

1 **Endemism increases species' risk to climate change in areas of global biodiversity**
2 **importance**

3

4

5 **Abstract**

6 Climate change affects life at global scales and across systems but is of special concern in
7 areas that are disproportionately rich in biological diversity and uniqueness. Using a meta-
8 analytical approach, we analysed >8,000 risk projections of the projected impact of climate
9 change on 273 areas of exceptional biodiversity, including terrestrial and marine
10 environments. We found that climate change is projected to negatively impact all assessed
11 areas, but endemic species are consistently more adversely impacted. Terrestrial endemics are
12 projected to be 2.7 and 10 times more impacted than non-endemic natives and introduced
13 species respectively, the latter being overall unaffected by climate change. We defined a high
14 risk of extinction as a loss of >80% due to climate change alone. Of endemic species, 34%
15 and 46% in terrestrial and marine ecosystems, and 100% and 84% of island and mountain
16 species were projected to face high extinction risk respectively. A doubling of warming is
17 projected to disproportionately increase extinction risks for endemic and non-endemic native
18 species. Thus, reducing extinction risks requires both adaptation responses in biodiversity
19 rich-spots and enhanced climate change mitigation.

20

21 **Keywords:** Extinction Risk, Biodiversity Hotspots, Global-200 Ecoregions, Introduced
22 Species

23 1. Introduction

24

25 Climate change is already impacting biodiversity and is likely to intensify over the next few
26 decades unless substantive mitigation efforts are implemented (IPCC 2018). Both modelling
27 and field observations suggest non-uniform extinction risks of wild species across geographic
28 regions and between taxa, even at low levels of warming (e.g. Urban 2015; Román-Palacios
29 & Wiens, 2020). This spatial variation in impacts shapes global biodiversity responses to
30 climate change. Despite the publication of many hundreds of studies on projected impacts of
31 climate change on species and ecological communities, it remains challenging to synthesize
32 clear patterns of risk across different levels of ecological organization (e.g. species and
33 community levels), between ecological realms (terrestrial, freshwater and marine), as a
34 function of ecological uniqueness (i.e. level of endemism), and as a function of policy-
35 relevant climate scenarios (low to high projected rates of climate change). Analysis to tease
36 out the importance of such factors would be valuable in informing our understanding of
37 climate risks to biodiversity, and in prioritising and developing adaptive responses.

38 Previous work suggests a range of expectations relevant to the factors mentioned
39 above. With respect to projected vulnerabilities across ecological realms, global level
40 assessments are rare. Marine communities are expected to show greater sensitivity to climate
41 change than terrestrial communities because the distribution of marine species is more
42 strongly governed by their thermal tolerances (Sunday et al. 2012) and thermal safety
43 margins are lower (Pinsky et al. 2019). As isotherms shift most strongly in marine equatorial
44 regions (Burrows et al. 2011) the combination of vulnerability and exposure predicts the
45 largest impacts there. In addition, there is a positive correlation between climatic and non-
46 climatic stressors in marine environments, whereas on land regions of strong climate change
47 tend to be those with low non-climatic impacts (Bowler et al. 2020). On land, subtropical to
48 temperate flatlands are projected to have the greatest climate velocities (Loarie et al. 2009,
49 Burrows et al. 2011), and are thus expected to show the greatest projected impacts.

50 Geographic range shifts, expansions and contractions are among the most common
51 responses of species to climate change (Poloczanska et al. 2013; Molinos et al. 2016; Saeedi
52 et al. 2017; Chaudhary et al. 2020; Yasuhara et al. 2020). Species with large geographic
53 ranges are expected to be less vulnerable, as they may find refugia in parts of their range
54 (Lucas et al. 2019). Introduced species that become invasive are expected to be less
55 vulnerable due to their adaptability to new environments (Oduor et al. 2016). In contrast, the
56 more restricted ranges of endemic species means that they are often at greater risk of
57 extinction from local impacts, including habitat loss and interactions with introduced species;
58 the effects of which are being exacerbated by changes in climate (Catford et al. 2012; IPCC
59 2019). Endemics have restricted geographic ranges, sometimes associated with a specialized
60 environmental niche, limited dispersal abilities, and reduced population size and adaptive
61 capacity (Chichorro et al. 2019; Staude et al. 2020). Therefore, areas of high endemism are
62 likely to be particularly vulnerable to climate change at both species- and community-levels
63 (Malcolm et al. 2006; Dirnböck et al. 2011; Enquist et al. 2019).

64 Biodiversity is unevenly distributed across the globe, and areas with exceptional
65 biodiversity are prioritized in conservation efforts (Brooks et al. 2006; Asaad et al. 2018;
66 Zhao et al. 2020). Biodiversity hotspots (Myers et al. 2000) and the Global-200 ecoregions
67 (Olson and Dinerstein 2002) together comprise 273 irreplaceable terrestrial, freshwater and
68 marine areas, with notable endemism, richness and/or unusual ecological or evolutionary
69 phenomena, hereafter called ‘rich-spots’. These areas are expected to experience severe
70 climatic change in the future (Beaumont et al. 2011; Bellard et al. 2014). If exceptional
71 biodiversity is due to long-term climatic stability (Dynesius and Jansson 2014; Senior et al.
72 2018), then endemic species of such areas may be particularly at risk of adverse impacts even
73 under less extreme climate scenarios.

74 The vulnerability of these rich-spots to climate change has been previously
75 investigated using coarse estimations based on modelling species-area relationships (*e.g.*
76 Brooks et al. 2002; Malcolm et al. 2006; Bellard et al. 2014; Habel et al. 2019). For example,
77 Malcolm et al. (2006) assessed the climate change impact on 25 rich-spots by modelling the
78 change in habitat area, and corresponding changes in biodiversity, likely as a result of future
79 biome distributions projected by global vegetation models. Similarly, Bellard et al. (2014)
80 modelled the effect of projected climate change on 34 rich-spots to examine the extent to
81 which they would experience novel climates and the proportion of endemic species affected
82 by this change, as well as the potential expansion of invasive species. However, such
83 previous studies have tended to produce approximations of the number of species that would
84 be adversely affected as climatic niche space is lost. Estimates based solely on area lack the
85 necessary sensitivity of species-specific parameters and do not incorporate the local context
86 of each different rich-spot, possibly biasing vulnerabilities towards larger areas (Brooks et al.
87 2006). A species-specific and community-level examination of vulnerability to climate
88 change would provide more robust evidence from which to estimate risks and on which to
89 base adaptation strategies.

90 We assessed over 8,000 projections of climate change impacts in 232 studies for
91 endemic, non-endemic native and introduced species and communities across terrestrial,
92 freshwater and marine environments, based on papers that account for their identity and local
93 context of different rich-spots. Through this extensive systematic review of the literature, we
94 aimed to test for differences in projected responses between endemic, non-endemic native
95 and introduced species; differences in projected responses of species and communities of
96 terrestrial and marine ecosystems; and how vulnerability is projected to vary among climate
97 zones, geographic regions, and across a representative range of climate change scenarios for
98 this century.

99

100

101 **2. Methods**

102

103 *2.1. Literature Search*

104 We performed an extensive literature search for papers that investigated the impacts of
105 climate change on biodiversity in global priority conservation areas. We considered two
106 conservation schemes: “Biodiversity Hotspots” (Myers et al. 2000, extended by Mittermeier
107 et al. 2004; Mittermeier et al. 2011; Williams et al. 2011; Noss et al. 2015), including 35
108 terrestrial regions; and “Global-200 Ecoregions” (Olson and Dinerstein 2002), including 195
109 terrestrial and freshwater regions and 43 marine regions (Supplementary Figure 2,
110 Supplementary Table 1). The Global-200 (Olson and Dinerstein 2002) are a set of
111 irreplaceable and distinctive ecoregions, which comprise areas of high endemism and/or
112 species richness, and/or unusual ecological or evolutionary phenomena. While biodiversity
113 hotspots represent a substantive fraction of global species richness on less than 16% of the
114 terrestrial surface area, the Global-200 ecoregions extend well beyond this area and are more
115 representative of all environments. The rich-spots included in this study comprise 48% and
116 17% of the world’s terrestrial and marine surfaces, respectively (Supplementary Table 1).
117 There is some overlap of approximately 14% between both conservation schemes on land
118 (Supplementary Figure 2). We searched for papers published since 2012 using “climate
119 change” AND “biodiversity” AND the names of each of the rich-spots. We aimed to
120 understand whether recent trends in biodiversity research have changed since the latest
121 reviews (IPCC 2014a; Urban 2015). We directed the search at peer-reviewed journal articles,
122 but included 10 scientific reports from research institutions where there were data gaps.

123 We found 395 publications that evaluated climate change on some aspect of biodiversity in
124 these rich-spots. From these, we only used 232 papers that established future projections of
125 climate change impacts with quantifiable risks upon biodiversity. According to the IPCC
126 WGII-AR5, risk is “the potential for consequences where something of value is at stake and
127 where the outcome is uncertain” (IPCC 2014b); i.e., any consequence brought about by
128 climate change for biodiversity (IPCC 2014c). If a paper provided risk projections for several
129 species or used several climate change scenarios, we gathered the information for all of them
130 as multiple data entries. Thus, we gathered risks for individual species or mean values for
131 species assemblages reported, compiling 8,158 risk projections (Supplementary Table 2).

132

133 *2.2. Data Analysis*

134 For each study, we classified the biodiversity rich-spots by (a) ecosystem, geographic region
135 and climatic zone; (b) major taxonomic group; (c) whether endemic (only present within the
136 rich-spot area), non-endemic native, or introduced species; and (d) type of impact on
137 biodiversity according to five commonly cited measures of species-level impacts, namely i)
138 population abundance (and catch potential of fisheries as a proxy for abundance), ii)
139 physiology and iii) increase or decrease in spatial range in species distribution; and of
140 community-level impacts, namely iv) diversity (species and taxonomic richness) and v)
141 habitat change (Supplementary Table 3). For conciseness, hereafter we use the term native
142 for non-endemic native species. We also classified climate change scenarios by their

143 projected warming levels (Supplementary Table 3), using IPCC (2018) thresholds, which
144 conclude that limiting global surface air temperature (Gsat) increase to 1.5°C above the pre-
145 industrial level would have a relatively muted (milder) impact on biodiversity, with
146 successively more adverse impacts projected with warming between 1.5-2°C (moderate), 2-
147 3°C (high) and increases in warming of >3°C (very high). For each study, we categorised
148 impacts by scenario used, and time frame over which impacts were projected. In cases in
149 which results were presented as mean values of multiple scenarios, these were categorised as
150 ‘ensemble’. In cases where authors did not follow recognised scenarios, and scenarios
151 described could not be placed within one of these categories (e.g. some studies applied
152 idiosyncratic, extreme scenarios or ad-hoc temperature and/or rainfall changes), these were
153 classified as ‘ambiguous’ and excluded from our main analysis (17 papers corresponding to
154 790 risk projections; Supplementary Table 2). Due to insufficient data, we excluded
155 introduced species in the marine ecosystem from this part of the analysis.

156 We determined an effect size quantified as the percent magnitude of change between
157 current and future time periods. Positive effect sizes represented increases in biodiversity
158 impact categories in the future whereas negative effect sizes represented decreases. For
159 example, a spatial change of 100% meant that a species was projected to double its
160 distribution area within the projected period. Neutral effect sizes indicated that no change in
161 biodiversity was projected to occur.

162 Because effect sizes were based on comparisons between varying time periods, we
163 standardised the effect size by dividing it by the number of years between the periods,
164 obtaining a projected annual incremental change. This standardized effect size allows direct
165 comparisons between studies (it cannot be inferred as an indication of actual change
166 occurring *per year*). Some of the papers did not explicitly specify the baseline current year of
167 the projections, and in these cases we extracted this information from the raw data used in
168 the model described in each paper’s methods (e.g., WorldClim database). We excluded
169 studies covering time spans of more than 150 years, because the calculated relative rates of
170 change are biased by time spans of observation. For instance, the negative power law
171 relationship between observed rates and time spans of observation leads to lower rate
172 estimates when time spans are long (Kemp et al. 2015).

173 We calculated extinction risks as the projected likelihood of extinction (i.e., disappearance of
174 the species within the rich-spot) in each geographic restriction and taxonomic group. We used
175 the International Union for Conservation of Nature (IUCN) criteria of $\geq 80\%$ abundance loss
176 characterizing critical endangerment, with extremely high risk of extinction (criteria A4,
177 IUCN 2012). For spatial change, we adopted the extinction risk criteria from Urban (2015) of
178 $\geq 80\%$ loss of geographic range. We also considered data that explicitly referred to
179 extirpation or extinction. For the extinction risk calculation, we only considered data that
180 presented risk projections for single species (6162 effect sizes for single species), since the
181 mean values presented for species assemblages could bias results. Therefore, we calculated
182 the number and proportion of species projected to have a positive response to climate change,
183 as well as those projected to be at risk of extinction.

184 All statistical analysis was performed in R version 4.0.3 (R Core Team 2020).
185 Because the different impact categories involve very different responses of either species,
186 communities or habitats, we decided to run separate generalised linear mixed-effects models
187 (GLMMs) to determine the significant ($\alpha = 0.05$) drivers of the standardised effect sizes of
188 each impact. The data were therefore subset into five groups, namely species-level impacts: i)
189 abundance, ii) physiology, iii) spatial change; and community-level impacts: iv) diversity, v)
190 habitat change. Because the standardised effect sizes clustered around the mean with higher
191 kurtosis than the Gaussian distribution for all data subsets, we corrected the distribution using
192 the LambertW package (Goerg 2016) thus reducing the effect of extreme outliers (Goerg
193 2011). These transformations were done individually for each effect group rather than overall
194 for the full dataset. The transformed standardised effect sizes were used in all GLMMs and
195 inferences are made using these. All GLMMs were run using the lmer function in the lme4
196 package (Bates et al. 2015) with Gaussian-identity distribution-links.

197 Saturated models for each impact category were built with the following predictor
198 variables included as fixed effects: ecosystem, climatic zone, taxonomic group, species
199 geographic restriction and warming level. Predictor variables were omitted from saturated
200 models if there was only one sub-category for that impact category (e.g., species' distribution
201 as endemic, native and introduced species was omitted from the physiology GLMM as there
202 were only native species in this impact category). The transformed standardised effect size
203 was included in all models as the response variable and the study's unique identity (DOI) was
204 included as a random effect. Once saturated models were constructed, a step-down model-
205 building approach was followed to simplify the models using the step function of the
206 lmerTest package (Kuznetsova et al. 2017). This approach requires the construction of a
207 saturated model followed by the automated removal of fixed effects and random effects that
208 do not contribute significantly ($\alpha = 0.05$ for fixed effects and $\alpha = 0.1$ for random effects) to
209 the intercept and slope of the model (Kuznetsova et al. 2017).

210 Once the final, simplified models (Table 1) for each impact category were obtained
211 from the step-down approach, the summary function of the lmerTest package was used to
212 obtain output tables for the GLMMs, with the model estimates and degrees of freedom using
213 the Satterthwaite's (Kenward-Roger's) approximations for the t test and the corresponding p
214 values (Kuznetsova et al. 2017). In addition to the summary tables, the emmeans function of
215 the emmeans package (Lenth 2019), which uses the Tukey post-hoc method, was used to
216 obtain pairwise comparisons of the sub-categories for each significant predictor variable in
217 the final model. From the summary tables and pairwise comparisons, inferences could be
218 made about the significance of each predictor variable in driving the respective impacts, as
219 well as the difference in the standardised effect sizes between the sub-categories for each
220 significant predictor variable.

221 We created the graphs using GraphPad Prism software version 8.0.1 (GraphPad
222 Software, San Diego, California USA, www.graphpad.com). We created the maps using
223 *tidyverse* and *sf* packages in R software (R Core Team 2019; Wickham et al. 2019; Pebesma
224 et al. 2018).

225

226 **3. Results**

227

228 *3.1. Study biases*

229 Literature on quantifiable climate impacts on biodiversity was unevenly distributed
230 worldwide. Some rich-spots appear very well assessed, with > 250 effect sizes each, namely
231 the Brazilian Atlantic Forest, Mesoamerica, Maputaland-Pondoland-Albany, Cape Floristic
232 Province and California Floristic Province, which together comprise 59 % of our data for
233 terrestrial effect sizes; and the Mediterranean Sea, which comprises 50 % of marine effect
234 sizes (Supplementary Figure 1; Supplementary Table 1). Despite our extensive literature
235 survey, we found no data for 49 % of the 273 rich-spots (Supplementary Figure 1;
236 Supplementary Table 1).

237 In our review, over 200 studies estimated climate change impacts on terrestrial
238 ecosystems, whereas only 34 studies focused on marine ecosystems, suggesting that the
239 ecological literature is biased towards biodiversity from terrestrial ecosystems. Only 14
240 studies assessed impacts over freshwater species, which were analysed within terrestrial due
241 to lack of data. We also found taxonomic bias in the literature towards birds and plants, with
242 over 1400 species each (Figure 4). Most studies considered a few selected threatened or
243 ecologically important species, some assessed only endemic species, and fewer studies
244 modelled all the species (> 100) within a taxonomic group, which reflects an inherent bias
245 towards local endemics in the global biota (e.g. Enquist et al. 2019). Of the species reviewed
246 in our analysis, 73% of the effect sizes referred to non-endemic natives, 17% endemics and
247 5% introduced species (plus <5% unclassified). Because this may under or overestimate their
248 proportions within each and overall study areas we have limited our interpretation to the
249 general direction of effects.

250

251 *3.2. Overall impacts*

252 Climate change is projected to have negative impacts on virtually all terrestrial species in all
253 rich-spots, with the exception of introduced species. This is in accordance with our previous
254 expectations that introduced species would be the least impacted by climate change (Figure
255 1). While this was also generally the case for marine endemic and native (i.e., non-endemic)
256 species, Arctic species were projected to increase their abundance and/or range (Figure 2).
257 When grouping species into climatic zones, and those inhabiting mainland, islands,
258 mountains and in the ocean (Figure 3), all impact categories projected negative effects due to
259 climate change, except in the case of introduced species. Biological measures of response
260 were also projected to be negatively affected, namely species abundance, diversity (including
261 of introduced species), spatial area, habitat area and physiology (Figure 3). Introduced
262 terrestrial species were projected to be significantly positively affected by climate change in

263 the subtropics, mountains and in terms of spatial change (Figure 3). There were insufficient
264 data on marine introduced species for analysis. Species of all groups of organisms and in
265 almost all geographic regions were negatively affected by climate change (Supplementary
266 Figure 3). Only non-endemic native amphibians in Central and South America were projected
267 to benefit from climate change, an unexpected result.

268

269 3.3. *Taxa*

270 All taxonomic groups, except for introduced species and non-endemic native amphibia, were
271 projected to be negatively affected by climate change both overall (Figure 4), and within
272 continents (Supplementary Figures 3, 4). Although amphibians had the highest average effect
273 size increase, meaning an overall positive impact, this average was elevated by a number of
274 native species with very high projected increases (Figure 4). At the same time, amphibians
275 were the group with one of the greatest number of species at risk of extinction (Figure 4). A
276 high number of native terrestrial plants may also face high extinction risk, even though
277 endemics were projected to be significantly more impacted (Figure 4, Table 1). Terrestrial
278 endemic birds were projected to be the most significantly impacted taxa (Figure 4, Table 1).
279 In marine ecosystems, the most impacted taxa appear to be seabed organisms, coral reefs, fish
280 and plants. Endemic marine fishes were projected to be significantly more impacted than
281 non-endemic native fishes (Table 1). Introduced species were positively impacted by climate
282 change, but the species evaluated were restricted to terrestrial plants and a few species of
283 freshwater benthos. Increased climate warming from 1.5 to 3°C increased the risk of species
284 extinctions except in the case of introduced species (Figure 5).

285

286 3.4. *Endemicity*

287 Terrestrial endemic species were projected to be significantly more adversely impacted by
288 climate change than terrestrial non-endemic native and introduced species (Figure 1).
289 Terrestrial endemic species were projected to be 2.7 times more impacted than native species
290 (negative mean standardised effect size of 0.34 % vs 0.92 %) and 10 times more impacted
291 than introduced species. Note that these values refer to the standardised effect sizes, which
292 when considering the time periods of these constant rates, a negative change of 1% can be
293 translated into losses of 80% by 2100. Introduced species were projected to be unresponsive
294 to or benefit from climate change overall. As in terrestrial regions, overall marine endemic
295 species were significantly more impacted than marine native species.

296 Endemic species were projected to be more impacted than natives in almost all
297 assessed rich-spots (with the exception of Cerrado, New Caledonia, Sundaland, Wallacea,
298 Polynesia-Micronesia and Himalaya for terrestrial ecosystems and Humboldt current for
299 marine ecosystems), while introduced species were projected to have either neutral or
300 positive impacts (Figure 1). This finding was supported by the GLMMs, where endemic

301 species were found to be significantly more affected than native and introduced species in
302 abundance, spatial change and diversity models (Table 1). The most prominent negative
303 impacts for endemic species were in South America, Africa and Oceania. In comparison,
304 native species were generally less negatively impacted than endemics, with a few native
305 species even showing small positive impacts (Supplementary Table 1; Table 1). Introduced
306 species were either neutrally or positively impacted, with only slight decreases in some rich-
307 spots (Figure 1, Figure 3, Supplementary Table 1, Table 1).

308 The greater adverse impact of projected climate change on endemic species was
309 evident across climatic zones and geographic regions (Figure 3). Endemic species were
310 projected to be the most sensitive to climate change in all climatic regions, showing higher
311 negative impacts than native or introduced species in tropical, subtropical and temperate
312 regions (Figure 3). Marine endemic species were projected to be more impacted than native
313 species in temperate regions, but not in tropical regions (Figure 3).

314 The five defined impact categories had different magnitudes of impacts on species.
315 Endemics were more impacted in terms of the abundance category than other categories, and
316 compared to native species in land and oceans (Figure 3, Table 1). Diversity was consistently
317 projected to be negatively impacted irrespective of species distribution. It was the only
318 impact category where introduced species were negatively impacted. In contrast, endemics
319 were the most significantly impacted (Table 1). Spatial area impacts were significantly
320 greater for terrestrial and marine endemics than natives and introduced species (Table 1).
321 These spatial area impacts were more prominent for marine species, whereas introduced
322 species are increasing their distributions despite climate change. Loss of habitat area for
323 terrestrial endemic and native species was similar, however marine habitat was more affected
324 for native species. Changes in physiology were more pronounced for marine species than
325 terrestrial.

326 Endemics were consistently projected to be more impacted than native and introduced
327 species under different warming intensities (Figure 5, Table 1). Although the average
328 projected negative mean impacts were constant with climate change intensification, the
329 proportion of species facing extremely high extinction risk increased considerably with
330 warming. The proportion of endemic species at risk of extinction rose tenfold, from 2 % to 20
331 % and 32 % in terrestrial and marine ecosystems, respectively, with a doubling of warming
332 from mild to very high (i.e., from below 1.5 to above 3 °C). Although the magnitude of
333 impact within the standardised time frames is higher for terrestrial than marine endemics (i.e.,
334 they reach high impacts within shorter time frames), the higher proportion of marine
335 endemics in the studies eventually amounts to projected impacts higher than an 80% loss, i.e.,
336 they face extinction risks (Figure 5).

337

338 *3.5. Extinction risk*

339 More than 60 % of tropical terrestrial endemic species were projected to be at risk of
340 extinction due to climate change alone. Endemic species from islands and mountain regions
341 had extremely high extinction risk (100 and 84 % of species, respectively), which was over
342 six times more than in mainland regions (12 %) (Figure 3). Of marine endemic species 54%
343 were at risk of extinction, and while most of these occurred in temperate regions note the
344 Mediterranean bias and paucity of tropical data in available studies (Figures 3, 4).

345 Overall, 92 % of terrestrial endemics were projected to be negatively affected as a
346 result of climate change, in comparison to 80 % and 48 % for terrestrial native and introduced
347 species, respectively. At the same time, 34 % of terrestrial endemic species were estimated to
348 be at extremely high risk of extinction, whereas this risk was 20 % for native and 0 % for
349 introduced species (Figures 3, 4). For marine species, 95 % of endemics and 87% of natives
350 were projected to be negatively impacted by climate change (Figure 4). We found significant
351 statistical differences between marine endemic and native species (Table 1). The proportion
352 of marine species at risk of extinction was more than twice as high for endemics (54 %) than
353 for natives (26 %) (Figure 3).

354 Most species assessed for risk of extinction were in Central and South America for
355 terrestrial (2,782), and the Mediterranean for marine ecosystems (576) (Supplementary
356 Figures 5, 6). However, Oceania, with its islands of high endemism, had the greatest
357 proportion (50 %) of terrestrial species projected to be threatened with extinction by climate
358 change, followed by 30 % in the Americas, Europe and Asia (Supplementary Figure 5). In
359 contrast, Oceania had no marine species projected to be at risk of extinction (Supplemental
360 Figure 6). In marine systems, the Mediterranean, an enclosed sea with high endemism, had
361 the highest number of marine species (25%) projected to have a high risk of extinction with
362 climate change (Supplemental Figure 6).

363

364 **4. Discussion**

365

366 *4.1. Key findings*

367 Our results demonstrate that endemic and native (i.e. indigenous non-endemics) species are
368 consistently more at risk from the adverse effects of climate change than introduced species
369 across both terrestrial and marine environments, geographic areas, climatic zones, taxonomic
370 groups and impact types, with endemics by far the most at-risk group. In contrast, introduced
371 species are projected to experience either neutral or beneficial impacts from changing climate
372 conditions. That introduced species are projected to increase despite climate change is an
373 additional concern within rich-spots. Because rich-spots have high diversity, uniqueness and
374 endemism, our findings are a cause for concern on a global scale.

375 Although the biodiversity rich-spots have been selected qualitatively based on a
376 mixture of criteria on data available at the time, recent analyses of plants support the

377 locations of terrestrial endemism and rarity (Enquist et al. 2019). Additional areas have been
378 proposed and/or protected for nature conservation and there exist many studies on the effects
379 of climate change on species outside these rich-spots that we have excluded from our
380 analysis. A quantitative biogeographic mapping across all biodiversity measures would
381 provide a more robust delimitation of rich-spots, as recently conducted for land plants
382 (Enquist et al. 2019) and the oceans (Zhao et al. 2020). In our analysis, we also found great
383 geographic bias in the sampling of rich-spots, which is a limitation that could skew results. It
384 is also important to note that climate change is one of several, often synergistic, threats to
385 these rich-spots, including habitat loss, overexploitation and pollution (Brook et al. 2008;
386 Albano et al. 2021), which were not considered here. However, the consistent projections of a
387 loss of biodiversity across geographic, taxonomic, and climate impact categories suggests
388 current knowledge is adequate to indicate the general risk of species extinctions, particularly
389 of endemic species.

390

391 *4.2. Introduced species*

392 Areas with high distinctiveness and endemism may be particularly vulnerable to invasion by
393 human introduced species (Ricciardi and Atkinson 2004; Berglund et al. 2009; Bellard et al.
394 2014), notably when native species are naive to introduced predators (Urban 2020). By
395 compressing the range of native species, invasive species may become a source of additional
396 pressure (Vila and Weiner 2004, Catford et al. 2012). Ultimately, the replacement of endemic
397 species by fewer, generalist and widespread opportunists would lead to homogenisation in
398 biodiversity rich-spots, causing ecosystem simplification (McKinney and Lockwood 1999).
399 This phenomenon could be masked initially by relatively unchanged local richness associated
400 with species turnover, but yet still contributing to a pattern of declining global biodiversity
401 (Thomas et al. 2013).

402 In our analysis, plants comprised the majority of introduced species within rich-spots.
403 Plants are some of the world's most proficient invasive species (Lowe et al. 2000). Future
404 climate change may exacerbate such invasions (Liu et al. 2016; Wang et al. 2019). Invasive
405 plants can outcompete native species under increased temperature and carbon dioxide
406 conditions (Van Kleunen et al. 2010; Davidson et al. 2011; Liu et al. 2016). Coastal and high
407 latitude regions have been identified to be most at risk from introduced plants as a result of
408 climate change (Wang et al. 2019). This is supported by our findings that introduced species
409 consistently responded positively to climate change in mountain and island systems.
410 Similarly, Bellard et al. (2014) projected that the biodiversity rich-spots most at risk from
411 invasive species are mainly islands or groups of islands, including Polynesia–Micronesia,
412 New Zealand and the Philippines.

413

414 *4.3. Endemic species*

415 Species adaptation can be enhanced by distributional shifts to habitats in suitable climatic
416 conditions, but this is less likely for endemic than for native species. Greater extinction risks
417 have already been associated with restricted range (rare and often endemic) species (Staudte et
418 al. 2020) in multiple taxonomic groups worldwide (Newbold et al. 2018). Bellard et al.
419 (2014) predicted that biodiversity rich-spots would experience an average 31% loss of current
420 climatic conditions by the 2080s, which would negatively impact an average of 25% of
421 endemic species per hotspot. We found that terrestrial endemic species from island and
422 mountain rich-spots were projected to be at much greater risk of climate change impacts than
423 mainland areas. Both are centres of endemism due to their geographic and environmental
424 isolation (Kier et al. 2009; Noroozi et al. 2018) and are more prone to species invasions than
425 mainlands (Bellard et al. 2014; Elsen and Tingley 2015). These areas have been projected to
426 experience proportionately higher rates of climate-induced range expansions of introduced
427 species (Lamsal et al. 2018, Wiens et al. 2019). Within mountain regions, upward shifts in
428 species elevational ranges (Chen et al. 2011) imply that many montane species will be limited
429 by future altitudinal space, although species responses depend on topographic complexity
430 (Elsen and Tingley 2015). Such consistent extinctions of endemics could disrupt the
431 ecological interactions that buffer ecosystems against disturbances (Mouillot et al. 2013;
432 Pires et al. 2018). Islands of the Caribbean, Madagascar, Indian Ocean Islands, Philippines,
433 Western Ghats and Sri Lanka, could lose all their endemic plants due to climate change by
434 2050, and African mountain rich-spots were also at risk of endemic plant loss (Habel et al.
435 2019).

436

437 *4.4. Island biota*

438 The very high extinction risk we discerned for islands reflects the geographic isolation, high
439 levels of endemism, narrow ranges and small population sizes of many insular species.
440 These factors limit range shifts and increase vulnerability to both stochastic and deterministic
441 threats (Manne et al. 1999). Old oceanic islands generally host orders of magnitude higher
442 levels of endemism (i.e., a higher percentage of all species are endemic) than continental
443 regions due to the greater levels of speciation arising from long periods of insular isolation
444 (Gallagher et al. 2020). However, lower genetic variation can be associated with this greater
445 degree of speciation, leading to poor adaptive, dispersal and defensive capacities and a high
446 vulnerability to extrinsic disturbances (Harter et al. 2015, Kumar and Taylor 2015).
447 Extinction risk of island endemics is further intensified when continuing loss, degradation
448 and fragmentation of habitats across already limited terrain are combined with a changing
449 climate, sea-level rise, extreme weather events and disproportionate prevalence of invasive
450 species (Bellard et al. 2014, Petzold and Magnan 2019). Given the high levels of endemism
451 on islands (Bellard et al. 2014, Petzold and Magnan 2019), the high extinction risk for insular
452 endemics found in our analyses indicates disproportionate risks for future global biodiversity.

453

454 *4.5. Adaptation*

455 This synthesis reveals that climate change is a widespread potential threat to biodiversity
456 rich-spots, regardless of climatic zone, geography or taxonomic grouping. Because
457 biodiversity rich-spots contain disproportionately more global biodiversity per unit area than
458 less rich regions, they are a priority for nature conservation. Importantly, their concentration
459 of endemic species implies particular vulnerability to the effects of climate change, based on
460 results presented here. Whereas a global synthesis also has suggested that endemism
461 increases species risk to climate change, the magnitude of this vulnerability was 6 % higher
462 for endemics than for non-endemics (Urban, 2015). Notably, our results indicate that endemic
463 species from rich-spots are at much higher vulnerability than non-endemics compared to
464 global averages, which reinforces their priority for conservation actions. The local extinctions
465 projected for non-endemic natives within rich-spots could be buffered by more heterogeneous
466 climate change impacts in other parts of their larger ranges. Additionally, they might be able
467 to disperse more readily than endemics, and track suitable climatic conditions, especially in
468 marine ecosystems (Lenoir et al. 2020).

469 The intensity and velocity of climate change can hinder species' ability to adapt to
470 such change (Visser 2008; Brito-Morales et al. 2020). Several measures hold promise for
471 reducing the species extinctions projected. These include implementing globally-networked
472 fully-protected areas on land and sea that are representative of habitats and environmental
473 conditions (Klein et al. 2015; Gray et al. 2016; Zhao et al. 2020). Addressing concomitant
474 stressors to biodiversity may also aid climate change adaptation by increasing resilience of
475 species and natural habitats subjected to degradation and disturbance (Bowler et al. 2019;
476 Travis 2003). For example, sustainable land and sea-use practices aid species persistence and
477 movement between natural habitats, such as provided by habitat connectivity through less-
478 transformed corridors, including multi-use landscapes and restricted seabed trawling.
479 Extending protected areas networks to include such biodiversity rich-spots, managing the
480 intensity of land and sea-use in their surroundings and addressing habitat degradation would
481 enhance their resilience (Bates et al. 2019). However, such protected areas would require
482 careful design to protect biodiversity at the present and under future conditions of climate
483 change (Vale et al. 2018; Hannah et al. 2007, Hannah et al. 2020), in order to facilitate
484 species range migration in response to climate change. Our analysis suggests that the design
485 and implementation of expanded protected area networks (e.g. Vale et al. 2018; Hannah et al.
486 2007, 2020) that prioritise endemic species would increase their efficacy under future
487 conditions of climate change. Focussed monitoring of endemic species' populations and
488 associated habitats would enable the early detection of negative trends in wild populations
489 and provide motivation for active interventions such as active habitat restoration and
490 translocation of populations (Segan et al. 2016).

491 The particular vulnerability of endemic species identified here suggests that even with
492 effective conservation, biodiversity rich-spots might remain at high extinction risk due to
493 increasing climate change alone (Bruno et al. 2018). Apart from our finding that mean effect
494 sizes are consistently negative regardless of warming level, the proportion of species at
495 extremely high risk of extinction increases considerably with temperature. Our results show
496 that even with successful conservation efforts, there remains still an extinction risk for 20 %

497 and 32 % of the terrestrial and marine endemics in biodiversity rich-spots at > 3 °C warming
498 without mitigating climate change. This finding supports previous studies that quantified the
499 benefits of mitigation (i.e., limiting warming) for biodiversity at the global scale (e.g.,
500 Warren et al. 2018, Nunez et al. 2019, Hannah et al. 2020, Hoegh-Guldberg et al. 2019).
501 Therefore, alongside enhanced conservation actions, efforts to mitigate climate change would
502 reduce risks to biodiversity considerably.

503

504 5. References

- 505 Albano, P. G. et al. Native biodiversity collapse in the eastern Mediterranean. *Proc. R. Soc.*
506 *B: Biol. Sci.* **288(1942)**, 20202469 (2021).
- 507 Asaad, I., Lundquist, C. J., Erdmann, M. V. & Costello, M. J. Delineating priority areas for
508 marine biodiversity conservation in the Coral Triangle. *Biol. Conserv.* **222**, 198–211
509 (2018).
- 510 Bates, A. E. et al. Climate resilience in marine protected areas and the ‘Protection Paradox’.
511 *Biol. Conserv.* **236**, 305–314 (2019).
- 512 Bates, D., Maechler, M., Bolker, B. & Walker, S. Fitting linear mixed-effects models using
513 *lme4*. *J. Stat. Soft.* **67(1)**, 1–48 (2015).
- 514 Beaumont, L. J. et al. Impacts of climate change on the world's most exceptional ecoregions.
515 *PNAS* **108(6)**, 2306–2311 (2011).
- 516 Bellard, C. et al. Vulnerability of biodiversity hotspots to global change. *Glob. Ecol.*
517 *Biogeogr.* **23**, 1376–1386 (2014).
- 518 Berglund, H., Järemo, J. & Bengtsson, G. Endemism predicts intrinsic vulnerability to
519 nonindigenous species on islands. *Am. Nat.* **174(1)**, 94–101 (2009).
- 520 Bowler, D. E. et al. Mapping human pressures on biodiversity across the planet uncovers
521 anthropogenic threat complexes. *People and Nature* **2(2)**, 380–394 (2020). Brito-
522 Morales, I. et al. Climate velocity reveals increasing exposure of deep-ocean biodiversity
523 to future warming. *Nat. Clim. Change* **10(6)**, 576–581 (2020).
- 524 Brook, B. W., Sodhi, N. S. & Bradshaw, C. J. Synergies among extinction drivers under
525 global change. *Trends Ecol. Evol.* **23(8)**, 453–460 (2008).
- 526 Brooks, T. M. et al. Habitat loss and extinction in the hotspots of biodiversity. *Conserv. Biol.*
527 **16(4)**, 909–923 (2002).
- 528 Brooks, T. M. et al. Global biodiversity conservation priorities. *Science* **313(5783)**, 58–61
529 (2006).

- 530 Bruno, J. F. et al. Climate change threatens the world's marine protected areas. *Nat. Clim.*
531 *Change* **8**, 499–503 (2018).
- 532 Burrows, M. T. et al. The pace of shifting climate in marine and terrestrial ecosystems.
533 *Science* **334**, 652–655 (2011).
- 534 Catford, J. A., Vesk, P. A., Richardson, D. M. & Pyšek, P. Quantifying levels of biological
535 invasion: towards the objective classification of invaded and invisable ecosystems. *Glob.*
536 *Change Biol.* **18(1)**, 44–62 (2012).
- 537 Chaudhary, C., Richardson, A. J., Schoeman, D. S. & Costello, M. J. Global warming is
538 causing a pronounced dip in marine species richness at the equator. *PNAS*, *in press*
539 (2021)
- 540 Chen, I. C., Hill, J. K., Ohlemüller, R., Roy, D. B. & Thomas, C. D. Rapid range shifts of
541 species associated with high levels of climate warming. *Science* **333(6045)**, 1024–1026
542 (2011).
- 543 Chichorro, F., Juslén, A. & Cardoso, P. A review of the relation between species traits and
544 extinction risk. *Biol. Conserv.* **237**, 220–229 (2019).
- 545 Davidson, A. M., Jennions, M. & Nicotra, A. B. Do invasive species show higher phenotypic
546 plasticity than native species and, if so, is it adaptive? A meta-analysis. *Ecol. Lett.* **14(4)**,
547 419–431 (2011).
- 548 Dirnböck, T., Essl, F. & Rabitsch, W. Disproportional risk for habitat loss of high-altitude
549 endemic species under climate change. *Glob. Change Biol.* **17(2)**, 990–996 (2011).
- 550 Dynesius, M. & Jansson, R. Persistence of within-species lineages: a neglected control of
551 speciation rates. *Evolution* **68(4)**, 923–934 (2014).
- 552 Elsen, P. R. & Tingley, M. W. Global mountain topography and the fate of montane species
553 under climate change. *Nat. Clim. Change* **5**, 772–776 (2015).
- 554 Enquist, B. J. et al. The commonness of rarity: Global and future distribution of rarity across
555 land plants. *Sci. Adv.* **5(11)**, 1–14 (2019).
- 556 Fox, J. & Weisberg, S. *An R Companion To Applied Regression*, Second Ed., Sage, Thousand
557 Oaks CA, 608 p. (2011).
- 558 Gallagher, R. V. et al. Global shortfalls in extinction risk assessments for endemic flora.
559 *bioRxiv* (2020).
- 560 Gaston, K. J., Blackburn, T. M., Greenwood, J. J., Gregory, R. D., Quinn, R. M., & Lawton,
561 J. H. Abundance-occupancy relationships. *J. Appl. Ecol.* **37**, 39–59 (2000).

- 562 Goerg, G. M. Lambert W random variables - a new family of generalized skewed
563 distributions with applications to risk estimation. *Ann. Appl. Stat.* **5(3)**, 2197c2230
564 (2011).
- 565 Goerg, G. M. LambertW: An R package for Lambert W x F Random Variables. R package
566 version 0.6.4 (2016).
- 567 Gray, C. L. et al. Local biodiversity is higher inside than outside terrestrial protected areas
568 worldwide. *Nat. Commun.* **7**, 12306 (2016).
- 569 Habel, J. C. et al. Final countdown for biodiversity hotspots. *Conserv. Lett.* **12(6)**, 1–9
570 (2019).
- 571 Hannah, L. et al. 30% Land Conservation and Climate Action Reduces Tropical Extinction
572 Risk By More Than 50%. *Ecography* **43(7)**, 1–11 (2020).
- 573 Hannah L. et al. Protected area needs in a changing climate. *Front. Ecol. Environ.* **5(3)**, 131–
574 138 (2007).
- 575 Harter, D. E. V. et al. Impacts of global climate change on the floras of oceanic islands –
576 Projections, implications and current knowledge. *Perspect. Plant Ecol.* **17(2)**, 160–183
577 (2015).
- 578 Hoegh-Guldberg, O. et al. The human imperative of stabilizing global climate change at 1.5
579 degrees C. *Science* **365(6459)**, eaaw6974 (2019).
- 580 IPCC. Summary for Policymakers. *In: Climate Change 2014: Synthesis Report. Contribution*
581 *of Working Groups I, II and III to the Fifth Assessment Report of the Intergovernmental*
582 *Panel on Climate Change [Core Writing Team, R.K. Pachauri and L.A. Meyer (eds.)].*
583 *IPCC, Geneva, Switzerland Cambridge, 1–32 (2014a).*
- 584 IPCC. Annex II: glossary. Agard J. et al. (eds) *Climate change 2014: impacts, adaptation, and*
585 *vulnerability. Part B: regional aspects. Contribution of Working Group II to the Fifth*
586 *Assessment Report of the Intergovernmental Panel on Climate Change. Cambridge*
587 *University Press, Cambridge, 1757–1776 (2014b).*
- 588 IPCC. Annex II: glossary. Mach, K. J., Planton, S. & von Stechow, C. (eds) *Climate change*
589 *2014: synthesis report. Contribution of Working Groups I, II and III to the Fifth*
590 *Assessment Report of the Intergovernmental Panel on Climate Change. In: Core Writing*
591 *Team, Pachauri RK, Meyer LA (eds). IPCC, Geneva, Switzerland, 117–130 (2014c).*
- 592 IPCC. Summary for Policymakers. *In: Climate Change and Land: An IPCC special report on*
593 *climate change, desertification, land degradation, sustainable land management, food*
594 *security, and greenhouse gas fluxes in terrestrial ecosystems [Shukla, P. R. et al. (eds.)].*
595 *World Meteorological Organization, Geneva, Switzerland, 1–43 (2019).*

596 IPCC. Summary for Policymakers. *In*: Global Warming of 1.5°C. An IPCC Special Report on
597 the impacts of global warming of 1.5°C above pre-industrial levels and related global
598 greenhouse gas emission pathways, in the context of strengthening the global response to
599 the threat of climate change, sustainable development, and efforts to eradicate poverty
600 [Masson-Delmotte, V. et al. (eds.)]. World Meteorological Organization, Geneva,
601 Switzerland, 1–32 (2018).

602 IUCN. Red List Categories and Criteria: Version 3.1. Second edition. Gland, Switzerland and
603 Cambridge, UK (2012). Available at:
604 <https://portals.iucn.org/library/sites/library/files/documents/RL-2001-001-2nd.pdf>

605 Kemp, D., Eichenseer, K. & Kiessling, W. Maximum rates of climate change are
606 systematically underestimated in the geological record. *Nat. Commun.* **6**, 8890 (2015).

607 Kier, G. et al. A global assessment of endemism and species richness across island and
608 mainland regions. *PNAS* **106(23)**, 9322–9327 (2009).

609 Klein, C. J. et al. Shortfalls in the global protected area network at representing marine
610 biodiversity. *Sci. Rep.* **5**, 17539 (2015)

611 Kumar, L. & Taylor, S. Exposure of coastal built assets in the South Pacific to climate risks.
612 *Nat. Clim. Change* **5**, 992–996 (2015).

613 Kuznetsova, A., Brockhoff, P.B. & Christensen, R. H. B. lmerTest Package: Tests in Linear
614 Mixed Effects Models. *J. Stat. Soft.* **82(13)**, 1–26 (2017).

615 Lamsal, P., Kumar, L., Aryal, A., & Atreya, K. (2018). Invasive alien plant species dynamics
616 in the Himalayan region under climate change. *Ambio* **47**, 697–710.

617 Lenth, R. Emmeans: Estimated Marginal Means, aka Least-Squares Means. Available at
618 <https://cran.r-project.org/package=emmeans> ; accessed in april, 2019.

619 Lenoir, J. et al. Species better track climate warming in the oceans than on land. *Nat. Ecol.*
620 *Evol.* **4**, 1044–1059 (2020).

621 Liu, Y. et al. Do invasive alien plants benefit more from global environmental change than
622 native plants? *Glob. Change Biol.* **23(8)**, 3363–3370 (2016).

623 Loarie, S. R. et al. The velocity of climate change. *Nature* **462(7276)**, 1052–1055 (2009).

624 Lowe, S., Browne, M., Boudjelas, S. & De Poorter, M. *100 Of The World's Worst Invasive*
625 *Alien Species: A Selection From The Global Invasive Species Database*. Auckland:
626 Invasive Species Specialist Group (2000).

627 Lucas, P. M., González-Suárez, M. & Revilla, E. Range area matters, and so does spatial
628 configuration: predicting conservation status in vertebrates. *Ecography* **42(6)**, 1103–1114
629 (2019).

- 630 Malcolm, J. R., Liu, C., Neilson, R. P., Hansen, L. & Hannah, L. Global Warming and
631 Extinctions of Endemic Species from Biodiversity Hotspots. *Conserv. Biol.* **20(2)**, 538–
632 548 (2006).
- 633 Manne, L. L., Brooks, T. M. & Pimm, S. L. Relative risk of extinction of passerine birds on
634 continents and islands. *Nature* **399(6733)**, 258–261 (1999).
- 635 McKinney, M. L. & Lockwood, J. L. Biotic homogenization: a few winners replacing many
636 losers in the next mass extinction. *Tree* **14(11)**, 450–453 (1999).
- 637 Mittermeier, R. A. et al. *Hotspots Revisited: Earth's Biologically Richest And Most*
638 *Endangered Terrestrial Ecoregions*. CEMEX, Agrupación Sierra Madre, SC (2004).
- 639 Mittermeier, R. A., Turner, W. R., Larsen, F. W., Brooks, T. M. & Gascon, C. Global
640 Biodiversity Conservation: The Critical Role of Hotspots. In: Zachos, F. & Habel, J.
641 (eds) *Biodiversity Hotspots*. Springer, Berlin, Heidelberg (2011).
- 642 Molinos, J. G., Halpern, B. S., Schoeman, D. S., Brown, C. J., Kiessling, W., Moore, P. J.,
643 Pandolfi, J. M., Poloczanska, E. S., Richardson, A. J. & Burrows, M. T. Climate velocity
644 and the future global redistribution of marine biodiversity. *Nat. Clim. Change* **6(1)**, 83–
645 88 (2016).
- 646 Mouillot, D. et al. Rare species support vulnerable functions in high-diversity ecosystems.
647 *PLoS Biol.* **11(5)**, 1–11 (2013).
- 648 Myers, N., Mittermeier, R. A., Mittermeier, C. G., Fonseca, G. A. B. & Kent, J. Biodiversity
649 hotspots for conservation priorities. *Nature* **403**, 853–858 (2000).
- 650 Newbold, T. et al. Widespread winners and narrow-ranged losers: Land use homogenizes
651 biodiversity in local assemblages worldwide. *PLoS Biol.* **16(12)**, e2006841 (2018).
- 652 Noroozi, J. et al. Hotspots within a global biodiversity hotspot - areas of endemism are
653 associated with high mountain ranges. *Sci. Rep.* **8**, 10345 (2018).
- 654 Noss, R. F. et al. How global biodiversity hotspots may go unrecognized: Lessons from the
655 North American Coastal Plain. *Divers. Distrib.* **21(2)**, 236–244 (2015).
- 656 Nunez, S., Arets, E., Alkemade, R., Verwer, C. & Leemans, R. Assessing the impacts of
657 climate change on biodiversity: is below 2 °C enough? *Clim. Change* **154**, 351–365
658 (2019).
- 659 Oduor, A. M. O., Leimu, R. & van Kleunen, M. Invasive plant species are locally adapted
660 just as frequently and at least as strongly as native plant species. *J. Ecol.* **104(4)**, 957–
661 968 (2016).
- 662 Olson, D. & Dinerstein, E. The global 200: Priority ecoregions for global conservation. *Ann.*
663 *Missouri Bot.* **89(2)**, 199–224 (2002).

- 664 Pebesma, E. Simple Features for R: Standardized Support for Spatial Vector Data. *The R*
665 *Journal* **10(1)**, 439–446 (2018).
- 666 Petzold, J. & Magnan, A. K. Climate change: Thinking small islands beyond Small Island
667 Developing States (SIDS). *Clim. Change* **152**, 145–165 (2019).
- 668 Pinheiro J., Bates D., DebRoy S. & Sarkar D., 2020. nlme: Linear and Nonlinear Mixed
669 Effects Models. R package version 3.1-147, <https://CRAN.R-project.org/package=nlme>.
- 670 Pinsky, M. L. et al. Greater vulnerability to warming of marine versus terrestrial ectotherms.
671 *Nature* **569(7754)**, 108–111 (2019).
- 672 Pires, A. P. F., Srivastava, D. S. & Farjalla, V. F. Is biodiversity able to buffer ecosystems
673 from climate change? What we know and what we don't. *BioScience* **68(4)**, 273–280
674 (2018).
- 675 Poloczanska, E. S., et al. Global imprint of climate change on marine life. *Nat. Clim. Change*
676 **3**, 919–925 (2013).
- 677 R Core Team. R: A language and environment for statistical computing. R Foundation for
678 Statistical Computing, Vienna, Austria, <http://www.R-project.org> (2020).
- 679 Román-Palacios, C. & Wiens, J. J. Recent responses to climate change reveal the drivers of
680 species extinction and survival. *PNAS* **117 (8)**, 4211–4217 (2020).
- 681 Ricciardi, A. & Atkinson, S. K. Distinctiveness magnifies the impact of biological invaders
682 in aquatic ecosystems. *Ecol. Lett.* **7(9)**, 781–784 (2004).
- 683 Saeedi, H., Basher, Z. & Costello, M. J. Modelling present and future global distributions of
684 razor clams (Bivalvia: Solenidae). *Helgol. Mar. Res.* **70**, 23 (2017).
- 685 Segan D. B., Murray, K. A. & Watson, J. E. M. A global assessment of current and future
686 biodiversity vulnerability to habitat loss–climate change interactions. *Glob. Ecol.*
687 *Conserv.* **5**, 12–21 (2016).
- 688 Senior, R.A., Hill, J.K., Benedick, S. & Edwards, D.P. Tropical forests are thermally buffered
689 despite intensive selective logging. *Glob. Change Biol.* **24(3)**, 1267–1278 (2018).
- 690 Staude, I. R., Navarro, L. M. & Pereira, H. M. Range size predicts the risk of local extinction
691 from habitat loss. *Global Ecol. Biogeogr.* **29(1)**, 16–25 (2020).
- 692 Sunday, J. M., Bates, A. E. & Dulvy, N. K. Thermal tolerance and the global redistribution of
693 animals. *Nat. Clim. Change* **2(9)**, 686–690 (2012).
- 694 Thomas, C. D. Local diversity stays about the same, regional diversity increases, and global
695 diversity declines. *PNAS* **110(48)**, 19187–19188 (2013).

696 Travis J. M. J. Climate change and habitat destruction: a deadly anthropogenic cocktail. *Proc.*
697 *R. Soc. B.* **270**, 467–473 (2003).

698 Urban, M. C. Accelerating extinction risk from climate change. *Science* **348**(6234), 571–573
699 (2015).

700 Urban, M. C. Climate-tracking species are not invasive. *Nat. Clim. Change* **10**, 382–384
701 (2020).

702 Vale, M. M., Souza, T. V., Alves, M. A. S. & Crouzeilles, R. Planning protected areas
703 network that are relevant today and under future climate change is possible: the case of
704 Atlantic Forest endemic birds. *PeerJ* **5**, 1–20 (2018).

705 Van Kleunen, M., Weber, E. & Fischer, M. A meta-analysis of trait differences between
706 invasive and non-invasive plant species. *Ecol. Lett.* **13**(2), 235–245 (2010).

707 Vilà, M. & Weiner, J. Are invasive plant species better competitors than native plant species?
708 – evidence from pair-wise experiments. *OIKOS* **105**(2), 229–238 (2004).

709 Visser, M. E. Keeping up with a warming world; assessing the rate of adaptation to climate
710 change. *Proc. R. Soc. B: Biol. Sci.* **275**(1635), 649–659 (2008).

711 Wang, C. J., Li, Q. F. & Wan, J. Z. Potential invasive plant expansion in global ecoregions
712 under climate change. *PeerJ* **7**, e6479 (2019).

713 Warren, R., Price, J., VanDerWal, J., Cornelius, S. & Sohl, H. The implications of the United
714 Nations Paris Agreement on climate change for globally significant biodiversity areas.
715 *Clim. Change* **147**, 395–409 (2018).

716 Wickham, H. et al. Welcome to the tidyverse. *J. Open Source Softw.* **4**(43), 1686 (2019).

717 Wiens, J. J., Litvinenko, Y., Harris, L. & Jezkova, T. Rapid niche shifts in introduced species
718 can be a million times faster than changes among native species and ten times faster than
719 climate change. *J. Biogeogr.*, **46**(9), 2115–2125 (2019).

720 Williams, K. J. et al. Forests of East Australia: The 35th Biodiversity Hotspot. *In: Zachos F.*
721 *& Habel J. (eds) Biodiversity Hotspots.* Springer, Berlin, Heidelberg (2011).

722 Yasuhara, M. et al. Past and future decline of tropical pelagic biodiversity. *Proceedings of the*
723 *National Academy of Sciences* **117** (23), 12891–12896 (2020).

724 Zhao, Q. et al. Where Marine Protected Areas would best represent 30% of ocean
725 biodiversity. *Biol. Conserv.* **244**, 108536 (2020).

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727

728 **Figure 1. Climate change impacts on species within terrestrial rich-spots.** Mean
729 standardised effect sizes of (a) all species, (b) endemic species, (c) native species and (d)
730 introduced species. The colour scale is standardised for all maps and ranges from Positive
731 (greater than 2%, blue) to Negative (less than -2%, red). Maximum and minimum values for
732 mean projected standardised effect sizes range from 3.2% (Non-endemic native -
733 Drakensberg Montane Woodlands and Grasslands) to -2.2% (Endemic - Caribbean Islands)
734 (Supplementary Table 1).

735

736 **Figure 2. Climate change impacts on species within marine rich-spots.** Mean standardised
737 effect sizes of (a) all species, (b) endemic species, (c) native species. The colour scale is
738 standardised for all maps and ranges from Positive (greater than 2%, blue) to Negative (less
739 than -2%, red). Maximum and minimum values for mean projected standardised effect sizes
740 range from 3.2% (Native - Drakensberg Montane Woodlands and Grasslands) to -2.2%
741 (Endemic - Caribbean Islands) (Supplementary Table 1).

742

743 **Figure 3. Climate change effects on species classified by climatic, geographic and**
744 **biological impact categories.** (a) Mean standardised effect sizes (mean \pm 95 % CI) represent
745 increases and decreases in impact categories. Comparisons between species with different
746 geographic distributions within impact categories are described in Table 1. (b) The proportion
747 of species that are positively or negatively impacted by climate change. Species groups with
748 risk projections higher than 80 % losses are considered at extremely high extinction risk
749 (endemics) and local extinction risk (non-endemic natives) within the rich-spots.

750

751 **Figure 4. Climate change effects on species with different geographic distributions**
752 **within different taxonomic groups.** (a) Mean standardised effect sizes (mean \pm 95 % CI)
753 representing increases and decreases. Comparisons between species with different
754 geographic distributions within impact categories are described in Table 1. (b) The number of
755 species that are positively or negatively impacted by climate change. Species with risk
756 projections higher than 80 % losses are considered at extremely high extinction risk
757 (endemics) and local extinction risk (non-endemic natives) within the rich-spots.

758

759 **Figure 5. The impact of warming level on species.** (a) Mean standardised effect sizes with
760 different warming levels where mild, moderate, high and very high levels correspond to <1.5
761 °C, 1.5-2 °C, 2-3 °C and >3 °C, respectively. (b) Diagram indicating the relative proportions
762 of species at extremely high extinction risk for each of the different warming levels. Species
763 with risk projections higher than 80 % losses are considered at extremely high extinction risk
764 (endemics) and local extinction risk (non-endemic natives) within the rich-spots.

765

766 **Table 1. Summary of the generalised linear mixed-effects models for the standardised**
767 **effect sizes for the projected impacts of climate change** on species and communities in
768 terrestrial and marine rich-spots globally. Models were run separately for each impact
769 category. Response variables and predictor variables included in the models are indicated. All
770 predictor variables were included in the models as fixed effects and the unique identity (DOI)
771 of each journal article was included in each model as a random effect. Parameter estimates,
772 standard errors (SE), degrees of freedom (df), t values and p values (computed using
773 Satterthwaite's method of approximation) for the models are given. Significant predictor
774 variables indicated in bold with significance given as * $p < 0.05$, ** $p < 0.01$ and *** $p < 0.001$.

775

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777