1	Habitat complexity and food item size modify the foraging behaviour of a freshwater
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25 Abstract

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The functional response describes the relationship between feeding rate and prey density, and 27 28 is important ecologically as it describes how the foraging behaviour of an animal changes in response to food availability. The effects of habitat complexity and food item size was 29 experimentally tested here on the foraging parameters and the functional responses of the 30 freshwater fish roach Rutilus rutilus (L.). Habitat complexity was varied through the 31 manipulation of substrate and turbidity, and food item size was varied by using fishmeal 32 33 pellets, in two sizes. As water turbidity and substrate complexity increased, the reaction distance and consumption rate of the fish significantly decreased. Increased food item size 34 significantly decreased consumption rates but had no influence on any other foraging 35 36 parameter. Analysis of the interactions between substrate complexity, turbidity and food item size revealed food item size had the greatest influence on consumption rate. Turbidity has the 37 least effect on all the foraging parameters tested. Across all experiments, the functional 38 39 responses were best described by the Type II response, a relatively consistent finding for R. *rutilus*. These outputs reveal that fish foraging behaviours and functional responses are highly 40 context dependent, varying with environmental parameters and the availability of prey of 41 different sizes. 42 43

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47 Key words: Foraging behaviour, Type II functional response, consumption rate, prey size,
48 abiotic, *Rutilus rutilus*

49 Introduction

The functional response is the relationship between the feeding rate of a forager and its prey 51 density, and is used to describe and model foraging behaviour (Solomon 1949; Holling 1959; 52 Holling 1966). It is an ecologically important metric as under conditions of limited resource 53 availability, individuals will attempt to maximise their energy intake whilst minimising the 54 55 costs associated with prey searching and handling (Galarowicz & Wahl, 2005; Oyugi et al. 2012a,b; Murray et al. 2013). Consequently, measuring how animals respond to variations in 56 food availability helps the understanding of how individuals optimise their foraging 57 58 behaviour (Werner et al. 1983; Galarowicz & Wahl, 2005; Murray et al. 2013). This provides knowledge to assist interpretation of the effect of prey availability on consumer condition, 59 growth and fitness (Werner et al. 1983). Moreover, functional responses provide insights into 60 61 the mechanics of consumer-prey relationships that can have cascading effects through the food web (Koski & Johnson, 2002). They have considerable ecological applications with, for 62 63 example, their use as important parameters within individual based models (e.g. Stillman 2008) and as explanatory variables in the success of invasive species (e.g. Bollache et al. 64 65 2008; Dick et al. 2013).

Due to how consumers can influence the structure and stability of their prey populations 67 (Alexander et al. 2013), it is ecologically important to distinguish the type of functional 68 69 response being exhibited (Murray et al. 2013). There are three major function response types: I, II and III (Hassell et al. 1977). Type I describes a linear increase in feeding rate with prey 70 71 density until it reaches a constant value at saturation (Jeschke et al. 2004) whereas the feeding rate of the Type II response increases at a decreasing rate with prey density until it reaches it 72 maximum value (Holling 1959; Murray et al. 2013). Type II is thus inversely density-73 74 dependent and so for the prey population, mortality risks decrease with increasing density

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(Jeschke & Hohberg, 2008). The Type III response describes a sigmoidal, density-dependent
relationship, where an initial increasing risk of prey mortality switches to a decreasing risk of
mortality as the prey density increases above a threshold level (Real 1979; Morgan & Brown,
1996).

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80 Despite their apparent simplicity, it has been established that functional responses are not fixed within specific predator-prey relationships; conversely, under different contexts, 81 foraging and anti-predator behaviours can shift and significantly alter the form of the 82 83 response (Holling 1959; Alexander et al. 2013). This may involve subtle changes in, for example, the ability of the consumer to detect and respond to the presence of prey items, or 84 may even involve a shift in the functional response type should there be, for example, a 85 86 substantial increase in the time spent foraging (Abrams 1982). Environmental variables that have been found to influence functional responses, e.g. temperature and light levels (Lipcius 87 & Hines, 1986; Koski & Johnson, 2002), and also habitat structure (Alexander et al. 2012). 88 89 Indeed, habitat structure and complexity have been found to both alter the search ability of the consumer (Savino & Stein, 1989; Heck & Crowder, 1991) and the refuge area of their 90 prey (Gotceitas 1990; Warfe & Barmuta, 2004; Alexander et al. 2012). Prey body size might 91 also be important in determining the values of foraging parameters, given trade-offs between 92 the ease of detection of larger items versus their increased handling time and energetic 93 94 profitability (Wankowski & Thorpe 1979; Bean & Winfield, 1983; Oksanen & Lundberg, 1995). Nevertheless, there remains a paucity of information on how foraging behaviours are 95 modified between simple and more complex habitats for specific taxonomic groups, such as 96 97 many benthic feeding freshwater fishes.

98 Consequently, the aim of this study was to determine how effects of habitat complexity 99 and prey item size, and their interactions, modify the foraging parameters and functional 100 response of a model freshwater fish, whose foraging behaviour is generally reliant on visual cues (Hielm et al. 2003; Bogacka-kapusta 2007). The model species selected was roach 101 Rutilus rutilus (Diehl 1988; Murray et al. 2013), a freshwater fish ubiquitous to many 102 103 temperate European freshwaters (e.g. Lappalainen et al. 2008) and invasive in others (e.g. Elvira & Almodovar, 2001; Winfield et al. 2011). Their ecological importance includes their 104 105 potential for invoking cascading effects on freshwater ecosystems through their high zooplankton grazing rates (e.g. Jeppesen et al. 2010) and thus understanding the context-106 dependency of their foraging behaviours and functional responses can be ecologically 107 significant. Previous studies have indicated R. rutilus can exhibit a Type II (e.g. Johanson & 108 Persson 1986; Persson 1987; Murray et al. 2013) and Type III (Winkler & Orellana, 1992) 109 110 response. These were, however, based on a range of field and experimental approaches, 111 making inter-study comparison of outputs difficult. Thus, here we build on the study of Murray et al. (2013) who used highly controlled experimental conditions to reveal that in a 112 simple environment R. rutilus demonstrated a Type II response. We tested the prediction that 113 increases in both habitat complexity and food item size will significantly modify R. rutilus 114 foraging behaviour through significantly reduced reaction distance and consumption rate, and 115 increased searching time, handling time and reaction distance, impacting their functional 116 117 response. 118

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123 Methods

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125 Ethical note

All animal work was conducted in accordance to national and international guidelines to minimize discomfort to animals. All regulated procedures completed under the Animals (Scientific Procedures) Act 1986 were licensed by the UK Home Office under project licence number PPL 30/2626. The Ethics Review Panel of the School of Applied Sciences of Bournemouth University approved this project licence.

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132 Experimental design overview

The experimental design used replicated groups of three R. rutilus individuals in 133 134 experimental arenas, exposed them to different numbers of prey items (10, 25, 50, 100 or 150) and captured their foraging behaviours using a two-camera videography system. One 135 camera was positioned horizontally, facing the side of the tank, with the second camera 136 positioned vertically, above the surface of the water. The actual positions and distances 137 moved by the fish were calculated using trigonometry based on footage from both cameras 138 (Murray et al. 2013). The specific details of the experimental arena, video capture, validation 139 of fish movement data, and the use of the Hollings Disc equation for the Type II functional 140 response are detailed in Murray et al. (2013). In summary, there were 12 behavioural arenas 141 142 (fish aquaria of $0.46 \times 0.31 \times 0.39$ m) in the experiments that were maintained at 18° C on a 12:12 hour light/ dark regime. Three randomly selected roach from a batch of 78 fish 143 (average length = $129 \text{ mm} \pm 2.5 \text{ mm}$; age 1+ years) were introduced into each arena and 144 145 allowed to acclimatize to the tanks for 14 days prior to the start of the experiments.

146 Throughout the acclimatization and experimental period, the food items used were 147 pelletized fish-meal ('pellets'). This was due to: (i) the experimental fish were originally 148 farmed fish that had been reared on pellets and so were used to consuming them; (ii) cyprinid fish (such as *R. rutilus*) tend to respond well to such pellets in foraging experiments in tanks 149 (e.g. Britton et al. 2012; Oyugi et al. 2012a,b); (iii) as a non-motile 'prey' item that can 150 151 neither actively select a refuge area, nor display evasive behaviour, measuring the effect of habitat complexity on the consumer would not be confounded by changes in the behaviour of 152 their prey; and (iv) pellets are available in different sizes so food item size could be easily 153 and accurately manipulated. Note that when compared to live prey, the use of pellets 154 precludes the display of more complex foraging behaviours under certain conditions, as prev 155 156 mobility has been shown to influence feeding rates both negatively, through the confusion effect (whereby large numbers of evasive prey can reduce attack rates and/or capture 157 efficiencies) (Ioannou et al. 2007; Tosh et al. 2009), or positively, with the movement of prey 158 159 items increasing predator reaction distance, especially in turbid environments (Utne-Palm 1999). 160

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During the experimental period, a single feeding trial, across all the tanks was conducted 162 every other day, with feeding on the day in between comprised of a maintenance ration of 163 approximately 1.5 % body weight. A maintenance ration was used rather than ad libitum to 164 ensure feeding motivation on the experimental days, given that functional responses relate to 165 optimal foraging and therefore behaviour seeking to maximise net energy gain should be 166 167 promoted. Thus, feeding on experimental days occurred 24 hours after the last exposure to the maintenance ration. Each feeding trial consisted of exposing each tank of fish, in turn, to 168 one randomly selected food density of 10, 25, 50, 100 or 150 pellets per tank (equivalent to 169 75, 187, 375, 750 and 1125 items m^{-2} respectively). By discounting a food density previously 170 used in a tank, eventually each of these food item densities was used across all 12 arenas, 171 with the process then being repeated once more (i.e. each food item density was used twice in 172

each tank). During the trials, the pellets were introduced to the tank across the entire surface
of the water with all pellets sinking through the water column and settling on the base of the
tank, with pellets being taken by the fish both as they fell through the water and once they
had settled on the bottom of the tank. On the release of the food, the filming of the foraging
behaviour commenced for 10 minutes (Oyugi et al. 2012a,b). At the end of this period, all
uneaten food was removed immediately using a siphon.

179

180 *Experimental treatments*

181 To test the effect of habitat complexity and food item size on the foraging parameters and the 182 functional response, the manipulated parameters were substrate complexity, water turbidity 183 and food item size. The effect of substrate complexity was tested first and then turbidity and 184 food item size.

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To test the effect of substrate complexity, the treatments were (1) arenas with no substrate (i.e. simply the glass bottom of the arena) (n=6) and (2) arenas with a layer of dark aquarium gravel (2 to 5 mm) of approximately 10 mm depth on the arena bottom to represent the complex substrate (n=6). Other than the change in substrate, the arenas were identical regarding water turbidly (clear) and food item size (1 mm pellets). These trials were completed separately from the trials of water turbidity and food item size, and used different fish. A table listing the experimental treatments is included in Online Resource 1.

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To test the effect of water turbidity and food item size, a two-factor experimental design was used as it enabled testing of the influence on foraging of both factors and their interactions. These two factors were used together as their interactions will be important in more natural systems where habitats are already complex and their interactions are likely to 198 have synergistic effects on a visual forager. Across the 12 arenas, 6 were used with clear water and 6 with water turbidity being increased through addition of a fine powder of 199 bentonite clay to the arena $(1g \pm 0.1g)$ 5 minutes prior to the experiments commencing. This 200 201 was as per Vollset & Bailey, (2011) who demonstrated the method had no harmful effects to the fish. At the end of each feeding trial, the water turbidity of each arena was quantified 202 through measurement with a turbidity meter (Hanna Instruments, HI 93703 Micro processor, 203 www.hannainst.co.uk), average turbidity in the increased turbidity areas being recorded as 204 equal to 3.41 ± 0.5 FTU, compared to the clear treatments (average = 0.01 ± 0.0 FTU), with 205 this difference being significant (ANOVA: $F_{1.528} = 110.43$, P < 0.01). As the clay settled out 206 of solution in approximately six hours, it was then able to be removed by siphoning. The 207 208 arenas used as clear and turbid water treatments remained constant throughout the 209 experiments.

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Across these 12 tanks of varying turbidity, two different sizes of sinking pellets were used: 211 1 mm and 2 mm; the numbers released across the trials were as per the substrate experiment 212 (cf. Experimental design). Whilst this meant at a given food density, the biomass of food 213 being introduced would differ between the sizes of pellet, this was justified through 214 functional response analyses generally being based on the consumption rate according to food 215 item density. During each experimental food exposure, the density of food items used was 216 217 selected randomly for each tank. Once each density had been tested, the trials were later repeated, i.e. each food item density was tested twice in each tank for both pellet sizes. The 218 actual size of pellet used alternated from one experimental food exposure to the next. A table 219 220 listing the experimental treatments is included in Online Resource 1.

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223 Data capture

The recorded footage of each feeding trial in both sets of experiments was analysed using a 224 purpose-built event-logger program (Event; Bournemouth University 2012). This allowed 225 226 frame-by-frame viewing and estimation of the position of objects in three dimensions, enabling parameter estimates of fish foraging behaviour to be measured that formed the basis 227 of the functional response equations (Holling 1959; Murray et al. 2013). These parameters 228 were: (i) swimming speed (s) whilst searching for food, characterised by relatively slow 229 swimming, with frequent changes in body orientation and leading to food item capture; (ii) 230 231 reaction distance (d), determined as the distance a fish would travel in a straight line directly towards a food item, quickly followed by capture of the food item, often following a change 232 in body orientation towards the food item; and (iii) handling time (h), determined as the time 233 234 taken to move towards and consume a food item, and then be ready to consume a further food item. Handling time was determined on occasions when food items were captured in rapid 235 succession and when no other behaviour was observed between food item capture. Other 236 237 parameters recorded, but not used in the functional response equations were: (i) Consumption rate, which was estimated directly, taken as the time between a fish taking its first and fifth 238 food item, and expressed as the number of items consumed per second (Murray et al. 2013). 239 By repeating across the range of food densities, the shape of functional response was able to 240 be described; and (ii) Searching time, recorded as a percentage proportion of the total 241 242 foraging time that was spent actively searching. This was used to gain insight into the level of risk-taking behaviour displayed by the fish. When perceived risk is reduced, it has been 243 shown that fish will spend a greater proportion of their time searching for food as a 244 245 compromise between energy intake and potential risks (Milinski & Heller, 1978; Oksanen & Lundberg, 1995). 246

248 Data analysis

Across the feeding trials in both experiments, there were insufficient data points related to 249 forager parameters collected for each individual fish to enable analyses of their foraging 250 251 behaviour at that level. Consequently, for the parameters of swimming speed, reaction distance and handling time, separate average parameter values were calculated, based on the 252 arithmetic mean, for each food density and treatment, whilst limiting the number of data 253 points collected for each parameter from any one fish to four, limiting the potential impact of 254 pseudo-replication (Hurlbert 1984). Any further potential impact on the experimental 255 outcomes through familiarisation and learning of optimum feeding behaviour at the 256 experimental food item densities was limited by the use of maintenance rations and time 257 between trials of the same density. Given that the effect of substrate was tested separately 258 259 with respect to turbidly and food item size, with substrate trials using the same fish population, its effect on the foraging parameters used repeated measures ANOVA. When two 260 factors were being tested (turbidity and food item size experiment) then linear mixed effects 261 models were used, with either food item size or turbidity as a random effect (depending on 262 the test). When comparing the proportion of time spent searching (as a percentage of total 263 experimental time) binomial generalised linear models (GLM) were used. 264

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The foraging behaviour parameters were used to parameterise both a Type I (Jeschke et al. 2002; Jeschke et al. 2004; Murray et al. 2013) and Type II (Holling 1959; Murray et al. 2013) functional response equation, after being tested for density independence (Murray et al 2013). These used the same variables of attack rate (derived from swimming speed and reaction distance) and handling time, together with food item density, differing only in how these parameters were treated. Note that the selection and parameterisation of the functional response models, and the estimation of the foraging parameters, are described in more detailin Murray et al. (2013). The Type I model was:

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$$F = \begin{cases} aD & \text{if } D \le \frac{1}{ah} \\ \frac{1}{h} & \text{if } D > \frac{1}{ah} \end{cases}$$
(Eqn 1)

275 The Type II model was Holling's Disc Equation (Holling 1959):

276
$$F = \frac{aD}{1+aDh}$$
(Eqn 2)

277 Where F = feeding rate (items s⁻¹), a = searching rate (i.e. search area per unit time) (m² s⁻ 278 ¹), D = food density (items m⁻²) and h = handling time (s) (Holling 1959). In both cases a was 279 defined as:

280

$$a = 2ds$$
 (Eqn 3)

281 Where s = swimming speed (ms⁻¹) and d = reaction distance (m).

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Thus, the outputs provided the predicted functional response of the fish according to Type I and II equations. These were then compared to the observed functional response i.e. the observed consumption rate data, taken directly from the recorded footage. The best fit between predicted models and observed functional response was then determined by its lower value of the akaike information criterion (AIC) using linear regression models, with each factor (substrate, turbidity and food item size) being tested separately.

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Finally, to assess the relative influence of all three factors on the consumption rate and the foraging parameters, as the experimental conditions were the same across both sets of experiments, the data were combined for further testing using linear mixed effects models. To test the relative effects of the factors on each foraging parameter, food density and body length of individual fish were the covariates and experimental arena number was set as a random effect (to account for the fact that different experimental arenas were used across the two experiments). Depending on the model, consumption rate and foraging parameters were the dependent variables and were fitted through stepwise removal of non-significant terms according to non-significant *P* values. All statistics and testing were completed in R (R version 2.15.1) (R Development Core Team 2012), where reported below, average values include standard deviation (\pm).

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302 **Results**

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304 Substrate complexity
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The effect of increasing the complexity of the substrate on the foraging behaviours was a 305 significantly decreased reaction distance between the no-substrate (average = 13.2 ± 5.5 cm) 306 307 and substrate treatment (average = 7.3 ± 3.9 cm) (ANOVA: F_{1.18} = 6.75, P < 0.05). There was also a significant difference in searching time, with fish in the substrate treatment searching 308 309 longer (average = 91.0 \pm 3.7 %) than the no-substrate treatment (average = 28.5 \pm 3.1 %; GLM: $F_{1.67} = 99.04$, P < 0.01). By contrast, there were no significant differences between the 310 treatments for swimming speed and handling time (ANOVA: $F_{1,28} = 0.91$, P > 0.05 and $F_{1,10} =$ 311 0.28, P > 0.05 respectively). See Online Resource 2 for a summary of the behavioural 312 parameter values. 313

314

The effect of substrate complexity on the consumption rate of the fish was significant, with reduced rates in the substrate treatment (ANOVA: $F_{1,16} = 6.21$, P < 0.05; Fig. 1). Comparison between observed functional response and that predicted by the foraging parameters fitted to equations 1, 2 and 3 revealed that a predicted Type II response was the better fit in both substrate and non-substrate treatments (adjusted $R^2 = 0.94 F_{1,3} = 48.84$, P <

320	0.01 and adjusted $R^2 = 0.96$, $F_{1,3} = 64.86$, $P < 0.01$ respectively) compared to a Type I
321	functional response (adjusted $R^2 = 0.92$, $F_{1,3} = 53.55$, $P < 0.01$; adjusted $R^2 = 0.94$, $F_{1,3} =$
322	72.52, $P < 0.01$). Similarly, the Type II functional response was a better fit when compared to
323	a simple linear increase (adjusted $R^2 = 0.91$, $F_{1,3} = 68.65$, $P < 0.01$; adjusted $R^2 = 0.92$ $F_{1,3} =$
324	77.87, $P < 0.01$). Lastly, the predicted Type II functional response was a better fit than Type I
325	for both the substrate and non-substrate treatment according to AIC (predicted Type II:
326	substrate AIC = -51.15 ; non-substrate AIC = -44.96 ; predicted Type I: substrate AIC = $-$
327	31.42; non-substrate AIC = -14.97).

328

329 (Figure 1)

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331 Water turbidity and food item size

When controlling for the effect of food item size, the effect of increased water turbidity was a 332 significant increase in searching time, with fish searching significantly longer (average = 85.0333 ± 3.2 %) than in the clear treatments (average = 25.0 ± 2.9 %) (GLM: F_{1.69} = 58.21, P < 0.01). 334 Its effect on consumption rate was also significant, with reduced rates in turbid conditions 335 (LMEM: $t_{74} = -4.37$, P < 0.01; Fig. 2). There were, however, no significant differences for 336 swimming speed, reaction distance or handling time between the turbid and clear conditions 337 (LMEM: $t_{48} = 1.43$, P = 0.13; $t_{89} = -2.92$, P = 0.06; $t_{87} = 0.149$, P = 0.88 respectively). See 338 339 Online Resource B for a summary of the behavioural parameter values.

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When controlling for the effects of turbidity, increasing food item size resulted in a significant reduction in consumption rate (LMEM: $t_{74} = 2.51$, P = 0.02; Fig. 3). There was, however, no significant effect on searching time, swimming speed, reaction distance or handling time (GLM: $F_{1,69} = 2.53$, P = 0.05; LMEM: $t_{48} = 1.22$, P = 0.18; $t_{89} = 2.90$, P = 0.06and $t_{87} = -1.57$, P = 0.11 respectively).

346

347 (Figures 2 and 3)

348

349 (**Table 1**)

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The effect of turbidity on functional response was analysed separately for both food item 351 352 sizes. Under turbid conditions, the functional response closely matched a Type II response using both 1 mm and 2 mm pellets (Table 1). Furthermore, the Type II functional response 353 was a better fit compared to a Type I functional response for both food item sizes in the 354 355 turbidity treatment (Table 1). Similarly, the Type II functional response provided a better fit when compared to a simple linear increase (Table 1). Lastly, the predicted Type II functional 356 response was seen to be a better fit than Type I through lower values using Akaike's 357 Information Criterion (Table 1). 358

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Under clear water conditions, the functional response for both food item sizes closely matched a Type II response in both food item size treatments (Table 1). Furthermore, the functional response was a better fit when compared to a Type I functional response and a simple linear increase (Table 1). When the models were compared, the lower AIC values were always for the predicted Type II response rather than predicted Type I (Table 1).

365

366 Factors influencing observed behaviour

The linear mixed effects model (LME) output for all experimental factors combined, revealedthat substrate and food item size tended to have the greatest consequences for the foraging

parameters (Table 2, Online Resource B). The most significant effect on consumption rate was food item size (t_{223} = 8.36, *P* < 0.01), and for reaction distance and handling time it was substrate complexity ($F_{1,50}$ = 12.3, *P* < 0.01) and handling time by substrate complexity (t_{526} = 5.19, *P* < 0.05). Within the model, the effects of turbidity on the foraging parameters were not significant.

374

375 (**Table 2**)

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377 Discussion

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The experiments demonstrated that changes in habitat complexity and food item size had 379 significant consequences for the foraging parameters and functional responses of R. rutilus. 380 Thus, aspects of their foraging behaviour were influenced by both their environment and food 381 382 resources, and this requires consideration in studies that relate to either foraging behaviours and or functional responses. These responses to changing conditions are likely to relate to 383 their foraging being strongly reliant on visual cues (e.g. Aksnes & Utne, 1997; Diehl 1988; 384 385 Podolsrky, Uiblein & Winkler, 1995; Wanzenbock et al. 1996). Regarding the type of functional responses elicited by *R. rutilus*, the best fitting functional response model in each 386 experiment was always Type II. This is a similar outcome to most other studies on R. rutilus 387 388 (e.g. Johanson & Persson 1986; Persson 1987; Murray et al. 2013). The only exception is Winkler & Orellana, (1992) where Type III functional response was measured, with this 389 likely to relate to the role of capture probability as a result of evasive behaviour displayed by 390 the live prey. 391

393 Testing of how water turbidity impacted the foraging parameters whilst controlling for the effect of food item size revealed that consumption rate and reaction distance were reduced as 394 turbidity increased, with this likely to be a result of visual foraging behaviour in roach. It was 395 396 not considered likely that it was related to changes in olfactory cues as bentonite clay is considered odourless (Vollset & Bailey, 2011; Zamor & Grossman, 2007), plus the role of 397 olfaction (compared to visual cues) is limited in roach foraging (Wanzenbock et al. 1996). 398 This outcome is in contrast to findings in three-spined sticklebacks Gasterosteus aculeatus 399 whose reaction distance and attack rate actually increased as turbidity increased (Vollset & 400 Bailey, 2011). This outcome was related to the altered conditions; both reducing the 401 perceived conspicuousness of the stickleback to potential predators and increasing prey item 402 403 contrast in the water column, increasing their visibility (Vollset & Bailey, 2011). 404 Notwithstanding, as the attack rate of G. aculeatus increased their capture success actually decreased, resulting in the consumption rate actually remaining the same and the energetic 405 costs of foraging increasing (Vollset & Bailey, 2011). The use of pelletized fish meal in our 406 407 study meant that there was a much more limited role for capture success in shaping the outcome of the foraging, as evasion behaviour was negligible in this experiment and prey 408 refuge was not available, when the gravel substrate was not used. Similarly, Wanzenböck 409 (1995) described the role of prey size selection in altering handling times during foraging by 410 0+ roach and bleak (Alburnus alburnus); whereby, fish and prey size affected handling time 411 412 as well as feeding rate, and ultimately, prey profitability (Wankowski & Thorpe 1979; Buckel & Stoner 2000). Furthermore, Wanzenböck (1995) showed handling time and prey 413 profitability were seen to vary over the feeding period, as feeding motivation changed. This 414 effect will be limited in our study due to the homogeneity of food item size used within each 415 trial, limiting the potential for prey item selection, altered handling times and profitability. 416

417 The presence of a gravel substrate within the experimental arenas inhibited the majority of the foraging parameters. Within this, and similar, previous experiments (Murray et al. 2013), 418 food pellets were used due to their being a non-motile food item. This was to eliminate the 419 420 potential confounding effect of the food items actively seeking refuge in more cryptic environments that could result in any shifts in the foraging parameters being due to prey 421 rather than fish behaviour. Indeed, other studies have revealed that functional responses are 422 significantly affected when the refuge area for live prey is increased, as this provides greater 423 opportunities for prey avoidance (e.g. Gotceitas 1990; Warfe & Barmuta, 2004; Alexander et 424 425 al. 2012). Nevertheless, in our study, observations on the reduced foraging performance of the R. rutilus in the substrate treatment indicated that the reduced consumption rate was 426 largely due to the increased difficulty of the fish being able to detect the pellets once they had 427 428 settled on the gravel, as the size of pellets allowed a proportion to settle into relatively inaccessible areas (i.e. they provided a 'prey' refuge). 429

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431 The outcomes of our study highlighted the respective roles of prey item visibility and environmental conditions in determining the foraging behaviours and parameters of a visually 432 foraging fish (Utne-Palm 1999; Sweka & Hartman, 2003). In natural environments, these 433 dynamic relations are important considerations in habitat selection and optimal foraging, 434 given that foragers will always seek to maximise their energy intake whilst minimising 435 energetic costs and risk of predation (Chick & McIvor, 1997). As such, the potential role of 436 factors other than prey item density that can influence feeding rate under more complex, 437 natural conditions should be understood within the context of the relationships described in 438 our study. For example, predation risk, competition and satiation may all affect the rate of 439 feeding (Mills 1982; Persson 1983; Werner et al. 1983; Brabrand & Faafeng, 1993; Henson 440 & Hallam, 1995; Elliot 2003; Vahl et al. 2005; Priyadarshana et al. 2006) and, potentially, 441

exert a greater influence on feeding rate than that of prey density. Furthermore, given the 442 need to observe feeding behaviour within this study, the 24 hour starvation period used to 443 ensure a high feeding motivation may then be higher than that of fish in the wild, that are less 444 likely to undergo starvation in the same manner (Simpson et al. 1996; Privadarshana et al. 445 2006). As such, it can be assumed that it is the short-term functional response of roach that 446 we describe here, rather than the longer term, daily functional response when time is also 447 allocated by individuals to non-feeding activities, and where foraging behaviours must also 448 consider trade-offs with predation risk and competition as described above. In order to more 449 450 accurately account for the role of feeding motivation in forager behaviours, some roach foraging models have included state-dependent variables that explicitly include the hunger-451 state of the foragers (Holker et al. 2002; Holker & Breckling 2005), where hunger-state can 452 453 dictate behaviours such as habitat selection in association with both resting and feeding activities. Nevertheless, despite these confounding factors, the basic relationships we reveal 454 here between environmental conditions and foraging rate remain valid, even when influenced 455 456 by further, complex, factors and trade-offs. The results described here will be especially useful in more simple systems, for example in individual-based models (Holker et al. 2002; 457 Railsback & Harvey 2002). 458

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Frequent changes in the environmental conditions of lowland riverine habitats (where the presence of *R. rutilus* tends to be ubiquitous across their range) are common, in response to prevailing weather conditions or more general shifts in lowland river management. This latter aspect is important given that many lowland river management techniques that are aligned to flood management works substantially modify fish habitats through, for example, removal of instream vegetation that tends to increase turbidity, decrease refugia and increase flow rates (e.g. Allouche & Gaudin, 2001; Copp 1997; Gregg & Rose, 1985; Grenouillet & Pont, 2001). This suggests that works such as these are likely to affect *R. rutilus* foraging performance, with adult roach switching to less productive filter feeding behaviour in response to environmental challenges. Such changes include increased flow rates and water turbidity (Van Den Berg et al. 2004; Bogacka-kapusta & Kapusta, 2007; Nurimen et al. 2010) in association with habitat refugia (Garner 1996), and so, the impact of river management on habitat conditions may potentially have substantial implications for fishery and fish population management.

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In conclusion, the investigation revealed that foraging parameters and functional responses of *R. rutilus* are modified by changing conditions, with increased complexity tending to decrease aspects of their foraging performance. Increased turbidity, substrate presence and larger food items significantly reduced their consumption rate. In combination, these outcomes suggest that the foraging performance of this species is context-specific; being subject to prevailing conditions and food item availability, and this requires consideration in all relevant applications of their foraging behaviour.

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- 681

Table 1 Outputs of the linear regression and AIC values, testing the fit of each predicted
model against the observed functional response. Model selection was based on the AIC
scores with tests performed separately for each factor.

Turbidity	Food Item Size	Model	df	R ²	F	P	AIC
Turbid	1mm	Type II	66	0.93	60.76	< 0.01	-30.26
Turbid	2mm	Type II	62	0.95	43.65	<0.01	-18.68
Turbid	1mm	Type I	66	0.91	65.94	<0.01	-20.65
Turbid	2mm	Type I	62	0.91	46.59	<0.01	-13.89
Turbid	1mm	Linear	66	0.92	38.3	<0.01	-5.65
Turbid	2mm	Linear	62	0.91	26.1	<0.01	-4.45
Clear	1mm	Type II	59	0.93	34.71	< 0.01	-8.23
Clear	2mm	Type II	69	0.97	66.01	<0.01	-3.17
Clear	1mm	Type I	59	0.90	38.36	< 0.01	11.47
Clear	2mm	Type I	69	0.92	72.6	< 0.01	13.84
Clear	1mm	Linear	59	0.91	49.2	< 0.01	15.29
Clear	2mm	Linear	69	0.95	26.1	< 0.01	16.38

Table 2 Outputs of the linear mixed effects models testing the effect of food item size, substrate presence and increased turbidity on consumption rate and foraging parameters. Fixed effects listed by the significance of their effect on each dependent variable. F =consumption rate; d = Reaction Distance; h = Handling Time and s = Swimming Speed.

		1st			2nd			3rd		
Dependent:	df	Factor:	F	Р	Factor:	F	Р	Factor:	F	Р
		Food								
F	223	Item Size	8.36	< 0.01	Substrate	0.13	0.73	Turbidity	2.75	0.87
					Food					
d	50	Substrate	12.32	0.01	Item Size	0.53	0.81	Turbidity	-	-
					Food					
h	526	Substrate	5.20	0.02	Item Size	2.16	0.11	Turbidity	0.01	0.97
		Food								
S	1976	Item Size	2.55	0.07	Substrate	0.28	0.86	Turbidity	1.73	0.18

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Fig.1 Comparison of observed functional responses for the no-substrate (filled squares) versus substrate treatments (clear circles), where the lines represent the modelled Type II functional response from Holling's Disc Equation parameterised using observed foraging parameters under the no-substrate (solid line), and the substrate treatments (dashed line)

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Fig.2 Comparison of observed functional responses for clear water (filled squares) and turbidity treatments (clear circles) using (a) 1mm pellets and (b) 2mm pellets. The lines represents the modelled Type II functional response from Holling's Disc Equation parameterised using observed foraging parameters, under clear (solid line) and turbid treatments (dashed line)

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Fig.3 Comparison of observed functional responses for differences in food item size, where filled squares represent 1 mm pellets and clear circles 2 mm pellets and under (a) clear conditions and (b) turbid conditions. The lines represents the modelled Type II functional response from Holling's Disc Equation parameterised using observed foraging parameters, under 1mm food item size (solid line) and 2mm food item size treatments (dashed line)