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1 **Unveiling trade-offs in resource selection of migratory caribou using a mechanistic**
2 **movement model of availability**

3

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18

19 **Abstract**

20 Habitat selection is a multi-level, hierarchical process that should be a key component in
21 the balance between food acquisition and predation risk avoidance (food-predation trade-
22 off). However, to date, studies have not fully elucidated how fine- and broad-scale habitat
23 decisions by individual prey can help balance food versus risk. We studied broad-scale
24 habitat selection by Newfoundland caribou (*Rangifer tarandus*), focusing on trade-offs
25 between predation risk versus access to forage during the calving and post-calving period.
26 We improved traditional measures of habitat availability by incorporating fine-scale
27 movement patterns of caribou into the availability kernel, thus enabling separation of
28 broad and fine scales of selection. Remote sensing and field surveys served to create a
29 spatio-temporal model of forage availability, whereas GPS telemetry locations from 66
30 black bears (*Ursus americanus*) and 59 coyotes (*Canis latrans*) provided models of
31 predation risk. We then used GPS telemetry locations from 114 female caribou to assess
32 food-predation trade-offs through the prism of our refined model of caribou habitat
33 availability. We noted that migratory movements of caribou were oriented mainly
34 towards habitats with abundant forage and lower risk of bear and (to a lesser extent)
35 coyote encounter. These findings were generally consistent across caribou herds and
36 would not have been evident had we used traditional methods instead of our refined
37 model when estimating habitat availability. We interpret these findings in the context of
38 stereotypical migratory behaviour observed in Newfoundland caribou, which occurs
39 despite the extirpation of wolves (*Canis lupus*) nearly a century ago. We submit that
40 caribou are able to balance food acquisition against predation risk using a complex set of
41 factors involving both finer and broader scale selection. Accordingly, our study provides

42 a strong argument for using refined habitat availability estimates when assessing food-
43 predation trade-offs.

44

45 **Key-words:** caribou (*Rangifer tarandus*), conservation biology, habitat selection,
46 mechanistic modelling, step-selection function, migration.

47 **Introduction**

48 Understanding the drivers of antipredator responses and the efficiency by which animals
49 trade-off food versus safety is crucial, since anti-predator behavioural modification can
50 have profound consequences on fitness, and ultimately, population dynamics (Gaillard et
51 al. 2010). Beyond their direct lethal impact, predators can increase physiological stress in
52 prey (Creel et al. 2009) and cause behavioural adjustments that contribute to the net effect
53 of predation (Creel and Christianson 2008, Lima and Bednekoff 1999, Schmitz 2008).
54 These anti-predator behavioural adjustments can also induce a reduction in foraging
55 efficiency (foraging cost of predation; Brown and Kotler 2004), ultimately leading prey
56 to compromise between food and safety. Prey are able to reduce the impact of predation
57 through various behavioural strategies, such as vigilance, grouping, and movement (Lima
58 1998, Lima and Dill 1990).

59 Movement is a central process in animal ecology, including in the study of
60 predator-prey interactions (Laundré 2010, Mitchell and Lima 2002, Nathan 2008).
61 Indeed, animals move in response to a variety of competing pressures such as the need to
62 feed, avoid predators, breed, and rear offspring (Brown et al. 1999, Cresswell 2008).
63 These competing demands give rise to trade-offs that individuals must mediate through
64 their space use and movements (Hebblewhite and Merrill 2009, Lima 1998), resulting in
65 distinctive patterns of habitat selection. Numerous studies have tried to unveil potential
66 trade-offs for prey through the process of habitat selection (e.g. Creel et al. 2005, Fortin
67 and Fortin 2009, Hebblewhite and Merrill 2009). Habitat selection is defined as the
68 disproportionate use of a habitat relative to its availability (Johnson 1980, Manly et al.
69 2002), and elucidating habitat selection determinants remains a central and unifying

70 concept bridging spatial and temporal scales (Mayor et al. 2007, Morris 2003). Indeed,
71 studies often have compared habitat selection across multiple scales (e.g. Dussault et al.
72 2005, Hebblewhite and Merrill 2009, van Beest et al. 2010). Especially owing to
73 prevalent and rapid environmental change, there is increasing interest in understanding
74 motivations associated with an animal's habitat selection, especially in the context of
75 revealing how such selection may be mismatched with current or future environmental
76 conditions (Middleton et al. 2013, Sih et al. 2011). Indeed, habitat selection is one of the
77 most studied concepts in ecology.

78 Despite such focused attention, habitat selection studies are frequently limited in
79 the insights they provide, due to: 1) absence of robust information (e.g., qualitative field
80 surveys or predator data), leading to a weak or simplified definition of available forage or
81 predation risk (Hebblewhite and Haydon 2010); 2) restrictive analysis of a single level of
82 selection (Boyce 2006); or 3) trivial or problematic comparison of use versus availability
83 to infer selection (Aarts et al. 2013). Notably, there remain substantive challenges in
84 understanding behavioural processes underlying habitat selection and the animal
85 motivation by which it is governed. This difficulty arises because of non-independent
86 behavioural processes and overlapping motivations across levels of selection, as well as
87 the conditional and statistical nature of 'selection'. This means that previous work often
88 addressed the question of resource selection on the basis of relatively simple (and
89 presumably imprecise) algorithms when defining habitat availability (Beyer et al. 2010).
90 It follows that such an approach may mask actual patterns and drivers of habitat selection
91 at a particular level due to artefacts of finer-scale processes also being considered in the
92 use-availability statistical comparison. Better integration of animal decisions that are

93 quantifiable on the basis of movement ecology should therefore be useful. Here, we
94 propose a refined approach for defining availability that considers finer scale selection
95 patterns and thereby improves the distinction between levels of selection, while also
96 providing insight into motivation underlying such selection.

97 Caribou (*Rangifer tarandus* L.) offer a unique system for studying food-predation
98 trade-offs in habitat selection, and on the island of Newfoundland, Canada, there are 14
99 major caribou herds with most exhibiting some degree of migratory behaviour involving
100 the annual use of traditional calving grounds by females. These herds are largely distinct
101 and spatially disjunct at calving (Bastille-Rousseau et al. 2013). Interestingly, during the
102 last 50 years Newfoundland caribou have undergone marked fluctuations in abundance,
103 with populations increasing rapidly during the 1980s to mid-1990s, and declining during
104 the 2000s (Mahoney and Schaefer 2002b, Mahoney et al. 2011).

105 We develop a refined model of availability to study broad-scale habitat selection,
106 with an emphasis on trade-offs between predation risk and foraging. More specifically,
107 we use a mechanistic model based on a step-selection function that approximates fine-
108 scale movement to create a refined sample of habitat availability. We use this model to
109 study selection of calving grounds (referred as second-order level of selection; Johnson
110 1980) as well as core areas within the calving grounds (referred as third-order level of
111 selection) in response to vegetation biomass and current predation risk (black bears
112 [*Ursus americanus* L.] and coyotes [*Canis latrans* Say.]). We chose to focus our analysis
113 at the herd level as caribou aggregate into groups during this period. Coyotes are non-
114 native predators that became widespread in Newfoundland in the 1990s. Considering the
115 high amount of caribou calf mortality during calving periods (Trindade et al. 2011) and

116 recent evidence of density-dependent population fluctuations (Mahoney and Schaefer
117 2002b, Mahoney et al. 2011), we first predicted that predation has a stronger effect on
118 habitat use and that selection of calving grounds would be mostly driven by an expression
119 of predation risk avoidance. Second, we predicted that access to rich foraging sites would
120 be the main factor driving habitat selection at the third order: i.e., the selection of core
121 areas within the calving ground. As a side contribution emanating from our analysis, we
122 compared insights obtained from our mechanistic definition of availability to the
123 traditional approach, and predicted that our refined model would provide insights into the
124 processes underlying caribou decisions vis-à-vis food-predation trade-offs that would not
125 otherwise be evident. We believe that our approach could provide a major shift in how
126 ecologists approach questions related to animal behavioural adjustments in response to
127 the subtle interplay between risks and rewards in their environment.
128

129 **Material and methods**

130 **Study area**

131 Newfoundland is a 108,860-km² island at the eastern extremity of Canada (47°44N,
132 59°28W to 51°44N, 52°38W), with humid-continental climate and substantial year-round
133 precipitation (Environment Canada 2013). Natural habitat consists mainly of coniferous
134 and mixed forests of balsam fir (*Abies balsamea*), black spruce (*Picea mariana*), and
135 white birch (*Betula papyrifera*), and in some locations substantial areas of bogs and heath
136 or barren habitats. Most of our analyses were based on a Landsat TM satellite imagery,
137 with a resolution of 25 m, classified into 6 different habitat types: wetland habitats
138 (Wetland), barren and other open habitats (Barren), mixed and coniferous open stand
139 (CO), mixed and coniferous dense stand (CD), open water (Water), and a category
140 (Other) comprised of rarer habitats such as broadleaf stands, herbs and bryoids (Wulder
141 et al. 2008). Anthropogenic disturbances are limited in caribou range in Newfoundland
142 but consist of logging, hydroelectric development, and roads. We restricted our analysis
143 to five important migratory herds located south of the main east-west highway that
144 crosses the island (Fig. 1).

145

146 **Animal capture and monitoring**

147 During 2006-2010, more than 200 caribou were captured, principally during winter, and
148 fitted with global positioning system (GPS) collars that obtained locations every 2 hours.
149 We focused on 114 adult females (271 caribou-years and 384,764 locations) that were
150 followed during 2007-2010 and that resided in 5 distinct herds (Buchans [n=17 caribou],
151 Lapoile [n=19], Middle Ridge [n=28], Pot Hill [n=18] and Gaff Topsails [n=32]). We

152 limited our movement analysis to the crucial, post-migratory period of calving and post-
153 calving (1 May - 1 August) when most caribou neonate mortalities occur. We also used
154 GPS locations of 66 adult male and female black bears (125 bear-years and 96,531
155 locations) and 47 adult male and female coyotes (59 coyote-years and 18,842 locations)
156 followed during the same period in the vicinity of our study area. Although most of the
157 study area contained radio-collared predators, the central portion of our study area was
158 under-represented in terms of predator locations, most notably for bears. We therefore did
159 not use the density of locations as a measure of predation risk (e.g. kernel density
160 estimate), but rather sought to quantify predation risk via habitat selection approaches.

161 We used caribou GPS locations to create a 95% bivariate kernel density estimate
162 using an *ad hoc* approach to estimate the smoothing parameter to roughly delineate the
163 areas used during calving and post-calving (hereafter, “calving grounds”) for each herd
164 (see Worton 1989). We then created a general study area of availability that encompassed
165 these five herds that was generally delineated by the Trans-Canada Highway to the north,
166 east and west, and by the coast to the south (Fig. 1). The study area and the herd calving
167 ground delineations represented our two levels of availability (second and third-order
168 selection, respectively; Johnson 1980).

169

170 **Definitions of availability**

171 1- Random model

172 Most resource selection analysis involving radio-telemetry is based on the use versus
173 availability design, where availability is sampled from locations drawn within an area
174 assumed to define what actually is available to the animal. However, defining habitat

175 availability has constituted a longstanding challenge in ecology (Beyer et al. 2010).
176 Specifically, ‘availability’ usually is identified by sampling habitats randomly within the
177 defined area and relying on the assumption that accessibility of different habitats is
178 similar across all individuals. This assumption depends on habitat connectivity and
179 animal movement (Dancose et al. 2011), and is less likely to be satisfied at higher orders
180 of selection (Johnson 1980). Our first definition of availability was based on this simple
181 definition (hereafter, "random model"). We generated 5 million random locations within
182 the study area and assigned each location evenly to one of 15,000 virtual individuals. We
183 also generated 1 million random locations within each herd’s calving ground and equally
184 associated them with one of 3,000 virtual individuals. We randomly assigned each
185 location to a specific day and each individual to a specific year (2007-2010 [2009-2010
186 for Middle Ridge]) corresponding to the radio-telemetry data for each herd. Associating
187 random locations to an individual, day, and year was necessary for subsequent analyses.

188

189 2- Mechanistic model

190 For fine-scale analyses of resource selection, realistic and restrictive definitions of
191 availability based on movement properties have been proposed (Aarts et al. 2013, Fieberg
192 et al. 2010, Fortin et al. 2005, Hjermann 2000, Matthiopoulos 2003), but for broad-scale
193 analyses, alternatives are still limited (see Arthur et al. 1996). Ecologists generally view
194 habitat selection as a hierarchical process; it is well accepted that fine-scale selection is a
195 function of resource availability at the same level, yet availability is defined by broad-
196 scale habitat selection (DeCesare et al. 2012, Mysterud and Ims 1998). The consequence
197 of such a view is that, when inferring motivation behind selection, each level is viewed as

198 independent. This view has been reinforced by the hierarchical habitat selection
199 hypothesis (HSS), proposed by Rettie & Messier (2000), where broad-scale selection
200 reflects the most relevant limiting factors (but see, Dussault et al. 2005, Hebblewhite and
201 Merrill 2009 for a critical discussion of this hypothesis). Use of specific resources, in
202 addition, should be seen as a summation of multiple processes operating at different
203 scales adding to the difficulty of interpreting scale-specific selection. Therefore, inferring
204 motivation behind such patterns often can be challenging (Beyer et al. 2010).

205 To understand the motivation behind caribou migration or other broad-scale
206 habitat selection patterns, researchers might compare locations used by animals to a set of
207 random locations within a larger area. However, mammals, and notably ungulates, are
208 known to display movements that balance both long-term and short-term motivations
209 (Mueller et al. 2011), and therefore a more refined habitat selection analysis should
210 reflect finer-scale decisions that are made when moving within the larger area. As we
211 seek to understand the motivation behind a level-specific behaviour as well as a realistic
212 estimate of habitat availability, we need to control for the influence of fine-scale selection
213 patterns. This can be achieved by refining our definition of availability to consider fine-
214 scale movements. In other words, we examine whether observed spatial patterns result
215 from actual differences in broad-scale space-use, or whether they are simply an artefact
216 of fine-scale movement choices. Refining the definition of availability therefore allows
217 for a more conservative estimate of broad-scale selection that improves the distinction
218 between levels of selection.

219 To get a more realistic (and restrictive) view of availability that considers fine-
220 scale animal movements, we built a spatially-explicit, mechanistic model that represented

221 between-patch transition in areas that could be occupied by caribou. At a minimum, a
222 suitable model of fine-scale movement should include step lengths and turning angles,
223 but also could incorporate a weighting function representing preference for specific
224 resources (Rhodes et al. 2005). Such a model would therefore include both reduction of
225 movement and biased movement to inform fine-scale selection patterns (Bastille-
226 Rousseau et al. 2010, Moorcroft and Barnett 2008).

227 We used a spatially-explicit mechanistic model, based on a step-selection
228 function, to provide our second definition of availability (hereafter, "mechanistic
229 model"). We randomly initiated this model within the study area to investigate selection
230 of caribou calving grounds (second-order selection) and to study third-order selection
231 within each of the five calving grounds. This model included movement parameters (step
232 length and turning angles) derived from collared caribou combined with a weighting
233 function translating between-habitat preference in inter-patch movements. Specifically,
234 we estimated habitat-specific step length and turning angle distributions using Weibull
235 and bivariate von-Mises distributions, respectively. The weighting function was
236 calculated by comparing an actual animal step originating in a specific habitat to 100
237 potential steps based on step length and turning angle distributions. Full details regarding
238 model formulation and estimation of parameters can be found in Potts *et al.* (2014). We
239 initiated 15,000 virtual individuals within the broader areas and 3,000 within each calving
240 ground, which were assigned locations every 2 hours and then processed similarly to
241 locations from the random model.

242

243 **Predation model**

244 We used a resource selection function (RSFs; Boyce et al. 2002, Manly et al. 2002) to
 245 describe the spatial relationship between the probability of occurrence of coyotes and
 246 black bears according to landscape attributes. We estimated RSFs by comparing habitat
 247 characteristics at observed and random locations with mixed-effects logistic regression
 248 models, with the identity of the individual as random factor (i.e. random intercept; Gillies
 249 et al. 2006, Hebblewhite and Merrill 2008) to account for spatial autocorrelation and
 250 differences in sample size. We drew random locations for a given individual within the
 251 99% utilization distribution evaluated from a Brownian bridge kernel approach (Horne et
 252 al. 2007). Random locations were drawn at a density of 2 points per km². Observed and
 253 random locations were characterized by dummy variables representing landcover types
 254 (with Wetland as the reference category), as well as elevation, slope, and proportion of
 255 each habitat category within a 5-km radius (except habitats classified as ‘Water’ and
 256 ‘Other’). Proportion of habitat within a buffer was used to account for the presence of a
 257 functional response in habitat selection (Moreau et al. 2012, Mysterud and Ims 1998),
 258 which may improve model fit, especially over large areas (Aarts et al. 2013). We
 259 therefore added an interaction term between coefficients for a specific habitat and its
 260 proportion (Aarts et al. 2013, McLoughlin et al. 2010).

261 The global RSF took the form:

$$262 \quad w(x) = \exp(\beta_1 x_1 + \dots + \beta_u x_{u_{ij}} + \beta_{u_{5k}} x_{(u_{5k})_{ij}} + \dots + \beta_u x_u * \beta_{u_{5k}} x_{(u_{5k})_{ij}} + \gamma_{0j}) \quad (1)$$

263 where $w(x)$ represented the RSF scores, β_u was the selection coefficient for resource x_u or
 264 for the elevation and the slope, $\beta_{u_{5k}}$ was the selection coefficient for proportion of the
 265 resource within a 5-km buffer $x_{(u_{5k})}$, and γ_{0j} was the random intercept for animal j . We
 266 tested for collinearity using the variance inflation factor (Graham 2003) and used AICc

267 selection criteria to identify the most parsimonious model (Burnham and Anderson 2002)
268 within the global model and subset of simpler models (Table S1, Supplementary
269 material). We then used k-fold cross validation to evaluate the robustness of RSFs (Boyce
270 *et al.* 2002). An RSF model based on 80% of the data was estimated, withholding the
271 remaining 20% for evaluation. Predicted scores of the model were placed in ten bins of
272 equal size that represented the percentile range of predicted scores. We then determined
273 the frequency of locations in the withheld data (20%) that fell into each bin. To evaluate
274 model performance, we calculated a Spearman rank correlation (r_s) between the
275 frequency of occurrence for the withheld 20% and the ranked RSF-availability bins
276 (Boyce *et al.* 2002). The process was repeated 20 times and we report the average r_s . We
277 used the validated RSFs to build island-wide maps of relative occurrence probabilities,
278 which we used to estimate encounter risk with both predator species. RSFs were
279 calculated using R statistical software (ver. 2.15.0, R Development Core Team 2008)
280 with the package lme4 (Bates *et al.* 2014) and adehabitatHR (Calenge 2006).

281

282 **Forage model**

283 To study caribou use of vegetation-rich areas, we created a spatiotemporally dynamic
284 model of forage biomass (similar to Hebblewhite *et al.* 2008). We considered only food
285 items that have been found in caribou feces during the spring-summer period. This model
286 was based on the five habitat categories, and field vegetation surveys linked to a temporally
287 dynamic forage availability model using MODIS Terra NDVI 250 m every 16 days.
288 Complete details of this model are given in Supplementary material (Appendix S1,
289 Supplementary material).

290

291 **Statistical analysis**

292 For every set of caribou locations (observed, random, and mechanistic), we extracted
293 habitat category, relative probability of occurrence of black bears and coyotes, and
294 vegetation biomass based on timing of the location. We estimated selection for each
295 habitat by computing resource selection ratios (w_i) and tested for overall selection using a
296 Chi-square test (Manly et al. 2002). We assessed selection for vegetation at a given scale
297 by comparing the yearly between-individual average value of vegetation biomass of each
298 herd with the average value for the set of available locations based on the random and
299 mechanistic models. For locations representing use, confidence intervals around the
300 average provide an indication of individual variation. Similarly, we tested for avoidance
301 of predation by comparing the average probability of occurrence of bears and coyotes for
302 each herd at actual caribou locations with average availability observed from each of our
303 four models of availability.

304 Lastly, to gain insight into the behavioural motivation behind migration, we
305 assessed trade-offs between vegetation and predation faced by caribou at the second-
306 order level of selection. We used the following linear model:

307
$$Biomass(x) = \beta_0 + \beta_{Bear} * x_{Bear} + \beta_{Coyote} * x_{Coyote} + \beta_{Interaction} * x_{Bear} * x_{Coyote} \quad (2)$$

308 where $Biomass(x)$ represents the vegetation biomass in a given location, β_0
309 represents the intercept, $\beta_{Predators}$ represents the slope between the risk from a predator
310 $x_{Predators}$ and biomass. A positive and statistically significant coefficient β indicates that
311 caribou would face a trade-off between the specific cause of predation and forage. An
312 interaction between bear and coyote relative probability of occurrence was added to

313 account for the presence of non-linearity in the influence. We estimated this model using
314 the actual set of locations, but also using the availability models generated within the
315 general study area based on the random and mechanistically simulated models. We used
316 bootstrapping to get more robust standard error estimates for the two availability models,
317 since these models are biased due to arbitrary determined sample sizes. More precisely,
318 we performed these regressions with a sub-sample of the random and mechanistic
319 datasets of available locations, sampling the same amount of individuals as the actual
320 data (n=271 individual-years). We repeated these steps 1000 times and used the average
321 standard errors in confidence interval calculation.
322

323 **Results**

324 **Predator occurrence and vegetation abundance**

325 For both black bears and coyotes, AICc model selection showed that the global model
326 with all habitat categories and presence of functional responses was most parsimonious
327 (AICc weights > 0.99, Table S1). Both black bears and coyotes displayed a functional
328 response in habitat selection, where preference for most habitats decreased as the
329 proportion of a given habitat in the area surrounding a location increased as revealed by
330 the negative coefficient for interactions terms. This response was stronger for coyotes
331 than for bears in the selection of Barren and Wetland habitats (Table 1). The two
332 predators responded differently to elevation and slope, with black bears avoiding sites
333 with higher elevations, but selecting sites with steeper slopes, and coyotes displaying the
334 opposite pattern, with selection favouring higher elevation and low slope. K-fold cross-
335 validation indicated these models were robust, with $r_s = 0.979$ for black bears and $r_s =$
336 0.930 for coyotes.

337 During the same period, Wetland and Coniferous Open supported the highest
338 vegetation biomass, followed by Barren and Coniferous Dense (Table 2). Correlations
339 between increases in NDVI Modis Tera satellite index and vegetation growth were strong
340 (average conditional $R^2 = 0.346$). As revealed by the magnitude of the slopes, changes in
341 NDVI had the strongest impact on changes in vegetation growth in Wetland and Barren
342 habitats, while having smaller influence in Coniferous Dense (Table 2). Complete details
343 of the spatio-temporal vegetation model are given in Appendix S1.

344

345 **Habitat selection**

346 Based on the random model of availability, female caribou (except for Pot Hill) displayed
347 selection for Barren and Wetland habitats at both second- and third-order levels.
348 Conversely, caribou tended to avoid Coniferous Open and Dense stands as well as Water,
349 at both scales. Surprisingly, the Pot Hill herd displayed the opposite pattern, with
350 preference for Coniferous Open stands and general avoidance of other habitats at both
351 scales (Table 3). Patterns of selection were qualitatively similar to those from the
352 mechanistic sampling model, although the proportion of statistically significant selection
353 ratios across habitats decreased from 68% to 53%. This decrease in statistical significance
354 would lead to different inferences regarding selection due to the more conservative nature
355 of the comparison between used- and mechanistically defined availability locations.

356

357 **Response of caribou to forage and predation**

358 Three herds displayed selection for sites with higher forage when choosing their calving
359 grounds based on the 2nd-order mechanistic definition of availability; all herds displayed
360 selection based on the random 2nd-order model. All herds also displayed selection for
361 vegetation when moving within the calving grounds based on the 3rd-order random
362 model, but only one herd (Gaff Topsails) displayed selection based on the 3rd-order
363 mechanistic model. Interestingly, the mechanistic model of availability indicated greater
364 access to forage than the random model, a pattern that was consistent across scales. This
365 indicates that no matter where caribou were moving, interpatch movement rules were
366 already providing access to sites with greater forage, but that the choice of calving
367 grounds and core areas within caribou calving grounds reinforced this selection (Fig. 2).

368 Only two herds (Buchans and Gaff Topsails) appeared to reduce risk of
369 encountering coyotes by migrating to their calving grounds. These two herds and the
370 Middle Ridge herd were also able to further reduce risk when moving within their calving
371 grounds. Individuals from two herds (Buchans and Lapoile) appeared to reduce risk of
372 encountering bears when migrating to their calving ground but when considering
373 carefully their potential exposure based on their fine-scale movement (mechanistic
374 model), all herds except Pot Hill appeared to reduce predation risk from bears via second-
375 order selection. Three herds also enhanced risk reduction when choosing core areas
376 within calving grounds. In all cases, the mechanistic model of availability showed higher
377 risk of predation than the random model, indicating that fine-scale movements could
378 increase risk for caribou (Fig. 2).

379

380 **Trade-offs between predation risk and forage**

381 If areas with high forage availability are associated (positively correlated) with an
382 increased risk of predation, caribou will face a trade-off between the two. In general,
383 available locations with higher forage biomass based on the random or mechanistic
384 models were associated with higher risk of predation from both bears and coyotes
385 (positive coefficient, Fig. 3). However, caribou were also exposed to higher risk from
386 both predators in their actual use of habitat, most notably regarding the relationship
387 between foraging sites and black bear predation risk (Fig. 3).

388

389

390 **Discussion**

391 Using an extensive dataset of telemetry locations of caribou and their predators, we
392 studied broad-scale habitat selection of five caribou herds with an emphasis on the trade-
393 offs between food acquisition and predation risk. We found that caribou movements are
394 oriented mainly toward increased access to forage and also reduction of encounter risk
395 with bears, and to a lesser extent, coyotes. This was somewhat contrary to our original
396 predictions in that we expected third order selection would be driven by an avoidance of
397 predation risk. Our refined definition of habitat availability, based on a mechanistic
398 model of caribou movements, provided different insights into the food-predation trade-off
399 faced by caribou and allowed us to consider behavioural motivation as a driving level-
400 specific force behind habitat selection. The fact that this refined analysis revealed patterns
401 of forage selection and predator avoidance that would not have been revealed using more
402 traditional approaches, speaks to the subtle factors underlying caribou movements and the
403 need to better identify what is considered as 'available' in use-availability studies.
404 Ultimately, our findings reveal how prey can integrate multiple levels of selection to
405 balance the importance of predation risk on foraging behaviour.

406 Our results showed that most caribou herds selected calving grounds on the basis
407 of the foraging opportunities that they provide. This observation was reinforced by
408 evidence of habitat selection at the movement paths between habitat patches (as shown by
409 the difference in vegetation exposure between our two models of availability). Following
410 Rettie and Messier's (2000) hypothesis that a hierarchy in limiting factors matches the
411 hierarchy in habitat selection, it appeared that foraging access was likely to be an
412 important limiting factor for some herds during the critical period of calving and post-

413 calving, with the cost of lactation for ungulate females and associated increasing need in
414 foraging (Hamel and Côté 2008). However, predation risk avoidance was not as clear
415 given that some herds were more responsive to risk exposure than were others, perhaps
416 reflecting local differences in cause-specific predation risk across the broader caribou
417 population. For instance, recent coyote colonization in Newfoundland may explain why
418 caribou tended to display less avoidance of this predator. Lastly, we contend that our
419 approach offers transparent and conservative results regarding selection because the
420 analyses summarized individual selection and then pooled the individual responses into
421 herds rather than a more uniform (and less appropriate) multi-herd pooling.

422 The Buchans herd appeared to be the most effective at avoiding predation, which
423 is interesting given that it is the herd that undergoes the longest annual migration to
424 calving grounds (Mahoney and Schaefer 2002a). This suggests that migratory caribou
425 may face a trade-off between migration distance and its expected benefit in terms of
426 reduced predation risk and increased foraging opportunities (Gunn et al. 2012); such a
427 trade-off is likely to exist in terrestrial species given the high costs associated with
428 migratory behaviour (Alerstam et al. 2003). Considering the observed variability in
429 Newfoundland caribou migratory movements (Rayl et al. 2014), it appears that this trade-
430 off may lead to variable migratory behaviour across herds. Some ungulates such as elk
431 (*Cervus elaphus* L.) and caribou exhibit partial migration with some populations
432 migrating and others being sedentary (Festa-Bianchet et al. 2011, Hebblewhite and
433 Merrill 2009, Middleton et al. 2013), but results from Rayl *et al.* (2014) as well as those
434 herein reveal a likely gradient of migratory behaviour in Newfoundland caribou.
435 Bergerud et al. (2008) concluded that migration for caribou herds in North America was

436 associated with wolf (*Canis lupus* L.) avoidance because migrating females typically had
437 access to lower quality forage than sedentary males. In addition, elsewhere in caribou
438 range, movements away from tree line likely reduce risk of wolf predation (Bergerud et
439 al. 2008). However, the relatively small size of Newfoundland island may impose spatial
440 constraints on migrating caribou compared to other populations, thereby reducing their
441 ability to escape predation by wolves (historically) or other carnivores (currently).
442 Indeed, migration in Newfoundland caribou may have originated both as a predation- and
443 foraging-oriented behaviour, which is supported by the observed behaviour among
444 female caribou in this study, almost a century after wolves were extirpated from the
445 island.

446 During the past 50 years, caribou herds on Newfoundland have undergone marked
447 changes in abundance, with population sizes being notably low during the 1960s and
448 1970s, increasing rapidly during the 1980s to mid-1990s, and declining precipitously
449 following the mid-late 1990s (Mahoney and Schaefer 2002b, Mahoney et al. 2011).
450 These fluctuations seem to be driven by a combination of factors, including decadal
451 trends in winter severity, density-dependent nutrition during summer, and predation on
452 neonates (Bastille-Rousseau et al. 2013, Schaefer and Mahoney 2013). However, if
453 migratory behaviour or habitat selection are mismatched with current predation risk and
454 forage availability, then reductions in productivity and survival are expected
455 (Hebblewhite and Merrill 2011, Middleton et al. 2013). To date, this potential source of
456 caribou population decline in Newfoundland had yet to be fully tested.

457 Our results do not support this hypothesis but rather show that habitat selection is
458 driven by improved foraging opportunities and predation risk reduction, implying that

459 food has been limiting, at least during the period of decline (see Fryxell and Avgar 2012,
460 Fryxell and Sinclair 1988). It seems that fine-scale interpatch movements may have
461 increased caribou exposure to predation risk while also providing increased access to
462 forage. It is understood that most prey species, notably ungulates (Creel et al. 2005),
463 avoid forage-rich areas when such areas also confer higher risk (leapfrog effect; Laundré
464 2010, Sih 1998). Because Newfoundland caribou do not avoid such habitats, this
465 disconnect may explain why high calf predation seems to be the main proximate factor
466 limiting the Newfoundland caribou population (Mahoney and Weir 2009). It follows that
467 low calf survival ultimately may be driven by risk-prone foraging by parturient females
468 under high nutritional stress.

469

470 **Refining the definition of availability to study behavioural trade-offs**

471 Habitat selection studies usually describe an animal as using certain areas within a rather
472 specific and narrow set of rules. Yet, this approach can be problematic because it fails to
473 provide an appropriate mechanism explaining habitat use patterns relative to what is
474 actually available to the animal (Aarts et al. 2013). We showed how a mechanistic model
475 of availability, mimicking fine-scale inter-patch movements, can be used to study broad-
476 scale selection and thereby improve our understanding of how caribou trade off food
477 acquisition versus predation risk. Our mechanistic model allows us to draw inferences
478 about multiple and perhaps paradoxical motivations, as was evident by the revelation that
479 female caribou make habitat-related decisions on the basis of foraging opportunities
480 despite resultant increase in predation risk. Specifically, we would have missed that
481 caribou are able to adjust their movements to reduce bear predation risk; such an

482 interpretation would not have been possible in the absence of our mechanistic model,
483 since we would not have detected that the majority of the herds displayed bear avoidance.
484 Accordingly, we suggest that our model offers an improvement over the random model
485 by restricting habitat availability to areas that are potentially usable by an individual on
486 the basis of its movement decisions. Other approaches have been proposed in this vein
487 (see notably Avgar et al. 2013), but our approach is unique in that we used a mechanistic
488 model of movement capturing fine scale selection to study broader scale patterns.
489 Spatially-explicit modelling therefore allowed us to isolate the selection process
490 occurring at a specific level, clarifying inferences about the motivation behind selection
491 and providing a refined understanding of how caribou handle food versus safety trade-
492 offs across levels of selection. Therefore, we infer that this refined assessment of habitat
493 availability will open up additional opportunities for testing new hypotheses related not
494 only to predator-prey interactions but to the general behavioural process of habitat
495 selection in relation to the several competing behavioural motivations underlying such
496 selection.
497
498

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516

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- 680
- 681
- 682

683 Table 1. Mixed-effects RSFs for black bears and coyotes, Newfoundland, 1 May - 1
 684 August, 2008-2010. Parameter estimates (β), standard errors (SE), and variance estimates
 685 of the random intercept are presented.
 686

Variables	Black bear		Coyote	
	β	SE	<i>B</i>	SE
Water	-1.103	0.035	-1.395	0.050
Barren	-0.113	0.032	-0.257	0.049
Coniferous Dense (CD)	0.799	0.024	0.331	0.038
Coniferous Open (CO)	0.570	0.017	0.195	0.028
Other	0.641	0.051	0.288	0.091
Elevation	-0.178	0.010	0.081	0.023
Slope	0.243	0.009	-0.105	0.012
Wetland within 5 km	0.050	0.014	-0.248	0.022
Barren within 5km	0.112	0.013	-0.255	0.023
Coniferous Open within 5km	-0.031	0.017	-0.345	0.025
Coniferous Dense within 5km	-0.074	0.016	-0.552	0.023
Wetland within 5km * Wetland	-0.146	0.014	-0.135	0.029
Barren within 5km * Barren	-0.264	0.020	0.023	0.030
Coniferous Open within 5km * CO	0.012	0.014	-0.076	0.023
Coniferous Dense within 5km * CD	-0.159	0.015	-0.350	0.032
Random effect	Variance: 1.168		Variance: 1.506	

687

688

689 Table 2. Relative abundance indices of vegetation biomass from vegetation surveys by
690 landcover type. The slope and coefficient of determination (conditional R²) represent the
691 relationship between vegetation biomass and NDVI values. See Supplementary
692 Information 1.
693

Habitat	Index of biomass	Slope	Conditional R ²
Barren	0.734	0.908	0.389
Wetland	1.000	0.912	0.465
Coniferous Open	0.990	0.902	0.380
Coniferous Dense	0.458	0.821	0.151

694

695

696 Table 3. Selection ratios (\pm 95% CI) of 134 caribou from five caribou herds,
697 Newfoundland, 1 May - 1 August, 2007-2010. Interpretation of selection ratios are
698 relative to one: Values higher than one indicate selection for a given resource whereas
699 values lower than one indicate avoidance of the resource. Selection ratios were computed
700 at two different levels: second-order and third-order (within calving grounds), using the
701 general random model of availability and a mechanistic model of availability. Chi² values
702 indicating overall presence of selection are also given. Statistically significant values
703 ($\alpha=0.05$) are presented in bold.

Random model

Herd	Level	Water	Barren	Wetland	Coniferous		Other	Chi2
					Open	Dense		
Buchans	3rd-order	0.264 \pm 0.178	1.971 \pm 0.376	1.531 \pm 0.262	0.700 \pm 0.129	0.622 \pm 0.276	0.929 \pm 2.238	82.476
	2 nd -order	0.242 \pm 0.163	3.366 \pm 0.642	1.700 \pm 0.291	0.603 \pm 0.111	0.652 \pm 0.290	0.251 \pm 0.605	157.178
Lapoile	3rd-order	0.308 \pm 0.201	1.377 \pm 0.278	1.236 \pm 0.222	0.971 \pm 0.154	0.458 \pm 0.275	0.87 \pm 1.244	37.046
	2 nd -order	0.257 \pm 0.167	3.103 \pm 0.626	1.589 \pm 0.285	0.738 \pm 0.117	0.368 \pm 0.221	0.71 \pm 1.015	130.663
Middle Ridge	3rd-order	0.414 \pm 0.191	2.215 \pm 0.862	1.492 \pm 0.219	0.817 \pm 0.121	0.877 \pm 0.445	2.724 \pm 3.158	51.602
	2 nd -order	0.497 \pm 0.229	1.027 \pm 0.400	2.061 \pm 0.303	0.809 \pm 0.120	0.509 \pm 0.258	1.075 \pm 1.247	68.687
Pot Hill	3rd-order	0.337 \pm 0.236	0.488 \pm 0.892	0.680 \pm 0.269	1.171 \pm 0.071	0.826 \pm 0.322	1.225 \pm 2.171	24.023
	2 nd -order	0.224 \pm 0.157	0.051 \pm 0.093	0.431 \pm 0.171	1.634 \pm 0.099	0.832 \pm 0.324	0.463 \pm 0.820	133.777
Gaff Topsails	3rd-order	0.147 \pm 0.138	1.119 \pm 0.288	1.372 \pm 0.250	1.016 \pm 0.142	0.898 \pm 0.348	0.392 \pm 0.804	42.535
	2 nd -order	0.125 \pm 0.118	2.117 \pm 0.545	1.558 \pm 0.284	0.861 \pm 0.121	0.843 \pm 0.326	0.346 \pm 0.710	81.911

Mechanistic model

Herds	Scale	Water	Barren	Wetland	Coniferous		Others	Chi2
					Open	Dense		
Buchans	3rd-order	0.663 \pm 0.445	2.137 \pm 0.408	1.493 \pm 0.255	0.605 \pm 0.112	0.572 \pm 0.254	1.031 \pm 2.486	79.488
	2 nd -order	0.601 \pm 0.404	3.412 \pm 0.651	1.610 \pm 0.275	0.533 \pm 0.098	0.630 \pm 0.280	0.315 \pm 0.760	144.079
Lapoile	3rd-order	0.633 \pm 0.412	1.396 \pm 0.281	1.180 \pm 0.212	0.866 \pm 0.138	0.457 \pm 0.274	0.954 \pm 1.364	21.728
	2 nd -order	0.638 \pm 0.416	3.144 \pm 0.634	1.506 \pm 0.27	0.653 \pm 0.104	0.356 \pm 0.214	0.892 \pm 1.275	115.613
Middle Ridge	3rd-order	1.005 \pm 0.464	2.289 \pm 0.891	1.420 \pm 0.209	0.703 \pm 0.104	0.877 \pm 0.445	3.819 \pm 4.428	42.600
	2 nd -order	1.232 \pm 0.568	1.041 \pm 0.405	1.952 \pm 0.287	0.715 \pm 0.106	0.492 \pm 0.250	1.352 \pm 1.567	62.433

Pot Hill	3rd-order	0.882 ± 0.617	0.561 ± 1.026	0.673 ± 0.267	1.080 ± 0.065	0.862 ± 0.336	1.622 ± 2.876	6.656
	2 nd -order	0.556 ± 0.389	0.051 ± 0.094	0.408 ± 0.162	1.444 ± 0.087	0.805 ± 0.314	0.582 ± 1.031	90.992
Gaff Topsails	3rd-order	0.356 ± 0.335	1.148 ± 0.295	1.312 ± 0.239	0.898 ± 0.126	0.864 ± 0.334	0.448 ± 0.919	15.654
	2 nd -order	0.311 ± 0.292	2.145 ± 0.552	1.476 ± 0.269	0.761 ± 0.107	0.815 ± 0.315	0.435 ± 0.893	52.952

704

705

706 **Figure legends**

707 Figure 1. Calving grounds and the larger study area for five caribou herds in
708 Newfoundland, Canada. These delineations were used to define habitat availability.
709

710 Figure 2. Average (\pm 95% C.I.) exposure to forage biomass, coyote encounter risk and
711 bear encounter risk for female caribou from five herds, Newfoundland. Actual exposure
712 (Use) is compared to availability represented by two scales of movement: (i) Selection of
713 a calving ground (2nd-order)) and (ii) movement within calving-ground (3rd-order)).
714 Availability at each scale was also defined using two approaches; (i) a random model
715 (Random) and (ii) simulated locations based on mechanistic modelling of fine-scale
716 movement (Mechanistic). Overall, selection is inferred when use is higher than
717 availability while avoidance is inferred otherwise.

718

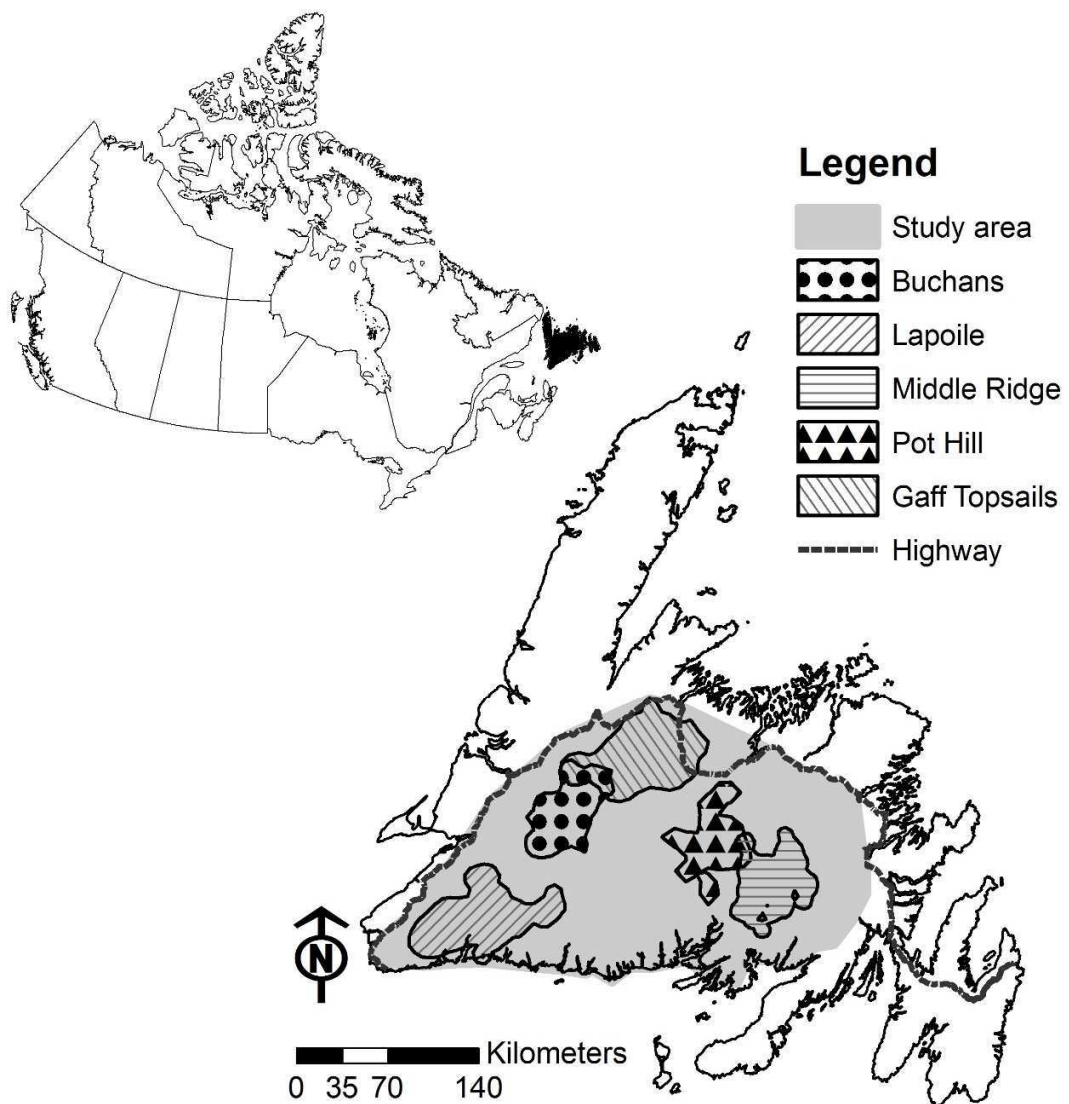
719 Figure 3. Linear models between vegetation biomass and predation risk from bear, coyote
720 and the interaction of the two showing potential trade-offs for caribou when choosing
721 their calving grounds. Models estimates are given with 95% confidence intervals. Models
722 were performed using the actual caribou locations (Use), but also using random sample of
723 availability (Random) within the study area and a simulated sample based on a
724 mechanistic model (Mechanistic) of fine scale movement for caribou.

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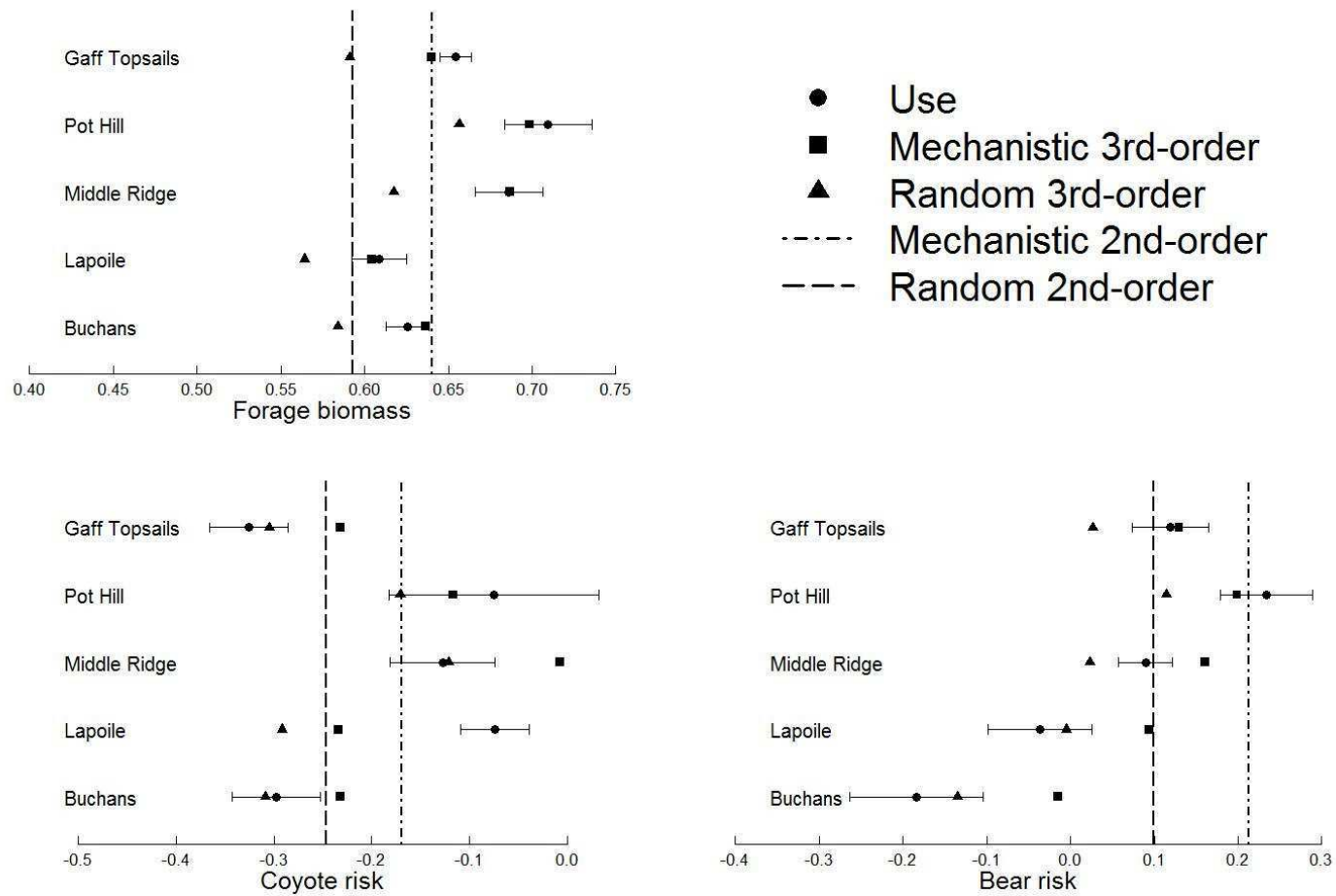
727 **Figure 1**

728



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730



731 **Figure 2**

Figure 3

