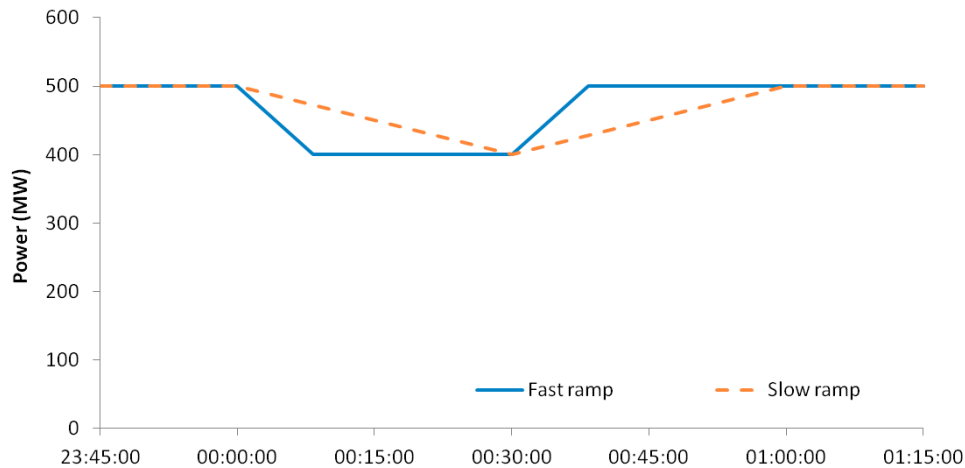


Figure 29 - Comparison of different ramp rate assumptions. The fuel consumption is a function of the area under the line, adjusted for changes in efficiency.

Le Pair then expands his analysis to account for the hysteresis effect (which he terms “cycling”) (le Pair, 2011b). This is the effect that additional fuel is required to change the output of a generator from a low level to a high level, and less fuel is required when ramping down. This effect has not been considered in the work by Thomson, although the analysis of similar networks in the USA by Kaffine was based on measured CO₂ emissions data, and therefore will implicitly include it (Kaffine et al., 2011; Thomson, 2014). Le Pair cites private communications that suggest that the fuel consumption of a CCGT plant that changes its output from 100% to 80% and back to 100% in an hour will consume 1% more fuel than it would have if the plant had continued running at full load (le Pair, 2011b). He then goes on to develop a “nett [sic] cycle loss” correction factor to apply to his quasi-stationary model to include this additional fuel consumption; his quasi-stationary model finds the fuel consumption for a 100% - 80% - 100% change is 85.2% of the full load fuel consumption, so the correction factor brings this value back to 101% of the full load fuel consumption (le Pair, 2011a). He argues that this is representative of a real system, as real empirical data from the Netherlands found that “the actual fuel use of the units doing the regulation and delivering the variable part of the power needed, nation wide [sic], was always some 0.3 – 0.5% higher than that calculated with the heat rate curves” (le Pair, 2011b). The fundamental error here is the confusion between the increase in fuel consumption being a proportion of the **full load fuel consumption**, or a proportion of the **fuel use calculated from heat-rate curves**. The empirical evidence, which shows that the fuel consumption is 0.3 – 0.5% higher than that calculated from heat-rate curves, suggests that the hysteresis effect should increase the fuel consumption **from his quasi-stationary model** by 0.3 – 0.5%: in the case of the 100% - 80% - 100% change, the fuel consumption will be 85.5 – 85.7% of the full load consumption. This figure agrees with observed operation of load-following CCGT plants. Furthermore, the correction factor applied by le Pair assumes that there is increased fuel consumption when a plant is both ramping up **and ramping down** – this doesn’t realistically reflect the dynamics of the plant operation. The most straightforward approach would simply be to increase the quasi-stationary figures by 0.5% to account for hysteresis, as shown in Table 10.

A 3.4 Life Cycle Energy Consumption

Normally the life cycle carbon emissions and the carbon displacement of a wind farm are compared by looking at the net emissions savings or the carbon payback time; however, in this analysis le Pair has chosen to assume that the energy consumption associated with construction, installation, grid connection and grid adaptation can be shown as a decrease in energy production. This is likely to underestimate the carbon impacts, as energy consumption for steel manufacture, for example, is usually from coal rather than gas.

In order to apply a more conventional methodology to these results the following method has been applied:

1. The annual CO₂ emissions displacements were calculated from the fuel saving corrected for the hysteresis effect (see Table 10), to get 73.3kt, 148.7kt and 224.1kt respectively for the 100MW, 200MW and 300MW farms.
2. The life cycle CO₂ emissions were calculated based on a capacity factor of 18.7% (derived from information in le Pair (2011a)), and two different carbon footprint estimates:
 - a. From the LCA Harmonization Project, the mean carbon footprint of 12g CO₂/kWh was used to calculate the life cycle CO₂ emissions per operational year: 2.0kt, 3.9kt and 5.9kt respectively (Dolan and Heath, 2012).
 - b. From Guezuraga (2012), the carbon footprint assuming construction in China of 38.33g CO₂/kWh was applied to give life cycle CO₂ emissions per operational year of: 6.3kt, 12.6kt and 18.9kt respectively .
3. The net annual emissions savings were calculated by subtracting the life cycle emissions from the annual emissions displacements.
4. These were then described as a proportion of full load CO₂ emissions, assuming that the CO₂ emissions from CCGT operation is 480g CO₂/kWh – the value quoted by le Pair (2011b).

This methodology found the impacts of life cycle CO₂ emissions on the net emissions savings to be much more optimistic, as shown in Table 10.

A 3.5 Corrected Results and Conclusions

The system modelled by le Pair is not a good representation of the National Grid in Great Britain, and the corresponding fuel mix; however, it is interesting to examine the performance of this model with more realistic assumptions.

Insufficient information precluded the re-analysis of the raw data, or correction of the “quasi-stationary” results to more accurately reflect ramping rates, so the starting point for this re-calculation was to assume that the quasi-stationary fuel savings are correct for the given model. More robust assumptions, as outlined in Sections A 3.3 and 0, are applied to account for the hysteresis effect and life cycle carbon emissions, as shown in Table 10 (note that two different life cycle impacts are assumed). The impact of operating OCGT plant is not considered, as this is unrealistic for the National Grid. Despite the simplicity of this model, it can be shown that the correction of calculation errors results in all scenarios showing a net reduction of carbon emissions.

Fuel/CO ₂ Saving (as a proportion of 100% CCGT consumption/emissions)	Wind Capacity (MW)		
	100	200	300
Quasi stationary results from le Pair (2011b)	3.5%	7.1%	10.7%
Corrected for "cycling"/hysteresis effect	3.48%	7.06%	10.65%
Corrected to account for typical life cycle impacts (Dolan and Heath, 2012)	3.39%	6.88%	10.37%
Corrected to account for pessimistic life cycle impacts (Guezuraga et al., 2012)	3.18%	6.47%	9.75%

Table 10 - Results of le Pair, 2011, adjusted for more robust assumptions

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