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# The Glasgow consensus on the delineation between pesticide emission inventory and impact assessment for LCA

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## Abstract

**Purpose:** Pesticides are applied to agricultural fields to optimise crop yield and their global use is substantial. Their consideration in Life Cycle Assessment (LCA) is affected by important inconsistencies between the emission inventory and impact assessment phases of LCA. A clear definition of the delineation

between the product system model (Life Cycle Inventory - LCI, technosphere) and the natural environment (Life Cycle Impact Assessment - LCIA, ecosphere) is missing and could be established via consensus building.

**Methods:** A workshop held 2013 in Glasgow, UK, had the goal of establishing consensus and creating clear guidelines in the following topics: 1) Boundary between emission inventory and impact characterisation model 2) Spatial dimensions and the time periods assumed for application of substances to open agricultural fields or in greenhouses, and 3) Emissions to the natural environment and their potential impacts. More than 30 specialists in agrifood LCI, LCIA, risk assessment, and ecotoxicology, representing industry, government, and academia from 15 countries and four continents met to discuss and reach consensus. The resulting guidelines target LCA practitioners, data (base) and characterisation method developers, and decision makers.

**Results and discussion:** The focus was on defining a clear interface between LCI and LCIA, capable of supporting any goal and scope requirements while avoiding double counting or exclusion of important emission flows/impacts. Consensus was reached accordingly on distinct sets of recommendations for LCI and LCIA respectively, recommending for example that buffer zones should be considered as part of the crop production system and the change in yield be considered. While the spatial dimensions of the field were not fixed, the temporal boundary between dynamic LCI fate modelling and steady-state LCIA fate modelling needs to be defined.

**Conclusions and recommendations:** For pesticides application, the inventory should report: pesticide identification, crop, mass applied per active ingredient, application method or formulation type, presence of buffer zones, location/country, application time before harvest and crop growth stage during application, adherence with Good Agricultural Practice, and whether the field is considered part of the technosphere or the ecosphere. Additionally, emission fractions to environmental media on-field and off-field should be reported. For LCIA, the directly concerned impact categories and a list of relevant fate and exposure processes were identified. Next steps were identified: 1) establishing default emission fractions to environmental media for integration into LCI databases, and 2) interaction among impact model developers to extend current methods with new elements/processes mentioned in the recommendations.

**Keywords:** Pesticides, consensus, life cycle inventory (LCI), life cycle impact assessment (LCIA), spatial boundary, temporal boundary, ecosphere, technosphere

## 1 Introduction

Pesticides, such as fungicides, herbicides, insecticides, plant growth regulators, and defoliant are applied to agricultural fields in order to optimise crop yield. Agricultural pesticide use is substantial and amounts globally to about 1.2 million metric tons of active ingredients (the biologically active part of a commercial pesticide formulation) per year between 2005 and 2010 (FAO 2013). In contrast to regulatory assessments, LCA is not used to analyse (or question) product risk or safety, but instead supports to establish and inform environmental performance profiles aiming to identify the most environmentally sustainable way(s) of providing a product or service between different options. These may, for example, range from various functionally equivalent pesticides applied in conventional agriculture to alternative solutions, such as organic or integrated farming practices. Whatever the choice of alternatives, LCA should be able to reliably identify the trade-offs between alternative solutions.

In practice, this ideal is currently hampered by significant inconsistencies between emission inventory modelling and impact assessment, two distinct phases of an LCA study. Notably, a clear definition of the

delineation between the product system model as represented in the Life Cycle Inventory (LCI) (i.e. technosphere) and the natural environment as represented by Life Cycle Impact Assessment (LCIA) characterisation models (i.e. ecosphere) is currently missing (Dijkman et al. 2012). The typical difficulty with respect to pesticides is to quantify the proportion emitted to the different media in the ecosphere, while usually only the amount applied to the agricultural field is known. This leads to the core question of considering the field as part of the ecosphere or technosphere, which is also related to different agricultural practices (e.g. organic vs. conventional) and represents a modelling choice that is currently inconsistently handled in LCA practice (Hellweg and Geisler 2003; Birkved and Hauschild 2006; Nemecek and Kägi 2007; Dijkman et al. 2012; van Zelm et al. 2014). As a consequence, several approaches and hypotheses have been proposed and used up to now, which leads to inconsistent results and may not always represent realistic conditions.

In 2003, the (Swiss) Discussion Forum on LCA approached the subject of pesticide emission and impact modelling in LCA, providing an overview of LCIA models for pesticides and showing directions for further development (Hellweg and Geisler 2003). Although, more than ten years have passed since then, several of their conclusions remain valid, stating e.g. that: 1) “the correlation between spatial variability on the one hand and regionally different crops and agricultural management practices on the other hand is not taken into account in LCA.”<sup>1</sup>; and 2) underlying assumptions and the model performances [of LCIA methods] need further validation and verification” p. 312 in Hellweg and Geisler (2003).

A wealth of information is available in the literature on fate and transport processes of pesticides after application to agricultural fields, such as spray drift including deposition, volatilisation, leaching, runoff, washoff, and plant uptake, as well as influencing factors such as buffer zones, application method, temperature, crop/soil type, and degradation, or special cases such as greenhouses. An overview (non-exhaustive) is given in Table 1.

(table 1 here)

These processes and conditions mainly influence how much of a pesticide applied to a field will be transported from the field into different environmental media, which is a first step in the environmental fate modelling of these chemicals. These processes depend on a number of factors and in particular on the active ingredient, the formulation containing it, the application method, weather conditions (before, during, and after application), soil conditions, crop characteristics, and whether and how irrigation is used. This complexity creates a substantial obstacle to a consistent quantification of agricultural emissions to environmental media.

In LCA, environmental fate of a chemical emission to the ecosphere is usually modelled in a generic way during the impact assessment phase. However, given the important influence of local conditions and agricultural practice on the first step of the environmental fate of pesticides (i.e. from field application to transport to off-field air, water, and natural soil), using generic fate factors leads to extremely high uncertainty. For this reason, the field may be considered as being part of the technosphere and the first step of fate modelling is thus considered as part of the inventory modelling. Although, inconsistencies exist between this first step of site specific and agricultural practice specific fate modelling included in the inventory phase and the next steps of generic fate, exposure, and effect modelling of toxic substances included in the impact assessment phase.

With the goal of establishing consensus, a full-day expert workshop was organised in a collaborative effort between the Division for Quantitative Sustainability Assessment of the Technical University of Denmark and Quantis on Saturday 11<sup>th</sup> May 2013 in Glasgow, UK, back to back with the 23<sup>rd</sup> SETAC Europe meeting. More

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<sup>1</sup> Our interpretation of this phrase is that it refers to the spatial variability of a number of (partially) correlated factors including local environmental conditions, crops produced, and agricultural management practices applied.

than 30 specialists, representing industry, government, and academia from 15 countries and four continents participated.

The objective of the workshop was to create consensus and clear guidelines for LCA practitioners, data (base) and characterisation method developers, and decision makers where the boundary between the product system model representing the technosphere (emission inventory) and the environmental characterisation model representing the ecosphere (impact assessment) should be set in all three spatial dimensions and time when considering application of substances to an open agricultural field or in greenhouses, and consequent emissions to the natural environment and their potential impacts. The workshop explicitly excluded discussions on *how* to quantify emissions or related impacts of pesticides, while focusing on clearly defining *what* should be quantified in the emission inventory of the product system and *what* should be quantified by the characterisation factors (CF) from the impact characterisation models, avoiding any overlap or double counting of chemical fate processes to be considered. In setting guidelines and recommendations, the workgroup focused on issues of science and scientific consensus and avoided recommendations that would be dependent on or endorse a specific model.

## 2 Methods

### 2.1 Life Cycle Inventory (LCI)

In order to at least provide meaningful emission estimates, LCA practitioners and developers proposed generic assumptions regarding varying percentages of applied active ingredient emitted to environmental media (Margni et al. 2002; Audsley et al. 2003; Panichelli et al. 2008) and in greenhouses (Antón et al. 2004; Juraske et al. 2007). Furthermore, a number of very different (i.e. inconsistent) approaches and assumptions are currently applied in quantifying life cycle emission inventories of pesticides in any LCA study involving agricultural systems, for example:

- 1) The widely used life cycle inventory database Ecoinvent adopted the assumption that the amount of pesticide applied is emitted 100% to agricultural soil (Nemecek and Kägi 2007);
- 2) In the U.S. Life Cycle Inventory Database (NREL 2012) pesticide emissions are inventoried as emissions to air and water based primarily on leaching and runoff data from (US EPA 1999; Kellogg et al. 2002);
- 3) The pesticide emissions model PestLCI 2.0 (Dijkman et al. 2012) employs a local fate model that estimates the amounts emitted (meaning here leaving the agricultural field) to air, surface water, and ground water through the soil, respectively, based on application methods, local meteorological conditions, crop types, and other influencing factors. This model considers the agricultural soil as part of the technosphere;
- 4) The recent U.S. field crop LCI database “USDA LCA Digital Commons” does not consider any fate and transport losses and clearly reports *release* of pesticides to agricultural soil, and also to air when relevant according to the application method inventoried (Cooper et al. (2013); <http://www.lcacommons.gov>);

Choosing among these available approaches will have an important influence on the results of the LCI as illustrated in Table 2 for a simplified example like the application of 2,4-D to corn.

(table 2 here)

### 2.2 Life Cycle Impact Assessment (LCIA)

All current LCIA characterisation models for toxicity, such as IMPACT 2002 (Pennington et al. 2005) – used in IMPACT 2002+ (Jolliet et al. 2003), USES-LCA (van Zelm et al. 2009) – applied in ReCiPe (Goedkoop et al. 2012), and USEtox (Rosenbaum et al. 2008) – adopted by ILCD (EC-JRC 2011) and TRACI 2.0 (Bare 2011), consider the fate (transport, distribution, and degradation) of a pesticide, once emitted to the ecosphere (= natural environment), i.e. air, water, and soil outside the agricultural field. Nevertheless, both USEtox and

USES-LCA differentiate between agricultural and natural soil as environmental media. The USEtox developers included agricultural soil arguing that “using two soil types, agricultural and natural soil, accounts for the fraction of agricultural soil relative to the total soil surface and also allows for specific (e.g. pesticide) emissions occurring on agricultural soil only” (Rosenbaum et al. 2008). For ecotoxicity, USEtox only accounts for freshwater ecotoxicity and does currently not provide factors for terrestrial ecotoxicity impacts on (or off) the agricultural field due to missing data and appropriate models that could provide a solid basis for a consensus. Therefore, only impacts outside the agricultural field and its soil (i.e. only effects occurring in the ecosphere) are characterised.

In practice this means that the practitioner has two options at the moment: 1) a pesticide emission model (like PestLCI 2.0) used in combination with CFs for emissions to off-field environmental media (e.g. natural soil, air, freshwater in the case of USEtox), or 2) the amount applied to the agricultural field is used and characterised with a CF for on-field emissions (e.g. to agricultural soil in USEtox – CFs for all other emission media are representing off-field emission media and are thus not applicable). In the absence of widely applicable pesticide emission models at the time, this was considered an approximate solution for pesticide emissions, but many important factors and fate processes such as wind drift, application method, buffer zones, plant deposit etc. are not modelled and a CF for an emission to agricultural soil in USEtox comes with considerable model uncertainty of at least two to three orders of magnitude (excluding uncertainty due to missing transport processes) (Rosenbaum et al. 2008).

With the inherent assumption of steady-state conditions, the dynamic fate processes taking place between the application and the moment the substance is transported outside the agricultural field are not explicitly modelled in LCIA, and are hence presently expected to be represented by LCI data. An implicit assumption thereby is that the field is part of the technosphere, which means that impacts to the environment/ecosystem on the field are not specifically considered. In consequence, all currently used toxicity LCIA models are incompatible with LCI data representing the amount applied to the field. This is true for ecosystem as well as human health LCIA models, although some exceptions to this general rule exist for specific impact pathway mechanisms. For example the human health characterisation model for pesticide residues in food crops by Fantke et al. (2011b), which requires information on the quantity applied to the field, rather than the consequent environmental emissions.

### **2.3 Workshop resume**

In order to preserve consistency with the characterisation of other chemical emissions, there seems to be limited flexibility to adapt LCIA characterisation modelling; hence, the most practical way to improve consideration of pesticide field applications in LCA may be to focus on clearly defining the inventory requirements for pesticide emissions as illustrated in Figure 1. Many definitions of delineations between LCI and LCIA are possible that respect the requirements that all flows between environmental media connect and the mass balance is conserved, but it is essential that all stakeholders (i.e. LCI/LCIA modelers and LCA practitioners) are using a consistent approach. Therefore, a consensus-based definition of a convention and related guidance for LCA practitioners and methodology developers is required and was the main objective of the Glasgow pesticide workshop.

(figure 1 here)

While a brief overview will be given hereafter, the complete minutes, details, and most slides presented during the workshop can be found in the Supporting Information. The workshop started with four presentations providing experiences and perspectives of LCA practitioners, emission inventory practice, and impact assessment developers.

Presenting a practitioner’s perspective, Sebastien Humbert (Quantis) pointed out that “Between 1 to 5 hours (max) can be allocated by practitioners to modelling inventory of pesticides for one crop. Therefore, practical and efficient solutions are needed.” Current issues encountered when conducting an LCA are: a) media of substance application and related fraction emitted to each medium to be considered; b) missing guidelines for greenhouses (Are they part of the technosphere or the ecosphere?); c) other issues related

to formulation additives and pesticide degradation products; d) potential double counting with land occupation (in some LCIA methods, characterisation factors for land occupation may already account for reduced biodiversity due to pesticides used); e) consideration of water quality in water footprinting (classically, for water pollution caused by pesticides only the fraction directly going to water is considered in water footprints, but if a pesticide is emitted to soil and air, a fraction of these emissions will also end up in water via e.g. deposition, runoff, or leaching. How should this be accounted for?); f) default values should be applied to generic conditions and the selected model should be flexible to adapt values representing local conditions.

Teunis Dijkman (DTU) presented and compared three examples of current practice in LCI modelling (Ecoinvent, U.S. LCI, and PestLCI 2.0, see Table 2), with a focus on the delineation between technosphere and ecosphere and transport processes considered in PestLCI 2.0. He concluded that a) different LCI approaches and system boundary choices lead to incomparable LCA studies where toxicity results can differ by orders of magnitude, depending on the emission model chosen; b) modelling fate processes in LCI does not lead to double counting of fate processes; c) clearly defining a technosphere-ecosphere boundary is essential.

Peter Fantke (DTU) provided an overview of current LCIA modelling practice. Among other he pointed out a number of potentially relevant agricultural chemical emissions which are not yet covered by LCIA methods. Examples are pesticide formulation by-products (adjuvants, solvents, etc.), phytohormones used as plant growth regulators, animal manure by-products (livestock antibiotics and hormones), sewage sludge constituents (metals, bacteria, etc.), and degradation products (metabolites). “None of these are covered in current LCIA practice, but can we nevertheless assume the same conditions regarding system boundaries?” He identified open questions regarding consistent system boundaries for LCIA: a) How to address species migrating between technosphere and ecosphere? b) How to address species living from the technosphere (e.g. birds feeding on insects living in agricultural fields)? c) How to consider applied mass for impacts from crop intake? He also named a number of LCIA important principles for pesticides: a) include predominant human exposure pathways: intake of treated crops; b) consider time dimensions for exposure via crop intake; c) allow comparison with exposed population sub-groups (e.g. bystanders, operators, etc.).

Rosalie van Zelm (Radboud University) presented work done in collaboration with Irstea (France) focusing on the boundary of LCI and LCIA so as to frame what modelling should be included in LCI and what in LCIA in order to prevent gaps and overlaps (van Zelm et al. 2014). She outlined the issues that we further and more thoroughly discussed in the meeting. Finally, a case study on bananas was presented applying the proposed framework. It showed the importance of a good integration of LCI-LCIA methodology allowing to quantify differences between efficient vs inefficient agricultural practices taking into account pre- and post-treatment pesticide flows (e.g. cleaning of equipment), use of buffer zones, and (over)dosing. This work was recently published (van Zelm et al. 2014), proposing a framework to bridge the gap and prevent overlaps between LCI and LCIA for toxicological assessments of pesticides.

With these inputs, the discussion was launched with a number of issues and open questions identified beforehand as starting points for the discussion:

- Is the ecosystem on the agricultural field (e.g. aerial such as insects and pollinators, terrestrial and deep soil ecosystems) part of the area of protection (natural environment) and should thus be addressed in the impact assessment or is it a highly controlled ecosystem anyway and essentially part of the technosphere and then (at least partially) covered in the land use impact category (which inherently includes changes in biodiversity due to pesticide application as part of some land-use types)?;
- Does it lead to double counting if both the inventory model and the impact model involve a fate modelling step or can these be distinguished clearly and consistently across substances and conditions?;
- Spatial dimensions: If the field is considered as part of the technosphere, can we define the agricultural field as a 3-dimensional box by height of air above the field, width and length of the field including

buffer zones, and soil depth? Then only the fraction(s) of an applied pesticide leaving this box to various environmental media would be accounted for in the emission inventory;

- Temporal dimensions: Pesticides are typically applied in pulses. Does the assumption of similarity between pulse and continuous emissions in the steady-state model of LCA also hold for such emission dynamics? How can it be determined, when steady-state assumptions are satisfactory and when not? Which periods of time from application to ecosphere emissions need to be taken into account in the emission inventory and in the impact assessment?;
- Mass applied vs. mass emitted – what information does the emission inventory need to provide? Which emission media are relevant to cover: air, surface water, ground water, deeper soil layer(s), others? Intake of pesticide residues in food products may be the dominating human exposure pathway in most cases, but requires the inventory to also report the quantity applied directly to the crop and not just the emission to the environmental media. Therefore, a clear definition of what needs to be reported and how in the life cycle inventory is needed;
- Which impact pathways (e.g. direct and indirect human exposure via food residues, inhalation, drinking water, or terrestrial/aquatic/aerial ecosystems on-/off-field, etc.) and indicators (human toxicity, ecotoxicity, species richness, functional diversity, etc.) should be covered for pesticides?;
- Are different recommendations required for specific cases such as greenhouses, irrigated vs. non-irrigated fields, etc.?

The discussions during the workshop were comprehensive, focused, and very productive, resulting in the recommendations presented in the following section.

### **3 Results and discussion**

A balanced assessment of toxicity effects due to pesticide emissions (i.e. to environmental media, which is not the same as the amount of pesticide used/applied) associated to their agricultural use in the context of life cycle assessment (LCA) is strongly desirable, allowing for consideration of both advantages and disadvantages of using pesticides or choosing an alternative approach instead. Recommended practice should enable the assessment of trade-offs between potential negative biodiversity impacts pesticides may cause on and around the field and positive indirect effects on biodiversity due to increased yield and thus less land-use compared to avoiding pesticide application.

In Table 3, the main results from the discussion around the essential core questions of allocation of the agricultural field to technosphere or ecosphere, as well as its spatial and temporal dimensions are presented. These questions were the starting point of the discussion during the workshop and the initial goal was to arrive at recommendations to answer the questions. However, it became clear that there is a strong dependency on goal and scope of an LCA study and that fixing these would not allow certain assessment goals. Instead, the focus was on defining a clear interface between LCI and LCIA. This should be capable of supporting any goal and scope requirements while avoiding double counting or excluding important emission flows (and their potential impacts). The detailed recommendations are given in the section Conclusions and recommendations.

(table 3 here)

### **4 Conclusions and recommendations**

The boundary between technosphere/ecosphere and allocation of the agricultural field to either technosphere or ecosphere and its spatial dimension was and should not be fixed on a general level, but depend on the goal and scope of the LCA study. Instead, we recommend a number of emission flows, both on and off the field, that need to be quantified in the LCI in order to characterise specific impacts, which may take place on or off the field.



Consensus was reached on distinct sets of recommendations for LCI and LCIA respectively, which are presented hereafter. This establishes a clearly defined interface between LCI and LCIA.

Life cycle inventory recommendations:

- For pesticides application, the LCI should report the following information (per functional unit):
  - Pesticide identification (CAS registry numbers and names of active ingredients per applied plant protection product, PPP);
  - Crop;
  - Mass applied of each active ingredient;
  - Application method (foliar spray, soil injection, drip irrigation) or formulation type (soluble concentrate, granule, wettable powder);
  - Presence of buffer zones (y/n);
  - Location/country (which identifies a default local scenario/site incl. climate, soil conditions, present environmental media, ...);
  - Application time in days before harvest (for human health impacts) and crop growth stage (for terrestrial/aerial ecosystem impacts) during application;
  - Adherence with Good Agricultural Practice(s) – GAP, such as discussed and defined by FAO (2003) (with adjustable parameters including pre- and post-treatment pesticide flows, e.g. equipment cleaning, tanks and cans residues, etc.), in order to be able to distinguish different levels of agricultural practice from good (i.e. efficient) to average to inefficient (e.g. overdosing, application during precipitation events, etc.);
  - The goal and scope section of an LCA should clearly state whether the agricultural field is assumed as part of the technosphere or the ecosphere.
- We recommend the explicit reporting of emission fractions, i.e. fractions of the applied pesticide emitted to environmental media via initial distribution from the applied mass based on models and/or measurements with mass fractions going directly to:
  - On-field:
    - soil surface (excluding plant deposit);
    - paddy water (excluding plant deposit), e.g. for paddy rice;
    - plant (plant deposit);
  - Off-field:
    - air (via volatilization on-field and wind drift, fraction not deposited close to the field);
    - surface water (via spray drift deposition, excluding runoff, avoiding double counting with air (of which a fraction eventually redeposits on surface water at long distance); distinguishing fresh and marine water as far as possible and meaningful;
    - soil (distinguishing types [natural, industrial, agricultural, etc.] if possible);
- The emission fractions to environmental media via initial distribution and the degraded fraction should always sum up to 100% of the quantity applied to the field in order to maintain conservation of mass and to allow for assessment of all potential impacts in the LCIA;
- There should be archetypical/default sets of values for these emission fractions to environmental media via initial distribution depending on conditions of applications that could be replaced if measured or more detailed information is available. These will be quantified and agreed upon in a follow-up workshop in Basel in May 2014;
- As a starting point, default values may correspond to current average agricultural practice that may be spatially differentiated;
- Good practice in greenhouses should lead to minimum emissions, but fractions on greenhouse soil surface (excluding plant deposit), plant, leaching through the top soil layer and emissions to air outside the greenhouse should be quantified;

- In case buffer zones are present, they should be considered as part of the crop (production system) and the change in yield per ha and the emission fractions to environmental media via initial distribution to off-field soil/surface water accounted for accordingly. Therefore, no specific CFs are required for buffer zones as these will be treated the same way as the rest of the agricultural field;
  - Runoff, accounted for in the impact assessment, would have to be modelled accordingly;
- Care should be taken to avoid double counting between the multimedia transfers of these emission fractions and the impact assessment fate model, thus the decision to only have the emission to environmental media via initial distribution in the inventory and subsequent transfers considered in the impact assessment. This is supported by furthermore distinguishing between initial distribution processes in LCI, where steady state conditions are typically not (yet) reached and a dynamic assessment is required (Rein et al. 2011; Fantke et al. 2013), and fate processes in LCIA, for which steady state is typically assumed based on continuous diffusive, advective, and degradation processes.

#### Life cycle impact assessment recommendations:

- Relevant impact categories (with temporal aspects built-in):
  - Land-use biodiversity impact on-field;
  - Toxicity biodiversity impact off-field (terrestrial, aerial, freshwater, ground-water);
  - Toxicity human health off-field (including residues in treated crops for subsequent human exposure via crop consumption as a potentially predominant exposure pathway, bystander);
  - Toxicity human health on-field, e.g. occupational exposure, should be further explored;
- LCIA methods should allow for the flexibility to define the agricultural field as part of the technosphere or the ecosphere in the goal and scope of a study (see Table 3) and CFs should be clearly labelled or documented as representing potential impacts in the ecosphere or the technosphere (e.g. terrestrial ecotoxicity on agricultural soil) if a specific assumption is required;
- Impact assessment should model:
  - Fate:
    - secondary drainage;
    - runoff;
    - degradation;
    - volatilization;
    - transfer (leaching) to ground-water and deeper soil layers on the field;
    - residues in entire crop of which some parts might remain on-field after harvest → leading to possible subsequent runoff/leaching and resulting background concentrations, or to exposure via fodder for animals or compost (with pesticide residues);
  - Human exposure, eventually including bystanders, applicators, field workers, etc.;
- The plant fraction washed off is not yet considered in impact assessment, only the fraction effectively transferred from root to soil could be added to the primary soil emission until properly modelled. Whether washoff can/should be included or not is a function of the scenario, i.e. at good or average agricultural practice, washoff is negligible, for application shortly before rain events, washoff may become relevant or even dominant;
- Compatibility between the dynamic fate modelling for the initial emission fractions and the subsequent steady-state fate modelling in LCIA needs to be acquired to avoid double counting or omissions of mass transport processes. This notably involves the consideration of the temporary boundary applied to the dynamic emission model;
- Impact assessment methods should also report which multi-media transfers are considered and recommend how the environmental medium may be linked to the mass applied;
- Open issues:
  - What is the influence of irrigation and how should it be modelled?;

- Care should be taken to avoid potential double counting between terrestrial ecotoxicity on agricultural soil and land use impacts on biodiversity due to agricultural soil use. Therefore, the characterisation models for these impact categories need to be harmonised (i.e. a decision is required on how and in which impact category these are modelled) when developing impact assessment methods for these impact categories;
- Further observation raised during the meeting:
  - Ecotoxicological impacts linked to background concentrations from earlier applications (e.g. with persistent substances) after crop harvest as function of crop rotation practice need to be accounted for;
  - For direct impact on the on-field soil, only the additional impact on ecosystem due to the PPP application (above the already considered effect of agricultural land use) should be considered when we are comparing different scenarios.

Next steps and future work discussed and identified during the workshop focused on the operationalisation of the recommendations in order to integrate them into LCA practice. Identified actions were:

- Establishment of recommended default emission fractions to environmental media via initial distribution including the definition of the underlying temporal dimension of the dynamic LCI modelling. The results should be integrated into LCI databases. This is the goal of a follow-up workshop on 10<sup>th</sup> May 2014 in Basel;
- Organisation of interaction with LCI database developers in order to help implement the recommendations;
- Interaction among LCIA developers to extend current methods with new elements/processes mentioned in the recommendations, including targeted technical workshops on “how to” model specific processes;
- Launching of a similar initiative to establish consensus about fertilizer emissions and impact assessment.

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## Tables

Table 1: Overview of available literature focusing on fate and transport processes of pesticides after application to agricultural fields

Process/factor	References
Spray drift including deposition	(Bilanin et al. 1989; Ganzelmeier et al. 1995; Holterman et al. 1997; Bird et al. 2002; Teske et al. 2002; Siebers et al. 2003; Gil and Sinfort 2005; Gil et al. 2007; Gil et al. 2008; Wang and Rautmann 2008; Teske et al. 2009; Butler-Ellis et al. 2010; Lebeau et al. 2011)
Volatilisation	(Kubiak et al. 1995; Rüdell 1997; Bedos et al. 2002a; Scholtz et al. 2002; Bedos et al. 2002b; Wolters et al. 2003; Guth et al. 2004; Leistra et al. 2005; van den Berg et al. 2008; Bedos et al. 2009; Davie-Martin et al. 2013b; Davie-Martin et al. 2013a)
Leaching	(Persicani 1996; Sorensen et al. 1998; Armstrong et al. 2000; Larsson and Jarvis 2000; Loll and Moldrup 2000; Harris et al. 2000; Geisler et al. 2004; Kruijne et al. 2004; Leterme 2006; Rosenbom et al. 2009; Akbar and Lin 2010; Vanderborgh et al. 2010; Vanderborgh et al. 2011)
Runoff	(Huber et al. 1998; Sexton et al. 2000; Kellogg et al. 2002; Inao et al. 2003; Miao et al. 2003; Liu et al. 2004; Berenzen et al. 2005; Numabe and Nagahora 2006; Schriever et al. 2007; Phong et al. 2011; Thuyet et al. 2012)
Washoff	(Smith and Carsel 1984; Cohen and Steinmetz 1986; Willis et al. 1994; Fisher et al. 2002)
Plant uptake	(Trapp and McFarlane 1995; Fryer and Collins 2003; Antón et al. 2004; Juraske et al. 2007; Juraske et al. 2009; Collins et al. 2011; Fantke et al. 2011a)
Buffer zones	(De Snoo and De Wit 1998; Matthews and Piggott 1999; Lin et al. 2002; Lacas et al. 2005; Madrigal et al. 2007; Passeport et al. 2011)
Application method, temperature, crop/soil type, and degradation	(González et al. 2001; Boesten et al. 2006; Kah et al. 2007; Fantke and Juraske 2013; Fantke et al. 2014)
Greenhouses	(Paul and Illing 1997; Antón et al. 2004; Garratt et al. 2007)

Table 2: Comparing four life cycle emission inventory approaches: percentages emitted to environmental media per unit of 2,4-D applied to corn

Approach	Ecoinvent	U.S. LCI	PestLCI 2.0	USDA LCA Digital Commons
Emission medium	%	%	%	%
Soil	100	–	–	100
Air	–	96	0.61	–
Surface water	–	4	0.02	–
Ground water	–	–	0.62	–
Degradation, crop uptake, etc.	–	–	98.7	–

Table 3: Recommendations for the definition of an interface between LCI and LCIA for agricultural pesticide emissions

Topic	Recommendation	Reasoning
Boundary between technosphere/ecosphere and allocation of the agricultural field to either technosphere or ecosphere	No fixed definition of the boundary or allocation of the agricultural field to either technosphere or ecosphere is defined. Instead, we recommend a number of emission flows, both on and off the field, that need to be quantified in the LCI in order to characterise specific impacts, which may take place on or off the field. Then the choice of characterisation factor defines which impacts are characterised. Furthermore, the consideration of the agricultural field as part of technosphere or ecosphere should be clearly defined in the goal and scope of a study.	Potential impacts may occur both on and off the agricultural field. The consideration of the agricultural field as part of technosphere or ecosphere therefore depends on the goal and scope of a LCA study and should not be fixed, but defined in each study correspondingly. Instead, the definition of emission flows to be quantified in the LCI (with specific impact pathways in mind) allows for LCIA developers to provide CFs for specific impact pathways, which may represent impacts on or off the field. This defines a clear interface between LCI and LCIA rather than fixing which impacts can be considered in an LCA.
Buffer zones	Buffer zones should be considered as part of the crop production system and the change in yield per ha and the emission fractions to environmental media via initial distribution to off-field soil/surface water accounted for accordingly.	Buffer zones are a measure to limit off-field emissions, but affect yield. Therefore the presence of buffer zones is an important factor affecting potential impacts that needs to be accounted for, especially when considering different agricultural practice (e.g. intensive vs. extensive).
Spatial dimensions of the agricultural field	Spatial dimensions are dynamic and may depend on various factors, such as crop and pesticide type and its use, the goal and	For both the spatial and temporal dimensions, the essential requirement is compatibility between LCI and LCIA modelling, which requires the definition of

	scope of the LCA, etc. Therefore, the spatial dimensions of the agricultural field should not be fixed, but clearly defined in the study's goal and scope.	an interface, since different actors are involved, which are typically disconnected. If the LCI requirements are clearly set, the LCIA developers can build on these to represent potential impacts based on available impact pathways and models. See "Emissions in the LCI" for further details on recommendations for a definition of this interface.
Temporal dimensions of the agricultural field	Temporal dimensions are dynamic and may depend on various factors, such as crop and pesticide type and its use, etc. The temporal dimensions of the agricultural field need to be fixed and will be defined in a follow-up workshop dealing with the definition of default emission fractions to environmental media via initial distribution.	The temporal dimensions of the agricultural field need to be fixed in order to enable the dynamic modelling of the initial distribution after application (see also under "Emissions in the LCI").
Greenhouses	Greenhouses are always part of the technosphere and emissions to air or soil outside the greenhouse should be quantified.	A greenhouse is a closed and controlled environment, accessible only to a very limited extent for species from its outside.
Emissions in the LCI	LCI emissions represent an initial distribution to environmental media both on and off the field.	Double counting between the fate modelling of on-/off-field emission fractions and the impact assessment fate model has to be avoided. Therefore, the emission to environmental media on-/off-field should be modelled dynamically via an initial distribution in the LCI for a specific time span after application (where steady state conditions are typically not (yet) reached and a dynamic assessment is required). Subsequent transfers between off-field environmental media are considered in the LCIA (where steady state is typically assumed based on continuous diffusive, advective, and degradation processes).

## Figures

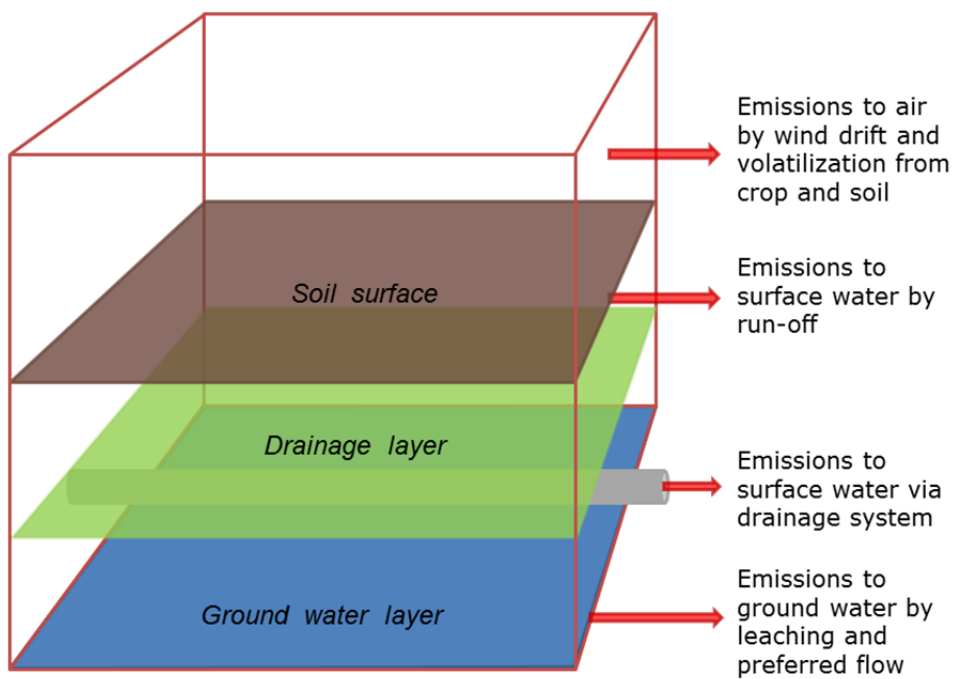


Figure 1: Conceptual representation of emissions to various environmental media occurring after application of a pesticide to a crop on an agricultural field (based on Birkved and Hauschild (2006))