



A study on the enhancement of sonochemical degradation of eosin B using other advanced oxidation processes

GOEL, Mukesh, DAS, Ashutosh, RAVIKUMAR, K. and ASTHANA, Abhishek

Available from Sheffield Hallam University Research Archive (SHURA) at:

<http://shura.shu.ac.uk/8797/>

This document is the author deposited version. You are advised to consult the publisher's version if you wish to cite from it.

Published version

GOEL, Mukesh, DAS, Ashutosh, RAVIKUMAR, K. and ASTHANA, Abhishek (2014). A study on the enhancement of sonochemical degradation of eosin B using other advanced oxidation processes. *Desalination and Water Treatment*, 52 (34-36), 6770-6776.

Repository use policy

Copyright © and Moral Rights for the papers on this site are retained by the individual authors and/or other copyright owners. Users may download and/or print one copy of any article(s) in SHURA to facilitate their private study or for non-commercial research. You may not engage in further distribution of the material or use it for any profit-making activities or any commercial gain.



Effects of process parameters during the sonochemical degradation of benzene and toluene in water

Journal:	<i>Environmental Technology</i>
Manuscript ID:	TENT-OA-2012-0520
Manuscript Type:	Original Article
Date Submitted by the Author:	17-May-2012
Complete List of Authors:	Goel, Mukesh; PRIST University, Environmental Engg Das, Ashutosh; PRIST University, Environmental Engg Asthana, Abhishek; Sheffield Hallam University, Chemical Engg K, Ravikumar; PRIST University, Environmental Engg
Keywords:	Sonication; , Cavitation; , Chemical Processes; , Pollution; , Remediation

SCHOLARONE™
Manuscripts

Only

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28

Effects of process parameters during the sonochemical degradation of benzene and toluene in water

¹Mukesh Goel*, ¹Ashutosh Das, ²Abhishek Asthana and K. Ravikumar¹

¹Centre for Environmental Engineering, PRIST University, Thanjavur, India – 613403

²Hallam Energy, Sheffield Hallam University, Sheffield, United Kingdom, S1 1WB

29 *Corresponding Author Tel: 0091-9600833057; Fax: 0091-04362-266960
30 Associate Professor, Centre for Environmental Engineering, PRIST University,
31 Thanjavur – 613403
32 Email: goelmukesh0@gmail.com
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

Abstract

This work describes the sonochemical (chemical effects of ultrasound) degradation of aqueous solution of benzene and toluene using ultrasonic probe (20 kHz frequency). The process variables studied include initial concentrations, temperature, and type of saturated gas. The effect of addition of H_2O_2 and chloride ions was also examined. The results were explained based on the hot spot theory. The reaction rate showed a reverse proportionality to the initial concentration, which may be attributed to the decrease in specific heat ratio with increase in initial concentration, thereby causing the decrease in temperature inside the cavitation bubbles and hence reduction in the reaction rate. However, H_2O_2 concentration and the type of the dissolved gas did not seem to affect the reaction rates.

Key words: Sonication; Cavitation; Chemical Processes; Reaction Engineering; Pollution; Remediation

1. Introduction

The rapid industrialization in the present society is leading to severe air, water and soil pollution. The disposal of huge amount of wastes in the aquatic streams has especially caused massive water degradation. These adverse effects are substantially affecting human health. The presence of even trace amount of dangerous chemicals like volatile organic compounds (VOCs) is sufficient to render water unusable for reasons of long term associated health risks [1, 2]. In fact, the Environmental Protection Agency (EPA) in the USA has listed more than 100 toxic pollutants present in water [3, 4]. Although clean and safe drinking water is indispensable for the sustained and healthy progress of the society, yet, the present set up precludes the possibility of avoidance of the generation of these wastes altogether.

Numerous water treatment technologies have been developed to eliminate toxic chemicals from water. An ideal waste water treatment technique should not only mineralize all the toxic species present in the water stream without leaving any hazardous residues, but also be cost effective. Since sonication is a chemical-free technology, there has been a growing interest in the sonochemical destruction of organic contaminants in water, especially in recent times. Several studies have been reported in the literature in this regard [5-7].

Mechanism of sono-chemical treatment is fairly well-agreed upon among researchers. The process of cavitation has been identified to be the primary mechanism for ultrasonic degradation. In fact, cavitation is the outcome of the expansion and contraction of

1
2
3 cavitation nuclei due to the presence of compression and rarefaction cycles in the
4 ultrasonic wave. Subsequent effects include bubble formation, rapid growth and finally
5 implosive collapse. The implosion of cavities causes an unusual reaction environment in
6 the vicinity of the bubbles [8, 9], resulting in intense heat and local hot spots. Thus, the
7 sound energy is converted into thermal energy, which is one of the major causes behind
8 much of the observable sonic degradations. Based on the pioneering study by Suslick
9 et al. [10], the temperature inside the cavity is expected to reach as high as 5,200 K,
10 the interfacial region (i.e., between the solution and the collapsing bubble) being
11 maintained at about 1,900 K or so. In essence, sonochemical effect does take place
12 either due to the pyrolytic decomposition at the gas-liquid interface or by the reduction
13 and oxidation due to the generation of H[•] and OH[•] radicals [11].
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31

32 The purpose of this work was to examine the effect of process variables such as
33 temperature, initial concentration, types of dissolved gas and addition of H₂O₂ on the
34 sonochemical reaction. The chemical compounds studied were benzene and toluene.
35
36
37
38
39
40

41 **2. Experimental Section**

42 *2.1 Materials*

43 Reagent-grade benzene (from Aldrich), toluene (from Merck) and H₂O₂ (from Merck)
44 were used as received. Aqueous solutions were prepared by dissolving the compounds
45 in deionized water.
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

2.2 Analytical Methods

Quantitative analysis of benzene and toluene were determined with HP 6890 (Hewlett-Packard Inc.) purge and trap Gas Chromatography (GC), equipped with an auto sampler, flame ionization detector and a column (HP-624, 30 m x 0.53 mm x 3 μ m). The analysis was conducted under the following conditions: injector temperature and detector temperature were both kept at 250°C, whereas the oven temperature was held steady at 110 °C. Samples (5.0 ml each) were drawn at various time intervals and transferred to GC sampling vial for analysis.

2.3 Apparatus

Sonication was done using an ultrasonic source (VC-750, Sonics and Materials), with the probe tip measuring 19 mm (in diameter) and the ultrasonic source (of frequency 20 kHz) employed at 50% amplitude. A water jacketed glass vessel (with Teflon cover) was used as reaction vessel. The volume of the solution taken for study was 200 ml, with almost practically no head-space (Fig. 1). The temperature was monitored periodically with the help of a thermocouple immersed in the reacting medium.

3. Results and Discussion

3.1 Influence of initial concentration

Air saturated aqueous solution of benzene and toluene were sonicated at four different initial concentrations (viz. 50, 100, 150 and 200 ppm, by volume). In all cases, the concentration was found to decrease exponentially with time, indicating first order or pseudo first order kinetics. In all cases, the experimental run was performed for 150 minutes. The degradation curve for benzene is shown in Fig. 2. Based on the results

1
2
3 obtained, the first order reaction rate seems to depend upon the initial concentration.
4
5 Accordingly, a decrease in rate constant (k) was observed with increase in the initial
6
7 concentration. A similar trend was also observed for toluene degradation. This
8
9 observation is very much similar to the reports by other investigators, as well. Jiang et
10
11 al. [12] found that the rate constants of chlorobenzene and 1,4-dichlorobenzene
12
13 decrease with the increase in initial concentration. They found that the reaction rate
14
15 constant (k) decreased from $k = 0.035 \text{ min}^{-1}$ (for initial chlorobenzene concentration of
16
17 $40 \text{ }\mu\text{M}$) to $k = 0.020 \text{ min}^{-1}$ (for initial chlorobenzene concentration of $500 \text{ }\mu\text{M}$).
18
19
20
21
22

23
24 The reaction condition showed distinct changes with change in concentration of the
25
26 organic compound. De Visscher et al. [9] derived the relationship between specific heat
27
28 ratio (γ) and initial concentration. For relatively small mole fractions of volatile
29
30 compounds in the cavitation bubble, they found that:
31
32

$$\gamma = \gamma_0 - KC_i \quad (1)$$

33
34
35
36
37
38 where γ is the specific heat ratio for pure gas-water mixture and K is a proportionality
39
40 constant. Since the specific heat ratio of gas-water mixture is larger than that of the
41
42 volatile compounds, K must be positive. Thus, specific heat ratio is expected to
43
44 decrease with the increase in initial concentration. Furthermore, according to hot spot
45
46 theory, the degradation rate is function of the temperature and pressure inside the
47
48 bubble. It is given by Noltingk and Neppiras [13], as follows:
49
50
51
52
53
54
55
56
57
58
59
60

The maximum temperature, T_{max} , obtained is:

$$T_{max} = T_0 \left[(\gamma - 1) \frac{p_a}{p_0} \right] \quad (2)$$

and, maximum pressure, p_{max} , obtained is:

$$p_{max} = p_0 \left[(\gamma - 1) \frac{p_a}{p_v} \right] \quad (3)$$

where, p_a is the acoustic pressure, p_0 is ambient pressure, p_v is the pressure in the bubble at its maximum size and T_0 is the ambient temperature.

The decrease in γ will cause the decrease in temperature inside the bubble according to equation 2 and hence reaction rate is expected to be lower. Fig. 3 compares the degradation rates for benzene and toluene for 50 ppmv initial concentration. Both being volatile and of similar nature, degradation follows the same pattern, as indicated by the plot (Fig.3).

3.2 Influence of dissolved gas

In addition to the vapors of volatile compound, the cavitation bubbles contain other gases, which are likely to have a significant effect on the cavitational collapse. To determine the effect of dissolved gas, the experiments were also carried out in argon saturated solution. Both the gases (Air and Argon) were bubbled for 1 hour before sonication (with intermittent shaking, to achieve proper mixing). No gas was added during the sonication, because of volatile nature of organics themselves. The reaction

1
2
3 was observed to be almost independent of the type of dissolved gas. In an argon
4 saturated solution, k_1 was estimated to be 0.0148 min^{-1} for benzene and 0.0125 min^{-1} for
5
6
7
8 toluene (at 50 ppmv).
9

10
11
12 The average specific heat ratio γ of the gas is an important parameter for the
13 degradation of organic compounds. In general, monoatomic gases like helium, argon,
14
15 krypton etc. are known to have the highest specific heat ratio ($\gamma = 1.67$) [14]. The final
16
17 collapse temperature for a monoatomic gas could be as much as twice the collapse
18
19 temperature for a triatomic gas. In fact, such wide variation in temperature makes the
20
21 specific heat of the gas a critical parameter. Besides, the sensitivity of the gas to
22
23 sonochemical degradation is also dependent on other factors such as thermal
24
25 conductivity and gas solubility.
26
27
28
29
30
31
32
33

34 The hot spot theory assumes adiabatic collapse of the bubbles. However, departure
35
36 from adiabatic collapse is likely to take place due to the heat conduction from the hot
37
38 bubble contents to the bulk liquid. The gas with low thermal conductivity is likely to
39
40 reduce the heat dissipation from cavitation site, thereby favoring high collapse
41
42 temperature. A highly soluble gas is expected to diffuse more easily into the bubbles
43
44 and cushion the implosion. Studies carried out herewith suggest that the reaction rate
45
46 is independent of the gas dissolved. This anomalous behavior might be due to the
47
48 higher solubility of argon (5.60 ml/100 ml H_2O) compared to the solubility of air (2.89
49
50 ml/100 ml H_2O) in water.
51
52
53
54
55
56
57
58
59
60

3.3 Influence of hydrogen peroxide

As seen in the fig. 4, the addition of hydrogen peroxide showed slight increase in the degradation rates for toluene, although it did not seem to alter the rate constant for benzene (Table 1). De Visscher and Langenhove [15] (1998) observed similar results for trichloroethylene (TCE) degradation, where they observed an increase in TCE degradation with the increase in H_2O_2 concentration. Teo et al. [16] also noted the increase in p-chlorophenol degradation with the addition of H_2O_2 .

This observation may be explained by the fact that benzene, toluene and TCE are highly volatile and hence have a strong tendency to diffuse into the gaseous bubbles. Thus, they are mostly degraded either in the interior of the bubbles or in the interfacial sheath, as supported by findings of Mahamuni and Adewuyi [17]. Therefore, the degradation does not seem to necessarily require OH^\bullet radicals, generated by the decomposition of hydrogen peroxide. On the other hand, since chlorophenol has a lower vapor pressure, it tends to remain in water and undergo oxidative degradation. Furthermore, the increase in hydrogen peroxide concentration would increase the concentration of hydroxyl radicals as well, which can attack the aromatic structures, thereby increasing the rates of reaction.

3.4 Influence of temperature

Generally for a chemical reaction following Arrhenius rate law, reaction rate increases with increase in temperature. However, sonochemical reaction rate depends on many factors, which causes irregularity in reaction rates. The material properties (namely,

1
2
3 viscosity, gas solubility, vapor pressure and surface tension) would also be affected by
4
5 the temperature. In literature also, there is hardly any report claiming consistency in
6
7 effects, caused by the impact of temperature on the decomposition of organic
8
9 compounds. Ku et al. [18] observed that decomposition was slightly lower at high
10
11 temperatures. Similarly, Ondruschka and Hoffmann [19] also indicated the increase in
12
13 sonochemical degradation of chlorobenzene with increase in temperature. On the other
14
15 hand, however, Jiang et al. [20] noted that p-chlorophenol degradation showed that
16
17 above 40 °C the degradation rate and temperature were inversely proportional whereas
18
19 below 40 °C the degradation rate increases with temperature. The present study clearly
20
21 indicated an increase in the degradation rate with increase in temperature. Fig. 5 and 6
22
23 present the decomposition of benzene and toluene respectively in the temperature
24
25 range 10-40 °C. The k_1 value is compared in Table 2.
26
27
28
29
30
31
32
33

34 The increase in temperature is known to increase the vapor pressure of the solution.
35
36 Consequently, the cavitation bubbles formed would be filled with water instantaneously,
37
38 leading to cushioning in the adiabatic collapse of the bubbles, and subsequent reduction
39
40 in the intensity of collapse (and hence reduction in the extremes of pressure and
41
42 temperature generated) [21]. This would cause the decrease in degradation rates.
43
44 However, according to Arrhenius law, the rate should increase with the increase in
45
46 temperature. Also, we know, the increased temperature is likely to result in the
47
48 reduction of gas solubility, which can have positive effect on degradation rate. In this
49
50 context, the experiments were chosen to be conducted at low concentration (50 ppmv)
51
52 because at higher concentrations, cavitation bubbles would be filled with large number
53
54
55
56
57
58
59
60

1
2
3 of gas molecules. This state of increase in size and number of bubbles, coupled with the
4
5 increased vapor pressure is likely to dampen the cavitation efficiency. On the other
6
7 hand, at low concentrations the decrease in cavitation efficiency could be offset by
8
9 increase in k .
10
11

12 13 14 15 *3.5 Influence of chloride ions*

16
17 As per the literature, sonochemical degradation has been known to be accelerated in
18
19 the presence of ions [22, 23]. In fact, ultrasonic irradiations normally result in the higher
20
21 mass-transfer, which is a function of specific parameters (viz. type and geometry of
22
23 reactors used for irradiation, gas flow rate, and power dissipation into the system) [24].
24
25 The presence of a salt (say sodium chloride) necessarily leads to enhanced mass
26
27 transfer rates (due to its non-combining nature in aqueous systems), thereby leading to
28
29 the formation and maintenance of smaller bubbles. The presence of salt has also been
30
31 found to decrease the vapor pressure of the medium and to increase the surface
32
33 tension, both resulting in more violent collapse of the cavities [24]. Besides, the
34
35 improvement in rate is also reported by Kumar et al. [23]. In fact, the use of Cl^- and
36
37 NO_2^- result in two to eight-fold enhancement in the rates of degradation of phenol and *p*-
38
39 nitrophenol. However, Gultekin and Ince, [25] had reported that the presence of ions is
40
41 less likely to increase the sonochemical degradation of bisphenol-A. Besides, the
42
43 addition of salt would also increase ionic strength of the aqueous phase, which can
44
45 drive the organic compounds to the bulk-bubble interface. Furthermore, dissolved
46
47 electrolytes also affect other properties such as viscosity and heat capacity. Thereby,
48
49 the resultant effect on degradation rate is observed to be non-linear and thus difficult to
50
51
52
53
54
55
56
57
58
59
60

1
2
3 estimate. Fig. 7 shows the effect of chloride concentration on degradation of benzene
4 and toluene, wherein a minor increase in the degradation rates was observed for
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

estimate. Fig. 7 shows the effect of chloride concentration on degradation of benzene and toluene, wherein a minor increase in the degradation rates was observed for volatiles. This may possibly be due to volatile degradation taking place in the interior of bubbles rendering the merit of presence of anion to be poor.

4. Conclusion

The application of ultrasound was studied for the decomposition of benzene and toluene.

The following conclusions can be drawn from the above study:

- 1) The reaction rate was observed to be a function of the initial concentration. It decreased with the increase in initial concentration. The volatile nature of compound causes the lowering of adiabatic collapse temperature at higher concentration.
- 2) The reaction rate was independent of the gas dissolved. The rate constant remains same for both argon and air saturated solution.
- 3) The addition of hydrogen peroxide did not seem to alter the reaction rates appreciably. It was almost constant for benzene, whereas toluene showed slight increase.
- 4) The reaction rate was found to be temperature dependent in the range 10-40 °C. Both, benzene and toluene showed increase in rate constant with respect to temperature.
- 5) The addition of anions like chloride results in minor increase in the degradation rate.

Nomenclature

p_a	Acoustic pressure (bar)
p_v	Pressure in the bubble at its maximum size (bar)
p_{\max}	Maximum temperature generated inside the bubble (bar)
T_{\max}	Maximum temperature generated inside the bubble ($^{\circ}\text{C}$)
k	First order rate constant (min^{-1})
T_0	Ambient temperature ($^{\circ}\text{C}$)
C_i	Initial concentration (mol/l)
K	Proportionality constant

Greek letters

γ	Specific heat ratio
----------	---------------------

CW Only

References

- [1] Lu M, Wu X and Wei X Chemical degradation of polyacrylamide by advanced oxidation processes, *Environmental Technology*, 33:9 (2012), 1021-1028
- [2] Li W, Lu S, Qiu Z and Lin K, UV and VUV photolysis vs. UV/H₂O₂ and VUV/H₂O₂ treatment for removal of clofibric acid from aqueous solution, *Environmental Technology*, 32:10 (2011), 1063-1071
- [3] Jalilnejad E., Mogharei A and Vahabzadeh F, Aerobic pretreatment of olive oil mill wastewater using *Ralstonia eutropha*, *Environmental Technology*, 32:10 (2011), 1085-1093
- [4] Shirzad Siboni M., Samadi M.T., Yang J.K. & Lee S.M., Photocatalytic reduction of Cr(VI) and Ni(II) in aqueous solution by synthesized nanoparticle ZnO under ultraviolet light irradiation: a kinetic study, *Env. Technol*, (2011), 1573-1579
- [5] Chen W-S and Huang G-C, Sonochemical decomposition of dinitrotoluenes and trinitrotoluene in wastewater. *J. Hazard. Mater*, (2009), 868–874.
- [6] Pétrier C, Torres-Palma R, Combet E, Sarantakos G, Baup S and Pulgarin C, Enhanced sonochemical degradation of bisphenol-A by bicarbonate ions. *Ultrason. Sonochem*, 17 (2010), 111–115.
- [7] Gogate P.R. and Pandit, A.B., A review of imperative technologies for wastewater treatment II: hybrid methods. *Adv. Env. Res.* 8 (2004): 553-597.
- [8] Sáez V, Esclapez M.D., Bonete P, Walton D.J., Rehorek A, Louisnard O and González-García J, 2011. Sonochemical degradation of perchloroethylene: The influence of ultrasonic variables, and the identification of products. *Ultrason. Sonochem.* 18: 104–113.
- [9] Ince N.H., Tezcanli G., Belen R.K. and Apikyan I.G, 2001. Ultrasound as a catalyzer of aqueous reaction systems: the state of the art and environmental applications. *Appl catal. B: Environ*, 29 : 167-176.
- [10] De Visscher A. and Langenhove H.V, 1998. Sonochemistry of organic compounds in homogenous aqueous oxidizing systems. *Ultrason. Sonochem*, 5: 87-92.

- 1
2
3 [11] Suslick K.S., Hammerton D.A. and Cline Jr. R.E, 1986. The sonochemical hot
4 spot., *J. Am. Chem. Soc.* 108: 5641-5642.
5
6
7 [12] Lee M and Oh J, 2010. Sonolysis of trichloroethylene and carbon tetrachloride in
8 aqueous solution. *Ultrason. Sonochem*, 17: 207–212.
9
10 [13] Jiang Y., Petrier C. and Waite T.D, 2002. Kinetics and mechanisms of ultrasonic
11 degradation of volatile chlorinated aromatics in aqueous solutions. *Ultrason.*
12 *Sonochem*, 9: 317-323.
13
14 [14] De Visscher A., Van Eenoo P., Drijvers D., and Langenhove H.V, 1996. Kinetic
15 model for the sonochemical degradation of monocyclic aromatic compounds *in*
16 *aqueous solution*. *J. Phys. Chem*, 100: 11636-11642.
17
18 [15] Noltinkg B.E and Neppiras E.A, 1950. Cavitation produced by ultrasonics. *Proc.*
19 *Phys. Soc*, 63B: 674-685.
20
21 [16] Riesz P. and Takashi F, 1992. Free radical formation induced by ultrasound and
22 its biological implications. *Free Rad. Biol. Med*, 13: 247-270.
23
24 [17] Teo K.C. Xu Y. and Yang C, 2001. Sonochemical degradation for toxic
25 halogenated organic compounds., *Ultrason. Sonochem*, 8: 241-246.
26
27 [18] Mahamuni N and Adewuyi Y.G, 2010. Advanced oxidation processes (AOPs)
28 involving ultrasound for waste water treatment: A review with emphasis on cost
29 estimation. *Ultrason. Sonochem*, 17: 990–1003
30
31 [19] Ku Y., Chen K. and Lee, K, 1997. Ultrasonic destruction of 2-chlorophenol in
32 aqueous solution. *Water Res*, 31: 929-935.
33
34 [20] Ondruschka B. and Hofmann J, 1999. Ultrasound in Environmental Engineering,
35 TUHH Reports on Sanitary Engineering, P-139.
36
37 [21] Mason T.J, 1991. Practical Sonochemistry user's guide to applications in
38 chemistry and chemical engineering, Ellis Horwood Series.
39
40 [22] Minero C, Pellizzari P., Maurino V., Pelizzetti E. and Vione D, 2007.
41 Enhancement of dye sonochemical degradation by some inorganic anions
42 present in natural waters. *Appl. Catal, B* 77: 308–316.
43
44 [23] Kumar, A. Gogate, P. R, Pandit, A. B., Delmas, H. Wilhelm, A.M, 2004.
45 Gas–Liquid Mass Transfer Studies in Sonochemical Reactors. *Ind. Eng. Chem.*
46 *Res*, 43: 1812-1819.
47
48
49
50
51
52
53
54
55
56
57
58
59
60

- 1
2
3 [24] Seymour, J.D. and Gupta, R.B., Oxidation of aqueous pollutants using
4 Ultrasound: salt-induced enhancement., Ind. Eng. Chem. Res. 1997; 36: 3453-
5 3457
6
7
8 [25] Gültekin I. and Ince N.H, 2008. Ultrasonic destruction of bisphenol-A: the
9 operating parameters, Ultrason. Sonochem, 15: 524–529.
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

For Peer Review Only

Table 1 First order rate constant at different H₂O₂ concentration for benzene (T = 20°C)

System	k₁ (min⁻¹)	r²
50 ppmv Benzene (Bn)	0.0146	0.985
50 ppmv Bn + 200 ppmv H₂O₂	0.0147	0.982
50 ppmv Bn + 400 ppmv H₂O₂	0.0147	0.98
50 ppmv Bn + 1000 ppmv H₂O₂	0.0146	0.984

Review Only

Table 2. First order rate constant at different temperature
(Initial concentration = 50 ppmv)

Temperature (°C)	Benzene		Toluene	
	k_1 (min ⁻¹)	r^2	k_1 (min ⁻¹)	r^2
10	0.0092	0.986	0.0087	0.982
20	0.0101	0.985	0.00906	0.979
30	0.0109	0.983	0.0102	0.98
40	0.012	0.983	0.0118	0.979

Review Only

Figure Captions

Fig. 1 Experimental set up.

Fig. 2 Effect of initial concentration on degradation of benzene at 20°C.

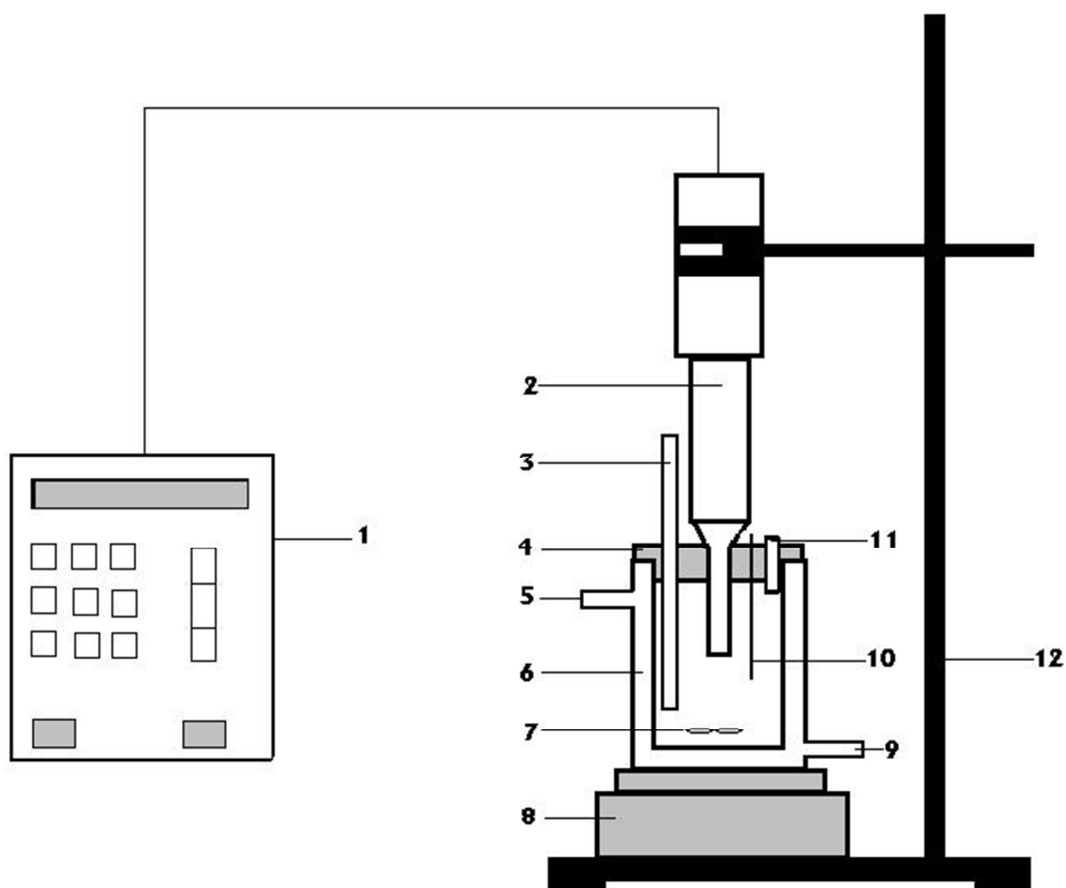
Fig.3 Comparing the degradation rate for benzene and toluene for 50 ppmv initial concentration.

Fig. 4 Effect of hydrogen peroxide on degradation of toluene at 20°C.
(initial conc. = 50 ppmv)

Fig.5 Effect of temperature on degradation of benzene at 50 ppmv initial concentration

Fig.6 Effect of temperature on degradation of toluene at 50 ppmv initial concentration

Fig. 7 Effect of Cl⁻ ions concentration on the degradation of benzene and toluene
(initial conc. = 50 ppmv)



- | | |
|-------------------------|-------------------------|
| 1. Probe system display | 7. Stirrer |
| 2. Ultrasonic probe | 8. Magnetic stirrer |
| 3. pH probe | 9. Cooling water inlet |
| 4. Teflon cover | 10. Thermocouple |
| 5. Cooling water outlet | 11. Septum for Sampling |
| 6. Water jacket | 12. Steel stand |

Fig. 1

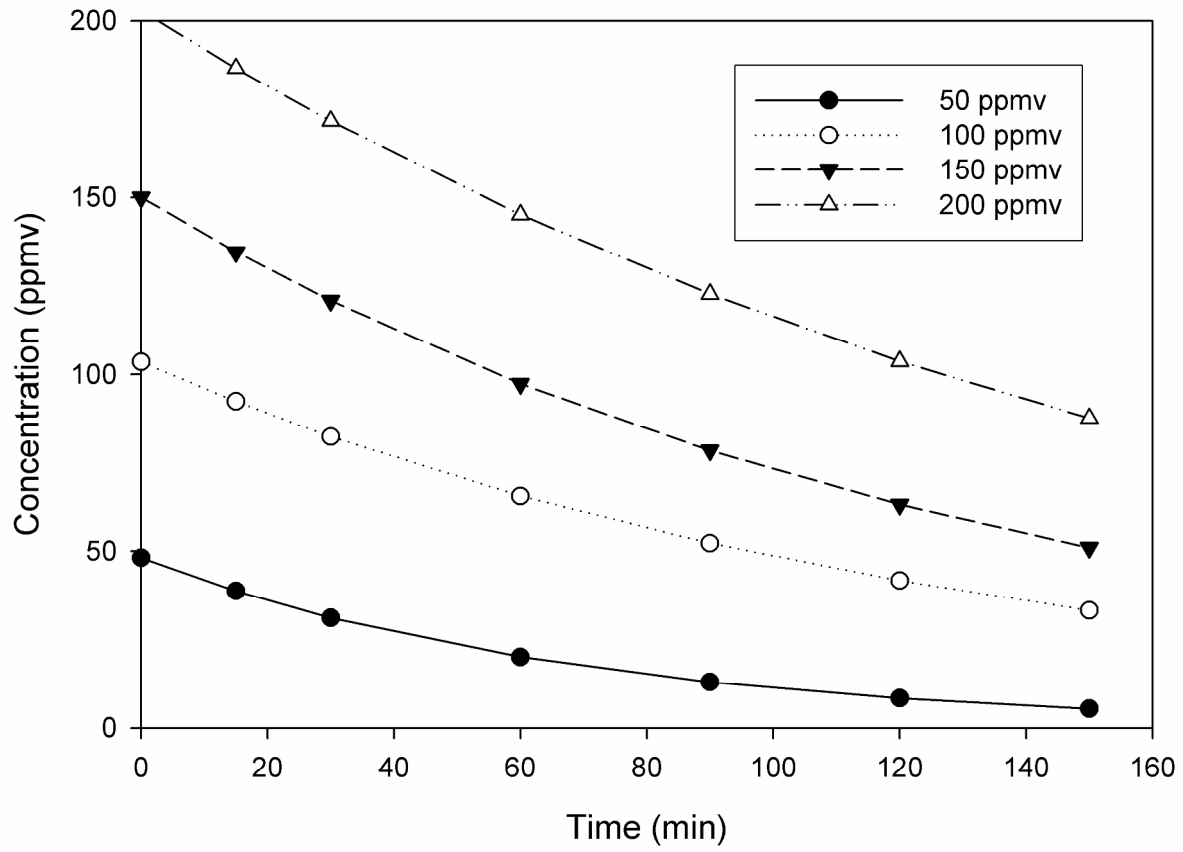


Fig. 2

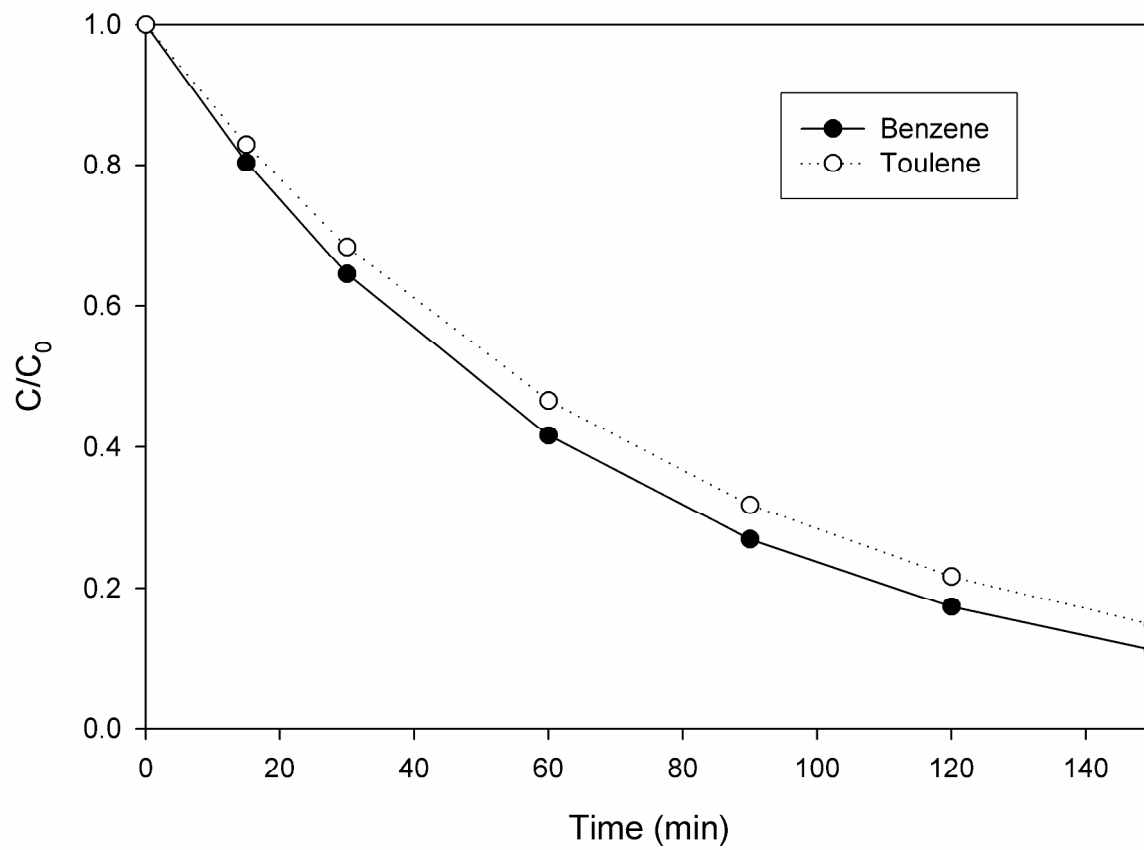


Fig. 3

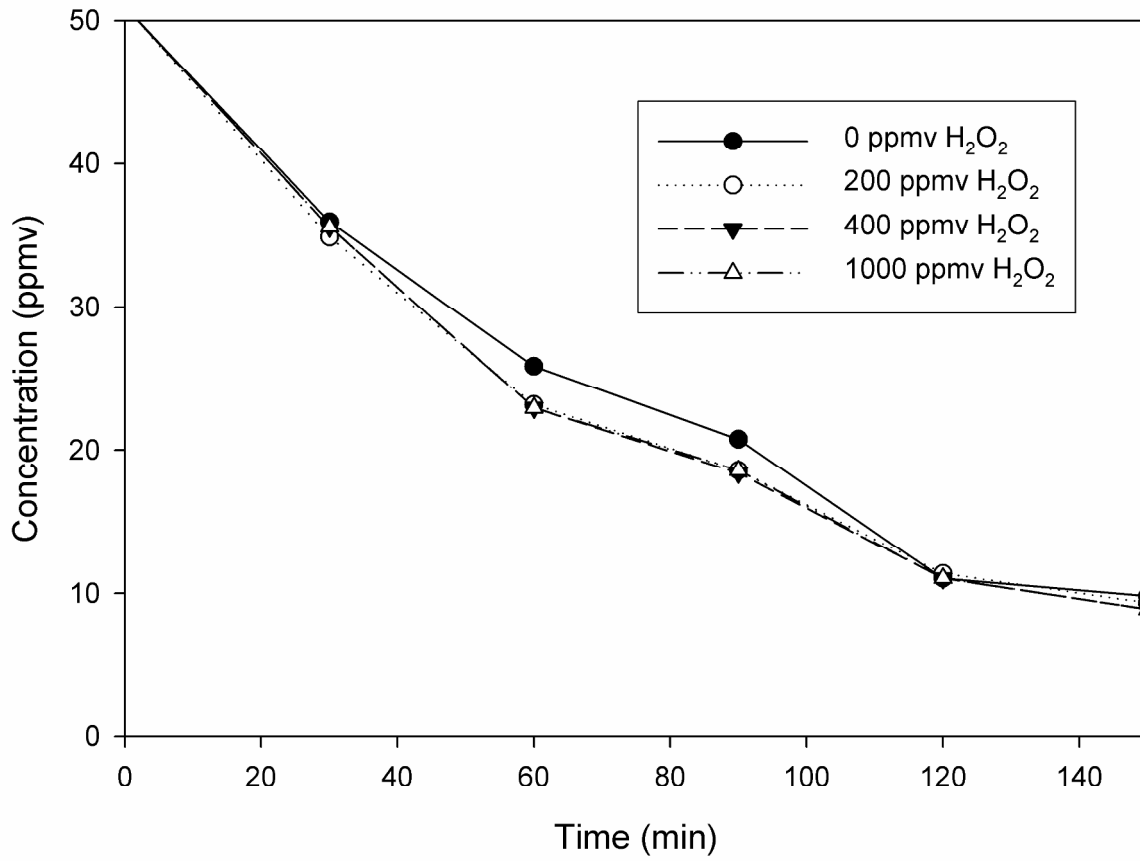


Fig. 4

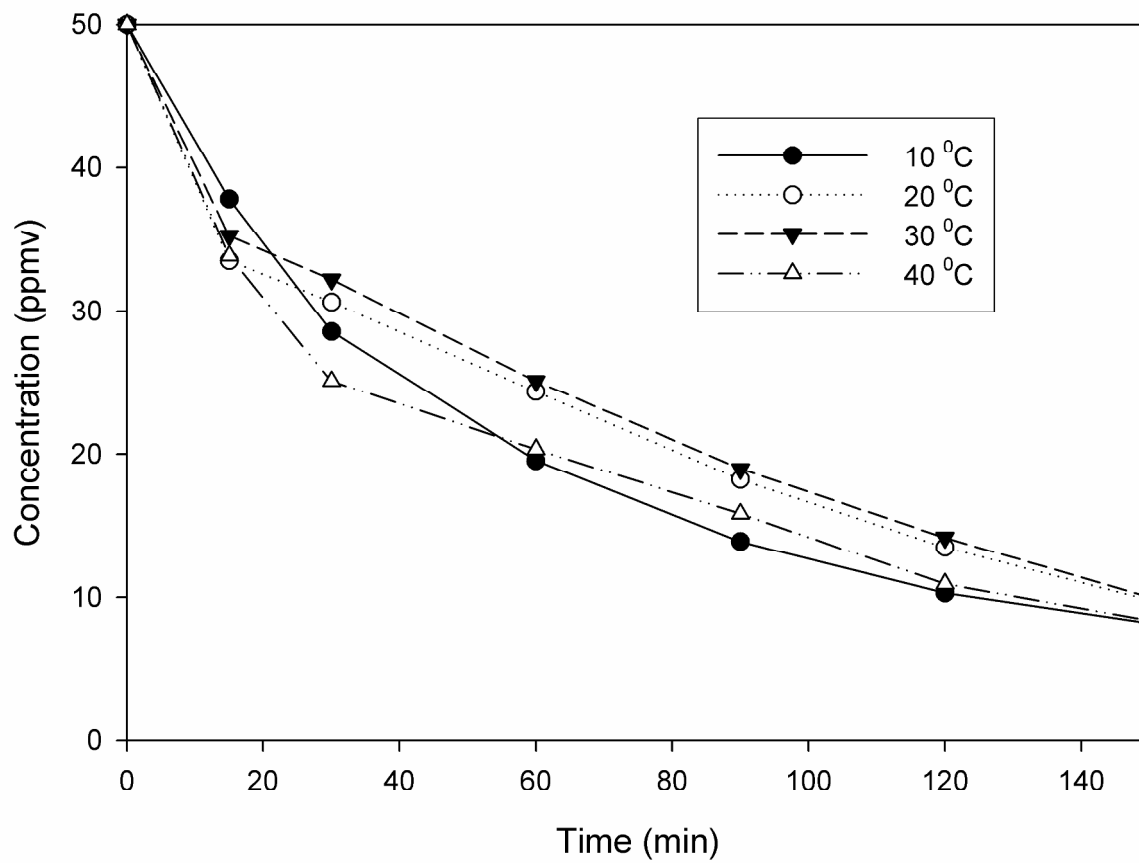


Fig. 5

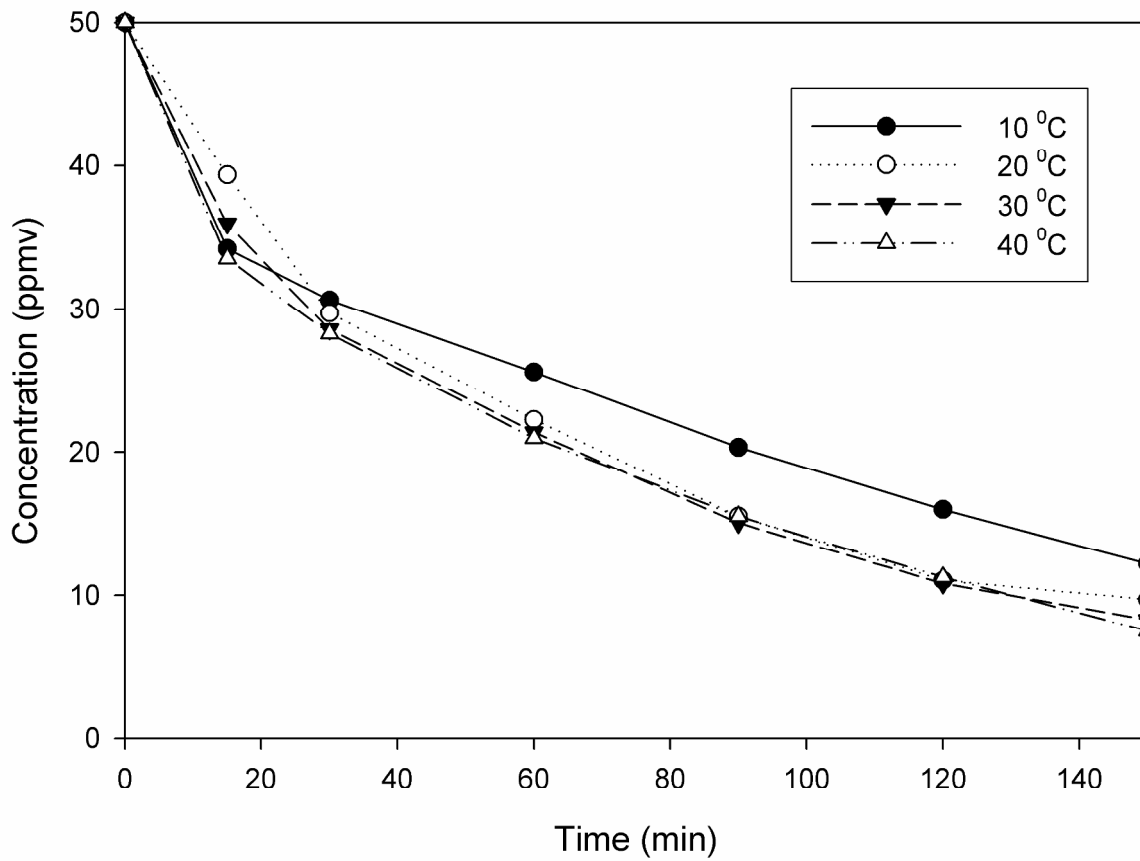


Fig. 6

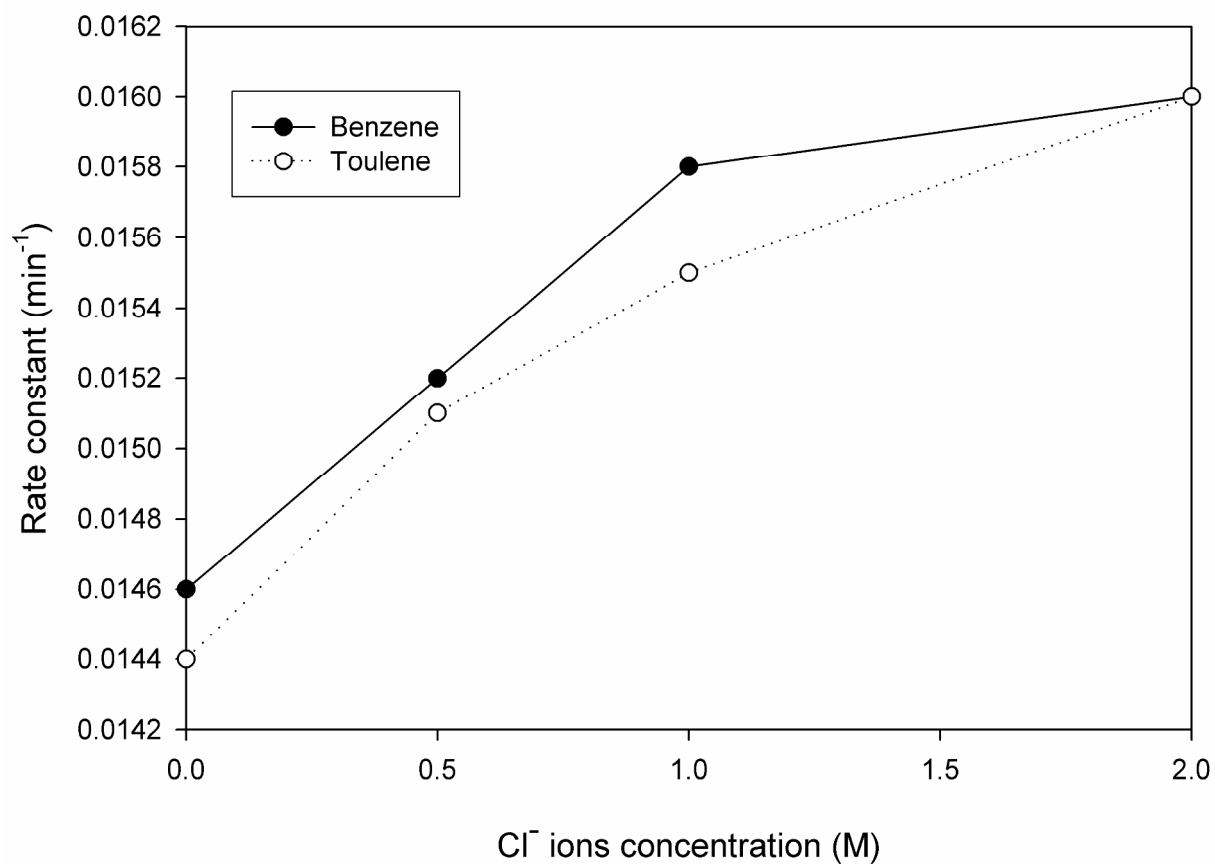


Fig. 7