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Comparison of Concentration Methods for Rapid Detection of Hookworm Ova in Wastewater Matrices Using Quantitative PCR

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## 27 Abstract

Hookworm infection contributes around 700 million infections worldwide especially in developing 28 29 nations due to poor sanitation. The effective recovery of hookworm ova from wastewater matrices is 30 difficult due to their low concentrations and heterogeneous distribution. In this study, we compared 31 the recovery rates of (i) four rapid hookworm ova concentration methods from municipal wastewater, and (ii) two concentration methods from sludge samples. Ancylostoma caninum ova were used as 32 33 surrogate for human hookworm (Ancylostoma duodenale and Necator americanus). Known 34 concentration of A. caninum hookworm ova were seeded into wastewater (treated and raw) and sludge 35 samples collected from two wastewater treatment plants (WWTPs) in Brisbane and Perth, Australia. The A. caninum ova were concentrated from treated and raw wastewater samples using centrifugation 36 (Method A), hollow fiber ultrafiltration (HFUF) (Method B), filtration (Method C) and flotation 37 (Method D) methods. For sludge samples, flotation (Method E) and direct DNA extraction (Method 38 39 F) methods were used. Among the four methods tested, filtration (Method C) method was able to recover higher concentrations of A. caninum ova consistently from treated wastewater (39-50%) and 40 raw wastewater (7.1-12%) samples collected from both WWTPs. The remaining methods (Methods 41 A, B and D) yielded variable recovery rates ranging from 0.2 to 40% for treated and raw wastewater 42 43 samples. The recovery rates for sludge samples were poor (0.02-4.7), although, Method F (direct DNA extraction) provided 1-2 orders of magnitude higher recovery rate than Method E (flotation). 44 45 Based on our results it can be concluded that the recovery rates of hookworm ova from wastewater 46 matrices, especially sludge samples, can be poor and highly variable. Therefore, choice of 47 concentration method is vital for the sensitive detection of hookworm ova in wastewater matrices. 48 49 50 51 52 **Keywords:** Hookworm ova, Wastewater and sludge, Concentration methods, Recovery rate, Public health 53

## 54 **1. Introduction**

55 It is estimated that almost a quarter of the world's population is infected with soil transmitted 56 helminths (STHs) (WHO, 2015). Among the STHs, hookworm infection contributes around 700 million infections worldwide, especially in tropical and subtropical regions of developing nations due 57 to poor sanitation and hygiene practices (Bethony et al., 2006; Brooker, 2010; Knopp et al., 2012). 58 59 Depending on the prevalence of infections in the community, high concentrations of viable hookworm 60 ova can be present in human wastewater. The presence of ova in wastewater does not pose a direct 61 health risks to humans. However, viable hookworm ova can be hatched into infective larvae (L3) under favorable conditions, and may survive up to 90 days in wastewater matrices (Ben Ayed et al., 62 63 2009). When wastewater is used as irrigation water for crop production, agricultural workers may get infected with hookworm larvae through skin penetration (Gupta et al., 2009; Sidhu and Toze, 2009; 64 65 Navarro and Jimenez, 2011). The infectious dose of hookworm is quite low (1 viable ovum) (WHO, 2006), and therefore, it is vital to detect and quantify these ova in wastewater matrices using 66 traditional or molecular methods. This is crucial for assessing the health risks of exposure to 67 68 ova/larvae contaminated wastewater matrices.

69 The distribution of hookworm ova in wastewater matrices could be patchy. Therefore, detection 70 and quantification of hookworm ova by traditional or molecular methods in wastewater matrices require concentration of the hookworm ova. Ideally, any concentration method should be rapid and 71 72 have the ability to consistently recover high concentrations of ova from wastewater matrices. The concentration method developed by the US EPA has been the most commonly used to recover 73 hookworm ova from wastewater and sludge samples (US EPA, 1999). The recovery rate of this 74 method can be ranged from 65-74% from wastewater samples (Maya et al., 2006). This method, 75 76 however, is laborious and time-consuming due to the requirement of multiple steps of washing and 77 concentrating the samples (Ferguson et al., 2004).

Several methods such as centrifugation (Whitmore and Carrington, 1993; Higgins et al., 2003),
hollow-fiber ultra filtration (HFUF) (Simmons et al., 2001; Ferguson et al., 2004; Hill et al., 2005;
Hill et al., 2007), filtration (Nieminski et al., 1995; Maya et al., 2006; Alli et al., 2011), and flotation

(Bowman et al., 2003; de Victorica and Galván, 2003; Bastos et al., 2013) have also been used to
recover various microorganisms including ova from water and soil samples. Some of these methods
are rapid and can potentially be used to concentrate hookworm ova from wastewater matrices.

84 The aim of this study was to evaluate the performance of various concentration methods to recover

85 hookworm ova from wastewater and sludge samples. For wastewater samples, (A) centrifugation, (B)

86 HFUF, (C) filtration, and (D) flotation, and for sludge samples, (E) flotation, and (F) direct DNA

87 extraction were chosen and compared. A newly developed rapid quantitative PCR (qPCR) assay was

88 developed in this study and used to measure the concentrations of seeded known concentrations of

89 Ancylostoma caninum ova in wastewater and sludge samples in order to identify the best performing

90 method(s).

## 91 **2. Materials and methods**

### 92 2.1. Isolation and enumeration of Ancylostoma caninum ova from dog fecal samples

We used dog hookworm (A. caninum ova) as a surrogate for human hookworm due to the low 93 prevalence of the latter in the Australian population. For the isolation of A. caninum ova, dog fecal 94 95 samples were collected from the School of Veterinary Science, University of Queensland, Gatton, Queensland, Australia. Ova were isolated from ~20 gm of dog fecal samples using the flotation 96 97 method described elsewhere (Bowman et al., 2003). After isolation, ova were preserved in 0.5% formalin and stored at 4°C. The concentrations of ova were estimated by microscopic observation 98 99 using a Sedgewick-Rafter Counting Chamber (Pyser-SGI, UK), and enumerated in each grid at  $40 \times$ magnification in triplicate. 100

## 101 2.2. Determination of ITS-1 rDNA gene copy concentrations in A. caninum ova

102 DNA was extracted from  $400 \pm 40$  (mean  $\pm$  standard deviation) ova in replicates (n = 6) using a MO 103 BIO Power Soil DNA Extraction Kit (Mo Bio, Carlsbad, CA) with minor modifications. All samples 104 were mixed with lysis buffer C1, and freeze-thawed for 10 min (repeated 5 times). In addition, the 105 protocol was amended to allow all the supernatant to be removed at each step, and therefore, 106 increased volumes of solutions C3 and C4 were added to compensate. Extracted DNA was eluted 107 through the spin filter membranes by adding 100 µL of Solution C6, and stored at -80°C until

- 108 processed. The concentrations of ITS-1 rDNA gene copies in A. caninum DNA samples were
- 109 determined using a qPCR assay (see below for methodological details).

## 110 **2.3. Sample preparation**

- 111 Ten liters of raw and treated wastewater samples were collected from two metropolitan wastewater
- treatment plants (WWTPs) in Brisbane, Queensland (WWTP-1) and Perth, Western Australia
- 113 (WWTP-2), Australia. The WWTP-1 is a large biological treatment facility, whereas the WWTP-2 is
- a ponding facility. Treated and raw wastewater samples were transported to the laboratory, and stored
- 115 at 4°C in the dark until processing. The pH of the wastewater samples were determined to be  $7.2 \pm 0.1$
- 116 (treated wastewater; WWTP-1),  $8.9 \pm 0.2$  (raw wastewater; WWTP-1) and  $7.2 \pm 0.1$  (treated
- 117 wastewater; WWTP-2),  $6.7 \pm 0.3$  (raw wastewater; WWTP-2). The turbidity values of the wastewater
- samples were determined to be  $86 \pm 8$  (treated wastewater; WWTP-1),  $197 \pm 17$  NTU (treated
- 119 wastewater; WWTP-1), and  $286 \pm 6$  (raw wastewater; WWTP-2),  $246 \pm 4$  NTU (raw wastewater;
- 120 WWTP-2). The pH and turbidity were measured using 90 FL-T field lab analyser (McVan
- 121 Instruments, Pty Ltd, Melbourne, Australia).

122 Sludge samples were collected from the dewatering belt from WWTP-1 and from the facultative

- pond from WWTP-2 in 500 mL sterile polyethylene zip-locked bags. Samples were then placed on ice
- 124 for transportation to the laboratory and kept at 4°C in dark until processing. The background levels of
- A. *caninum* ITS-1 rDNA gene copies ova in all samples (treated wastewater, raw wastewater and
- sludge) were determined using a qPCR assay (see below). All samples were determined to be free of
- 127 A. caninum ITS-1 rDNA. Approximately,  $400 \pm 40$  A. caninum ova were seeded into 1 L of treated
- 128 wastewater, raw wastewater and sludge (~ 4 gm dry weight) samples. Three repeat trials were
- 129 undertaken, and all samples were tested in triplicates in each trial.

## 130 **2.4.** Ova recovery from wastewater matrices

131 Ova concentration methods flow chart is shown in Fig 1. These methods are referred to as Method A

- 132 [centrifugation (Whitmore and Carrington, 1993)], Method B [HFUF (Hill et al., 2005)], Method C
- 133 [filtration (Hawksworth et al., 2012)], Method D [flotation (Bowman et al., 2003)] for wastewater
- 134 matrices, and Method E [flotation (Bowman et al., 2003)], and Method F [Direct DNA extraction
- 135 (Ahmed et al., 2015)] for sludge samples.

Method A began with the centrifugation of each sample (1 L) in a bucket at 5,200 g for 30 min (Allegra  $\times$ -15R, Beckman Coulter, USA) in two consecutive steps. The pellet was then transferred into a 50 mL polycarbonate tube, further centrifuged at 5,200 g for 10 min, and stored at -20°C until DNA was extracted.

140 Method B involved amending the sample with sodium hexametaphosphate (NaPO<sub>3</sub>) (Sigma Aldrich, Australia) to achieve a final concentration in the water samples of 0.01%. Each water sample 141 was pumped with a peristaltic pump in a closed loop with sterile high-performance, platinum-cured 142 L/S 36 silicone tubing (Masterflex, Cole-Parmer Instrument Co.). Tubing was sterilized by soaking in 143 10% bleach for 30 min, washed with sterile distilled water, and autoclaved at 121°C for 15 min prior 144 to use. A Fresenius Hemoflow F80A polysulfone dialysis filter with a surface area of  $1.8 \text{ m}^2$  and a 145 fiber inner diameter of 200 µm (Fresenius Medical Care, Lexington, MA) was used to process the 146 147 treated and raw wastewater samples. A new filter cartridge was used for each sample. The sample (1 L) was concentrated to approximately 150-200 mL, depending on the turbidity. A 500-mL elution 148 solution consisting of 0.01% Tween 80, 0.01% NaPP, and 0.001% Antifoam A was recirculated 149 through the filter for 5 min, and then allowed to concentrate to 150 mL (Hill et al., 2007). This elution 150 151 solution was added to the concentrated sample to achieve a final volume of approximately 300-350 mL. Secondary concentration of A. caninum ova from the HFUF concentrated samples was performed 152 by centrifugation at 5,200 g for 15 min. After the centrifugation, the supernatant was discarded and 153 154 the pellet was stored at -20°C for DNA extraction.

Method C began with filtering a sample through series of sieves (800-38 µm pore size) (Rowe 155 scientific Pty Ltd, Australia) with the help of a stream of tap water. Particles including ova retained in 156 the smallest pore sized sieve (38 µm) were collected in a 50 mL polycarbonate tube and centrifuged at 157 5,200 g for 15 min to obtain a pellet. The pellet was then stored at -20°C until DNA was extracted. 158 Method D began with centrifuging treated and raw wastewater samples (1 L) to achieve a pellet. 159 The pellet was then transferred into a 50 mL polycarbonate tube and approximately 40-45 mL 160 flotation solution (MgSO<sub>4</sub>) was added. The pellet was mixed with the flotation solution by vortexing. 161 The mixture was centrifuge for 3 min at 800 g and the materials present in the top 10 mL were 162

transferred into a 15 mL polycarbonate tube. Water was added to make up the volume to 15 mL andfurther centrifuged at 800 g for 10 min to obtain a pellet.

## 165 **2.5. Ova recovery from sludge**

- 166 Ova from sludge samples were concentrated using Methods E and F. Method E began with
- 167 centrifugation of ova spiked sludge (~ 4 gm dry weight) samples at 800 g for 10 min. The supernatant
- 168 was discarded, and 40-45 mL flotation solution was added in each samples. The mixture was then
- 169 centrifuged for 3 min at 800 g and floated materials were transferred into 15 mL polycarbonate tube.
- 170 Water was added to make up the volume to 15 mL and further centrifuged at 800 g for 10 min to
- 171 obtain a pellet. For Method F, direct DNA extraction was performed from ova spiked sludge samples
- 172 (~ 4 gm dry weight) using a MO Bio Power Max® Soil DNA Extraction Kit as described below.

## 173 **2.6. DNA extraction**

- 174 DNA was extracted from each pellet obtained through all Methods (A, B, C and D) using the MO Bio
- 175 Power Max<sup>®</sup> Soil DNA Extraction Kit with minor modification. In brief, pellets were mixed with
- 176 lysis buffer C1 and freeze-thawed for 10 min (repeated 5 times). Extracted DNA samples were eluted
- through the spin filter membranes by adding 2 mL solution C6 and stored at -80°C until processed.
- 178 DNA was extracted from each pellet using a MO Bio Power Max® Soil DNA Extraction Kit with
- 179 minor modification.

## 180 2.7. PCR inhibition

Previously published assay (Sketa22) was used to determine the presence of PCR inhibitors in the
extracted DNA samples from treated wastewater, raw wastewater and sludge samples (Ahmed et al.,
2015).

#### 184 **2.8. Preparation of standard curves**

185 DNA was extracted from the larvae using DNeasy Blood and Tissue<sup>®</sup> Kit (Qiagen, Valencia, CA).

- 186 qPCR standards were prepared by cloning the purified amplicons into the pGEM-T Easy Vector
- 187 System (Promega, Madison, WI, USA). Plasmid DNA was extracted using Plasmid Mini Kit
- 188 (Qiagen). Standards were prepared from the plasmid DNA (Yun et al. 2006; Ahmed et al. 2014).
- 189 Serial dilutions were prepared ranging from  $10^5$ - $10^0$  gene copies per  $\mu$ L and used as standard curves.
- 190

- 191 2.9. qPCR assay for the quantification of ITS-1 rDNA
- 192 For qPCR assay, newly designed primers (F: 5'-TTT GCT AAC GTG CAC TGA ATG-3' and R: 5'-
- 193 GAA ACA CCG TTG TCA TAC TAG CC-3'), and a probe (P: FAM-5'-AAC TCG TTG TTG CTG
- 194 CTG AA-3'-TAMRA) targeting the 5.8S ITS-1 rDNA genes were used. The qPCR amplification was
- 195 performed in 25  $\mu$ L reaction mixtures containing 12.5  $\mu$ L iQ<sup>TM</sup> Supermix (Bio-Rad Laboratories, CA,
- 196 USA), 250 nM of each primer, 400 nM of probe, 3  $\mu$ L of template DNA and UltraPure<sup>TM</sup>
- 197 DNase/RNase-free distilled water (Life Technologies, Australia). The thermal cycler program
- 198 consisted of 15 min at 95°C, 15s at 95°C and 1 min at 59°C. The qPCR assays were performed using
- the Bio-Rad CFX96 thermal cycler (Bio-Rad Laboratories, USA). All qPCR reactions were
- 200 performed in triplicate. The qPCR assay performance criteria such as efficiency (E), slope, intercept,
- 201  $R^2$  and lower limit of quantification (LLOQ) were determined by analyzing the standard curves over
- the course of the study.

## 203 2.10. qPCR lower limit of quantification

- 204 The qPCR lower limit of quantification (LLOQ) was determined from the Ct values obtained for
- standards range ( $3 \times 10^5 3$  gene copies). The lowest amount of diluted standards detected in 100%
- 206 triplicates assays was considered as qPCR LLOQ.

# 207 2.11. Recovery rate determination

- 208 The recovery rate of hookworm ova in the wastewater and sludge samples by the different
- 209 concentration methods was calculated as follows:
- 210 Recovery rate (%) = (Quantified gene copies/spiked gene copies)  $\times 100$ .

## 211 **2.12. Quality control**

- 212 To minimize qPCR contamination, DNA extraction and qPCR set up were performed in separate
- 213 laboratories. A method blank was included for each batch of treated wastewater, raw wastewater, and
- sludge samples. A reagent blank was also included during DNA extraction to account for any
- 215 contamination during extraction. For each qPCR experiment, standards (also served as a positive
- control) and triplicate negative controls (UltraPure<sup>TM</sup> water) were included.
- 217
- 218

#### 219 **2.13. Statistical analysis**

- 220 GraphPad Prism 6 (GraphPad Software, CA, USA) was used to conduct the statistical analysis. A
- one-way ANOVA was performed to determine the differences between the  $C_T$  values obtained for O.
- keta DNA suspended in UltraPure<sup>TM</sup> water and *O. keta* seeded DNA samples extracted from
- 223 wastewater matrices. ANOVA was also used to assess whether the concentration of A. caninum gene
- 224 copies obtained through Methods (A-D) for treated and raw wastewater samples were statistically
- 225 different within and between WWTPs. A paired T- test was used to assess the significant difference
- between Methods (E and F) for sludge samples within and between WWTPs. Statistical significance
- 227 was determined at  $\alpha = 0.05$ .

## 228 **3. Results**

## 229 **3.1. qPCR standards and lower limit of quantification (LLOQ)**

qPCR standards were analysed to determine the reaction efficiencies. The standards had a linear range 230 of quantification from  $10^5 - 10^1$  gene copies per  $\mu$ L of plasmid DNA. The slope of the standards 231 ranged from -3.31 to -3.38. The amplification efficiencies ranged from 100.7% to 108.2%, and the 232 correlation coefficient ( $R^2$ ) ranged from 0.96-0.98. The intercepts for the qPCR standards were 35.8 to 233 38.4 (Fig 2). LLOQ of qPCR assays were determined using the standards. The qPCR LLOQ was 30 234 235 gene copies for all triplicate samples. The intra-assay and inter-assay Coefficient of Variation (CV) of the standards were also determined. These values were less than 1% and 3% respectively, indicating 236 237 high reproducibility of the qPCR assay.

#### 238 **3.2. PCR inhibition**

Sketa22 assay was used to determine the presence of PCR inhibitors in the extracted DNA samples. The mean  $C_T$  value and standard deviation for the *Oncorhynchus keta* seeded UltraPure<sup>TM</sup> water was 28.5 ± 0.2. The  $C_T$  values for *O. keta* seeded treated and raw wastewater DNA samples from WWTP-1 processed through all methods (A-D) were similar to *O. keta* seeded UltraPure<sup>TM</sup> water, indicating the DNA samples were free of PCR inhibitors (Table 1). However, PCR inhibition was observed in DNA samples extracted from treated wastewater (WWTP-2) processed through Methods A and B. Raw wastewater DNA samples from WWTP-2 processed through Methods A and C also had PCR

246 inhibitors. Sludge DNA samples (WWTP-1) processed through Method E had no PCR inhibitors. In contrast, sludge DNA samples from WWTP-2 processed through Method E had PCR inhibitors. None 247 of the sludge DNA samples (both WWTPs) processed using Method F showed PCR amplification. 248 Samples that showed the sign of PCR inhibitors were then serially diluted (10-fold) to relieve PCR 249 250 inhibitors, and re-analysed by seeding O. keta DNA. The mean  $C_T$  values and standard deviations of O. keta for the 10-fold diluted treated wastewater, raw wastewater and sludge samples indicated the 251 removal of PCR inhibition (Table 1). Further ANOVA analysis on the  $C_T$  values for O. keta seeded 252 UltraPure<sup>TM</sup> water, undiluted DNA and those 10-fold diluted DNA samples did not differ 253 significantly. Based on the results, all the samples that passed PCR inhibition test were used for qPCR 254 255 assays. 3.3. Recovery rate of A. caninum from wastewater matrices 256

To obtain the recovery rates for each method,  $400 \pm 40$  ova (corresponds to  $3.3 \times 10^7 \pm 8.5 \times 10^6$  gene 257 copies as determined by the qPCR) were seeded into each wastewater and sludge samples. The mean 258 concentration of A. caninum gene copies recovered from treated wastewater did not vary significantly 259 (P > 0.05) among the methods. The concentrations ranged from  $4.6 \times 10^5$  (Method A) to  $1.3 \times 10^6$ 260 (Method D) for wastewater sample collected from WWTP-1 (Fig. 3a). Similar results were also 261 obtained for WWTP-2. However, the mean concentration of gene copies  $(3.5 \times 10^3)$  recovered through 262 Method D was 2-3 orders of magnitude lower than the other Methods (A-C). Furthermore, this 263 difference was significant (P < 0.05). 264

For raw wastewater samples, the mean concentration of A. caninum gene copies recovered using 265 Method C was the highest  $(3.8 \times 10^5)$  followed by Method D  $(2.3 \times 10^5)$  for WWTP-1 (Fig. 3b). 266 However, Methods A and B yielded 2 orders of magnitude lower concentrations of gene copies 267 compared to Methods C and D, and this difference was significant (P < 0.05). For WWTP-2, Method 268 B yielded the highest concentration  $(1.1 \times 10^6)$  of gene copies followed by Methods D and C, although 269 Methods B, C and D did not differ significantly (P > 0.05). However, the mean concentration of gene 270 copies  $(1.5 \times 10^4)$  recovered through Method A was 1-2 orders of magnitude lower than the other 271 methods (P < 0.05). 272

273	For the sludge samples collected from WWTP-1, Methods E $(7.8 \times 10^2)$ and F $(2.7 \times 10^3)$ yielded
274	similar concentrations of gene copies (Fig. 3c), that were not significantly ( $P > 0.05$ ) different. Sludge
275	samples collected from WWTP-2 also yielded similar concentrations of gene copies for Method E
276	$(1.2 \times 10^5)$ and F $(1.5 \times 10^5)$ , and the difference was not statistically significant ( $P > 0.05$ ). Both
277	Methods were able to recover ~ 2 orders of magnitude higher gene copies from WWTP-2 samples
278	compared to WWTP-1 samples ( $P < 0.05$ ).
279	For treated wastewater, Method D outperformed all other methods except Method C, yielding a
280	recovery rate of $40 \pm 57\%$ for WWTP-1 (Table 2). Interestingly, for WWTP-2, Method C performed
281	better than the others, yielding a recovery rate of 50 $\pm$ 39%. For raw wastewater, Method C (12 $\pm$
282	10%) and D (7.1 $\pm$ 2.0%) had much better recovery rate than Methods A (0.3 $\pm$ 0.2%) and B (0.3 $\pm$
283	0.4%) for WWTP-1. For WWTP-2, the recovery rate of Method B outperformed all other methods.
284	For sludge samples, the recovery rates of hookworm ova were poor compared to treated and raw
285	wastewater samples. For both WWTPs Method F yielded 1-2 orders of magnitude higher $(3.7 \pm 9.0\%)$
286	WWTP-1; 4.7 $\pm$ 6.2%, WWTP-2) recovery rate than Method E (0.02 $\pm$ 0.03%, WWTP-1; 0.10

287 ±0.15%, WWTP-2).

# 288 4. Discussion

A reliable, sensitive and rapid method is needed in order to detect low concentrations (1-10 ova) of 289 290 helminth ova in the wastewater matrices. Various methods have been used to recover hookworm ova 291 from wastewater matrices with variable degrees of success (Bowman et al., 2003; McCuin and Clancy, 2005; Maya et al., 2006; Ensink et al., 2008;). In light of this, we have evaluated several rapid 292 concentration methods for the recovery of hookworm ova from wastewater matrices including sludge 293 samples. For the methods evaluation, wastewater and sludge samples were collected from two 294 295 WWTPs with variable characteristics. Method A (centrifugation) used in this study was originally 296 developed to separate helminth ova from environmental water samples with low turbidity (Whitmore and Carrington, 1993). The results obtained in this study suggest that the recovery rate of the Method 297 298 A was 1-2 orders of magnitude higher for treated wastewater than raw wastewater samples. Raw 299 wastewater samples generally contain large amount of heavy particles and grease that may potentially

300 bind to ova (Kuczynska and Shelton, 1999). As a result it is possible that DNA extraction lysis buffer 301 may not have penetrated the cell wall, which may have led to inefficient DNA extraction. Method B (HFUF) has been widely used to concentrate bacterial, viral and protozoa pathogens 302 simultaneously from environmental water samples (Hill et al., 2005; Hill et al., 2007). The recovery 303 304 rates of the HFUF from treated wastewater samples were slightly better than the centrifugation (Method A). However, the recovery rates from raw wastewater were highly variable (0.3-35%) 305 between the WWTPs. Such discrepancy again could be attributed to the variable solid contents 306 present in wastewater samples in time and space. The turbidity of raw wastewater collected from both 307 the WWTPs were much higher (246-286 NTU) than the treated wastewater (86-197 NTU). Several 308 studies have demonstrated the efficacy of the HFUF system to recover higher concentrations (up to 309 86%) of Giardia cysts and Cryptosporidium oocysts from surface water samples (Simmons et al., 310 311 2001; Ferguson et al., 2004). Perhaps, HFUF method is suitable for concentrating protozoa when the turbidity of the water samples is low. Mull and Hill (2012) and Ferguson and collegues (2004) 312 demonstrated that the turbidity of water samples is inversely proportional with the recovery rates. 313 314 Method C (filtration) used in this study is based on retaining hookworm ova on a filter through a series of sieves. This method is simple, involves only few steps, and because of that, has the potential 315 316 to recover higher concentrations of ova from wastewater samples. Our results indicated that the 317 recovery rate of Method C was as high as 50% for treated wastewater and 12% for raw wastewater 318 samples. This is comparable to a 26% recovery rate of Ascaris from treated wastewater reported by 319 Maya et al. (2006), and 9-49% recovery rate of Cryptosporidium oocysts and Giardia cysts from 320 environmental waters reported by Nieminski et al. (1995) using a similar methodology. One drawback of this method is the potential clogging of the sieve with large solid wastewater particles. This may 321 322 leave behind a portion of ova attached to the solid particles on the sieve (Nieminski et al., 1995; Zarlenga and Trout, 2004). 323 The flotation method (Method D) separates helminth ova by selecting their specific gravity while 324

325 other denser particles present in a sample sink to the bottom for removal (Dryden et al., 2005;

Goodman et al., 2007). Thus, this method is more suitable to recover helminth ova from highly turbid

327 samples like raw wastewater and sludge. Studies have shown that the flotation method can provide

variable recovery rates (12%-32%) from wastewater samples (Maya et al., 2006). This is in agreement
with the findings of this study. The recovery rate obtained through Method D for the treated
wastewater collected from WWTP-1 was high, although the result was not consistent for both
WWTPs. Treated wastewater samples from the WWTP-2 contained large amount of blue green algae,
which may have affected the recovery rate. However, more studies would be required to determining
the effect of blue green algae on ova recovery rate possibly from large number of samples from
different ponding facilities.

The flotation method (Method E) has also been used to recover hookworm ova from sludge 335 samples. The result of this study indicated that the recovery rates of this method were very poor (0.02-336 3.7%). McCuin and Clancy (2005) could not recover any Cryptosporidium oocysts from lime 337 stabilized sludge samples using flotation method. In contrast, several studies reported 26-82% 338 339 recovery rate of helminth ova from different sludge samples using flotation method (Bowman et al., 2003; Maya et al., 2006). Several factors such as sample matrix, sample volume and the 340 341 concentrations of ova present in samples may influence the recovery rate, therefore, making direct 342 comparison among the studies is difficult.

343 It has been reported that direct DNA extraction from water samples may yield better recovery of viruses as it bypasses the concentration procedure (Ahmed et al., 2015). In view of this, we used 344 Method F, which involved direct DNA extraction from sludge samples. Method F was indeed able to 345 346 recover higher numbers of ova from sludge samples than Method E. However, the DNA samples obtained through this method had PCR inhibitors present, despite the DNA extraction kit used in this 347 study being equipped with inhibitor removal technology. PCR inhibitors are known to be matrix 348 associated, and a wide array of PCR inhibitors with varying concentration could be present in sludge 349 350 samples (Schrader et al., 2012). Our results also indicated that the 35% of DNA samples extracted from wastewater matrices had PCR inhibitors. This is a formidable barrier for downstream PCR 351 detection or quantification of hookworm ova. Based on our data, we recommend that DNA samples 352 extracted from wastewater matrices should be checked for the presence of PCR inhibitors prior to 353 PCR/qPCR analysis. In the present study, we simply assumed that the DNA extraction efficiency of 354 355 the MO Bio Power Max® Soil DNA Extraction Kit was 100% in order to calculate the concentrations

of seeded ova in wastewater matrices. Further work would be required to determine the extractionefficiency of the DNA extraction kit.

## 358 **5. Conclusions**

- From the results obtained in this study, it appears that the recovery rates of *A. caninum* ova
  from wastewater matrices can be highly variable and matrix-specific.
- The results indicated that centrifugation (Method A), HFUF (Method B), filtration (Method
   C), and flotation (Method D) were able to yield better recovery rates from treated wastewater
   samples than raw wastewater. The recovery rates obtained through flotation (Method E) and
   direct DNA extraction (Method F) from sludge samples were low compared to treated and
   raw wastewater samples.
- Among the four concentration methods tested, filtration (Method C) was able to recover
   higher concentrations of *A. caninum* ova consistently from treated wastewater and raw
   wastewater samples collected from both WWTPs. The performances of Methods B (HFUF)
   and D (flotation) were reasonable, although, the results were not consistent for both WWTPs.
- Both methods (Methods E and F) failed to recover *A. caninum* ova efficiently from sludge
   samples. Further method development would be required in order to improve the recovery
   rate of hookworm ova from sludge samples.
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# 384 **Reference**

- Ahmed, W., Gyawali, P., Sidhu, J., Toze, S., 2014. Relative inactivation of faecal indicator bacteria
   and sewage markers in freshwater and seawater microcosms. Lett Appl Microbiol. 59 (3),
   348-354.
- Ahmed, W., Harwood, V. J., Gyawali, P., Sidhu, J. P., Toze, S., 2015. Concentration methods
   comparison for quantitative detection of sewage-associated viral markers in environmental
   waters. Appl Environ Microbiol. 81 (6), 2042-2049.
- Alli, J., Abolade, G., Kolade, A., Salako, A., Mgbakor, C., Ogundele, M., Oyewo, A., Agboola, M.,
   2011. Prevalence of intestinal parasites on fruits available in Ibadan Markets, Oyo State,
   Nigeria. Acta Parasitol Globalis. 2 (1), 6-10.
- Bastos, V. K., Cutolo, S. A., Doria, M. C. O., Razzolini, M. T. P., 2013. Detection and quantification
  of viable *Ascaris* sp. and other helminth eggs in sewage sludge. Int J Environ Health Res 23
  (4), 352-362.
- Ben Ayed, L., Schijven, J., Alouini, Z., Jemli, M., Sabbahi, S., 2009. Presence of parasitic protozoa
  and helminth in sewage and efficiency of sewage treatment in Tunisia. Parasitol Res. 105 (2),
  393-406.
- Bethony, J., Brooker, S., Albonico, M., Geiger, S. M., Loukas, A., Diemert, D., Hotez, P. J., 2006.
  Soil-transmitted helminth infections: ascariasis, trichuriasis, and hookworm. Lancet. 367
  (9521), 1521-1532.
- Bowman, D. D., Little, M. D., Reimers, R. S., 2003. Precision and accuracy of an assay for detecting
   *Ascaris* eggs in various biosolid matrices. Water Res. 37 (9), 2063-2072.
- Brooker, S., 2010. Estimating the global distribution and disease burden of intestinal nematode
  infections: adding up the numbers-a review. Int J Parasitol. 40 (10), 1137-44.
- 407 De Victorica, J., Galvan, M., 2003. Preliminary testing of a rapid coupled methodology for
   408 quantitation/viability determination of helminth eggs in raw and treated wastewater. Water
   409 Res. 37 (6), 1278-1287.
- 410 Do, T. T., Molbak, K., Phung, D. C., Dalsgaard, A., 2007. Helminth infections among people using
  411 wastewater and human excreta in peri-urban agriculture and aquaculture in Hanoi, Vietnam.
  412 Trop Med Int Health. 12 Sup (2), 82-90.
- 413 Dryden, M. W., Payne, P. A., Ridley, R., Smith, V., 2005. Comparison of common fecal flotation
   414 techniques for the recovery of parasite eggs and oocysts. Vet Ther. 6 (1), 15-28.
- Ensink, J. H., Blumenthal, U. J., Brooker, S., 2008. Wastewater quality and the risk of intestinal
  nematode infection in sewage farming families in hyderabad, India. Am J Trop Med Hyg. 79
  (4), 561-567.
- Ferguson, C., Kaucner, C., Krogh, M., Deere, D., Warnecke, M., 2004. Comparison of methods for
  the concentration of *Cryptosporidium* oocysts and *Giardia* cysts from raw waters. Can J
  Microbiol. 50 (9), 675-682.
- Goodman, D., Haji, H. J., Bickle, Q. D., Stoltzfus, R. J., Tielsch, J. M., Ramsan, M., Savioli, L.,
  Albonico, M., 2007. A comparison of methods for detecting the eggs of *Ascaris*, *Trichuris*,
  and hookworm in infant stool, and the epidemiology of infection in Zanzibari infants. Am J
  Trop Med Hyg. 76 (4), 725-731.
- Gupta, N., Khan, D. K., Santra, S. C., 2009. Prevalence of intestinal helminth eggs on vegetables
  grown in wastewater-irrigated areas of Titagarh, West Bengal, India. Food Cont. 20 (10),
  942-945.
- Haugland, R. A., Siefring, S. C., Wymer, L. J., Brenner, K. P., Dufour, A. P., 2005. Comparison of
   *Enterococcus* measurements in freshwater at two recreational beaches by quantitative
   polymerase chain reaction and membrane filter culture analysis. Water Res. 39 (4), 559-568.
- Hawksworth, D., Archer, C., Rodda, N., Smith, M., Appleton, C., Buckley, C., 2012. An improved method for the recovery of *Ascaris* ova from solid waste from urine diversion toilets.
  <u>http://www.ewisa.co.za/literature/files/155\_107%20Hawksworth.pdf</u>. Accessed 6 March 2015.

435	Higgins, J. A., Trout, J. M., Fayer, R., Shelton, D., Jenkins, M. C., 2003. Recovery and detection of
430	<i>Cryptosportatum parvum</i> oocysis from water samples using continuous flow centrifugation.
437	Water Kes. 37 (15), 3551-3560.
438	Hill, V. R., Kahler, A. M., Jothikumar, N., Johnson, T. B., Hahn, D., Cromeans, T. L., 2007.
439	Multistate evaluation of an ultrafiltration-based procedure for simultaneous recovery of
440	enteric microbes in 100-liter tap water samples. Appl Environ Microbiol. 73 (13), 4218-4225.
441	Hill, V. R., Polaczyk, A. L., Hahn, D., Narayanan, J., Cromeans, T. L., Roberts, J. M., Amburgey, J.
442	E., 2005. Development of a rapid method for simultaneous recovery of diverse microbes in
443	drinking water by ultrafiltration with sodium polyphosphate and surfactants. Appl Environ
444	Microbiol. 71 (11), 6878-6884.
445	Hotez, P. J., Bethony, J., Bottazzi, M. E., Brooker, S., Buss, P., 2005. Hookworm: "The Great
446	Infection of Mankind". PLoS Med. 2 (3), e67.
447	Knopp, S., Steinmann, P., Keiser, J., Utzinger, J., 2012 Nematode infections: soil-transmitted
448	helminths and <i>Trichinella</i> . Infect Dis Clin North Am. 26 (2), 341-358.
449	Kuczynska, E., Shelton, D. R., 1999. Method for detection and enumeration of Cryptosporidium
450	parvum oocysts in feces, manures, and soils. Appl Environ Microbiol. 65 (7), 2820-2826.
451	Maya, C., Jimenez, B., Schwartzbord, J., 2006. Comparison of techniques for the detection of
452	helminth ova in drinking water and wastewater. Water Environ Res. 78 (2), 118-124.
453	Mccuin, R. M., Clancy, J. L., 2005. Methods for the recovery, isolation and detection of
454	<i>Cryptosporidium</i> oocysts in wastewaters. J Microbiol Methods, 63 (1), 73-88.
455	Mull, B., Hill, V. R., 2012, Recovery of diverse microbes in high turbidity surface water samples
456	using dead-end ultrafiltration. I Microbiol Methods 91 (3) 429-433
457	Navarro I Jimenez B 2011 Evaluation of the WHO helminth eggs criteria using a OMRA
/58	approach for the safe reuse of wastewater and sludge in developing countries. Water Sci
450 159	Technol 63 (7) 1499-505
455	Nelson K I Darby I I 2001 Inactivation of viable Ascaris eggs by reagents during enumeration
400	Appl Environ Microbiol 67 (12) 5453 0
401	Nieminski E C Schoofer E Orgerth I E 1005 Comparison of two methods for detection of
402	Cigridia syste and Cryptognaridium acousts in yeater Appl Environ Microbiol. 61 (5), 1714
405	1710
404	1/17. Schreder C. Schielles A. Ellerhreut I. Johns D. 2012 DCD inhibitors, accurrence, properties and
405	schladel, C., Schleike, A., Eheroroik, L., Johne, K., 2012. FCK inhomore-occurrence, properties and
466	removal. J Appi Microbiol. 113 (5), 1014-1020.
467	Sidnu, J. P. S., Toze, S., 2009. Human pathogens and their indicators in biosonds: A intreature review.
468	Environ International. 35 (1), 187-201.
469	Simmons, O. D., Sobsey, M. D., Heaney, C. D., Schaefer, F. W., Francy, D. S., 2001. Concentration
470	and detection of <i>Cryptosporidium</i> oocysts in surface water samples by method 1622 using
4/1	ultrafiltration and capsule filtration. Appl Environ Microbiol. 67 (3), 1123-1127.
4/2	Traub, R. J., Hobbs, R. P., Adams, P. J., Behnke, J. M., Harris, P. D., Thompson, R. C., 2007. A case
4/3	of mistaken identity-reappraisal of the species of canid and felid hookworms (Ancylostoma)
474	present in Australia and India. Parasitol. 134 Pt (1), 113-119.
475	United State Environment protection Agency, 1999. Control of pathogens and vector attraction in
476	sewage sludge. USEPA Environmental Regulations and Technology. Office of Research and
477	Development, EPA/623/R-92/013, Washington DC.
478	United State Environment protection Agency, 2003. Control of pathogens and vector attraction in
479	sewage sludge. USEPA Environmental Regulations and Technology. Office of Research and
480	Development, EPA/625/R-92/013. Washington DC.
481	Vuong, T. A., Nguyne, T. T., Klank, L. T., Phung, D. C., Dalsgaard, A., 2007. Faecal and protozoan
482	parasite contamination of water spinach (Ipomoea aquatica) cultivated in urban wastewater in
483	Phnom Penh, Cambodia. Trop Med Int Health. 12 Sup (2), 73-81.
484	Whitmore, T. N., Carrington, E. G., 1993. Comparison of methods for recovery of Cryptosporidium
485	from water. Water Sci Technol. 27 (3-4), 69-76.
486	World Health Organisation, 2006. Guidelines For the safe use of Wastewater, Excreta and Greywater
487	in agriculture, Vol. 2 and Vol. 4. WHO Library Cataloguing-in-Publication Data, p.213.

488 Geneva, Switzerland.

489 490 491	World Health Organisation, 2015. Investing to overcome the global impact of neglected tropical diseases: third WHO report on neglected diseases. WHO Press, World Health Organization, 20 Avenue Appia, 1211 Geneva, Switzerland.
492 493 494 495 496	<ul> <li>Yun, J. J., Heisler, L. E., Hwang, I. I., Wilkins, O., Lau, S. K., Hyrcza, M., Jayabalasingham, B., Jin, J., Mclaurin, J., Tsao, M. S., Der, S. D., 2006. Genomic DNA functions as a universal external standard in quantitative real-time PCR. Nucleic Acids Res. 34 (12), e85.</li> <li>Zarlenga, D. S., Trout, J. M., 2004. Concentrating, purifying and detecting waterborne parasites. Vet Parasitol. 126 (1-2), 195-217.</li> </ul>
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Table 1: Sketa22 real-time PCR assay for the evaluation of PCR inhibition in ova spiked raw wastewater, treated wastewater, and sludge DNA samples as opposed to UltraPure<sup>TM</sup> water samples. UltraPure<sup>TM</sup> water samples, undiluted and diluted DNA samples were spiked with 10 pg of *Oncorhynchus keta* DNA 508

Concentrations	Sample types	Mean $\pm$ standard deviation of threshold cycle (C <sub>T</sub> ) values for Sketa22 PCR assay			
methods					
		Undiluted D	NA samples	10-fold diluted	d DNA samples
		WWTP-1	WWTP-2	WWTP-1	WWTP-2
Method A	Treated wastewater	$27.8\pm0.2$	$31.2\pm1.9$	NA	$29.0 \pm 1.7$ <sup>b</sup>
	Raw wastewater	$28.0\pm0.1$	$31.4 \pm 1.2$	NA	$27.0 \pm 0.2^{b}$
Method B	Treated wastewater	$27.7\pm0.1$	$30.5 \pm 0.2$	NA	$29.2 \pm 1.8^{b}$
	Raw wastewater	$28.0\pm0.1$	30.0 ± 0.1	NA	NA
Method C	Treated wastewater	$28.1\pm0.1$	$29.9 \pm 0.1$	NA	NA
	Raw wastewater	$28.3\pm0.4$	$33.0 \pm 1.5$	NA	$27.0 \pm 0.1$ <sup>b</sup>
Method D	Treated wastewater	$28.1\pm0.2$	$29.8\pm0.1$	NA	NA
	Raw wastewater	$28.2\pm0.1$	$29.8\pm0.1$	NA	NA
Method E	Sludge	$28.2\pm0.2$	$31.3 \pm 2.0$	NA	$27.9 \pm 1.7^{\ b}$
Method F	Sludge	No amplification	No amplification	$29.2 \pm 0.1^{a}$ ,	$27.1 \pm 0.1^{b}$

509 Mean  $\pm$  standard deviation of  $C_T$  values for UltraPure<sup>TM</sup> water samples =  $28.5 \pm 0.2$ 

510 NA: Not applicable.

Table 2: Evaluation of recovery rate of *A. caninum* ova from raw wastewater, treated wastewater, and
 sludge samples from six concentration methods (A-F)

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Concentration	Sample types	Mean and standard deviation of recovery rate (%)	
methods		WWTP-1	WWTP-2
Method A	Treated wastewater	$14 \pm 35$	$7.6 \pm 14$
	Raw wastewater	$0.3 \pm 0.2$	$0.5 \pm 1.4$
Method B	Treated wastewater	$18 \pm 26$	$17 \pm 20$
	Raw wastewater	$0.3 \pm 0.4$	$35 \pm 30$
Method C	Treated wastewater	$39 \pm 26$	$50 \pm 39$
	Raw wastewater	$12 \pm 10$	$7.1 \pm 13$
Method D	Treated wastewater	$40 \pm 57$	$0.2 \pm 0.1$
	Raw wastewater	$7.1 \pm 2.0$	$7.4 \pm 31$
Method E	Sludge	$0.02\pm0.03$	$3.7 \pm 9.0$
Method F	Sludge	$0.10\pm0.15$	$4.7 \pm 6.2$

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- Fig 1: Hookworm ova concentration methods for raw wastewater, secondary treated wastewater and sludge 668 samples. Method A = centrifugation, Method B = HUFU, Method C = Filtration, Method D = Floatation (for 669
- 670 wastewater samples), Method E = Floatation (for sludge samples) and Method F = Direct DNA extraction.
- Fig 2: A standard curves generated using the plasmid DNA. The concentrations of gene copies are plotted 671 against  $C_T$  values. The  $C_T$  is the cycle number at which the fluorescence signal increased above the defined 672 threshold value, calculated by the real-time PCR software. 673
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- Fig 3: Mean and standard deviation of the concentrations of gene copies recovered through different 675
- methods tested from A. caninum ova seeded into (a) treated wastewater, (b) raw wastewater, and (c) sludge 676 samples
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# Highlights:

- The distribution of hookworm ova in wastewater matrices could be patchy.
- A rapid concentration method is required for the detection of ova from wastewater matrices.
- Six rapid methods were compared to identify the best performing method to recover ova from wastewater matrices.
- Recovery rates of *A. caninum* ova from wastewater matrices especially sludge samples can be highly variable.
- Further method development would be required in order to improve the recovery rate of hookworm ova from sludge samples.