1 Title:

# 2 The spatial organisation and microbial community structure of an

## 3 epilithic biofilm

4

5 Nick A. Cutler<sup>a\*</sup>, Dominique L. Chaput<sup>b</sup>, Anna E. Oliver<sup>c</sup>, Heather A. Viles<sup>d</sup>

6

- 7 a Geography Department, University of Cambridge, Downing Place, Cambridge, CB2 3EN,
- 8 UK
- 9 b Department of Mineral Sciences, Smithsonian Institution, National Museum of Natural
- 10 History, 10<sup>th</sup> & Constitution NW, Washington, DC 20560-119, USA
- 11 ° Centre for Ecology and Hydrology, Maclean Building, Benson Lane, Crowmarsh Gifford,
- 12 Wallingford, OX10 8BB, UK
- d School of Geography and the Environment, Oxford University Centre for the Environment,
- 14 South Parks Road, Oxford, OX1 3QY, UK

15

16 \* corresponding author

17

- 18 Address: Churchill College, Cambridge, CB3 0DS, UK
- 19 E-mail: nac37@cam.ac.uk
- 20 Telephone: +44 1223 336202
- 21 Fax: +44 1223 336180

#### **Abstract**

232425

26

27

28

29

30

31

32

33

34

35

36

37

38

39

40

Microbial biofilms are common on lithic surfaces, including stone buildings. However, the ecology of these communities is poorly understood. Few studies have focussed on the spatial characteristics of lithobiontic biofilms, despite the fact that spatial structure has been demonstrated to influence ecosystem function (and hence biodegradation) and community diversity. Furthermore, relatively few studies have utilised molecular techniques to characterise these communities, even though molecular methods have revealed unexpected microbial diversity in other habitats. This study investigated 1) the spatial structure and 2) the taxonomic composition of an epilithic biofilm using molecular techniques, namely amplicon pyrosequencing and terminal restriction fragment length polymorphism (TRFLP). Dispersion indices and Mantel correlograms were used to test for the presence of spatial structure in the biofilm. Diversity metrics and rank-abundance distributions (RADs) were also generated. The study revealed spatial structure on a centimetre-scale in eukaryotic microbes (fungi and algae), but not the bacteria. Fungal and bacterial communities were highly diverse; algal communities much less so. The RADs were characterised by a distinctive 'hollow' (concave up) profile and long tails of rare taxa. These findings have implications for understanding the ecology of epilithic biofilms and the spatial heterogeneity of stone biodeterioration.

41 42

Keywords: amplicon pyrosequencing; TRFLP; lithobiontic microbes; green algae; bacteria; fungi; mantel correlograms; rank-abundance distributions

#### 1 Introduction

 Microbial biofilms are a common feature on both natural rock outcrops and stone buildings. Particular attention has been focused on the biofilms that form on stone buildings, as these communities can have a detrimental impact on both the appearance and physical structure of the stone (Warscheid & Braams, 2000, McNamara & Mitchell, 2005, Scheerer, *et al.*, 2009). However, whilst laboratory work has revealed some of the mechanisms by which biofilmforming microbes contribute to the biodeterioration of stone, very little is known about the ecology of these organisms. In particular, the spatial structure of stone-dwelling (lithobiontic) microbial communities at sub-metre scales (i.e. the size of masonry building components) is poorly understood, as most previous studies have focussed on the micro- or field/landscape scale (Franklin & Mills, 2010). Furthermore, relatively few studies have used molecular (DNA) techniques to characterise the microbial communities of building stone. To address this knowledge gap, we conducted a high-resolution, molecular study of an epilithic biofilm in order to investigate both small- (centimetre) scale spatial structure and microbial community composition.

Studies of epilithic biofilms have largely neglected the spatial dimension of lithobiontic communities even though a) many biofilms on buildings are patchy on a scale visible to the naked eye and b) distribution of organisms in space (spatial structure) has been demonstrated to have a profound impact on ecosystem function (Franklin, 2005, Nunan, *et al.*, 2010) and hence the patterns and processes of biodeterioration. Microbial communities can obviously vary on very small spatial scales. Gleeson et al. (2005), for example, found that lithobiontic fungal communities can vary on the scale of individual mineral grains.

Studies have investigated spatial variation on the scale of individual buildings (Rindi & Guiry, 2003, Rindi, 2007, Cutler, *et al.*, 2013) or on larger (regional) scales (Gaylarde & Gaylarde, 2005). However, there is almost no published work that looks at the intermediate scale of centimetres - metres, i.e. at a level that is of interest to building conservators (Franklin & Mills, 2010). Understanding the factors that structure microbial communities at this scale may shed light on the fundamental ecological processes that affect subaerial, lithobiontic biofilms and their influence on the deterioration of stone.

Previous studies have described the composition of lithobiontic microbial communities (see, e.g., Gleeson, et al., 2005, Gorbushina, 2007, Gleeson, et al., 2010). Most of this work has focussed on individual taxonomic groups (e.g. Cyanobacteria or fungi) and has been based on techniques such as microscopy and *in vitro* culturing of environmental samples. However, these techniques can only capture a tiny proportion of environmental microbial diversity. For

example, it is thought that only 1-10% of environmental microbes can be successfully cultured (Pace, 1997). Modern molecular techniques have been used in a variety of habitats including soil (Acosta-Martinez, et al., 2008, Jones, et al., 2009, Tedersoo, et al., 2010) and marine habitats (Pommier, et al., 2010, Gaidos, et al., 2011), but, with a few exceptions, they have not been systematically applied to building stone (Cutler & Viles, 2010). Typically, molecular techniques reveal levels of microbial diversity much higher than those based on cultured environmental samples.

Our study was broad in that it encompassed the three main groups of microorganism (hereafter, microorganism types) living on building stone i.e. algae, fungi and bacteria. We applied two, widely used, molecular techniques (amplicon pyrosequencing and terminal restriction fragment length polymorphism (TRFLP)) to describe community structure, concentrating on the two-dimensional arrangement of microbes in a thin biofilm (i.e. predominantly epilithic microbes). In terms of spatial structure, we anticipated that samples close to each other would be more similar than widely-spaced samples in terms of community composition, due to centimetre scale patchiness in the biofilm. We also anticipated low taxonomic diversity, but high taxonomic richness, particularly in fungal and bacterial communities.

### 2 Materials and Methods

We used molecular methods to describe the spatial structure and composition of a biofilm on a sandstone slab. Broadly, sample-specific TRFLP was used to ascertain the spatial structure of each microbial group and amplicon pyrosequencing was used to establish microbial community composition.

## 2.1 Sampling

In order to assess potential variation in microbial communities at a centimetre-scale, we sampled a smooth, fine-grained sandstone slab at a high level of spatial resolution. The slab, which had been kept in an exposed location in a stonemason's yard in Oxford, UK, had obvious biological growth on its surfaces but was homogeneous in terms of surface texture and composition. Whilst sampling was carried out during the development of a non-destructive sampling technique for epilithic biofilms described in Cutler et al. (2012) the following research is based on new analyses of these samples. We focussed on the epilithic microbial community because surface-dwelling microorganisms are easiest to sample and are likely to have the greatest impact on the appearance of the stone. The sampling regime is described in detail in Cutler et al. (2012). Briefly, the surface of the slab was divided into a

119 6 x 4 grid of sample locations (24 samples in total), each measuring 30 x 30 mm and 120 separated from neighbouring sample locations by a buffer zone of 25 mm. A section of sterile 121 adhesive tape was applied to the surface of the stone to collect microbial cells from the 122 biofilm. The biological material on the tape was then analyzed using molecular techniques. 123 124 2.2 Amplicon pyrosequencing and TRFLP 125 A modified CTAB extraction method was used to extract DNA from the tape samples: refer to 126 Cutler et al. (2012) for a detailed protocol. Briefly, subsamples (20 µl) were taken from the 127 DNA extracted from each of the 24 tape samples. These subsamples were pooled and 128 cleaned with a Powerclean kit (MoBio Laboratories Inc, Carlsbad, CA) in accordance with the 129 manufacturer's instructions. The pooled sample was then standardised to a DNA 130 concentration of 20 ng/µl and analysed via tag-encoded FLX amplicon pyrosequencing, 131 utilising a Roche 454 FLX instrument (454 Life Sciences, Branford, CT). The primer set 132 euk516F (5'-GGAGGCAAGTCTGGT-3'), euk1055R (5'-ARCGGCCATGCACCACC-3') was 133 used for eukaryotes (primarily to characterise algae); the primers ITS1F (5'-134 CTTGGTCATTTAGAGGAAGTAA-3') and ITS4 (5'-TCCTCCGCTTATTGATATGC-3') were 135 used for fungi and the primers 104F (5'-GGACGGGTGAGTAACACGTG-3'), 530R (5'-136 GTATTACCGCGGCTGCTG-3') for bacteria. The pyrosequencing was performed at the 137 Research and Testing Laboratory (RTL, Lubbock, TX) based upon RTL protocols. 138 139 Mothur 1.32.1 (Schloss, et al., 2009) was used to process raw sequence data generated by 140 the amplicon pyrosequencing, following the pipelines described in Cutler et al. (2014). Briefly, 141 bacterial 16S rRNA, eukaryotic 18S rRNA and fungal ITS flow files were trimmed and 142 denoised with the mothur implementation of PyroNoise (Quince, et al., 2009). Bacterial and 143 eukaryotic rRNA seguences were aligned using the corresponding SILVA reference 144 alignments (Quast, et al., 2013) and only sequences spanning the targeted regions were 145 kept. Data were denoised by clustering together sequences with 1 bp mismatch per 100 bp, 146 and chimeras were removed using the mothur implementation of uchime (Edgar, et al., 147 2011). Bacterial and eukaryotic rRNA sequences were classified against the SILVA 148 reference databases using the Wang method (Wang, et al., 2007), with a cutoff value of 60% 149 for taxonomic assignment. Sequences were also clustered into operational taxonomic units 150 (OTUs) at the 97% similarity level, which corresponds approximately to the species level. We 151 screened the eukaryotic data and removed OTUs identified as Fungi, Embrophyta and 152 Metazoa, as well as all unclassified Viridiplantae and Charophyta (i.e. only OTUs that could 153 be confidently, if broadly, classified as algae remained).

155 For fungal ITS sequences, following denoising, the ITS1 region was extracted using the ITS 156 Extractor tool on the PlutoF Workbench (Abarenkov, et al., 2010, Nilsson, et al., 2010) and 157 sequences shorter than 100 bp were discarded. Chimeras were removed using the mothur 158 implementations of uchime. OTU clustering was carried out from a distance matrix 159 constructed in mothur using pairwise distance values. Sequences were classified against the 160 UNITE+INSDC fungal ITS database (Abarenkov, et al., 2010), modified as previously 161 described (Cutler, et al., 2014), with a cutoff value of 50% for taxonomic assignment. 162 163 All sequence data were uploaded with MIMARKS-compliant metadata to the NCBI Sequence 164 Read Archive under Bioproject number PRJNA260418. 165 166 Alongside amplicon pyrosequencing, we also conducted a fresh analysis of TRFLP data 167 associated with the same samples: refer to Cutler et al. (2012) for a full description of the 168 original TRFLP protocols. In principle, amplicon pyrosequencing, or a similar, direct 169 sequencing technique, could be used to analyse spatial structure in microbial communities. 170 However, the large number of samples required for a fine-grained, statistically significant 171 study meant that these techniques were considered too expensive when the analysis was 172 initially performed. 173 174 2.3 Data Analysis 175 Because the TRFLP data were sample-specific, they could be used to investigate spatial 176 variation on a scale of a few centimetres. Patchiness in the abundance of dominant taxa in 177

each microorganism type was assessed by calculating dispersion indices. Morisita's Index,  $I_{M}$ , is a simple, global statistic that quantifies spatial patchiness. If the abundance of an OTU varies randomly in space, the observations follow a Poisson distribution; in this case  $I_M$ approximates unity. If abundance has a clumped (under-dispersed) pattern,  $I_M > 1$ . Conversely, in a regular (over-dispersed) arrangement,  $I_M < 1$ . The deviation from random expectation ( $I_M$  = 1) can be tested using critical values of the Chi-squared distribution with n-1 degrees of freedom. In this way, Morisita's index may be used to detect the presence of spatial structure.

186 The spatial structure of the epilithic community was also investigated with correlograms.

178

179

180

181

182

183

184

185

191

187 Correlograms indicate the degree of spatial correlation between samples at different 188

separation distances (Legendre & Legendre, 1998). In this case, the degree of correlation

189 between samples at different locations was summarised by the Mantel statistic (Mantel,

190 1967). The Mantel statistic indicates the degree of correlation between two matrices,

specifically, the correlation between a matrix of ecological dissimilarity and a matrix of

geographical distances (i.e. the physical separation between sample locations) derived from the same community. The significance of the Mantel statistic was estimated by means of permutation tests involving 999 randomisations of the ecological distance matrix. The Mantel correlograms were implemented using the *mantel.correlog* function in the *vegan* package (Oksanen, *et al.*, 2011).

Percent similarity (*PS*) was calculated from presence-absence data extracted from the TRFLP analysis and used to compare community composition across all the sampling locations for each microorganism type separately. Percent similarity is a metric based on similarities in community composition (Faith, *et al.*, 1987). In principle, *PS* figures provide an indicator of spatial heterogeneity in the microbial communities. High mean *PS* between samples implies a spatially homogeneous community (essentially the same taxa are observed in different locations). Conversely, low mean *PS* implies a relatively high degree of spatial variation in community composition, i.e. there are marked differences in community composition from location to location.

The overall taxonomic diversity for each microorganism type, expressed in terms of the Shannon index, H, and Simpson's index, J, was calculated from the pyrosequencing analysis (Hill, *et al.*, 2003). Community structure was investigated with rank-abundance plots which illustrated both taxonomic richness and evenness. Rank-abundance distributions (RADs) have been used to infer the mechanisms that underlie community diversity (Tokeshi, 1993, McGill, *et al.*, 2007). Microbial communities have been found to follow several different distributions. Curves describing distributions frequently reported in published literature (i.e. geometric, lognormal and power law models) were fitted to the pyrosequencing data using the *rad.fit* function in the *vegan* package (Oksanen, *et al.*, 2011).

#### 3 Results

### 3.1 Spatial structure

Each microorganism type was dominated by just a handful of taxa: for example, the taxa illustrated in Fig. 1 accounted for 41% (algae), 68% (fungi) and 40% (bacteria) of total abundance in their respective categories. The abundance of the dominant OTUs varied across the surface of the slab (Fig. 1). Morisita indices were calculated for each of these OTUs (values given on Fig. 1). In each case, a highly significant degree of under-dispersion (clumping) was observed (p < 0.001), with the most marked clumping in the eukaryotic OTUs. At a community level, diversity varied across the surface of the slab in a similar

229 fashion (results not shown). Fungal and bacterial diversity were positively correlated: sites 230 with high fungal diversity also tended to have high bacterial diversity (Spearman rank 231 correlation: r = 0.4, p = 0.05). Algal diversity was uncorrelated with bacterial and fungal 232 diversity (Spearman rank correlation: algal-fungal r = 0.3, p = 0.14; algal-bacterial r = -0.1, p 233 = 0.69). The mean PS figure for the algal samples was very high ( $\sim$ 90%). Bacterial 234 communities exhibited intermediate levels of PS (64%). The figure for fungal communities 235 was lower (50%). 236 237 Fig. 1 238 239 The Mantel correlograms based on the TRFLP analysis indicated significant spatial structure 240 in algal and fungal samples (p < 0.05) (Fig. 2). Positive correlation was observed at the 241 smallest spatial scale (< 8 cm) in these samples (Fig. 2a, b). No significant spatial structure 242 was apparent in the bacterial communities (Fig. 2c). 243 244 Fig. 2 245 246 3.2 Community composition 247 A total of 74 algal OTUs, 244 fungal OTUs and 486 bacterial OTUs were identified by 248 amplicon pyrosequencing (Table 1). The proportion of singletons was high in each case 249 (34% of algal reads, 60% of fungal reads and 54% of bacterial reads). 250 251 In total, 2648 reads were classified as algae (Table 2). The sequences were divided between 252 two phyla, the Chlorophyta and the Charophyta. The majority of sequences were associated 253 with the Chlorophyta; sequences from the class Trebouxiophyceae were particularly 254 abundant, accounting for 93.5% of the Chlorophyte reads. Most of these sequences 255 belonged to the Prasiolales. In the Charophyta, the Klebsormidiophyceae were particularly 256 abundant. 257 258 A total of 2626 fungal sequences were analysed. Two fungal phyla, the Ascomycota and 259 Basidiomycota, accounted for 82.1% of these sequences, with each phylum having 260 approximately equal representation (43.9% Ascomycetes, 38.2% Basidiomycetes). Overall, 261 17.9% of fungal reads were unclassified at a phylum level. The sequences from the 262 Ascomycota were overwhelmingly from the subphylum Pezizomycota; the majority of these 263 reads (811 out of a total of 1147) were unclassified below this level. Of the Ascomycete

reads classified at a higher level of resolution, sequences associated with the family

Capnodiales (class Dothideomycetes) were prominent (e.g. the diametiaceous fungi

264

266 Cladosporium sp. and Batcheloromyces sp.) In the Basidiomycota, most reads were 267 associated with the Agaricomycotina (915 reads out of a total of 999 Basidiomycetes). The 268 Agaricomycetes accounted for the bulk of these sequences. Lichenised fungi were not 269 detected. 270 271 There were 2401 bacterial reads of which 98.5% were resolved to at least phylum level. 272 Eleven bacterial phyla were represented, however, only three phyla, the Acidobacteria, 273 Actinobacteria and Proteobacteria, accounted for >5% of the total (Table 2). Almost all the 274 Actinobacterial reads were associated with the order Actinomycetales; the sub-orders 275 Frankineae and Pseudonocardinae were notably abundant. Proteobacterial reads were 276 dominated by the Alphaproteobacteria, with the Rhizobiales and Sphingomonadales 277 prominent. However, levels of taxonomic resolution were generally low and most taxa were 278 rare. 279 Table 1 280 281 282 Table 2 283 284 The RADs derived from the pyrosequencing analysis were characterised by high levels of dominance, a distinctive 'hollow' (concave up) profile and long tails of rare species (Fig. 3). 285 286 Omitting singletons from the rank-abundance plots did not fundamentally change this shape, 287 which closely approximated a zipf (power law) distribution. 288 289 Fig. 3 290 291 292 4 Discussion 293 294 Spatial variability 295 The Baas-Becking hypothesis, often summarised as "everything is everywhere - the 296 environment selects" has been a remarkably persistent concept in microbial biogeography 297 (Martiny, et al., 2006). However, the hypothesis does not explicitly address issues of spatial 298 scale and the factors driving the distribution of microorganisms at intermediate scales (from a 299 few centimetres to a few metres) are largely unknown. If the 'everything is everywhere' 300 model applied to our sample slab, then a spatially random distribution of microbes might be

expected, given the apparent environmental homogeneity of this surface (in terms of surface

texture, meteorological conditions, etc.) However, this was not the case for eukaryotic microbes.

303304305

306

307

308

309

310

311

312

313

314

315

316

317

318

319

320

321

322

323

324

325

326

327

302

The mean PS figure for the algae was high, indicating a spatially homogeneous community (at least in terms of species composition): the same taxa frequently co-occurred. Despite this, spatial patchiness over short distances (<8 cm, or approximately equal to the spacing between adjacent sampling locations) was observed. In other words, at this scale each sampling location was positively correlated with adjacent sampling locations, on average. In general terms, spatial pattern in ecological communities can arise due to environmental patchiness (e.g. resource availability or disturbance), contagious ecological processes (e.g. dispersal, reproduction) or a combination of both (Fortin & Dale, 2005) In this case, the observed spatial patchiness may have been related to the contagious spread of an abundant species. In the pyrosequencing analysis, the most abundant OTU was associated with a Klebsormidium species, a filamentous alga which could structure the algal community by contagious spread. However, it is more likely that the spatial structure observed represents differences in OTU abundance (i.e. patch intensity), rather than the presence/absence of particular taxa, because community structure was similar across the sampling locations. In other words, what we observed was a compositionally homogeneous community that varied in relative abundance of different species. The contour plot of the dominant algal OTUs (Fig. 1), which features pronounced hotspots of abundance, and the  $I_M$  values for the same taxa supported this view. Rindi and Guiry (2003) and Cutler et al. (2013) noted that green algal communities can exhibit small-scale spatial heterogeneity. Our results are consistent with these observations. The presence of green algal patches on stone surfaces has been linked to geochemical and geophysical changes in stone (Cutler, et al., 2013), so understanding characteristic scales of variation may assist in understanding patterns of biodeterioration/bioprotection.

328329

330

331

332

333

334

335

336

337

338

The fungal community had a low mean *PS* figure. This indicated considerable differences in community composition between sampling locations. The distribution of the most abundant taxa was also highly patchy (Fig. 1). Spatial variation in fungal communities has been described previously (Cutler, *et al.*, 2013) and, as with algae, may be a function of environmental or ecological variation (or both). Like the algae, the fungi exhibited short-range spatial autocorrelation (Fig. 2b). When viewed in the context of low mean *PS*, this indicates sharp changes in community composition and species abundance across short spatial distances. Spatial pattern in the fungal samples could be due to the short-range development of patchy surface molds, although the low level of taxonomic resolution in many OTUs made it impossible to be definitive. As with algae, fungal activity has been closely linked with stone

biodeterioration, so it is likely that the patchy structure of abundant fungal OTUs is closely correlated with patchy discoloration and/or surface degradation.

The bacterial community had a moderate mean PS value as there was a high level of cooccurrence amongst the dominant OTUs. The  $I_M$  values calculated for these OTUs
suggested spatial clumping. However, the associated contour plots indicated that the
patches were larger and less intense than those observed for the eukaryotes (Fig. 1). The
Mantel correlogram, which was based on the whole bacterial community, did not indicate
significant spatial structure at the scale of measurement (Fig. 2c). It is possible that spatial
structure was present at a scale smaller than the sampling unit. Whereas filamentary
microorganisms (e.g. mycelium-forming fungi and certain algal species) may spread widely,
the spatial range of assemblages of isolated bacterial cells may be limited. The absence of
pattern in the bacterial community contrasted with the algal and fungal communities and
suggested homogeneity in species composition and patch intensity. Interestingly, diversity in
the heterotrophic fungal and bacterial communities was positively correlated, suggesting
diversity 'hotspots' across the surface of the slab. Further study would be required to
establish the factors that promote diversity in the heterotrophic community.

This study only concentrated on the two-dimensional arrangement of microbes in a thin biofilm. Whilst epilithic microbes have an extremely important impact on stone conservation (not least because they can cause surface discolouration), endolithic microbes also play a significant role in stone biodeterioration. Future studies could seek to establish the three-dimensional structure of lithobiontic communities, perhaps by analysing thin, stone sections from different depths at each sampling location.

#### 4.2 Community composition and structure

In taxonomic terms, the microbial community was broadly similar to those previously reported for other stone substrates in temperate climates (e.g. John, 1988, Flores, *et al.*, 1997, Burford, *et al.*, 2003, Rindi & Guiry, 2004). The distinctive hollow shape of the RADs derived from the pyrosequencing data was also familiar from previous studies of microbial communities in different settings (e.g., Gans, *et al.*, 2005, Pommier, *et al.*, 2010, Inceoglu, *et al.*, 2011). Therefore, the high-resolution molecular study, whilst it provided additional detail on community composition and diversity, was largely consistent with previous models of microbial community structure.

#### 4.2.1 Taxonomic composition

375 A review of literature on epilithic algal communities in Western Europe suggests a relatively 376 small pool of widely dispersed, cosmopolitan species (Cutler, et al., 2013). Our results are 377 consistent with this scenario. In common with previous studies, green algae from the 378 Chlorophyta (primarily of the order Trebouxiophyceae) were dominant. Microscopic 379 Charophyta, notably from the order Klebsormidiophyceae, were also abundant. 380 Klebsormidium spp. have been routinely reported on stone surfaces in humid habitats (e.g. 381 Ortega-Calvo, et al., 1993, Rindi & Guiry, 2003). 382 383 Several previous studies of lithobiontic fungal communities have indicated dominance by the 384 Ascomycetes (e.g. Gleeson, et al., 2010). This was not the case in our study: the 385 representation of Ascomycetes and Basidiomycetes was only slightly in favour of the former. 386 The relatively high proportion of Basidiomycetes and non-lithobiontic Ascomycetes may 387 indicate a high level of allochthonous material (e.g. spores) across the surface of the slab 388 (this material might be preferentially sampled when adhesive tape is used). Lichenized fungi 389 were unexpectedly absent. Lichens are often found in lithic habitats and poorly 390 developed/degraded (and unidentifiable) lichen thalli were present on the surface of the slab. 391 However, these areas were not directly sampled and it may be that the tape sampling 392 method we used is poorly suited to collecting material from firmly-adhered lichen thalli. In 393 contrast, ubiquitous, diametiaceous hyphomycetes were relatively abundant. Taxa such as 394 Cladosporium spp. are frequently observed in lithobiontic habitats (Burford, et al., 2003); 395 Batcheloromyces sp. is reported less commonly. 396 397 McNamara and Mitchell (2005) observed that bacterial communities on stone are typically 398 dominated by taxa drawn from five phyla, i.e. the Proteobacteria, Actinobacteria, 399 Bacteroidetes, Acidobacteria and low-GC Firmicutes group. All these phyla were present on 400 the slab, although only Actinobacteria and Proteobacteria were abundant. Bacterial 401 communities on stone are often closely related to soil communities and this was the case in 402 our study. Several authors have commented that Actinobacteria are particularly abundant in 403 temperate biofilms and this observation was consistent with our results (Scheerer, et al., 404 2009). The Proteobacterial taxa found on the slab are cosmopolitan and have been reported 405 in a wide range of habitats, so it is difficult to infer much from their presence. 406 407 4.2.2 Community structure 408 As expected, the bacteria were the most diverse group, followed by the fungi. Algal diversity

409

410

411

was relatively low. In both the bacteria and the fungi, the diversity metrics generated from the

pyrosequencing study were much higher than the equivalent figures derived from TRFLP

data (by a factor of two: Cutler, et al., 2012). This was unsurprising, as DNA fingerprinting

techniques such as TRFLP are known to underestimate microbial community richness (Lalande, et al., 2013). Interestingly, Shannon and Simpson diversity for the algae were remarkably similar for both the pyrosquencing and TRFLP data. It appeared that in the case of algae, rare species (undetectable by TRFLP) did not contribute greatly to overall diversity. This suggests that algal community structure may be adequately captured by TRFLP in certain circumstances.

A number of previous studies have suggested generic models of microbial RADs, including geometric, lognormal and power law models. Jackson et al. (2001), for example, reported that a geometric relationship best described early successional bacterial communities in an aquatic biofilm. However, it is likely that simple geometric models only apply where the members of the community are all competing for the same niche (Dunbar, et al., 2002) and this model was a poor fit for our data. Dunbar et al. (2002) proposed that the lognormal distribution is better suited to functionally and phylogenetically diverse assemblages and proposed the use of this distribution as a null model for microbial communities. Lognormal RADs can arise from the multiplicative effects of biotic and abiotic factors and this distribution does not necessarily depend on specific biological/ecological mechanisms. Lognormal distributions have been reported in several bacterial communities (Dunbar, et al., 2002, Doroghazi & Buckley, 2008) but did not capture the essential characteristics of the lithobiontic RADs. Our results are most consistent with the power-law distributions that have been observed in a range of microbial communities from different settings (Gans, et al., 2005, Pommier, et al., 2010, Inceoglu, et al., 2011).

#### 5 Conclusions

Lithobiontic communities, especially those dominated by subaerial green algae, have been characterised as low-diversity assemblages (John, 1988). However, our study demonstrated that microbial communities on building stone can be heterogeneous, both in terms of spatial distribution and taxonomic composition. Different components of the microbial community exhibited different spatial patterns. If the results of our study apply more widely, lithobiontic eukaryotes should exhibit spatial structure over intermediate (centimetre) spatial scales, as well as the large- (metre-) scale patchiness often found to be associated with varying aspect and exposure. Spatial structure in lithobiontic bacterial communities, if it exists, is likely to be at a smaller scale than our sampling interval. DNA fingerprinting techniques, despite their inability to detect rare taxa, may be adequate for profiling green algae in these settings. These findings have implications for understanding spatial heterogeneity in the biodeterioration of stone as the observed patchiness of fungal and algal varieties is likely to

be correlated with centimetre-scale variation in stone degradation and soiling. Further study is required to elucidate the ecological relationships between the species that comprise these communities and the factors that generate spatial patchiness in eukaryotes, but not prokaryotic microbes.

Funding
This work was supported by the Engineering and Physical Sciences Research Council (grant no. EP/G011338/1).

Acknowledgements

The authors are grateful for the helpful comments made by two anonymous reviewers. They

are also grateful for the assistance of APS Masonry, Oxford, during the preparation of this

paper. The Authors confirm there are no conflicts of interest.

#### References

- 464 Abarenkov K, Nilsson RH, Larsson K-H, et al. (2010) The UNITE database for molecular
- identification of fungi recent updates and future perspectives. New Phytol 186: 281-285.
- 466 Abarenkov K, Tedersoo L, Nilsson RH, et al. (2010) PlutoF-a web based workbench for
- ecological and taxonomic research, with an online implementation for fungal ITS sequences.
- 468 Evolutionary Bioinformatics 6: 189-196.
- 469 Acosta-Martinez V, Dowd S, Sun Y & Allen V (2008) Tag-encoded pyrosequencing analysis
- of bacterial diversity in a single soil type as affected by management and land use. Soil Biol
- 471 Biochem 40: 2762-2770.
- 472 Burford EP, Fomina M & Gadd GM (2003) Fungal involvement in bioweathering and
- 473 biotransformation of rocks and minerals. *Min Mag* 67: 1127-1155.
- 474 Cutler N & Viles H (2010) Eukaryotic microorganisms and stone biodeterioration.
- 475 *Geomicrobiol J* 27: 630-646.
- 476 Cutler NA, Chaput DL & van der Gast CJ (2014) Long-term changes in soil microbial
- 477 communities during primary succession. Soil Biol Biochem 69: 359-370.
- 478 Cutler NA, Oliver AE, Viles HA & Whiteley AS (2012) Non-destructive sampling of rock-
- dwelling microbial communities using sterile adhesive tape. J Microbiol Methods 91: 391-
- 480 398.
- 481 Cutler NA, Viles HA, Ahmad S, McCabe S & Smith BJ (2013) Algal 'greening' and the
- 482 conservation of stone heritage structures. Sci Total Environ 442: 152-164.
- 483 Cutler NA, Oliver AE, Viles HA, Ahmad S & Whiteley AS (2013) The characterisation of
- 484 eukaryotic microbial communities on sandstone buildings in Belfast, UK, using TRFLP and
- 485 454 pyrosequencing. Int Biodeter and Biodegr 82: 124-133.
- 486 Doroghazi JR & Buckley DH (2008) Evidence from GC-TRFLP that bacterial communities in
- 487 soil are lognormally distributed. PLOS ONE 3.
- 488 Dunbar J, Barns SM, Ticknor LO & Kuske CR (2002) Empirical and theoretical bacterial
- diversity in four Arizona soils. Appl Environ Microbiol 68: 3035-3045.
- 490 Edgar RC, Haas BJ, Clemente JC, Quince C & Knight R (2011) UCHIME improves sensitivity
- and speed of chimera detection. Bioinformatics 27: 2194-2200.
- 492 Faith DP, Minchin PR & Belbin L (1987) Compositional dissimilarity as a robust measure of
- 493 ecological distance. Plant Ecol 69: 57-68.
- 494 Flores M, Lorenzo J & Gómez-Alarcón G (1997) Algae and bacteria on historic monuments
- 495 at Alcala de Henares, Spain. Int Biodeter and Biodegr 40: 241-246.
- 496 Fortin M-J & Dale M (2005) Spatial Analysis A Guide for Ecologists. Cambridge University
- 497 Press, Cambridge.
- 498 Franklin JF (2005) Spatial pattern and ecosystem function: reflections on current knowledge
- 499 and future directions. Ecosystem Function in Heterogeneous Landscapes, (Lovett GM, Jones
- 500 CG, Turner MG & Weathers KC, eds), pp. 427-441. Springer, New York.
- 501 Franklin RB & Mills AL (2010) The importance of microbial distribution in space and spatial
- scale to microbial ecology. The Spatial Distribution of Microbes in the Environment, (Franklin
- 503 RB & Mills AL, eds), pp. 1-30. Springer, Dordrecht.
- Gaidos E, Rusch A & Ilardo M (2011) Ribosomal tag pyrosequencing of DNA and RNA from
- benthic coral reef microbiota: community spatial structure, rare members and nitrogen-
- 506 cycling guilds. *Environ Microbiol* 13: 1138-1152.
- 507 Gans J. Wolinsky M & Dunbar J (2005) Computational improvements reveal great bacterial
- diversity and high metal toxicity in soil. Science 309: 1387-1390.
- 509 Gaylarde C & Gaylarde PM (2005) A comparative study of the major microbial biomass of
- 510 biofilms on exteriors of buildings in Europe and Latin America. *Int Biodeter and Biodegr* 55:
- 511 131-139.
- 512 Gleeson DB, Clipson N, Melville K, Gadd GM & McDermott FP (2005) Characterization of
- fungal community structure on a weathered pegmatitic granite. *Microb Ecol* 50: 360-368.
- 514 Gleeson DB, Melville K, McDermott FP, Clipson N & Gadd GM (2010) Molecular
- 515 characterization of fungal communities in sandstone. *Geomicrobiol J* 27: 559-571.
- Gorbushina AA (2007) Life on the rocks. Environ Microbiol 9: 1613-1631.

- 517 Hill TCJ, Walsh KA, Harris JA & Moffett BF (2003) Using ecological diversity measures with
- 518 bacterial communities. FEMS Microbiol Ecol 43: 1-11.
- 519 Inceoglu O, Abu Al-Soud W, Salles JF, Semenov AV & van Elsas JD (2011) Comparative
- analysis of bacterial communities in a potato field as determined by pyrosequencing. *PLOS ONE* 6.
- Jackson CR, Churchill PF & Roden EE (2001) Successional changes in bacterial
- assemblage structure during epilithic biofilm development. *Ecology* 82: 555-566.
- John DM (1988) Algal growths on buildings: a general review and methods of treatment.
- 525 Biodeterioration Abstracts 2: 81-102.
- Jones RT, Robeson MS, Lauber CL, Hamady M, Knight R & Fierer N (2009) A
- 527 comprehensive survey of soil acidobacterial diversity using pyrosequencing and clone library 528 analyses. *ISME J* 3: 442-453.
- Lalande J, Villemur R & Deschenes L (2013) A New Framework to Accurately Quantify Soil
- Bacterial Community Diversity from DGGE. *Microb Ecol* 66: 647-658.
- Legendre P & Legendre L (1998) *Numerical Ecology*. Elsevier, Oxford.
- Mantel N (1967) The detection of disease clustering and a generalized regression approach.
- 533 Cancer Res 27: 209-220.
- Martiny JBH, Bohannan BJM, Brown JH, et al. (2006) Microbial biogeography: putting
- 535 microorganisms on the map. *Nat Rev Microbiol* 4: 102-112.
- 536 McGill BJ, Etienne RS, Gray JS, et al. (2007) Species abundance distributions: moving
- beyond single prediction theories to integration within an ecological framework. *Ecol Lett* 10:
- 538 995-1015.
- 539 McNamara CJ & Mitchell R (2005) Microbial deterioration of historic stone. Front Ecol
- 540 Environ 3: 445-451.
- Nilsson RH, Veldre V, Hartmann M, et al. (2010) An open source software package for
- automated extraction of ITS1 and ITS2 from fungal ITS sequences for use in high-throughput
- community assays and molecular ecology. *Fungal Ecol* 3: 284-287.
- Nunan N, Young IM, Crawford JW & Ritz K (2010) Bacterial interactions at the microscale:
- linking habitat to function in soil. The Spatial Distribution of Microbes in the Environment,
- 546 (Franklin RB & Mills AL, eds), pp. 61-85. Springer, Dordrecht.
- Oksanen J, Blanchet FG, Kindt R, et al. (2011) vegan: Community Ecology Package.
- 548 <a href="http://CRAN.R-project.org/package=vegan">http://CRAN.R-project.org/package=vegan</a>.
- Ortega-Calvo JJ, Hernadez-Marine M & Saiz-Jimenez C (1993) Experimental strategies for
- 550 investigating algal deterioration of stone. Proceedings of 7th International Congress of
- Deterioration and Conservation of Stone, Vol. 1 (Rodrigues JD, Henriques F & Jeremias FT,
- 552 eds), pp. 541-549. Lisbon, Portugal.
- Pace NR (1997) A molecular view of microbial diversity and the biosphere. Science 276:
- 554 734-740.
- Pommier T, Neal PR, Gasol JM, Coll M, Acinas SG & Pedros-Alio C (2010) Spatial patterns
- of bacterial richness and evenness in the NW Mediterranean Sea explored by
- 557 pyrosequencing of the 16S rRNA. Aquat Microb Ecol 61: 212-224.
- Quast C, Pruesse E, Yilmaz P, et al. (2013) The SILVA ribosomal RNA gene database
- project: improved data processing and web-based tools. *Nucleic Acids Res* 41: D590-D596.
- Quince C, Lanzen A, Curtis TP, et al. (2009) Accurate determination of microbial diversity
- from 454 pyrosequencing data. *Nat Methods* 6: 639-U627.
- Rindi F (2007) Diversity, distribution and ecology of green algae and cyanobacteria in urban
- habitats. *Algae and Cyanobacteria in Extreme Environments*, (Seckbach J, ed) pp. 619-638.
- 564 Springer, Dordrecht.
- 565 Rindi F & Guiry MD (2003) Composition and distribution of subaerial algal assemblages in
- 566 Galway City, western Ireland. *Cryptogamie Algologie* 24: 245-267.
- 567 Rindi F & Guiry MD (2004) Composition and spatial variability of terrestrial algal
- assemblages occurring at the bases of urban walls in Europe. *Phycologia* 43: 225-235.
- Scheerer S, Ortega-Morales O & Gaylarde C (2009) Microbial deterioration of stone
- 570 monuments an updated overview. Adv Appl Microbiol 66: 97-139.

571 572 573 574 575 576 577 578 579 580	Schloss PD, Westcott SL, Ryabin T, et al. (2009) Introducing mothur: open-source, platform-independent, community-supported software for describing and comparing microbial communities. <i>Appl Environ Microbiol</i> 75: 7537-7541.  Tedersoo L, Nilsson RH, Abarenkov K, et al. (2010) 454 Pyrosequencing and Sanger sequencing of tropical mycorrhizal fungi provide similar results but reveal substantial methodological biases. <i>New Phytol</i> 188: 291-301.  Tokeshi M (1993) Species abundance patterns and community structure. <i>Adv Ecol Res</i> 24: 111-186.  Wang Q, Garrity GM, Tiedje JM & Cole JR (2007) Naive Bayesian classifier for rapid assignment of rRNA sequences into the new bacterial taxonomy. <i>Appl Environ Microbiol</i> 73:
581 582	5261-5267. Warscheid T & Braams J (2000) Biodeterioration of stone: a review. <i>Int Biodeter and Biodegr</i>
583 584	46: 343-368.
585 586	
300	
	Z'O'
	. (7)

587	Figure captions				
588					
589	Fig. 1: Contour plots indicating the varying abundance (measured in fluorescence units) of				
590	dominant OTUs for each microorganism type. $I_M$ = Morisita index; in each case the value is				
591	significantly > 1 (p < 0.001), indicating a patchy distribution. The white dots indicate sampling				
592	locations.				
593					
594	Fig. 2: Mantel correlograms for each microorganism type; filled circles indicate significant (p				
595	< 0.05) correlation in a given distance class.				
596					
597	Fig. 3: Rank-abundance plots for algal, fungal and bacterial OTUs (each OTU is represented				
598	by a point), based on amplicon pyrosequencing. Singletons have been omitted. The taxon				
599	ranked 1 is the most abundant in each case. The plots have been fitted with a zipf (power				
600	law) model, indicated by a red line.				
601					

## 602 Tables

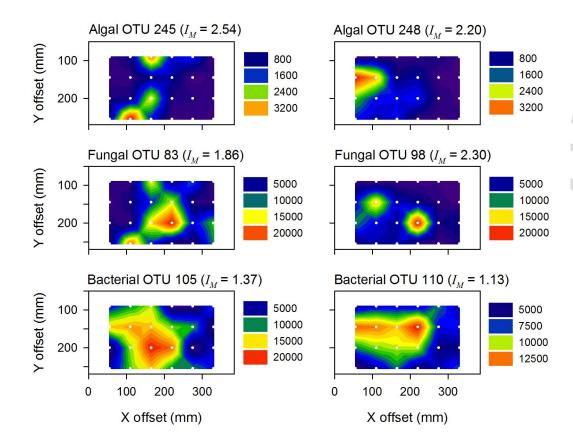
	No. OTUs	Shannon's diversity, <i>H</i>	Simpson's diversity, $J$
Algae	74	2.3	0.81
Fungi	244	3.0	0.83
Bacteria	486	4.2	0.89

Table 1: Diversity metrics derived from amplicon pyrosequencing

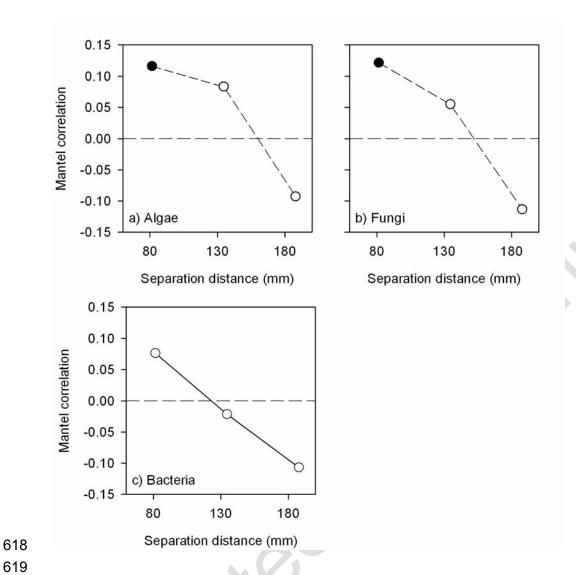
Alexan			Proportion of reads
Algae		No. reads	(%)
Chlorophyta			
Trebouxiophyceae		1454	54.9
Others		101	3.8
Charophyta			
Klebsormidiophyceae		1087	41.0
Others		6	0.2
	Totals	2648	100.0
Fungi			
Ascomycota			
Dothideomycetes <sup>a</sup>		314	12.0
Others <sup>b</sup>		840	32.0
Basidiomycota			
Agaricomycetes		965	36.7
Others		39	1.5
Unclassified at phylum level		468	17.8
	Totals	2626	100.0
Bacteria			
Acidobacteria			
Acidobacteria		127	5.3
Others		5	0.2
Actinobacteria			
Actinobacteridae		440	18.3
Others		55	2.3
Proteobacteria			
Alphaproteobacteria		1579	65.8
Others		88	3.7
Other bacterial phyla		107	4.5
	Totals	2401	100.0

Table 2: Summary of pyrosequencing results; phyla are indicated with bold text

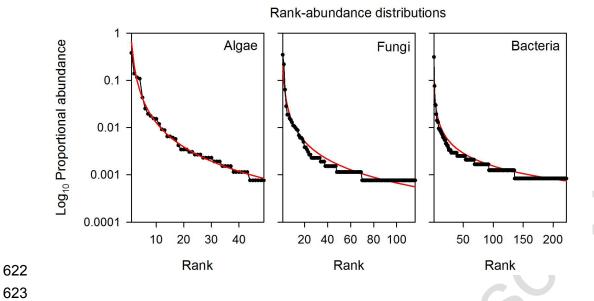
# **Figures**



616 Fig. 1



620 Fig. 2 



624 Fig. 3