

1 Particle Emission Characteristics of a Gas Turbine 2 with a Double Annular Combustor

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24 ABSTRACT

25 The total climate, air quality and health impact of aircraft black carbon (BC) emissions
26 depends on quantity (mass and number concentration), as well as morphology (fractal
27 dimension and surface area) of emitted BC aggregates. This study examines multiple BC
28 emission metrics from a gas turbine with a double annular combustor, CFM56-5B4-2P. As a
29 part of the SAMPLE III.2 campaign, concurrent measurements of particle mobility, particle
30 mass, particle number concentration and mass concentration, as well as collection of
31 transmission electron microscopy (TEM) samples, allowed for characterization of the BC
32 emissions. Mass- and number-based emission indices were strongly influenced by thrust
33 setting during pilot combustion and ranged from <1 to 208 mg/kg-fuel and 3×10^{12} to 3×10^{16}
34 particles/kg-fuel, respectively. Mobility measurements indicated that mean diameters ranged
35 from 7-44 nm with a strong dependence on thrust during pilot-only combustion. Using
36 aggregation and sintering theory with empirical effective density relationships, a power law
37 relationship between primary particle diameter and mobility diameter is presented. Mean
38 primary particle diameter ranged from 6-19 nm, however, laser induced incandescence (LII)
39 and mass-mobility calculated primary particle diameters demonstrated opposite trends with
40 thrust setting. Similarly, mass-mobility-calculated aggregate mass specific surface area and
41 LII-measured surface area were not in agreement, indicating both methods need further
42 development and validation before use as quantitative indicators of primary particle diameter
43 and mass-specific surface area.

44 1. INTRODUCTION

45 Aircraft gas turbine engines emit particulate matter (PM) arising from incomplete combustion
46 of fuel, lubrication oil and the conversion of fuel sulfur compounds (Timko et al., 2010).
47 Non-volatile carbonaceous PM is referred to as soot of which, the fraction that is light-
48 absorbing is referred to as black carbon (BC) (Petzold et al., 2013). BC emitted by aircraft
49 engines has a positive direct radiative forcing (Lee et al., 2010; Stettler et al., 2013) and
50 emitted BC particles are a significant source of ice nuclei, which affect the formation of
51 contrails (Kärcher et al., 2009; Schumann et al., 2002; Schumann et al., 2013) and aviation
52 induced cloudiness (Lee et al., 2010). These indirect climate effects are potentially
53 significant, yet remain highly uncertain due to poor understanding of the effect of BC particle
54 composition and morphology on ice nucleation (Bond et al., 2013).

55 Aircraft emissions during landing and takeoff lead to elevated ambient concentrations of PM,
56 particularly in the vicinity of airports (Westerdhal et al., 2008; Zhu et al., 2011). As gas
57 turbine soot aggregates typically have a mobility diameter less than 100 nm (Kinsey et al.,
58 2010), health effects are potentially elevated as ultra-fine PM (<100 nm) could have greater
59 health effects than PM_{2.5} (<2.5 µm) (Cassee et al., 2013). PM surface area impacts the
60 reactivity of particles in the upper atmosphere and influences the uptake of sulfuric acid
61 (Zhang et al., 2008). In addition to the size characteristics of PM, the toxicity of PM may
62 depend upon the composition, surface chemistry and surface charge (Bakand et al., 2012).
63 Modelling studies have shown that morphology can affect the deposition of soot aggregates
64 in the human respiratory tract (Broday et al., 2011).

65 Emissions of soot from gas turbine engines emanate from the incomplete combustion of fuel
66 in the combustion chamber, the combustor. In a conventional combustor, soot is formed in
67 the region into which fuel is sprayed, initially by PAH inception and then surface growth
68 mechanisms (Hall et al., 1997; Wen et al., 2003). Downstream of this region, soot is
69 consumed by oxidation processes as fuel and air mixing and addition of dilution air increase
70 the air-to-fuel ratio. The difference between these two processes determines the concentration
71 of soot in the engine exhaust (Cumpsty, 2003; Lefebvre et al., 2010). The rate of soot
72 formation increases with combustion temperature, which is influenced both by the combustor
73 inlet temperature and local air-to-fuel ratios (Wen et al., 2003). Combustor inlet temperature
74 increases with increasing engine thrust setting and in conventional combustors, combustion
75 temperatures generally increase concomitantly, as evidenced by higher NO_x emissions at
76 higher engine thrust settings (EASA, 2012).

77 Existing measurements of modern gas turbine PM emissions have focused on PM mass and
78 number emissions indices (EI), emissions normalized by fuel burnt, and show that the mass
79 EI is greatest at higher engine thrust settings (Lobo et al., 2015; Lobo et al., 2008; Timko et
80 al., 2010; Wey et al., 2006). These existing measurements correspond to engines with
81 conventional annular combustors. Using high resolution transmission electron microscopy
82 (HRTEM), Vander Wal et al. (2014) reported that the nanostructure of the aggregate primary
83 particles is amorphous at low engine thrust settings and becomes more ‘graphitic’ at higher
84 engine thrust settings, suggestive of different soot growth mechanisms at different
85 combustion temperatures. Also using TEM, Liati et al. (2014) showed that the primary
86 particle size of soot aggregates was dependent on the engine thrust setting; the mode of the
87 primary particle size distribution increased from 13 to 24 nm from 7% to 100% of maximum
88 engine thrust setting. Durdina et al. (2014) showed that BC aggregate effective density is a
89 function of engine thrust setting for a given aggregate mobility diameter and that the mass-

90 mobility exponent ranged from 2.37 to 2.64 for 3-5% and 50-100% engine thrust settings
91 respectively.

92 In contrast to conventional combustors, double annular combustors (DACs) have two stages
93 of operation: a pilot stage in the outer annulus of the combustor, and a main stage in the inner
94 annulus. Only the outer (pilot) stage is fueled during light-off and at low power and is
95 characterized by low local air-to-fuel ratios and low through-flow velocity to achieve good
96 ignition and low CO and HC emissions. The main stage is characterized by high local air-to-
97 fuel ratios and high velocity to provide a lean flame and lower combustion temperatures
98 (Stickles et al., 2013). Compared the conventional combustor on the CFM56-5B4 engine, the
99 DAC combustor operating with the main stage reduces NO_x emissions by ~40% (EASA,
100 2012).

101 Soot aggregate morphology also affects the particle's scattering and radiative properties.
102 Radney et al. (2014) showed that while the mass specific absorption cross section is
103 independent of aggregate morphology, there is increased scattering for a more compacted
104 soot morphology and a concomitant increase in mass specific extinction cross section.
105 Furthermore, Yon et al. (2014) have shown that multiple scattering effects can influence
106 optical absorption measurements using laser induced incandescence (LII).

107 An analysis of the morphology of gas turbine soot and the dependence on engine operating
108 conditions is vital to improved understanding of the climate and health impacts of aircraft PM
109 and also to the correct interpretation of measurements using optical techniques. This paper,
110 therefore, aims to quantify the PM mass and number EI, as well as provide an analysis of the
111 morphology of solid particulate matter exhausted from a DAC gas turbine. Multiple in-situ
112 and ex-situ analysis techniques are compared to measure fundamentally distinct parameters
113 of the soot aerosol. Combinations of measurements taken as a part of the SAMPLE III.2
114 campaign are used to determine morphology metrics that are critical in understanding the
115 atmospheric and human health impacts of turbine particle emissions. The specific
116 morphology metrics measured and inferred within this study are aggregate mobility
117 distribution, mean particle specific surface area, and mean primary particle diameter as a
118 function of aggregate mobility diameter and engine thrust setting.

119 2. APPROACH

120 The SAMPLE III.2 campaign was conducted at the SR Technics turbine engine test facility in
121 Zurich, Switzerland from April 23rd to May 4th, 2012. The campaign consisted of “piggy-
122 back” tests of turbines being validated after maintenance procedures as well as “dedicated”
123 turbine engine testing that is the focus of this work. The dedicated test engine was a CFM
124 International CFM56-5B4-2P engine (120 kN thrust) with double annular staged combustion
125 fueled with Jet A-1 with an estimated sulfur concentration of 300 ppm to 800 ppm. Further
126 details of the testing approach and apparatus are described by Crayford et al. (2012). This
127 study focused on characterizing solid particulate matter and therefore all measurements and
128 sampling were taken downstream of a catalytic stripper (CS) or volatile particle remover
129 (VPR) (Giechaskiel et al., 2010; Giechaskiel et al., 2008; Khalek et al., 1995; Swanson et al.,
130 2010).

131 132 2.1 EXPERIMENTAL

133 The experimental apparatus used to collect, condition, and transport the aircraft turbine
134 exhaust to the aerosol characterization instruments on each sample line is shown in Figure 1.
135 Data included herein was from April 28th, 29th and 30th, 2012 test dates. Geometric and

136 operational details of the sampling and transport components are described by Crayford et al.
137 (2012). The characterization instruments and measurement techniques employed during the
138 campaign were as follows and have been previously been reviewed in this context by Petzold
139 et al. (2011).

140 *Aerosol thermal conditioning* (catalytic stripper and VPR). Semi-volatile material was
141 removed by using a catalytic stripper (CS) or volatile particle remover (VPR). The CS
142 contained two geometrically dissimilar catalyzed ceramic substrates: an oxidizing catalyst
143 and a sulfur trap both heated to 350°C. The purpose of the oxidation catalyst is to remove the
144 semi-volatile hydrocarbon particles and vapor. The sulfur trap removes sulfur species by
145 adsorption. The VPR approach is similar to the CS in intent but different in methodology. It
146 includes a 100:1 dilution of the exhaust with air heated to 150°C, a heated section with a wall
147 temperature in the range 350°C, a room temperature dilution section to reduce the particle
148 concentration to less than approximately 10,000 particles/cm³, and a particle number counter
149 (condensation particle counter) with 50% detection efficiency of 23 nm.

150
151 *Particle concentration measurement.* Particle concentration was determined by use of
152 condensational particle counters (CPCs) with 5, 10 and 23 nm cut points, where cutpoint is
153 defined as the particle diameter at which the particle detection efficiency is 50% (D₅₀). Both
154 TSI (Model 3775, 5 nm D₅₀; Model 3772, 10 nm D₅₀; and Model 3010, 23 nm D₅₀) and
155 Grimm (Model 5435, 10 nm D₅₀) CPCs were used in the study of two separate lines, the
156 FOCA and SAMPLE. Additional particle concentration information is given by mobility
157 measurement devices, DMS500 (Cambustion) and SMPS (TSI), but are used as a secondary
158 indicator of particle concentration for purposes of this study. All particle concentrations used
159 in this study were measured downstream of a volatile particle remover, VPR (AVL APC489-
160 CS).

161
162 *Electrical mobility sizing.* Particle mobility (“size”) distributions were measured using TSI
163 scanning mobility particle sizers (SMPS) (Wang et al., 1990) with 3085 nano-DMAs and
164 3081 long-DMAs both configured with 10:1 sheath/aerosol flowrates. The SMPS were
165 sometimes located in the secondary dilution vent line downstream of an AVL APC
166 (Giechaskiel et al., 2010; Giechaskiel et al., 2008). A DMS500 (Biskos et al., 2005; Reavell
167 et al., 2002) was used with its standard configuration.

168 *Mass mobility sizing* (Centrifugal particle mass analyzer, CPMA). The CPMA classifies
169 particles by their mass-to-charge ratio by balancing the electrostatic and centrifugal forces
170 between two concentric cylinders in motion relative to each other (Olfert et al., 2005). To
171 determine the real-time effective particle density, particles with a given mass-to-charge ratio
172 were transferred to the (modified, “m”) DMS500, which classified particles by their electrical
173 mobility (Biskos et al., 2005; Reavell et al., 2002) as described by Crayford et al. (2012).
174 Multiple charge correction was used in interpreting the combined CPMA and DMS results.

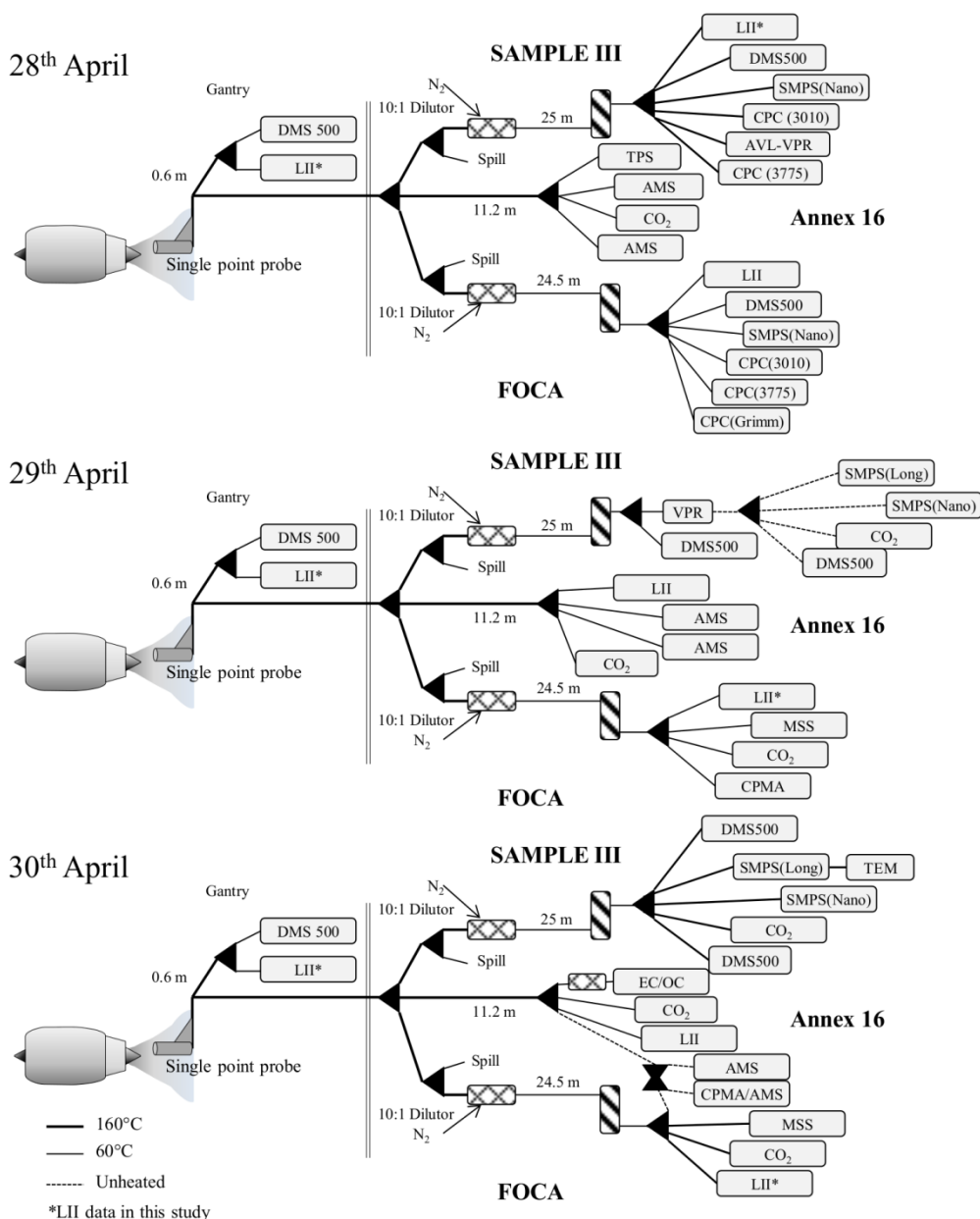
175 *Laser-induced incandescence.* Artium Technologies’ LII-300 measures the thermal emission
176 (incandescent light) from particles heated by a pulsed laser to temperatures in the 2500 K to
177 4500 K range (Snelling et al., 2005), making it appropriate for measuring the solid particles
178 produced by a combustion source. The selectivity is due to the fact that the solid particles are
179 primarily “black,” such that they absorb laser radiation and incandesce over a broad spectral
180 range. With careful control of the laser fluence, the instrument heats the particles to the point
181 of sublimation but not beyond, so that there is no significant mass loss while still achieving
182 the high temperatures necessary for the incandescence to be detected.

183 *Soot-particle aerosol mass spectrometer (SP-AMS)*. The SP-AMS is used for size and
184 composition analysis of solid and semi-volatile submicron aerosol (Jayne et al., 2000; Onasch
185 et al., 2012). Aerosols are sampled at ambient pressure by an aerodynamic lens that contracts
186 and expands the sampled air stream through a series of orifices. Solid particles are vaporized
187 by a continuous-beam laser operated at 1064 nm. The resulting vapor is then ionized by
188 electron impact at 70 eV; ions are then mass analyzed within a high-resolution time-of-flight
189 chamber. Comparisons of aggregate AMS vacuum aerodynamic diameter with mobility
190 diameter were used to determine the dynamic shape factor, χ , in accordance with DeCarlo et
191 al. (2004). The resulting shape factors, shown in SI Figure 9, were equal to or less than 1, and
192 were concluded as biased due to the insensitivity of the AMS to aggregates with a vacuum
193 aerodynamic diameter less than 50 nm. Further work is needed to improve the sensitivity of
194 AMS for aggregates with diameters less than 50 nm in order to make definitive
195 measurements of dynamic shape factors for gas turbine particles.

196 *Transmission electron microscopy*. Particles were collected onto 3 mm lacey carbon (Cu
197 Holey carbon film 400 mesh, Agar Scientific) TEM grids using thermophoresis (Just, 2012)
198 and electrophoresis techniques (Fierz et al., 2007). The flow of the thermophoretic sampler (1
199 L min⁻¹, ± 50 cm³ min⁻¹) resulted in impaction of large particles and thus an oversampling of
200 large aggregates. Therefore, the TEM-measured primary particle data was corrected using a
201 relation between aggregate and primary particle diameter in accordance with Dastanpour and
202 Rogak (2014). An empirical power correlation between volume-area primary particle
203 diameter, d_{va} , and aggregate mobility diameter, d_m , were fitted ($d_{va} = k_{TEM} d_m^{D_{TEM}}$, see
204 §2.2) to the entire TEM data base. Then mean mobility diameters from multiple mobility
205 measurements were used to determine the volume-area primary particle diameter for the 28th
206 April at each RPM test point. The resulting TEM-determined volume-area primary particle
207 diameters were compared with measurements from mass-mobility relations and LII.

208 *Line loss correction*. The particle mass and number correction factors for line losses were
209 determined in accordance with ASME E-31 committee's procedure of using the downstream
210 measured particle number mobility distribution with a mobility-dependent line loss curve to
211 determine the initial particle mobility distribution at the exit plane of the engine. The line loss
212 penetration was calculated using the United Technologies Research Centre (UTRC) model
213 which contains conventional aerosol theory diffusion, thermophoretic and inertial losses
214 (Liscinsky et al., 2010), and accounted for the 25 m line length and line temperature of
215 160°C. For purposes of this study both the upstream and downstream distributions were
216 assumed to fit lognormal distribution mass and number profiles. The downstream DMS500-
217 measured total number concentration (N_∞), geometric mean (d_{pg}) and geometric standard
218 deviation (σ_g) were used within an iterative routine to determine a lognormal upstream
219 distribution that when accounting for line losses results in a best fit ($R^2 > 0.9$) to the
220 downstream lognormal particle number and mass distributions characterized by the measured
221 N_∞ , d_{pg} and σ_g . DMS500 measurements on the SAMPLE III line were used in conjunction
222 with the measured particle effective density provided by Johnson et al. (In Press 2014) to
223 infer mass distributions from particle number distributions. The line loss correction approach
224 was used to determine the upstream to downstream particle number and mass ratio, as well as
225 upstream mean geometric mobility diameter and mean geometric standard deviation. The
226 particle number and mass line loss correction factors for the various thrust settings are shown
227 in the supporting information. All reported particle number and mass emission indices in the
228 results section have been corrected for line losses.

229



231
 232 Figure 1: Schematic of April 28th, 29th and 30th 2012 sampling system during SAMPLE III.2
 233 campaign.

234
 235 2.2 THEORETICAL

236 Aggregate aerosols are characterized by numerous metrics, many of which originate from the
 237 fundamental characteristics measured by analysis techniques. Spherical particles are most
 238 readily characterized by their geometric diameter (d_p), mass (m_p), volume (V) and the
 239 relation via density (ρ), $m_p = \rho \pi d_p^3 / 6$, $m_p = \rho V$. For non-spherical particles, similar
 240 parameters are used to define effective metrics of diameter, mass and volume. This study
 241 employs and tests the following analytical and semi-empirical constructs to facilitate

242 comparison between the different measurements and infer particle metrics beyond those that
 243 are measured fundamentally by each device.

244

245 *Mass and mobility metrics*

246 Agglomerates of soot with fractal-like structures are characterized by a variety of metrics,
 247 often related to the method of measurement. The agglomerate mobility diameter, d_m , as
 248 measured in a differential mobility analyzer, is related to aggregate mass by primary particle
 249 diameter, d_{pp} , mass, m , and the mass-mobility exponent, D_{fm} by the relation

$$250 \quad m = k_{fm} \left(\frac{d_m}{d_{pp}} \right)^{D_{fm}} \quad \text{Eq. 1}$$

251 where k_{fm} is the mass-mobility prefactor with units of appropriate mass (Park et al., 2004).
 252 Soot agglomerates can also be defined by their effective density, ρ_{eff} ; the ratio of the
 253 agglomerate mass to equivalent volume based on the mobility diameter,

$$254 \quad \rho_{eff} = \frac{m}{\pi d_m^3/6} = k d_m^{D_{fm}-3} \quad \text{Eq. 2}$$

255 where $k = \frac{6}{\pi} \frac{k_{fm}}{d_{pp}^{D_{fm}}}$ (McMurry et al., 2002). The power law relationship described by Eq. 2
 256 has been shown to fit well with experimental data using a constant prefactor k , despite the
 257 potential for varying primary particle size, for a variety of engines, including the engine
 258 studied here (Johnson et al., In Press 2014). The number of primary particles within a soot
 259 aggregate is related to the overall aggregate mobility and the primary particle diameter, d_{pp} ,
 260 by the power law relation

$$261 \quad n_{pp} = k_a \left(\frac{d_m}{d_{pp}} \right)^{2D_\alpha}, \quad \text{Eq. 3}$$

262 where k_a and D_α are the pre-exponential and power law exponent, respectively. Eq. 1 can be
 263 related to Eq. 3 where $m = n_{pp} \rho \pi d_{pp}^3/6$ and thus, $k_a = 6k_{fm}/(\rho \pi d_{pp}^3)$. For non-ideal
 264 aggregates, i.e. partially sintered, it is appropriate to define the primary particle diameter as a
 265 volume area equivalent primary particle diameter, $d_{va} \equiv \frac{6v}{a}$ and, thus $n_{va} = \frac{v}{\pi d_{va}^3/6}$, where v
 266 and a are the aggregate volume and surface area respectively. By taking the primary particle
 267 diameter and number of primary particles as their volume area equivalent, $d_{pp} = d_{va}$ and
 268 $n_{pp} = n_{va}$, Eq. 3 can be solved for volume-area primary particle diameter as a function of
 269 measured d_m and m ,

$$270 \quad d_{va} = \left(\frac{k_a \pi \rho}{6m} (d_m)^{2D_\alpha} \right)^{\frac{1}{2D_\alpha-3}}. \quad \text{Eq. 4}$$

271 Eggersdorfer et al. (2012) have shown that Eq. 4 is valid with constant values of $k_a=0.998$
 272 and $D_\alpha=1.069$ for a polydisperse mix of primary particles regardless of the sintering
 273 mechanism or state of sintering. For particle sources where an empirical relationship has been
 274 determined between the particle mass and the particle mobility, such as described by the
 275 effective density relationship in Eq. 2, the mass term in Eq. 4 can be replaced with a function
 276 of mobility. By eliminating the mass term with an empirical effective density formulation, a

277 power law relationship between the volume average primary particle diameter and the
278 mobility diameter can be derived

$$279 \quad d_{va} = k_{va} d_m^{D_{va}}, \quad \text{Eq. 5}$$

280 where $k_{va} = (\rho k_a/k)^{\frac{1}{2D_{va}-3}}$ and $D_{va} = \frac{2D_{\alpha}-D_{fm}}{2D_{\alpha}-3}$. By including empirical relations for particle
281 mass within analytical fractal scaling laws, the physical significance of the pre-exponential
282 constants is lost. As above, Eq. 5 assumes a constant value of k that is independent of
283 primary particle diameter, the validity of which is tested in the results section below. This
284 relationship can be used to relate the surface area primary particle diameter with the mean
285 mobility diameter for each mobility distribution.

286 A relation for the particle mass-specific surface area can be derived from the definition of
287 volume area equivalent primary particle diameter,

$$288 \quad \frac{a}{m} = \frac{6}{\rho d_{va}}. \quad \text{Eq. 6}$$

289 When particle measurements of both mass and mobility are available, Eq. 4 may be used to
290 determine d_{va} , whereas Eq. 5 may be used if an empirical relationship is known for the
291 aggregate effective density.

292 Dobbins et al. (1994) report a value of 1.86 g/cm³ for diesel and quote six other works, in the
293 range of 1.82 to 2.05 g/cm³ that have a mean of 1.92 g/cm³. An elemental soot density of $\rho =$
294 1.9 g/cm³ will be used for this study.

295 *LII primary particle size*

296 As the LII 300 instrument does not allow significant particle sublimation, the dominant
297 cooling mechanism for the particles is conduction to the surrounding gas, associated with the
298 surface area of the particles. Assuming monodisperse primary particles allows a direct
299 relationship between the surface area and the primary particle diameter. During conduction
300 cooling, the temperature difference between the particles, T_p , and the ambient gas, T_g , decays
301 steadily with a near-single exponential behavior. An equation of the form

302
$$T_p - T_g = A \cdot e^{-\tau} \quad \text{Eq. 7}$$

303 is fit to the temperature data (measured by the instrument with two-colour pyrometry) to
304 determine τ , the time constant of the exponential decay, and where A is a constant (Snelling
305 et al., 2002). This method requires a priori knowledge of the ambient gas temperature, which
306 is determined by thermocouple in the sample cell. The primary particle diameter, d_{pp} , is
307 determined directly from the decay of the LII signal, using the relation derived from McCoy
308 et al. (1974),

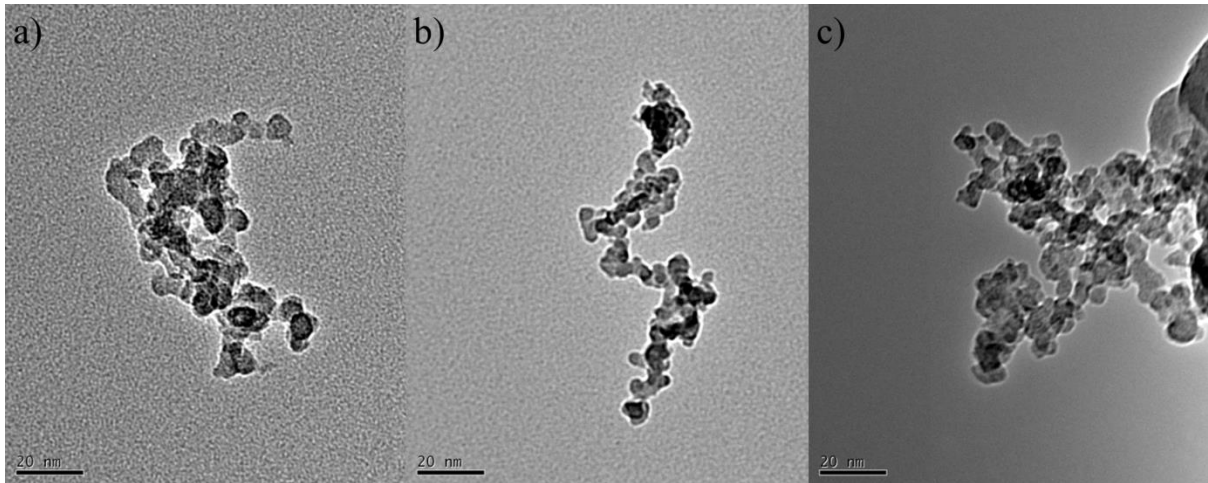
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$$d_{pp} = \frac{12 k_g \alpha \tau}{G \lambda_g c_p \rho_p} \quad \text{Eq. 8}$$

310 where k_g is the thermal conductivity of the ambient gas, α is the thermal accommodation
311 coefficient, G is a geometry-dependent heat transfer coefficient, λ_g is the mean free path in
312 the ambient gas, and c_p and ρ_p are the specific heat and material density of the particle,
313 respectively.

314 The assumption of monodisperse primary particles maximizes the surface area to volume
315 ratio; in reality, there is a distribution of primary particle diameters, and these primary
316 particles are formed into aggregates, for which there is a distribution of aggregate sizes. Both
317 of these effects have an impact in terms of interpreting the temperature decay rates, such that
318 the reported primary particle diameter is an effective heat transfer primary particle diameter
319 for an equivalent population of monodisperse primary particles (Liu et al., 2006).

320
321 **3 RESULTS**

322
323 Engine emissions from the CFM56-5B4-2P were sampled on multiple days and different
324 sample lines with varying sample-to-nitrogen dilution ratios, as depicted in Figure 1.
325 Mobility-selected samples were collected downstream of a long DMA and imaged in an
326 HRTEM. Figure 2 shows representative images of a compact, (a), and linear, (b), 15-nm
327 mobility diameter aggregate, as well as a 50-nm, (c), mobility diameter aggregate. In all cases
328 the particles are seen to be composed of many (> 30) primary particles. In several cases, such
329 as Figure 2c, the presence of higher contrast particles was observed on the surface of the
330 lighter contrast soot. EDX analysis of these samples showed the presence of metals, such as
331 vanadium, silicon and titanium, indicating a likelihood of ash within the particles.
332 Quantitative EDX analysis across many particles was not conducted due to the lack of
333 statistically significant quantities of mobility-selected particles.
334



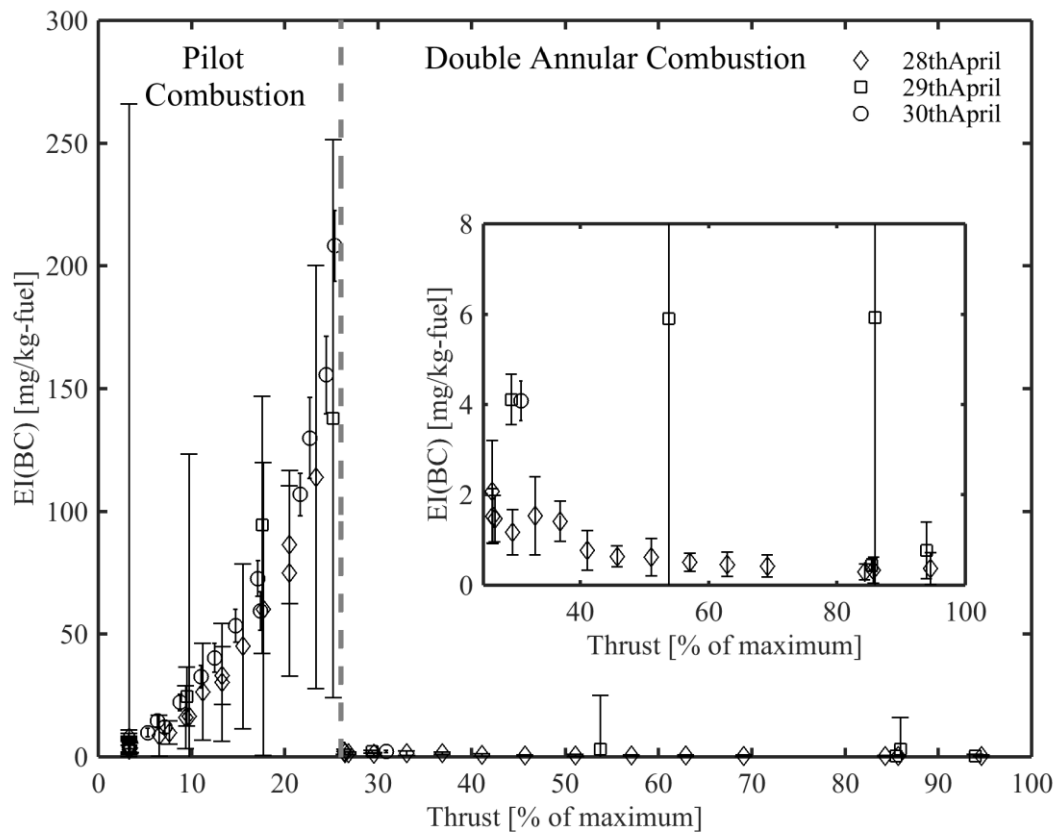
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336 Figure 2: TEM images of 15 nm, (a) and (b), and 50 nm, (c), mobility-selected soot
 337 aggregates.

338 The LII measurements of mass concentration were conducted in parallel with CO₂
 339 measurements on respective sample lines. The mass concentrations were transformed into
 340 emission indices in accordance with SAE methodology (SAE Aerospace, 2009). Analysis of
 341 the emissions indices indicated that the variability (range of measured values) between
 342 instruments and lines was less than the measured variability within a given instrument at a
 343 specific test point (see SI Figure 1). Therefore, only data from the LII noted with an asterisk
 344 in Figure 1 is presented here for clarity and consistency.

345

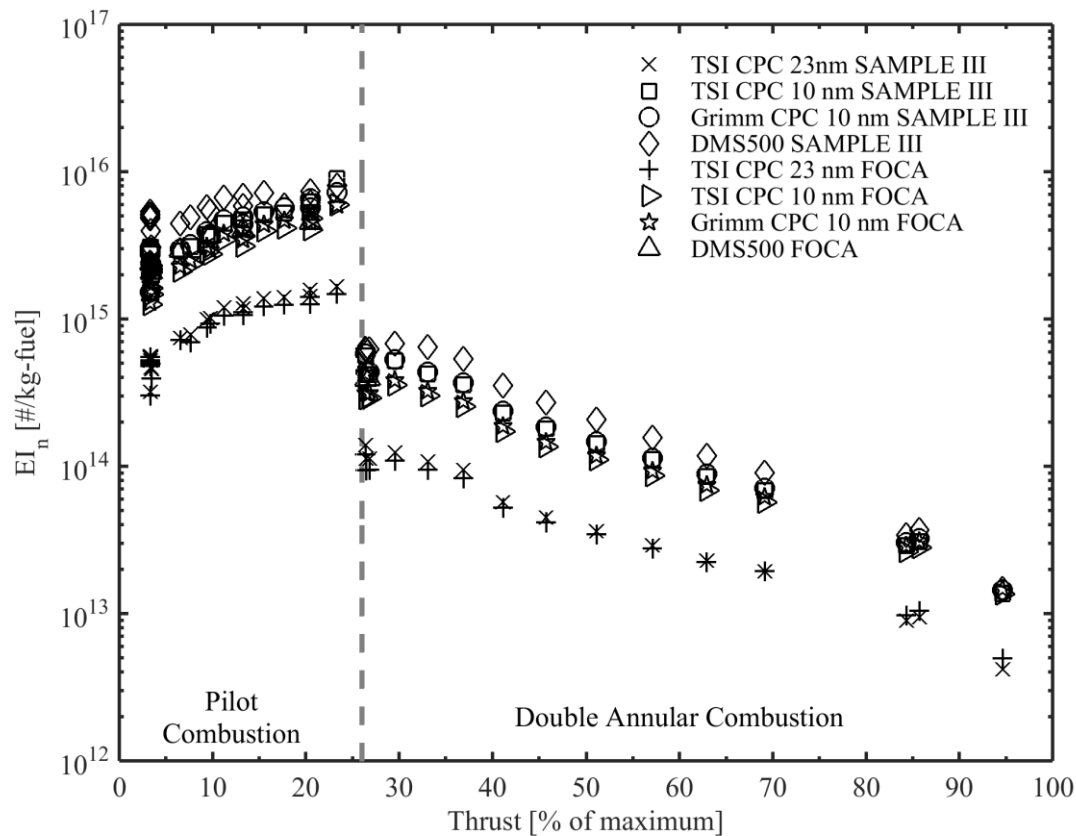
346 As shown in Figure 3, the mass emission index for the engine increases with increasing
 347 engine thrust setting while under pilot combustion. During low thrust settings only the single
 348 pilot combustor is fueled. The global stoichiometry is lean (measured global air to fuel ratio
 349 is shown in SI Figure 3), but the local stoichiometry within the pilot combustor is rich and
 350 only mixes with excess air downstream of the pilot combustion zone. The rich combustion
 351 ensures stability of the flame, but results in mass emission indices that are greater than at
 352 higher thrust conditions. At higher thrust settings where double annular combustion occurs,
 353 the mass emissions indices are less than 7 mg/kg-fuel (see Figure 3 inset) with a majority of
 354 emissions less than 2 mg/kg-fuel. The trend of mass emission index with engine thrust of the
 355 CFM56-5B4-2P is atypical when compared to conventional single-mode combustors, which
 356 tend to increase with increasing thrust setting (greater than 30%) and have EI(BC) in the
 357 range of 33 to 611 mg/kg-fuel for thrust settings greater than 45% of maximum thrust
 358 (Stettler et al., 2013). Thus, for thrust settings less than 15%, the measured mass emission
 359 indices of the CFM56-5B4-2P are typical of other engines (1-108 mg/kg-fuel). At
 360 intermediate thrust settings (15-25%) the CFM56-5B4-2P mass emission indices are greater
 361 than other engines (9-47 mg/kg-fuel), and at high thrust setting (>25%) are considerably
 362 lower than other measured engines.



363

364 Figure 3: Black carbon mass emission index, EI(BC), for CFM56-5B4-2P as measured by LII
 365 on three separate lines for various thrust settings. Error bars represent the 90% variability
 366 interval within a given thrust setting.

367 Particle number concentrations were measured on the 28th April 2012 from both the
 368 SAMPLE III and FOCA lines. In Figure 4, particle number based emissions indices (EI_n)
 369 calculated according to SAE methodology (SAE Aerospace, 2009) are shown as a function
 370 of engine thrust setting for the CFM56-5B4-2P engine. Measurements by the CPCs with D₅₀
 371 = 10 nm are in good agreement with the measurements from the DMS500s on both the
 372 SAMPLE III and FOCA sample lines. However, CPCs with D₅₀ = 23 nm measure
 373 significantly lower particle number concentration, counting between 18-38% of the total
 374 particles counted by the D50 = 10 nm CPC from the same manufacturer (TSI), which
 375 indicates that a majority of particles are less than 23 nm. The variability in engine emissions
 376 at each test point is greater than the variability across different sample lines for the DMS500
 377 and D₅₀ = 10 nm CPCs. As with the mass-based emissions index, EI_n increases with engine
 378 thrust setting during pilot combustion up to a maximum of 3×10¹⁶ particles/kg-fuel. After the
 379 engine transitions to use double annular combustion, EI_n reduces by an order of magnitude
 380 and decreases with increasing thrust to a minimum of 3×10¹² particles/kg-fuel at the highest
 381 recorded thrust.

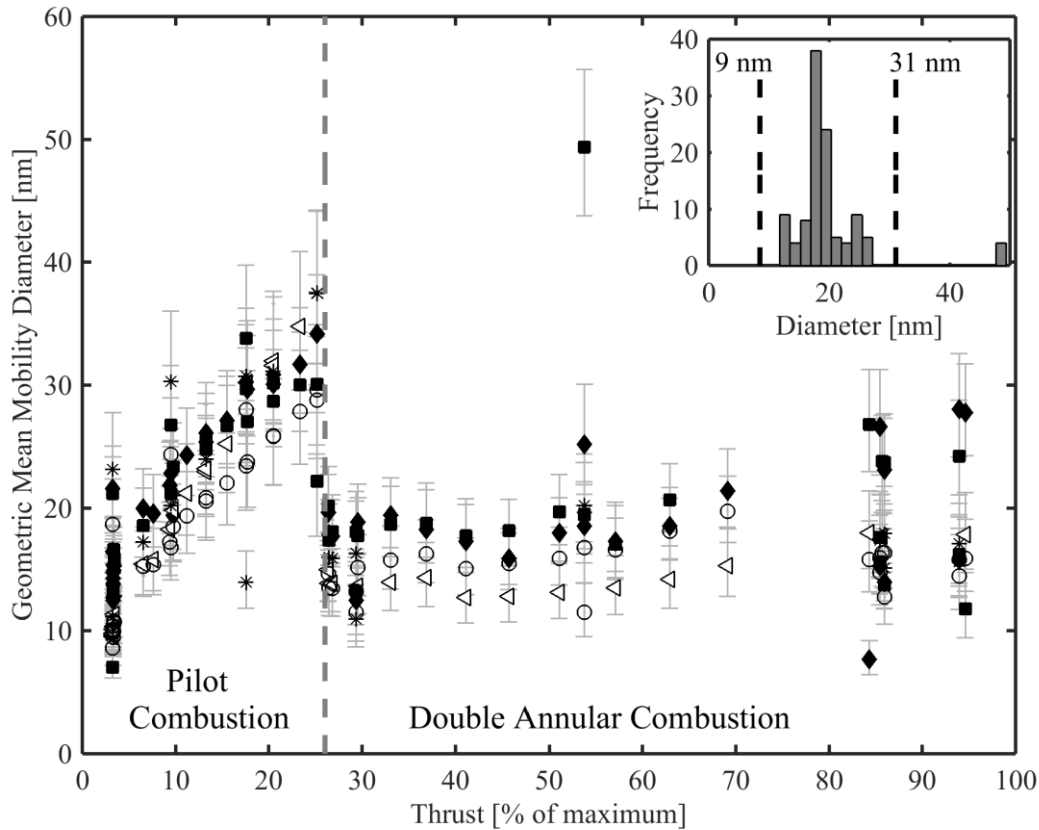


382

383 Figure 4: Black carbon number emission index, EI_n , for CFM56-5B4-2P as measured by
 384 CPCs with D_{50} cut points of 10 nm (open symbols) and 23 nm (cross and plus sign) on two
 385 separate lines (FOCA and SAMPLE III) for various thrust settings. The 90% variability was
 386 not significant are were omitted for legibility of symbols, but are shown SI Figure 2.

387 At each thrust setting mobility measurements were taken on different sample lines and by
 388 different instruments as shown in Figure 1. The resulting measurements (over 10^5 separate
 389 mobility scans) were compiled and averaged over the entire test condition for each mobility
 390 instrument. The geometric mean mobility diameters were determined for each thrust setting,
 391 as shown in Figure 5. The aggregate geometric mean mobility diameter from the CFM56-
 392 5B4-2P generally varied from 7 to 44 nm (two outliers excluded). The mobility diameter
 393 increased with increasing thrust within the single pilot combustion stage from 12 nm (90%
 394 variability interval, VI, 8-16 nm) at 4% maximum thrust to 33 nm (90% VI, 24-43 nm),
 395 coinciding with the higher mass concentrations shown in Figure 3. Aggregate diameters for
 396 particles produced during double annular combustion had less variation in size throughout the
 397 entire range of thrusts with a mean particle diameter of 17 nm (90% VI, 8-26 nm, Figure 5
 398 inset). The measurement of aggregate mobility diameter at the gantry without dilution (open
 399 circles) typically resulted in smaller measured aggregate diameters at thrusts with high
 400 emission indices (10-25% maximum thrust) when compared to the other thrust settings. The
 401 largest mobility measurements were typically recorded by the SMPS systems (closed square
 402 and triangle) which varied in their line placement. While not shown, the DMS500
 403 measurements at times measured mobility distributions that appeared bimodal, whereas the
 404 SMPS measurements almost exclusively measured a single mode. As with the emission index
 405 measurements, the variability within a given thrust setting as measured by a given instrument
 406 was greater than the variability between instruments and lines at most settings. For higher

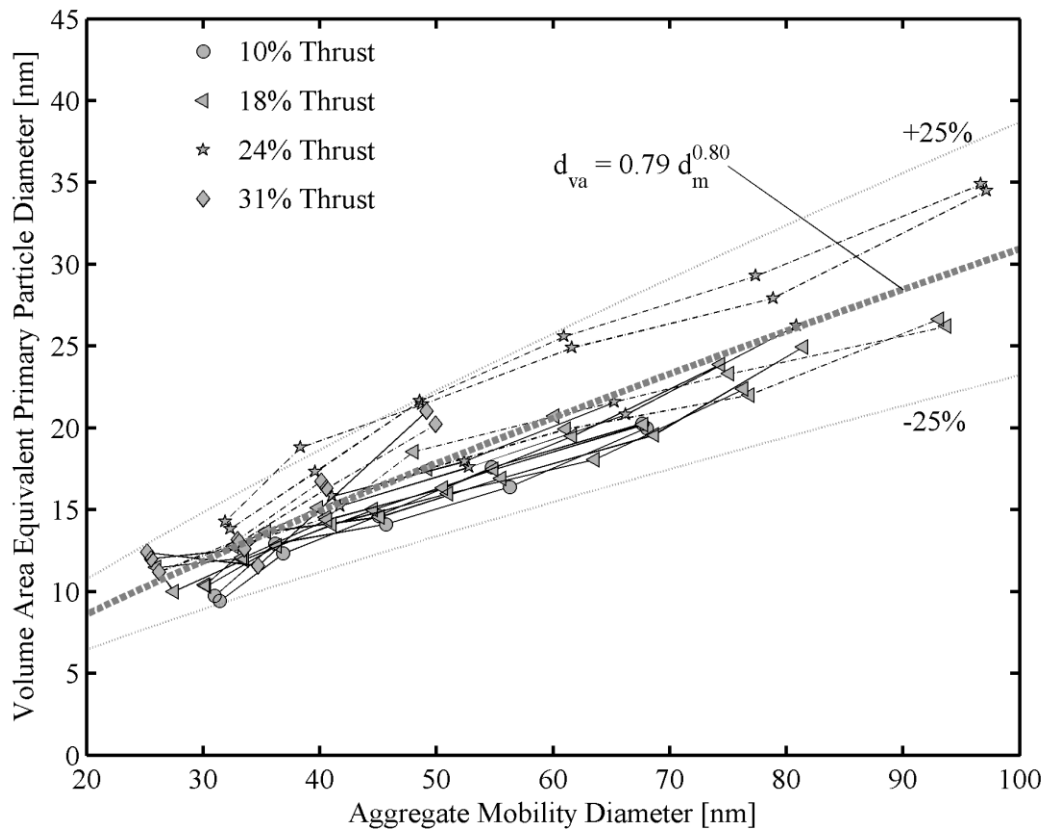
407 thrust settings, where the variability was greatest among instruments, the low concentrations
 408 of measured particles resulted in higher variability in the measurements.
 409



410
 411 Figure 5: Geometric mean diameter as measured by multiple mobility instruments over a
 412 range of engine thrusts. The measurements were taken from 28-30th April, 2012 on lines as
 413 depicted in Figure 1 from different mobility instruments DMS500 (FOCA – asterisk,
 414 SAMPLE – triangle, Gantry - Circle), nanoSMPS (SAMPLE – diamond), longSMPS
 415 (SAMPLE/FOCA – square). The inset shows the frequency distribution of the mean
 416 aggregate diameters for the DAC thrust settings. Error bars indicate the mean geometric
 417 standard deviation of the mobility distribution across all measurements within a given test
 418 condition.

419 As shown in §2.2, $d_{va} = \left(\frac{k_a \pi \rho}{6m} (d_m)^{2D_\alpha} \right)^{\frac{1}{2D_\alpha - 3}}$. Eq. 4 relates the volume area equivalent
 420 primary particle diameter within an aggregate to the measured aggregate mass and mobility.
 421 Mean aggregate mobilities were measured for a range of different selected masses within a
 422 subset of engine thrusts as reported by Johnson et al. (In Press 2014). The resulting data set
 423 (reproduced in SI Figure 4) allowed for determination of the primary particle size by
 424 analytical methods. The resulting volume area equivalent primary particle diameters are
 425 shown in Figure 6 as a function of aggregate mobility diameter for the denuded samples,
 426 where primary particle size is determined according to $d_{va} = \left(\frac{k_a \pi \rho}{6m} (d_m)^{2D_\alpha} \right)^{\frac{1}{2D_\alpha - 3}}$. Eq. 4. As
 427 shown, volume area equivalent primary particle diameter increases with aggregate mobility,
 428 whereby a power-law relationship of $d_{va} = 0.79 d_m^{0.8} \pm 25\%$ encapsulates all but one of the

429 measured data points. Fits to each individual thrust setting are shown in SI Figure 6. The
 430 value of the power law exponent, $D_{va} = 0.8$, can be compared to the result calculated using
 431 the effective density results reported by Johnson et al. (In Press 2014), $D_{fm} = 2.76$ which
 432 when used with a constant $D_{\alpha} = 1.069$, results in a power law exponent as defined in Eq. 5 of
 433 $D_{va} = 0.72$. The discrepancy in the two D_{va} values is a result of the difference in least squares
 434 regression (see SI Figure 7). The trend observed here is consistent with the correlation of
 435 primary particle size with aggregate size obtained from TEM analysis of different combustion
 436 sources; however the value of the power law exponent measured by this method is
 437 considerably larger than those reported by Dastanpour and Rogak (2014).



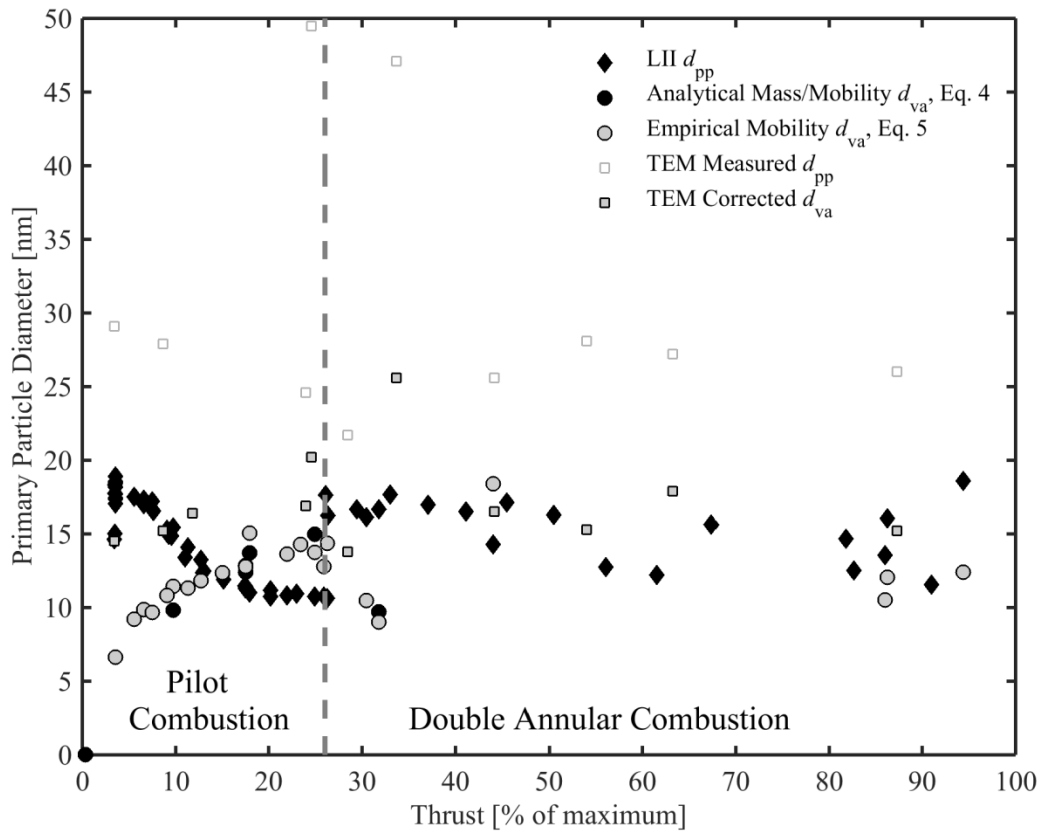
438

439 Figure 6: Volume area equivalent primary particle diameter as a function of aggregate
 440 mobility diameter as measured by mass and mobility analysis. The grey lines correspond to
 441 the empirical fit with the power law form of Eq. 5 ($R^2=0.86$) and a $\pm 25\%$ interval.

442 The volume area equivalent primary particle diameter, as measured by TEM and mass-
 443 mobility techniques, are plotted in Figure 7 as they relate to thrust setting, along with the LII
 444 effective heat transfer primary particle diameter. Primary particle diameters as measured by
 445 LII and mass-mobility vary from 6 to 19 nm, while TEM-measured primary particle
 446 diameters were considerably larger (18 to 47 nm). When corrected for oversampling of larger
 447 aggregates, the TEM-measured volume area equivalent primary particle diameters were in
 448 closer agreement (15 to 26 nm) with the range determined by mass-mobility relations.
 449 Primary particle diameters within the pilot combustion stage demonstrate a noticeable change
 450 with thrust setting, whereas the primary particle diameters produced during double annular
 451 combustion show no noticeable trend with thrust setting. The LII-measured primary particle
 452 diameter decreases from 19 to 10 nm with increasing thrust setting from 0 to 26% full thrust,

453 whereas the mass-mobility and corrected TEM primary particle diameters increased over the
454 same thrust range. As shown in Figure 5, the aggregate diameter increases with increasing
455 thrust setting within the pilot combustion stage, indicating that the average primary particle
456 diameter also likely increases over that range. As the aggregate mobility diameters increase,
457 the effective density decreases while mass increases (see SI Figures 4 and 7), which affects
458 the radiative and convective heat removal from the aggregate surface after heating within the
459 LII beam. The influence of effective density is not accounted for within the current LII
460 primary particle calculation, but it is known that primary particle measurement from the LII
461 signal decay is in better agreement within larger, less dense aggregates (Schulz et al., 2006).
462 The impact of effective density is hypothesized to dominate measurements of primary
463 particle size for compact aggregates and may account for the discrepancy in LII
464 measurements. Further work is needed to accurately account for effective density effects on
465 LII-determined primary particle diameter. Estimates of error within these measurements and
466 derived quantities are provided within the supporting information, where it is shown that the
467 TEM measured diameter is ± 2 nm and derived d_{va} has an uncertainty of $\pm 26.6\%$ based on
468 the current theoretical formulation. Current error estimates for the measured LII d_{pp} are not
469 available, and is an active area of research.

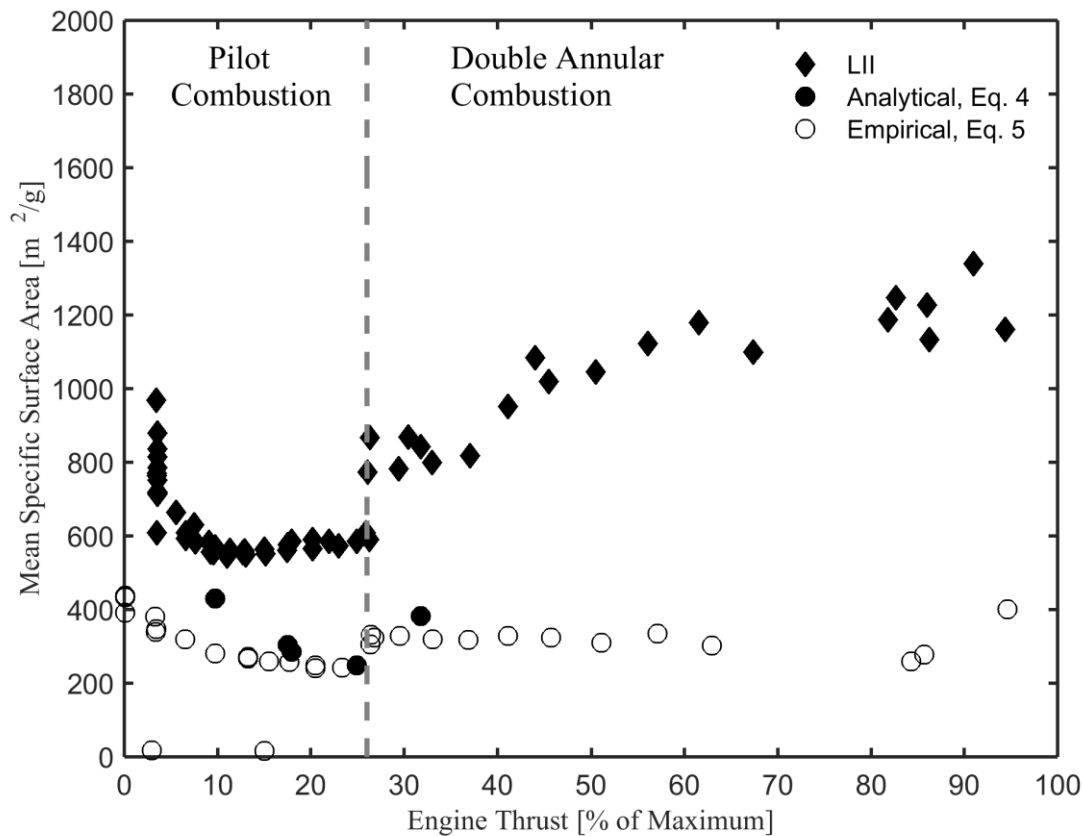
470 The observed increase in corrected TEM-measured volume area equivalent primary particle
471 diameter from 15 to 21 nm over the pilot combustion stage and primary particle diameters
472 typically below 18 nm during double annular combustion corroborate trends previously
473 observed between primary particle diameter and combustion temperature for conventional
474 combustors (Liati et al., 2014; Vander Wal et al., 1999). Lean combustion and lower
475 temperatures during double annular combustion are the likely cause of the observed reduction
476 in BC mass and number emissions and smaller primary particle diameters compared to
477 conventional combustors.



478

479 Figure 7: Primary particle diameter as determined by TEM which reports number mean d_{pp}
 480 and d_{va} ; mass and mobility measurements which report number mean d_{va} and LII which
 481 represents an effective heat transfer primary particle diameter.

482 LII gives a measure of mass specific surface area which is active for heat transfer. The heat
 483 transfer specific surface area can be compared to the total surface area calculated from mass
 484 and mobility relations as given by Eq. 5 and Eq. 6. The aggregate specific surface area is
 485 plotted as a function of engine thrust in Figure 8, where the LII-measured specific surface
 486 area is seen to vary from 552 to 1339 m^2/g . The mass-mobility determined specific surface
 487 area varied over a smaller range from 240 to 347 m^2/g . Both LII measurements and mass-
 488 mobility determined specific surface areas remained relatively constant in pilot combustion
 489 mode while above the 5% full thrust setting. During double annular combustion the LII
 490 specific surface areas showed an increase with thrust setting that was more pronounced than
 491 the mass-mobility determined surface area. Previously reported soot specific surface area are
 492 in better agreement with the mass-mobility determined specific surface area, where Ishiguro
 493 et al. (1997) measured specific surface areas of 52 to 296 m^2/g where higher surface areas
 494 corresponded to more oxidized samples. Popovitcheva et al. (2000) report an aggregate
 495 specific surface area for aircraft soot of 47-100 m^2/g as measured by N_2 thermodesorption
 496 spectroscopy. Given the high value of LII-measured specific surface area when compared
 497 with previous measurements, it is likely that the results are influenced by other factors, such
 498 as effective density, as discussed previously.



499

500 Figure 8: Aggregate specific surface area as determined by LII, and analytical (Eq. 4) and
 501 empirical (Eq. 5) mass and mobility relationships as defined in Eq. 6.

502 4 SUMMARY AND CONCLUSION

503 The BC emissions from a gas turbine with a double annular combustor, CFM56-5B4-2P,
 504 were measured as a part of the SAMPLE III.2 campaign. TEM images indicated that the soot
 505 consisted of aggregates composed of many (>30) primary particles for aggregate mobility
 506 diameters as low as 15 nm. Mass-based emission indices demonstrate a unique trend from
 507 single stage-combustion engines, whereby the EI(BC) increases rapidly with thrust setting
 508 during pilot only combustion reaching a maximum of 80-208 mg/kg-fuel at 20-25% of
 509 maximum thrust. At higher thrusts settings where double annular combustion occurs, the mass
 510 emissions indices are significantly less than single-stage combustors, with measured EI(BC)
 511 less than 8 mg/kg-fuel. Particle number emissions, EI_n, increase with engine thrust setting
 512 during pilot combustion up to a maximum of ~10¹⁶ particles/kg-fuel. During double annular
 513 combustion EI_n reduces by an order of magnitude and decreases with increasing thrust setting
 514 to a minimum of ~10¹³ particles/kg-fuel at the highest recorded thrust setting.

515 The aggregate geometric mean mobility diameter corresponds to other modern gas engines
 516 with diameters ranging from 7 to 44 nm. As with the emissions indices, there was a positive
 517 correlation for mobility diameter with increasing thrust within the single pilot combustion
 518 stage resulting in diameters ranging from 12 nm to 33 nm. Thrust setting had less impact on
 519 the aggregate mobility diameters produced during double annular combustion where mean

520 particle diameters were 17 nm (90% VI, 8-26 nm). Concurrent aggregate mass and mobility
521 measurements also allowed for calculation of aggregate volume average primary particle
522 diameters, which were seen to increase with mobility diameter according to the empirical
523 power-law relationship $d_{va} = 0.79d_m^{0.8}$. Assuming this relationship holds for this engine at all
524 thrust settings, the primary particle diameters as determined by LII, TEM and mass-mobility
525 relations were compared. The primary particle results show conflicting trends, particularly
526 between the LII and mass-mobility determined primary particle diameters. It is hypothesized
527 that the effective density may play a role in the effective heat transfer surface area from
528 aggregates, which will serve to bias LII results of larger aggregates. Further work is needed
529 for accurate measurement primary particle diameters. Measures of aggregate mass specific
530 surface area were compared between LII and mass-mobility calculated values. While neither
531 method is a recognized standard for determining surface area, the mass-mobility relations
532 were closer to measures in other studies. Further work is needed to refine and validate LII-
533 determined surface area and primary particle diameter.

534

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544 5 BIBLIOGRAPHY

545 Bakand, S., Hayes, A., Dechsakulthorn, F. (2012). Nanoparticles: a review of particle
546 toxicology following inhalation exposure. *Inhalation Toxicology*. 24:125-135.

547

548 Biskos, G., Reavell, K., Collings, N. (2005). Unipolar diffusion charging of aerosol particles
549 in the transition regime. *Journal of Aerosol Science*. 36:247-265.

550

551 Bond, T. C., Doherty, S. J., Fahey, D. W., Forster, P. M., Berntsen, T., DeAngelo, B. J.,
552 Flanner, M. G., Ghan, S., Kärcher, B., Koch, D., Kinne, S., Kondo, Y., Quinn, P. K., Sarofim,
553 M. C., Schultz, M. G., Schulz, M., Venkataraman, C., Zhang, H., Zhang, S., Bellouin, N.,
554 Guttikunda, S. K., Hopke, P. K., Jacobson, M. Z., Kaiser, J. W., Klimont, Z., Lohmann, U.,
555 Schwarz, J. P., Shindell, D., Storelvmo, T., Warren, S. G., Zender, C. S. (2013). Bounding
556 the role of black carbon in the climate system: A scientific assessment. *Journal of*
557 *Geophysical Research: Atmospheres*. n/a-n/a.

558

559 Broday, D. M., Rosenzweig, R. (2011). Deposition of fractal-like soot aggregates in the
560 human respiratory tract. *Journal of Aerosol Science*. 42:372-386.

561
562 Cassee, F. R., Héroux, M.-E., Gerlofs-Nijland, M. E., Kelly, F. J. (2013). Particulate matter
563 beyond mass: recent health evidence on the role of fractions, chemical constituents and
564 sources of emission. *Inhalation Toxicology*. 25:802-812.

565
566 Crayford, A., Johnson, M., Marsh, R., Sevcenco, Y., Walters, D., Williams, P., Petzold, A.,
567 Bowen, P., Wang, J., Lister, D., (2012). SAMPLE III: Contribution to Aircraft Engine PM
568 Certification Requirement and Standard Second Specific Contract– Final Report,
569 easa.europa.eu.

570
571 Cumpsty, N. A., (2003). *Jet Propulsion: A simple guide to the aerodynamic and*
572 *thermodynamic design and performance of jet engines*. Cambridge University Press.

573
574 Dastanpour, R., Rogak, S. N. (2014). Observations of a Correlation between Primary Particle
575 and Aggregate Size for Soot Particles. *Aerosol Science and Technology*.00-00.

576
577 DeCarlo, P. F., Slowik, J. G., Worsnop, D. R., Davidovits, P., Jimenez, J. L. (2004). Particle
578 Morphology and Density Characterization by Combined Mobility and Aerodynamic
579 Diameter Measurements. Part 1: Theory. *Aerosol Science and Technology*. 38:1185-1205.

580
581 Dobbins, R. A., Mulholland, G. W., Bryner, N. P. (1994). Comparison of a Fractal Smoke
582 Optics Model with Light Extinction Measurements. *Atmospheric Environment*. 28:888-897.

583
584 Durdina, L., Brem, B. T., Abegglen, M., Lobo, P., Rindlisbacher, T., Thomson, K.,
585 Smallwood, G. J., Hagen, D. E., Sierau, B., Wang, J. (2014). Determination of PM mass
586 emissions from an aircraft turbine engine using particle effective density. *Atmospheric*
587 *Environment*. 99:500-507.

588
589 EASA, (2012). ICAO Engine Emissions Databank [online], Available from:
590 <http://easa.europa.eu/environment/edb/aircraft-engine-emissions.php>.

591
592 Eggersdorfer, M. L., Kadau, D., Herrmann, H. J., Pratsinis, S. E. (2012). Aggregate
593 morphology evolution by sintering: Number and diameter of primary particles. *Journal of*
594 *Aerosol Science*. 46:7-19.

595
596 Fierz, M., Kaegi, R., Burtscher, H. (2007). Theoretical and experimental evaluation of a
597 portable electrostatic TEM sampler. *Aerosol Science and Technology*. 41:520-528.

598
599 Giechaskiel, B., Chirico, R., DeCarlo, P., Clairotte, M., Adam, T., Martini, G., Heringa, M.,
600 Richter, R., Prevot, A., Baltensperger, U. (2010). Evaluation of the particle measurement
601 programme (PMP) protocol to remove the vehicles' exhaust aerosol volatile phase. *Science of*
602 *The Total Environment*. 408:5106-5116.

603
604 Giechaskiel, B., Dilara, P., Sandbach, E., Andersson, J. (2008). Particle measurement
605 programme (PMP) light-duty inter-laboratory exercise: comparison of different particle
606 number measurement systems. *Measurement Science and Technology*. 19:095401.

607
608 Hall, R., Smooke, M., Colket, M. (1997). Predictions of soot dynamics in opposed jet
609 diffusion flames. *Physical and Chemical Aspects of Combustion: A Tribute to Irvin*
610 *Glassman*.189-230.

611
612 Ishiguro, T., Takatori, Y., Akihama, K. (1997). Microstructure of diesel soot particles probed
613 by electron microscopy: First observation of inner core and outer shell. *Combustion and*
614 *Flame*. 108:231-234.

615
616 Jayne, J. T., Leard, D. C., Zhang, X., Davidovits, P., Smith, K. A., Kolb, C. E., Worsnop, D.
617 R. (2000). Development of an aerosol mass spectrometer for size and composition analysis of
618 submicron particles. *Aerosol Science & Technology*. 33:49-70.

619
620 Johnson, T. J., Olfert, J. S., Symonds, J. P. R., Johnson, M., Rindlisbacher, T., Swanson, J. J.,
621 Boies, A. M., Thomson, K., Smallwood, G., Walters, D., Sevcenco, Y., Crayford, A.,
622 Durdina, L., Bahk, Y. K., Brem, B., Wang, J., Dastanpour, R., Rogak, S. N. (In Press 2014).
623 Effective density and mass-mobility exponent of aircraft particulate matter. *Journal of*
624 *Propulsion & Power*.

625
626 Just, B. G., (2012). Characterization of ultrafine particulate matter from traditional and
627 improved biomass cookstoves, *Mechanical Engineering*. University of British Columbia,
628 <https://circle.ubc.ca/handle/2429/43053>.

629
630 Kärcher, B., Yu, F. (2009). Role of aircraft soot emissions in contrail formation. *Geophys.*
631 *Res. Lett.* 36:L01804.

632
633 Khalek, I. A., Kittelson, D. B. (1995). Real Time Measurement of Volatile and Solid Exhaust
634 Particles Using a Catalytic Stripper. SAE Technical Paper. 950236.

635
636 Kinsey, J. S., Dong, Y., Williams, D. C., Logan, R. (2010). Physical characterization of the
637 fine particle emissions from commercial aircraft engines during the Aircraft Particle
638 Emissions eXperiment (APEX) 1–3. *Atmospheric Environment*. 44:2147-2156.

639
640 Lee, D. S., Pitari, G., Grewe, V., Gierens, K., Penner, J. E., Petzold, A., Prather, M. J.,
641 Schumann, U., Bais, A., Bernsten, T. (2010). Transport impacts on atmosphere and climate:
642 *Aviation*. *Atmospheric Environment*. 44:4678-4734.

643
644 Lefebvre, A., Ballal, D., (2010). *Gas turbine combustion, alternative fuels and emission*, 3rd
645 edition. Boca Raton: Taylor & Francis.

646
647 Liati, A., Brem, B. T., Durdina, L., Vögtli, M., Arroyo Rojas Dasilva, Y., Dimopoulos
648 Eggenschwiler, P., Wang, J. (2014). Electron microscopic study of soot particulate matter
649 emissions from aircraft turbine engines. *Environmental Science & Technology*. 48:10975-
650 10983.

651
652 Liscinsky, D. S., Bhargava, A., Colket, M. B., Hautman, D. C., Hollick, H. H., True, B.
653 (2010). Effect of Particle Sampling Technique and Transport on Particle Penetration at the
654 High Temperature and Pressure Conditions Found in Gas Turbine Combustors and Engines.
655 NASA/CR-2010eNNC07CB03C.

656
657 Liu, F., Yang, M., Hill, F. A., Snelling, D. R., Smallwood, G. J. (2006). Influence of
658 polydisperse distributions of both primary particle and aggregate size on soot temperature in
659 low-fluence LII. *Applied Physics B*. 83:383-395.

660
661 Lobo, P., Hagen, D. E., Whitefield, P. D., Raper, D. (2015). PM emissions measurements of
662 in-service commercial aircraft engines during the Delta-Atlanta Hartsfield Study.
663 *Atmospheric Environment*. 104:237-245.

664
665 Lobo, P., Whitefield, P. D., Hagen, D. E., Trueblood, M. B., Mundis, N. L., Magdits, I. P.,
666 Herndon, S. C., Onasch, T. B., Jayne, J. T., Miake-lye, R. C., Eberhard, W. L., Wayson, R.
667 (2008). Delta - Atlanta Hartsfield (UNA-UNA) Study.

668
669 McCoy, B., Cha, C. (1974). Transport phenomena in the rarefied gas transition regime.
670 *Chemical Engineering Science*. 29:381-388.

671
672 McMurry, P. H., Wang, X., Park, K., Ehara, K. (2002). The Relationship between Mass and
673 Mobility for Atmospheric Particles: A New Technique for Measuring Particle Density.
674 *Aerosol Science and Technology*. 36:227-238.

675
676 Olfert, J., Collings, N. (2005). New method for particle mass classification—the Couette
677 centrifugal particle mass analyzer. *Journal of Aerosol Science*. 36:1338-1352.

678
679 Onasch, T., Trimborn, A., Fortner, E., Jayne, J., Kok, G., Williams, L., Davidovits, P.,
680 Worsnop, D. (2012). Soot particle aerosol mass spectrometer: development, validation, and
681 initial application. *Aerosol Science and Technology*. 46:804-817.

682
683 Park, K., Kittelson, D. B., Zachariah, M. R., McMurry, P. H. (2004). Measurement of
684 inherent material density of nanoparticle agglomerates. *Journal of Nanoparticle Research*.
685 6:267-272.

686
687 Petzold, A., Marsh, R., Johnson, M., Miller, M., Sevcenco, Y., Delhaye, D., Ibrahim, A.,
688 Williams, P., Bauer, H., Crayford, A., Bachalo, W. D., Raper, D. (2011). Evaluation of

689 Methods for Measuring Particulate Matter Emissions from Gas Turbines. *Environmental*
690 *Science & Technology*. 45:3562-3568.

691
692 Petzold, A., Ogren, J. A., Fiebig, M., Laj, P., Li, S.-M., Baltensperger, U., Holzer-Popp, T.,
693 Kinne, S., Pappalardo, G., Sugimoto, N., Wehrli, C., Wiedensohler, A., Zhang, X.-Y. (2013).
694 Recommendations for reporting “black carbon” measurements. *Atmos. Chem. Phys.*
695 13:8365–8379.

696
697 Popovitcheva, O., Persiantseva, N., Trukhin, M., Rulev, G., Shonija, N., Buriko, Y. Y.,
698 Starik, A., Demirdjian, B., Ferry, D., Suzanne, J. (2000). Experimental characterization of
699 aircraft combustor soot: Microstructure, surface area, porosity and water adsorption. *Physical*
700 *Chemistry Chemical Physics*. 2:4421-4426.

701
702 Radney, J. G., You, R., Ma, X., Conny, J. M., Zachariah, M. R., Hodges, J. T., Zangmeister,
703 C. D. (2014). Dependence of soot optical properties on particle morphology: measurements
704 and model comparisons. *Environmental Science & Technology*. 48:3169-3176.

705
706 Reavell, K., Hands, T., Collings, N., (2002). A fast response particulate spectrometer for
707 combustion aerosols. SAE Technical Paper.

708
709 SAE Aerospace, (2009). Aircraft Exhaust Nonvolatile Particle Matter Measurement Method
710 Development, in: Report, A.I. (Ed.), <http://www.sae.org>.

711
712 Schulz, C., Kock, B. F., Hofmann, M., Michelsen, H., Will, S., Bougie, B., Suntz, R.,
713 Smallwood, G. (2006). Laser-induced incandescence: recent trends and current questions.
714 *Applied Physics B*. 83:333-354.

715
716 Schumann, U., Arnold, F., Busen, R., Curtius, J., Karcher, B., Kiendler, A., Petzold, A.,
717 Schlager, H., Schröder, F., Wohlfrom, H. (2002). Influence of fuel sulfur on the composition
718 of aircraft exhaust plumes: The experiments SULFUR 1–7. *Geophys. Res. Lett.* 40:2867-
719 2872.

720
721 Schumann, U., Jeßberger, P., Voigt, C. (2013). Contrail ice particles in aircraft wakes and
722 their climatic importance. *Geophys. Res. Lett.* 40:2867–2872.

723
724 Snelling, D. R., Smallwood, G. J., Liu, F., Gülder, Ö. L., Bachalo, W. D. (2005). A
725 calibration-independent laser-induced incandescence technique for soot measurement by
726 detecting absolute light intensity. *Applied optics*. 44:6773-6785.

727
728 Snelling, D. R., Smallwood, G. J., Sawchuk, R. A., Neill, W. S., Gareau, D., Clavel, D. J.,
729 Chippior, W. L., Liu, F., Gülder, Ö. L., Bachalo, W. D., (2002). In-situ real-time
730 characterization of particulate emissions from a diesel engine exhaust by laser-induced

731 incandescence. SAE Technical Paper, Diesel Particulate Emissions: Landmark Research
732 1994-2001.

733

734 Stettler, M. E. J., Boies, A. M., Petzold, A., Barrett, S. R. H. (2013). Global Civil Aviation
735 Black Carbon Emissions. *Environmental Science & Technology*. 47:10397-10404.

736

737 Stickles, R., Barrett, J., (2013). TAPS II Combustor Final Report. CLEEN Program. General
738 Electric, Washington DC.

739

740 Swanson, J., Kittelson, D. (2010). Evaluation of thermal denuder and catalytic stripper
741 methods for solid particle measurements. *Journal of Aerosol Science*. 41:1113-1122.

742

743 Timko, M. T., Onasch, T. B., Northway, M. J., Jayne, J. T., Canagaratna, M. R., Herndon, S.
744 C., Wood, E. C., Miake-Lye, R. C., Knighton, W. B. (2010). Gas Turbine Engine
745 Emissions—Part II: Chemical Properties of Particulate Matter. *J. Eng. Gas Turbines Power*.
746 132:061505–061505–061515.

747

748 Vander Wal, R. L., Bryg, V. M., Huang, C.-H. (2014). Aircraft engine particulate matter:
749 Macro- micro- and nanostructure by HRTEM and chemistry by XPS. *Combustion and Flame*.
750 162:602-611.

751

752 Vander Wal, R. L., Tichich, T. M., West, J. R. (1999). Laser-induced Incandescence Applied
753 to Metal Nanostructures. *Appl. Opt.* 38:5867-5879.

754

755 Wang, S. C., Flagan, R. C. (1990). Scanning electrical mobility spectrometer. *Aerosol*
756 *Science and Technology*. 13:230-240.

757

758 Wen, Z., Yun, S., Thomson, M., Lightstone, M. (2003). Modeling soot formation in turbulent
759 kerosene/air jet diffusion flames. *Combustion and Flame*. 135:323-340.

760

761 Westerthal, D., Fruin, S., Fine, P., Sioutas, C. (2008). The Los Angeles International Airport
762 as a source of ultrafine particles and other pollutants to nearby communities. *Atmospheric*
763 *Environment*. 42:3143-3155.

764

765 Wey, C., Anderson, B., Hudgins, C., Wey, C., Li-Jones, X., Winstead, E., Thornhill, L. K.,
766 Lobo, P., Hagen, D., Whitefield, P. (2006). Aircraft particle emissions experiment (APEX).
767 ARL-TR-3903. NASA.

768

769 Yon, J., Liu, F., Bescond, A., Caumont-Prim, C., Rozé, C., Ouf, F.-X., Coppalle, A. (2014).
770 Effects of multiple scattering on radiative properties of soot fractal aggregates. *J. Quant.*
771 *Spectrosc. Radiat. Transf.* 133:374-381.

772
773 Zhang, R., Khalizov, A. F., Pagels, J., Zhang, D., Xue, H., McMurry, P. H. (2008).
774 Variability in morphology, hygroscopicity, and optical properties of soot aerosols during
775 atmospheric processing. *Proceedings of the National Academy of Sciences*. 105:10291-
776 10296.

777
778 Zhu, Y., Fanning, E., Yu, R. C., Zhang, Q., Froines, J. R. (2011). Aircraft emissions and local
779 air quality impacts from takeoff activities at a large International Airport. *Atmospheric*
780 *Environment*. 45:6526-6533.

781

782

783