1	Reframing landscape fragmentation's effects on ecosystem services
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3	Matthew G.E. Mitchell ¹ , Andrés F. Suarez Castro ¹ , Maria Martinez-Harms ² ,
4	Martine Maron ¹ , Clive McAlpine ¹ , Kevin J. Gaston ³ , Kasper Johansen ¹ , Jonathan
5	Rhodes ^{1,2}
6	
7	¹ School of Geography, Planning and Environmental Management, The University
8	of Queensland, Brisbane, QLD 4072, Australia
9	² Australian Research Council Centre of Excellence for Environmental Decisions,
10	School of Biological Sciences, The University of Queensland, Brisbane, QLD
11	4072, Australia
12	³ Environment and Sustainability Institute, University of Exeter, Penryn,
13	Cornwall TR10 9FE, United Kingdom

14 Abstract

15 Landscape structure and fragmentation have important effects on ecosystem 16 services, with a common assumption that fragmentation reduces service 17 provision. This is based on fragmentation's expected effects on ecosystem 18 service supply, but ignores how fragmentation influences the flow of services to 19 people. Here, we develop a new conceptual framework that explicitly considers 20 the links between landscape fragmentation, the supply of services, and the flow 21 of services to people. We argue that fragmentation's effects on ecosystem service 22 flow can actually be positive or negative and use our framework to construct 23 testable hypotheses about the effects of fragmentation on final ecosystem service 24 provision. Empirical efforts to apply and test this framework are critical to 25 improve landscape management for multiple ecosystem services. 26 27 **Keywords:** landscape fragmentation, ecosystem services, ecosystem service

28 flow, ecosystem service supply, biodiversity

29 Landscape fragmentation: the need to reconceptualize for ecosystem

30 services

31 Humans continue to heavily modify natural ecosystems around the world, 32 with negative consequences for biodiversity (see Glossary) and natural capital 33 [1,2]. At the same time, demand for ecosystems to provide benefits, or services, 34 to society is growing rapidly [3]. This has significantly increased the need to 35 understand and manage landscapes simultaneously for ecosystem services and 36 biodiversity. Recently, the potential of managing landscape structure [4-6], and 37 in particular landscape fragmentation [7,8], for these multiple goals has been highlighted. Interest in landscape fragmentation - the breaking apart of areas of 38 39 natural land cover into smaller pieces independent of a change in the amount of 40 natural land cover - has a long history in ecology [9]. Consequently, a well-41 developed understanding exists of its effects on biodiversity and ecosystem 42 functioning [10]. However, the shift in research interest from biodiversity 43 towards the concept of ecosystem services has recast what before were solely 44 ecological questions into social-ecological ones [11-13]. This recasting means 45 that predictions about the ecological effects of landscape fragmentation on 46 biodiversity and ecosystem functioning are unlikely to translate directly into 47 ecosystem service provision. This will be especially true if fragmentation has 48 contrasting effects on people and how they interact with ecosystems to produce 49 ecosystem services compared to biodiversity and ecosystem functioning. It is 50 therefore critical to rethink how fragmentation alters all of the components of 51 ecosystem service provision in order to improve landscape management for 52 multiple services.

53	Ecosystem service provision depends on three elements: supply, demand,
54	and <i>flow</i> (Figure 1), each of which can respond differently to landscape
55	fragmentation. Ecosystem service supply is the potential for natural capital to
56	generate a benefit for people, irrespective of it being realized or used [14]. In
57	turn, ecosystem service demand is the level of service provision desired or
58	required by people, and is influenced by human needs, values, cultures,
59	institutions, and built capital [15]. Finally, for ecosystem service provision to be
60	realized, people must interact with ecosystems to gain a benefit. This interaction
61	connects service supply with demand to produce a service flow: the actual
62	delivery of a service to people to be used or enjoyed [15].
63	Here, we argue that the effects of fragmentation on ecosystem service
64	supply and flow can either complement or oppose each other, leading to
65	contrasting net effects on service provision. Ecosystem service supply depends
66	on the presence of particular species, ecosystems, or ecological processes that
67	are often negatively affected by fragmentation. In contrast, most ecosystem
68	service flows depend on the distribution and movement of organisms, matter,
69	and people between areas of natural and anthropogenic land cover. For example,
70	fragmentation of forests from logging, road construction, or agricultural and
71	urban expansion can alter plant species composition and growth, negatively
72	affecting water quality regulation and carbon sequestration [16,17]. At the same
73	time, this fragmentation can improve forest access, increasing timber harvesting,
74	hunting, wild food foraging, and park visits [18,19]. Thus, by altering the
75	arrangement of areas of service supply and demand, or humans and natural
76	capital across a landscape, fragmentation can modify ecosystem service supply,
77	movements critical for service flow, and ultimately service provision.

78	That landscape fragmentation simultaneously affects ecosystem service
79	supply and flow has not, thus far, been widely acknowledged in the development
80	and application of the ecosystem service concept. The majority of ecosystem
81	service studies that consider fragmentation focus only on service supply [4,20]
82	and disregard demand and flow. Similarly, most ecosystem service decision-
83	support and quantification tools focus on service supply and have limited ability
84	to determine flow [21]. While tools such as InVEST
85	(naturalcapitalproject.org/InVEST.html) and ARIES (ariesonline.org) aim to
86	better quantify service flows across landscapes, integration of this information
87	into decision-making remains limited and is still mainly focused on service
88	supply. Consequently, predictions about how landscape fragmentation will affect
89	ecosystem service provision are likely to be incorrect. This has important
90	implications for landscape planning to optimize service provision.
91	To spur research in this area, we present a conceptual framework that
92	links fragmentation explicitly with ecosystem service supply and flow, and use it
93	to make testable predictions about the effects of landscape fragmentation on
94	ecosystem service provision. We discuss how fragmentation could drive
95	tradeoffs and synergies among services, highlighting the implications for policy
96	and planning, and identify future research priorities for investigating the role of
97	landscape fragmentation on ecosystem service provision.
98	
99	Linking fragmentation to ecosystem service supply and flow
100	Here, we identify specific mechanisms by which landscape fragmentation,

- 101 independent of the loss of natural land cover, affects service supply and flow
- 102 (Figure 1), and the ultimate consequences of these relationships for service

103 provision. A planning issue of critical importance in many human-dominated 104 landscapes is how spatially to arrange areas of natural land cover within the 105 human-dominated matrix [22,23]. While we recognize that alteration of the 106 spatial arrangement of natural land cover also has important consequences for 107 landscape heterogeneity, our framework simplifies this complexity by focusing 108 on fragmentation of natural land cover. We feel this is a necessary first step to 109 better develop a spatially explicit landscape-scale understanding of ecosystem 110 services.

111

112 Fragmentation and ecosystem service supply

113 Fragmentation tends to drive biodiversity loss and shifts in ecosystem 114 function [24,25], although a variety of responses can occur, especially at low or 115 intermediate levels of fragmentation [9]. Fragmentation often reduces the ability 116 of plant and animal species to move across landscapes, interrupting daily 117 movements between foraging and breeding habitat, dispersal events, and 118 migration [10]. In addition, smaller habitat patches support fewer species, 119 contain smaller populations that are at greater risk of extinction [26], and have 120 increased edge effects that can negatively affect the persistence of native species 121 [27]. Each of these different effects of fragmentation can result in degradation of 122 the natural capital and biodiversity that contribute to service supply (Figure 1). 123 There is widespread evidence that biodiversity influences or is strongly 124 correlated with the supply of many ecosystem services [28,29]. For example, 125 increased tree species richness [30] and plant diversity [6] are each associated 126 with an increased supply of multiple ecosystem services. In particular, 127 biodiversity is increasingly important as the number of services considered

128 increases [31]. Thus, if biodiversity declines with landscape fragmentation, as is 129 commonly observed [10], ecosystem service supply will also likely be lost. 130 Pollination and pest regulation are among the best-studied examples 131 where landscape fragmentation drives this relationship. Increased species and 132 functional diversity in pollinator or arthropod predator communities can 133 increase service supply [32,33]. In turn, this diversity can be enhanced by 134 increased forest and grassland connectivity or increased landscape complexity (smaller fields, more hedgerows) across agricultural landscapes [34,35]. 135 136 Fragmentation can also affect forest plant diversity and the supply of carbon storage and sequestration [17,36], although this effect is not universal [37]. 137 138 Similarly, fragmentation of marine ecosystems and rivers can have significant 139 effects on aquatic biodiversity and fish abundance important for commercial 140 fisheries [38,39]. Unfortunately, most of these examples only quantify service 141 supply and not actual flows to people, which might be affected very differently by 142 fragmentation.

143

144 Fragmentation and ecosystem service flow

145 For most ecosystem services, their flow depends on the movement of 146 organisms, matter, energy, and/or people across landscapes to connect spatially 147 separate locations of supply and demand (Figure 1)[20]. For example, pollination 148 depends on the movement of native pollinators from fragments of non-crop 149 vegetation into fields [40], drinking water provision relies on the flow of above-150 and below-ground water to areas of collection or consumption [41], and the 151 movement of people to fishing locations or parks is needed for fisheries and 152 recreation [42]. Conversely, some services depend on ecosystems restricting

flows of organisms or matter. For example, flood regulation is provided when
ecosystems restrict or delay water flow [43], disease regulation when the
movements of disease vectors to people are limited [44], and water quality
regulation when ecosystems capture or transform excess nutrients, sediments or
pollutants [41].

158 Because ecosystem service flow relies on facilitating or restricting 159 movement, landscape fragmentation can affect the magnitude and spatial pattern 160 of these flows (Box 1)[20]. Importantly, fragmentation increases the 161 interspersion of natural and anthropogenic lands, reducing distances between 162 areas of service supply and demand, and potentially increasing service flow. At 163 the same time, fragmentation affects the number, size, shape, spatial 164 arrangement, and isolation of patches of natural land cover, which in turn can 165 positively or negatively affect the flow of soil, water, energy, and organisms 166 across landscapes [4]. Thus, fragmentation can have either negative or positive 167 effects on service flow, depending on the service in question, the process of 168 landscape fragmentation, and the resulting landscape structure (Box 1). In 169 addition, the flow of some ecosystem services will be insensitive to 170 fragmentation. For example, carbon sequestration and storage provides climate 171 regulation globally regardless of its spatial location or the location of 172 beneficiaries. 173

174 How fragmentation affects ecosystem service flow

175 Increased interspersion of natural and anthropogenic lands

Expansion of human land-use resulting in the fragmentation of natural
land cover can place areas of service supply and demand in closer proximity to

178 one another. For services that rely on the juxtaposition of ecosystems and 179 people, this can increase service flows (Figure 2A). Services provided by mobile 180 organisms often fall into this category. For example, interspersion of remnant 181 forests and grasslands with cropland can maximize both pollination and pest 182 regulation services [45]. Small reservoirs of regularly-placed natural land cover 183 that provide shelter and nesting resources can more evenly distribute pollinators 184 across agricultural landscapes and are predicted to maximize the flow of pollination services [22]. Similarly, regularly-spaced forest patch and hedgerow 185 186 reservoirs of arthropod predators are needed to ensure an even flow of pest 187 regulation to agricultural fields [46,47].

188 Increased fragmentation can also improve people's access to ecosystems 189 to obtain recreational and health benefits. Increased visitation to parks and 190 previously inaccessible wilderness areas when roads and trails are built can 191 increase fishing, hunting, timber harvesting, and land clearing [18,19]. Similarly, 192 in urban areas having nearby green spaces increases accessibility and can 193 improve human health and well-being [48,49]. We predict that these effects of 194 fragmentation on patterns of human movement, while often overlooked in the 195 literature [4], will be as common and important for ecosystem service flow as 196 those on the movement of other organisms.

Increased interspersion of people, their activities, and ecosystems can
also increase flows of ecosystem disservices (damages or costs to people from
ecosystems). For example, the spread of human diseases via biotic vectors is
often greater when human habitation occurs in close proximity to natural areas.
For Lyme's disease in North America, increased interspersion of people and
forests is highly correlated with disease prevalence [50,51].

204 Increased isolation of patches of natural land cover

205 By isolating patches of natural land cover and reducing patch sizes, 206 fragmentation can have negative effects on the movement of organisms and 207 matter (Figure 2B). This is especially true if the intervening matrix impedes 208 movement between patches. For services provided by mobile organisms [52], 209 including pollination and seed dispersal, isolation can negatively affect service 210 flow. For example, seed dispersal can be highly sensitive to forest fragmentation 211 by agriculture, especially the loss of small forest patches that maintain landscape 212 connectivity [53]. Services that rely on the movement of water can also be 213 disproportionately affected. The presence of dams has fragmented most of 214 Earths' major river systems, reducing water flow and the movement of people 215 along these rivers, altering water provision to people, water quality regulation 216 [54], and opportunities for recreation [55,56].

217

218 Decreased patch size and increased edge

219 Reduced patch size can decrease visitation rates and ecosystem service 220 flows, for both organisms and people (Figure 2C). For example, smaller fields 221 often experience less pollinator visitation compared to larger fields, with 222 consequences for pollination and other services provided by mobile organisms 223 [34,57]. Similarly, small parks attract fewer visitors from surrounding urban 224 areas [58], reducing recreation [59] and other cultural services. 225 For those services that depend on restricting movement, increases in edge 226 and edge: area ratio can have a variety of effects, either reducing or increasing

service flow to people (Figure 2D). For example, fragmentation of areas of

natural land cover by agriculture can result in greater vegetation-field edge and
increased soil erosion [60,61] and nutrient loss [62,63], with consequences for
downstream water quality. Contrastingly, linear patches of vegetation such as
hedgerows can fragment the cropland matrix of agricultural landscapes,
intercepting pesticides and odors and increasing air quality regulation [64,65].
Other directionally-provided ecosystem services, such as storm protection and
flood regulation might also be improved by more linear wetlands [66].

236 Consequences for ecosystem service provision

237 The varied processes by which fragmentation affects landscape structure 238 and heterogeneity, and thereby service flow, means that fragmentation's effects 239 on supply and flow can be in parallel or opposition. We argue that this will result 240 in a variety of landscape-scale fragmentation effects on the provision of different 241 services, and hypothesize that three broad categories of effects are possible (Box 242 2). For example, when the effects of fragmentation on supply and flow oppose 243 each other, service provision will peak at intermediate levels of fragmentation 244 (Figure 3F). These three categories of relationships provide testable predictions 245 of the effects of fragmentation on service provision.

The diverse effects of fragmentation on service provision will also drive positive and negative relationships between services in fragmented landscapes as each responds differently to the modified landscape structure, even if the services themselves do not interact strongly [67]. Importantly, our framework predicts that tradeoffs and synergies between ecosystem services might not always be unidirectional or constant, but could vary depending on the level of landscape fragmentation. Thus, we predict that managing landscape structure

for ecosystem services does not simply involve minimizing fragmentation, but
requires a much more complete understanding of the effects of landscape
structure on service provision.

256

257 Challenges for ecosystem service science and policy

258 The challenge of incorporating the ecosystem services paradigm into 259 environmental policy and landscape planning is increasingly being recognized 260 [68,69]. The next major challenge is to develop a body of predictive theory to 261 support policy and planning activities, similar to that currently present in biodiversity-fragmentation research. In this context, ecosystem service research 262 263 needs to move away from simply quantifying and mapping the biophysical 264 supply of services [70], and towards identifying locations of service demand, and 265 potential pathways and magnitudes of service flow [15,20]. Understanding these 266 different aspects of service provision and what features of landscape structure, 267 fragmentation, and heterogeneity control them will significantly improve the 268 ability to manage landscapes for ecosystem services. Our framework is a first 269 step towards a more robust theory linking landscape structure with ecosystem 270 services.

We propose that ecosystem service supply will decline with increasing fragmentation, but that the flow of ecosystem services to beneficiaries can increase or decrease. Thus, fragmentation of the landscape can either enhance or degrade ecosystem service provision (Box 2). We also argue that the responses of ecosystem service flow to fragmentation are driven by: (a) increased interspersion of anthropogenic and natural lands, (b) increased isolation of patches of natural land cover, and (c) reduced patch sizes and increased amounts

278 of edge. These predictions reflect a number of important gaps in current 279 knowledge and highlight a number of key research questions that will best 280 address them (Box 3). In particular, testing our hypotheses across landscape 281 gradients of fragmentation by quantifying the supply, demand, and flow of 282 multiple services is an essential next step. Only in this way will the mechanisms 283 by which fragmentation drives both service provision and tradeoffs between 284 services be identified. Describing the precise form of the relationships between fragmentation and service provision, and identifying if distinct classes of 285 286 relationships exist, similar to those in our framework, are also critical questions 287 for future research.

288 Landscape planning almost always involves decisions about the spatial 289 arrangement of conflicting land-uses that influence the level of landscape 290 fragmentation (e.g. [71]). Active urban and rural landscape planning could 291 benefit substantially from a more nuanced understanding of the relationships 292 between landscape fragmentation and heterogeneity, and ecosystem service 293 provision. Yet implications for other globally relevant policy challenges are 294 equally important. Understanding when and why fragmentation inhibits or 295 enhances ecosystem service provision is central to the land sparing versus land 296 sharing (or wildlife-friendly farming) debate [23,72]. This is also true for 297 designing rules to improve the effectiveness and co-benefits from trades in 298 carbon markets (e.g. REDD+)[73], biodiversity (e.g. offsetting, agri-environment 299 schemes) [5,74], and other ecosystem services (e.g. water quality). Market-based 300 approaches to stimulate desirable land-use outcomes are also increasingly 301 incorporating effects of spatial configuration [75], but currently incorporate only 302 a simple understanding of the consequences of fragmentation. Thus,

303 understanding the effects of fragmentation on ecosystem services is of critical304 importance for developing effective policy mechanisms.

305

306 Concluding Remarks

307 Our conceptual framework highlights the vital importance of 308 understanding how fragmentation of natural land cover affects service supply 309 and flow and the different ecological and social components of ecosystem service 310 provision. Incorporating these effects into ecosystem service assessments is 311 critical to develop effective tools that can help structure landscapes to provide 312 multiple ecosystem services. In many ways, the field of ecosystem services is 313 ideally placed to address this challenge; many studies already work at large 314 spatial scales across landscapes with different levels of fragmentation, and 315 incorporate data from a diversity of sources, including ecological, remote 316 sensing, and social survey data. What is needed now is increased empirical 317 research into the exact nature of the relationships between fragmentation and 318 ecosystem service supply and flow. As the ecosystem services concept is 319 increasingly incorporated into decision-making and planning activities, the need 320 to improve understanding of ecosystem service provision at the landscape-scale 321 is fundamentally important.

322

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332 Figure 1. A conceptual diagram of the effects of landscape fragmentation on the 333 provision of ecosystem services. Fragmentation alters ecosystem service supply 334 by affecting natural capital. This occurs when fragmentation affects the 335 movement and distribution of organisms, matter, and energy across a landscape, 336 with consequences for the biodiversity and ecosystem functions that are 337 important for service provision. Fragmentation also affects patterns of human 338 distribution, activities, and movement across the landscape. Combined, these 339 effects influence the magnitude and spatial pattern of ecosystem service flows 340 that connect areas of service supply to areas of demand. Thus, ecosystem service 341 flows, and ultimately service provision, depend on how landscape fragmentation 342 and the resulting landscape structure affect the movement and distribution of 343 both ecosystems and people. In turn, the benefit derived from an ecosystem 344 service affects service demand by altering human wellbeing and needs. This 345 demand then drives human activities that alter landscape fragmentation (dashed 346 arrow). Ecosystem service provision can also directly affect natural capital 347 (dashed arrow) through over-exploitation. Adapted from [14].

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Box 1. What is landscape fragmentation and how does it affect ecosystem

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350 service flow?
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Landscape fragmentation is the breaking up of larger areas of natural land cover into smaller, more isolated patches, independent of a change in the total area of natural land cover (Figure 2). Landscape fragmentation causes three main interconnected changes to patches of natural land cover across a landscape: (i) an increase in the isolation of patches and their interspersion with

555 landscape. (1) an increase in the isolation of patches and then interspersion with

the surrounding human-dominated land (e.g., agricultural or urban areas), (ii) an

357 increase in the number of patches and the amount of patch edge, and (iii) a 358 decrease in average patch area [9]. Simultaneously, the surrounding human-359 dominated portion of the landscape can become more connected as 360 fragmentation proceeds, with important consequences for the movement and abundance of species that inhabit this portion of the landscape [52,76]. 361 362 Thus, landscape fragmentation results in a number of interrelated effects 363 for landscape structure, including changes to landscape configuration and 364 heterogeneity. This means that a variety of mechanisms and effects on ecosystem 365 service flow are possible (Figure 2). Fragmentation affects ecosystem service 366 flows by facilitating or interrupting movement of organisms, matter, energy, and 367 people across landscapes. This includes the daily movements of mobile 368 organisms like pollinators and insect predators across agricultural landscapes; 369 long-distance migrations; directional overland flows of water and the nutrients, 370 pollutants, and eroded soil it contains; ocean and atmospheric currents at 371 multiple spatial scales; and the movement of people across landscapes. The final 372 effect of fragmentation on service provision will depend heavily on these processes and the key species, ecosystem functions, biophysical flows, and 373 374 human activities that underlie each service, as well as the exact form and amount 375 of landscape fragmentation that takes place. Additionally, the scale at which 376 fragmentation takes place relative to ecosystem service flow will also change 377 how it affects service provision. 378

380 change in the area of natural land cover, can affect ecosystem service flow.

379

381 Locations of natural land cover and ecosystem service supply (*green* areas)

17

Figure 2. The mechanisms by which landscape fragmentation, independent of a

382 provide ecosystem service flows (red arrows) and benefits (red areas) to the 383 human-dominated matrix (*light brown* areas) that are affected by landscape 384 fragmentation. Ecosystem service flows of organisms and people (arrows) can 385 depend on proximity to natural areas (A) and will therefore be influenced by the 386 interspersion of natural and anthropogenic land cover across the landscape (e.g., 387 recreation, pollination, waste treatment, pest regulation). At the same time, 388 increased isolation of patches and reduced connectivity (**B**), as well as decreased 389 patch size (C), can decrease service flow in fragmented landscapes (e.g., 390 pollination, seed dispersal, cultural services, watercourse recreation, water 391 provision and regulation). Finally, for services that depend on restricting 392 movement across landscapes, increased edge amounts with fragmentation (**D**) 393 can have positive (e.g., storm protection, air quality regulation) or negative (e.g., 394 water quality or soil erosion regulation) effects on ecosystem service flow. In 395 each panel, the area of natural land cover and ecosystem service supply is 396 unchanged between intact and fragmented landscapes. Adapted from [66]. 397 398 Box 2. Combining the effects of fragmentation on ecosystem service supply 399 and flow 400 Our conceptual framework predicts that a range of relationships between 401 landscape fragmentation and final ecosystem service provision are possible 402 depending on the specific processes by which fragmentation affects service 403 supply and flow (Figure 3). While a range of effects is likely, we identify three 404 general categories of effects: 405 (1) *Double Whammy*: fragmentation negatively affects both supply and flow, 406 resulting most often in rapid and dramatic decreases in ecosystem service

407 provision with fragmentation. We predict this relationship for services 408 where reduced connectivity and decreased patch size drive reductions in 409 service flow (e.g., water provision and regulation, watercourse recreation, 410 and pollination and pest regulation at high levels of landscape 411 fragmentation). 412 (2) *Compensating*: the effects of fragmentation on flow oppose those on 413 supply, resulting in increased service provision at intermediate levels of fragmentation. The exact level of fragmentation that maximizes service 414 415 provision depends on the strength and shape of the relationship between 416 fragmentation and service flow. Services where interspersion of natural 417 land cover and human-dominated areas determines service flow should 418 respond in this way (e.g., recreation, cultural and aesthetic services, 419 genetic resources, pollination, and pest regulation) 420 (3) *Supply Driven*: ecosystem service flows are insensitive to fragmentation, 421 therefore final service provision is simply a function of the effects of 422 fragmentation on service supply. Examples include carbon sequestration, 423 carbon storage, and the existence value of biodiversity. 424 Because there is a wide range of possible patterns of ecosystem service provision 425 with fragmentation, this will drive synergies and tradeoffs between services in 426 fragmented landscapes. For example, services that respond in 'Double Whammy' 427 or 'Supply Driven' ways to fragmentation might show positive relationships 428 across landscapes as fragmentation varies. Of course, variation in the strength of 429 these relationships will also occur (e.g., *blue* versus *red* lines in Figure 3E). 430 Contrastingly, tradeoffs might occur among services following a 'Compensating' 431 relationship. Here, the strength of the trade-offs between services will depend on

the level of fragmentation and resulting landscape structure. Tradeoffs and
synergies between services and switches between the two could also occur
within the 'Compensating' category as levels of fragmentation vary (e.g., *green dashed* versus *blue solid* line in Figure 3F). Thus, our framework predicts that
tradeoffs and synergies between services might not always be unidirectional or
constant, but will vary depending on the level of landscape fragmentation.

438

Figure 3. Effects of landscape fragmentation on the supply and the flow of 439 440 ecosystem services will affect the final relationship between landscape 441 fragmentation and ecosystem service provision. Landscape fragmentation, by 442 reducing biodiversity and ecosystem function, is (A) predicted to reduce 443 ecosystem service supply (three alternative possible trajectories are shown: red, 444 green, and blue lines). At the same time, the amount of flow per unit of ecosystem 445 service supply to beneficiaries can also be affected (**B**) negatively, (**C**) positively, 446 or (**D**) be insensitive to landscape fragmentation (e.g., carbon sequestration), 447 with a range of relationships possible (e.g., *solid*, *dashed*, and *dotted lines*). 448 Combining ecosystem service supply and flow multiplicatively (E,F,G) will result 449 in distinct relationships between landscape fragmentation and ecosystem 450 service provision. Each of the trend lines in (**E**,**F**,**G**) is a combination of the lines 451 in the plots above. Note that some lines overlap in (E) and for clarity not all 452 possible combinations of supply and flow are shown; the *grey* lines in (E) show 453 what provision would be if flow was insensitive to fragmentation. 454 455 Box 3. Outstanding questions about the effects of fragmentation on

456 ecosystem services

(1) What are the specific relationships between landscape fragmentation and
ecosystem service supply and flow for different services? While there is
likely wide variation in the form of these relationships, this has yet to be
quantified. This is a key first step to creating landscape management tools
for ecosystem services that deal with fragmentation.

462 (2) What are the important mechanisms by which fragmentation affects
463 service flow for different ecosystem services, and do these vary
464 depending on spatial scale considered? We identify four potential
465 mechanisms, but their relative importance across different services is
466 largely unknown. Understanding these mechanisms is key to creating a
467 predictive framework for the effects of landscape fragmentation on
468 ecosystem service provision.

(3) Can the relationships between fragmentation and ecosystem service flow
and final provision be generalized for specific categories of services?
While we identify three broad potential categories (Figure 3), there might
be additional categories or there might be instances where relationships
between services and fragmentation are idiosyncratic depending on the
scale of fragmentation or other biophysical and social factors. While we
hypothesize that this is unlikely, it has yet to be tested.

476 (4) How are positive or negative relationships between ecosystem services
477 affected by landscape fragmentation? Our framework predicts that these
478 relationships might not be constant, but could vary across gradients of
479 fragmentation or landscape structure. The prevalence and actual form of
480 these relationships need to be tested in real landscapes.

481	(5) How can the effects of fragmentation on ecosystem service provision be
482	effectively integrated into decision-making? The causes of fragmentation
483	across landscapes are varied and it can often be driven by external factors
484	such as demand for ecosystem services from distant locations. Therefore,
485	effectively integrating knowledge about the effects of fragmentation into
486	landscape planning will likely be difficult and effective paths to do this are
487	yet to be explored.
488	(6) What is the most important component of ecosystem service provision

489 (i.e., supply or flow) to understand with respect to landscape planning?

490 With limited resources available to investigate how fragmentation affects

491 both service supply and flow, determining which is most important for

492 landscape management is critical to efficient decision-making.

- 493 Glossary
- 494

495	Benefit: the ways in which ecosystems improve human wellbeing via the
496	provision of ecosystem services. Constituents of human wellbeing include
497	materials essential for life, and contributions to health, security, social relations,
498	and freedom of choice and action [77].
499	Biodiversity: the variability among living organisms from all sources including,
500	inter alia, terrestrial, marine and other aquatic ecosystems and the ecological
501	complexes of which they are part; this includes diversity within species, between
502	species and of ecosystems. Defined here following the 1993 Convention on
503	Biological Diversity (CBD) meaning of 'biological diversity', which we assume is
504	equivalent to 'biodiversity' (www.cbd.int/convention/articles).
505	Connectivity : the degree to which a landscape facilitates the movement of
506	organisms and matter [78]. We use the term to include both biotic connectivity
507	(movement of organisms) and abiotic connectivity (movement of water,
508	nutrients, and soil) across landscapes.
509	Ecosystem function: the flow of energy and materials through the arrangement
510	of biotic and abiotic components of an ecosystem that allow or could allow
511	natural systems to provide ecosystem services [79].
512	Ecosystem service: defined broadly, the biophysical and social conditions and
513	processes by which people, directly or indirectly, obtain benefits from
514	ecosystems that sustain and fulfill human life [77].
515	Ecosystem service demand: the level of service provision desired or required
516	by people. Demand is influenced by human needs, values, institutions, built

517 capital, and technology [15].

518	Ecosystem	service flo	w : the actual	delivery to	or realization	of an	ecosystem
	5			5			5

service by people. Ecosystem service flow depends on both the supply of and

520 demand for a service [14,15] as well as the movement of organisms, matter, and

521 people [4].

522 **Ecosystem service supply**: the full potential of ecological functions or

523 biophysical elements in an ecosystem to provide a given ecosystem service,

524 without consideration of whether humans recognize, use, or value that function

525 or element [14,15].

526 Landscape: a heterogeneous area composed of interacting ecosystems that is

527 repeated in similar form throughout, including both natural and anthropogenic

528 land covers, across which humans interact with their environment [80].

529 Landscape fragmentation: the breaking apart of areas of natural land cover

530 into several smaller areas within a human-dominated matrix, without any

531 change in the area of natural land cover [9].

532 **Landscape heterogeneity**: the amount of variation in landscape structure

533 (composition and configuration) at a particular spatial scale across a landscape.

534 Landscape heterogeneity is affected by landscape fragmentation through

535 changes to patterns of spatial complexity.

536 Landscape structure: the arrangement of land covers and land uses across a

537 landscape. Broadly, it includes landscape composition (how much of each land

538 cover or land use that exists), configuration (the spatial pattern of these land

539 cover or land use types), and connectivity.

540 **Landscape matrix**: the surrounding portion of the landscape in which

541 fragments of natural land cover are located. In most cases we consider the matrix

to be the human-dominated or disturbed areas of the landscape (e.g., agricultural

- 543 fields, urban areas, cleared land). Characteristics of the matrix can be important
- 544 for determining landscape connectivity and ecosystem service flow.
- 545 **Natural capital**: the stock of natural ecosystems, including all of their biological
- 546 and physical features that supply flows of ecosystem services to people.

547 **References**

- 548 1 Butchart, S.H.M. *et al.* (2010) Global biodiversity: indicators of recent
 549 declines. *Science* 328, 1164–1168
- 550 2 Foley, J.A. *et al.* (2011) Solutions for a cultivated planet. *Nature* 478, 337–
- 551 342
- 552 3 Carpenter, S.R. *et al.* (2009) Science for managing ecosystem services:
- Beyond the Millennium Ecosystem Assessment. *Proc. Natl. Acad. Sci. U.S.A.*106, 1305–1312
- 555 4 Mitchell, M.G. *et al.* (2013) Linking landscape connectivity and ecosystem
- service provision: current knowledge and research gaps. *Ecosystems* 16,
 894–908
- 558 5 Ekroos, J. et al. (2014) Optimizing agri-environment schemes for
- biodiversity, ecosystem services or both? *Biol. Conserv.* 172, 65–71
- 560 6 Werling, B.P. *et al.* (2014) Perennial grasslands enhance biodiversity and
- 561 multiple ecosystem services in bioenergy landscapes. *Proc. Natl. Acad. Sci.*
- 562 *U.S.A.* 111, 1652–1657
- 563 7 Ziter, C. *et al.* (2013) Functional diversity and management mediate
- aboveground carbon stocks in small forest fragments. *Ecosphere* 4, art85
- 565 8 Dobbs, C. *et al.* (2014) Multiple ecosystem services and disservices of the
- urban forest establishing their connections with landscape structure and
 sociodemographics. *Ecol. Indic.* 43, 44–55
- 568 9 Fahrig, L. (2003) Effects of habitat fragmentation on biodiversity. *Annu. Rev.*569 *Ecol. Evol. S.* 34, 487–515
- 570 10 Fischer, J. and Lindenmayer, D.B. (2007) Landscape modification and habitat
 571 fragmentation: a synthesis. *Global. Ecol. Biogeogr.* 16, 265–280
 - 26

- Hughes, T.P. *et al.* (2005) New paradigms for supporting the resilience of
 marine ecosystems. *Trends Ecol. Evol.* 20, 380–386
- 574 12 Chapin, F.S. *et al.* (2010) Ecosystem stewardship: sustainability strategies
 575 for a rapidly changing planet. *Trends Ecol. Evol.* 25, 241–249
- 576 13 Mace, G.M. (2014) Whose conservation? *Science* 345, 1558–1560
- 577 14 Tallis, H. *et al.* (2012) A global system for monitoring ecosystem service
 578 change. *BioScience* 62, 977–986
- 579 15 Villamagna, A.M. et al. (2013) Capacity, pressure, demand, and flow: A
- 580 conceptual framework for analyzing ecosystem service provision and
- 581 delivery. *Ecol. Complex* 15, 114–121
- 582 16 Edwards, D.P. *et al.* (2014) Maintaining ecosystem function and services in
 583 logged tropical forests. *Trends Ecol. Evol.* 29, 511-520
- 584 17 Pütz, S. *et al.* (2014) Long-term carbon loss in fragmented Neotropical
 585 forests. *Nature Commun.* 5, 5037
- 586 18 Peres, C.A. and Lake, I.R. (2003) Extent of nontimber resource extraction in
- tropical forests: accessibility to game vertebrates by hunters in the Amazon
 basin. *Conserv. Biol.* 17, 521–535
- Trombulak, S.C. and Frissell, C.A. (2000) Review of ecological effects of roads
 on terrestrial and aquatic communities. *Conserv. Biol.* 14, 18–30
- 591 20 Bagstad, K.J. et al. (2013) Spatial dynamics of ecosystem service flows: a
- 592 comprehensive approach to quantifying actual services. *Ecosystem Services*
- 593 4, 117–125
- 594 21 Bagstad, K.J. *et al.* (2013) A comparative assessment of decision-support
- 595 tools for ecosystem services quantification and valuation. *Ecosystem Services*
- 596 5, 27–39

- 597 22 Brosi, B.J. *et al.* (2008) Optimal design of agricultural landscapes for
- 598 pollination services. *Conserv. Lett.* 1, 27–36
- Lin, B.B. and Fuller, R.A. (2013) Sharing or sparing? How should we grow
 the world's cities? *J. Appl. Ecol.* 50, 1161–1168
- 601 24 Leibold, M.A. *et al.* (2004) The metacommunity concept: a framework for
- 602 multi-scale community ecology. *Ecol. Lett.* 7, 601–613
- 603 25 Gonzalez, A. et al. (2009) Biodiversity as spatial insurance: the effects of
- habitat fragmentation and dispersal on ecosystem functioning. In
- 605 *Biodiversity, Ecosystem Functioning, and Human Wellbeing* (Naeem, S. et al.,
- 606 eds), pp. 134–146, Oxford University Press
- 607 26 Dobson, A. *et al.* (2006) Habitat loss, trophic collapse, and the decline of
- 608 ecosystem services. *Ecology* 87, 1915–1924
- 609 27 Harper, K.A. *et al.* (2005) Edge influence on forest structure and
- 610 composition in fragmented landscapes. *Conserv. Biol.* 19, 768–782
- 611 28 Cardinale, B.J. *et al.* (2012) Biodiversity loss and its impact on humanity.
- 612 *Nature* 486, 59–67
- 613 29 Balvanera, P. *et al.* (2006) Quantifying the evidence for biodiversity effects
- on ecosystem functioning and services. *Ecol. Lett.* 9, 1146–1156
- 615 30 Gamfeldt, L. *et al.* (2013) Higher levels of multiple ecosystem services are
- found in forests with more tree species. *Nature Commun.* 4, 1340
- 61731Isbell, F. *et al.* (2011) High plant diversity is needed to maintain ecosystem
- 618 services. *Nature* 477, 199–202
- 619 32 Hoehn, P. *et al.* (2008) Functional group diversity of bee pollinators
- 620 increases crop yield. *P. Roy. Soc. B-Biol. Sci.* 275, 2283–2291
- 621 33 Letourneau, D.K. *et al.* (2009) Effects of natural enemy biodiversity on the

- 622 suppression of arthropod herbivores in terrestrial ecosystems. *Annu. Rev.*
- 623 *Ecol. Evol. S.* 40, 573–592
- Klein, A.-M. *et al.* (2012) Wild pollination services to California almond rely
 on semi-natural habitat. *J. Appl. Ecol.* 49, 723–732
- 626 35 Chaplin-Kramer, R. *et al.* (2011) A meta-analysis of crop pest and natural
- 627 enemy response to landscape complexity. *Ecol. Lett.* 14, 922–932
- 628 36 Numata, I. *et al.* (2011) Carbon emissions from deforestation and forest
- fragmentation in the Brazilian Amazon. *Environ. Res. Lett.* 6, 044003
- 630 37 Ziter, C. *et al.* (2014) Temperate forest fragments maintain aboveground
- 631 carbon stocks out to the forest edge despite changes in community632 composition. *Oecologia* 176, 893-902
- 633 38 Ziv, G. *et al.* (2012) Trading-off fish biodiversity, food security, and
- hydropower in the Mekong River Basin. *Proc. Natl. Acad. Sci. U.S.A.* 109,
 5609–5614
- 636 39 Edwards, H. *et al.* (2010) Incorporating ontogenetic dispersal, ecological
 637 processes and conservation zoning into reserve design. *Biol. Conserv.* 143,
- 638 457-470
- 639 40 Ricketts, T.H. *et al.* (2008) Landscape effects on crop pollination services:
 640 are there general patterns? *Ecol. Lett.* 11, 499–515
- 641 41 Brauman, K.A. *et al.* (2007) The nature and value of ecosystem services: an
- 642 overview highlighting hydrologic services. *Annu. Rev. Env. Resour.* 32, 67–98
- 643 42 Bagstad, K.J. *et al.* (2014) From theoretical to actual ecosystem services:
- 644 mapping beneficiaries and spatial flows in ecosystem service assessments.
- 645 *Ecol. Soc.* 19, art64
- 646 43 Luck, G.W. *et al.* (2009) Protecting ecosystem services and biodiversity in

- 647 the world's watersheds. *Conserv. Lett.* 2, 179–188
- 648 44 Plantegenest, M. *et al.* (2007) Landscape epidemiology of plant diseases. *J.*
- 649 *Roy. Soc. Interface* 4, 963–972
- 45 Tscharntke, T. *et al.* (2005) Landscape perspectives on agricultural
- 651 intensification and biodiversity ecosystem service management. *Ecol. Lett.*652 8, 857–874
- 653 46 Bianchi, F.J.J.A. *et al.* (2010) Spatial variability in ecosystem services: simple
- rules for predator-mediated pest suppression. *Ecol. Appl.* 20, 2322–2333
- 655 47 Otieno, M. *et al.* (2011) Local management and landscape drivers of
- pollination and biological control services in a Kenyan agro-ecosystem. *Biol. Conserv.* 144, 2424–2431
- 48 Takano, T. *et al.* (2002) Urban residential environments and senior citizens'
- 659 longevity in megacity areas: the importance of walkable green spaces. *J.*
- 660 *Epidemiol. Commun. H.* 56, 913–918
- 49 Wolch, J. *et al.* (2011) Childhood obesity and proximity to urban parks and
- recreational resources: a longitudinal cohort study. *Health Place* 17, 207–
- 663 214
- 50 Jackson, L.E. *et al.* (2006) Towards landscape design guidelines for reducing
 Lyme disease risk. *Int. J. Epidemiol.* 35, 315–322
- 666 51 Li, S. *et al.* (2012) Consequences of landscape fragmentation on Lyme
- disease risk: a cellular automata approach. *Plos One* 7, e39612
- 668 52 Kremen, C. et al. (2007) Pollination and other ecosystem services produced
- by mobile organisms: a conceptual framework for the effects of land-use
 change. *Ecol. Lett.* 10, 299–314
- 671 53 Bodin, O. *et al.* (2006) The value of small size: loss of forest patches and

- 672 ecological thresholds in southern Madagascar. *Ecol. Appl.* 16, 440–451
- 673 54 Nilsson, C. et al. (2005) Fragmentation and flow regulation of the world's
- large river systems. *Science* 308, 405–408
- 675 55 Loomis, J. (2002) Quantifying recreation use values from removing dams
- and restoring free-flowing rivers: A contingent behavior travel cost demand
- 677 model for the Lower Snake River. *Water Resour. Res.* 38, 2–1 2–8
- 678 56 Whittaker, D. and Shelby, B. (2002) Evaluating instream flows for
- 679 recreation: Applying the structural norm approach to biophysical
- 680 conditions. *Leisure Sci.* 24, 363–374
- 57 Isaacs, R. and Kirk, A.K. (2010) Pollination services provided to small and
- large highbush blueberry fields by wild and managed bees. *J. Appl. Ecol.* 47,
 841–849
- 684 58 Cohen, D.A. *et al.* (2010) Parks and physical activity: why are some parks
 685 used more than others? *Prev. Med.* 50, S9–12
- 686 59 Van Dyck, D. *et al.* (2013) Associations of neighborhood characteristics with
- 687 active park use: an observational study in two cities in the USA and Belgium.
- 688 Int. J. Health Geogr. 12, 26–34
- 60 Ouyang, W. *et al.* (2010) Soil erosion dynamics response to landscape
 pattern. *Sci. Total Environ.* 408, 1358–1366
- 691 61 Ziegler, A.D. *et al.* (2007) Hydrological consequences of landscape
- 692 fragmentation in mountainous northern Vietnam: buffering of Hortonian
- 693 overland flow. *J. Hydrol.* 337, 52–67
- 694 62 Gergel, S.E. (2005) Spatial and non-spatial factors: When do they affect
- landscape indicators of watershed loading? *Landscape Ecol.* 20, 177–189
- 696 63 Gémesi, Z. et al. (2011) Effects of watershed configuration and composition

697		on downstream lake water quality. J. Environ. Qual. 40, 517–527
698	64	Lazzaro, L. et al. (2008) Role of hedgerows in intercepting spray drift:
699		Evaluation and modelling of the effects. Agr. Ecosyst. Environ. 123, 317–327
700	65	Tyndall, J. and Colletti, J. (2007) Mitigating swine odor with strategically
701		designed shelterbelt systems: a review. <i>Agroforest. Syst.</i> 69, 45–65
702	66	Fisher, B. et al. (2009) Defining and classifying ecosystem services for
703		decision making. <i>Ecol. Econ.</i> 68, 643–653
704	67	Bennett, E.M. et al. (2009) Understanding relationships among multiple
705		ecosystem services. <i>Ecol. Lett.</i> 12, 1394–1404
706	68	Bateman, I.J. et al. (2013) Bringing ecosystem services into economic
707		decision-making: land use in the United Kingdom. Science 341, 45–50
708	69	Perrings, C. et al. (2011) The biodiversity and ecosystem services science-
709		policy interface. <i>Science</i> 331, 1139–1140
710	70	Seppelt, R. <i>et al.</i> (2011) A quantitative review of ecosystem service studies:
711		approaches, shortcomings and the road ahead. J. Appl. Ecol. 48, 630–636
712	71	Phalan, B. et al. (2011) Reconciling food production and biodiversity
713		conservation: land sharing and land sparing compared. Science 333, 1289-
714		1291
715	72	Fischer, J. et al. (2014) Land sparing versus land sharing: moving forward.
716		Conserv. Lett. 7, 149–157
717	73	Jantz, P. et al. (2014) Carbon stock corridors to mitigate climate change and
718		promote biodiversity in the tropics. Nat. Clim. Change 4, 138–142
719	74	Wendland, K.J. et al. (2010) Targeting and implementing payments for
720		ecosystem services: Opportunities for bundling biodiversity conservation
721		with carbon and water services in Madagascar. Ecol. Econ. 69, 2093–2107

722	75	Polasky, S. et al. (2014) Implementing the optimal provision of ecosystem
723		services. Proc. Natl. Acad. Sci. U.S.A. 111, 6248–6253
724	76	Watling, J.I. et al. (2011) Meta-analysis reveals the importance of matrix
725		composition for animals in fragmented habitat. Global Ecol. Biogeogr. 20,
726		209–217
727	77	Millennium Assessment (2005) Ecosystems and human well-being: synthesis,
728		Island Press Washington, DC.
729	78	Taylor, P.D. et al. (1993) Connectivity is a vital element of landscape
730		structure. <i>Oikos</i> 68, 571–573
731	79	Diaz, S. and Cabido, M. (2001) Vive la difference: plant functional diversity
732		matters to ecosystem processes. Trends Ecol. Evol. 16, 646–655
733	80	Forman, R.T. and Godron, M. (1981) Patches and structural components for
734		a landscape ecology. <i>BioScience</i> 31, 733–740

Figure 1

Socioecological system





