



Title	Calcium concentration in leaf litter alters the community composition of soil invertebrates in warm-temperate forests
Author(s)	Ohta, Tamihisa; Niwa, Shigeru; Agetsuma, Naoki; Hiura, Tsutom
Citation	Pedobiologia, 57(4-6), 257-262 <a href="https://doi.org/10.1016/j.pedobi.2014.07.003">https://doi.org/10.1016/j.pedobi.2014.07.003</a>
Issue Date	2014-11
Doc URL	<a href="http://hdl.handle.net/2115/58011">http://hdl.handle.net/2115/58011</a>
Type	article (author version)
File Information	Pedobiologia_rev3.pdf



[Instructions for use](#)

1 Title: Calcium concentration in leaf litter alters the community composition of soil

2 invertebrates in warm-temperate forests.

3

4 Authors: Tamihisa Ohta<sup>1</sup>, Shigeru Niwa<sup>2</sup>, Naoki Agetsuma<sup>3</sup> and Tsutom Hiura<sup>1</sup>

5

6 *Affiliations:*

7 <sup>1</sup>Tomakomai Research Station, Field Science Center for Northern Biosphere, Hokkaido

8 University, Takaoka, Tomakomai, Hokkaido 053-0035, Japan

9 <sup>2</sup>Network Center of Forest and Grassland Survey in Monitoring Sites 1000 Project,

10 Japan Wildlife Research Center, Takaoka, Tomakomai, Hokkaido 053-0035, Japan

11 <sup>3</sup>Wakayama Experimental Forest, Field Science Center for Northern Biosphere,

12 Hokkaido University, Hirai, Kozagawa-cho Higashimuro-gun, Wakayama, 649-4563,

13 Japan

14

15 Running title: Calcium in leaf litter and soil organisms

16

17 \*Corresponding Author: Email: tammyohta@gmail.com Phone: +81-0144-33-2171,

18 Fax: +81-0144-33-2173

19 Email:

20 S. Niwa: sniwa@fsc.hokudai.ac.jp

21 N. Agetsuma: agetsuma@fsc.hokudai.ac.jp

22 T. Hiura: hiura@fsc.hokudai.ac.jp

23

24

25

26

27

28

29

30

31

32

33 ABSTRACT

34

35 Many studies have shown the effects of aboveground plant species on soil organisms  
36 due to differences in litter quality. However, the calcium concentration in soil has  
37 received less attention as a controlling factor of soil invertebrate communities, despite  
38 its being an essential element for many animals, especially crustaceans. Litter of  
39 Japanese cedar (*Cryptomeria japonica*) plantations, which account for 19% of the  
40 forested area in Japan, has a higher calcium concentration compared to other taxa such  
41 as broad-leaved trees. We predicted that *C. japonica* plantations affect soil invertebrates  
42 by altering calcium availability. We compared soil properties including exchangeable  
43 calcium concentration and soil invertebrate communities between *C. japonica*  
44 plantations and natural broad-leaved forests. Exchangeable calcium was significantly  
45 higher in soil from cedar plantations than in that from broad-leaved forests. The  
46 invertebrate community composition differed between the two forest types and was best  
47 explained by the exchangeable calcium concentration. In particular, two major taxa of  
48 soil crustaceans (*Talitridae* and *Ligidium japonicum*) were found only in cedar

49 plantations. Our results suggest that calcium concentrations in soil are altered in *C.*

50 *japonica* plantations and that this affects soil invertebrate communities.

51

52 *Keywords:* Evergreen broad-leaved trees, *Cryptomeria japonica*, forest management,

53 crustaceans, *Ligidium japonicum*, Talitridae

54

55

56

57

58

59

60

61

62

63

64

65 **Introduction**

66

67 Soil organisms can be affected by differences in aboveground vegetation (Bardgett  
68 and Wardle 2010), often driven by the quality of the litter types (Swift et al. 1979; Berg  
69 and McLaugherty 2003). Differences in litter quality among plant species can influence  
70 the chemical properties of soil and act as determinants of the community structure of  
71 soil invertebrates (Widden and Hsu 1987; Wardle et al. 2006). Because calcium is a  
72 major structural component of the proteins forming animals, ambient calcium  
73 concentration related to the animal densities in calcium poor environment (Alstad et al.  
74 1999; Hessen et al. 2000; Ohta et al. 2014). Similarly, soil calcium concentrations can  
75 have an important influence on soil invertebrate communities (Springett and Syers  
76 1984). For example, the abundance of some soil invertebrates increase with the  
77 available soil calcium concentration (Hotopp 2002; Reich et al. 2005; Skeldon et al.  
78 2007). Global pattern of soil calcium concentration is governed by geological change,  
79 acid deposition and annual precipitation (Potter and Conkling 2012; Binkley and Fisher  
80 2013). For example, low rainfall areas (central North America) tend to have grater soil

81 calcium than humid Eastern United States (Potter and Conkling 2012). Calcium  
82 availability and cycling in regional scale are governed by numerous factors including  
83 forest vegetation dynamics, atmospheric deposition, soil mineral weathering, and losses  
84 due to leaching (Likens et al. 1998; McLaughlin and Wimmer 1999; Dijkstra and Smits  
85 2002). However, much research over the last half century has focused on the leaching of  
86 calcium due to acid deposition (Likens et al. 1996; Driscoll et al. 2001), with much less  
87 emphasis on other factors, such as changes in forest vegetation, even though the calcium  
88 concentration in leaf litter varies greatly among tree species.

89 Litter of members of the Cupressaceae has a higher concentration of calcium  
90 compared with other plant families (Kiilsgaard et al. 1987; Ohta et al. 2014). In this  
91 study, we focused on Japanese cedar (*Cryptomeria japonica*, Cupressaceae) because its  
92 leaf litter contains ~3% calcium (Xue and Luo 2002; Baba et al. 2004), more than three  
93 times the amount in many other taxa, such as fir (*Abies* spp.) and many broad-leaved  
94 trees (Kiilsgaard et al. 1987; Reich et al. 2005; Ohta et al. 2014). Japanese cedar  
95 plantations cover 12% of the total land area and 19% of the forested area in Japan  
96 (Forestry Agency 2011). Because soil organic matter in forests is derived mainly from

97 plant litter, the chemical properties of litter affect soil chemical properties (Reich et al.  
98 2005). Indeed, the soil in Japanese cedar plantations has a calcium content that is three  
99 to four times higher than that in evergreen broad-leaved forests in some parts of Japan  
100 (Tsutsumi 1987; Ohta et al. 2014). Ohta et al. (2014) showed that the calcium  
101 concentration in soil and streams, and the density and survival of dominant aquatic  
102 crustaceans, were significantly higher in *C. japonica* plantations compared with  
103 evergreen broad-leaved forests. However, Ohta et al. (2014) did not assess the effects of  
104 forest vegetation on soil animal community through alteration of calcium availability.  
105 Soil crustaceans that are frequently dominant decomposers in soil systems (O'hanlon  
106 and Bolger 1999), and contain large amount of calcium in their body (Greenaway 1985).  
107 Terrestrial crustaceans mainly take calcium from their food (e.g. leaf litter) and soil  
108 water. Therefore, we anticipate that the calcium concentration in litter affects the  
109 community structure of soil organisms in calcium-poor environment.

110 Addition of inorganic calcium often increases soil pH (Likens et al. 1996; Driscoll  
111 et al. 2001; Warby et al. 2009), and therefore, higher calcium concentrations in soil due  
112 to differences in forest vegetation are also likely to increase soil pH (Reich et al. 2005).



113 Alteration of soil pH also causes changes in the abundance of soil invertebrates (Hågvar  
114 1990; Myrold 1990; Kaneko and Kofuji 2000). Therefore, plantations of *C. japonica*  
115 may affect the community structure of soil invertebrates via increased soil pH.

116 We examined the effect of Japanese cedar (*C. japonica*) plantations on the  
117 community structure of soil invertebrates, particularly the density of crustaceans. We  
118 conducted field surveys in six plots that differed in surrounding forest vegetation. We  
119 predicted (1) that the calcium concentration and soil pH would be higher in *C. japonica*  
120 plantations compared to evergreen broad-leaved forests, and (2) that crustacean density  
121 would be higher in *C. japonica* plantations than in evergreen broad-leaved forests.

122

## 123 **Methods**

124

### 125 *Study area*

126

127 We conducted field surveys in the Wakayama Experimental Forest of Hokkaido  
128 University (33°40'N, 135°40'E; 428 ha; annual mean temperature: 15.2°C) on the

129 Southern Kii Peninsula of Japan. The geological structure in this region consists of  
130 sandstone and mudstone formed during the middle Tertiary (Tateishi 1976). Because of  
131 the highly acidic soil and high annual rainfall (~4000 mm), the area is extremely poor in  
132 calcium (Kihira et al. 2005). The forest soils are extremely thin, nearly exposing the  
133 bedrock. Japanese cedar was planted in much of the area beginning in the 1960s, and  
134 remnant natural evergreen broad-leaved forests are patchy.

135 We established a sampling plot (50 × 50 m) in each of six different catchments of  
136 the Wakayama Experimental Forest. The plots were located on relatively flat forest  
137 floors and separated by 0.2–1.5 km. Three of the six catchments were mostly covered  
138 by evergreen broad-leaved forests ‘evergreen’, and the other three were covered by  
139 Japanese cedar plantations ‘cedar’. Forests in the ‘evergreen’ plots were dominated by  
140 *Quercus acuta*, *Quercus myrsinifolia*, *Quercus sessilifolia*, *Neolitsea aciculata*, *Eurya*  
141 *japonica*, and *Machilus thunbergii* (Ohta et al. 2014). The *C. japonica* trees in the  
142 ‘cedar’ plots were planted 30–82 years prior to this study. Calcium concentration in the  
143 litter of *C. japonica* (3.4%) is about three times higher than in the evergreen  
144 broad-leaved species (0.8–1.5%) at this study site (Ohta et al. 2014). Carbon, nitrogen,

145 phosphorus, and magnesium concentrations do not differ significantly among the  
146 species, whereas potassium is about three times lower in *C. japonica* compared to the  
147 broad-leaved species (Ohta et al. 2014).

148

#### 149 *Sampling*

150

151 On 24 July 2012, we collected five samples at each plot from the litter and soil layers  
152 using core samplers (soil layer: 50 mm in diameter and 50 mm in height, litter layer:  
153 113 mm in diameter and 40 mm in height) to measure soil chemical properties and mass  
154 of the litter layer. To determine the soil crustacean density, we established five sampling  
155 quadrats separated by over 10 m in each plot. We collected crustaceans within the  
156 sampling quadrats (25 × 25-cm) to a depth of 3 cm (including litter and surface soil  
157 layers) on 17 May and 21 September 2013. Soil crustaceans were separated from soil by  
158 hand-sorting and placed in 99% ethanol. To examine the community structure of  
159 ground-dwelling macroinvertebrates, we established two subplots (20 × 20 m) in each  
160 plot. We collected ground-dwelling invertebrates using five pitfall traps (8 cm in

161 diameter and 6 cm in depth) per subplot. We placed the pitfall traps in each subplot ~2  
162 m apart on 17 May, 19 July, 19 October, and 18 November 2013, and collected them 3  
163 days later. We counted and identified all invertebrates found in the traps at least to the  
164 ordinal level following Aoki (1999) and Ueno et al. (1985).

165

#### 166 *Sample processing*

167

168 To measure soil nitrate and exchangeable calcium, we shook ( $160 \text{ rev min}^{-1}$ ) a 0.5 g  
169 (air-dried mass) subsample of each soil sample in 100 ml of 1 M KCl solution for 1 h,  
170 filtered the sample through filter paper (No. 5C; Advantec, Tokyo, Japan), and then  
171 stored the suspension at  $-30^{\circ}\text{C}$  until analysis. We analyzed the soil extracts for calcium  
172 and nitrate concentration per unit air-dried mass using an inductively coupled plasma  
173 (ICP) atomic emission spectrometer (ICPE-9000; Shimadzu, Kyoto, Japan) and the  
174 absorptiometric method (Sakata 2000). We placed a 5-g (air-dried mass) subsample of  
175 each soil sample in 25 ml 1 M KCL and measured the pH using a pH meter (TOA-DKK,  
176 HM-30V; TOA Electronics, Tokyo, Japan). We dried soil subsamples in a drying oven

177 at 60°C for 24 h and then analyzed 50-mg dried soil samples for carbon and nitrogen  
178 concentrations per dry mass using a CN analyzer (Sumigraph NC-900; Sumika  
179 Chemical Analysis Service, Osaka, Japan). We dried a 1-g fresh subsample at 60°C for  
180 48 h to calculate the soil water content as the difference in mass before and after  
181 desiccation.

182

183 *Statistical analysis*

184

185 The soil properties (mass of the litter layer, water content, pH, C:N ratio,  
186 exchangeable calcium, total carbon, total nitrogen, and nitrate concentration) were fit to  
187 linear mixed models with forest vegetation type as a fixed factor and plot identity as a  
188 random factor. The statistical significance of the effect of the fixed factor in each model  
189 was evaluated by a likelihood ratio test ( $\alpha = 0.05$ ).

190 We performed canonical correspondence analysis (CCA) to explore the  
191 relationships between the soil invertebrate composition and soil properties. The  
192 invertebrate data from the five pitfall traps on all four sampling dates were pooled for

193 each subplot. Before conducting the CCA ordination, the abundance data for each taxon  
194 were standardized to unit variance, and the most important explanatory variables from  
195 all soil properties were determined by forward stepwise selection based on Akaike's  
196 information criteria and Monte Carlo permutation tests. All statistical analyses were  
197 conducted with R version 2.9.2 software (R Development Core Team 2011).

198

## 199 **Results**

200

201 Soil exchangeable calcium concentration (likelihood ratio test:  $\chi^2 = 9.13$ , d.f. = 1,  $P$   
202 = 0.002) and C:N ratio (likelihood ratio test:  $\chi^2 = 8.90$ , d.f. = 1,  $P = 0.003$ ) were  
203 significantly higher, and nitrate concentration (likelihood ratio test:  $\chi^2 = 8.53$ , d.f. = 1,  $P$   
204 = 0.003) was significantly lower in 'cedar' than 'evergreen'. In particular, soil in 'cedar'  
205 plots had ~2.5 times more exchangeable calcium than the 'evergreen' plots (Table 1,  
206 Fig. 1). The other soil properties including the mass of litter layer that was selected  
207 explanatory variable in community composition of soil invertebrates did not differ  
208 significantly between the forest types although soil pH was marginally higher in 'cedar'

209 than ‘evergreen’ (likelihood ratio test:  $\chi^2 = 3.05$ , d.f. = 1,  $P = 0.08$ ). We found two taxa  
210 of crustaceans, Talitridae (Amphipoda) and *Ligidium japonicum* (Isopoda: Ligiidae) in  
211 ‘cedar’ plots by hand-sorting, but we found no crustaceans in ‘evergreen’ plots (Fig. 2).  
212 Abundances of Talitridae and *L. japonicum* did not differ significantly among ‘cedar’  
213 plots between the two sampling months (generalized linear models assuming Poisson  
214 distribution and likelihood ratio tests:  $P > 0.05$ ).

215 Crustaceans, spiders, ants, beetles, lepidopteran larvae, and millipedes accounted for  
216 23, 19, 9, 19, 10, and 9%, respectively, of the total invertebrates captured by pitfall traps  
217 (Table 2). Crustaceans (mainly Talitridae and *L. japonicum*) dominated the  
218 ground-dwelling invertebrate communities in ‘cedar’ plots, while they were remarkably  
219 scarce, and millipedes (Paradoxosomatidae) and beetles were relatively abundant in  
220 ‘evergreen’ plots. Talitridae and *L. japonicum* were only collected in ‘cedar’ plots using  
221 pitfall traps, and a few individuals of *Venezillo* sp. (Isopoda: Armadillidae) were  
222 collected not only in ‘cedar’ but also ‘evergreen’ plots (Table 2). Taxonomic  
223 compositions of ground invertebrates differed distinctively between ‘cedar’ and  
224 ‘evergreen’ plots (Table 2, Fig. 3). From all the soil properties, exchangeable calcium

225 concentration and mass of the litter layer explained the most variation among the 12  
226 subplots in invertebrate community composition as determined through the forward  
227 selection process of CCA. The first and second axes explained 17.17 and 14.04%,  
228 respectively, of the variation in community composition (Monte Carlo permutation test:  
229  $P < 0.05$ ). The CCA ordination showed that community composition in the ‘cedar’ plots  
230 was distinctively different from that in the ‘evergreen’ plots along the first CCA axis,  
231 which corresponded to the gradient of exchangeable calcium concentration, and the  
232 ‘cedar’ plots had much lower variation than the ‘evergreen’ plots (Fig. 3).  
233 Compositional difference within ‘evergreen’ was distinctively indicated by second axis,  
234 which weakly correlated with mass of litter layer (Fig. 3). Talitridae and *L. japonicum*  
235 had large negative values on the first axis, and their high abundance characterized the  
236 ‘cedar’ community.

237

## 238 **Discussion**

239

240 Our results show that forest vegetation might affect the community structure of soil



241 invertebrates by altering calcium availability. Our field survey showed that soil  
242 exchangeable calcium was ~2.5 times higher and soil pH was not significantly higher in  
243 *C. japonica* plantations compared with that in evergreen broad-leaved forests (these  
244 partly supported prediction 1). The major taxa of soil crustaceans (Talitridae and *L.*  
245 *japonicum*) were found only in *C. japonica* plantations, whereas only a few individuals  
246 of a minor crustacean species (*Venezillo* sp.) occurred in broad-leaved forests  
247 (supporting prediction 2). The community structure of soil invertebrates varied with  
248 forest vegetation types, and calcium in the soil layer was the most important  
249 environmental variable explaining the variation in community composition (supporting  
250 prediction 2). Furthermore, both intra- and inter-plot variation in the soil invertebrate  
251 community structure in *C. japonica* plantations were lower than in natural broad-leaved  
252 forests, suggesting that the homogeneous environment created by the monoculture  
253 plantation caused a large decrease in micro- and local-scale  $\beta$ -diversity of soil  
254 invertebrates. This might be caused by uniform increase in the abundance of Talitridae  
255 and *L. japonicum* in *C. japonica* plantations and difference in the mass of litter layer  
256 within 'evergreen' plots (Table 1, Fig. 3).

257       The difference in calcium concentration in leaf litter produces significant  
258 differences in soil calcium (Morrison 1985; Kloeppel and Abrams 1995). In fact, Ohta  
259 et al. (2014) showed that the calcium concentration in *C. japonica* litter was about three  
260 times higher than that of dominant evergreen broad-leaved trees in Wakayama  
261 Experimental Forest. We found that total calcium in the litter layer and  
262 water-extractable calcium in the soil at ‘cedar’ sites were three to four times higher than  
263 at ‘evergreen’ sites. These results support our finding that calcium supplied by *C.*  
264 *japonica* litter increased the calcium concentration in the soil of our study plots.

265       Field manipulations at the Hubbard Brook Experimental Forest in the northeastern  
266 United States indicated that adding  $\text{CaSiO}_3$  to a catchment area increased calcium  
267 concentrations in soil (Juice et al. 2006; Minocha et al. 2010; Nezat et al. 2010) and  
268 altered the community structure of terrestrial snails, which have high demand for  
269 calcium (Skeldon et al. 2007). Hotopp (2002) showed that the abundance of sugar  
270 maple, a calciphilic species (Likens and Bormann 1970), was positively correlated with  
271 terrestrial snail density. As in snails, crustaceans must ingest a lot of calcium  
272 (Greenaway 1985). Because terrestrial crustaceans lose 20% of their body calcium

273 through exuviae (Ziegler et al. 2007), they need adequate calcium to calcify their  
274 exoskeleton rapidly after exuviation. Terrestrial crustaceans in soil get calcium from  
275 soil water and litter (Greenaway 1985; Glötzner and Ziegler 2000). Therefore, the  
276 ‘cedar’ plots, which have high calcium levels in their litter and soil layers, are well  
277 suited for the survival of crustaceans (Figs. 1, 2). Indeed, both natural and artificial *C.*  
278 *japonica* forests in central Japan have high densities of crustaceans (Ikeda et al. 2005).

279 Calcium addition to acidic soil increases soil pH (Groffman et al. 2006; Groffman  
280 and Fink 2011). However, our result showed the soil pH was not significantly higher in  
281 *C. japonica* plantations that increase the calcium concentration of the soils (Table 1).

282 Although previous studies showed soil pH may also affect the structure of invertebrate  
283 communities by altering forest vegetation (e.g. Kaneko and Kofuji 2000), forward  
284 selection of the CCA analysis selected exchangeable soil calcium concentration and not  
285 pH in our result. Calcium concentration may have a greater impact on the community  
286 structure of soil invertebrates in our study site.

287 The litter of other members of Cupressaceae, such as *Chamaecyparis* and  
288 *Sequoiadendron*, also have high calcium content comparable to that of *C. japonica*

289 (Kiilsgaard et al. 1987; D'Amore et al. 2009). This might mean change in soil system  
290 through difference in calcium concentration in leaf litter is caused in other forest types.  
291 Furthermore, Reich et al. (2005) showed that exchangeable calcium in soils and the  
292 density of earthworms were significantly higher in forests dominated by *Acer*  
293 *pseudoplatanus* and *Tilia cordata* compared to forests dominated by *Larix* and *Pinus*,  
294 even though calcium concentrations were ~40% lower in the litter of *A. pseudoplatanus*  
295 and *T. cordata* than in *C. japonica*. This implies that the change in soil systems is  
296 caused by the alteration of forest vegetation with low calcium content relative to *C.*  
297 *japonica*.

298 Soil crustaceans, such as Talitridae and *L. japonicum*, are powerful litter  
299 decomposers (Zimmer 2002). Their existence influences litter decomposition rates  
300 (O'Hanlon and Bolger 1999), and their abundance can lead to increased turnover rates  
301 of organic matter. Furthermore, soil crustaceans affect the dynamics of organic matter  
302 by incorporating organic material from the forest floor into deeper soil horizons  
303 (Mattson 2012) and may also enhance microbial biomass (Escher et al. 2000). Reich et  
304 al. (2005) showed that plantations of tree species with high calcium concentrations in

305 their litter cause an increase in exchangeable soil calcium and earthworm densities.  
306 Therefore, especially in calcium-poor environments, transformation of forest vegetation  
307 might change the densities of key decomposers such as crustaceans and earthworms,  
308 thereby affecting the decomposition rates of soil organic matter and nutrient dynamics.

309

### 310 Acknowledgments

311 We thank staff members of Wakayama Experimental Forest, Hokkaido University, for  
312 their support during the study. This study was partly supported by Grant-in-Aid from  
313 JSPS (12J07244 to TO and 2566011103, 25281053 to TH) and from the Ministry of  
314 Environment (S-9-3 to TH).

315

316

317

318

319

320

321 References

322 Alstad, N.E.W., Skardal, L., Hessen, D.O., 1999. The effect of calcium concentration on  
323 the calcification of *Daphnia magna*. *Limnol. Oceanogr.* 44, 2011-2017.

324 Aoki, J., 1990. *Soil Animals of Japan: Manual with Keys and Illustrations* (in Japanese).  
325 Tokai University Press, Tokyo.

326 Baba, M., Kato, M., Sugiura, T., Kobayashi, H., 2004. Calcium accumulation alleviates  
327 soil acidification in Japanese cedar (*Cryptomeria japonica*) stands. *Soil Sci.*  
328 *Plant Nutr.* 50, 403-411.

329 Bardgett, R.D., Wardle, D.A., 2010. *Aboveground-Belowground Linkages. Biotic*  
330 *Interactions, Ecosystem Processes and Global Change.* Oxford university  
331 press, Oxford.

332 Berg, B., McClaugherty, C., 2003. *Plant litter: Decomposition, Humus Formation and*  
333 *Carbon Sequestration.* Springer, Berlin.

334 Binkley, D. Fisher R.F., 2013. *Ecology and Management of Forest Soils* 4<sup>th</sup> ed.  
335 Wiley-Blackwell, New York.

336 D'Amore, D.V., Hennon, P.E., Schaberg, P.G., Hawley, G.J., 2009. *Adaptation to*

337 exploit nitrate in surface soils predisposes yellow-cedar to climate-induced  
338 decline while enhancing the survival of western redcedar: A new hypothesis.  
339 Forest Ecol. Manag. 258, 2261-2268.

340 Dijkstra, F.A., Smits, M.M., 2002. Tree species effects on calcium cycling: The role of  
341 calcium uptake in deep soils. Ecosystems 5, 385-398.

342 Driscoll, C.T., Lawrence, G.B., Bulger, A.J., Butler, T.J., Cronan, C.S., Eagar, C.,  
343 Lambert, K.F., Likens, G.E., Stoddard, J.L., Weathers, K.C., 2001. Acidic  
344 deposition in the northeastern United States: Sources and inputs, ecosystem  
345 effects, and management strategies. Bioscience 51, 180-198.

346 Escher, N., Käch, B., Nentwig, W., 2000. Decomposition of transgenic *Bacillus*  
347 *thuringiensis* maize by microorganisms and woodlice *Porcellio scaber*  
348 (Crustacea: Isopoda). Basic Appl. Ecol. 1, 161-169.

349 Forestry Agency, 2011. 2011 Forestry Census. Forestry Agency, Tokyo.

350 Glötzner, J., Ziegler, A., 2000. Morphometric analysis of the calcium-transporting  
351 sternal epithelial cells of the terrestrial isopods *Ligia oceanica*, *Ligidium*  
352 *hypnorum*, and *Porcellio scaber* during molt. Arthropod Struct. Dev. 29,





368 Ikeda, H., Homma, K., Kubota, K., 2005. Biotic and abiotic factors affecting the  
369 structures of ground invertebrate communities in Japanese cedar dominant  
370 forests. *Eur. J. For. Res.* 8, 1-13.

371 Juice, S.M., Fahey, T.J., Siccama, T.G., Driscoll, C.T., Denny, E.G., Eagar, C., Cleavitt,  
372 N.L., Minocha, R., Richardson, A.D., 2006. Response of sugar maple to  
373 calcium addition to Northern Hardwood Forest. *Ecology* 87, 1267-1280.

374 Kaneko, N., Kofuji, R., 2000. Effects of soil pH gradient caused by stemflow  
375 acidification on soil microarthropod community structure in a Japanese red  
376 cedar plantation: An evaluation of ecological risk on decomposition. *J. For.*  
377 *Res.* 5, 157-162.

378 Kihira, T., Yoshida, S., Hironishi, M., Miwa, H., Okamoto, K., Kondo, T., 2005.  
379 Changes in the incidence of amyotrophic lateral sclerosis in Wakayama,  
380 Japan. *Amyotroph. Lateral. Sc.* 6, 155-163.

381 Kiilgaard, C.W., Greene, S.E., Stafford, S.G., 1987. Nutrient concentrations in litterfall  
382 from some western conifers with special reference to calcium. *Plant Soil* 102,  
383 223-227.

- 384 Kloeppe, B.D., Abrams, M.D., 1995. Ecophysiological attributes of the native *Acer*  
385 *saccharum* and the exotic *Acer platanoides* in urban oak forests in  
386 Pennsylvania, USA. *Tree Physiol.* 15, 739-746.
- 387 Likens, G.E., Bormann, F.H., 1970. Chemical analyses of plant tissue from the Hubbard  
388 Brook Ecosystem in New Hampshire. *Yale Univ. Sch. For. Bull.* 79, 1–25.
- 389 Likens, G.E., Driscoll, C.T., Buso, D.C., 1996. Long-term effects of acid rain: Response  
390 and recovery of a forest ecosystem. *Science* 272, 244-246.
- 391 Likens, G.E., Driscoll, C.T., Buso, D.C., Siccama, T.G., Johnson, C.E., Lovett, G.M.,  
392 Fahey, T.J., Reiners, W.A., Ryan, D.F., Martin, C.W., Bailey, S.W., 1998.  
393 The biogeochemistry of calcium at Hubbard Brook. *Biogeochemistry* 41,  
394 89-173.
- 395 Mattson, W.J., 2012. *The Role of Arthropods in Forest Ecosystems (Proceedings in Life*  
396 *Sciences)*. Springer, Berlin.
- 397 McLaughlin, S.B., Wimmer, R., 1999. Tansley Review No. 104 - Calcium physiology  
398 and terrestrial ecosystem processes. *New Phytol.* 142, 373-417.
- 399 Minocha, R., Long, S., Thangavel, P., Minocha, S.C., Eagar, C., Driscoll, C.T., 2010.

400 Elevation dependent sensitivity of northern hardwoods to Ca addition at  
401 Hubbard Brook Experimental Forest, NH, USA. *Forest Ecol. Manag.* 260,  
402 2115-2124.

403 Morrison, I.K., 1985. Effect of crown position on foliar concentrations of 11 elements  
404 in *Acer Saccharum* and *Betula Alleghaniensis* trees on a till soil. *Can. J.*  
405 *Forest Res.* 15, 179-183.

406 Myrold, D.D., 1990. Effect of Acidic Deposition on Soil Organisms. In *Mechanisms of*  
407 *Forest Response to Acidic Deposition*. Lucir A.A. and Haines S.G. (eds.),  
408 Springer, New York, pp. 163-187.

409 Nezat, C.A., Blum, J.D., Driscoll, C.T., 2010. Patterns of Ca/Sr and Sr-87/Sr-86  
410 variation before and after a whole watershed CaSiO<sub>3</sub> addition at the Hubbard  
411 Brook Experimental Forest, USA. *Geochim. Cosmochim. Ac* 74, 3129-3142.

412 O'hanlon, R.P., Bolger, T., 1999. The importance of *Arcitalitrus dorrieni* (Hunt)  
413 (Crustacea: Amphipoda: Talitridae) in coniferous litter breakdown. *Appl. Soil*  
414 *Ecol.* 11, 29-33.

415 Ohta, T., Niwa, S., Hiura, T., 2014. Calcium concentration in leaf litter affects the

416 abundance and survival of crustaceans in streams draining warm-temperate  
417 forests. *Freshw. Biol.* 59, 748-760.

418 Potter, K.M., Conkling, B. L., 2012. Forest health monitoring: 2008 national technical  
419 report. Gen. Tech. Rep. SRS-158. Asheville, NC: U.S. Department of  
420 Agriculture Forest Service, Southern Research Station. 179 p.

421 R Development Core Team, 2011. R: A Language and Environment for Statistical  
422 Computing. R Foundation for Statistical Computing, Vienna, Austria.

423 Reich, P.B., Oleksyn, J., Modrzyński, J., Mrozinski, P., Hobbie, S.E., Eissenstat, D.M.,  
424 Chorover, J., Chadwick, O.A., Hale, C.M., Tjoelker, M.G., 2005. Linking  
425 litter calcium, earthworms and soil properties: a common garden test with 14  
426 tree species. *Ecol. Lett.* 8, 811-818.

427 Sakata, T., 2000. The assay of nitrate nitrogen in the soil extract using the UV  
428 absorbance (in Japanese). *Jpn. J. For. Environment.* 42, 53-55.

429 Skeldon, M.A., Vadeboncoeur, M.A., Hamburg, S.P., Slum, J.D., 2007. Terrestrial  
430 gastropod responses to ecosystem-level calcium manipulation a northern  
431 hardwood forest. *Can J. Zool.* 85, 994-1007.

- 432 Springett, J.A., Syers, J.K., 1984. Effect of pH and calcium content of soil on  
433 earthworm cast production in the laboratory. *Soil Biol. Biochem.* 16,  
434 185-189.
- 435 Swift, M.J., Heal, O.W., Anderson, J.M., 1979. *Decomposition in Terrestrial*  
436 *Ecosystems*. Backwell, Oxford.
- 437 Tateishi, M., 1976. The Muro Group in the southwestern part of the Muro Belt of the  
438 Shimanto Terrain (in Japanese). *J. Geol. Soc. Jpn.* 82, 395–407.
- 439 Tsutsumi, T., 1987. *Material Cycling in Forest* (in Japanese). University of Tokyo Press,  
440 Tokyo.
- 441 Ueno, S., Kurosawa, Y., Sato, M., 1985. *The Coleoptera of Japan in color, vol. II* (in  
442 Japanese). Hoikusha Press, Osaka.
- 443 Warby, R.A.F., Johnson, C.E., Driscoll, C.T., 2009. Continuing acidification of organic  
444 soils across the north-eastern USA: 1984–2001. *Soil Sci. Soc. Am. J.* 73,  
445 274–284.
- 446 Wardle, D.A., Yeates, G.W., Barker, G.M., Bonner, K.I., 2006. The influence of plant  
447 litter diversity on decomposer abundance and diversity. *Soil. Biol. Biochem.*

448                   38, 1052-1062.

449   Widden, P., Hsu, D., 1987. Competition between Trichoderma Species - Effects of  
450                   Temperature and Litter Type. Soil Biol. Biochem. 19, 89-93.

451   Xue, L., Luo, S., 2002. Seasonal changes in the nutrient concentrations of leaves and  
452                   leaf litter in a young Cryptomeria japonica stand. Scand. J. Forest Res. 17,  
453                   495-500.

454   Ziegler, A., Hagedorn, M., Ahearn, G.A., Carefoot, T.H., 2007. Calcium translocations  
455                   during the moulting cycle of the semiterrestrial isopod *Ligia hawaiiensis*  
456                   (Oniscidea, Crustacea). J. Comp. Physiol. B 177, 99-108.

457   Zimmer, M., 2002. Nutrition in terrestrial isopods (Isopoda: Oniscidea): an  
458                   evolutionary-ecological approach. Biological Reviews 77, 455-493.

459

460

461

462

463

464 Table 1 Soil properties (mean  $\pm$  SE) at each plot. Significant differences between cedar  
 465 plantations (cedar) and evergreen broad-leaved forests (evergreen) are denoted in the  
 466 last column: \*\* $P < 0.01$ , n.s.  $P > 0.05$  (likelihood ratio tests).

Plot name	cedar1	cedar2	cedar3	evergreen1	evergreen2	evergreen3	Significant differences between 'cedar' and 'evergreen' plots
Mass of litter layer (g)	10.42 $\pm$ 0.527	11.04 $\pm$ 0.561	10.28 $\pm$ 1.560	11.70 $\pm$ 1.240	7.47 $\pm$ 0.569	7.80 $\pm$ 0.845	n.s.
Water content (g/g)	0.62 $\pm$ 0.025	0.63 $\pm$ 0.064	0.57 $\pm$ 0.057	0.53 $\pm$ 0.045	0.64 $\pm$ 0.023	0.60 $\pm$ 0.036	n.s.
pH	4.34 $\pm$ 0.166	4.81 $\pm$ 0.143	5.20 $\pm$ 0.217	4.26 $\pm$ 0.182	4.55 $\pm$ 0.226	3.58 $\pm$ 0.085	n.s. ( $P = 0.08$ )
Exchangeable calcium (mg/g)	1.81 $\pm$ 0.256	1.83 $\pm$ 0.349	2.21 $\pm$ 0.971	0.63 $\pm$ 0.183	1.06 $\pm$ 0.413	0.43 $\pm$ 0.112	**
Total carbon (mg/g)	154.03 $\pm$ 23.188	117.63 $\pm$ 33.999	174.65 $\pm$ 58.206	139.58 $\pm$ 28.232	257.24 $\pm$ 47.917	201.00 $\pm$ 30.483	n.s.
Total nitrogen (mg/g)	10.33 $\pm$ 1.823	6.76 $\pm$ 1.770	9.90 $\pm$ 3.079	9.90 $\pm$ 1.908	19.0 $\pm$ 3.641	14.15 $\pm$ 1.880	n.s.
Nitrate (mg/g)	0.39 $\pm$ 0.021	0.52 $\pm$ 0.059	0.58 $\pm$ 0.047	0.73 $\pm$ 0.099	1.01 $\pm$ 0.113	0.82 $\pm$ 0.081	**
C:N ratio	15.89 $\pm$ 1.990	17.00 $\pm$ 0.781	17.36 $\pm$ 1.140	14.27 $\pm$ 0.775	13.52 $\pm$ 0.638	14.13 $\pm$ 0.473	**

485 Table 2 Abundance (5 traps<sup>-1</sup> 12 days<sup>-1</sup>; mean ± SE) of ground-dwelling  
 486 macroinvertebrates sampled by pitfall traps. ‘cedar’ and ‘evergreen’ mean cedar  
 487 plantation and evergreen broad-leaved forest.

	cedar	evergreen
<b>Gastropoda</b>	0.67 ± 0.67	0.00 ± 0.00
<b>Oligochaeta</b>		
Haplotaxida	0.33 ± 0.33	0.00 ± 0.00
<b>Diplopoda</b>		
Polydesmida		
Paradoxosomatidae	0.00 ± 0.00	3.33 ± 1.76
Xystodesmidae		
<i>Xystodesmus</i> sp.	0.33 ± 0.33	0.00 ± 0.00
Julida		
Julidae		
<i>Anaulaciulus</i> sp.	0.00 ± 0.00	0.33 ± 0.33
<b>Arachnida</b>		
Araneae	4.00 ± 0.58	4.00 ± 1.00
<b>Crustacea</b>		
Amphipoda		
Talitridae	4.33 ± 0.88	0.00 ± 0.00
Isopoda		
Ligiidae		
<i>Ligidium japonicum</i>	4.67 ± 0.88	0.00 ± 0.00
Armadillidae		
<i>Venezillo</i> sp.	0.67 ± 0.33	0.33 ± 0.33
<b>Insecta</b>		
Archaeognatha		
Machilidae	0.33 ± 0.33	0.33 ± 0.33
Orthoptera		
Rhaphidophoridae	0.00 ± 0.00	0.67 ± 0.33
Lepidoptera (larva)	2.33 ± 0.88	2.00 ± 0.00
Hymenoptera		
Formicidae		
<i>Pachycondyla chinensis</i>	0.67 ± 0.67	0.00 ± 0.00
<i>Aphaenogaster famelica</i>	0.33 ± 0.33	1.00 ± 1.00
<i>Paratrechina flavipes</i>	0.33 ± 0.33	0.67 ± 0.67
<i>Formica hayashi</i>	1.00 ± 0.58	0.33 ± 0.33
<i>Camponotus obscuripes</i>	0.00 ± 0.00	0.67 ± 0.67
Coleoptera (larva)	0.00 ± 0.00	0.67 ± 0.67
Coleoptera (adult)		
Geotrupidae		
<i>Phelotrupes laevistriatus</i>	0.33 ± 0.33	0.33 ± 0.33
<i>Phelotrupes auratus</i>	0.33 ± 0.33	0.00 ± 0.00
Scarabaeidae		
<i>Panelus parvulus</i>	0.00 ± 0.00	0.33 ± 0.33
<i>Onthophagus nitidus</i>	0.33 ± 0.33	0.67 ± 0.67
Staphylinidae		
<i>Bolitobius</i> sp.	0.33 ± 0.33	0.33 ± 0.33
Staphylininae	0.00 ± 0.00	0.33 ± 0.33
Carabidae		
<i>Carabus iwawakianus</i>	0.00 ± 0.00	0.67 ± 0.67
<i>Chlaenius costiger</i>	0.67 ± 0.67	0.00 ± 0.00
<i>Stomis prognathus</i>	0.00 ± 0.00	0.33 ± 0.33
<i>Pterostichus (Rhagadus)</i> sp.	0.33 ± 0.33	1.33 ± 0.88
<i>Rupa japonica</i>	1.00 ± 0.58	0.33 ± 0.33
<i>Synuchus picicolor</i>	0.00 ± 0.00	0.33 ± 0.33



489 Fig. 1 Concentration (mean  $\pm$  SE) of exchangeable calcium in soil in each plot. White  
490 and black bars indicate evergreen broad-leaved forests (evergreen) and cedar plantations  
491 (cedar), respectively. Significant differences between vegetation types are denoted by  
492 different letters (likelihood ratio tests,  $P < 0.05$ ).

493

494 Fig. 2 Abundance (mean  $\pm$  SE) of Talitridae (a) and *Ligidium japonicum* (b) in each plot  
495 by the hand-sorting method. White and black bars indicate evergreen broad-leaved  
496 forests (evergreen) and cedar plantations (cedar), respectively.

497

498 Fig. 3 Canonical correspondence analysis (CCA) ordination of soil invertebrate  
499 community composition in 12 subplots by the pitfall trap method. Explanatory variables  
500 selected by forward selection are shown as arrows: Ca, exchangeable calcium  
501 concentration and Litter, mass of the litter layer. White and black symbols indicate  
502 subplot scores (mean  $\pm$  SE) of evergreen broad-leaved forests (evergreen) and cedar  
503 plantations (cedar), respectively. Invertebrate taxa are abbreviated by alphabets: G,  
504 Gastropoda; H, Haplotaxida; P, Paradoxosomatidae; Xs, *Xystodesmus* sp.; As,  
505 *Anaulaciulus* sp.; A, Araneae; T, Talitridae; Lj, *Ligidium japonicum*; Vs *Venezillo* sp.;  
506 L, Lepidoptera (larva); Pc, *Pachycondyla chinensis*; Af, *Aphaenogaster famelica*; Pf,  
507 *Paratrechina flavipes*; Fh, *Formica hayashi*; Co, *Camponotus obscuripes*; C,  
508 Coleoptera (larva); Pp, *Panelus parvulus*; On, *Onthophagus nitidus*; Pl, *Phelotrupes*  
509 *laevistriatus*; Pa, *Phelotrupes auratus*; Bs, *Bolitobius* sp.; St, Staphylininae; Ci,  
510 *Carabus iwawakianus*; Cc, *Chlaenius costiger*; Sp, *Stomis prognathous*; Ps,

511 *Pterostichus (Rhagadus) sp.*; Rj, *Rupa japonica*; Syp, *Synuchus picicolor*; M,

512 Machilidae; R, Rhabdophoridae. Taxa indicated by boldface are crustaceans.

513

514

515

516

517

518

519

520

521

522

523

524

525

526

527

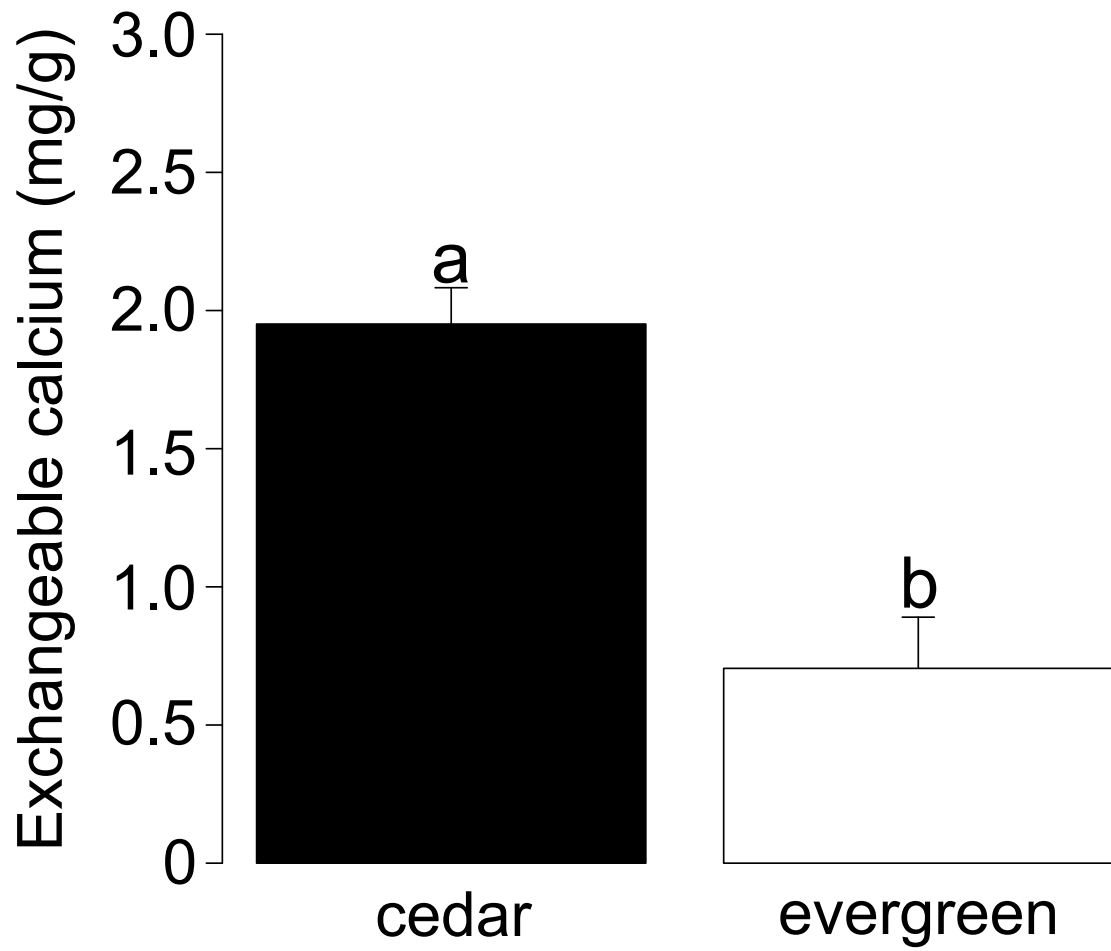
528

529

530

531

532



534

535

536

537

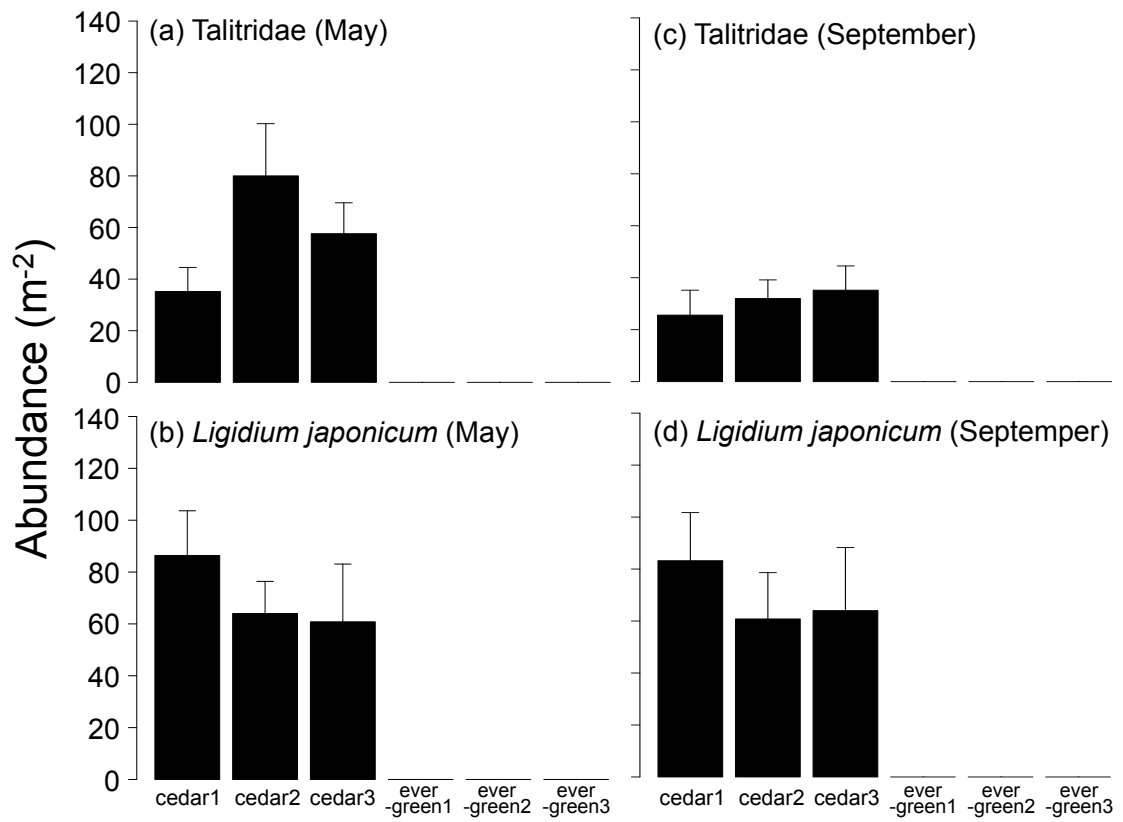
538

539

540

541

542



544

545

546

547

548

549

550

551

552

553

554

