1	PHYSICAL CHEMISTRY OF WATER TREATMENT PROCESSES
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3	Removal of Fe and Mn from Polluted Water Sources in Lesotho
4	using Modified Clays
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6	Lebea N. Nthunya ^{1,2*} , Mosotho George ¹ , Monaheng L. Masheane ^{1,2} , Méschac-Bill Kime ³ ,
7	Sabelo D. Mhlanga ^{4*}
8	
9	¹ Department of Chemistry and Chemical Technology, Faculty of Science, National University of Lesotho,
10	P.O. Roma 180, Maseru 100, Lesotho
11	Email: <u>nthunyalebea@gmail.com</u>
12	² Department of Applied Chemistry, University of Johannesburg, P.O. Box 17011, Doornfontein, 2028,
13	Johannesburg, South Africa
14	³ Department of Metallurgy, University of Johannesburg, P.O. Box 17011, Auckland Park, 2028, South
15	Africa
16	⁴ Nanotechnology and Water Sustainability Research Unit, College of Science, Engineering and
17	Technology, University of South Africa, Florida, 1709, Johannesburg, South Africa
18	Email: <u>mhlansd@unisa.ac.za</u>
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20 21 22 23 24 25 26 27 28 29 30 31 32 33 34 35 36	Abstract - This paper reports the use of unfunctionalized and phenylalanine functionalized clays as an alternative cost effective, environmental friendly and efficient sorbent for the removal of Mn and Fe from polluted drinking water sources in Lesotho. Mn and Fe were found to be the most prevalent metal pollutants in rivers supplying three drinking water treatment plants situated in Maseru, Roma and Quthing. The Mn and Fe metals were adsorbed on two different clays (clay A – a black clay and clay B – a yellow clay) collected from Ha-Teko in Maseru (clay A) and Phoqoane in Mafeteng district (clay B) in Lesotho. As a common assessment for adsorption materials, comprehensive batch test studies were performed to assess the effect of pH, stirring time, initial concentration and kinetic studies of the adsorption of these metals. The adsorption of the metals was greater at higher pH and equilibrium was reached at pH 8.0 after 30 min of stirring. The phenylalanine functionalized clays displayed improved adsorption efficiency of up 100 % (Fe adsorption using modified clay A in 30 min) while the unfunctionalized clays gave relatively low adsorption of up to 70 % (Fe and Mn adsorption using modified clay A). This was observed on real and simulated water samples. The findings of this work suggest that the clays, which are present in abundance in Lesotho, can be effectively used for the removal of Fe and Mn from drinking water sources in Lesotho.
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INTRODUCTION

Sewage discharges, urban runoffs, agricultural wastes, industrial effluents and open pit 41 latrines located nearby water sources are known to introduce a lot of harmful pollutants in water 42 sources [1, 2]. This presents a real threat as heavy metals may also be entrained to water 43 treatment systems. The presence of high concentrations of Fe and Mn ions in water sources in 44 Lesotho is a growing concern. These metals can cause death when present at high concentrations 45 due to their toxicity. Fe^{2+} and Mn^{2+} ions that are not removed prior to distribution of treated 46 water to the communities are oxidized in the presence of oxygen and chlorine to form a dark-47 48 brown colour which recedes the aesthetic appearance of potable water [3]. Assimilation of high concentrations of Mn (higher than 0.4 mgL^{-1} in potable water) has been reported to cause brain 49 malfunctioning and results in neurotic disorder, while concentrations of Fe above 3 mgL⁻¹ can 50 51 cause diseases like diabetes, cardiopathy and kidney failure [4, 5].

52 The most commonly used drinking water treatment processes include coagulation, flocculation, filtration, ion exchange, membrane filtration and adsorption. These processes have 53 been reported to have challenges related to the removal of high concentrations of heavy metals 54 55 and emerging contaminants in the water [6]. Cost effective and environmentally friendly water purification technologies continue to be sought and developed by many researchers worldwide 56 [7]. Water purification by adsorption has been extensively explored due to its high efficiency, re-57 usability, availability, low cost, low toxicity and ease of handling [8]. In particular, clay 58 materials as adsorbents have attracted much interest due to their high specific surface area and 59 60 high adsorption capacity [9]. Lesotho is characterized by clay minerals found in abundance at 61 both the lower and upper catchment of the country [10]. Others have shown that these clay minerals are rich in zeolites, silicates and iron-titanium oxide and these minerals possess 62 adsorption capacities that can be used for removal of heavy metals in water [11]. In this work, 63 the adsorption of Fe and Mn using locally available clay minerals was investigated. 64

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69	EXPERIMENTAL
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71	Sample Collection
72	The clay samples were collected from brick producing companies in Lesotho and purified
73	before adsorption studies. The water samples were collected from Roma, Maseru central, and
74	Quthing water treatment plant. Water samples were preserved by adding 0.2 M HNO_3 (1 mL) to
75	each water sample before storage. The sampling locations are shown in the Lesotho map given in
76	Fig. 1. The study area was chosen due to production of water with poor quality.



Fig. 1. Water sampling points and clay collection sites of the study.

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Sample Preparation

The clay samples were crushed and sieved using 20 µm pore sized sieves. 20 g of each clay sample was added to a separate beaker containing 50 mL of de-ionized water and stirred for 30 min. The pH of the resulting mixture was measured using a pH meter. All mixtures were acidified with 3 M H₂SO₄. The strong acid was chosen to remove all the metal ions which are naturally adsorbed in the cavities of the clay, thereby conditioning them for better adsorption. The acidified clay solutions were stirred for 30 min and allowed to settle. The supernatants were decanted and the clays were washed several times with acidified distilled water until the supernatant liquid gave a negative test for Fe, Mn and other metals. This was done to remove the metals which are naturally bound on the clays in order to increase their adsorption capacity. The pH of the clays was adjusted to their natural pH using CH₃COOH/CH₃COONa buffer solution with pH 6.5 as well as 1 % HNO₃ and 1 % NaOH solutions. The clays were filtered and dried in an oven at 105 °C. They were then labeled as clay A = black clay and clay B = yellow clay.

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Modification of Clays

A procedure reported by Ray (2006) was followed to modify the clays. Typically, 95 distilled water (25 mL), phenylalanine (Phe) (2 g), H₂SO₄ (4 mL) were added into the beaker and 96 97 stirred to form a homogeneous solution. Phenylalanine was found to be the best solvent to use for modification of the clays since it has binding sites that can be used for adsorption of heavy 98 metals. The other reason to use it was that it is not toxic and does not leach into solution [12]. 99 The mixture was stirred for 15 min and the reaction vessel was allowed to cool to room 100 temperature. Thereafter, HNO₃ (2 mL) was added drop wise to the mixture with constant 101 stirring. The resulting mixture was cooled in an ice bath to 0 °C and 10 g of clay A or B was 102 added. The mixture was further stirred for 30 min at 40 °C and the resulting modified clay was 103 filtered and dried in an oven at 105 °C for 48 h. 104

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Characterization of Clays

The chemical composition of the clays was ascertained using X-ray fluorescence with a 107 Axios^{mAX} wavelength dispersive X-ray fluorescence (WDXRF) spectrometer. The particle size 108 of the clays was analyzed using a hydrometer. Sodium hexametaphosphate $Na_6(PO_3)_6(5\%)$ was 109 used as the dispersing solution. The blank of the dispersing solution was prepared by diluting 5 110 111 % Na₆(PO₃)₆ to a desired dilution. 30 g of clay and 100 mL of 5 % Na₆(PO₃)₆ were added into a beaker. This mixture was stirred for 2 min and allowed to settle. The suspension was transferred 112 to a measuring cylinder. The hydrometer readings were taken at 40 sec and 412 min. The amount 113 of clay, silt and sand were calculated using the following equations. 114

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116 % clay =
$$\frac{\text{corrected hydrometer reading at 412 min}}{\text{weight of sample}} X100 \%$$
 (1)

118% silt =
$$\frac{corrected hydrometer reading at 40 sec}{weight of sample}$$
 X 100 % - % clay(2)119110% sand = 100 % - (% clay + % silt)(3)120% sand = 100 % - (% clay + % silt)(3)121The functional groups of the clays were analyzed using Perkin Elmer Spectrum 100 FTIR123spectrometer. X-ray diffraction (XRD) was conducted on dry powdered samples using a Rigaku124Ultimate IV diffractometer. The XRD patterns were analyzed qualitatively using "PDX1."125software, provided with JCPDS-PDF2' database.126A known amount of purified clay was added to a 500 mL beaker. To this, 50 ml of the129aqueous solutions of known concentrations of MnO2 or FeCl2 was added. The pH of the solutions130adjusted using 8 M NaOH and 2 M HCl and then stirred for 20 min with magnetic stirrer. The131solutions were then filtered under vacuum. Fe and Mn concentrations were determined from the132feed and filtrate using the UV-Vis Spectrophotometer (DR/2000).133Determination of the Effect of pH on the Adsorption of Fe and Mn Using Clays A and B135A 25 mL solution of FeCl2 (5 ppm) and clay A were added to five beakers. The pH of the139mixtures was adjusted to 3, 5, 8, 10 and 12. Each solution was stirred for 30 min and then filtered140under vacuum pressure. The concentration of Fw aus measured from the feed and filtrate. The138same procedure was conducted for adsorption of M using MnO2 (5ppm) solution. Clay B was139also tested for adsorption of Fe and Mn using the same procedure. Adsorption efficiency of the140clay w

147 final mass (m_f) = final concentration x volume used

(7)

(8)

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Effect of Stirring Time on Adsorption Fe and Mn Using Clays A and B

Mass of analyte per mass of clay $(m/m) = \frac{\text{mass of adsorbate}}{\text{mass of adsorbent}}$

A 25 mL solution of FeCl₂ (5 ppm) and clay A were added to six beakers. The pH of 153 these mixtures was adjusted to pH 8 (using the buffer solution of pH 8, NaOH and HNO₃) since 154 it was found to be the average pH for optimal adsorption. The solutions were then stirred for 5 155 156 min (solution in beaker 1), 10 (solution in beaker 2), 15 min (solution in beaker 3), 20 min (solution in beaker 4), 25 min (solution in beaker 5) and 30 min (solution in beaker 6) and 157 158 filtered under vacuum. The concentration of Fe was measured from the feed and filtrate using a 159 UV-Vis spectrophotometer. The same procedure was conducted for adsorption of Mn using MnO₂ (5 ppm) solution. The similar experiment was conducted using clay B for adsorption of Fe 160 and Mn. The adsorption efficiencies were then calculated. 161

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Effect of Initial Concentration on the Adsorption of Fe and Mn Using Clays A and B

164 Clay A, distilled water and varying concentrations of Fe solutions (5 M, 10 M, 15 M, 20 165 M, 25 M, 30 M) were added to six beakers. The pH of these mixtures was adjusted to pH 8 and 166 all the solutions were stirred for 30 min and filtered under vacuum. The concentration of Fe was 167 determined from the feed and filtrate using a UV-Vis spectrophotometer. The same procedure 168 was conducted for adsorption of Mn using MnO_2 (5 M) solution as well as the adsorption on clay 169 B for both metals. Adsorption efficiencies were then calculated.

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Determination of the Adsorbent Holding Capacity

172 Clay A and Fe solution were added to a beaker. The pH of the solution was adjusted to 173 pH 8. The mixture was stirred for 30 min and filtered under vacuum. The Fe concentration was 174 measured from the feed and filtrate. The mass of adsorbate (Fe) adsorbed per mass of the 175 adsorbent (clay) was calculated using Eq. (8). The same procedure was conducted for adsorption 176 of Mn using MnO₂ (5 M) solution. The similar experiment was conducted using clay B on 177 adsorption of both metals.

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179	Adsorption of Fe and Mn from Real Polluted Samples
180	The adsorption efficiencies of the modified clays were done on samples collected from
181	three different locations (Fig. 1.). The sampling sites were selected due to high concentration of
182	Fe and Mn detected in their water samples. Adsorption was done at pH 8 with stirring time of 30
183	min.
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185	RESULTS AND DISCUSSION
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187	Characterization of Clays
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189	X-Ray Fluorescence Analyses
190	The chemical composition of the clays was determined using a wavelength dispersive X-
191	ray fluorescence (WDXRF) spectrometer. All the clays were found to have a high content of
192	silica and a moderate content of alumina, which are known to be active sites for the adsorption of
193	metals (Table 1). Other metal oxides were found at trace levels. The high chemical composition
194	of silica and alumina explains the ability of these clays to remove high concentrations of Fe and
195	Mn metals contaminants due to the presence of silanol (Si-O) and aluminol (Al-O) groups which
196	are capable of forming inner sphere complexes with the metals at particle edges of the clay [13].
197	MnO is the chemical compound responsible for the appearance (colour) of black clay while V_2O_5
198	is responsible for the appearance of yellow clay.
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Mineral	Black clay, A (%)	Yellow clay, B (%)
SiO ₂	71.8	71.8
Al ₂ O ₃	16.4	14.5
Na ₂ O	0.10	0.10
K ₂ O	2.96	2.45
MgO	0.80	0.70
MnO	1.10	<0.01
Fe ₂ O ₃	1.42	4.56
CaO	0.22	0.23
TiO ₂	1.05	0.97
P ₂ O ₅	0.03	0.01
Cr ₂ O ₃	<0.01	<0.01
V ₂ O ₅	0.01	0.11
SO ₃	<0.01	<0.01
Cl	<0.01	<0.01

209 Table 1: The chemical composition of minerals in the clays.

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Particle Size Determination

The particle sizes of the clays are recorded in Table 2. The composition of the clays was found to be 15.8% for black clay and 15.3% for yellow clay and the rest was silt and sand. The silt and sand components play an important role in water filtration as they increase the porosity; hence reduce energy needed for water filtration [14, 15]. It must be noted however that the materials were classified as clays because of the physical properties they exhibited when mixed with water.

218	Table 2: The particle size and	the other components	of clays.
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Composition	Black clay, A (%)	Yellow clay, B (%)
clay (< 2 µm)	15.8	15.3
silt (2 < 60 µm)	54.4	52.3
sand (> 60 µm)	29.8	32.3

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FT-IR Analysis of the Clays

The FT-IR analysis of the clays was conducted to confirm their functionalization using 221 phenylalanine (Phe) (Fig. 2 and Fig. 3). From Fig. 2 (a), the broad peak at 3006 cm⁻¹ is assigned 222 to a carboxylic O-H stretches present in phenylalanine. The characteristic vibration bands of 223 phenvlalanine can also be observed at 1622 cm⁻¹ (C=O), 1073 cm⁻¹ (C-N), 3512 cm⁻¹ (N-H), 224 1128 cm⁻¹ (C-O) stretches. The characteristic absorption bands of black clay were observed at 225 998 cm⁻¹, 796 cm⁻¹, 694 cm⁻¹, 3100 cm⁻¹ belonging to Si-O, Mg-Al-OH, Al-O and O-H 226 vibrations respectively (Fig. 2 (b). The characteristic absorption bands of yellow clay were 227 observed at 794 cm⁻¹, 1020 cm⁻¹ and 3216 cm⁻¹ belonging to Si-O, Mg-Al-OH and O-H stretches 228 respectively (Fig. 3. (b). The vibration bands at 998 cm⁻¹ (Si-O) and 694 cm⁻¹ (Al-O) on black 229 clay and 1017 cm⁻¹ (Si-O) on yellow clay confirm the presence of chelating groups (Al-O⁻ and 230 Si-O⁻) on modified clays. However, a shift of the C-O vibration to 1415 cm⁻¹ on the black clay 231 and 1206 cm⁻¹ on the yellow clay confirmed the addition of Si-O and Al-O to the electrophilic 232 carbon of the carboxylic group on phenylalanine (Fig. 2 (c). The FTIR spectra of pure and 233 functionalized yellow clay are presented in Fig. 3. 234







Fig. 3. FT-IR spectra of (a) Phe, (b) unfunctionalized yellow clay and (c) Phe-f-yellow clay.

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XRD Analysis of the Clays

The XRD patterns of unnfunctionalized and phenylalanine functionalized black and 242 yellow clays are shown in Figs 4 and 5. The diffraction patterns show that the crystalline nature 243 of the silicate and aluminate in the clays was not affected by the modification with phenylalanine 244 on both clays. This was confirmed by the peaks at $2\theta = 26.178^{\circ}$ and 36.740° that were assigned to 245 SiO₂ and AlO₂ on the black clay (Fig. 4) and $2\theta = 26.681^{\circ}$ and 36.791° assigned SiO₂ and AlO₂ 246 on the yellow clay (Fig. 5). Both clays were found to have high content of silica (60-80%) at a d-247 spacing of 152 nm. The black clays were found to be composed of illite (3.8%), muscuvite 248 (28%), kaolinite (7.8%), and montmorillonite (1.6%) at d-spacings of 12 nm, 5 nm, 1 nm, and 12 249 nm respectively (Table 3). The yellow clay was found to consist of illite (13.1%), muscovite 250 (0.8%), kaolinite (5.7%), montmorillonite (0.03%) and bentonite (0.9%) at d-spacings of 12 nm, 251 5 nm, 1 nm, 12 nm and 12 nm respectively (Table 2). The XRD results thus confirmed the 252

chemical composition and functional groups of both groups obtained from XRF and FT-IRrespectively.





Fig. 4. XRD patterns of (a) Phe, (b) unfunctionalized black clay and (c) Phe-f-black clay.





261 Table 3: Clay phases, their d-spacing and quantitative analysis.

Phase name	Formula	Content (%)		Space grou	Space groups	
		Black	Yellow	Black	Yellow	
		Clay	Clay	Clay	Clay	
quartz	SiO ₂	58.8	79.4	152	125	
Illite	K _{0.7} Al ₂ (Si,Al) ₄ O ₁₀ (OH) ₂	3.8	13.1	12	12	
Muscovite	K Al($Mg_{0.2}$ Al _{0.8})(Al _{0.42} Si _{3.58})O ₁₀ (OH) ₂	28.0	0.8	5	5	
Kaolinite	$Al_2Si_2O_5(OH)_4$	7.8	5.7	1	1	
Montmorillonite	$Li_{0.62}(Al_3Fe_{0.45}Mg_{0.54})((Si_8Al_{0.2})O_{20}(OH)_4)$	1.6	0.03	12	12	
Bentonite	$(Ca_{0.2}Na_{0.2}K_{0.3})(Al_{1.6}Fe_{0.06}Mg_{0.3})(Al_{0.3}Si_{3.7}O_{10}(OH)_2)$	0	0.9	1	12	

265 Adsorption Studies 266 267 The Effect of pH on the Adsorption Capacities of Clay A and B The pH of the solution containing the metal to be adsorbed is one of the most important 268 269 factors that determine the efficiency of the adsorbent. This is because pH has the potential to interact with the electronic properties of the active sites of the adsorbent and can make it less 270 271 effective as an adsorbent. The effect of pH on the adsorption of Mn and Fe ions on the two different unmodified and modified black and yellow clays was studied and the results are shown 272 in Fig. 6 and Fig. 7. 273 274



Fig. 6. Effect of pH on the adsorption of Fe and Mn using unmodified clay A and B.

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Fig. 7. Effect of pH on the adsorption of Fe and Mn using modified clay A and B.

The adsorption efficiencies of clays were found to increase with an increase in pH to 281 282 neutral and slightly basic pH. Further increase in pH led to a decrease in adsorption. At lower pH, functional adsorption groups dissociate and hydronium ions compete with the metal ions at 283 284 the active surface area of the clay. As the pH increases, the hydronium ion concentration decreases and neucleophilicity of the centers for adsorption of metals increases. At higher pH 285 286 (*i.e.* pH > 8), there is formation of metal hydroxides which reduce the adsorption of heavy metals, thereby decreasing the adsorption efficiency of clay. Adsorption of Fe was greater than 287 288 that of Mn on both modified and unmodified clay A (Fig. 6 and Fig. 7). As pH was increased and neutralization occurred, both Fe and Mn formed aqua complexes and water exchange rate of Mn 289 was greater than that of Fe which reduced adsorption of Mn ions. A study conducted by Kamel et 290 al. (2004) reported a greater affinity of Fe ion adsorption on kaolinite clay than that of Mn ion 291 292 [16]. However, Fe adsorption was very poor on clay B probably due to lack of ion exchange and poor mass transfer from the bulk solution to the active sites of adsorption since diffusion 293 coefficient of Fe is smaller than that of Mn on adsorption in most adsorption [17]. 294

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Effect of Stirring on the Adsorption Kinetics for Clays A and B

The effect of stirring time of Fe and Mn on the adsorbents was also studied. The adsorption efficiencies of both modified and unmodified clays increased with increase in stirring time (Fig. 8 and Fig. 9). The adsorption of the metals increased sharply in the first 5 min as there were many vacant sites available for adsorption. Further increase in contact time resulted to a
slower adsorption rate because the number of active surface sites had decreased. An equilibrium
(where no more adsorption) occurred at 20-30 min of stirring time for both clays. The adsorption
of Fe and Mn was found to be higher on the modified clays than that of unmodified clays.
Addition of phenylalanine increased chelating groups which increased adsorption efficiency of
clays.





Fig. 8. Effect of stirring time on the adsorption of Fe and Mn using unmodified clay A and B.



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Fig. 9. Effect of stirring time on adsorption of Fe and Mn using modified clay A and B.

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Effect of Initial Concentration on Adsorption of Clays A and B

314 The effect of initial concentration of the adsorption of Fe and Mn was also studied. 315 Solutions containing high concentrations of Fe and Mn increased the competition for adsorption due to saturation at active surface sites [1]. Therefore, increasing initial concentration decreased 316 317 the adsorption efficiency (Fig. 10 and Fig. 11). This is contradictory to other studies where increase in initial concentration increases adsorption efficiency of adsorbent [18]. However, a 318 different trend was observed in unmodified black clay where adsorption efficiency increased 319 with an increase in initial concentration. This was explained by the high diffusion rate from the 320 321 bulk solution into the adsorption active sites of the adsorbent. Also more active sites available since black clay has more silicate and aluminate content than the clay B which are responsible 322 for chelation of metal ion. 323



Fig. 10. Effect of initial concentration on the adsorption of Fe and Mn using unmodified clay Aand B.

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Fig. 11. Effect of initial concentration on the adsorption of Fe and Mn using modified clay A andclay B.

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The proposed mechanism for the adsorption of the metals on the clays is as follows: When clays are purified with H_2SO_4 , The metal ions including Fe^{3+} and Mn^{2+} which are naturally bound to the adsorption sites of the clays are washed away (Fig. 12). The positively charged

protons have strong electrostatic interaction with the electron rich \equiv SiO⁻ and \equiv AlO⁻, as a result, the protons replace the metal ions from the surface of the clay adsorbent. To activate the adsorbent, NaOH is added to neutralize the protons bound on the surface of the adsorbent, hence increase the electron density on the electron rich atoms (Fig. 12). The active surface is then used for adsorption purposes.



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Adsorption Capacity of the Clays

The adsorption capacities of unmodified and modified clay (A and B) were evaluated for adsorption of Fe and Mn from simulated samples at optimum conditions. The adsorption was found to increase significantly on functionalized clays (Table 4 and Table 5). Functionalization of these clays with phenylalanine increased the binding sites of the metals due to introduction of chelating groups (amino and carboxylic groups); hence improve metal removal from solutions.

Analyte	Trials	Black Clay (A)		Yellow Clay (B)		
		mass of clay (g)	madsorbate/madsorbent	mass of clay (g)	madsorbate/madsorbent	
			(mg/g)		(mg/g)	
Mn	1	1.0156		1.0124		
	2	1.0124	0.113 ± 0.000	1.0216	0.121 ± 0.002	
	3	1.0139		1.0094		
Fe	1	1.0374		1.0221		
	2	1.0286	0.030 ± 0.001	1.0315	0.011 ± 0.001	
	3	1.0412		1.0267		

Table 4: Holding capacity of unmodified black (A), and yellow (B) clay at optimized parameters.

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17/	I able 5' Holding	CADACITY OF	тоаннеа ріаск	(A) and (A)	venow(B)	CIAV ALO	biimizea i	parameters.
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Analyte	Trials	Clay A		Clay	
		Mass of clay (g)	$M_{adsorbate}/m_{adsorbent}$	Mass of clay	$M_{adsorbate}/m_{adsorbent}$
			(mg/g)	(g)	(mg/g)
Mn	1	1.0142		1.0134	
	2	1.0241	0.128 ± 0.0011	1.0156	0.138 ± 0.003
	3	1.0182		1.0125	
Fe	1	1.0169		1.0165	
	2	1.0112	0.0398 ± 0.0003	1.0158	0.030 ± 0.001
	3	1.0125		1.0119	

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Adsorption on Real Water Samples

Following the optimization the clay samples and their high capacity to adsorb Fe and Mn ions from simulated water samples, the clays were used in a batch reactor to remove these metals from real water samples. The adsorption efficiency of the clay obtained from Maseru sample was found to be lower than that of clays obtained from Quthing and Roma (Fig. 13 and Fig. 14). This observation was attributed to the saturation of the adsorption sites of the clays from Maseru samples because the water contained higher concentrations of Fe and Mn than the other two samples.



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Fig. 13. Adsorption of Mn and Fe on real samples using unmodified clay A and B.



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Fig. 14. Adsorption of Mn and Fe on real samples using modified clay A and B.

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368 Clay A was observed to have the highest adsorption of Fe and Mn compared to clay B on 369 real samples obtained from the three water sources (Fig. 13). This could be due to high level of 370 adsorption active surface sites present in clay A relative to clay B. Functionalization of all these 371 clays with phenylalanine increased their adsorption efficiency (Fig. 14) as shown by the 372 increased removal of Fe and Mn from the water samples.

373	Adsorption was significant enough to the give treated water an acceptable aesthetic
374	appearance (Fig. 14). The concentrations of Fe detected in Maseru, Roma and Quthing samples
375	were found to be 3.15 ppm, 2.99 ppm and 2.31 ppm respectively. The adsorption efficiencies
376	using modified clay A were found to be 97.8%, 98.1% and 98.4% while using clay B, they were
377	found to be 85.5%, 94.2%, 98.2%. The concentrations of Mn detected in Maseru, Roma and
378	Quthing samples were found to be 1.201 ppm, 1.527 ppm and 1.754 ppm respectively. The
379	following adsorption efficiencies were obtained 98.5%, 97.9%, 100% when using clay A and
380	97.6%, 96.1%, 98.9% when using clay B. The amount of Fe and Mn adsorbed and the adsorption
381	efficiencies on Fig 14 are key explainations for the aesthetic appearance of the treated water (Fig.
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407 18, 19). The R² values of Pseudo second order kinetics were generally higher than those of the 408 Pseudo first order kinetics except for the adsorption of Fe on unmodified red (clay B), Fe on 409 modified black clay (clay A) and Mn on modified red clays (clay A) (Table 6). This meant that 410 the adsorption of Mn and Fe on both clays (modified and unmodified) was favoured by the 411 Pseudo second order kinetics except for the adsorption of Fe on unmodified red (clay B), Fe on 402 modified black clay (clay A) and Mn on modified red clays (clay A) which was favoured by 413 Pseudo first order kinetics.

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426 Fig. 16. The Pseudo first order kinetics for the adsorption of Mn and Fe on unmodified clay A427 and clay B.



Fig. 17. The Pseudo second order kinetics for the adsorption of Mn and Fe on unmodified clay A
and B. The plot of Fe on unmodified clay B is not shown because no adsorption of Fe occurred
in the first 15 min.



Fig. 18. The Pseudo first order kinetics for the adsorption of Mn and Fe on modified clay A andclay B.



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Fig. 19. The Pseudo second order kinetics for the adsorption of Mn and Fe on modified clay A
and B. The plot of Fe on modified clay B is not shown because no adsorption of Fe occurred in
the first 15 min.

Table 6: The correlation coefficients for the kinetic studies of clay A and B.

Clays Mn ²⁺		Fe ²⁺		
	R ² 1st order	$R^{2}_{2nd order}$	R ² 1st order	$R^{2}_{2nd order}$
Aunmodified	0.9594	0.9973	0.7281	0.9997
$B_{unmodified}$	0.9067	0.9993	0.9650	0.2101
A _{modified}	0.8067	0.9958	0.8777	0.8630
B _{modified}	0.777	0.3135	0.6643	0.9867

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CONCLUSIONS

This study was aimed at determining the adsorption efficiency of Fe and Mn onto two 447 different types of clays (black and yellow clays) obtained from three sites in Lesotho. These 448 clays were demonstrated to be suitable for removal of Fe and Mn metals from drinking water 449 450 sources. The clays are relatively cheap and more abundant that conventional adsorbents used such as activated carbon, aluminium sulfate, aluminium hydroxide, etc, in Lesotho. The 451 functionalization of the clays with phenylalanine resulted to their improved adsorption efficiency 452 and the water produced presented acceptable aesthetic appearance with very low untreated metal 453 concentration; within the international guidelines. It was also found that increasing pH of the 454 455 water samples and the stirring time resulted to increased adsorption efficiency of the clays until equilibrium was reached. Increasing the initial concentration decreased adsorption efficiency due 456 to the fast depletion of vacant active surface sites for adsorption. The clays investigated were 457 found to be effective on removal of Mn and Fe from the real water samples with up to 99.7% 458 459 after modification with phenylalanine. The use of these clays can be a viable solution to low income economies since the clays are present in abundance. The clays could be mined and used 460 461 as adsorbent based filters for domestic water supply in Lesotho.

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