1							
2	TRENDS IN OZONE CONCENTRATIONS IN THE IBERIAN						
3	PENINSULA BY QUANTILE REGRESSION AND CLUSTERING						
4 5	A. Monteiro ^{*1} , A. Carvalho ¹ , I. Ribeiro ¹ , M. Scotto ² , S. Barbosa ³ , A. Alonso ⁴ , J. M.						
6	Baldasano ^{5,6} , M. T. Pay ⁵ , A.I. Miranda ¹ , C. Borrego ¹						
7	¹ CESAM & Department of Environment and Planning, University of Aveiro, Aveiro, Portugal						
8	² Department of Mathematics, University of Aveiro, Aveiro, Portugal						
9	³ Center of Geophysics, IDL, University of Lisbon, Lisbon, Portugal						
10	⁴ Departamento de Estadística, Universidad Carlos III de Madrid, Spain						
11 12	⁵ Earth Science Department, Barcelona Supercomputing Center, Jordi Girona 29, Edificio Nexus II, Barcelona, Spain ⁶ Environmental Modeling Lab, Technical University of Catalonia, Barcelona, Spain						
13	*Corresponding author: A. Monteiro, e-mail: <u>alexandra.monteiro@ua.pt;</u> Tel: +351 234 370220, Fax: +351 234						
14	570309						
16	Abstract						
17	In this paper, 10-years of ozone (O_3) hourly concentrations collected over the period 2000–2009 in the						
18	Iberian Peninsula (IP) are analyzed using records from 11 background sites. All the selected monitoring						
19	stations present an acquisition efficiency above 85%. The changes in tropospheric ozone over the Iberian						
20	Peninsula are examined by means of quantile regression, which allows to analyse the trends not only in						
21	the mean but in the overall data distribution. In addition, the ozone hourly concentrations records are						
22	clustered on the basis of their resulting distributions.						
23	The analysis showed that high altitude stations (> 900 m) have higher background O_3 concentrations (~80						
24	μ g.m ⁻³). The same magnitude of background O ₃ concentrations is found in stations near the						
25	Mediterranean Sea. On the other hand, the rural stations near the Atlantic coast present lower background						
26	values (~50-60 μ g.m ⁻³) than those of Mediterranean influence. The two sub-urban stations exhibit the						
27	lowest background concentrations (~45 μ g.m ⁻³). The results of the quantile regression show a very						
28	distinct behaviour of the data distribution, the slopes for a fixed quantile are not the same over IP,						
29	reflecting the spatial dependence of O_3 trends. Hence the rate of temporal change is not the same for all						
30	parts of the data distribution, as implicitly assumed in ordinary regression. The lower quantile (percentile						
31	5) presents higher rates of change than the middle (percentile 50) and the upper quantile (percentile 95).						
32	The clustering procedure reveals what has been already detected in the quantile regression. The station						
33	with highest rates of decrease on the O ₃ concentrations (easternmost station of IP) is isolated and then						
34	other clusters are formed among the moderately positive/negative O ₃ trends around the IP. The clustering						
35	procedure highlighted that the largest trends are found for the lower ozone O ₃ values, with largest						
36	negative trend at the easternmost station of IP, and also in northern and mainland stations, and an opposite						
37	behaviour, with positive O_3 trends, is observed at the Atlantic coast stations.						
38							
39							
40							

Keywords: Iberian Peninsula; tropospheric ozone; spatial and temporal analysis; quantile
 regression; cluster procedure

43

44 **1. INTRODUCTION**

Tropospheric ozone (O_3) is a key determinant of the atmospheric oxidation state and a major constituent of photochemical smog which impacts air quality at urban and regional scale. The production of elevated levels of O_3 at ground level is of particular concern because it is known to have adverse effects on human health, vegetation, and a variety of materials (EA, 2010a). There is a high interest in quantifying surface O_3 concentrations and associated trends, as they serve to indirectly quantify the impacts of the anthropogenic precursor reductions and to evaluate the effects of emission control strategies (Tang et al., 2006; Sicard et al., 2009).

There have been a few studies on the analysis of surface O_3 trends in different regions of Europe (Brönnimann et al., 2002, Jenkin, 2008; Sicard et al., 2009). Over the Iberian Peninsula (IP) where high surface O_3 concentrations are monitored each year from April to September (EEA, 2010b), several analyses of surface O_3 concentrations have been carried out. However, they were limited to a single location restricted to a region of the IP and adopting an ordinary regression approach or based on the median/mean and high percentile O_3 analysis (Gimeno et al., 1999; Millán et al., 2002; Ribas and Peñuelas, 2004; Adame et al., 2008).

59 Observations from background monitoring stations have revealed that baseline surface O_3 60 concentrations in the northern hemisphere have been increasing over the past three decades 61 (Marenco et al., 1994), with average increases of approximately 0.5–2% per year at northern mid-latitudes (Vingarzan, 2004). The observed increasing trend in baseline O3 concentrations is 62 63 believed to be driven by emissions and processing of O_3 precursors on a global scale (Jaffe et 64 al., 2003; Honrath et al., 2004; Derwent et al., 2006, 2007). Although this hemispheric baseline 65 influences O₃ concentrations throughout the IP, the observed concentrations can be further 66 modified by processes occurring on regional- and local-scales, which can both increase and 67 decrease O₃ levels. Therefore such processes occurring on local, regional, and global scales 68 have an influence on whether O_3 air quality standards at a given location are achieved (Jenkin, 69 2008). Although the progressive control of O₃ precursors emissions -like volatile organic 70 compounds (VOC) and nitrogen oxides (NO_x) - within the European Community since the early 71 1990s (CEC, 1991) have influenced the magnitude of the O_3 regional- and local- scale effects 72 (Derwent et al., 2003; Jonson et al., 2005; Vautard et al., 2006), the observed O_3 trends is 73 determined from the net trend of the global-, regional- and local-scale effects, the relative 74 contributions of which can vary both spatially and temporally. 75 Anthropogenic emissions of the main air pollutants across Europe have decreased continuously

between 1990 and 2008 in Europe (EEA, 2010a). Reported European emissions of NO_x and NMVOC have both decreased by 39% and 51%, respectively, since 2000. Concerning Spain

and Portugal, the O_3 precursor emissions have also been reduced, namely for NO_x (8% and 7%, respectively) and NMVOC (21% and 34%, respectively).

80 The understanding of the O_3 budget and trends in the troposphere over the IP is required to (a) 81 properly identify the various mechanisms that contribute to the observed hourly average 82 concentration distribution; and to (b) develop and test models capable of simulating and 83 predicting atmospheric chemical and physical processes (Lefohn et al., 2008). It is also 84 important to characterize the changes in the distribution of hourly average O₃ concentrations 85 which provide (a) quantitative feedback on the effects of emission reductions on O_3 86 concentrations; (b) insights concerning the long-range transport of O_3 outside IP and possible 87 impacts of climate change; and (c) important information on which processes dominate during a 88 specific time of the year and which processes are more likely to influence particular portions of 89 the distribution (Oltmans et al., 2006).

90 Robust statistical procedures can be applied to investigate the spatial and temporal evolution of 91 the O_3 concentrations over a region from historical datasets. This study adopts the method 92 introduced by Barbosa et al. (2011) which combines quantile regression and clustering 93 procedures in order to better assess the spatial and the temporal evolution of the hourly O_3 94 measurements over the IP. On the one hand, quantile regression (Koenker and Hallock, 2001) 95 provides the rate of change not only in the mean, as in ordinary regression, but also in all parts 96 of the data distribution. In this sense, the quantile regression quantifies the variability structure 97 of the hourly O_3 concentrations and assesses the changes in the data distribution. On the other 98 hand, cluster analysis is an adequate procedure to spatially characterize the regional variability 99 on the O₃ data and it has been widely used in different analysis of environmental processes 100 (Alonso et al., 2006; Scotto et al., 2009; Barbosa et al., 2011; Carvalho et al., 2011).

101 This work focuses on investigating the temporal and spatial trends of the hourly surface O_3 102 concentrations at background environment over the IP for the last decade (2000-2009). The 103 remainder of this paper is laid out as follows. Section 2 discusses the O_3 concentrations acquired 104 at the background monitoring stations used in this study. Section 3 describes the application of 105 the quantile regression approach and the clustering procedure. Results are presented in Section 106 4. Finally, in Section 5 the results are discussed and main conclusions are summarized.

107

108 **2. O**₃ **DATA OVER THE IBERIAN PENINSULA**

109 A total of 11 O_3 monitoring stations within the IP are selected taking into account their 110 background influence and the efficiency data collection (> 85%) during the 10-years period 111 (2000-2009) as shown in Fig. 1 and Table 1. The spatial coverage is suitable over the IP, with 112 two stations located in Portugal and the remaining 9 stations located over Spain.

- 113
- 114

115	
116	(Table 1)
117	
118	Ambient O3 concentrations are reported on an hourly basis and were obtained from the
119	Portuguese Air Quality Database (<u>www.qualar.org</u>) and the EMEP monitoring network
120	(www.emep.int).
121	Fig. 2 shows the distribution of hourly O3 concentrations by year. Different groups of stations
122	can be distinguished in terms of background values (median) and minimum/peak values.
123	
124	(Figure 2)
125	
126	High altitude stations (> 900 m) (CPB, VIZ and PEN) show high background concentration
<mark>127</mark>	(~80 μ g.m ⁻³) due to higher O ₃ levels in elevated terrains. The same range of background O ₃
128	concentrations are found in stations near the Mediterranean Sea (CCR, ZAR, and TOR). O_3
129	atmospheric dynamics in the Spanish Mediterranean areas is affected by mesoscale and local
130	meteorological processes but also regional factors, such as (Baldasano et al., 1994, Millán et al.,
131	1997; Toll and Baldasano, 2000; Martin-Vide and Olcina, 2001; Soriano et al., 2001; Pérez et
132	al., 2004): (1) the influence of the Azores high-pressure system, (2) the costal ranges
133	surrounding the Mediterranean coast, (3) the influence of the Iberian and Saharan thermal lows
134	causing weak pressure gradients over the Mediterranean (4) the intense breeze action along the
135	Mediterranean coast favoured by the prevailing low advective conditions, (5) the scarce summer
136	precipitation, and (6) the intense seasonal contrast concerning temperature, humidity and
137	rainfall. All these facts favour the photochemical formation of O_3 and contribute to the
138	accumulation and recirculation of aged air masses which contain O ₃ . The two rural stations
139	closest to Portugal and located under 506 m of altitude -BAR and SEV - register lower median
140	(~50-60 μ g.m ⁻³) than those of Mediterranean influence, also presenting O ₃ peaks (P95) less than
<mark>141</mark>	120 µg.m ⁻³ . The NIB station, in the northern IP, also presents low median concentrations along
<mark>142</mark>	the decade (~50-60 μ g.m ³) due to the influence of large plumes coming from power plants
<mark>143</mark>	located in northeaster Spain (Pay et al., 2011) containing high NO_x concentration that affects O_3
<mark>144</mark>	chemistry in this region. Such episodes happen under the influence of westerly winds which are
145	relatively frequent (Jorba et al., 2004). The two suburban stations (CUS and PP) exhibit the
146	lowest median (~45 μ g.m ⁻³) and the minimum O ₃ concentrations (P5) (~ 0 μ g.m ⁻³), explained by
147	the O_3 destruction by NO (emitted by road-traffic and shipping in the urban and suburban areas
148	of Oporto and Lisbon) mainly at nigh-time (Seinfeld and Pandis, 1998).
149	During the study period the most critical years in terms of O ₃ peaks/episodes were 2005 and

150 2006 for the majority of the stations. The summer period of these two years was characterized

151 by meteorological conditions very favourable for photochemical activity (Monteiro et al., 2005,

152

155 **3. STATISTICAL METHODS**

156 Quantile regression is a well-defined statistical technique for regression on quantiles rather than 157 regression on the mean. Although it was first introduced in econometrics by Koenker and Basset 158 (1978), quantile regression is being applied in various geoscience contexts (e.g. Koenker and 159 Schorfheide, 1994; Cade and Noon, 2003; Baur et al., 2004; Elsner et al., 2008; Barbosa et al., 160 2011). We outline here the essential of the quantile regression approach. The starting point is a 161 random variable Y with cumulative continuous distribution function $F_{Y}(y)$ (by definition: FY(y) = P(Y \le y)). The quantile τ is defined as the value $Q_Y(\tau)$ such that $P(Y \le Qy(\tau)) = \tau$, for $0 \le \tau \le 1$. 162 163 The quantile function $Q_{Y}(\tau)$ is defined from the cumulative distribution function $F_{Y}(y)$ as $Q_{Y}(\tau)$ $= F_{Y}^{-1}(y)$. Then considering the conditional distribution of Y given X=x, the conditional quantile 164 function $Q(Y|X)(\tau|x)$ verifies $P(Y \le Q(Y|X)(\tau|x)|X=x) = \tau$. Whereas ordinary regression is based 165 166 on the conditional mean function E(Y|X)=x and minimization of the respective residuals, 167 quantile regression is based on the conditional quantile function and minimization of the sum of

2007). The year 2003 was also a particular critical year in terms of photochemical activity (and

high O₃ values) due to the occurrence of a strong heat wave over the IP (Ordonez et al., 2010).

168 asymmetrically weighted absolute residuals $\sum_{i\geq 1} \rho(\tau)(y_i - Q_{(y|x)}(\tau | x = x_i))$, where $\rho(.)$ represents 169 the tilted absolute value function. For further details see Koenker (2005).

170 The time series clustering procedure proposed to classify the time series of O_3 hourly 171 concentrations based on the corresponding distributions for quantile slopes at lower, middle and 172 upper quantiles is as follows: firstly, for a fixed (but arbitrary) quantile, the algorithm starts with 173 the estimation of the distribution corresponding to quantile slope estimates; second, the 174 corresponding *dissimilarity* matrix is computed. To this extend, an adequate metric between 175 univariate distribution functions is required. In the present setting the weighted L2-Wasserstein 176 distance between two quantile slope distributions is adopted. Finally, a dendrogram based on the 177 application of classical cluster techniques to the dissimilarity matrix is built and that provides 178 the different clusters formed by the distributions of the quantile slopes. In particular, 179 agglomerative hierarchical methods with nearest distance (single linkage), furthest distance 180 (complete linkage) and unweighted average distance (average linkage) are used as grouping 181 criteria. In order to summarise those distributions, the average linkage procedure is applied to 182 obtain dendrograms of slopes for quantiles 0.05, 0.5 and 0.95. Similar conclusions are obtained 183 using the single linkage and the complete linkage methods.

184

185 **4. RESULTS**

186 In this section, quantile regression is applied for the hourly O_3 concentrations in order to 187 describe the temporal variability of different quantiles of the O_3 distribution over IP. The

188	quantile slopes and corresponding standard errors are derived using the algorithm of Koenker						
189	and D'Orey (1987). The clustering procedure is also discussed.						
190	The results for all the stations are shown in Fig. 3, along with the quantile slopes at quantiles						
191	0.05, 0.5 and 0.95 , corresponding respectively to the lowest 5%, 50% (median) and 95% of the						
192	ordered observations.						
193							
194	(Figure 3)						
195							
196	Several O ₃ trends over the last decade can be identified in this group of stations. A significant						
197	negative trend is only exhibited by CCR station, especially for lowest quantiles (P5). The same						
198	tendency was found by Ribas and Peñuelas (2004) for a coastal station (Begur) in northeastern						
199	Spain. CCR is a costal station located in the northeastern extreme of the IP. This site presents						
200	strong north-westerly winds (tramontane and mistral) channelled by Pyrenees and Central						
<mark>201</mark>	Massif throughout the Gulf of Lyon. The flow crosses the Carcasone gap into the Mediterranean						
<mark>202</mark>	which can transport new pollutants into the area that are added to local emissions and re-						
<mark>203</mark>	circulated within the coastal breezes at eastern Iberian (Gangoiti et al., 2001). CPB, PEN –						
204	located in the northern Spanish plateau - and SEV show a slightly negative slope, mainly for the						
205	lower quantiles. BAR and ZAR monitoring sites don't show any significant trend for the three						
206	quantiles.						
207	By contrast, the NIB coastal station in the northern IP presents the largest positive trends, even						
208	larger for lower concentration (P5). Similar trends are found in TOR and VIZ, sited under the						
209	Mediterranean influence, and in a lesser extend at the two suburban stations at Oporto and						
210	Lisbon cities (CUS and PP, respectively).						
211	A more complete description of the quantile regression results is displayed in Fig. 4 which						
212	displays the quantile slopes and the corresponding standard errors computed for quantiles 0.1 to						
213	0.9 in steps of 0.02.						
214							
215	(Figure 4)						
216							
217	Fig. 4 clearly shows a distinct pattern for the different monitoring sites. However, there are						
218	similarities between specific stations in terms of the sign and the distribution over the different						
219	quantiles. CCR shows the highest negative slopes over all the analysed sites (from -28 to -19						
220	μ g.m ⁻³ /decade), with a higher decrease observed for the lower quantiles. A negative slope over						
221	the all ranges of concentrations is also registered for the north mainland stations - CPB, SEV						
222	and PEN - with similar magnitudes (around -5 and -10 μ g.m ⁻³ /decade) of the quantile						
223	distribution pattern. A slight negative slope (> -2.5 μ g.m ⁻³ /decade) is also verified for ZAR and						
224	BAR, but only for the lower quantiles.						

225 On the opposite, the Atlantic coastal stations - NIB, CUS and PP - have positive slopes over the 226 all concentrations range with the lower increasing at a much faster rate than the middle and 227 upper values. Besides a similar quantile distribution, the magnitude of the slope is significantly 228 different, higher for NIB (> 18 μ g.m⁻³/decade) and lower for CUS and PP (~ 5-15 μ g.m⁻ 229 ³/decade). Positive slopes are also found for the VIZ and TOR stations (4-12 μ g.m⁻³/decade),

both presenting specific and unique quantile distribution.

231 For all cases the derived slopes vary with the quantiles and are distant to the original ordinary

232 least squares slope, indicating that the distribution of the ozone values is not symmetric and the 233 rate of change is not the same for all parts of the data distribution (lower, middle and upper 234 quantiles behave differently).

In summary over the last decade a group of stations – CCR, CPB, ZAR, BAR, SEV and PEN – registered a decrease mainly on the lower quantiles of O_3 data distribution which reflect the minimum (nocturnal) values over these areas. On the opposite, the rest of the monitoring sites – NIB, PP, CUS, TOR and VIZ – exhibit a high positive slope on these lower quantiles, indicating

an increase over the background values of ozone.

Furthermore, the results of the clustering procedure, together with the spatial representation ofthe quantile slopes, are shown in Fig. 5.

(Figure 5)

- 242
- 243 244

245 The dendrogram for the lower quantile (P5) clearly discriminates three groups: stations with 246 larger negative slopes, \cong - 28 µg.m⁻³/decade (CCR), slight negative slopes (BAR, SEV, CPB, 247 PEN and ZAR) and the remaining stations with positive slopes (NIB, PP, CUS, TOR, VIZ). 248 These results corroborate the previous analysis, namely in what concerns the different trend on 249 the background ozone values registered over Iberian Peninsula. The second cluster, with 250 positive slopes, further distinguishes the station with the highest slope, $> 18 \ \mu g.m^{-3}/decade$ 251 (NIB) from the other stations. The third cluster, with negative slopes, further subdivides into 252 sites with moderate slopes and stations with very small or non-significant trends (BAR). A 253 similar pattern is found in the dendrogram for the median quantile (P50), with the same groups 254 identified.

The dendrogram for the upper quantile (P95) continues to distinguish the CCR station with the highest negative slopes, > -20 μ g.m⁻³/decade. Within the remaining stations, and differing from the previous dendograms, the major subdivision clusters include (1) the rural stations with positive trend (TOR, NIB and VIZ) and (2) all the other stations with negative slopes and the two suburban stations. This last cluster is then subdivided into two clusters of slight/moderate slopes (ZAR, CUS, BAR and PP) and a cluster of stations with high absolute negative slopes,

261 typically < - 4 μ g.m⁻³/decade (SEV, PEN and CPB).

263 5. DISCUSSSION AND CONCLUSIONS

Quantile regression and clustering analysis are applied to study changes in hourly O_3 data over the Iberian Peninsula on the last decade (2000-2009). Ozone data was collected from 11 background monitoring stations, spatially distributed along the IP, characterized by different background values that goes from 30 µg.m⁻³ (suburban stations on the coast of Portugal) to 80 µg.m⁻³ (stations located in centre and east of IP).

269 Quantile regression allows computing trends at different quantiles of the O_3 data distribution 270 within a well-defined statistical framework. In addition, the classical clustering procedure 271 allows summarising the resulting distributions of sample quantile slopes. As in ordinary 272 regression, the slopes for a fixed quantile are not the same over IP, reflecting the spatial 273 dependence of O_3 trends. The results for all monitoring sites show different slopes for the 5%, 274 50% and 95% percentiles, indicating a different rate of temporal change for all parts of the data 275 distribution, as implicitly assumed in ordinary regression. Lower (P5), middle (P50) and upper 276 (P95) quantiles behave differently, with the lower quantiles of O_3 data distribution 277 increasing/decreasing at a much faster rate than the middle and higher quantiles.

278 For example, the CCR station located in the eastern extreme of IP, under influence of different 279 climatic patterns and topographic features, exhibit a very distinct behaviour, with a strong 280 negative trend (< -20 μ g.m⁻³/decade) over all the data distribution, with a higher decrease observed for the lower quantiles (background values) (~ -28 µg.m⁻³/decade). CPB, SEV and 281 282 PEN – located in the interior north part of IP – show a slight negative slope mainly for the lower 283 quantiles (-10 μ g.m⁻³/decade). On the other hand, a positive slope (8-18 ug.m⁻³/decade) can be identified for the stations - NIB, CUS and PP - sited over the Atlantic Ocean coast and also 284 285 TOR and VIZ (4-12 μ g.m⁻³/decade), sited over the Mediterranean influence, and mainly on the 286 lower quantiles of O₃ data distribution (background values). This larger trend in the lower 287 quantiles than in the central and upper part of the data distribution was not found in studies 288 conducted over North America where higher hourly average O₃ concentrations decrease faster 289 than the mid- and lower-values (Lefohn et al., 2008).

290 The analysis of the clusters for different quantiles reflects the differences existent mainly 291 between the lower/middle and the upper quantile. The dendrograms for the lower and median 292 clearly discriminate three groups: stations with larger negative slopes (CCR), slight negative 293 slopes (BAR, SEV, CPB, PEN and ZAR) and the remaining stations with positive slopes (NIB, 294 PP, CUS, TOR, VIZ). The dendrogram for the upper quantile displays a distinct picture: 295 continues to distinguish the CCR station with the highest negative slopes, but the remaining 296 stations are classified in several sub-clusters with minor significance. In fact, the minor gradient 297 of spatial variability occurs at the 95% quantile, with slopes ranging from -8 μ g.m⁻³/decade to 8 298 $\mu g.m^{-3}/$ decade.

In summary, this complementary analysis pointed out that the largest trends are found for the lower O_3 values, with the largest negative trend at the easternmost station of the IP (CCR), and also in northern and mainland stations (BAR, SEV, CPB and PEN), and an opposite behaviour

- 302 is detected at the Atlantic coastal stations (NIB, CUS and PP) with positive O_3 trends.
- 303

304 ACKNOWLEDGEMENTS

The authors gratefully acknowledge the Portuguese 'Ministério da Ciência, da Tecnologia e do Ensino Superior' for the BIOGAIR (PTDC/AAC-AMB/103866/2008) project, for the Ph.D grant of Isabel Ribeiro (SFRH/ BD/60370/2009) and the post-doc grant of Alexandra Monteiro (SFRH/BPD/63796/2009). The acknowledge is extended to the CRUP and the Ministerio de Ciencia e Innovación of Spain by the support of the Integrated Action E 122-10/PT2009-0029. The Spanish Ministry of Science and Innovation is also thanked for the Formación de Personal Investigador (FPI) doctoral fellowship held by María Teresa Pay (CGL2006-08903).

312

313 **REFERENCES**

Adame J.A., Lozano A., Bolívar J.P., De la Morena B.A., Conteras J., Godoy F., 2008.
Behavior, distribution and variability of surface ozone at an arid region in the south of Iberian
Peninsula (Sevilla, Spain). Chemosphere 70, 841-849.

Alonso A.M., Berrendero JR, Hernández A, Justel A., 2006. Time series clustering based on
forecast densities. Comput. Statist. Data Anal., 51 (2), 762-776.

319 Baldasano J.M., Cremades L., Soriano C., 1994. Circulation of air pollutants over the Barcelona

geographical area in summer. Proceedings of Sixth European Symposium Physic-Chemical
 Behaviour of Atmospheric Pollutants. Varese, Italy, 18-22 October. Report EUR 15609/1 EN:

- 322 474-479.
- 323 Barbosa S.M., Scotto M.G., Alonso A.M., 2011. Summarising changes in air temperature over
- 324 central Europe by quantile regression and clustering. Nat. Hazards Earth Syst. Sci. (in press).
- Baur D., Saisana M., Schulze N., 2004. Modelling the effects of meteorological variables on
- 326 ozone concentration a quantile regression approach. Atmos. Environ., 38, 4689-4699.
- Brönnimann S., Buchmann B., Wanner H., 2002. Trends in near-surface ozone concentrations
 in Switzerland: the 1990s. Atmos. Environ., 36, 2841-2852.
- 329 Cade B., Noon B., 2003. A Gentle introduction to quantile regression for ecologists. Frontiers in
- Ecology and the Environment, 1, 412-420.

- 331 Carvalho A.C., Carvalho A., Martins H., Marques C., Rocha A., Borrego C., Viegas D.X.,
- Miranda A.I. (2011). Fire weather risk assessment under climate change using a dynamical
 downscaling approach. Environmental Modelling and Software 26 (9), 1123-1133.
- CEC, 1991, Commission of the European Communities CEC, 1991. Council directive amending
 directive 70/220/EEC on the approximation of the laws of member states relating to the
 measures to be taken against air pollution by emissions from motor vehicles. 91/441/EEC.
 Journal of the European Communities, L242/1-L242/106.
- Derwent R., Jenkin M., Saunders S., Pilling M., Simmonds, P., Passant, N., Dollard, G.,
 Dumitrean P., Kent A., 2003. Photochemical ozone formation in north west Europe and its
 control. Atmos. Environ., 37, 1983–1991.
- 341 Derwent R.G., Simmonds P.G., O'Doherty S., Stevenson D.S., Collins W.J., Sanderson M.G.,
- 342 Johnson C.E., Dentener F., Cofala J., Mechler R., Amann M., 2006. External influences on

343 Europe's air quality: baseline methane, carbon monoxide and ozone from 1990 and 2030 at

- 344 Mace Head, Ireland. Atmos. Environ., 40, 844–855.
- Derwent R.G., Simmonds P.G., Manning A.J., Spain T.G., 2007. Trends over a 20-year period
 from 1987 to 2007 in surface ozone at the atmospheric research station, Mace Head, Ireland.
 Atmos. Environ., 41, 9091–9098.
- EEA, 2010a. The European Environment. State and outlook 2010. Air pollution. European
 Environmental Agency. Luxembourg, Pulication Office of the European Union. ISBN 978-929213-152-4. doi: 10.2800/57792.
- EEA, 2010b. Air pollution by ozone across Europe during summer 2009. Overview of
 exceedances of EC ozone threshold values for April-September 2009. Technical Report 2/2010.
 Euroepan Environmental Agency.
- Elsner J.B., Kossin J.P., Jagger T.H., 2008. The increasing intensity of the strongest tropical
 cyclones. Nature 455, 92-95.
- Gangoiti G., Millán M.M., Salvador R., Mantilla E., 2001. Long-range transport and recirculation of pollutants in the western Mediterranean during the project Regional Cycles of Air
 Pollutions in the West-Central Mediterranean Area. Atmos. Environ., 35, 6267-6276.
- Gimeno L., Hernández E., Rúa A., García R., Martín I., 1999. Surface ozone in Spain.
 Chemosphere, 38, 3061-3074.

- Honrath R.E., Owen R.C., Val Martin M., Reid J.S., Lapina K., Fiahlo P., Dziobak M.P.,
 Kleissel J., Westphal D.L., 2004. Regional and hemispheric impacts of anthropogenic and
 biomass burning emissions on summertime CO and O₃ in the North Atlantic lower free
 troposphere. J. Geophys. Res., 109, D24310.
- Jaffe D., Price H., Parrish D., Goldstein A., Harris J., 2003. Increasing background ozone during
 spring on the west coast of North America. Geophys. Res. Lett., 30(12), 1613,
 doi:10.1029/2003GL017024.
- Jenkin M., 2008. Trends in ozone concentration distributions in the UK since 1990: Local,
 regional and global influences. Atmos. Environ.., 42, 5434–5445.
- Jorba, O., Pérez, C., Rocadenbosch, F., Baldasano, J. M., 2004. Cluster Analysis of 4-Day Back
- 371 Trajectories Arriving in the Barcelona Area, Spain, from 1997 to 2002. Journal of Applied
- 372 Meteorology, 43(6), 887-901.
- Jonson J. E., Simpson D., Fagerli H., Solberg S., 2005. Can we explain the trends in European
 ozone levels? Atmos. Chem. Phys., 6, 51–66.
- 375 Koenker R., 2005. Quantile regression. Cambridge University Press, New York.
- 376 Koenker R., Basset G., 1978. Regression quantiles. Econometrica, 46, 33-50.
- 377 Koenker R., D'Orey V., 1987. Computing regression quantiles. Applied Statistics, 36, 383-393.
- Koenker R., Schorfheide F., 1994. Quantile spline models for global temperature change.
 Climatic Change, 28, 395-404.
- Koenker R., Hallock, K., 2001. Quantile Regression. Journal of Economic Perspectives, 15,
 143–156.
- 382 Lefohn A., Shadwick D., Oltmans S., 2008. Characterizing long-term changes in surface ozone
- 383 levels in the United States (1980–2005). Atmos. Environ., 42, 8252–8262.
- 384 Martín-Vide J., Olcina J., 2001. Climas y tiempos de España. Ed. Alianza, Madrid, 258 pp.
- 385 Marenco A., Gouget H., Nedelec P., Pages J.-P., 1994. Evidence of a long term increase in
- tropospheric ozone from Pic du Midi data series: Consequences: positive radiative forcing. J.
- 387 Geophys. Res., 99, 16617-16632.

- Millán M.M., Salvador R., Mantilla E., 1997. Photooxidant dyanmics in the Mediterranean
 basin in summer: Results from European research projects. J. Geophys. Res., 102(D7), 88118823.
- Millán M., Sanz M.J., Salvador R., Mantilla, E., 2002. Atmospheric dynamics an ozone cycles
 related to nitrogen deposition in the western Mediterranean. Environ. Poll., 118, 167-186.
- 393 Monteiro, A., Vautard, R., Borrego, C., Miranda, A.I., 2005. Long-term simulations of photo
- 394 oxidant pollution over Portugal using the CHIMERE model. Atmos. Environ., 39, 3089-3101.
- Monteiro, A., Miranda, A.I., Borrego, C., Vautard, R., 2007. Air quality assessment for Portugal. Science of the Total Environment 373, 22-31.
- 397 Oltmans S.J., Lefohn A.S., Harris J.M., Galbally I., Scheel H.E., Bodeker G., Brunke E., Claude
- 398 H., Tarasick D., Johnson B.J., Simmonds P., Shadwick D., Anlauf K., Hayden K., Schmidlin F.,
- 399 Fujimoto T., Akagi K., Meyer C., Nichol S., Davies J., Redondas A., Cuevas E., 2006. Long-
- 400 term changes in tropospheric ozone. Atmos. Environ., 40 (17), 3156-3173.
- 401 Ordonez C., Elguindi N., Stein O., Huijnen V., Flemming J., Inness A., Flentje H., Katragkou
- 402 E., Moinat P., Peuch V., Segers A., Thouret V., Athier G., van Weele M., Zerefos C., Cammas
- 403 J., Schultz M., 2010. Global model simulations of air pollution during the 2003 European heat
- 404 wave. Atmos. Chem. Phys., 10, 789–815.
- 405 Pay, M.T., Jiménez-Guerrero, P., Jorba, O., Basart, S., Pandolfi, M., Querol, X., Baldasano,
- 406 J.M., 2011. Spatio-temporal variability of levels and speciation of particulate matter across
- 407 Spain in the CALIOPE modeling system. Submitted in Atmos. Environ.
- 408 Pérez C., Sicard M., Jorba O., Comerón A., Baldasano J.M., 2004. Summertime re-circulations
- 409 of air pollutants over the north-eastern Iberian coast observed from systematic EARLINET lidar
- 410 measurements in Barcelona. Atmos. Environ., 38, 3983-4000.
- 411 Ribas A., Peñuelas J., 2004. Temporal patterns of surface ozone levels in different habitats of
 412 the north western Mediterranean basin. Atmos. Environ., 38, 985-992.
- 413 Scotto M.G., Barbosa S.M., Alonso A.M, 2009. Model-based clustering of Baltic sea-level.
- 414 Appl. Ocean Res., 31, 4-11.
- 415 Seinfeld J.H., Pandis S.N., 1998. Atmospheric Chemistry and Physics: From air pollution to 416 climate change. John Wiley & Sons, New York. ISBN: 9780471178163.

- 417 Sicard P., Coddeville P., Galloo J., 2009. Near-surface ozone levels and trends at rural stations
- 418 in France over the 1995-2003 period. Environ. Monit. Assess., 156, 141-157.
- 419 Soriano C., Baldasano J.M., Buttler W.T., Moore K., 2001. Circulatory patterns of air pollutants
- 420 within the Barcelona air basin in a summertime situation: liar and numerical approaches.
- 421 Bound.-Lay. Meteorol., 98, 33-55.
- Tang G., Li X., Wang Y., Xin J., Ren X., 2006. Surface ozone trend details and interpretations
 in Beijing, 2001–2006. Atmos. Chem. Phys., 9, 8813-8823.
- Toll I., Baldasano J.M., 2000. Modeling of photochemical air pollution in the Barcelona area
 with highly disaggregated anthropogenic and biogenic emissions. Atmos. Environ., 34, 30603084.
- 427 Vautard R., Szopa S., Beekmann M., Menut L., Hauglustaine D. A., Rouil L., Roemer M., 2006.
- 428 Are decadal anthropogenic emission reductions in Europe consistent with surface ozone 429 observations?, Geophys. Res. Lett., 33, L13810, doi:10.1029/2006GL026080.
- 430 Vingarzan R., 2004. A review of surface ozone background levels and trends. Atmos. Environ.,
 431 38, 3431-3442.
- 432

Table captions
Table 1. Selected O ₃ background monitoring stations over the IP.
Figure captions
Figure 1. Map of the IP showing the locations of the background monitoring stations considered
in the present analysis (Table 1). The bullet size indicates the altitude and the round/square
shape the type of the background station (rural/suburban).
Figure 2. Whisker plots of the hourly O ₃ concentrations, measured at the selected sites over the
IP, depicting the median (P50), the P5-P95 range and the non-outliers range.
Figure 3. Time series of hourly O_3 concentrations changes per decade ([µg.m ³]/decade) (solid
grey line) and trends for quantiles 0.05 (dashed line), 0.5 (solid line) and 0.95 (dotted line).
Figure 4. Quantile slopes (O_3 concentration [μ g.m ⁻³]/decade) and corresponding standard errors
for the selected group of stations. The horizontal dashed line represents the usual ordinary least
squares slope.
Figure 5. Dendrogram for 5%, 50% and 95% quantile slopes (right) and the spatial
representation of the quantile slopes (left).











Altitude (m)

Code	Station name	Country	Lat	Lon	Туре	Data collection (%)	Altitude (m)
CUS	Custóias	Portugal	41.21	-8.65	Suburban	89.2%	100
PP	Paio Pires	Portugal	38.63	-9.08	Suburban	89.3%	46
VIZ	Víznar	Spain	37.23	-3.53	Rural	97.2%	1265
NIB	Niembro	Spain	43.44	-4.85	Rural	97.6%	134
CPB	Campisábalos	Spain	41.28	-3.14	Rural	95.5%	1360
CCR	Cabo de Creus	Spain	42.32	3.32	Rural	96.6%	23
BAR	Barcarrota	Spain	38.48	-6.92	Rural	97.0%	393
ZAR	Zarra	Spain	39.08	-1.10	Rural	96.8%	885
PEN	Peñausende	Spain	41.28	-5.87	Rural	91.9%	985
TOR	Els Torms	Spain	41.40	0.72	Rural	89.7%	470
SEV	O Saviñao	Spain	42.64	-7.71	Rural	85.0%	506