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High resolution HPLC-MS confirms overestimation of urea in soil by the diacetyl monoxime (DAM) colorimetric method

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14 ABSTRACT

15 Urea represents a common form of organic nitrogen (N) which is added to agricultural soils in 16 large quantities in both cropping (e.g. fertiliser) and livestock (e.g. urine) systems. In addition, 17 there is a small, dynamic ambient pool of urea in soil associated with metabolic functioning in 18 the microbial community. The diacetyl monoxime (DAM) colorimetric method is routinely 19 used to quantify urea in soil, however, it lacks specificity due to the potential to react with the 20 ureido group (R_1 NHCONH R_2), a common structural moiety in soil organic matter. The aim of 21 this study was therefore to critically evaluate the accuracy of this method for urea determination 22 in soil. Using the DAM assay, we demonstrated significant cross-reactivity with a range of 23 ureido compounds, many of which are ubiquitous in soil. We conclude therefore that the DAM 24 assay is highly likely to overestimate urea concentrations in environmental materials. Such 25 overestimation was confirmed using high resolution HPLC-Orbitrap MS to quantify urea in 26 grassland soils using standard addition and the concentrations compared with those of the DAM 27 assay. The results obtained show the DAM colorimetric method overestimated urea 28 concentration by between 7.2 and 58 times for the sites studied. This significant overestimation 29 of urea emphasises the need to validate the colorimetric method with reference to the LC-MS 30 assay to ensure the robustness of measured urea concentrations. On this basis we recommend 31 that reporting of the results from the DAM colorimetric method as "urea" concentration be curtailed and reported as "ureido-N" to recognise the contribution of unknown and variable 32 33 contributions from other compounds. Indeed, given the problems with quantitatively assessing 34 the latter contributions we would recommend the DAM method is now avoided in surveys of urea concentrations in soil and the wider environment. 35

Keywords: Analytical method, Dissolved organic nitrogen, Nitrogen cycling, Urea
determination, Ureido-N

38

39 1. Introduction

40 Urea occurs ubiquitously in the environment, deriving from both natural and anthropogenic 41 sources. It is an important compound in key biological processes and metabolic pathways of 42 all living organisms (Berman and Bronk, 2003; Remsen, 1971). Due to the high abundance of 43 the microbial enzyme urease and plant uptake mechanisms, urea is rapidly turned over in soils 44 under ambient conditions, resulting in a small but dynamic urea pool, which is an important 45 part of the soil N-cycle (Harder Nielsen et al., 1998; Lloyd and Sheaffe, 1973; Solomon et al., 2010). Elevated urea concentrations in the environment can indicate anthropogenic inputs, 46 47 predominantly from agriculture (as fertiliser or from livestock urine) and wastewater treatment systems (Galloway and Cowling, 2002; Glibert et al., 2006). It is important to determine these 48 49 anthropogenic inputs into the environment to assess human impacts. High concentrations of 50 soluble low molecular weight nitrogen (N) compounds, such as urea, can have detrimental 51 impacts on the environment, such as eutrophication and denitrification, leading to the release 52 of the greenhouse gas, nitrous oxide (N₂O) (Galloway and Cowling, 2002). The latter also 53 catalyses the production of ozone in the troposphere, which damages agricultural crops and 54 natural ecosystems, alongside negative impacts on human and animal health (Galloway and 55 Cowling, 2002; Heil et al., 2016). Such negative impacts on ecosystems, and on human and 56 animal health, alongside the increasing importance of urea globally as a fertiliser, means it is 57 important to quantify N inputs, particularly the urea component, into the soil environment 58 (Gilbert et al., 2006).

There are two indirect methods commonly used to quantify urea in environmental matrices: enzymatic and colorimetric. The enzymatic method uses the bacterial enzyme urease to convert urea to ammonium, which is subsequently quantified using other enzymatic or colorimetric methods (McCarthy, 1970). A drawback of this method is the considerable background levels of ammonium in the environment (i.e. high blanks). Furthermore, the urease 64 inhibitors used in combination with urea fertilisers in agriculture reduce the accuracy of these65 determinations (Lambert et al., 2004).

The diacetyl monoxime (DAM) colorimetric method provides an often-used alternative 66 67 for urea detection and quantification in both biological and environmental matrices (Douglas and Bremner, 1970; Stepanauskas et al., 2000; Taylor and Vadgama, 1992). The latter method 68 69 is preferred for environmental matrices due to its sensitivity, with modified methods able to detect urea down to 0.03 µM N concentrations (Alam et al., 2017; Goeyens et al., 1998). It has 70 71 been applied to a wide range of environmental matrices, including river water (Burford et al., 72 2011; Revilla et al., 2005; Satoh and Katoh, 1989; Stepanauskas et al., 2000), seawater (Cochlan and Bronk, 2001; Mulvenna and Savidge, 1992; Remsen, 1971), estuarine waters and 73 74 sediments (Alam et al., 2017; Nakas and Litchfield, 1977; Torres-Valdés and Purdie, 2006; 75 Twomey et al., 2005), rainwater (Cornell et al., 1998), and soil extracts (Daigh et al., 2014; 76 Greenan et al., 1995; Sullivan and Havlin, 1991).

77 Interestingly, the DAM assay was first developed for the determination of the 78 amino acid citrulline and subsequently adapted to detect urea, diacetyl. The assay relies 79 on the production of diacetyl from diacetyl monoxime under acidic conditions, which 80 reacts with urea to form an unknown chromogen, by an unknown mechanism, which is 81 detected at 520 nm (Archibald, 1945; Butler and Walsh, 1982; Fearon, 1939; Lugosi et al., 1972; Veniamin and Vakirtzi-Lemonias, 1970). The reaction has been adapted for 82 83 automation using continuous flow analysis and flow injection analysis (FIA), enabling 84 high sample throughput and high reproducibility, at low reagent costs. However, such colorimetric methods are known to be non-specific, and the non-specificity of this 85 86 method for urea detection has been noted previously in biological fluids (Archibald, 1945). However, the consequences of this have not yet been investigated in soil. 87 Critically, this method is commercially marketed as a method to detect urea in 88

89 environmental samples, with few caveats to acknowledge potential interference or 90 cross-reactivity in such settings. Other colorimetric methods, such as the molybdenum 91 blue method for orthophosphate have addressed the potential for interference, e.g. 92 known silicate interference prompted method development to address this (Murphy and Riley, 1962; Neal et al., 2000). In the urea method, no such adaptation is currently 93 94 available, and the specificity/non-specificity of the method for determination of urea in 95 solution is undefined. Significantly, citrulline and alloxan are expected to cross-react 96 under the reaction conditions and adaptions of the DAM colorimetric method have been 97 used to quantify these compounds in biological fluids (Archibald, 1945; Fearon, 1939; 98 Satoh and Katoh, 1989). Citrulline interference has been evaluated in natural waters 99 with the chromogen produced with the DAM colorimetric reaction conditions separated 100 from urea by liquid chromatography (LC), however, the question of non-specificity with 101 other compounds in relation to determined urea concentrations was not addressed (Satoh 102 and Katoh, 1989). As far as we are aware, a thorough investigation of environmentally 103 relevant compound classes cross-reacting under the reaction conditions used to produce 104 a response at 520 nm has not been undertaken.

105 One approach to assessing the potential for cross-reactivity in the DAM assay would 106 be to submit known concentrations of candidate compounds to the assay and determine their 107 responses relative to known concentrations of urea determined under the same assay 108 conditions. An alternative approach would be to employ an independent reference assay for 109 urea and compare the results of this assay to the results of DAM based analyses. The reference 110 assay of choice is HLPC-MS, with the latest generation of high resolution (HR) MS 111 instruments, e.g. Orbitraps, offering the possibility for assessing urea concentrations in 112 complex biological and environmental matrices at previously unattainable selectivity and specificity (Clark et al., 2007; Kind et al., 2007; Warren, 2014). Use of such an HPLC-HRMS 113

114 method for urea in the environment would allow unambiguous identification and 115 quantification, facilitating rigorous evaluation of the DAM colorimetric method.

116 Herein, we assess the potential for a range of potentially cross-reacting compounds to 117 interfere in the "off-the-shelf" DAM colorimetric method. The compounds chosen are likely to be present in environmental matrices and represent a range of compound classes that will allow 118 119 a better understanding of the reasons underlying the non-specificity of this method. The results 120 obtained are assessed in relation to previous determinations undertaken for biological matrices, 121 but which up to now are lacking for environmental matrices likely to contain cross-reacting compounds. A reference HPLC-HRMS assay is then described which allows direct 122 123 determination of urea in soil extracts. This allowed rigorous assessment of the degree of 124 overestimation by the DAM colorimetric method and demonstrated an effective reference 125 method is vital for cross-validation of the less specific and selective automated and low-cost colorimetric method. 126

127

128 **2. Methods**

129 2.1. Colorimetric detection of urea

Urea concentrations were determined using the Skalar San⁺⁺ system (Skalar Analytical 130 B.V., The Netherlands), a continuous flow analyser with a series 1050 sampler and data 131 collected and analysed on the San⁺⁺ FlowAcessTM V3 data acquisition Windows® software 132 133 package. All reagents and standards were from Merck (Germany) and were prepared according 134 to Skalar method for the analysis of urea (Catnr 612-001 issue 080714/MH/99290607). A urea stock standard solution was prepared monthly (100 mg N l⁻¹) and working standards prepared 135 daily (0-1.0 mg N l⁻¹). The colour reagent was prepared with diacetyl monoxime (DAM; 4.125 136 g l⁻¹), semi carbazide hydrochloride (0.05 g l⁻¹), manganese (II) chloride (MnCl₂.4H₂O, 13.075 137 g l⁻¹) and potassium nitrate (KNO₃, 0.25 g l⁻¹). A saturated sodium chloride solution was also 138

prepared with Brij 35 (3 ml l^{-1}). The acid reagent was prepared from sodium dihydrogen phosphate (NaH₂PO₄.H₂O, 50g l^{-1}) and sulfuric acid (H₂SO₄, 97 % ν/ν , 588 ml l^{-1}). All analyses were carried out in triplicate.

142

143 2.2. Analysis of standards to identify cross-reactivity

144 A range of commercially available compounds were used to test the specificity of the DAM colorimetric method. Initial standard solutions were prepared to identify compounds 145 146 which showed a positive response. Nitrogenous compounds tested were: allantoin, xanthine, 147 hypoxanthine, biuret, alloxan monohydrate, glucosamine, galactosamine, mannosamine, 148 adenine, cytosine, guanine, thymine and uracil (Sigma-Aldrich, UK). Sugars, including ribose, 149 glucose, sucrose, fructose, rhamnose, lyxose, xylose, arabinose, galactose and fucose (Acros 150 Organics, UK) were also tested. Other compounds tested included myo-inositol and oxalic acid (Sigma Aldrich, UK). Solutions at a concentration of 2 mg l⁻¹ were prepared in purified water 151 152 (Millipore). Standard solutions were analysed using the described DAM colorimetric method 153 and compounds which showed a positive response (allantoin, alloxan, biuret, hydantoin), or 154 interference (xanthine, hypoxanthine, adenine, cytosine, guanine, thymine, uracil, fructose) were analysed at a range of concentrations 0.01–10 mg N l⁻¹ or 0.1-10 mg l⁻¹, prepared weekly 155 from stock solutions (100 mg N l^{-1} / 100 mg l^{-1}). Whilst this is higher than expected 156 157 concentrations of such compounds in environmental matrices, this concentration range was 158 selected to ensure compounds exhibiting cross-reactivity would be identified and the degree of 159 response over a range of concentrations detectable once cross-reactivity had been identified.

160

161 2.3. UV-Vis absorbance

162 UV-Vis absorption measurements of the standard solutions (concentrations shown in
163 table 1) were carried out on a Cary 60 UV-Vis spectrophotometer (Agilent Technologies Inc.,

US) between 200-800 nm and recorded with Cary WinUV Software (Agilent Technologies
Inc., US). The absorption spectra were background corrected, and instrument wavelength
accuracy confirmed using a holmium perchlorate standard (Starna Scientific, USA).

167

168 2.4. Sample collection and extraction

169 Soil samples were collected from eleven grassland sites in the UK (see table S1 for further site information). Three replicate soil samples from each site were collected at random, 170 171 bulked and stored at 4 °C until extraction. Prior to extraction, plant material, macrofauna and 172 stones were removed by sieving (2 mm). Triplicate soil samples (2 g) were treated with double 173 distilled water (soil: double distilled water (DDW) ratio of 1:10 w/v) and shaken at room 174 temperature (1 h at 400 rev min⁻¹). Following this, the soils were centrifuged (10 min, 1690 x 175 g), the supernatant was removed, and residues washed with a further 5 ml of DDW and the 176 supernatant combined to give an overall soil to extractant ratio of 1:12.5 w/v. Soil extracts were 177 stored at 4 °C for 24 h prior to analysis, and were diluted with purified water (Millipore) at a 178 ratio of 1:2 w/v soil extract to DDW for colorimetric urea determination.

179

180 2.5. Reference LC-MS assay

181 To directly determine urea concentration in soil extracts, a reference LC-MS assay was 182 used. The LC separation was performed using a 150 mm x 2.1 mm i.d., 1.7 µm Synchronis HILIC column (Thermo Scientific) operated by a Dionex Ultimate 3000 HPLC system 183 184 (Thermo Scientific). The mobile phase comprised of solvent A, 0.1 % (v/v) formic acid in water and solvent B, 0.1 % (ν/ν) formic acid in acetonitrile (ACN). The LC mobile phase was held at 185 80 % B for 2 min, followed by a linear gradient to 20 % B (15 min), 1 min ramp to 80 % B and 186 a 15 min re-equilibration at 80 % B. The flow rate was 300 µl min⁻¹, column temperature was 187 188 30 °C and injection volumes were 20 µl. The column was directly interfaced to an Orbitrap

189 Elite mass spectrometer (Thermo Scientific). The FTMS was operated in full scan mode (m/z) 190 50 to 400; 60,000 mass resolution) and to improve sensitivity to urea, the ion trap was set to 191 scan a range of m/z 60 to 62. The ion source parameters were: sheath gas (N₂) flow 60 arbitrary 192 units (arb), the auxiliary gas (N_2) flow and the sweep gas (N_2) flow both set to 0 arb. The 193 capillary temperature was 275 °C. The electrospray was in positive ionisation mode, with a 194 source voltage of 3.00 kV. The data was acquired and analysed using Thermo Xcalibur (version 195 3.0). To ensure instrument performance, urea standards (80 % ACN; 0.1 to 5 mg l⁻¹) were analysed and subsequent analyses only accepted when the calibration graph had $R^2 > 0.99$. 196 Between each triplicate analysis of samples, a urea standard (1 mg l⁻¹) was analysed to check 197 198 instrument performance.

The soil extracts were prepared as above, and 5 ml sub-samples concentrated 10-fold using a stream of nitrogen, and brought to 80 % ACN. Quantification of urea was achieved using standard addition as ion suppression interfered with quantification using an external calibration curve. A urea standard in purified water (Millipore) was prepared (8 mg ml⁻¹) and added to the concentrated extracts to give three spike concentrations (0.4, 2.0 and 4.0 mg l⁻¹) to allow quantification by standard addition and analysed in triplicate.

205

206 2.6. Statistical analysis

All statistical analyses were performed using SigmaPlot 13.0 (Systat Software Inc.). For all statistical analyses and regressions, data was tested for normality (Shapiro-Wilk) and homogeneity of variance (Brown-Forsythe). A one-way ANOVA was used to evaluate differences between treatments and the significance level was set at $P \le 0.05$ for all statistical analyses.

212

213 **3. Results**

214 *3.1. Cross reactivity and interference of standards with the DAM colorimetric method*

215 The reference compounds (as shown in Table 1) were analysed at concentrations above expected environmental concentrations to determine any cross-reactivity under the 216 217 experimental conditions for the DAM colorimetric determination of urea. Compounds which 218 showed a response at this concentration (allantoin, alloxan, biuret, fructose, hydantoin, sucrose) were subsequently analysed at a range of concentrations (0.1 to 10 mg N l⁻¹ or mg l⁻¹ compounds 219 220 containing no N) to determine the linearity and degree of response under the experimental 221 conditions. The other compounds tested did not exhibit cross-reactivity or interference with the 222 detection of urea by this method, in agreement with Douglas and Bremner (1970).

223 The purine and pyrimidine derivatives (hypoxanthine, xanthine, adenine, uracil, 224 thymine, guanine and cytosine) were analysed separately at concentrations between 0.01 to 5 mg N l⁻¹. It was determined that this group of compounds did not exhibit a linear response 225 under the experimental conditions, however, they appeared to cause baseline instability. A 226 227 similar baseline instability was observed for the sugars fructose and sucrose (0.1 to 10 mg l⁻¹), 228 which had been previously observed (Prescott and Mangnall, 1976). It was not possible to 229 quantify the interference, as the baseline interference caused by these compounds was 230 irreproducible over triplicate determinations and separate analytical runs. Allantoin, alloxan, 231 biuret and hydantoin showed a linear response with increased concentration under the 232 conditions used in the DAM colorimetric determination of urea, as shown in Fig. 1. The 233 response of these compounds relative to urea at the same concentrations varies across the 234 concentration range tested as shown in Table 2.

To confirm there was no intrinsic absorbance at 520 nm for the selected standards which would result in a response under the analysis conditions without cross-reaction with the reagents, the reference compounds were analysed using UV-Vis at the concentrations shown in Table 1. There was no absorbance observed at 520 nm for any of the compounds tested, indicating no intrinsic interference at this wavelength. Therefore, any interference or crossreactivity observed with these compounds is solely due to a response generated by the analytical reagent conditions.

242

243 *3.2. Stability of colour reagent*

244 During analysis of standards to determine cross-reactivity and interference, it was noted that there was a decrease in determined concentration over the course of a week (as shown in 245 246 Fig. S1). Therefore, repeated determinations over the course of one week with the same colour 247 reagent and working reference compounds made up daily were undertaken to assess the change 248 in response with colour reagent age for urea and the cross-reacting compounds. For urea, the 249 decrease in response after 4 d was within analytical error (less than 0.01 mg N l⁻¹), however, 250 for allantoin and hydantoin, the decrease in response on day 4 was greater than the acceptable 251 analytical error ($\pm 0.05 \text{ mg N}$ l⁻¹ for allantoin and alloxan; $\pm 0.01 \text{ mg N}$ l⁻¹ for hydantoin). For 252 allantoin, alloxan and hydantoin, an analysis of variance showed the effect of reagent age on 253 the determined concentration was significant (one-way ANOVA; P < 0.05), although the 254 determined concentrations for urea and biuret did not differ significantly with reagent age (P > 0.05). The analyses were repeated using colour reagent prepared daily and no decrease in 255 256 response was observed after 4 d, indicating the colour reagent degraded over time and must be 257 prepared daily to ensure there is no decrease in response due to reagent degradation.

258

3.3. Urea concentration in soil extracts determined using the DAM colorimetric method and
LC-MS

Urea was directly detected in soil extracts using LC-MS following HILIC separation, eluting at 2.1 min, after the initial column break-through. Fig. 2 shows a typical extracted ion chromatogram at m/z 61.08 (±0.1 Da), the [M+H]⁺ ion for urea, in a soil extract. Urea was added at three concentrations to enable urea quantification in the soil extracts by standard addition (Fig. 2e). It was determined that the peak area for the $[M+H]^+$ ion was linear with respect to urea concentration for both urea standards (concentration range from 0 to 12 mg l⁻¹) and in spiked soil extracts, therefore it was not necessary to use other ions (such as m/z 121.2 $[2M+H]^+$) for quantification.

269 The determined urea concentrations in soil extracts for the DAM colorimetric method and the LC-MS assay are shown in Table 3. Determined urea concentrations for the DAM 270 colorimetric method were between 1.5 (S.D. 0.08) and 17.5 (S.D. 0.1) µg g⁻¹ soil whilst the 271 272 urea concentration determined by the LC-MS method ranged between 0.067 (S.D. 0.0007) and 0.71 (S.D. 0.01) µg g⁻¹ soil. To evaluate the degree of overestimation observed across the sites, 273 274 the fold overestimation was calculated (Table 3) for the eleven grassland soils studied. UV-Vis 275 determinations for soil extracts (soil-extract solution 1:12.5 w/v) showed no significant 276 absorbance at 520 nm, indicating no intrinsic absorbance at the wavelength at which the 277 chromogen is measured for the DAM colorimetric method. The limit of detection (LOD) for the LC-MS method was determined to be $0.1 \text{ mg } l^{-1}$, in line with similar LC-MS methods for 278 279 urea (note soil extracts were concentrated 10 fold prior to analysis, thus the LOD accounting for the concentration step is 0.01mg l⁻¹; Warren, 2014). 280

281

282 4. Discussion

283 4.1. Cross-reactivity and interference with the DAM colorimetric method

The compounds tested for cross-reactivity and interference with the DAM colorimetric method are all compounds which might be expected in environmental matrices from a range of natural and anthropogenic sources (Table S2 shows the variety of compound classes investigated). Interference caused by purine and pyrimidine bases, found in the environment in the form of DNA, RNA, ATP, ADP, etc. and purine degradation products reduce base-line stability and therefore the reliability of urea concentrations determined by this method for environmental matrices. It is not practically possible to quantify the interference in environmental matrices, due to varying concentrations of the interfering compounds. Interference at less than 10 times the concentration of urea N has not been previously reported (such as for uracil) indicating different reaction conditions (e.g. colour reagent concentration, stabilising agent and acid reagent composition) may offer an opportunity to reduce this baseline instability using continuous flow and flow injection analyses (Price and Harrison, 1987).

296 Current protocols for the "off-the-shelf" method suggest the colour reagent is stable for 297 one week, however, we have demonstrated here that colour reagent degradation occurs within 298 3 days of preparation, and a decrease in response is observed for cross-reacting compounds. 299 The observed decrease in response for analyses of urea standards is within acceptable analytical 300 error for this compound. However, it should be noted, the observed drop in the determined 301 concentration for cross-reacting compounds is greater than the acceptable analytical error (0.05 302 mg N l⁻¹). The concentration of cross-reacting compounds in environmental matrices is 303 unknown, and variable, therefore the changing response of these compounds with colour 304 reagent age reduces the reproducibility of ureido-N concentrations determined by the 305 colorimetric method. Therefore, it is recommended that the colour reagent is made up daily 306 and used for a maximum of 24 h to prevent degradation of the reagent to an extent that results 307 in variation in the measured urea (or ureido) concentration in environmental matrices.

308 Cross-reactivity for this method has been demonstrated before in biological samples 309 (Archibald, 1945, 1944; Fearon, 1939), however, the degree of cross-reactivity for the "off-310 the-shelf" DAM colorimetric method has not been assessed for these compounds and others of 311 the same compound class, which this work has undertaken. The key structural moiety present 312 in these compounds is the ureido group (highlighted in Fig. 3). Compounds with the general 313 structure $R_1NHCONHR_2$ will react with diacetyl monoxime under acidic conditions to yield a 314 chromogen which absorbs at 520 nm. The degree of absorption at this wavelength depends on 315 the substituents R₁ and R₂. It is important to note that whilst the chromogen structure has been 316 hypothesised (e.g. substituted 1,2,4-triazine (Beale and Croft, 1961; Lugosi et al., 1972), 317 imidazole ring (Lugosi et al., 1972; Veniamin and Vakirtzi-Lemonias, 1970), or a skipped 318 diene susceptible to oxidation to generate a carbonium ion under acidic reaction conditions 319 (Butler and Walsh, 1982)) it has not been confirmed. Elucidating the mechanism of chromogen 320 formation and structure in future research may help understanding the relative response of 321 cross-reacting ureido compounds compared to urea.

322 The ureido group is a common feature in biological compounds, such as citrulline and 323 allantoin, shown in Fig. 3b, accounting for between 4-8 % of ruminant urinary-N (Bristow et 324 al., 1992). Furthermore, compounds containing the ureido group are also used in agriculture, 325 with biuret an impurity in urea fertiliser and utilised as a non-protein nitrogen (NPN) in animal feed and hydantoin derivatives applied as pesticides and fungicides (e.g. miprothin and 326 327 iprodione). Allantoin, biuret and hydantoin can all be expected in inputs from agricultural 328 settings, alongside urea, and therefore will contribute to urea concentrations determined by the 329 DAM colorimetric method due to cross-reactivity in such sites. They are also likely to be 330 available to the biota to support primary production, in much the same way as urea is. The size 331 and composition of the ureido-N pool will vary with different land uses due to different inputs 332 of ureido compounds (such as an arable crop, which may receive urea inputs as fertiliser, 333 compared to a grazed grassland, which would have urea and allantoin inputs from ruminant 334 urine). Subsequently, the different levels of response relative to urea of ureido compounds (Table 2) means there will be varying degrees of overestimation of urea depending on the 335 336 concentration of individual compounds within the ureido-N pool, which extends beyond the 337 compounds identified in this study due to the ubiquity of this moiety in both natural and anthropogenic compounds. Therefore, a direct method of determination for urea was required 338

to confirm the presence of cross-reactivity by comparison of urea concentrations determined
by the colorimetric method and a direct LC-MS assay. Such analyses are needed to ensure the
commercial "off-the-shelf" method has been sufficiently validated for the environmental
application (Glibert et al., 2006).

- 343
- 344 4.2. LC-MS assay confirms overestimation of urea

The LC-MS assay described enabled direct determination urea in environmental 345 346 matrices (using standard addition) with no interference as observed with the DAM colorimetric 347 method. This provides a specific determination for urea, enabling investigation into the degree 348 of overestimation of urea by the DAM colorimetric method. This is illustrated in Fig. 2, where 349 the extracted ion chromatogram (m/z 61.08) for urea at three spike concentrations was used to 350 construct the standard addition calibration curve and determine the urea concentration in the 351 soil extracts. It was necessary to quantify using standard addition due to the presence of ion 352 suppression in the area of elution in the chromatogram for urea. This was confirmed by spiking a soil extract with urea at a concentration equivalent to 12 mg l^{-1} with a spike recovery of 18 % 353 354 (S.D. 3.2 %) indicating suppression of MS response. Ion suppression is widely recognised phenomenon which arises in the analysis of complex mixtures due to the presence of non-355 356 volatile analytes which reduce the efficiency of ionisation for the analyte of interest (in this 357 case urea), Quantification via standard addition is advantageous as it can correct for ion 358 suppression as the same sample matrix is present and a calibration is based on the known added 359 urea concentration. Since ion suppression affects the analyte and calibrant urea equally the 360 absolute ratio is maintained, hence, accurate quantification of urea in soil extracts is achievable. 361 Overestimation by the DAM colorimetric method is confirmed using the LC-MS reference assay, which detects urea based on m/z 61.08, due to lower concentrations observed 362 by the direct method compared to the indirect method (Table 3). The degree of overestimation 363

364 varies across the grassland sites, with the urea concentration determined by the DAM 365 colorimetric method between 7.2 and 58.9 times higher than the directly determined urea 366 concentration using LC-MS. The degree of overestimation is not consistent with land cover (all 367 sites were grasslands), soil texture or geographical location, and it is likely this variation in 368 overestimation can be attributed to the varying composition of the ureido pool in soils, in both 369 concentration and compounds. This has implications for our understanding of the importance of urea in soils as it is overestimated to different degrees in different settings and therefore urea 370 371 concentrations determined by this method are not comparable, unless reported as ureido-N. 372 Possible implications of this would be overestimation of the concentration of anthropogenic 373 urea, or underestimation of urea utilisation by soil microbes. Ideally, it would be useful to 374 characterise the nature of the of the ureido compounds present, however, this is impractical in 375 such a complex matrix. It was for this reason we undertook to use known natural and 376 anthropogenic compounds, representative of those likely to be present, to demonstrate that it is 377 the ureido group that is responsible for DAM cross-reactivity. Other compounds with the 378 ureido group include pharmaceuticals, mammalian metabolites, antibacterial and antifungal 379 agents in personal care products and plastics, insecticides, pesticides and food products 380 (Brausch and Rand, 2011; Lewis et al., 2016). Due to the widespread occurrence of the tested 381 compounds and related ureido compounds, it can be expected a number of ureido compounds 382 will be present in environmental matrices, contributing an unknown and variable degree to urea 383 concentration when analysed using the DAM colorimetric method.

384

385 *4.3. Recommendations for the DAM colorimetric method for urea determinations*

The results presented above highlight considerable concerns in the use of the DAM colorimetric method for urea determinations in environmental matrices. Our findings using model compounds unequivocally confirm the cross-reactivity of the ureido group, a common 389 moiety in natural and anthropogenic compounds, alongside interferences from sugars and 390 purine and pyrimidine derivatives. The colorimetric method should be considered a useful 391 technique to quantify ureido-N as it is an inexpensive and quick tool (extraction and analysis 392 of 60 soil extracts per day) relative to the more intensive and time-consuming LC-MS method (extraction and analysis of *ca.* 16 soil extracts per day as 3 analytical runs are required per soil 393 394 for standard addition). The type of studies where this would be relevant include studies of the 395 fate of added urea on a temporal or spatial scale, where the cross-reactivity can be accounted 396 for by proper use of controls. It is also a useful tool for comparison of the ureido-N pool 397 between natural systems, which is an environmentally relevant and biologically available pool. 398 This is also the case for existing studies comparing urea in soil extracts, which should be 399 considered as reporting ureido-N due to the unknown contribution of this pool to the reported 400 urea concentration.

Given the ubiquitous occurrence of these compounds containing the ureido group in the environment the DAM method is an inappropriate choice for quantitative analysis of urea in environmental matrices. Due to the ubiquity and unpredictability of ureido and interfering compounds in environmental samples it is unfeasible to identify and quantify their contributions to the urea concentrations determined by the DAM method, negating any ideas of applying "corrections". It is therefore suggested that:

407 (i) Results reported from the DAM colorimetric method should include a caveat
 408 recognising the presence of cross-reactivity and interferences in environmental
 409 matrices, reporting determined concentrations as "ureido-N".

(ii) It is vitally important to monitor the stability of reagents used in the DAM colorimetric
method since deterioration results in changing responses affecting reproducibility of
ureido-N concentrations.

413 (iii) If accurate determinations of concentrations of urea are required, then HPLC-HRMS414 is the preferred approach.

415 (iv) It is important to properly validate the DAM method in all the environmental matrices
416 this method is currently applied to.

417 We will be discussing our findings with the marketers of the commercial assay to help them to 418 re-design their user guidelines to ensure reagent stability and sufficient method validation for 419 the DAM colorimetric method for the matrix under analysis.

420

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432 **References**

433

434 Alam, M.S., Casareto, B.E., Suzuki, Y., Sultana, R., Suzuki, T., 2017. Optimization of

dissolved urea measurements in coastal waters with the combination of a single reagent

436 and high temperature. Journal of Oceanography 73, 249–258. doi:10.1007/s10872-016-

437 0400-2

- 438 Archibald, R.M., 1945. Colorimetric Determination of Urea. The Journal of Biological
 439 Chemistry 157, 507–518.
- 440 Archibald, R.M., 1944. Determination of citrulline and allantoin and demonstration of
 441 citrulline in blood plasma. Journal of Biological Chemistry 156, 121–142.
- 442 Beale, R.N., Croft, D., 1961. A sensitive method for the colorimetric determination of urea.

443 Journal of Clinical Pathology 14, 418–24. doi:10.1136/JCP.14.4.418

- Berman, T., Bronk, D.A., 2003. Dissolved organic nitrogen: a dynamic participant in aquatic
 ecosystems. Aquatic Microbial Ecology 31, 279–305.
- 446 Brausch, J.M., Rand, G.M., 2011. A review of personal care products in the aquatic
- 447 environment: Environmental concentrations and toxicity. Chemosphere 82, 1518–1532.
- 448 doi:10.1016/j.chemosphere.2010.11.018
- 449 Bristow, A.W., Whitehead, D.C., Cockburn, J.E., 1992. Nitrogenous constituents in the urine
- 450 of cattle, sheep and goats. Journal of the Science of Food and Agriculture 59, 387–394.
- 451 doi:10.1002/jsfa.2740590316
- 452 Burford, M.A., Revill, A.T., Palmer, D.W., Clementson, L., Robson, B.J., Webster, I.T.,
- 453 2011. River regulation alters drivers of primary productivity along a tropical river-
- 454 estuary system. Marine and Freshwater Research 62, 141–151. doi:10.1071/MF10224
- 455 Butler, A.R., Walsh, D., 1982. Colorimetric non-enzymic methods for the determination of
- 456 urea. TrAC Trends in Analytical Chemistry 1, 120–124.
- 457 doi:http://dx.doi.org/10.1016/0165-9936(82)80011-3
- 458 Clark, S., Francis, P.S., Conlan, X.A., Barnett, N.W., 2007. Determination of urea using
- 459 high-performance liquid chromatography with fluorescence detection after automated
- 460 derivatisation with xanthydrol. Journal of Chromatography A 1161, 207–213.
- 461 doi:http://dx.doi.org/10.1016/j.chroma.2007.05.085
- 462 Cochlan, W.P., Bronk, D.A., 2001. Nitrogen uptake kinetics in the Ross Sea, Antarctica.

- 463 Deep Sea Research Part II: Topical Studies in Oceanography 48, 4127–4153.
- 464 doi:10.1016/S0967-0645(01)00083-2
- 465 Cornell, S.E., Jickells, T.D., Thornton, C.A., 1998. Urea in rainwater and atmospheric
- 466 aerosol. Atmospheric Environment 32, 1903–1910. doi:http://dx.doi.org/10.1016/S1352467 2310(97)00487-1
- 468 Daigh, A.L., Savin, M.C., Brye, K., Norman, R., Miller, D., 2014. Urea persistence in
- 469 floodwater and soil used for flooded rice production. Soil Use and Management 30,

470 463–470. doi:10.1111/sum.12142

- 471 Douglas, L.A., Bremner, J.M., 1970. Extraction and Colorimetric Determination of Urea in
- 472 Soils. Soil Science Society of America Journal 34, 859.
- 473 doi:10.2136/sssaj1970.03615995003400060015x
- 474 Fearon, W.R., 1939. The carbamido diacetyl reaction: a test for citrulline. Biochemistry
 475 Journal 33, 902–907.
- 476 Galloway, J.N., Cowling, E.B., 2002. Reactive Nitrogen and The World: 200 Years of
- 477 Change. AMBIO: A Journal of the Human Environment 31, 64–71. doi:10.1579/0044478 7447-31.2.64
- 479 Glibert, P.M., Harrison, J., Heil, C., Seitzinger, S., 2006. Escalating Worldwide use of Urea -
- 480 A Global Change Contributing to Coastal Eutrophication. Biogeochemistry 77, 441–
 481 463. doi:10.1007/s10533-005-3070-5
- 482 Goeyens, L., Kindermans, N., Abu Yusuf, M., Elskens, M., 1998. A Room Temperature
- 483 Procedure for the Manual Determination of Urea in Seawater. Estuarine, Coastal and
- 484 Shelf Science 47, 415–418. doi:10.1006/ecss.1998.0357
- 485 Greenan, N.S., Mulvaney, R.L., Sims, G.K., 1995. A microscale method for colorimetric
- 486 determination of urea in soil extracts. Communications in Soil Science and Plant
- 487 Analysis 26, 2519–2529. doi:10.1080/00103629509369465

| 488 | Harder Nielsen, T., Bonde, T.A., Sørensen, J., SÃ, rensen, J., 1998. Significance of microbial |
|-----|--|
| 489 | urea turnover in N cycling of three Danish agricultural soils. FEMS Microbiology |
| 490 | Ecology 25, 147–157. doi:10.1111/j.1574-6941.1998.tb00468.x |
| 491 | Heil, J., Vereecken, H., Brüggemann, N., 2016. A review of chemical reactions of |
| 492 | nitrification intermediates and their role in nitrogen cycling and nitrogen trace gas |
| 493 | formation in soil. European Journal of Soil Science 67, 23–39. doi:10.1111/ejss.12306 |
| 494 | Kind, T., Tolstikov, V., Fiehn, O., Weiss, R.H., 2007. A comprehensive urinary metabolomic |
| 495 | approach for identifying kidney cancer. Analytical Biochemistry 363, 185–195. |
| 496 | doi:10.1016/j.ab.2007.01.028 |
| 497 | Lambert, D.F., Sherwood, J.E., Francis, P.S., 2004. The determination of urea in soil extracts |
| 498 | and related samples - A review. Australian Journal of Soil Research 42, 709–717. |
| 499 | doi:10.1071/SR04028 |
| 500 | Lewis, K.A., Tzilivakis, J., Warner, D., Green, A., 2016. An international database for |
| 501 | pesticide risk assessments and managment. Human and Ecological Risk Assessment: An |
| 502 | International Journal 22, 1050–1064. |
| 503 | doi:http://dx.doi.org/10.1080/10807039.2015.1133242 |
| 504 | Lloyd, A.B., Sheaffe, M.J., 1973. Urease activity in soils. Plant and Soil 39, 71-80. |
| 505 | doi:10.1007/BF00018046 |
| 506 | Lugosi, R., Thibert, R.J., Holland, W.J., Lam, L.K., 1972. A study of the reaction of urea |
| 507 | with diacetyl monoxime and diacetyl. Clinical Biochemistry 5, 171–181. |
| 508 | doi:http://dx.doi.org/10.1016/S0009-9120(72)80028-6 |
| 509 | McCarthy, J.J., 1970. A urease method for urea in seawater. Limnology and Oceanography |
| | |

- 510 15, 309–313. doi:10.4319/lo.1970.15.2.0309
- 511 Mulvenna, P.F., Savidge, G., 1992. A modified manual method for the determination of urea
- 512 in seawater using diacetylmonoxime reagent. Estuarine, Coastal and Shelf Science 34,

| 513 4 | 429–438. (| doi:http://d | x.doi.org/10 | 0.1016/S0272 | 2-7714(05 | 6)80115-5 |
|-------|------------|--------------|--------------|--------------|-----------|-----------|
|-------|------------|--------------|--------------|--------------|-----------|-----------|

- 514 Murphy, J., Riley, J.P., 1962. A modified single solution method for the determination of
- 515 phosphate in natural waters. Analytica Chimica Acta 27, 31–36. doi:10.1016/S0003-
- 516 2670(00)88444-5
- 517 Nakas, J.P., Litchfield, C.D., 1977. Application of the diacetyl-monoxime thiosemicarbazide
- 518 method to the analysis of urea in estuarine sediments. Estuarine and Coastal Marine

519 Science 5, 143–150. doi:10.1016/0302-3524(77)90079-2

- 520 Neal, C., Neal, M., Wickham, H., 2000. Phosphate measurement in natural waters: two
- 521 examples of analytical problems associated with silica interference using
- 522 phosphomolybdic acid methodologies. Science of The Total Environment 251–252,
- 523 511–522. doi:10.1016/S0048-9697(00)00402-2
- 524 Prescott, A., Mangnall, D., 1976. Interference by sucrose in the chemical determination of
 525 citrulline with diacetyl monoxime. Analytical Biochemistry 76, 551–555.
- 526 doi:http://dx.doi.org/10.1016/0003-2697(76)90348-1
- 527 Price, N.M., Harrison, P.J., 1987. Comparison of methods for the analysis of dissolved urea
- 528 in seawater. Marine Biology 94, 307–317. doi:10.1007/bf00392945
- Remsen, C.C., 1971. The distribution of urea in coastal and oceanic waters. Limnology and
 Oceanography 16, 732–740. doi:10.4319/lo.1971.16.5.0732
- 531 Revilla, M., Alexander, J., Glibert, P.M., 2005. Urea analysis in coastal waters: comparison
- of enzymatic and direct methods. Limnology and Oceanography: Methods 3, 290–299.
- 533 doi:10.4319/lom.2005.3.290
- 534 Satoh, Y., Katoh, T., 1989. Evaluation of the degree of citrulline interference in the
- 535 spectrophotometry of urea with diacetyl monoxime in natural waters. Japanese Journal
- of Limnology (Rikusuigaku Zasshi) 50, 39–43. doi:10.3739/rikusui.50.39
- 537 Solomon, C., Collier, J., Berg, G., Glibert, P., 2010. Role of urea in microbial metabolism in

- 538 aquatic systems: a biochemical and molecular review. Aquatic Microbial Ecology 59,
- 539 67–88. doi:10.3354/ame01390
- 540 Stepanauskas, R., Laudon, H., Jørgensen, N.O.G., 2000. High DON bioavailability in boreal
- 541 streams during a spring flood. Limnology and Oceanography 45, 1298–1307.
- 542 doi:10.4319/lo.2000.45.6.1298
- 543 Sullivan, D.M., Havlin, J.L., 1991. Flow injection analysis of urea nitrogen in soil extracts.
- 544 Soil Science Society of America Journal 55, 109–113.
- 545 Taylor, A.J., Vadgama, P., 1992. Analytical Reviews in Clinical Biochemistry: The
- 546 Estimation of Urea. Annals of Clinical Biochemistry: An International Journal of
- 547 Biochemistry and Laboratory Medicine 29, 245–264.
- 548 doi:10.1177/000456329202900301
- 549 Torres-Valdés, S., Purdie, D.A., 2006. Nitrogen removal by phytoplankton uptake through a
- temperate non-turbid estuary. Estuarine, Coastal and Shelf Science 70, 473–486.
- 551 doi:10.1016/j.ecss.2006.06.028
- 552 Twomey, L.J., Piehler, M.F., Paerl, H.W., 2005. Phytoplankton uptake of ammonium, nitrate
- and urea in the Neuse River Estuary, NC, USA. Hydrobiologia 533, 123–134.
- 554 doi:10.1007/s10750-004-2403-z
- 555 Veniamin, M.P., Vakirtzi-Lemonias, C., 1970. Chemical Basis of the Carbamidodiacetyl
- 556 Micromethod for Estimation of Urea, Citrulline, and Carbamyl Derivatives. Clinical557 Chemistry 16.
- 558 Warren, C.R., 2014. Development of liquid chromatography mass spectrometry method for
- analysis of organic N monomers in soil. Soil Biology and Biochemistry 78, 233–242.
- 560 doi:10.1016/j.soilbio.2014.08.008
- 561

563 **Figure legends**

564

Fig. 1 Calibration curves for (a) allantoin, (b) alloxan, (c) hydantoin, and (d) biuret under the
DAM colorimetric reaction conditions at concentrations between 0.1 to 10 mg N l⁻¹.

567

- **Fig. 2** Extracted ion chromatograms for urea $[M+H]^+$ (*m*/*z* 61.08) for a soil extract from Merddwr, Conwy with (a) no added urea spike, and (b) 0.4 mg l⁻¹, (c) 2.0 mg l⁻¹ and (d) 4.0 mg l⁻¹ added urea spikes used for quantification by standard addition and (e) the standard addition calibration curve used to determine the urea concentration in the soil extract (0.709 µg g⁻¹).
- Fig. 3 (a) Ureido group responsible for the cross-reactivity with the DAM colorimetric method,
 and environmentally relevant compounds containing the ureido group which cross-react: (b)
 allantoin; (c) hydantoin and (d) biuret.