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3 **Effect of stocking strategy on distribution and recapture rate of common carp in a large**
4 **and shallow temperate lake: implications for recreational put-and-take fisheries**
5 **management**

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16 Short title: Common carp stocking strategies for angling

17

18 **Summary**

19 It is hypothesized that stocking procedure influences survival, growth and distribution of
20 introduced fishes; however, there is still limited information on the effect of various stocking
21 strategies on recaptures in natural freshwaters. The present study aimed to investigate how the
22 rate and distribution of anglers' common carp (*Cyprinus carpio*) catches vary with the
23 stocking season (i.e. spring, summer and autumn), lake area, method (i.e. shore and offshore
24 releases) and fish size (i.e. ≤ 500 g and > 500 g) in large and shallow Lake Balaton, Hungary.
25 In 2010, 4500 two-summer old, individually tagged common carp were stocked to test 36
26 releasing set-ups (i.e. three seasons \times three lake areas \times two methods \times two size groups).
27 Anglers reported date, location and fish size (standard length and weight) on 787 recaptures
28 within two years after the release. Recapture rate was highest in summer and lowest in
29 autumn stockings, but it was not affected by the stocking area, method and fish size.
30 Recaptures dispersed most in space in stockings carried out in autumn and the centre of the
31 lake, but movement of fish was not influenced by stocking method and fish size. To conclude,
32 in summer, stocking quotas should be evenly distributed along the entire shore line, while
33 early spring stockings may be optimized for transport cost and concentrated by each lake
34 basin. Late autumn stockings should be avoided, and the capacity of effective wintering ponds
35 developed. This study also provides a good framework for testing fisheries management
36 alternatives in other intensively fished habitats.

37

38 **Keywords:** angling, *Cyprinus carpio*, fisheries management, game fishes, mark and recapture,
39 stocking strategy.

40 **Introduction**

41 Recreational fishery (i.e. angling) is displacing commercial fishery from freshwater habitats
42 in developed countries, especially in Central and Western Europe and North America
43 (Hickley, 2009). Intensive angling requires specific fisheries management. In general, anglers
44 are selective and fish for a few valuable species, putting unbalanced pressure on fish
45 assemblages (Vostradovský, 1991; Arlinghaus and Mehner, 2003; Hickley, 2009).
46 Consequently, maintaining high angling activity in freshwaters generally requires stocking of
47 the most important game fishes, especially in areas where captured fish are not returned to the
48 water and natural recruitment is unsatisfactory.

49 In some European and Asian countries, common carp, *Cyprinus carpio* L., is a much
50 preferred commercial and game fish (e.g. Czech Republic: Vostradovský, 1991; Germany:
51 Arlinghaus and Mehner, 2003; United Kingdom: Linfield, 1980; Poland: Wolos et al., 1998).
52 Unfortunately, due to human-induced habitat alterations, wetland draining, floodplain
53 isolation by dykes, overfishing and intensive stocking of domesticated strains, most native
54 wild common carp populations have become endangered (e.g. River Danube subpopulation
55 listed as critically endangered in the IUCN Red List; Kottelat, 1996). Therefore, to maintain
56 common carp populations and establish conditions of intensive fisheries utilization, extended
57 stocking programmes have been implemented mainly in Central Europe.

58 In fisheries oriented stocking programmes, most important indicators of efficiency are the
59 recapture rate, distribution of recaptures in space and time, and size of fish at recapture in
60 relation to resources invested (i.e. number of fish stocked and cost of the project; Patterson
61 and Sullivan, 2013). Generally, in the temperate region, fish stocked in spring have better
62 survival rate than those stocked before the winter (Kennedy et al., 1982; Vostradovský, 1991),
63 and post-stocking survival and recapture rates positively correlate with size of fish at release
64 (Kennedy et al., 1982; Michaletz et al., 2008). Further, place of stocking can also influence

65 survival and recapture rates (Balfry et al., 2011; Vostradovský, 1991; Michaletz et al., 2008),
66 and movement and distribution of the released fish may vary between size-groups and sexes
67 (Young et al., 1999; Stuart and Jones, 2006). However, although common carp is one of the
68 most important game fishes and is stocked in large amounts in freshwater systems, there is
69 still limited information on the effect of different stocking strategies on recaptures in natural
70 waters (Vostradovský, 1991). In addition, to the best of our knowledge, no study has yet
71 investigated the relationship between stocking strategy and recapture rate and distribution of
72 common carp in large water bodies ($>100 \text{ km}^2$).

73 Common carp is native to Hungary including Lake Balaton, but it has lost the majority of
74 its natural spawning and nursery areas due to flood and wetland regulations in the last 100-
75 150 years. Thence, in order to maintain good angling conditions for the 40 000 anglers
76 visiting Lake Balaton each year, 250 to 350 tons of two and three-summer old (1+-2+ age
77 groups) common carp are stocked annually. Up to the present, stockings were implemented
78 primarily on the basis of anglers' catch statistics and available stocking resources (i.e. the
79 capacity of rearing ponds and the income from angling licences). By now, annual stockings
80 have reached their maximum and it has also become evident that traditional catch statistics
81 have little relevance and cannot be used effectively for fisheries management planning.
82 Consequently, strong motivation has arisen from anglers and fisheries managers to examine
83 how the recapture rate and its distribution in space and time could be improved under the
84 given stocking rate and angling activity pattern.

85 Accordingly, the main objectives of the present study were to investigate how the rate and
86 distribution of common carp catches by anglers vary with the stocking season (i.e. spring,
87 summer and autumn), lake area (i.e. Keszthely, Szemes and Siófok), method (i.e. shore and
88 offshore releases) and size of fish at release (i.e. $\leq 500 \text{ g}$ and $>500 \text{ g}$) in Lake Balaton. It was
89 hypothesised that: (i) recapture rate will be higher in spring and summer than autumn

90 stockings when winter mortality is expected before the next angling season (from the end of
91 March to the end of October in common carp), will be positively affected by fish size at
92 release, and will be similar across stocking areas and methods; (ii) mean time between the
93 release and recapture will be shorter in spring and summer than autumn stockings (which is
94 trivial because carp fishing is ineffective in winter), for shore than offshore releases when fish
95 should move to preferred inshore habitats to be exposed to fishing, will be negatively affected
96 by size of fish at release, and will be similar across stocking areas; (iii) fish will travel longer
97 distances and (iv) recaptures will cover larger areas in autumn, central (Szemes) basin and
98 offshore stockings than in other stocking set-ups, and the distribution of fish will be positively
99 affected by size of fish at release.

100

101 **Materials and methods**

102 **Study area**

103 Lake Balaton is the largest shallow lake (surface area: 593 km²; mean depth: 3.2 m) in Central
104 Europe, located at 46° 42' - 47° 04' N, 17° 15' - 18° 10' E and 104.8 m above sea level. The
105 lake is meso-eutrophic with mean annual chlorophyll-*a* concentrations of 3.6-18.7 mg m⁻³
106 (Istvánovics et al., 2007). At present, only 47% of the lake shore is in a natural or semi-
107 natural state and these sections are covered by reed grass *Phragmites australis*. Submerged
108 macrophytes occur sparsely in the littoral zone. A majority of the lake area (>85%) is largely
109 homogeneous open water providing mainly zooplankton and benthic chironomids as food for
110 fishes. This area is inhabited by a species-poor fish assemblages dominated by bleak,
111 *Alburnus alburnus* (L.), common bream, *Abramis brama* (L.), razor fish, *Pelecus cultratus*
112 (L.) and introduced hybrid Asian carp, *Hypophthalmichthys molitrix* (Valenciennes) × *H.*
113 *nobilis* (Richardson). The littoral zone is more heterogeneous and inhabited by a diverse fish
114 assemblage including the common carp (Specziár et al., 2013). Angling for common carp is

115 allowed throughout the year and individuals above 300 mm standard length (*SL*) may be kept
116 by anglers up to maximum of three individuals per day.

117

118 **Tagging and release of fish**

119 In 2010, 4500 two-summer old (1+-2+ age-groups, depending on the season of stocking),
120 fully scaled, less domesticated (c.f. Klefoth et al., 2013) common carp were tagged with
121 Floy[®] FD-68BC T-Bar Anchor Tags (2 × 38 mm; www.floytag.com) of orange colour and
122 marked with unique tag numbers as well as the name and address of the Balaton Limnological
123 Institute. All experimental fish were hatched and reared in the fish farm of the Balaton Fish
124 Management Non-Profit Ltd. and belonged to the same strain. Tagging was performed in the
125 fish farm, near the pond from which experimental fish were obtained by seine netting. Only
126 fish in good condition and with no visible signs of disease or injury were used. Fish were
127 anesthetized in groups in a 0.1 g l⁻¹ solution of clove oil prior to tagging. Each fish was
128 measured for *SL* and body mass (*M*) to the nearest 1 mm and 1 g, respectively. Tagging was
129 made with the maximum accuracy in order to achieve the best possible tag retention. All tags
130 were inserted by the same long-experienced person (B. Turcsányi, with 15 years of tagging
131 experience) between the pterygiophores of the dorsal fin with a Floy[®] tagging gun
132 (www.floytag.com). As such, the T-bar leaned against two neighbouring pterygiophores and
133 from the other side the tag body hung out of the fish. In each individual, the inserted tag was
134 pulled slightly to ensure that it was fixed properly, and thereby minimize initial tag loss
135 (Booth and Weyl, 2008). Tagged fish were transported from the fish farm to the stocking site
136 or the nearest harbour (for offshore stocking) by tanker truck in oxygenated water, at a
137 biomass density of <160 kg m⁻³ (each experimental group of 250 individuals was transported
138 in a separate tank filled with 1 m³ culture pond water).

139 Tagged fish were released seasonally (i.e. in spring, summer and late autumn), at three
140 lake areas (Keszthely, Szemes and Siófok) and both from the shore and offshore, at standard
141 locations (Table 1, Fig. 1). Accordingly, common carp were released in groups of 250 tagged
142 individuals representing 18 stocking trials (Table 1). Upon arrival to the lake, fish were
143 acclimated to ambient water temperature. At shore sites, fish were released to the water
144 directly from the tanker truck by a flexible tube. For offshore stocking, tagged fish were taken
145 by a boat equipped with tanks suitable for safely carrying fish, and then fish were released to
146 the water 2 km offshore by buckets. Before being released, fish in each tank were checked
147 again for viability, and the bottom of the tank was searched for lost tags. No post-handling tag
148 loss and injury were observed among the tagged fish.

149

150 **Recapture of fish and data processing**

151 Tagged fish were recaptured and reported by the anglers. The experiment, its goal and a guide
152 of how the tagged fish should be processed and reported were advertised in written and
153 electronic media and an instruction was also enclosed to each angling licence. However, no
154 information was provided to the anglers about the study design, including date and size at
155 which fish were released. Anglers were instructed to report (either by mail, email or phone)
156 tagged fish irrespective of size, but after measurement undersized fish (i.e. <300 mm *SL*) were
157 to be released back into the lake. Anglers were asked to provide information about the date
158 and location (i.e. nearest settlement and street, estimated distance from the closest
159 unambiguous geographical point or GPS coordinate) of the catch and the size (either *SL* or *M*,
160 preferably both) of the fish at capture. Anglers were distinctly instructed to indicate also if
161 they were not able to provide precise data and were rewarded identically. Ambiguous data
162 were handled as missing information. In order to certify the recapture, anglers were asked to
163 cut the tag and send it (by mail or personally) to the Institute. Consequently, multiple

164 recaptures could not be monitored. It has been shown that an adequate reward significantly
165 increases reporting rate (Denson et al., 2002). Therefore, anglers reporting tagged fish
166 received a bonus (3000 HUF \approx 10 EUR, which is about one half of the average daily net wage
167 in Hungary) to be added to their next licence.

168 MapSource version 6.16.3. software (Garmin Ltd.; www.garmin.com) and the NaviGuide
169 Hungary version 6.51 map layer (Navi-Gate Ltd.; www.garmin.hu) were used to process
170 distribution data of recaptured fish. Distance (i.e. shortest route in the water) between the
171 release and recapture locations was assessed for each fish. The distribution of fish by release
172 strategy was characterised with the length of shore line (i.e. potential angling area), measured
173 separately for the northern and southern shore lines, covered by the first 50, 75 and 90% of
174 recaptures according to their distance from the release point.

175

176 **Statistical analysis**

177 The present study relies on assumptions that reporting rate (i.e. number of reports sent per
178 number of actually captured tagged fish) was similar for all size-groups, capture season and
179 area and that tag retention rate did not vary between releasing strategies. Note that, under the
180 applied experimental setup, testing the study hypotheses does not require any information on
181 the absolute value of the tag retention and reporting rates and on the level and distribution
182 (either in time and space) of the fishing effort (for more details see discussion).

183 Reported recapture rate (thereafter 'recapture rate') and the distribution of recaptures in
184 time and space were tested by using analysis of variance (ANOVA) to the second degree of
185 factor interactions except hypothesis (iv) where only main effects could be analysed. For
186 statistical analysis, fish were grouped into two size-groups (≤ 500 g and > 500 g) based on their
187 M at release and supported by the preliminary experiments (Specziár, 2010). Accordingly, 36
188 releasing set-ups were differentiated on the basis of four explanatory factors: season (i.e.

189 spring, summer and late autumn), area (i.e. Keszthely, Szemes and Siófok basins), method
190 (i.e. shore and offshore) and fish size (i.e. ≤ 500 g and > 500 g *M*) of release. To test study
191 hypotheses, separate ANOVAs were performed evaluating effects of the four explanatory
192 factors (i) on the recapture rate, (ii) number of days fish spent in lake, (iii) distance between
193 the release and recapture sites, and (iv) shore line length covered by the first 50, 75 and 90%
194 of recaptures, based on their distance from the release site. Percent recapture data were arcsin
195 square-root transformed, whereas the other response variables were $\log_{10}(x+1)$ transformed
196 prior to analysis. In case of significant factor effect ($P < 0.05$), ANOVA was completed with
197 Tukey HDS post hoc test. In order to ensure comparability across releasing strategies, only
198 data of fish recaptured within two years (730 days) after release were considered in the above
199 analyses.

200

201 **Results**

202 **Reported recapture rate**

203 Altogether 787 recaptures (17.5% of the fish released) were reported within 730 days after
204 release (Table 1). Recapture rate varied between stocking seasons, but not between lake areas,
205 methods and fish size-groups and all combinations of second degree factor interaction were
206 also unimportant (Table 2). Recapture rates were unequivocally highest in summer
207 ($25.9\% \pm 6.6\%$; mean \pm SD; Tukey HDS post hoc test, $P < 0.05$), followed by spring
208 ($17.0\% \pm 3.8\%$) and least in late autumn stockings ($10.4\% \pm 3.3\%$).

209

210 **Time in lake**

211 Recaptures started with four months delay in the late autumn, but immediately in spring (two
212 days after release) and summer stockings (on the day of the release) (Fig. 2). Catches showed
213 a marked seasonality; very few common carp were caught at < 8 °C water temperature (i.e.

214 from November to March) and most recaptures were reported from June to September at 16-
215 25 °C (Fig. 2). Depletion of recaptures was rapid; 88.6%±5.9% of the 730 days recaptures
216 occurred within the first year after release of fish (i.e. within 365 days after release) and only
217 11.4%±5.9% in the second (Student's t-test, $t_{70}=55.9$, $P<0.001$).

218 The number of days from release to recapture was evidently influenced by the stocking
219 season (Table 2) and common carp released in late autumn spent much longer periods
220 (271±119 days) in lake than those released in spring (152±106 days) and especially in
221 summer (123±146 days; Tukey HDS, $P<0.05$).

222

223 **Distribution of fish**

224 The majority of recaptures were reported from the littoral zone and only 19 common carp
225 were caught >300 m offshore and 15 in southern inflows. Individual fish showed remarkable
226 differences in their movements; some specimens travelled >60 km within 60-150 days, but
227 others were recaptured <3 km distance from the release point even after two years in lake
228 (Fig. 3). Therewith, the mean distance of recaptures from the release points varied
229 significantly between stocking seasons and areas and was affected by the interaction of season
230 and area and season and method of stocking, but not by the method of release and the size of
231 fish at release (Table 2). Namely, common carp released in autumn were recaptured at highest
232 mean distance (23.0±17.1 km) from the stocking point, followed by spring (18.5±15.3 km)
233 and summer-released fish (10.5±11.5 km; Tukey HDS, $P<0.05$). Individuals stocked at
234 Szemes moved a longer mean distance (18.2±11.9 km) than those released at Siófok
235 (13.9±13.4 km) and Keszthely (14.3±18.0 km; Tukey HDS, $P<0.05$). Further, in the summer
236 stockings, fish released offshore were captured at a slightly longer mean distance (11.7±13.3
237 km) from the stocking site than those released from the shore (9.4±9.7; Tukey HDS, $P<0.05$).

238 Lake area covered by 50, 75 and 90% of recaptures also varied significantly between
239 releasing strategies, and it was influenced by the season and area of release. Recaptures
240 dispersed in space more in autumn and spring and Szemes stockings than in summer and
241 Siófok and Keszthely stockings (Table 3, Fig. 4).

242

243 **Discussion**

244 The present study is the first reporting relative effects of different stocking strategies for large
245 lake put-and-take recreational common carp fishery. In Lake Balaton, total reported recapture
246 rate was 17.5%, which falls within the upper range observed by other tagging studies. Tagged
247 common carp were recaptured at a rate of 16.6% in the Lipno reservoir and 15.0% in the
248 Vlatava River, Czech Republik (Vostradovský, 1991), 8.3% in the Monelho Reservoir, Czech
249 Republik (Baruš et al., 1997) and 8.8% in an Australian river (Stuart and Jones, 2006). The
250 total 787 reports of recaptured fish along with the symmetrical study design provided an
251 excellent opportunity to evaluate the relationship between particular stocking strategies (i.e.
252 season \times lake area \times method \times fish size-group) and recaptures by anglers. It was shown that
253 stocking strategy influenced both recapture rate and the extent of the lake area over which
254 recaptures distributed.

255 Fish released in summer and spring were much more likely to be recaptured than those
256 released in late autumn. This result is in accordance with findings of Vostradovský (1991) in
257 Lipno Reservoir and indicates an overwintering risk on recapture. In winter, Lake Balaton is
258 not a benign habitat, especially not for freshly-stocked fish unfamiliar with the environment.
259 This large, shallow and open lake is heavily exposed to strong winds disturbing the water to
260 the bottom. Pond-reared common carp stocked into the lake in late autumn have little chance
261 to find appropriate wintering sites rapidly and thus have to waste their energy reserves
262 moving against the exposed, turbid water in ice-free periods (c.f. Ruuhijärvi et al., 1996). In

263 addition, fish are generally released along the southern shore line of the lake, which is the
264 shallowest and most wind-exposed part of the lake with very few wintering sites and poor
265 food resources. Most of the wintering sites occur along the northern shore line, which is less
266 exposed to dominant northern winds and has ideal water depth (1.5-3 m) and richer food
267 resources. However, those sites are located far from the culture ponds established on the
268 southern inflows of Lake Balaton, and therefore it would be much more costly to transport
269 and release fishes there. On the contrary, common carp stocked in spring and especially in
270 summer can immediately feed, and thus, are less exposed to starvation and related mortality.

271 Although the offshore area of Lake Balaton is not a preferred habitat of common carp
272 (Specziár et al., 2013), individuals released there were recaptured at the same rate as those
273 released from the shore. This observation along with the finding that common carp were
274 captured only sporadically offshore indicated that stocked fish moved rapidly to the littoral
275 zone. Dispersion of stocked fish is very important for consistent angling facility. Since most
276 anglers knowing habitat preference of common carp fish it in the littoral zone, and thus,
277 fisheries managers assumed that offshore stocking could provide more time to fish dispersing
278 before reaching the preferred angling areas. Therefore, a part of the annual stocking quota
279 now is released offshore. However, individuals released offshore did not disperse over larger
280 areas than those released from the shore, except in summer when recaptures were most rapid.
281 Accordingly, instead of the costly offshore stockings, it would be more beneficial to distribute
282 stocking quotas among more shore sites. Prevalence of the latter strategy is also supported by
283 that there was no difference in the recapture rate between stocking areas.

284 Since body size is a crucial factor in most ecological processes influencing mortality of
285 fish (Houde, 1997; Schultz and Conover, 1999), we supposed that fish size will have a
286 positive effect on the recapture rate. However, results did not prove a definite size-specific
287 pattern within the studied size range. Present results also contradict our preliminary

288 observations based on series of individual stocking trials (with limited comparability) with
289 much more variable size-groups of common carp (Specziár, 2010), which indicated a strongly
290 decreasing recapture rate below 500 g *M* stocking size and with practically no recaptures at
291 <150 g *M*. Similarly, size of fish at release did not affect the time of recapture or the
292 distribution of individuals.

293 Recaptures depleted at a rate of 89% year⁻¹, which is concordant with the results of all
294 preliminary common carp tagging experiments in Lake Balaton (Specziár, 2010). Similarly,
295 rapid depletion of recaptures was observed in other preferred game fishes, such as for instance
296 in the white bass, *Morone chrysops* (Rafinesque), in the Brazos River and Whitney Reservoir,
297 Texas USA (Muoneke, 1992) and in the red drum, *Sciaenops ocellatus* (L.), in South Carolina
298 and Georgia estuaries, USA (Denson et al., 2002). Although tag loss (discussed below) and
299 natural mortality might also contribute to this pattern, in our opinion this primarily indicated
300 the extremely high angling pressure on common carp in Lake Balaton.

301 The mean distance between the release and recapture locations and the extent of the lake
302 area over which recaptures distributed depended clearly on the mean time that stocked
303 common carp spent in lake. Namely, fish moved longer distances and dispersed over larger
304 areas in cold water stockings (i.e. in late autumn and spring) than in summer stocking when
305 mass recaptures started immediately. Moreover, common carp released in the middle of the
306 lake (i.e. at Szemes) dispersed over larger areas than those released at either end (i.e. at
307 Keszthely or Siófok) of the elongated-shaped lake. In the present study, common carp
308 dispersed over larger distances than in most other experiments. For example, Stuart and Jones
309 (2006) found that 66% of the recaptured common carp remained within 1 km of their release
310 location and only 20% of fish were recaptured at >5 km distance over 1107 days of
311 monitoring in the Murray River, Australia. However, in that study, experimental fish were
312 caught from the river near the release site, and therefore those wild individuals were familiar

313 with their environment and might have defined home ranges to which they held to after
314 release as well. Site fidelity is characteristic for common carp (Jones and Stuart, 2007) and its
315 strength was also proved by the homing behaviour of displaced wild-caught individuals
316 (Schwartz, 1987). On the other hand, Vostradovská and Vostradovský (1980) found that
317 pond-reared common carp stocked into the River Vltava also stayed in <5 km vicinity of the
318 release site and few specimens migrated up to 20 km either upstream or downstream.
319 Accordingly, further studies are needed to explore factors influencing the dispersion of
320 stocked and wild common carp.

321 Unavailability of detailed data on the rate and pattern of tag loss, anglers' reporting level
322 and fishing effort generally limits evaluation of the processes underlying the observed
323 recapture patterns in larger waterbodies (Denson et al., 2002; Booth and Weyl, 2008).
324 However, in the present study, thanks to the symmetrical study design and problems tested, it
325 was not essential to know the reporting level and fishing effort. Note, that this study did not
326 ask any questions about stock size, mortality rate, vulnerability of fish to angling and total
327 angling catch per unit effort or yield; simply, it was asked which stocking strategy yields most
328 recaptures from the largest area and time interval under the given and not exactly known
329 fishing activity. The only required assumption in this respect was that the reporting level did
330 not change systematically among stocking strategies tested, a criterion that was very likely to
331 be met considering the short time-frame and the design of the experiment. Nevertheless, it
332 remains a weakness of this study that individual effects of biotic (i.e. mortality and
333 behavioural variations of fish) and human factors (i.e. variations of fishing effort) could not
334 be identified just their resultant. On the other hand, tag loss could influence the result at least
335 in one respect. Namely, fish released in late autumn spent significantly longer periods in lake
336 and thus were more exposed to eventual tag loss than those released in spring and especially
337 in summer. This bias might cause an underestimation of recapture rate for fish released in late

338 autumn. However, recapture rates differed so much between the late autumn and spring and
339 summer stockings (1.7 and 2.5 times, respectively) that the observed deviation was very
340 unlikely to be an artefact caused only by the tag loss. In addition, several studies proved that
341 with well-experienced staff and proper insertion of tag and handling of experimental fish,
342 very high tag retention up to 89-96% year⁻¹ can be achieved with T-bar anchor type tags
343 (Ebener and Copes, 1982; Buzby and Deegan, 1999; Livings et al., 2007). High tag retention
344 rate was also supported by lifespan recaptures of asp, *Aspius aspius* (L.) and tench, *Tinca*
345 *tinca* (L.) in Lake Balaton (Specziár, 2010). However, a little bit lower, 71%, tag retention
346 rate was assessed for wild captured common carp in an Australian river (Stuart and Jones, 2006).

347 To conclude, the present study provided valuable information for improvement of
348 recreational fisheries management in Lake Balaton, especially how to disperse annual
349 stocking quotas across seasons and lake areas to achieve highest rate and distribution of
350 recaptures. It is suggested that the more costly offshore stocking might be beneficial only in
351 summer, when it may facilitate initial dispersion of common carp before recaptures. It is also
352 suggested that in summer fish should be released at several sites distributed evenly along both
353 the southern and the northern shore lines of the lake. On the contrary, cold water spring
354 stockings may be optimized for transport cost and made at fewer easily accessible sites in
355 each lake basin along the southern shore line close to rearing ponds. However, late autumn
356 stockings should be avoided, and the capacity of effective wintering ponds is suggested to be
357 developed. The present study also provides a good framework for testing fisheries
358 management alternatives in other intensively fished habitats.

359

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366

367 **References**

368 Arlinghaus, R.; Mehner, T., 2003: Socio-economic characterisation of specialized common
369 carp (*Cyprinus carpio* L.) anglers in Germany, and implications for inland fisheries
370 management and eutrophication control. *Fish. Res.* **61**, 19-33.

371 Balfry, S.; Welch, D. W.; Atkinson, J.; Lill, A.; Vincent, S., 2011: The effect of hatchery
372 release strategy on marine migratory behaviour and apparent survival of Seymour River
373 steelhead smolts (*Oncorhynchus mykiss*). *PLoS ONE* **6**, e14779.

374 Baruš, V.; Prokeš, M.; Peňáz, M., 1997: Dispersion and density assessment of the common
375 carp (*Cyprinus carpio* m. *domestica*) in the Mohelno reservoir Czech Republic. *Folia*
376 *Zool.* **46**, 315-324.

377 Booth, A. J.; Weyl, O. L. F., 2008: Retention of T-bar anchor and dart tags by a wild
378 population of African sharptooth catfish, *Clarias gariepinus*. *Fish. Res.* **92**, 333-339.

379 Buzby, K.; Deegan, L., 1999: Retention of anchor and passive transponder tags by arctic
380 grayling. *N. Am. J. Fish. Manage.* **19**, 1147-1150.

381 Denson, M. R.; Jenkins, W. E.; Woodward, A. G.; Smith, T. I. J., 2002: Tag-reporting levels
382 for red drum (*Sciaenops ocellatus*) caught by anglers in South Carolina and Georgia
383 estuaries. *Fish. Bull.* **100**, 35-41.

384 Ebener, M. P.; Copes, F. A., 1982: Loss of Floy anchor tags from lake whitefish. *N. Am. J.*
385 *Fish. Manage.* **2**, 90-93.

386 Hickley, P., 2009: Recreational fisheries – social, economic and management aspects. In:
387 Fisheries, sustainability and development. Eds: H. Ackefors; M. Cullberg; P. Wramner,
388 Stockholm: Royal Swedish Academy of Agriculture and forestry, pp. 137-157.

389 Houde, E. D., 1997: Patterns and consequences of selective processes in teleost early life
390 histories. In: Early life history and recruitment of fish populations. Eds: R.C. Chambers;
391 E. A. Trippel, London: Chapman and Hall, pp. 173-196.

392 Istvánovics, V.; Clement, A.; Somlyódy, L.; Specziár, A.; G.-Tóth, L; Padisák, J., 2007:
393 Updating water quality targets for shallow Lake Balaton (Hungary), recovering from
394 eutrophication. *Hydrobiologia* **581**, 305-318.

395 Jones, M. J; Stuart, I. G., 2007: Movements and habitat use of common carp (*Cyprinus*
396 *carpio*) and Murray cod (*Maccullochella peelii peelii*) juveniles in a large lowland
397 Australian river. *Ecol. Freshw. Fish* **16**, 210-220.

398 Kennedy, G. J. A.; Strange, C. D., O'Neill, G. O., 1982: Tagging studies on various age
399 classes of brown trout (*Salmo trutta* L.). *Fish. Mgmt.* **13**, 33-41.

400 Klefoth, T.; Pieterek, T.; Arlinghaus, R., 2013: Impacts of domestication on angling
401 vulnerability of common carp, *Cyprinus carpio*: the role of learning, foraging behaviour
402 and food preference. *Fish. Manage. Ecol.* **20**, 174-186.

403 Kottelat, M., 1996: *Cyprinus carpio* (River Danube subpopulation). In: IUCN 2012, IUCN
404 Red List of Threatened Species, Version 2012.2. www.iucnredlist.org (Downloaded on 14
405 March 2013)

406 Linfield, R. S. J., 1980: Catchability and stock density of common carp, *Cyprinus carpio* L. in
407 a lake fishery. *Fish. Mgmt.* **11**, 11-22.

408 Livings, M. E.; Schoenebeck, C. W.; Brown, M. L., 2007: Long-term anchor tag retention in
409 yellow perch, *Perca flavescens* (Mitchill). *Fish. Manage. Ecol.* **14**, 365-366.

410 Michaletz, P. H.; Wallendorf, M. J.; Nicks, D. M., 2008: Effects of stocking rate, stocking
411 size and angler catch inequality on exploitation of stocked channel catfish in small
412 Missouri impoundments. *N. Am. J. Fish. Manage.* **28**, 1486-1497.

413 Muoneke, M. I., 1992: Loss of Floy anchor tags from white bass. *N. Am. J. Fish. Manage.* **12**,
414 819-824.

415 Patterson, W. F.; Sullivan, M. G., 2013: Testing and refining the assumptions of put-and-take
416 rainbow trout fisheries in Alberta. *Hum. Dimens. Wildl.* **18**, 340-354.

417 Ruuhijärvi, J.; Salminen, M.; Nurmio, T., 1996: Releases of pikeperch (*Stizostedion*
418 *lucioperca* (L.)) fingerlings in lakes with no established pikeperch stock. *Ann. Zool. Fenn.*
419 **33**, 553-567.

420 Schultz, E. T.; Conover, D. O., 1999: The allometry of energy reserve depletion: test of a
421 mechanism for size-dependent winter mortality. *Oecologia* **119**, 474-483.

422 Schwartz, F., 1987: Homing behavior of tagged and displaced carp, *Cyprinus carpio*, in
423 Pymatuning Lake, Pennsylvania/Ohio. *Ohio J. Sci.* **87**, 15-22.

424 Specziár, A., 2010: Fish fauna of Lake Balaton: stock composition, living conditions of fish
425 and directives of the modern utilization of the fish stock. *Acta Biol. Debr. Suppl. Oecol.*
426 *Hung.* **23** (Hydrobiol. Monogr. vol. **2**): 7-185. (In Hungarian with an English summary)

427 Specziár, A.; György, Á. I.; Erős, T., 2013: Within-lake distribution patterns of fish
428 assemblages: the relative roles of spatial, temporal and random environmental factors in
429 assessing fish assemblages using gillnets in a large and shallow temperate lake. *J. Fish*
430 *Biol.* **82**, 840-855.

431 Stuart, I. G.; Jones, M. J., 2006: Movement of common carp, *Cyprinus carpio*, in a regulated
432 lowland Australian river: implications for management. *Fish. Manage. Ecol.* **13**, 213-219.

- 433 Vostradovská, M.; Vostradovský, J., 1980: Fish tagging in the Vltava River in Prague.
434 *Živočišná Výroba* **25**, 863-870. (In Czech with summaries in Russian, English and
435 German)
- 436 Vostradovský, J., 1991: Carp (*Cyprinus carpio* L.) “put-and-take” fisheries in the
437 management of angling waters in Czechoslovakia. In: Catch effort sampling strategies.
438 Their application in freshwater fisheries management. Ed.: I. G. Cowx, Oxford: Fishing
439 New Books, Blackwell Scientific Publications, pp. 100-107.
- 440 Wolos, A.; Theodorowicz, M.; Brylski, H., 1998: Socio-economic analysis of recreational
441 fisheries in two departments of the Polish Anglers Association, based on the results of the
442 registration of anglers’ catches. In: Recreational fisheries: Social, economic and
443 management aspects. Eds: P. Hickley; H. Tompkins, Oxford: Fishing News Books,
444 Blackwell Scientific Publications, pp. 36-47.
- 445 Young, G. C.; Wise, B. S.; Ayvazian, S. G., 1999: A tagging study on tailor (*Pomatomus*
446 *saltatrix*) in Western Australian waters: their movement, exploitation, growth and
447 mortality. *Mar. Freshw. Res.* **50**, 633-642.

448 Table 1

449 Specifications of stocking trials including season, area and method of release, water temperature (T), number (N_r), standard length (SL) and body
 450 mass (M) of fish released and total number of reported recaptures within 730 days by anglers (N) in Lake Balaton, Hungary.

Release										Recapture	
No.	Date	Area	Method	T ($^{\circ}\text{C}$)	N_r	SL (mm)		M (g)		$N_{730\text{days}}$	%
						Mean \pm SD	Range	Mean \pm SD	Range		
1.	30.03.2010	Keszthely	shore	8.2	250	254 \pm 19	200-300	457 \pm 85	243-710	38	15.2
2.	30.03.2010	Keszthely	offshore	8.2	250	255 \pm 18	190-300	434 \pm 81	225-625	42	16.8
3.	30.03.2010	Szemes	shore	8.2	250	259 \pm 20	200-310	485 \pm 92	274-781	43	17.2
4.	31.03.2010	Szemes	offshore	8.2	250	254 \pm 20	200-310	492 \pm 99	236-831	43	17.2
5.	31.03.2010	Siófok	shore	8.2	250	250 \pm 18	200-285	478 \pm 99	234-769	30	12.0
6.	31.03.2010	Siófok	offshore	8.2	250	255 \pm 21	190-310	506 \pm 117	240-968	54	21.6
7.	13.07.2010	Keszthely	shore	23.6	250	265 \pm 21	220-330	554 \pm 122	300-897	67	26.8
8.	13.07.2010	Keszthely	offshore	23.6	250	256 \pm 27	190-340	463 \pm 151	187-1025	61	24.4
9.	13.07.2010	Szemes	shore	23.6	250	268 \pm 21	220-320	565 \pm 127	300-960	65	26.0
10.	12.07.2010	Szemes	offshore	23.3	250	269 \pm 21	200-330	586 \pm 123	193-890	43	17.2
11.	12.07.2010	Siófok	shore	23.3	250	272 \pm 20	230-330	601 \pm 117	350-922	80	32.0
12.	12.07.2010	Siófok	offshore	23.3	250	274 \pm 18	230-330	616 \pm 124	385-1015	74	29.6
13.	23.11.2010	Keszthely	shore	3.4	250	233 \pm 25	170-300	367 \pm 117	174-764	23	9.2
14.	23.11.2010	Keszthely	offshore	3.4	250	234 \pm 28	170-320	375 \pm 133	158-825	31	12.4
15.	23.11.2010	Szemes	shore	3.4	250	245 \pm 27	200-350	421 \pm 138	222-1110	31	12.4
16.	07.12.2010	Szemes	offshore	3.3	250	228 \pm 27	180-330	355 \pm 138	172-1435	22	8.8
17.	07.12.2010	Siófok	shore	3.3	250	225 \pm 26	180-310	335 \pm 110	160-698	27	10.8
18.	07.12.2010	Siófok	offshore	3.3	250	234 \pm 30	180-340	378 \pm 149	179-1020	13	5.2
Total					4500					787	17.5

451

452 Table 2

453 Results of the ANOVA on the effect of stocking season, lake area, method (i.e. shore and
 454 offshore stockings) and fish size on arcsin square root transformed 730 days recapture rate by
 455 anglers (%), and $\log_{10}(x+1)$ transformed days spent in lake and distance travelled between
 456 releasing and recapturing sites of tagged common carp in Lake Balaton.

	Recapture rate			Days in lake			Distance travelled		
	d.f.	F	P	d.f.	F	P	d.f.	F	P
Season	2	36.1	<0.001	2	70.7	<0.001	2	41.3	<0.001
Area	2	1.2	0.327	2	1.6	0.209	2	19.5	<0.001
Method	1	0.1	0.821	1	0.1	0.807	1	0.0	0.901
Fish size	1	2.9	0.109	1	0.7	0.397	1	1.4	0.242
Season × area	4	1.5	0.257	4	2.9	0.021	4	3.6	0.007
Season × method	2	2.2	0.146	2	0.7	0.508	2	5.4	0.005
Area × method	2	1.2	0.341	2	2.4	0.092	2	0.4	0.649
Season × fish size	2	0.1	0.894	2	0.2	0.787	2	0.2	0.828
Area × fish size	2	1.5	0.248	2	1.5	0.220	2	2.0	0.135
Method × fish size	1	0.6	0.437	1	0.0	0.979	1	2.5	0.115
Error	16			767			760		

457

458 Table 3

459 Results of the ANOVA on the effect of stocking season, lake area and method (i.e. shore and
 460 offshore stockings) on $\log_{10}(x+1)$ transformed shore line length data covered by the first 50,
 461 75 and 90% of common carp recaptures according to their distance from the release point in
 462 Lake Balaton, Hungary. Results of the Tukey HDS post hoc tests are reported for significant
 463 single factor effects (at $P < 0.05$) in total shore line. Note that the effect of fish size could not
 464 be tested due to limited sample sizes, but fish ≤ 500 and > 500 g body mass were recaptured at
 465 equal mean distances from the release point (Table 2).

	d.f.	Southern shore line		Northern shore line		Total shore line		Tukey HDS ($P < 0.05$)
		F	P	F	P	F	P	
50% of recaptures								
Season	2	3.6	0.060	8.5	0.005	8.6	0.005	Autumn, spring > summer
Area	2	3.9	0.050	5.2	0.023	6.4	0.013	Szemes > Keszthely
Method	1	0.8	0.391	0.0	0.978	0.3	0.592	
Error	16							
75% of recaptures								
Season	2	9.2	0.004	12.8	0.001	13.0	<0.001	Autumn, spring > summer
Area	2	17.1	<0.001	9.5	0.003	14.6	<0.001	Szemes > Siófok, Keszthely
Method	1	1.7	0.210	3.9	0.072	3.4	0.090	
Error	16							
90% of recaptures								
Season	2	2.4	0.131	25.0	<0.001	11.1	0.001	Autumn, spring > summer
Area	2	9.0	0.004	14.9	<0.001	13.9	<0.001	Szemes > Siófok, Keszthely
Method	1	4.1	0.067	0.0	0.855	2.1	0.175	
Error	16							

466 **Figure captions**

467

468 Fig. 1. Map of Lake Balaton (Hungary) and the major tributaries, with indication of shore
469 (white circles) and offshore (grey circles) stocking sites.

470

471 Fig. 2. Seasonal recapture dynamics of common carp stocked in spring (a), summer (b) and
472 autumn (c) in relation to water temperature (d) in Lake Balaton, Hungary. In each stocking
473 season 1500 tagged fish were released (Table 1). Stocking events are indicated by arrows.

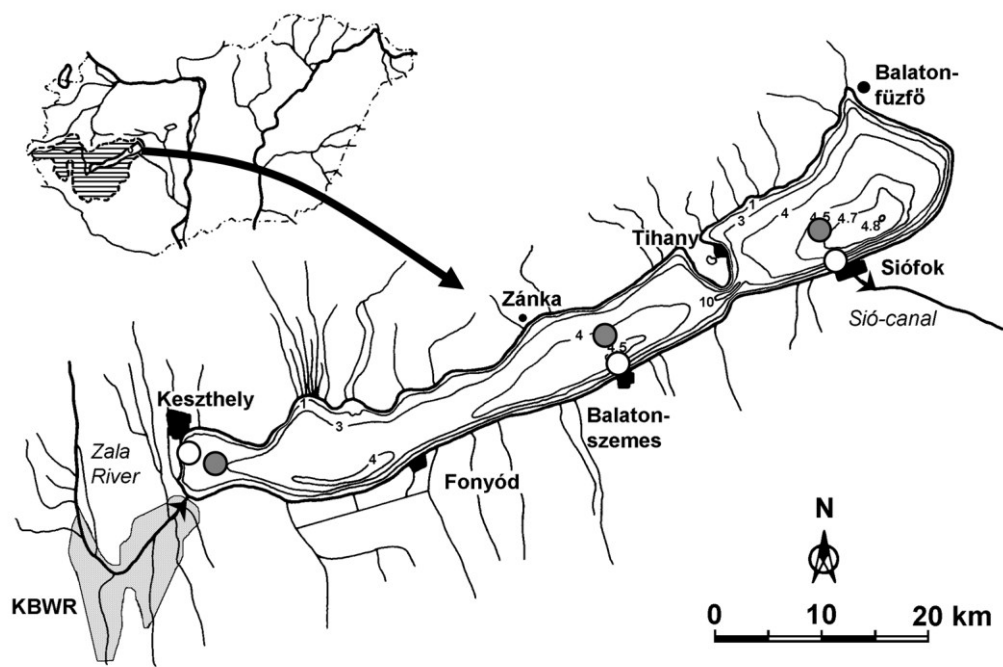
474

475 Fig. 3. Distance between the release (i.e. Keszthely, Szemes, Siófok) and recapture locations
476 of stocked common carp in relation to days spent in Lake Balaton, Hungary. Note that Szemes
477 is located approximately at the middle of the longitudinal axis of the lake and thus fish
478 released here can move away maximum 40 km.

479

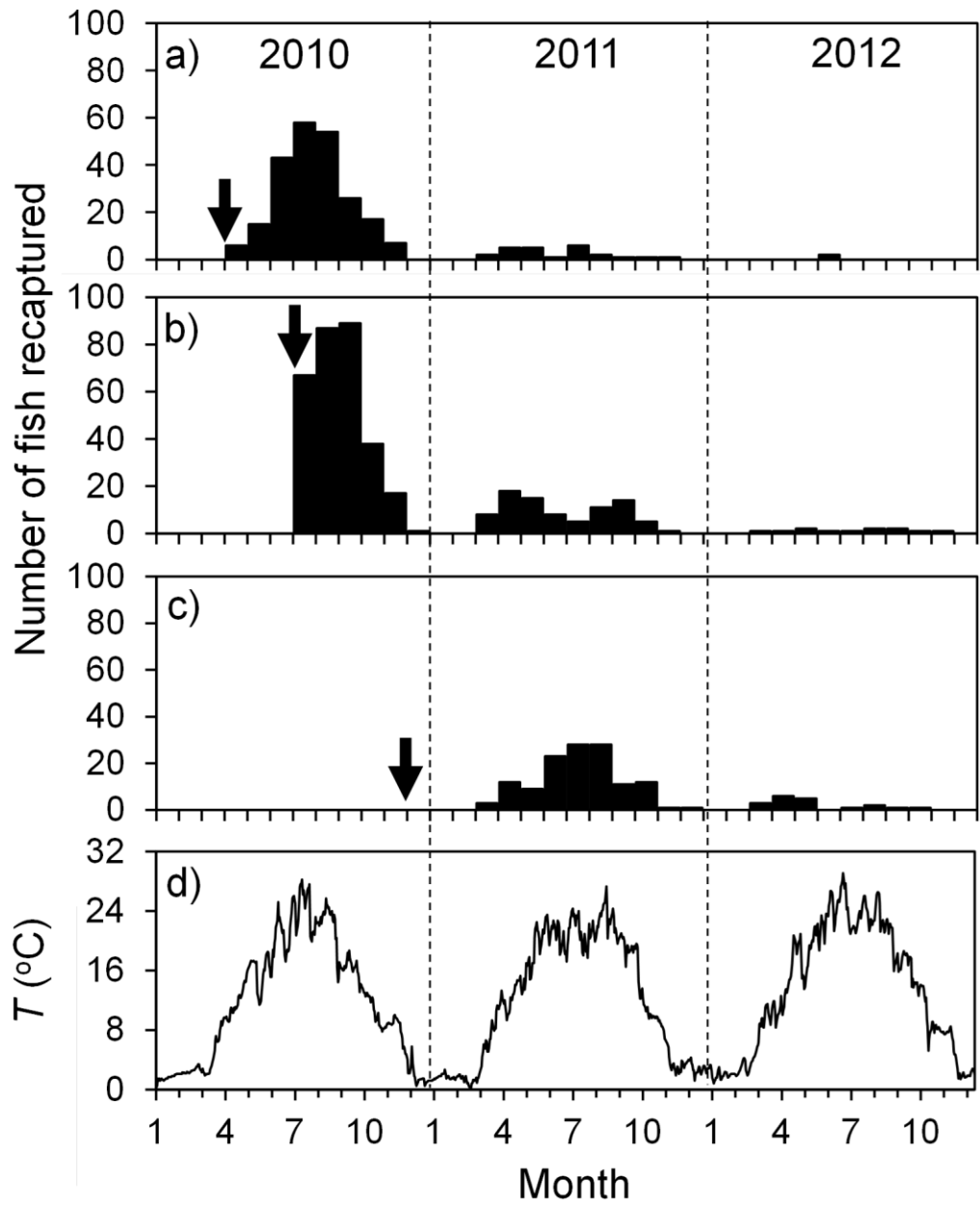
480 Fig. 4. Length (km) of northern and southern shore lines (i.e. fishing area) covered by the first
481 50, 75 and 90% of recaptures according to their distances from the release site in Lake
482 Balaton, Hungary. Results indicated longer dispersion distances in spring and autumn than in
483 summer and for fish released at Szemes (Sz) than at Keszthely (K) and Siófok (S) (for
484 statistics see Table 3).

485



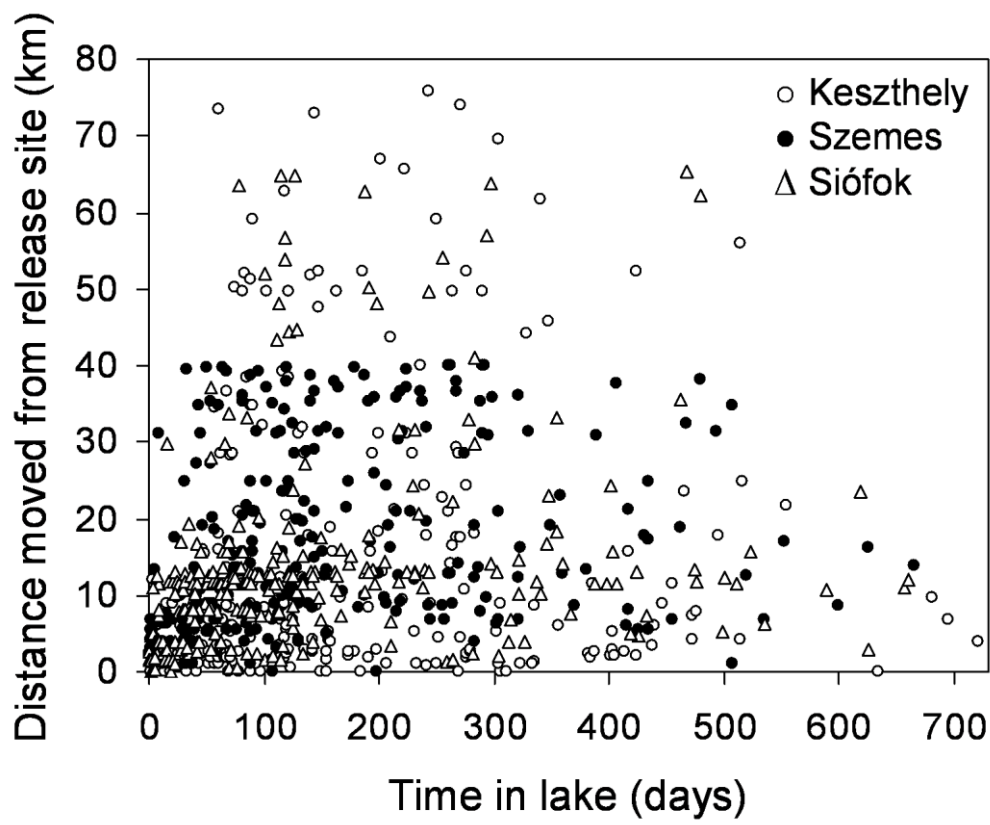
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