

Accepted Manuscript

Title: Fractionation of microbial populations in a PHA accumulating mixed culture and associated PHA content and composition

Author: Om Murugan Janarthanan Kenn Lu Bronwyn Laycock Alan Werker Steven Pratt



PII: S0141-8130(14)00282-7
DOI: <http://dx.doi.org/doi:10.1016/j.ijbiomac.2014.04.055>
Reference: BIOMAC 4322

To appear in: *International Journal of Biological Macromolecules*

Received date: 22-1-2014
Revised date: 23-4-2014
Accepted date: 28-4-2014

Please cite this article as: O.M. Janarthanan, K. Lu, B. Laycock, A. Werker, S. Pratt, Fractionation of microbial populations in a PHA accumulating mixed culture and associated PHA content and composition, *International Journal of Biological Macromolecules* (2014), <http://dx.doi.org/10.1016/j.ijbiomac.2014.04.055>

This is a PDF file of an unedited manuscript that has been accepted for publication. As a service to our customers we are providing this early version of the manuscript. The manuscript will undergo copyediting, typesetting, and review of the resulting proof before it is published in its final form. Please note that during the production process errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain.

1 **Fractionation of microbial populations in a PHA accumulating mixed culture and**
2 **associated PHA content and composition**

3

4 Om Murugan Janarthanan ^a, Kenn Lu ^b, Bronwyn Laycock ^a, Alan Werker ^c, Steven Pratt ^a

5 ^a School of Chemical Engineering, The University of Queensland, Brisbane, QLD 4072,
6 Australia

7 ^b Advanced Water Management Centre, The University of Queensland, Brisbane, QLD 4072,
8 Australia

9 ^c AnoxKaldnes AB, Klosterängsvägen 11A, SE-226 47 Lund, Sweden

10

11 *Corresponding author:*

12 Dr Steven Pratt (s.pratt@uq.edu.au); phone +61 7 33654943; School of Chemical
13 Engineering, The University of Queensland, Brisbane, QLD 4072, Australia

14

15 **Abstract**

16 The uniformity of PHA composition and content across groups of organisms in mixed
17 cultures was considered. An activated sludge microbial community, with an average PHA
18 content of 20 wt%, was fractionated by Percoll assisted buoyant density separation. The
19 microbial community in the two principal fractions was characterised using amplicon
20 pyrosequencing. While organisms were common to both fractions, the relative abundances of
21 species were found to be different between the two fractions. The average PHA content in
22 one of the fractions was found to be higher (24 wt%) than the other (16 wt%); separation was
23 considered to be in part driven by the density difference associated with PHA content, but
24 also by other factors such as cell dimension and cellular morphology. But while differences
25 in PHA content were observed, the PHA composition in both fractions was found to be
26 approximately the same (43 – 44 mol% HV), which shows that distinct groups of microbial
27 populations within mixed cultures may generate PHA with similar average copolymer
28 composition.

29

30

30

31 **Keywords**

32 Polyhydroxyalkanoates; PHA; Mixed cultures; Buoyant density gradient centrifugation;
33 pyrosequencing.

34

35 **1. Introduction**

36 Polyhydroxyalkanoates (PHAs) have drawn increasing attention because of their
37 biodegradability, biocompatibility and potential for production from renewable resources [1].
38 More than 300 species of Gram-positive and Gram-negative bacteria as well as a wide range
39 of archaea synthesize PHA as intracellular insoluble cytoplasmic inclusion bodies in the
40 presence of excess carbon source and limited amounts of nutrients such as phosphorus or
41 nitrogen. The production and material properties of PHA have been widely reviewed,
42 e.g. Laycock et al. [2]. The use of mixed cultures, potentially derived from activated sludge
43 used in the treatment of wastewater, might offer a cost effective approach to PHA synthesis
44 as it avoids requirement for sterilisation and allows for the use of fermented residuals,
45 industrial process waters and wastes as feedstocks [3, 4].

46

47 Many biologically produced PHAs, from both pure and mixed cultures, have a broad
48 compositional distribution and are in fact blends of copolymers of differing hydroxyvalerate
49 (HV) and hydroxybutyrate (HB) content rather than a single copolymer of narrow
50 composition [5, 6]. It is known that PHBV copolymers in blends will at least partially phase
51 separate if the difference in HV content between components is more than 12% [7].
52 Therefore the macro-scale architecture, and consequently the mechanical properties, of such
53 blends are anticipated to be influenced by the polymer make-up.

54

55 Fractionation studies show that commercial PHA materials from pure cultures can be a
56 mixture of random copolymers [8, 9]; it is expected that this co-polymer blend nature of
57 products from mixed cultures, enriched on mixed feeds, will be similarly evident. In the
58 presence of multiple substrates, microbial populations specialised in those different substrates
59 are selected, coexisting in the reactor by adapting to different survival niches [10]. The

60 structure, physio-chemical properties, monomer composition and the number and size of the

61 PHA granules apparently vary depending on the organism and the carbon source utilized by
62 the microorganisms [11]. So it is possible that populations with distinct substrate affinities
63 will produce different co-polymers and consequently blends of PHA copolymers will result
64 during the polymer recovery from a mixed-culture biomass. However, the reason for broad
65 copolymer compositional distributions in both pure and mixed cultures is likely to be more
66 complex than this, being potentially based on intracellular differences with respect to
67 substrate utilisation within the same bacterial family. Overall, the origin of complex
68 monomer distribution in bacterial copolyesters has not yet become clear in either pure or
69 mixed cultures, but from these fractionation studies it can be understood that it is of interest
70 to gain more fundamental understanding of the inter- and intra-cellular nature of the co-
71 polymer distribution and thereby identify the principal source of variability in PHA co-
72 polymer distribution that have been identified in PHA recovered from mixed culture biomass.

73

74 One method to follow specific groups of populations in mixed cultures is by isolating each
75 bacterial species (based on tagging or by flow fractionation etc.), followed by pure culture
76 cultivation. However, this approach may not be effectively achieved for the microbial
77 community in activated sludge obtained from wastewater treatment plants (WWTPs) because
78 only small quantities of individual species would be obtainable and many of the
79 microorganisms are anticipated to be non-culturable under laboratory conditions [12-14]. So,
80 methods that rely on selective separation of the biomass are required to discriminate distinct
81 population fractions. Buoyant density-gradient centrifugation is a useful technique for
82 separating cells or organelles of different densities. It has even been trialled as a method to
83 concentrate PHA rich biomass in mixed cultures: Oshiki et al. [15, 16] concentrated the
84 number of PHA rich cells in mixed cultures through density gradient centrifugation. In one
85 study, the authors considered two mixed cultures, each with about 20% of cells being rich in
86 PHA [16]. After density gradient centrifugation (using a Percoll solution of 1.087 g mL^{-1})
87 they obtained pellets with 50-60% of cells being rich in PHA. The approach was based on

88 the separation principle that PHA granules have a higher density of 1.15 to 1.25 g mL⁻¹ [17]
89 than activated sludge at 1.02 to 1.06 g mL⁻¹ [18].

90

91 In this study we aimed to separate a PHA accumulating mixed cultures by buoyant density
92 gradient centrifugation and thereby produce distinct microbial fractions. The work builds on
93 that presented by Oshiki et al. by examining the microbial community as well as the PHA
94 content and composition within the respective fractions. These were analysed with a view to
95 gaining a deepened understanding of the homogeneity or heterogeneity of distribution for
96 PHA within a PHA-rich mixed culture biomass.

97 2. Materials and Methods

98

99 2.1 Materials

100 Acetic acid and propionic acid were obtained from Sigma Aldrich and were of 98% purity.
101 Chloroform was of HPLC grade (99.9% purity) and was obtained from Sigma Aldrich.
102 Percoll consists of colloidal silica particles of 15-30 nm diameters (23 wt% in water) that
103 have been coated with polyvinylpyrrolidone (PVP). Percoll was obtained from Sigma Aldrich
104 with a density of $1.130 \pm 0.005 \text{ g mL}^{-1}$. All other chemicals were obtained from Sigma
105 Aldrich and were of at least 98% purity.

106

107 2.2 PHA accumulation and chemical analysis

108 Waste Activated Sludge (WAS) was obtained from Luggage Point Wastewater treatment
109 plant (WWTP), Queensland, Australia, and was kept at $4 \text{ }^{\circ}\text{C}$ before being warmed up to room
110 temperature before use ($21\text{--}25 \text{ }^{\circ}\text{C}$) and aerated until reaching stable dissolved oxygen
111 concentrations close to saturation. PHA was accumulated in this WAS without enrichment of
112 the population [3], using a pulse feeding sequence with the feed consisting of a solution
113 containing acetic acid (35.71 g L^{-1}) and propionic acid (6.67 g L^{-1}), with a chemical oxygen
114 demand (COD) ratio of 80:20 acetic acid: propionic acid. Accumulation was undertaken
115 under aerobic conditions at room temperature for 8 h in a laboratory-scale batch reactor with
116 a working volume of 2 L. Air was supplied by an air pump at the rate of $2 \text{ vvm (L L}^{-1} \text{ min}^{-1})$.
117 pH was monitored but not controlled. Experiments were managed using the Opto 22 PLC
118 controlled program based on dissolved oxygen (DO), and feeding was based on the use of a
119 DO control point set at 1.8 mg L^{-1} to initiate the addition of the carbon source following
120 depletion of the previous dose of feed. DO returning to the set point (1.8 mg L^{-1}) indicated
121 that readily available substrate had been taken up by the biomass and so the next dose of
122 substrate ($5 \text{ mL at } 50 \text{ g L}^{-1}$) was pumped in automatically. This pulse-wise feeding strategy
123 was used because it has been observed that PHA accumulation is enhanced when the carbon
124 source is added in a number of controlled pulses rather than in a single, larger pulse [19-21].

125

126 During the course of the accumulation, grab samples were taken at selected time points
127 between 0 to 8 h for Gas Chromatography (GC), Volatile Fatty Acids (VFA), Total
128 Suspended Solids (TSS) and Volatile Suspended Solids (VSS) analysis. For the

129 measurements of $\text{NH}_4\text{-N}$ and phosphate phosphorus ($\text{PO}_4\text{-P}$) concentrations, a flow injection
130 analyser (FIA) was (Lachat QuickChem8000). Total Suspended Solids (TSS) and Volatile
131 Suspended Solids (VSS) were measured gravimetrically by drying a filtered and washed,
132 known volume in a $100\text{ }^\circ\text{C}$ oven overnight (for TSS), and for 2 h in a $550\text{ }^\circ\text{C}$ furnace (for
133 VSS).

134

135 PHA was measured by GC analysis: 10 mL samples were prepared according to the hot
136 chloroform method [22], with some modifications [23]. 10 mL samples were taken, with 10
137 drops of 40% formaldehyde added to stop all activity. The samples were centrifuged at 3750
138 rpm ($3267 \times g$) for 5 mins. The supernatant was removed and the pellet dried in an oven at
139 $110\text{ }^\circ\text{C}$ for 24 h (overnight). Next, 2 mL of acidified methanol with benzoic acid as internal
140 standard (3 wt% of H_2SO_4) and 2 mL of chloroform were added followed by digestion of the
141 sample for 20 h at $100\text{ }^\circ\text{C}$. After cooling, 1 mL of milli-Q water was added to promote
142 methanolysis. After 1 h settling, the organic phase was transferred to a vial and 3 \AA molecular
143 sieves added (to remove water) in preparation for GC analysis. 1 mL of the chloroform phase
144 (bottom phase) was then removed for analysis on a Perkin-Elmer gas chromatograph (GC). 3
145 μL of this chloroform phase was then analysed using a flame ionisation detector (FID) at 300
146 $^\circ\text{C}$ and a DB-5 nonpolar capillary column ($30\text{ m} \times 0.25\text{ mm} \times 0.25\text{ }\mu\text{m}$) at a range of $80\text{-}270$
147 $^\circ\text{C}$. Biologically sourced PHBV copolymer (70 mol% HB) (Sigma Chemicals, USA) was
148 used as a standard.

149

150 VFAs were determined by means of a Perkin-Elmer GC. The column was a DB-FFAP 15 m
151 $\times 0.53\text{ mm} \times 1.0\text{ }\mu\text{m}$ (length ID film) at $140\text{ }^\circ\text{C}$, while the injector and flame ionisation
152 detector (FID) were operated at $220\text{ }^\circ\text{C}$ and $250\text{ }^\circ\text{C}$, respectively. High purity helium was used
153 as a carrier gas at a flow rate of 17 mL min^{-1} .

154

155 **2.3 Buoyant density separation of PHA accumulating cells**

156 Mixed microbial cells containing PHA were harvested by centrifugation at 3750 rpm ($3267 \times$
157 g) for 10 min (wet biomass) using a Beckman SX4750 rotor in an Allegra X-12 Series
158 centrifuge. To the wet biomass, 25 mL distilled water was added to make a cell suspension
159 for separation studies. Percoll was used for density gradient centrifugation because of its
160 nontoxicity, low viscosity, ease of preparation at the desired osmolarity and pH [24]. The cell

161 suspension (9.6 mL) was layered on top of a 40% Percoll gradient working solution that
162 contained 10.8 mL of Percoll, 1.2 mL of 2 M NaCl, and 14.4 mL of distilled water. The
163 mixture was centrifuged for 10 min at 3750 rpm (3267 x g). In the same fashion, different
164 selected concentrations of Percoll ranging from 10% to 45% were prepared.

165

166 **2.4 Quantitative determination of PHA in the separated cells**

167 PHA content and composition of the cells separated by buoyant density gradient were
168 analysed through the following procedure. The samples were centrifuged, again at 3750 rpm
169 (3267 x g) for 10 min, and the pellet and the float were pipetted out in to fresh new tube. Both
170 were then washed with milliQ water, re-centrifuged, and the supernatant decanted. The
171 separated cells (pellet and float) were washed repeatedly to effectively rinse off all the
172 Percoll. All samples were then dried at 100 °C overnight. Approximately 20 mg of separated
173 cells were added to 2 mL of chloroform and 2 mL of an acidified methanol solution
174 containing 100 mg L⁻¹ of sodium benzoate as an internal standard were added. The samples
175 were heated to 100 °C for 20 h and the GC analysis performed as above.

176

177 **2.5 Microscopic analysis of the separated cells**

178 Samples were divided into two fractions (pellet and float) after buoyant density gradient
179 centrifugation, with the middle section not being analysed due to it being too disperse. To
180 view PHA inclusion bodies, samples were heat fixed, stained with 1% (wt vol⁻¹) Nile blue A
181 (NBA) (Sigma, NSW, Australia) for 15 min at 55 °C, destained for 30 s in 8% (vol%) acetic
182 acid, water washed, air dried, and viewed through an Olympus BX51 fluorescent microscope
183 equipped with a CCD DP70 (Olympus, Japan) camera. To determine the overall abundance
184 of PHA-accumulating microorganisms in the cells, dual chemical staining with NBA and
185 DAPI was performed. Because PHA granules emit strong fluorescence when they are bound
186 with NBA, PHA-accumulating cells are easily identified by fluorescence microscopy staining
187 with NBA [25].

188

189 **2.6 Microbial characterization of the separated cells**

190 The microbial characterization of the separated cells was accomplished through
191 pyrosequencing. DNA was extracted from the control and the two fractions after buoyant
192 density separation using the Q-Biogene Fast DNA SPIN kit for soil (MP Biomedicals, Seven
193 Hills, NSW, Australia) according to the manufacturer's instructions. DNA integrity was

7

194 checked by agarose gel electrophoresis in 1% agarose gels and quantified using a
195 Nanodrop™ ND1000 spectrophotometer (Nanodrop Technologies Inc., Wilmington, DE,
196 USA). DNA extracted from separated cells was amplified through PCR. PCR amplification
197 of the 16S rDNA was performed using the eubacterial primers 926f and 1392r. The PCR
198 mixture consisted of 1 µL of DNA template, 1 µL of each primer 926f (5'-
199 AAACYAAAKGAATTGACGG-3') and 1392r (5'-ACGGGCGGTGT-GTAC-3') [26], 5 µL
200 of 10X PCR buffer (Applied Biosystems, VIC, Australia), 1 µL each of a 200 µM solution of
201 dATP, dCTP, dGTP, and dTTP, and 34.5 µL of sterile water. Each reaction mixture was
202 overlaid with filter-sterilized mineral oil. The reaction mixtures were heated to 100 °C,
203 followed by the addition of 0.5 µL of AmpliTaq Gold Taq polymerase (5 U µL⁻¹; Applied
204 Biosystems, VIC, Australia.). The PCR was conducted on a MyCycler Thermal Cycler
205 System (Bio-Rad Laboratories Inc., Hercules, CA, USA), with a programme comprising an
206 initial 5-min denaturation step at 95 °C, followed by 25 cycles of 94 °C for 1 min, 55 °C for 1
207 min and 72 °C for 2 mins, and a final 10-min extension step at 72 °C [27]. Equal volumes of
208 triplicate PCR products were pooled and unincorporated primers and reaction components
209 were removed using a Qiaquick PCR purification kit (Qiagen, Australia) according to the
210 manufacturer's instructions and the purified products were resuspended in DNase-free water.
211 For each set of PCR amplifications, a control reaction without template was performed to
212 check the kit and solution purity. Amplicons were pooled in equimolar concentrations and
213 sequenced using a 454 GS FLX Titanium sequencer (Roche) as per the manufacturer's
214 protocol. Pyrosequencing results were analysed through a local implementation of the ACE
215 Pyrosequencing Pipeline (<https://github.com/Ecogenomics/APP>) in which sequence reads
216 were split according to the barcode in QIIME [28]. De-multiplexed sequences were then
217 trimmed to 250 bp length and de-noised by ACACIA [29], a tool for homopolymer error-
218 correction that has greater scalability than existing tools and is used to maintain sensitivity
219 without compromising genuine signal in the data [30]. Sequences with 97% similarity were
220 assigned to operational taxonomic units (OTUs) by CD-HIT-OTU [31] and aligned by *Pynast*
221 [28]. Each sequence was then assigned to the taxonomy with BlastTaxonAssigner in QIIME
222 through greengenes database (2013 Aug release). Finally the non-normalized OTUs table and
223 rarefaction curve were generated by QIIME. Normaliser
224 (<https://github.com/minillim/Normaliser>) was used to find a centroid normalised OTUs
225 table with 2300 reads per sample. The sequences recovered from the dominant 10 OTUs were
226 also compared with other sequences previously deposited in GenBank

227 (<http://www.ncbi.nlm.nih.gov>) using the Basic Local Alignment Search Tool (BLAST) and
228 genus level classification were assigned (if >99% identity were obtained). The normalised
229 OTUs table was then summarised to phylum level and used to predict the variance of
230 microbial community. Further, the normalised OTUs table was imported to R (version 2.8.1).
231 After a log transformation, a heat map was drawn up including the top 30 abundant OTUs.
232 OTUs were clustered based on Euclidean distance in the heat map.

233

234 **3. Results and Discussion**

235

236 **3.1 PHA accumulation**

237 PHA was accumulated in non-enriched waste activated sludge. The substrate for
238 accumulation was a mix of acetic and propionic acids (80:20 on a gCOD basis). PHA
239 accumulation was run for 8 hours. After this time, consumption of further carbon and hence
240 PHA accumulation effectively ceased and a PHA content of 20 wt% (gPHA gTSS⁻¹), and
241 composition of 57 mol% HB – 43 mol% HV was obtained.

242

243 **3.2 Biomass fractionation: PHA content and composition**

244 Percoll solutions with densities of 1.013, 1.025, 1.037, 1.049 and 1.055 g mL⁻¹ based on
245 10%, 20%, 30%, 40% and 45% dilute Percoll solutions respectively were trialled in this
246 study, with the 1.049 g mL⁻¹ (40%) solution being selected for full fractionation as it
247 delivered the most distinct fractions.

248

249 The material balance showing the fractionation in terms of PHA and solids in the 40%
250 Percoll separated samples is shown in Table 1. After buoyant density separation a fraction of
251 the biomass settled at the bottom of the centrifuge tube as a pellet (50%), a fraction floated
252 (30%) and a residual remained in suspension (20%). Two distinct fractions were considered
253 for analysis: the pellet and the float. Table 1 shows that the PHA content of the pellet was
254 higher than of the float.

255

256 [TABLE 1 near here]

257

258 The PHA content observed in the pellet and float was 24 wt% and 16 wt%, respectively.
259 Considering that PHA is denser than activated sludge, separation was likely in part driven by

260 the density difference associated with PHA content. Density gradients have been applied to
261 collect PHA rich cells from cultures containing cells with varying PHA content [16].
262 However, while Oshiki et al. [16] managed to concentrate the number of PHA rich cells using
263 density gradient centrifugation, the authors found that PHA recovery in the pellet was
264 actually reduced as the density of the solution was increased. The authors recovered less than
265 20% of PHA in the pellet when the optimal density of 1.087 g mL^{-1} was used. They
266 attributed this to the aggregation of PHA rich and sparse cells; an additional factor may have
267 been inconsistencies between the concentration of PHA rich cells and mass fractions. The
268 results of the present investigation and associated unpublished activities also suggest that the
269 density gradient separation methods are not universally applicable to industrial processes for
270 recovering PHA rich biomass. In this work, separation was not based on PHA alone,
271 evidenced by the fact that in low concentration Percoll solutions (10%, 20% and 30%) most
272 of the sludge material settled, indicating, based on assumed densities of PHA and activated
273 sludge, that most of the biomass had a PHA content of greater than approximately 15%. At
274 40% Percoll solution, effective separation was achieved and higher PHA content in the pellet
275 and lower in the float was observed. But in high Percoll solution (45%) all the sludge
276 material floated, erroneously indicating, based on assumed densities of PHA and activated
277 sludge, that none of the biomass had greater than approximately 20% PHA content; GC
278 analysis showed that it did. By contrast, in an associated unpublished experiment using
279 biomass with a PHA content of 56 wt%, separation occurred at a Percoll concentration of
280 only 25%, and the float layers (74% of separated biomass) had elevated PHA content (63.30
281 $\pm 0.01 \text{ wt\%}$, $n=8$) compared to the pellet (6% of biomass, which had a PHA content of 56
282 wt%, $n=1$). These results highlight that factors additional to PHA content may contribute to
283 fractionation in a density gradient separation. In mixed culture biotechnology, density
284 separation has been applied to separate microbial populations, for example, for differentiation
285 of methanotrophic bacteria [32], separation of bacteria from freshwater sediments [33],
286 separation of subpopulations with different gene expression patterns [34], separation of cells
287 of different viability or activity [35], separation of respiring bacterial cells from mixed
288 populations in natural seawater [36], and separation of polyphosphate-accumulating cells
289 [37]. However, in the present investigation we find that PHA content is not a sufficiently
290 dominating attribute of the biomass to control fractionation in density gradient separation
291 when using a relatively low speed centrifuge, although it would be worth exploring if
292 separation would be more effective if ultracentrifugation is used.

293

294 The monomer composition in pellet and float were effectively the same, being measured as
295 P(3HB-co-44% 3HV) in the pellet and P(3HB-co-43% 3HV) in float. It has been
296 demonstrated through fractionation studies that these copolymers are more than likely
297 comprised of blends of P(3HB-co-3HV) of differing HV contents [38]. Originally this result
298 was attributed to the difference in composition of PHA accumulated in different bacterial
299 species present in the community. However, the fact that we have obtained very similar bulk
300 compositions in two fractions that we later show (Section 3.3) to have differing community
301 composition may point towards a more complex explanation, such as contribution of
302 intracellular variation to distribution of co-polymer blends.

303

304 3.3 Microbial communities

305 Pyrosequencing recovered total 8380 raw reads, grouped into 362 OTUs from two fractions
306 (pellet and float) and the original biomass. The errors in pyrosequencing reads have been
307 estimated as approximately 100–250 base pairs and 5–10 errors kb⁻¹ [39], which is not
308 considered to be significant with respect to classification of phyla [40]. Figure 1 shows that
309 the phyla of the pellet were similar to that of the float, but the relative abundances were
310 different. *Proteobacteria* dominated both the pellet (52.3% of the total sequences) and the
311 float (41.6%). In the pellet the remainder of the community consisted of *Bacteroidetes*
312 (17.8%), *Chloroflexi* (9.8%), *Actinobacteria* (4.0%), *Firmicutes* (3.3%), *Planctomycetes*
313 (2.4%), and *Acidobacteria* (1.1%), while in the float it consisted of *Chloroflexi* (18.8%),
314 *Bacteroidetes* (12.4%), *Actinobacteria* (9.5%), *Firmicutes* (6.0%), *Planctomycetes* (2.4%),
315 and *Acidobacteria* (1.4%).

316

317 [FIGURE 1 near here]

318

319 Figure 2 shows that at the OTUs level the differences are even more pronounced, with the
320 community of the pellet being more closely aligned with the original than it was to the float.
321 The heat map shows *Candidatus Competibacter phosphatis* was the most abundant OTU
322 present, representing 28.3% of all classified sequences in the pellet and 17.7% in the float.
323 *Candidatus Competibacter phosphatis* is a non-culturable glycogen accumulating organism
324 of the class *Gammaproebacteria*, with all phylotypes having the ability to anaerobically
325 store volatile fatty acids (VFAs) as PHA without observed polyphosphate cycling. These

11

326 organisms are commonly found during post-FISH chemical staining for PHA in full-scale
327 EBPR wastewater treatment plants [41]. The distribution of other OTUs, including
328 *Actinomycetales* and *Dokdonella*, varied substantially in relative magnitude. These
329 organisms were 50 to 100% more prevalent in the float (4.1%, 4.5%) than the pellet (2.4%,
330 2.7%).

331

332 [FIGURE 2 near here]

333

334 **3.4 Microscopy**

335 Microscopic investigation with staining using Nile blue A and DAPI was undertaken to
336 confirm the presence of PHA in the fractions. The cells in the pellet contained PHA granules
337 whereas the float possessed some cells apparently not containing any granules (data not
338 shown).

339

340

341

342

343 **4 Conclusions**

344 In this work a mixed culture population with PHA content of 20 wt% and copolymer
345 composition of 43 mol% HV was separated into two fractions (pellet and float) containing
346 distinctions in relative abundances of microbial populations. The pellet and float possessed
347 different PHA content, but density separation was apparently not due solely to PHA content
348 showing that buoyant density separation is not necessarily effective for concentrating PHA
349 rich cells as has been previously been suggested in the literature. Observed differences in net
350 PHA content could be a result of the measured differences in the microbial communities or
351 simply differences in stages of the growth and accumulation cycles within individual
352 communities. The fact that PHA composition in both fractions was found to be similar (43 –
353 44 mol% HV) indicates the latter might be most likely, but shows that groups of microbial
354 populations within mixed cultures do not necessarily generate PHA with unique composition,
355 although at this low level of PHA content the degree of separation may not be enough to
356 discriminate significant differences.

357

358 **5 Acknowledgements**

359 The authors would like to acknowledge the Australian Research Council for funding this
360 work through project LP0990917.

361

362 References

- 363 [1] S. Bengtsson, A. Werker, M. Christensson, and T. Welander, "Production of
364 polyhydroxyalkanoates by activated sludge treating a paper mill wastewater," *Bioresource*
365 *Technology*, vol. 99, pp. 509-516, Feb 2008.
- 366 [2] B. Laycock, P. Halley, S. Pratt, A. Werker, and P. Lant, "The chemomechanical properties of
367 microbial polyhydroxyalkanoates," *Prog. Polym. Sci.*, vol. 38, pp. 536-583, 2013.
- 368 [3] M. V. Arcos-Hernandez, S. Pratt, B. G. Laycock, P. Johansson, A. Werker, and P. A. Lant,
369 "Waste activated sludge as biomass for production of commercial-grade
370 polyhydroxyalkanoate (PHA)," *Waste and Biomass Valorization*, 2012.
- 371 [4] J. Nikodinovic-Runic, M. Guzik, S. T. Kenny, R. Babu, A. Werker, and O. C. K. E, "Carbon-rich
372 wastes as feedstocks for biodegradable polymer (polyhydroxyalkanoate) production using
373 bacteria," *Adv Appl Microbiol*, vol. 84, pp. 139-200, 2013.
- 374 [5] Y. Arai, A. Cao, N. Yoshie, and Y. Inoue, "Studies on comonomer compositional distribution
375 and its effect on some physical properties of bacterial poly(3-hydroxybutyric acid-co-3-
376 hydroxypropionic acid)," *Polymer International*, vol. 48, pp. 1219-1228, Dec 1999.
- 377 [6] N. Yoshie and Y. Inoue, "Chemical composition distribution of bacterial copolyesters,"
378 *International Journal of Biological Macromolecules*, vol. 25, pp. 193-200, Jun-Jul 1999.
- 379 [7] N. Yoshie and Y. Inoue, "Cocrystallization and phase segregation in blends of two bacterial
380 polyesters," *Macromolecular Symposia*, vol. 224, pp. 59-70, Apr 2005.
- 381 [8] N. Yoshie, H. Menju, H. Sato, and Y. Inoue, "Complex composition distribution of poly(3-
382 hydroxybutyrate-co-3-hydroxyvalerate)," *Macromolecules*, vol. 28, pp. 6516-6521, Sep 11
383 1995.
- 384 [9] B. Laycock, M. V. Arcos-Hernandez, A. Langford, S. Pratt, A. Werker, P. J. Halley, *et al.*,
385 "Crystallisation and fractionation of selected polyhydroxyalkanoates produced from mixed
386 cultures," *New Biotechnology*, vol. in press, pp. 1-12, 2013.
- 387 [10] M. G. E. Albuquerque, G. Carvalho, C. Kragelund, A. F. Silva, C. M. T. Barreto, M. A. M. Reis,
388 *et al.*, "Link between microbial composition and carbon substrate-uptake preferences in a
389 PHA-storing community," *ISME J.*, vol. 7, pp. 1-12, 2013.
- 390 [11] S. P. Valappil, S. K. Misra, A. R. Boccaccini, I. Keshavarz, C. Bucke, and I. Roy, "Large-scale
391 production and efficient recovery of PHB with desirable material properties, from the newly
392 characterised *Bacillus cereus* SPV," *Journal of Biotechnology*, vol. 132, pp. 251-258, Nov 1
393 2007.
- 394 [12] J. T. Staley and A. Konopka, "Measurement of insitu activities of nonphotosynthetic
395 microorganisms in aquatic and terrestrial habitats," *Annual Review of Microbiology*, vol. 39,
396 pp. 321-346, 1985.
- 397 [13] R. Amann, "In-situ identification of microorganisms by whole cell hybridization with rRNA-
398 targeted nucleic acid probes," *Molecular Microbial Ecology Manual*, vol. 3, pp. 331-345,
399 1995.
- 400 [14] P. Hugenholtz, B. M. Goebel, and N. R. Pace, "Impact of culture-independent studies on the
401 emerging phylogenetic view of bacterial diversity," *Journal of Bacteriology*, vol. 180, pp.
402 4765-4774, Sep 1998.
- 403 [15] M. Oshiki, M. Onuki, H. Satoh, and T. Mino, "PHA-accumulating microorganisms in full-scale
404 wastewater treatment plants," *Water Science and Technology*, vol. 58, pp. 13-20, 2008.

- 405 [16] M. Oshiki, M. Onuki, H. Satoh, and T. Mino, "Separation of PHA-accumulating cells in
406 activated sludge based on differences in buoyant density," *Journal of General and Applied*
407 *Microbiology*, vol. 56, pp. 163-167, Apr 2010.
- 408 [17] J. Mas, C. Pedrosalio, and R. Guerrero, "Mathematical-model for determining the effects of
409 intracytoplasmic inclusions on volume and density of microorganisms," *Journal of*
410 *Bacteriology*, vol. 164, pp. 749-756, 1985.
- 411 [18] E. E. Dammel and E. D. Schroeder, "Density of activated-sludge solids," *Water Research*, vol.
412 25, pp. 841-846, Jul 1991.
- 413 [19] M. M. Cai, H. Chua, Q. L. Zhao, S. N. Shirley, and J. Ren, "Optimal production of
414 polyhydroxyalkanoates (PHA) in activated sludge fed by volatile fatty acids (VFAs) generated
415 from alkaline excess sludge fermentation," *Bioresource Technology*, vol. 100, pp. 1399-1405,
416 Feb 2009.
- 417 [20] N. Gurieff, "Production of Biodegradable Polyhydroxyalkanoate Polymers Using Advanced
418 Biological Wastewater Treatment Process Technology " PhD Thesis, School of
419 Engineering, The University of Queensland, Brisbane, 2007.
- 420 [21] L. S. Serafim, P. C. Lemos, R. Oliveira, and M. A. M. Reis, "Optimization of
421 polyhydroxybutyrate production by mixed cultures submitted to aerobic dynamic feeding
422 conditions," *Biotechnology and Bioengineering*, vol. 87, pp. 145-160, Jul 20 2004.
- 423 [22] G. Braunegg, B. Sonnleitner, and R. M. Lafferty, "Rapid Gas-Chromatographic Method for
424 Determination of Poly-Beta-Hydroxybutyric Acid in Microbial Biomass," *European Journal of*
425 *Applied Microbiology and Biotechnology*, vol. 6, pp. 29-37, 1978.
- 426 [23] A. Oehmen, B. Keller-Lehmann, R. J. Zeng, Z. G. Yuan, and E. Keller, "Optimisation of poly-
427 beta-hydroxyalkanoate analysis using gas chromatography for enhanced biological
428 phosphorus removal systems," *Journal of Chromatography A*, vol. 1070, pp. 131-136, Apr 8
429 2005.
- 430 [24] H. Pertoft, "Fractionation of cells and subcellular particles with Percoll," *Journal of*
431 *Biochemical and Biophysical Methods*, vol. 44, pp. 1-30, Jul 10 2000.
- 432 [25] M. Oshiki, H. Satoh, and T. Mino, "Acetate uptake by PHA-accumulating and non-PHA-
433 accumulating organisms in activated sludge from an aerobic sequencing batch reactor fed
434 with acetate," *Water Science and Technology*, vol. 62, pp. 8-14, 2010.
- 435 [26] P. H. Dang, P. D. Jensen, and D. J. Batstone, "Methanosarcinaceae and acetate oxidising
436 pathways dominate in high-rate thermophilic anaerobic digestion of waste activated
437 sludge.," *Applied and Environmental Microbiology*, 2013.
- 438 [27] R. J. Newton, S. E. Jones, M. R. Helmus, and K. D. McMahon, "Phylogenetic ecology of the
439 freshwater Actinobacteria acl lineage," *Applied and Environmental Microbiology*, vol. 73, pp.
440 7169-7176, Nov 2007.
- 441 [28] J. G. Caporaso, J. Kuczynski, J. Stombaugh, K. Bittinger, F. D. Bushman, E. K. Costello, *et al.*,
442 "QIIME allows analysis of high-throughput community sequencing data," *Nature Methods*,
443 vol. 7, pp. 335-336, May 2010.
- 444 [29] L. Bragg, G. Stone, M. Imelfort, P. Hugenholtz, and G. W. Tyson, "Fast, accurate error-
445 correction of amplicon pyrosequences using Acacia," *Nature Methods*, vol. 9, pp. 425-426,
446 May 2012.
- 447 [30] L. Bragg, G. Stone, M. Imelfort, P. Hugenholtz, and G. W. Tyson, "Fast, accurate error-
448 correction of amplicon pyrosequences using Acacia," *Nat. Methods*, vol. 9, pp. 425-426,
449 2012.
- 450 [31] T. H. Wu, E. Ayres, R. D. Bardgett, D. H. Wall, and J. R. Garey, "Molecular study of worldwide
451 distribution and diversity of soil animals," *Proceedings of the National Academy of Sciences*
452 *of the United States of America*, vol. 108, pp. 17720-17725, Oct 25 2011.

- 453 [32] K. P. Putzer, L. A. Buchholz, M. E. Lidstrom, and C. C. Remsen, "Separation of
454 Methanotrophic Bacteria by Using Percoll and Its Application to Isolation of Mixed and Pure
455 Cultures," *Applied and Environmental Microbiology*, vol. 57, pp. 3656-3659, Dec 1991.
- 456 [33] A. L. D. Furtado and P. Casper, "Different methods for extracting bacteria from freshwater
457 sediment and a simple method to measure bacterial production in sediment samples,"
458 *Journal of Microbiological Methods*, vol. 41, pp. 249-257, Aug 2000.
- 459 [34] H. Makinoshima, A. Nishimura, and A. Ishihama, "Fractionation of *Escherichia coli* cell
460 populations at different stages during growth transition to stationary phase," *Molecular
461 Microbiology*, vol. 43, pp. 269-279, Jan 2002.
- 462 [35] A. S. Whiteley, M. R. Barer, and A. G. O'Donnell, "Density gradient separation of active and
463 non-active cells from natural environments," *Antonie Van Leeuwenhoek International
464 Journal of General and Molecular Microbiology*, vol. 77, pp. 173-177, Feb 2000.
- 465 [36] D. A. Siegele, M. Almirón, and R. Kolter, "Approaches to the study of survival and death in
466 stationary-phase *Escherichia coli*," in *Starvation in Bacteria*, S. Kjelleberg, Ed., ed New York:
467 Plenum Press, 1993, pp. 151-169.
- 468 [37] J. L. Zilles, J. Peccia, M. W. Kim, C. H. Hung, and D. R. Noguera, "Involvement of Rhodocyclus-
469 related organisms in phosphorus removal in full-scale wastewater treatment plants," *Applied
470 and Environmental Microbiology*, vol. 68, pp. 2763-2769, Jun 2002.
- 471 [38] B. Laycock, M. V. Arcos-Hernandez, A. Langford, S. Pratt, A. Werker, P. J. Halley, *et al.*,
472 "Crystallisation and fractionation of selected polyhydroxyalkanoates produced from mixed
473 cultures," *N Biotechnol*, 2013.
- 474 [39] C. L. Wang, Y. Mitsuya, B. Gharizadeh, M. Ronaghi, and R. W. Shafer, "Characterization of
475 mutation spectra with ultra-deep pyrosequencing: Application to HIV-1 drug resistance,"
476 *Genome Research*, vol. 17, pp. 1195-1201, Aug 2007.
- 477 [40] E. B. Hollister, A. S. Engledow, A. J. M. Hammett, T. L. Provin, H. H. Wilkinson, and T. J.
478 Gentry, "Shifts in microbial community structure along an ecological gradient of hypersaline
479 soils and sediments," *Isme Journal*, vol. 4, pp. 829-838, Jun 2010.
- 480 [41] G. R. Crocetti, J. F. Banfield, J. Keller, P. L. Bond, and L. L. Blackall, "Glycogen-accumulating
481 organisms in laboratory-scale and full-scale wastewater treatment processes," *Microbiology-
482 Sgm*, vol. 148, pp. 3353-3364, Nov 2002.

483

484 Table 1. Buoyant Density separations showing different PHA content and monomer
 485 composition in the pellet and float for 40% Percoll separated samples

486

Fractions	PHA Content* (wt% ±s.d.)	TSS (g)	PHA Composition*	
			(mol% ± s.d.)	
			HB	HV
Control ¹	19.8±0.2	0.101	57.5 ±0.7	43.5±0.7
Pellet	23.8±0.6	0.053	56.5±0.7	44.5±0.7
Float	16.1±4.9	0.037	57.5±2.1	43.5±2.1

487

488

¹ Control: 0% Percoll solution, same wash procedure as others

489

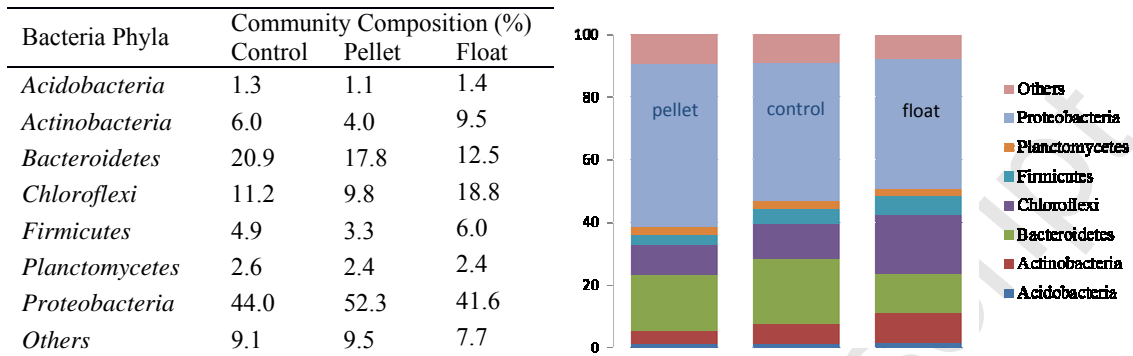
* duplicate samples for PHA content and composition

490

491

492

493



494

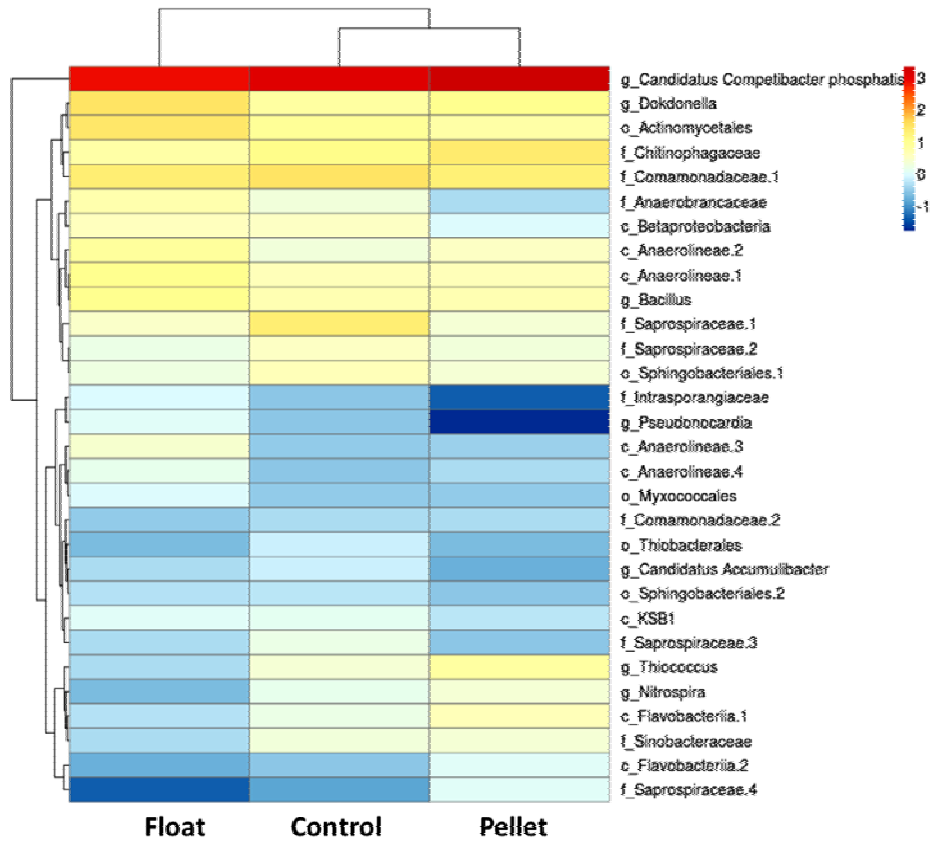
495 Figure 1. Phylum level microbial community of pellet and float sample separated from 40%
 496 Percoll and control from 0% Percoll.

497

498

499

500



501

502

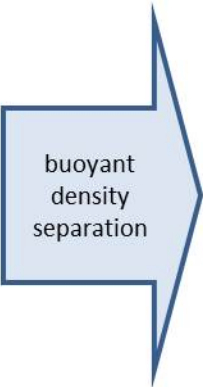
503 Figure 2. Heat map for relative abundances (after transformation) of top 30 abundant OTUs.

504 Each OTU is labelled to the lowest known classification (g: genus, f: family, o: Order, c:

505 class) from greengenes (or genebank) database. Clusters based on Euclidean distance.



**mixed
culture**



buoyant
density
separation

FLOAT

sub-community 1

16 wt% PHA

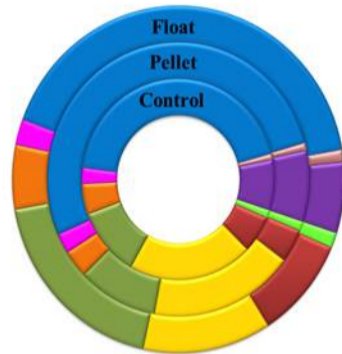
57% HB

PELLET

sub-community 2

24 wt% PHA

56% HB



■ *Acidobacteria*

■ *Actinobacteria*

■ *Bacteroidetes*

■ *Chloroflexi*

■ *Firmicutes*

■ *Planctomycetes*

■ *Proteobacteria*

■ *Caldithrix*

■ *Others*