

## **Bio-cultural heritage and biodiversity: emerging paradigms in conservation and planning**

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# *Bio-Cultural Heritage & Biodiversity - emerging paradigms in conservation and planning*

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**Abstract** Long-term studies across Europe have clarified the eco-cultural nature of landscapes and their biodiversity, and the importance of bio-cultural heritage. This raises issues of the nature of ‘*Nature*’ for example, and of how perceptions of ‘*natural*’ landscapes may be misleading. Indeed, the lack of understanding of how ecological systems and their biodiversity relate to the cultural nature of landscapes is hugely problematic. Whilst wilder ‘*futurescapes*’ offer many benefits, the underlying concepts frequently confuse abandonment of ‘*eco-cultural landscapes*’ with ‘*re-wilding*’. The ending of traditional and customary uses and utilisation of landscapes mistakenly seen as re-naturing or re-wilding, and inherently a beneficial change, may threaten the conservation of important bio-cultural heritage. The reality of landscape heritage is that much biodiversity relates to long-term, predictable, sustainable, traditional uses. The ending of such traditions has now happened in many regions and taking place rapidly across much of Europe. Sudden, dramatic and often unexpected changes occur and massive declines of biodiversity result.

With environments transformed by human activity, the eco-cultural landscapes in traditional or customary management hold much of the most significant wildlife resources. The ending of traditional and customary management, termed ‘*cultural severance*’ (Rotherham 2009), is probably the most serious threat for nature conservation in the twenty-first century, at least in the medium-term, the impacts exceeding those of climate change. The transformations now happening also have major implications for rural human communities and their economies.

Observational studies and cross-disciplinary research across Europe highlights the urgent need to recognise the eco-cultural nature of landscapes and to establish inventories and conservation programmes for important bio-cultural heritage. This paper results from long-term historical research, scientific analysis of case studies, and international researcher collaborations to present ideas and paradigms relating to emerging concepts and visions.

**Key words** bio-cultural heritage, biodiversity, eco-cultural landscapes, cultural severance, traditional and customary management

## **Introduction**

A renewal of interest in new ways to address conservation problems through radical, novel approaches followed seminal texts by Adams (2003), Taylor (2005), and Vera (2000). ‘*Wilding*’ and ‘*wilder*’ landscapes, applied effectively and sensitively, offer huge, exciting benefits for biodiversity, bio-cultural heritage, and amenity. However, there are significant pitfalls if implementation lacks a broad, multi-disciplinary approach, with careful planning and design. The ‘*eco-cultural nature*’ of landscape (Rotherham 2014a), resulting from long-term, intimate interactions between people and ecologies is important. Often the interplay of humanity with nature creates the construct of ‘*place*’ and of local distinctiveness (e.g. Westland (ed.) 1997). Across Europe in particular, twenty-first century depopulation means rural landscapes ‘*abandoned*’ not ‘*wilded*’, with ecology, communities and economies potentially devastated. Alongside urbanisation of rural landscapes, these socio-economic and demographic changes cause ‘*cultural severance*’ (Rotherham 2008, 2013b), and this leads to long-term, often rapid, loss of biodiversity and landscape quality. Furthermore, from urban to

52 remote, rural areas, attitudes to, and perceptions of, 'alien' invasive species challenge to  
53 attempts to 'wild' the landscape. Feral species, exotic plants and animals, and invasive  
54 natives forming recombinant biodiversity (Rotherham & Lambert 2001; Rotherham 2014a),  
55 but 're-wilding' discussions rarely mention feral and exotic. Current thinking may even place  
56 a positive spin on a future with invasive, alien species (e.g. Pearce 2015). Central to  
57 'futurescapes' and 're-wilding' are ideas and perceptions of 'wild', 'wildness', 'wilderness',  
58 'nature', and 'natural', and importantly there is an imperative need to both learn from bio-  
59 cultural heritage and to seek to conserve and safeguard what remains. The implications and  
60 scale of human impacts on nature were raised by authors such as Rachel Carson in *Silent*  
61 *Spring* (1962), with a legacy as discussed by Jameson (2012), and then more recently by, for  
62 example, McKibben (1990), in *The End of Nature*. Yet, despite these major contributions to  
63 debates on nature and humanity, the eco-cultural landscape and its bio-cultural heritage often  
64 remain overlooked or misunderstood. McKibben (1995) presented ideas and suggested  
65 solutions to remediate adverse human impacts, and it is perhaps from some of these and other  
66 similar writings, that the ideas of 'wild' have emerged. Fundamental drivers within these eco-  
67 cultural landscapes have frequently been misunderstood (e.g. Hardin 1968; Ostrom 1990;  
68 Appell 1993).

69

## 70 **Bio-cultural resources**

71

72 Bio-cultural heritage, both tangible and intangible, is threatened with loss and decline on a  
73 scale that is unprecedented and it is widely recognised that traditional nature conservation is  
74 failing (Adams 2003; Rotherham 2014a). Within eco-cultural landscapes, the heritage  
75 associated with ancient woods, historic forests and long-lived trees provides a good example  
76 of the types of issues and problems (Muir 2005; Fowler 2002; Rackham 1976, 1980, 1986,  
77 1996; Hayman 2003; Rotherham et al. 2012; Smout 2000). The remarkable resources of  
78 ancient trees, including open-grown veterans such as the Major Oak in Sherwood Forest  
79 (Nottinghamshire, England), the fuelwood pollards at Burnham Beeches (Buckinghamshire,  
80 England) or the old coppice limes of Whitwell Wood (Derbyshire, England), provide  
81 examples of irreplaceable bio-cultural heritage. In many cases, these trees and associated  
82 anthropogenic features in their landscapes are many centuries old; in some instances, such as  
83 the ancient lime coppices, they may be a thousand years or more.

84

85 [Figure 1]

86

87 An example of the type of threats and issue for such unique heritage was the loss of the Great  
88 Oak of Pontfadog, in Wrexham, North Wales, which blew down on 18<sup>th</sup> April 2013. The  
89 National Tree of Wales, estimated to be somewhere between 1,181 years and 1,628 years old,  
90 was a culturally modified veteran; a biological feature but altered by human usage over  
91 centuries. The tree could have survived for many more centuries with just a little help in  
92 terms of supporting wooden struts to help bear the load. Long-since isolated in its landscape  
93 as the forest in which it belonged had been cleared, the tree, though iconic, was allowed to  
94 collapse. Over northern parts of Britain, there are so-called 'orchards' of veteran pollard trees  
95 aged perhaps between 400 years and 600 years in age and managed historically to provide  
96 leaf fodder for domestic animals. Neglected for around 150 years since the practices largely  
97 died out, surviving re-grown pollards are top-heavy and vulnerable to catastrophic collapse  
98 and the same fate as the Pontfadog Oak. The first steps in rational, planned conservation, are  
99 to survey and audit the tree resources.

100

101 Writers in Cronon (ed) (1996) considered the interface between people and environment in  
102 relation to the human role in nature. As Cronon notes, 'Nature is not nearly as natural as it  
103 seems', and he examines the relationships in North America in Cronon (2003). Furthermore,

104 authors such as Hoffman (2014) demonstrate the cultural influence in landscapes, in for  
105 example, Europe. The importance of the human-nature interaction in forming landscapes and  
106 even national identity was highlighted by, for example, Schwartz (2006) in the case of post-  
107 communist Latvia. Rackham (1996) and Grove, & Rackham (2003), considered the cultural  
108 importance of Mediterranean landscapes, and Smout (for example Smout 2000), describes the  
109 human influence in Scottish environments. However, as noted by Agnoletti et al. (2007),  
110 even though academic interest has grown in terms of identifying and valuing cultural  
111 landscapes and traditional management practices, there is only limited, integration of this in  
112 policy and landscape planning. Currently, few countries have effectively addressed the  
113 complex processes underlying landscape bio-cultural heritage and have established  
114 appropriate conservation methods and guidance. There are specific moves for example with  
115 forest management in France (Agnoletti et al. 2007), where forest landscapes and traditional  
116 forest techniques are considered '*heritage*' to be conserved. In Italy too, there has been  
117 significant progress by the Italian government and regional administrations such as in  
118 Tuscany, in the recognition and then conservation of unique cultural landscapes such as the  
119 chestnut groves (Agnoletti 2008, 2013). The bio-cultural resource of the landscapes, the  
120 ancient trees, and the associated infrastructure of human features such as drying barns and  
121 trackways, may then be managed as an integrated whole. Similarly, projects in both Austria  
122 (Johann 2013), and in Switzerland (Burgii & Stuber 2013), have raised awareness of the  
123 heritage and the issues, and some steps are being taken to achieve management which is more  
124 sustainable. However, major challenges still arise at national and European levels, and much  
125 planning and practical management neglects or even damages bio-cultural heritage  
126 (Rotherham 2014a). Integrating, preserving and enhancing social and cultural dimensions of  
127 sustainable landscape management remains problematic and austerity measures across the  
128 continent since 2008 have compounded the situation (Rotherham 2014a).

129  
130 In order to change approaches to become more sustainable, it is important to raise awareness  
131 and to grow education and research in key fields relating to bio-cultural heritage and its  
132 conservation and that of intangible cultural heritage (i.e. oral traditions, traditional landscape-  
133 related knowledge). This was raised at workshops and conferences such as Forestry and Our  
134 Cultural Heritage (Sunne 2005), IUFRO All Division 9 Conference Sarajevo (2012), Linking  
135 Biological and Cultural Diversity in Europe, Florence (2014), Frontiers in Historical Ecology,  
136 Zurich (2011), and The End of Tradition?, Sheffield (2010). An aim has been to encourage  
137 appropriate approaches to be included in landscape management strategies. To this end, more  
138 countries should investigate the complex processes relating to bio-cultural heritage,  
139 conservation methods, and principles. Additionally, there is urgent need to collate inventories  
140 of landscape-related cultural sites and bio-cultural heritage resources in pragmatic and usable  
141 ways (such as GIS, GPS, LIDAR topography datasets, and databased information). These  
142 databases can then be used to assist and guide site managers to avoid or at least minimise  
143 adverse impacts of site works on bio-cultural assets. (Rotherham & Ardron developed a  
144 pioneering approach to this process in wooded landscapes for Ecclesall Woods in  
145 Sheffield, UK; see Rotherham 2007a, and Rotherham & Ardron 2006).

146  
147 [Figure 2]

148  
149 There is a need to recognise the issues for conservation of bio-cultural resources and then to  
150 apply the developed approaches more widely across the landscape. In some regions, such as  
151 the Alps for example, relevant policies are already in place. The declaration, *People and*  
152 *Culture*, as part of the *Alpine Convention* (2009) and its protocols, drew attention to people-  
153 centred, sustainable development policies focussing on the needs, desires, and views opinions  
154 of the communities in the region. Issues around the conservation and protection of both  
155 tangible and intangible cultural heritage, and the maintenance and development of physical

156 and non-physical cultural heritage, including traditional knowledge, were considered in the  
157 convention. In particular, the document addressed traditional methods of landscape  
158 management, architectural, built and artistic heritage, traditional forestry, rural crafts, and  
159 related industrial activities. It is important that research be undertaken to highlight, for  
160 example, existing tangible and intangible cultural heritage of traditional knowledge. This is in  
161 order that practical conservation is better-informed, so bio-cultural heritage is more  
162 effectively safeguarded and maintained in the future.

163

164 However, it is necessary to recognise also, that perceptions and ideas relating to ‘*wild*’ and to  
165 ‘*nature*’ have changed and evolved over time (Oelschlaeger 1991; Giblett 1996). They also  
166 vary from place to place, and from region to region.

167

## 168 **Dilemmas & contentions**

169 A major issue with respect to Europe’s traditional cultural landscapes is their dependence on  
170 varying degrees of human impact, in many cases, traditional or customary rural practices.

171 Many of these landscapes include richly diverse areas of significant local and regional  
172 distinctiveness, and with associated high levels of biodiversity. These landscapes have  
173 evolved slowly over time. They reflect and often depend on the imprint of human utilisation  
174 in traditional or customary ways. When existing land uses are extensified or abandoned,  
175 traditional cultural landscapes may be displaced by spontaneous ecological successions with  
176 resulting vegetation and landscape change. However, this severance of human customary  
177 utilisation is often greeted as a positive phenomenon, and as re-wilding (Rotherham 2014b).  
178 Yet mostly, this is not ‘*wilding*’ but abandonment and dereliction (Rotherham 2009, 2014b).  
179 Interestingly, these changes are happening at a time when the provision of benefits, so-called  
180 ‘*ecosystem services*’ by nature to humanity, are finally being recognised and valued (e.g.  
181 Juniper 2013, 2015; Maier 2012).

182

183 However, in terms of landscape management and bio-cultural resources, conservation is on  
184 the horns of a dilemma. Whilst abandonment is bad, the converse is also true in that  
185 intensified human utilisation especially with macro-mechanisation and industrial agriculture  
186 or forestry, displaces traditional landscapes and imposes simplified ones of limited ecological  
187 diversity. At local, regional and national scales, human cultural values and attachments, bio-  
188 cultural resources and assets, and associated biodiversity are progressively degraded. Such  
189 change can also affect future rural economic function, through for example, reduced tourism  
190 and leisure value. As traditional rural economies decline, and governments or planners seek  
191 to replace or at least supplement agriculture and forestry with tourism income, the value of  
192 the leisurely landscape is expected to grow. In this context, the particular character of the  
193 local or regional cultural landscape requires investigation in terms of economic development  
194 through tourism expansion. This in turn might help engender the conditions necessary to  
195 promote valorisation and conservation of bio-cultural. History, tradition, tangible and  
196 intangible heritage, and local cultural knowledge combine as regional identity factors and  
197 potentially a distinctive tourist product. However, the major challenge is to convert economic  
198 and social benefits, for tourism businesses, into support for local communities to maintain  
199 locally traditional land uses. The risk is that tourism develops as a parasitic business  
200 opportunity and fails to support and enhance local traditional values (Rotherham 2008).

201

202 Bio-cultural heritage and assets exist within the ‘*cultural landscape*’ of the physical and  
203 natural environment interacting with and modified by human activities. The term the ‘*eco-*  
204 *cultural landscape*’ (Rotherham 2008, 2013b, 2013d, 2014a), refers to the intimate  
205 relationships between people and nature which form ‘*landscapes*’ over long periods. The  
206 cultural landscape has a duality of meaning in relation to lands, which themselves have been

207 altered by human activity, and to interactions between people seeking to subsist within a  
208 physical setting in which they live and work. Over time, such human activities shape the  
209 landscape and modify its form to become a significant factor in regional or local character,  
210 and to provide communities with their own identity. Such shared identities generate feelings  
211 of unity between, for example, localities at human, economic, spatial, and cultural levels. The  
212 images generated may result in the external projection and presentation of a shared. Today,  
213 this may attract investment and promote both external and internal networks of co-operation  
214 and coordination; important factors in tourism development (Capriello & Rotherham 2008).  
215 To better promote and exploit these benefits requires recognition, enhanced understanding,  
216 and above all, effective conservation of the underpinning bio-cultural heritage. Corporate  
217 awareness of these cultural and environmental phenomena is necessary in order to exemplify  
218 and diffuse local and regional identities expressed through shared cultural and historic roots.  
219 These are imprinted into landscapes as patterns of land-use and spatial occupation. Since the  
220 1980s, such regional adaptations have been recognised increasingly in terms of significance  
221 for indigenous knowledge, sustainable development, and hence for nature conservation and  
222 agriculture. In this respect, there have been programmes, projects, policies, and strategies to  
223 grow synergies of traditional insight and modern scientific knowledge to explore solutions for  
224 shared problems. Despite this, as observed by Agnoletti et al. (2007), development based on  
225 sound scientific and historic understanding of cultural landscapes is still lacking across most  
226 of Europe.

227

228

## 229 **Methodologies**

230

231 This paper draws on the findings of long-term observational studies of key habitats and their  
232 wildlife; long-term observational studies of rural economies and tourism; analysis of long-  
233 term trends of political, social, economic and ecological influences in the landscape.  
234 Essentially, the paper is a review of much current thinking on issues of '*futurescapes*', on  
235 bio-cultural heritage, on eco-cultural landscapes, and on matters of cultural severance and re-  
236 wilding.

237

238 The work addresses relationships between traditional and customary management of natural  
239 resources and combines with the findings of long-term observational studies to highlight the  
240 connection between management and ecology (e.g. Rotherham 2007a, 2008, 2013b, 2014a).  
241 In terms of environmental context, an historic time-line has been generated to support the  
242 concept that many key wildlife habitats and their ecologies have descended from analogues in  
243 the primeval European environment (Vera 2000, 2005, 2009; Rotherham 2014a).  
244 Importantly, when considering nature conservation, biodiversity, bio-cultural heritage and  
245 sustainability, these unique communities, which have been maintained in traditional  
246 landscapes. This is the essence of the eco-cultural nature of the landscape as the space  
247 occupied by bio-cultural heritage.

248

249 The methodological approach involves the creation of the conceptual timeline from analysis  
250 and primary observation, and populate of the resulting tapestry of landscape mosaics by case  
251 studies at local and regional levels. Ecological and social changes at spatial levels then reflect  
252 the drivers of environmental, economic, and political factors. The timeline approach provides  
253 a framework to enhance understanding and to facilitate future actions.

254

## 255 **Results & Discussion**

256 Recent studies (Rotherham et al. 2012; Rotherham 2012, 2013a & b, 2014a) have sought to  
257 place the ideas and concepts of Frans Vera (Vera 2000, 2005, 2007, 2009), into an ecological  
258 and historical timeline. The basic ideas of the timeline were presented in Rotherham (2014a),  
259 and it is suggested that the biodiversity and ecological characteristics of the open and fluid  
260 landscape suggested by Vera for primeval Europe, are present in the modified land use  
261 patterns of the medieval period. In England, the act of Commons in 1235 is suggested as a  
262 watershed moment in the ‘fixing’ of rural patterns and their habitats (Rotherham 2012). Plant  
263 and animal communities present in the mosaic landscape suggested for the Vera model are in  
264 part taken into medieval land use patterns and both maintained or modified by traditional and  
265 customary usage. The major losses of diversity of ecosystems and of associated species then  
266 follow much later with the processes of enclosure and improvement, of urbanisation, and  
267 industrialisation, both urban and rural. A major trigger for the decline of traditional  
268 landscapes and ecologies is the advent of large-scale energy subsidies from fossil fuels, firstly  
269 mineral coal and then oil. For the first time in history, human communities were de-coupled  
270 from their local environmental resources, with catastrophic implications for bio-cultural  
271 heritage.

272

### 273 **Misunderstood cultures & overlooked impacts**

274 The occurrence and especially the extent of traditional, customary, ‘*cultural*’ use of natural  
275 resources and the consequent impacts on landscape and ecology of often overlooked  
276 (Agnoletti 2006, 2007; Rotherham 2008, 2014a). Indeed, the interpretation of the naturalness  
277 of landscapes and ecologies has frequently been misjudged and sometimes on a colossal  
278 scale. For example, major landscape-scale effects, such as the role of medieval peat cutting to  
279 supply Norwich and other areas with fuel, in the formation of England’s Norfolk Broads, was  
280 overlooked until the 1950s (Lambert et al. 1961). Ecosystems and landscapes are considered  
281 ‘*free-willed*’, natural and wild even when they are deeply eco-cultural and features are often  
282 considered to be ‘*natural*’ (Rotherham 2014b).

283

284 [Figure 3]

285

286 An example of such misunderstanding is in the British uplands where mountain and  
287 moorland are deemed wild and natural landscapes, whereas in reality they have been affected  
288 by grazing and other farming activities, often by industry, and especially by fuel utilisation  
289 such as peat turbarry (Ardron 1999; Rotherham 1999a, 2005). Rotherham et al. (1997, 2004)  
290 described how medieval and later peat cutting transformed many British upland landscapes.  
291 Similarly, ancient woodlands are rated as some of Britain’s most valued conservation sites  
292 (Rackham 1986, 1989, 2006; Peterken 1981), and yet their histories and eco-cultural natures  
293 are widely misunderstood. Management as coppice for fuelwood, charcoal, whitecoal, and  
294 pit-props, is often forgotten and the transformed soils and vegetation are considered to be  
295 natural phenomena (Rotherham 2007b; Perlin 1989). Key drivers of change leading to  
296 contemporary ecologies are unseen. Furthermore, the implications for important bio-cultural  
297 heritage following cultural severance of wooded landscapes and their modern emergence as  
298 ‘*leisurely landscapes*’ are ignored. Much contemporary site management is based on limited  
299 understanding of history and heritage and a misplaced view of ecology; and this applies  
300 across a wide range of vegetation and habitat types.

301

302 Medieval woods, heaths, commons, and bogs across western and Mediterranean Europe  
303 supplied most people with fuel, building materials and food over many centuries (Rackham,  
304 1980, 1986, 1990, 2003; Warde 2005). Additionally, along with providing energy for  
305 domestic use, medieval coppice woods and peatland turbarries, also fuelled early industry.

306 These sometimes-intensive uses had major impacts on soils and vegetation and often  
307 transformed landscapes, and different fuels (mineral coal, wood, charcoal, and peat or turf),  
308 varied in their effects. Although some of these changes were major determinants in the  
309 landscape, their impacts are rarely considered (Rotherham et al. 2004; Rotherham 2005).

310  
311 The multiplicity of landscape uses over history made many of these areas contested spaces  
312 with different, rival stakeholders and actors vying for resource rights. In England there are  
313 descriptions of conflicts over resource use, with for example, medieval iron masters accused  
314 of destroying woods and affected both local, common usage and the Crown's need for  
315 shipbuilding timber (Rotherham & Egan 2005). Indeed, to the untutored eye a wood managed  
316 as '*sustainable*' coppice appears devastated, whereas re-grown high forest seems pristine and  
317 almost primeval, '*ancient*' woodland or old growth forest; both misunderstanding of the  
318 landscapes, their ecologies, and their histories. Today, such woodlands are believed to be  
319 either modern plantations or even remnants of primeval '*wildwood*'; both equally  
320 misconceived ideas. Medieval and later parklands, often descended from ancient wood  
321 pasture, were often contested spaces (Harding, & Rose 1986; Liddiard 2003), and even  
322 recently, were considered of limited conservation significance.

323

324

### 325 **Landscape utilisation**

326 Subsistence use and later industrial exploitation of landscape resources over centuries  
327 generates character-defining change and this varies at local and regional levels. Particular  
328 environmental conditions and resources, and the drivers of economic, political and social  
329 forces, together with interaction and competition, are vital determinants in land-use and  
330 landscape. Eco-cultural landscapes show continuums often punctuated by crises for  
331 community and environment (Rotherham 2005). Interactions between community, resource  
332 utilisation, and environment drove the development of the landscapes we inherit today but the  
333 fundamental relationship, its sophistication and totality, is rarely appreciated. Subsistence  
334 communities with traditional and customary uses interacted intimately with the environment;  
335 medieval landscapes being like the traditional family pig, with everything used except the  
336 squeak. Almost all landscapes were modified, some were transformed, and others, especially  
337 during social, economic, or environmental crises, were devastated, but the intimacy and the  
338 totality of past utilisation are things, which now elude us. Like strangers in a foreign land  
339 (Lowenthal 1985), as visitors in our leisurely landscapes, we gaze into the past unfettered by  
340 knowledge or insight.

341

### 342 **Local, utilitarian subsistence**

343

344 European traditional agrarian, early industrial or subsistence communities depended on the  
345 local environment for most of their immediate resource needs. Indeed, until relatively  
346 recently only the affluent in more sophisticated economies had access to imported or luxury  
347 goods. Local communities and their individuals or households, for essential arable, pasture,  
348 fuel and building materials, relied on the limited resources of land, which they owned, or for  
349 which rights were held in common. Traditional and customary utilisation developed over  
350 centuries, which by medieval times, had a sophisticated web of cultural customs, legal rights  
351 and restrictions controlling them. The local and regional systems of usage adapted to  
352 particular conditions and generally provided the effects of changing populations, vagaries of  
353 weather, and for the impacts of other catastrophes such as disease. Customary rights  
354 developed in order to protect shared resource rights within communities whose stakeholders  
355 were not equals. Furthermore, in the pre-industrial and pre-urban society, if these controls



356 failed, and the system was unreliable and unsustainable, the community was at risk; people  
357 were deeply immersed in their local environment (Rotherham 2013b).

358  
359 The long-term impacts of such uses on environments and ecology are etched indelibly into  
360 pre-industrial landscapes and today have major implications for future management. Indeed,  
361 understanding the implications of land-use and long-term human influences (both drastic and  
362 subtle), on soils, water, and vegetation are essential if a futurescape vision is to be rooted in  
363 reality. Yet as noted earlier, much contemporary site management for nature conservation is  
364 not informed by any in-depth knowledge of past uses. Furthermore, the implications of  
365 cultural severance and the problems for bio-cultural heritage with intensified or abandoned  
366 systems in unappreciated. Abandoned pastoral landscapes for example, with rapidly declining  
367 biodiversity, deterioration in aesthetic qualities, and loss of indigenous local communities, are  
368 greeted as the expansion of '*forest landscapes*' and an inherently '*good*' thing (Rotherham  
369 2014b). Much of this management (or lack of management), does not pass any test of  
370 sustainability for environmental, economic or social values, and yet this is overlooked or  
371 ignored. However, with the interrogation of ecological systems at landscape level with  
372 understanding of cultural severance implications challenges current conservation  
373 management.

374

### 375 **Examples of British landscape impacts in uplands & lowlands**

376 The effects of customary landscape utilisation can be assessed for particular resources,  
377 products or materials, like fuel, or foodstuffs. For example, many British upland landscapes  
378 evolved over centuries through influences of environmental factors and human exploitation  
379 for fuel and grazing. By the 1800s, management for game, particularly red grouse and deer,  
380 affected many areas. Subsistence utilisation was ended, in England by Parliamentary  
381 enclosures, and in the Highlands of Scotland, by the '*Clearances*', both through the  
382 eighteenth and nineteenth centuries, and both leading to cultural severance. These became  
383 strongly contested spaces with hotly disputed resources.

384

385 The processes of exploitation and utilisation transformed vegetation and soil, changing  
386 landscape colour and texture large areas of the British uplands. Wetland drainage combined  
387 with removal of deep layers of thick peat or shallower organic turf, reduced water-holding  
388 capacity and changed water behaviour in vast areas of landscapes in catchments below the  
389 upland zones.

390

391 [Figure 4]

392

393 Lowland areas of England, like the Humberhead Levels, the Cambridgeshire Fens, the  
394 Norfolk Broads, and the Somerset Levels, were progressively changed throughout the  
395 medieval period (Rotherham 2013c). There was exploitation and drainage of marginal  
396 peatlands and conversion to agriculture. In the Norfolk Broads, peat removal to fuel  
397 commoners and ecclesiastical centres created vast, open turbaries, but by the late medieval  
398 these were being inundated by floodwaters and industrial exploitation was abandoned.  
399 However, the dramatic changes came from the 1600s onwards across the other regions with  
400 huge drainage and land improvement schemes, displacement of local common rights and of  
401 commoners, and the removal of almost all the peat resources. By the late nineteenth and into  
402 the twentieth century, the drivers in these lands were industrial, modern farming and  
403 industrial harvesting of the remaining deep peats. In many areas, removal of peatlands was so  
404 complete that even their memory was quickly lost. Landscapes, ecologies and communities  
405 were transformed beyond recognition (Rotherham 2013c).

406

407 Lowland and upland fringe sites were exploited and progressively destroyed during the  
408 sixteenth, seventeenth, and eighteenth centuries; much of this removal associated with  
409 Parliamentary and private ‘enclosures’ of heath, moor, common, bog and ‘waste’  
410 (Gimingham 1972; Webb 1998, 1986; Rotherham 2011a). Whilst agrarian and then early  
411 industrial communities often maintained vital environmental resources through complex  
412 social, economic, and legal mechanisms, capital-based exploitation was often more  
413 destructive (Rotherham, 2013b). Therefore, although the productive landscapes created, in  
414 the medieval period reflected social need, with mechanisms for conservation providing  
415 continuity over long periods, increasing technological industrialisation, urbanisation, and  
416 rural de-population transformed or exhausted the resources. The essentially conservative  
417 cultural landscapes were abandoned, transformed, and swept away on a tide of ‘improvement’  
418 (Rotherham 2014a). With industrial and urban areas increasingly techno-centric, agriculture  
419 became industrial, fuelled by coal and then oil. With social and technological innovation  
420 during the 1700s and 1800s, landscapes were freed of dependence on local sustainability and  
421 the needs of local subsistence. By the late twentieth century, across much of Britain, the  
422 cultural landscape was a largely forgotten, archaic relict of a lost existence.

423

424 [Figure 5]

425

### 426 **Heath, common, fen, & bog**

427

428 Peatland landscapes and their vegetation make a good example of the processes of traditional  
429 utilisation and subsequent cultural severance and either improvement or abandonment.  
430 Formerly widespread and abundant across much of Western Europe these habitats are now  
431 drastically reduced (Rotherham 2011b). Exploitation of these resources transformed both  
432 landscapes and sometimes the economy too (see De Vries 1974; De Zeeuw 1978). Those  
433 areas that remain are often in poor condition, and such landscapes have been transformed by  
434 agricultural intensification and land ‘improvement’. Britain holds a globally significant  
435 resources of these sites, and the scale of destruction, abandonment, and resulting  
436 fragmentation of lowland heaths has long been recognised (e.g. Gimingham 1972; Webb  
437 1986). However, even these authors substantially under-estimated the wider extent and the  
438 scale of loss. Indeed, this destruction or abandonment so completely decimated the resource,  
439 that even expert researchers missed much of it (Rotherham 1995, 1996, 2009). This has  
440 erased landscape, ecology, and community from the countryside, with resulting reduced  
441 biodiversity and lost bio-cultural heritage.

442

443 The bulk of the people were cottagers, labourers, farm servants, and squatters, and the heath  
444 or common was a resource at the centre of their lives and existence. Cottagers either owned  
445 or occupied cottages, to which ancient custom attached rights of commonage on the ‘wastes’.  
446 Such rights were of various kinds including the pasturing of animals on the common, cutting  
447 of turf and extraction of fuel, or of building stone. Widespread conversion of heath, moor,  
448 waste, fen, bog, and marsh, to arable and to enclosed pasture, abruptly ended such utilisation  
449 by the rural population especially the rural poor and the poorer commoners.

450

451 [Figure 6]

452

453 The nature and scale of induced landscape changes are basic to understanding their present-  
454 day condition and character, much the result of long-term use for subsistence farming. The  
455 scale of such impacts has been significantly under-appreciated. Separation of these  
456 landscapes into lowlands (enclosed and improved) and uplands (unenclosed and unimproved)  
457 often masks the human element which is often less evident in peripheral zones. Where it has  
458 been environmentally possible and economically feasible, enclosure, liming, cutting, and

459 drainage, followed by agricultural intensification or creeping urbanisation, changed most  
460 lowland areas almost beyond recognition. In the twenty-first century, the loss of cultural,  
461 subsistence impacts has big implications for nature conservation management. With their  
462 plagio-climax communities, abandoned heaths, commons, grasslands, and other habitats  
463 progress speedily through successional changes to tall herb, scrub, and woodland. Ecologies  
464 change and many conservation priority species decline.

465

466 Even with recognition of the need for effective conservation, this is frequently without an  
467 appreciation of the scale of the impacts of cultural severance and the loss or change in  
468 biodiversity. Very often, conservation is primarily protection with relatively small areas of  
469 targeted, generally grant-aided management. Conservation managers have a range of tools to  
470 achieve their aims, and these include varying mixes of grazing by domestic stock and regimes  
471 of cutting or harvesting tall vegetation in order to halt or deflect ecological successions.  
472 Whilst these bring some benefits to biodiversity conservation, they often omit important  
473 aspects of the traditional processes that they attempt to mimic. An important part of the bio-  
474 cultural heritage that is lost with cultural severance is the intangible heritage of local  
475 knowledge of process and product. Furthermore, not driven by economic utilisation,  
476 conservation is vulnerable to short-term grant availability and problems of spatial, temporal,  
477 and cultural continuity. The results for ecology may also be disastrous (Denton 2013, 2014).

478

#### 479 **Local subsistence to feed people & livestock**

480 Pre-industrial European subsistence landscapes had, of necessity, direct importance and  
481 relevance to local people. Landscapes generate distinctive local and regional products and the  
482 distinctive capabilities and character of the land led to particular land use patterns and  
483 resulting ecologies (Rotherham (ed.) 2013). Coastal wetlands for example, were valued for  
484 dairy products like cheese, and for high quality beef, and other meats such as mutton. Such  
485 otherwise marginal lands might be maintained when less valued areas were lost to  
486 ‘*improvement*’ (Rippon 2000). Salt marshes dotted with salterns produced salt to flavour and  
487 preserve food, and again were highly valued. For landscapes unaffected by modern  
488 technologies, interactions between production of food and drink were especially intimate. As  
489 recently as the late 1800s and early 1900s, countries like Britain were characterised by  
490 distinctive zones of productive farming landscapes, with characteristic local and regional  
491 foods, drinks, cultures, traditions, and biodiversity. The bio-cultural heritage generated by the  
492 distinctive regions was often related to such utilisations, and the cultural heritage derived  
493 from seasonal celebrations in regional farming or land use calendar.

494

495 In Britain, there were broad zones with lowland arable farming, coastal grazing marsh,  
496 western beef, dairy, and upland sheep and beef cattle. Each zone reflected particular  
497 environmental constraints, but in a pre-petrochemical era with dependence on animal power,  
498 all areas produced mixed crops including oats, hay and grass. With their wet conditions,  
499 upland zones favoured oats over other grain, but also maintained hay meadows and pastures  
500 to feed livestock and draught animals. The utilisation of long periods generated distinctive  
501 ecologies, landscapes and land use patterns, and associated cultural traditions and behaviours.  
502 Together these have become a complex of tangible and intangible bio-cultural heritage and  
503 biodiversity. Typical birds in the British uplands for example, included twite and corncrake,  
504 both associated with traditionally managed landscapes. The two species became very rare and  
505 in many regions extinct, following cultural severance and the ending of traditional  
506 management. Similarly, with the ending of traditional uses, rural depopulation meant the  
507 communities of these landscapes dissipated, the patterns of meadows and pastures were lost,  
508 and associated ecologies declined. These patterns of use plus the plants and animals, the built

509 structures and human communities, amount to the bio-cultural heritage characteristic of each  
510 region.

511

512 Food production and harvesting, together with hunting for sport and for food, have influenced  
513 rural landscapes across Europe over thousands of years. Hunting has been hugely significant  
514 in many European landscapes and affected species and cultural heritage both directly  
515 (through exploitation) and indirectly (through landscape management). Alongside the  
516 tangible heritage, hunting activities have extensive intangible heritage too. Land maintained  
517 and administered for hunting greatly influenced landscape development at both and regional  
518 levels. England for example, had extensive royal forests, private parks and chases established  
519 from late Saxon times with elements of some persisting into the modern countryside  
520 (Rotherham, 2007b; Liddiard 2003, Liddiard (ed.) 2007; Tubbs 1986). Many hunting parks  
521 led directly or indirectly to grand ornamental parks of the 1700s and 1800s, which in turn  
522 became defining features in many British landscapes (e.g. Harding, & Wall (eds) 2000).

523

524 [Figure 7]

525

526 Alongside obvious modern landscape drivers of agriculture and forestry, there have been  
527 numerous intimate and subtle interactions between communities and the countryside for  
528 sustenance, subsistence, for hunting and sport. The traditional and customary uses have  
529 created rich bio-cultural resources but in many parts of Europe, these have ended over the last  
530 two centuries, but the decline has accelerated through the late twentieth century. Whilst rural  
531 bio-cultural heritage and diversity is a living dynamic, and so changes and evolves over time,  
532 sudden and rapid industrialisation or abandonment of traditional uses threatens both biological  
533 and cultural aspects. As uses end, even their memory, and the intangible, cultural heritage is  
534 lost too. For nature conservation, the problem seems to be that understanding of what  
535 generated and maintained the biodiversity that we want to safeguard or enhance, an intangible  
536 heritage, has been lost. This situation then presents serious problems for planners and  
537 managers when restoration projects, for example, fail.

538

539 Local and regional exploitation of landscapes and their resources varies with environmental,  
540 socio-political, and economic pressures. During climatic deterioration, for example, upland  
541 zones may be abandoned and low-lying peatlands may be vulnerable to catastrophic flooding.  
542 Political and economic pressures may tip the balance of spatial disputes and resource  
543 competition, or may push communities back to subsistence use of environmental resources as  
544 more sophisticated assets become scarce or unaffordable. Over-use and exhaustion of a  
545 particular resource, or access to it restricted for social or political reasons, may mean  
546 alternatives, including less suitable materials, have to be found. Competition or restriction on  
547 use might be through the influence of different or alternative requirements like oak timber for  
548 the late medieval navy affected by wood harvested for charcoal-fuelled iron smelting. Both  
549 these competed with wood use for fuel by both rich and poor; but poorer commoners and  
550 peasants were the worst affected. Competition between commoner, peasant, and major  
551 landowners, and between industrial exploitation and domestic use has affected many  
552 landscapes and importantly, their bio-cultural heritage today. In some cases, later industrial  
553 uses have removed entire palimpsests of eco-cultural landscapes and their associated heritage  
554 resources, both tangible and intangible.

555

556

### 557 **Cultural severance & bio-cultural heritage**

558

559 Human resource use in the natural landscape is a fundamental driver (Rotherham, 1999; De  
560 Moor et al. 2002; Agnoletti et al. (eds.) 2005). It interacts with the ecology and other

561 environmental factors through complex social, legal, economic, and political mechanisms,  
562 facilitating and constraining usage (e.g. Tubbs 1986). Almost all the landscapes observed  
563 across Europe and many other parts of the world are eco-cultural, often managed in  
564 traditional ways for centuries. In this context, observations the impacts of people over time  
565 have been described for 1) wooded and forested landscapes, 2) marsh, meadow and fen, 3)  
566 heath, bog and common, and for 4) cultivated landscapes such as field systems (Rotherham  
567 2014a).

568  
569 Traditional and customary '*cultural*' utilisation, whilst not always sustainable, generated and  
570 drove many landscapes we now value so highly (e.g. Rotherham 2007b, 2013b). The reasons  
571 and mechanisms were discussed earlier, ranging from direct environmental impacts (like  
572 lowering nutrient levels and creating micro-disturbance), to indirect effects through social  
573 and economic impacts (allowing people the means to remain on the land or in a region). With  
574 the agricultural and industrial revolutions, the long process of severance in European  
575 landscapes began in earnest with people and supplies of food, fuel, building materials or other  
576 resources increasingly separated. Timing and impacts vary but the trends are the same across  
577 most European countries, though lagging in some eastern and Mediterranean regions.

578  
579 As industry and agriculture move more towards technological processes, resources and  
580 solutions, nature and local landscapes are less important. Concomitant with severance is the  
581 massive shift from rural to urban populations, a phenomenon that continues to accelerate in  
582 the twenty-first century. For many regions, rural working communities in subsistence  
583 landscapes were displaced to become urban poor. The traditional rural environment becomes  
584 a disputed space and local people squeezed out; the landscape abandoned to become a  
585 backdrop to tourism, and the affluent seeker of rural recreation (Rotherham, 2014a).

586  
587 [Figure 8]

588  
589 However, there is a problem and it is one of the most serious threats, perhaps to  
590 environmental sustainability, and certainly to nature conservation. With the abandonment of  
591 traditional uses and practices, many sites have been lost or fragmented. Those that remain  
592 now have little or no management, and more-or-less quickly pass through successional  
593 change. Not '*natural*' but '*cultural*' landscapes, these have ecologies evolved over centuries  
594 of locally distinct and generally predictable exploitation driven by economic need. Attempts  
595 to now conserve and manage the remaining fragments are often far too little and far too late.  
596 Importantly too, they generally omit key parts of the traditional process, and have no long-  
597 term economic viability, or at least not one connected to land management processes.

598  
599 Economy and landscape are separated at the level of cultural tradition and subsistence, to be  
600 replaced by a '*sticking plaster*' approach through targeted grant aid. This is often laudable  
601 and in the short-term may be essential if sites, species and even some traditions are not to be  
602 totally lost. However, it is not a long-term solution, and it may be dangerous indeed to  
603 believe that it is. There is a widespread myth that release from farming in many areas will  
604 lead to '*re-wilding*' or '*re-naturing*' of landscapes, and so will be inherently good for wildlife  
605 (Monbiot 2013a, 2013b; Fisher 2006, 2013; Carver 2014). Some of the critical issues were  
606 discussed by Dudley (2011) and by Elliot (1997) in considering both the issues of so-called  
607 '*faking nature*', and the often-unquestioned '*ethics*' of restoration projects. Despite concerns,  
608 it is clear that restoration, however defined, offers many opportunities for improvement (e.g.  
609 Egan et al. (eds) 2011; Hall (ed.) 2009). In the case of re-wilding or re-naturing, it is true that  
610 some species will benefit, although these will ebb and flow as successional changes move on.  
611 However, the abandonment of a cultural or working landscape will in many cases simply  
612 amount to dereliction. This is seen increasingly across the Mediterranean, as rural areas are

613 de-populated and both social and environmental problems result (e.g. Grove, & Rackham  
614 2003). Favourably located landscapes may acquire a veneer of tourism affluence or  
615 commuter-belt sophistication, but most areas go into steep decline. With a derelict landscape  
616 and no working rural population, a degraded ecology and an abandoned cultural heritage,  
617 most regions will hold little appeal for the tourism or the leisure visitor (Rotherham 2013a,  
618 2013b).

619

620 Progress in modern economies usually means socio-economic development, rural  
621 depopulation, and urban growth. It involves technological provision of needs, and separation  
622 of people from nature. The process forms part of human cultural evolution, but with serious  
623 environmental consequences. With severance of people and landscape, there is rapid de-  
624 coupling of communities from local environmental resources and ending of traditional land  
625 uses. For individual sites, the ecological consequences can be especially problematic:

626

- 627 1. Eutrophication due to non-removal of biomass (for fuel, animal bedding, fodder)
- 628 2. Lack of micro-disturbance from grazing or other working animals, and from  
629 subsistence activities (including transhumance use *etc*)
- 630 3. Lack of propagule dispersal, particularly seeds through grazing stock moving from  
631 site to site
- 632 4. Successional change due to abandonment (the rate varying with the landscape and its  
633 location, so upland zones in the UK for example are more resilient than lowland one)
- 634 5. Decreased value for local communities and abandonment or replacement by other  
635 uses (building development *etc*)
- 636 6. Fragmentation and isolation

637

638 At regional levels, there is serious risk of losing unique cultural heritage. Furthermore, this  
639 not only drove the ecology of former landscapes, but also may be a vital link to heritage  
640 tourism economies in the future. That gross changes are driven by economic '*progress*' but  
641 responses are generally not, remains a huge challenge since most conservation initiatives are  
642 cosmetic rather than economic. Many successful conservation and environmental projects  
643 across Britain and throughout Europe are to be celebrated and encouraged, but their scale is  
644 insufficient to redress the balance of on-going losses. Additionally, most conservation  
645 projects merely address their ecological or 'biodiversity' components not the wider, more  
646 fundamental, bio-cultural heritage.

647

648 There are examples of good practice such as the work of bodies such as England's National  
649 Trust in Cumbria. Here the operations are closing the gap between nature conservation and  
650 the local economy. However, this limited success is against a backdrop of cultural landscape  
651 abandonment probably unprecedented in human history, and it does not necessarily address  
652 core issues of bio-cultural heritage.

653

654 As local cultural knowledge is lost or not recognised, we no longer know how landscapes  
655 were managed even fifty or so years ago. This process of severance is happening rapidly  
656 across Europe, especially around the Mediterranean, and in former Eastern Bloc countries. It  
657 has occurred in Britain too (e.g. Rotherham 2007a), and recent ethnological research with  
658 older farmers in the Peak National Park showed how rapid abandonment of family farms is  
659 leading to loss of the intangible heritage of local knowledge about countryside. Those  
660 wishing to conserve such lands and their unique wildlife heritage frequently have little  
661 understanding of how the ecology evolved through the eco-cultural nature of the landscape  
662 (Rotherham 2014b).

663

664

665 **Sherwood Forest in Nottinghamshire as an example of process & problem for bio-**  
666 **cultural heritage**

667

668 The example of wooded or forested landscapes and their trees was given earlier as an  
669 example for which there are rich and diverse resources of biodiversity inter-twined with bio-  
670 cultural heritage (Rackham 1986; Peterken 1996). Sherwood Forest displays the  
671 characteristics of the consequences of cultural severance. The highest level of bio-cultural  
672 heritage of the area is in the great and ancient trees of global iconic status. Furthermore, the  
673 trees hold uniquely rich biodiversity from rare fungi to threatened saproxylic invertebrates  
674 (see for example, Rotherham 2007b). Additionally, some of the individual trees have famous  
675 stories and cultural heritage attached to them, and all the big trees tell a story of the  
676 remarkable landscape history of the Forest. There is much more to add, but this sufficiently  
677 illustrates the point.

678

679 The Forest mixed historic uses as grazing and commonland for peasants and as a royal  
680 hunting preserve. The great trees were standing in an expansive landscape of grazed wood  
681 pasture with heath, bog, grassland and scrub, which most of all was open. The world-famous  
682 Major Oak is one of the largest of its type in the world, and tells of growing as an open-  
683 grown tree in a grazed landscape. Some of the trees may be '*shreds*' and others may be  
684 '*pollards*'.

685

686 Probably around 100 to 150 years ago, the grazing management reduced considerably, and  
687 the royal forest had long-since been abandoned. From the 1920s onwards, and accelerating  
688 during the 1950s and 1960s, much of the area was afforested with exotic conifers, and other  
689 areas were military training grounds. With stock grazing now ended, the ecology began  
690 progressive successional change to scrub and then young birch woodland. Severance  
691 occurred with removal of traditional management but also with the loss of the local  
692 commoners and others with a stake in the grazed landscape. Areas were either abandoned to  
693 succession or planted with conifers, both resulting in the '*shrouding*' of the bio-cultural  
694 heritage resource of veteran trees, and the associated biodiversity. Some of the great trees  
695 survived though with diminished vigour, but many slipped silently and unnoticed into death.  
696 By the 1980s, it was realised that this nationally and in parts, globally, significant landscape  
697 and its ecology was in serious decline. Therefore, steps have been taken to reverse  
698 successional change and to remove conifer plantations. The work has achieved some success,  
699 though many ancient trees have been lost. Ironically, grant-funded projects to remove  
700 conifers from around veteran oaks (haloing) succeeded in actually killing many because the  
701 work was undertaken too quickly and the sudden change in microclimate was too much of a  
702 shock.

703

704 The restoration work is dependent on grant aid but this is justified because of the huge  
705 significance of the area for its heritage, its ecology, but also the economic significance of its  
706 tourism. There are even ideas of reintroducing large grazing herbivores to areas of the forest  
707 and the heath, though this has to be aid for whereas historically it was an economic driver.  
708 Growing populations of both red deer and fallow deer are already spreading across the area  
709 having escaped from nearby aristocratic parks. Overall, the future looks reasonably positive  
710 for the Forest, though government support and funding are always in doubt. However, over  
711 the last century or more, a huge and irreplaceable bio-cultural heritage has been lost. Some of  
712 the tangible heritage survives but most of the intangible, cultural knowledge has been lost.

713

714

715 **Futurescapes & re-wilding**

716

717 Approaches to conservation or landscape management, that suggest either or both  
718 'abandonment' or 're-wilding', can in fact present significant problems. Firstly, with  
719 declining biodiversity over the last fifty years or more, abandonment may be the final straw  
720 for many species. In terms of bio-cultural heritage, abandonment is undoubtedly disastrous.  
721 Furthermore, from an historic perspective, abandonment-style 're-wilding' is itself a  
722 misnomer since it implies a reversion to a former 'natural' state (Monbiot 2013a, 2013b;  
723 Fisher 2006, 2013; Carver 2014). In reality, this is a myth and the example of regenerating  
724 the Great Forest of Caledon exemplifies this. It is a great idea, which gains a strong emotional  
725 response. If the Great Forest had existed, then the idea would be even better. The reality is  
726 that most landscapes lacking trees in northern Scotland have done so for thousands of years.  
727 The history is that these were not 'wild', 'natural' areas but settled populated landscapes  
728 (Rotherham 2014b).

729

730 Taking people out of the landscape and separating people from nature is wrong on many  
731 social, ethical, economic and political levels; and damaging to ecology, biodiversity, and bio-  
732 cultural heritage. It is not inherently wrong to create patches of Caledonian Pine Forest  
733 habitats now and in the future, if that is what we want for nature conservation. However, to  
734 claim this is somehow re-creating past landscapes is misleading and misinformed. The other  
735 aspect of the wilding debate is the idea of 'feral', 'free', 'self-willed' nature unfettered by  
736 human constraints (Monbiot 2013a, 2013b; Fisher 2006, 2013; Carver 2014; Taylor 2005).  
737 However, is it a problem if abandonment to feral nature leads to colonisation by invasive  
738 Sitka spruce, Japanese knotweed, bracken, Himalayan balsam, or rhododendron? We decide,  
739 and the result is eco-cultural not natural. Feral future nature might be significantly populated  
740 by competitive, exotic, globalising species, something which the advocates of 're-wilding'  
741 rarely discuss, but according to Pearce (2015) for example, might be inevitable. However,  
742 large herbivore projects such as at Knepp in southern England (Taylor 2006) or Wild  
743 Ennerdale in northwest England (Browning & Gorst 2013; Taylor 2010), offer alternative  
744 visions of a wilder futurescape that resonates with many historic landscapes. Coastal zones  
745 also offer major opportunities for designed wilder environments (e.g. May et al. 2006).  
746 Unfortunately, the ideas and opportunities of wilder landscapes are frequently misunderstood  
747 (e.g. Vidal 2006), with suggestions of unsuitable and unsustainable herbivores for example,  
748 'wilded' into inappropriate environments. Furthermore, this is often in landscapes perceived  
749 as devoid of people.

750

751

## 752 **Nature with & without people**

753

754 To remove people from the natural world is not natural since we are a part of nature. For  
755 many reasons, in the twenty-first century it may be beneficial to have wilder landscapes than  
756 we have had in the recent past. However, these are designed scenarios led and determined by  
757 humanity. The skill in nature conservation and a challenge for the future may be to embed  
758 humanity in the natural world but in ways less damaging and more positive than throughout  
759 the twentieth century. There is a toolkit of options and techniques that can be applied as  
760 various types of land management and deliberate conservation interventions. Approaches  
761 may involve grazing by wild, feral, or domesticated stock of various types applied in different  
762 densities and seasons (e.g. Rotherham 2013a). These interventions influence the ecological  
763 outcomes, and sometimes the favoured approach may be non-intervention. Each intervention  
764 or non-intervention leads to ecological successional changes and, based on knowledge of site  
765 environmental conditions these are predictable (Rotherham 2014b).

766

767 This is not new since people have intervened in nature for many diverse reasons over many  
768 centuries. In times past, such as with the great English landscape designers of the 1700s and



769 the wild gardeners of the 1800s, people have created ‘*wilderness*’. This was often through  
770 removal of people, application of grazing regimes, and ‘wild’ exotic species; manufactured,  
771 romantic wilderness but not natural. These wild landscapes were to be viewed from the  
772 outside rather than lived in. In designing futuroscapes, wilder landscapes have much to offer,  
773 alongside other traditional and indeed, radical conservation approaches (Rotherham 2014a,  
774 2014b).

775

776 In a rapidly changing world, approaches to landscape conservation need planning and design.  
777 They must apply science and the insights of history. Failure to do this effectively risks  
778 continuing downwards spiral of environmental quality and declining species diversity,  
779 alongside rural depopulation and deteriorating rural economies, both farming and tourism.  
780 Because of massive human impacts on the environment over countless centuries, whether we  
781 like it or not, we are custodians of the countryside. The responsibility for future landscapes is  
782 ours. The decisions we make and how those decisions are reached may be debated, but  
783 simply abandoning landscapes, (and even de-populating them), are not viable options.  
784 However, ‘*wilder by design*’ and large-scale, imaginative, wilding projects in appropriate  
785 locations, offer great possibilities, but these are not wilderness but wilder eco-cultural  
786 landscapes (Rotherham 2014b).

787

788

### 789 **Futurescapes visions & free nature**

790

791 Where does bio-cultural heritage sit within a vision of free, self-willed nature? How should  
792 conservation bodies respond to remnant biodiversity and priority species lost when a site is  
793 ‘*freed*’? Even if we accept that, ‘*the loss of a few species is a price worth paying for a wilder*  
794 *nature*’ (Carver 2014), who decides? If we intervene, then who does it, why do they do it,  
795 what do they do, where do they do it, and when do they do it, and who decides and pays?  
796 (Rotherham 2014b) Over centuries, people have shifted environmental baselines so  
797 significantly that whether we choose to intervene or not, the outcomes are culturally  
798 determined within eco-cultural landscapes. Even the decision not to intervene is a positive  
799 intervention; people and nature trapped within our humanity, as a part of nature. Therefore,  
800 the critical paradigms are concerned with 1) the type of human interventions in nature and the  
801 responses that follow the changed parameters; and 2) how might these be managed and  
802 manipulated to free nature for a wilder landscape; and 3) how does bio-cultural heritage fit  
803 within these conceptual frameworks. History and science inform likely trajectories for future,  
804 wilder nature, but it may be a rocky road ahead.

805

806

### 807 **Conclusions**

808

809 Bio-cultural heritage needs to be placed firmly at the forefront of conservation, as a link  
810 between people, history and biodiversity. In Europe most coppice woods and associated  
811 ground flora, birds like nightingales, and woodland butterflies have gone. With ancient wood  
812 pastures abandoned, we lose 1,000-year-old oaks, unique saproxylic invertebrates, lichens,  
813 and fungi. Heathlands and grasslands like meadows and pastures, are essentially eco-cultural;  
814 severed from human tradition, they become rank, eutrophic communities of little ecological  
815 interest aside from catholic, competitive, opportunists. All this is widely known, and  
816 predicted in the plant strategy work of Grime et al. 2007) and Hodgson (1986), and by  
817 specialists like Webb (1986, 1990, 1998) and Chadwick (1982), considering European  
818 heathlands. With areas abandoned, landscapes become contested spaces; local, traditional  
819 peoples squeezed out by capital-intensive land-uses, absentee landowners, and leisure or  
820 recreation (Rotherham 2014a). Whilst traditional management transformed ecologies, local

821 economic dependence fosters sustainable uses unless other factors tip the balance. Incomers  
822 may bring fresh ideas and fresh funds for environmental management and innovation, but  
823 across Europe, traditional landscapes morph into either abandonment or into leisurely  
824 landscapes detached from most ecosystem functions. With a few exceptions such as the work  
825 of the National Trust in Britain, little of the leisure and tourism economic activity feeds back  
826 into land management or conservation.

827

828 Following abandonment, biomass increase, eutrophication, and intensive recreational use or  
829 urbanisation, many areas become vulnerable to rampant wildfires. From California, to  
830 Australia, from Greece, Spain, and Italy to France, and from Dorset heaths to Peak District  
831 moors, such fires are predictable results of cultural severance and abandonment. Traditional  
832 peoples often used regular fires to manage their landscapes, re-cycling and releasing precious  
833 nutrients, and providing essential grazing at the right time of year. When European  
834 imperialists populated the planet, they generally viewed native, indigenes as ignorant and  
835 primitive, and suppressed local fire management of landscapes. Today's catastrophic  
836 wildfires are direct consequences and descendants of past cultural severance (Rotherham  
837 2008, 2013b, 2014b; Pyne 2001).

838

839 A major challenge now is to record local cultural knowledge and insight, to re-build and  
840 celebrate local connectivity with nature, to value local traditions and uses, and to apply the  
841 knowledge in a meaningful way. Bio-cultural heritage is at the core of such ideas. It is neither  
842 possible nor desirable (socially and economically) to stop the clock, but we need to find long-  
843 term economically sustainable solutions to these problems. The approaches must be more  
844 ambitious and more radical than anything that we have achieved so far. Webb (1986, 1998)  
845 considered the issues and conservation management options for the European heathland  
846 component of this discussion, and his prognosis was less than positive.

847

848 With dynamic landscapes and fluid ecologies (Rotherham 2014a) replaced by fixed locations,  
849 habitat fragmentation, isolation, and soils and water altered by eutrophication, environmental  
850 conditions today are not '*natural*'. Regular micro-disturbance, vital for many species, is  
851 replaced by unpredictable macro-disturbances. Domestic grazing herbivores or wild / wilded  
852 stock may be either beneficial or calamitous for conservation target species, depending on  
853 what, how, and when (Rotherham & Lambert 2011; Rotherham 2014b). Introducing large  
854 herbivores into small, isolated sites does not produce ecological benefits since they lack the  
855 dynamics of larger-scale ecosystems. Animal behaviour is not '*natural*' without large  
856 carnivores influencing and directing herbivore feeding patterns and movement. Abandoned to  
857 '*re-wilding*' without either or both large herbivores or carnivores is not '*natural*' but  
858 attenuated ecology lacking keystone fauna or traditional management (Rotherham 2013a).  
859 Ecological successions are then predictable but no more '*natural*' than other options, and to  
860 intervene or not, a management decision for already highly modified landscapes. Informed by  
861 history, ecological visions must look forwards to new futurescapes to conserve and enhance  
862 not only biodiversity, but also bio-cultural heritage. The realities of such visions are  
863 beginning to be addressed by environmental writers such as Marris (2011), but there is a  
864 genuine risk that bio-cultural heritage may simply be overlooked and lost.

865

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867

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Figure 1. Ancient open grown tree - the world-famous Major Oak of Sherwood Forest - a landscape hunted by kings but grazed by commoners

Figure 2. The Wood-Wain by Birkett Foster 'The Wood-Wain' showing a managed English wood in the early nineteenth century as eco-cultural countryside

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