

VARIABILITY, SPECIATION AND PHYTOREMEDIATION OF SOIL ARSENIC AT CATTLE DIP SITES IN NSW, AUSTRALIA

NABEEL KHAN NIAZI M.Sc. (Hons) Agri

A THESIS SUBMITTED TO THE FULFILMENT OF THE REQUIREMENTS FOR THE DEGREE OF DOCTOR OF PHILOSOPHY

> Faculty of Agriculture, Food and Natural Resources The University of Sydney, Sydney New South Wales Australia



CERTIICATE OF ORIGINALITY

I hereby declare that the text of this thesis is my own work, and that, to the best of my knowledge and belief, it contains no material that has been previously published or written by another person, nor any material that has been accepted as part of the requirements for any other degree or diploma in any university or other institute of higher learning, unless due acknowledgement has been made.

I also declare that the intellectual content of this thesis is original and the result of my own research and to the best of my knowledge and belief, any assistance I received in the experimentation presented, and all sources of information cited have been duly acknowledged.

TO MY PARENTS AND FAMILY

ABSTRACT

Arsenic (As) contamination of soils is a major environmental problem due to its toxic and carcinogenic nature. Historical use of As-containing pesticides has resulted in the contamination of soils with high and variable concentrations of As in many parts of Australia. Phytoremediation using As-hyperaccumulating ferns can be potentially utilised as an environmental friendly and low-cost remediation technology to phytoextract As from soils at sites containing elevated and varying concentration of As.

The spatial variability of total and phosphate-extractable As concentrations was evaluated in soil adjacent to a cattle-dip site located at Wollongbar in northern NSW, Australia. The results from the linear mixed model showed that total (0–0.2 m) and phosphate-extractable (0–0.2, 0.2–0.4 and 0.4–0.6 m depths) As concentrations in the soil adjacent to the dip site varied greatly and increased significantly (P = 0.004–0.048) toward the dip site, indicating that As variability in soil was spatially correlated with distance from the dip. The data suggest that 5 samples should be required to assess the soil contamination level (mean = 826 mg kg⁻¹) and 15 samples would be required to evaluate the effects of phytoremediation of As-contaminated site. The proposed guidelines on sampling requirements are important to estimate the variability in As contamination levels around other cattle-dip sites and to monitor changes in soil As content from phytoremediation activities.

Ensuing study compared the phytoremediation potential of *Pityrogramma calomelanos* var. *austroamericana* (gold dust fern) against the well-known *Pteris vittata* (Chinese brake fern) over a 27-month duration grown at the cattle-dip site described earlier. The ferns were planted in January 2009 and harvested following 10, 22 and 27 months of growth. After 10 months of growth (short-term data), P. calomelanos var. austroamericana produced significantly higher frond dry biomass, possessed higher frond As concentration and removed more As in fronds (mean = 130 g plant⁻¹, 887 mg kg⁻¹ and 124 mg plant⁻¹, respectively; P < 0.05) than P. vittata $(\text{mean} = 81 \text{ g plant}^{-1}, 674 \text{ mg kg}^{-1}, 57 \text{ mg plant}^{-1})$. Further samplings up to 27 months (longterm data) confirmed the earlier results that the mean frond dry biomass, As concentration and As uptake were significantly higher in *P. calomelanos* var. *austroamericana* than *P. vittata*. In the three harvests over the 27-month period, P. calomelanos var. austroamericana removed (8,053 mg As) 2.65 times higher As than P. vittata (3,042 mg As). For the surface (0-20 cm) and subsurface (40-60 cm) layers, the mean total soil As content was significantly (P < 0.05) reduced by 49% and 63%, respectively, using P. calomelanos var. *austroamericana*; and 17% and 15%, respectively, by *P. vittata* (P > 0.05). It is estimated that P. calomelanos var. austroamericana would take approximately 6 years to decrease mean total As content below the ecological investigation level (EIL; 20 mg kg⁻¹) limit in the surface and subsurface soils, whereas P. vittata would require 13-15 years to achieve this target. The field experiment results suggest that P. calomelanos var. austroamericana is better suited than P. vittata for the phytoremediation of As-contaminated soils under the experimental conditions existing at the site.

The potential of mid infrared (MIR) spectroscopy in combination with partial least squares (PLS) regression was investigated to estimate the total As content in a large number of soil samples collected from a highly variable As-contaminated dip site. The MIR-PLS calibration model developed excluding spectral outliers (n = 149) was robust with an acceptable reliability (coefficient of determination; $R_c^2 = 0.73$; residual prediction deviation; RPD_c = 1.94) to estimate total soil As content. The validation of calibration model using a separate set of unknown soil samples (n = 149; validation set) showed R_v^2 and RPD_v values of 0.63 and

1.66, respectively. The results indicate an acceptable prediction of total As content in unknown samples, suggesting that MIR-PLS based model is capable of estimating total soil As and possibly be used in certain situations; for example to estimate soil As concentration at a highly variable site, where a large number of samples needs to be analysed.

The solid-phase speciation and plant availability of As in contaminated soils was determined using combination of a sequential extraction procedure (SEP), X-ray absorption near edge structure (XANES) spectroscopy and As plant uptake using *Brassica juncea* as a test plant. Arsenic was found to be predominantly associated with amorphous Fe oxides in arsenate (As^{V}) form; in few samples As was present in arsenite (As^{III}) form. The concentration of As in plant shoots showed significant (*P* < 0.001–0.05) correlations with the exchangeable As (r = 0.85), and amorphous Fe oxides associated As evaluated by the SEP (r = 0.67) and XANES spectroscopy (r = 0.51). The results suggest that As in these fractions is readily available for plant uptake and may pose potential risk to the environment. Such detailed analysis for As speciation and phytoavailability is vital for the management and rehabilitation of As-

* * *

ACKNOWLEDGEMENTS

I am thankful to the Higher Education Commission (HEC) of Pakistan for awarding me a PhD scholarship to start my postgraduate studies here at The University of Sydney in Australia. I am highly grateful to my supervisor, Associate Professor Balwant Singh who always travelled above and beyond his responsibility as a supervisor in all aspects of my candidature. He has been deeply involved and dedicated in my research endeavours at all times – from the very first day when I met him in Pakistan and until the end of my thesis submission. I thank you for your sincere and substantial efforts to initiate this project, guiding and encouraging me, and most importantly, for maintaining a high level of contact with me throughout the last years of my PhD. I must say that I have learnt it from you to offer high level of passion and devotion to science. I would like to sincerely thank Dr Lukas Van Zwieten (Associate Supervisor) for his substantial contribution in successfully continuing the phytoremediation experiment at the field site in Wollongbar, northern New South Wales (NSW). I am thankful to Lukas and his colleagues, Stephen Kimber, Joshua Rust, Scott Petty, Victor Warren, Desmond Cook and George Nastase for their help during the field work at the site. My heartily thanks to Dr Anthony G. Kachenko (Nursery and Garden Industry Australia) who introduced me to the intriguing world of ferns and offered his sincere contribution in my field experiment. I am grateful to Dr Thomas F. A. Bishop (The University of Sydney) who assisted me in learning geostatistics and provided me with useful suggestions for the work I did in my PhD research. My sincere thanks to Dr Pushan Shah (Environmental Protection Authority, Adelaide) who came with me (and Balwant) to use synchrotron at the Australian National Beamline Facility at Tsukuba, Japan. Both Balwant and Pushan introduced me to the fascinating world of synchrotron science and I feel honoured to have worked with such passionate scientists, and to have had the opportunity to conduct a beamline experiment and visit the Australian Synchrotron. My thanks to Dr Michael Cheah who assisted at the beamline for spectroscopy work. The bulk of this research would not have been possible without generous funding from several sources. I would like to acknowledge the financial support for this project by the NSW Government through its environmental trust. I wish to thank the Australian Synchrotron Research Program, for enabling me to travel to the Australian National Beamline Facility in Tsukuba (Japan) for performing my experiment (Project AS093/ANBF1851).

I am thankful to Dr Edith Lees for providing constructive comments on review chapter of this thesis. My thanks go to Dr Euan Smith (University of South Australia, Adelaide) for providing me three soil samples from As-contaminated soils in South Australia, and Dr Budiman Minasny (The University of Sydney) for assistance in MIR data analysis. Thanks are extended to the technical staff, Ms Loraine Watson and Ms Iona Georgy for their assistance. I wish to thanks Tom Savage for his invaluable assistance in ICP-AES analyses, and Michael Nelson and Peter Geelan-Small who provided much needed statistical advice. Many thanks to the administration staff, Ms Pamela Stern, Ms Prue Winkler, Ms Fortunée Cantrell and Ms. Nancy Cheng for providing me with overwhelming levels of logistical support throughout the course of my study. I am also thankful to all my friends at FAFNR and in my research group for their assistance whenever I needed.

I am thankful to Almighty God Who has blessed me with this much of strength, passion and courage to complete my studies. I am thankful to my wife, Irshad Niazi for her tremendous support who with her own PhD (and our little son, Aaish Khan) assisted and encouraged me in the last few months of my thesis writing. And finally, I would like to convey my deepest love and heartfelt thanks to my parents, brothers and sister.

PUBLICATIONS

Refereed scientific journals (Published/submitted)

- 1. Niazi NK, Singh B, Van Zwieten L, Kachenko AG (2011) Phytoremediation potential of *Pityrogramma calomelanos* var. *austroamericana* and *Pteris vittata* L. grown at a highly variable arsenic contaminated site. *International Journal of Phytoremediation* 13 (9):912–932.
- 2. Niazi NK, Singh B, Shah P (2011) Arsenic speciation and phytoavailability in contaminated soils using a sequential extraction procedure and XANES spectroscopy. *Environmental Science & Technology* 45 (17):7135–7142.
- 3. Niazi NK, Bishop TFA, Singh B, (2011) Evaluation of spatial variability of soil arsenic adjacent to a disused cattle-dip site, using model-based geostatistics. *Environmental Science & Technology* 45 (24):10463–10470.
- 4. Niazi NK, Singh B, Van Zwieten L, Kachenko AG (2011) Phytoremediation of an arsenic contaminated site using *Pteris vittata* L. and *Pityrogramma calomelanos* var. *austroamericana*: a long-term study. *Environmental Science and Pollution Research* (Submitted).

Refereed scientific journals (In preparation)

5. Niazi NK and Singh B (2011) Using mid infrared (MIR) spectroscopy and partial least-squares regression to predict total soil arsenic content at a highly variable site. *Environmental Science & Technology* (In preparation).

Conference paper (refereed)

6. Niazi NK, Singh B, Zwieten LV and Kachenko, AG (2010) Arsenic hyperaccummulation by ferns: a field study in northern NSW. *Published paper in the Proceedings of the 19th World Congress of Soil Science on Soil Solutions for the Changing World*. Brisbane, Australia, 1–6 August 2010, pp. 5–8.

Commentary (Article published in a refereed journal)

7. Niazi NK and Kachenko AG (2011) Letter to the Editor regarding, "First evidence on different transportation modes of arsenic and phosphorus in arsenic hyperaccumulator *Pteris vittata*" by Lei et al. (2012). *Environmental Pollution* (Article in press) doi: 10.1016/j.envpol.2011.11.031.

Conference proceedings (abstracts)

- 8. Niazi NK, Singh B, Shah, P (2010) Speciation and bioavailability of arsenic in contaminated soils using X-ray absorption near edge structure spectroscopy and a sequential extraction procedure. *Proceedings of ConSoil International Conference on the Management of Groundwater, Soil & Sediments*. Salzburg, Austria, 21–24 September, 2010 (Oral presentation).
- 9. Niazi NK, Bishop T, Singh B, (2010) Comparative study for the arsenic hyperaccumulation by ferns: a model-based geostatistical approach. *Proceedings of ConSoil International Conference on the Management of Groundwater, Soil & Sediments*. Salzburg, Austria, 21–24 September, 2010 (Poster presentation).
- 10. Niazi NK, Singh B, Zwieten LV and Kachenko, AG (2010) A comparative study to evaluate the phytoremediation potential of two fern species for the remediation of an arsenic contaminated site under field conditions. 7th International Conference on Phytotechnologies. Parma, Italy, 26–29 September, 2010 (Oral presentation).*
- 11. Niazi NK, Singh B, Bishop, T (2010) A geostatistical model based approach to evaluate spatial variability of arsenic in soil and to compare arsenic hyperaccumulation efficiency of two fern species. 7th International Conference on Phytotechnologies. Parma, Italy, 26–29 September, 2010 (Poster presentation).*
- 12. Niazi NK, Singh B, Zwieten LV and Kachenko, AG (2011) Phytoremediation potential of *Pityrogramma calomelanos* var. *austroamericana* and *Pterris vittata* L. grown at a highly variable arsenic contaminated site. *Proceedings of the International Annual Conference of ASA, CSSA and SSSA*. Fundamental for Life: Soil, Crop & Environmental Sciences. San Antonio, Texas, USA, 16–19 October, 2011 (Oral presentation).
- 13. Niazi NK, Bishop TFA and Singh B (2011) Using model-based geostatistics to develop design-based sampling guidelines for estimating arsenic contamination around cattle-dip sites. *Proceedings of Pedometrics- Innovations in Pedometrics.* Czech Republic, 31st August–2 September, 2011 (Poster presentation).

Poster presentations

- 14. Niazi NK, Singh B, Zwieten LV and Kachenko, AG (2010) Field evaluation of hyperaccumulating ferns to remediate an arsenic contaminated soil. *Nursery & Garden Industry National Conference*. Darwin, Australia, 19–22 April 2010.
- 15. Niazi NK, Singh B, Bishop T, Zwieten LV and Kachenko, AG (2010) Environmentally friendly approach to clean-up arsenic contaminated soils. *The 'Centenary Research Symposium'* of The Faculty of Agriculture, Food and Natural Resources, The University of Sydney, 4 June, 2010.

* The 3rd best prize was awarded for the poster and oral presentation at the 7th International Conference on Phytotechnologies, Parma, Italy, 26–29 September, 2010.

TABLE OF CONTENTS

ABSTRACT	iii
ACKNOWLEDGEMENTS	vi
PUBLICATIONS	viii
TABLE OF CONTENTS	Х
LIST OF FIGURES	xviii
LIST OF TABLES	xxii

<u>Ger</u>	neral introduction	1
1.1	Introduction	1
1.2	Aims	4
1.3	Objectives	4
1.4	References	5

<u>Rev</u>	riew of	literature	8	
2.1	Histo	ory of As and its position in the periodic table	8	
2.2	Distr	ibution of As in the soil environment	9	
2.3	Sour	ces	16	
	2.3.1	Natural sources	16	
	2.3.2	Anthropogenic sources	17	
		2.3.2.1 Mining and smelting activities	18	
		2.3.2.2 Industrialisation and urbanisation	18	
		2.3.2.3 Irrigation	19	
		2.3.2.4 Agriculture activities	19	
2.4	Speci	Speciation of As in soil		
	2.4.1	Sequential extraction procedure	22	
	2.4.2	X-ray absorption fine structure (XAFS) spectroscopy	23	
	2.4.3	Combination of SEP and XAFS spectroscopy	24	
2.5	Avai	lability of As in soil	25	
2.6	Arsenic accumulation in plants		30	
	2.6.1	Tolerance strategies in plants	31	
	2.6.2	Arsenic uptake mechanisms	34	
	2.6.3	Factors affecting plant As uptake	36	
2.7	Reme	ediation of As-contaminated soils	40	
	2.7.1	Phytoremediation	41	
2.8	Phyte	pextraction of As using hyperaccumulating fern species	44	
	2.8.1	Background on P. calomelanos var. austroamericana	48	
2.9	Refe	rences	50	

Evaluation of spatial variability of soil arsenic adjacent to a	
cattle-dip site	

Abstract			64
3.1	Introd	luction	65
3.2	Materials and methodology		67
	3.2.1	Study area and dataset	67
	3.2.2	Geostatistical analysis	69
	3.2.3	Sample size estimation	72
3.3	Results and discussion		74
	3.3.1	Spatial variability in soil As	74
	3.3.2	Estimation of sample size	82
3.4	Concl	usions	86
3.5	Refer	ences	88

64

		ediation potential of <i>Pteris vittata</i> L and <i>Pityrogramma</i> os var. <i>austroamericana</i> at a highly variable arsenic	
		ited site — short-term data	91
Abst	ract		91
4.1	Intro	duction	92
4.2	Mate	rials and methodology	94
	4.2.1	Fern species	94
	4.2.2	Experimental Site	94
	4.2.3	Experimental set up	94
	4.2.4	Soil and plant sampling	97
		4.2.4.1 Total and phosphate-extractable As in soil	97
	4.3	Statistical analysis	98
	4.4	Calculations to estimate the remediation time-frame based on	
		As uptake data	99
4.5	Resu	lts	101
	4.5.1	Soil As concentrations	101
		4.5.1.1 Total As concentration in the surface soil	101
		4.5.1.2 Phosphate-extractable soil As concentrations for three depth	hs101
	4.5.2	First plant sampling	103
	4.5.3	Major plant harvest	105
		4.5.3.1 Fern dry matter yield	105
		4.5.3.2 Fern As concentration	107
		4.5.3.3 Fern As uptake	107
		4.5.3.4 Fern bioconcentration factors (BFs)	110
4.6	Discu	ission	112
4.7	Time	estimation for remediation	116
4.8	Conc	lusions	118
4.9	9 References		119

Phy	torem	ediation of an arsenic contaminated site using <i>P. vit</i>	tata L.
and	P. cal	omelanos var. austroamericana — long-term data	123
Abst	ract		123
5.1	Intro	duction	124
5.2	Mate	rials and methodology	126
	5.2.1	Fern species and experimental site	126
	5.2.2	Ferns transplantation and experimental set up	126
	5.2.3	Frond harvests	126
	5.2.4	Soil sampling	127
5.3	Statis	stical analysis	127
5.4	Results and discussion		
	5.4.1	Frond harvests	129
		5.4.1.1 Dry matter yield	129
		5.4.1.2 Arsenic concentration in ferns	129
		5.4.1.3 Fern As uptake	131
	5.4.2	Total and phosphate-extractable soil As concentrations	
		for three depths	135
5.5	Time	estimation for remediation of the experimental site	138
5.6	Conc	lusions	139
5.7	Refe	rences	140

		red reflectance spectroscopy and partial least-squares	
<u>regr</u>	ession	to predict total As content in soil	143
Abst	ract		143
6.1	Intro	duction	144
6.2	Mate	rials and methodology	147
	6.2.1	Soil samples	147
	6.2.2	Soil digestion to measure total As	147
	6.2.3	MIR spectroscopy	148
	6.2.4	Principal component analysis and partial least-squares	148
6.3	Resu	Its and discussion	153
	6.3.1	MIR spectral features	153
	6.3.2	Principal component analysis	157
	6.3.3	Calibration model using MIR spectra and PLS regression	157
	6.3.4	Validation of MIR-PLS model	159
6.4	Conc	lusions	160
6.5	Refe	rences	162

_		and phytoavailability of arsenic in contaminated sc P and XANES spectroscopy	oils <u>166</u>
Abst	ract		166
7.1	Intro	duction	167
7.2	Mate	rials and methodology	169
	7.2.1	Soil sampling	169
	7.2.2	Physicochemical analysis	171
	7.2.3	Phytoavailability pot experiment	171
	7.2.4	Sequential extraction procedure	174
	7.2.5	XANES spectroscopy	175
	7.2.6	Data analysis	177
7.3	Results and discussion		178
	7.3.1	Total As in soil	179
	7.3.2	Arsenic fractionation in soils using a SEP	179
	7.3.3	Arsenic speciation using XANES spectroscopy	183
	7.3.4	Arsenic phytoavailability to <i>B. juncea</i> in relation to SEP	
		and XANES spectroscopy	190
7.4	Conc	lusions	194
7.5	Refer	rences	196

201
205

LIST OF FIGURES

CHAPTER 2

Figure 2.1	Global As cycle (adapted from Matschullat 2000).	10
Figure 2.2	A model for determining available As concentration in soil	27
	(modified from Morel 1997).	
Figure 2.3	A dose response curve for the essential and non-essential (As in this thesis)	
	elements in plants (modified after Morel 1997).	31
Figure 2.4	(a,c,e) Tolerance strategies in plants in relation to increasing heavy	
	metal(loid) (As in this study) concentrations in soil; (modified after Baker 1981);	
	(b,d,f) the plant growth in response to increasing heavy metal(loid) concentration	IS
	(in soil and plant), are also presented (Kachenko 2008).	32
Figure 2.5	Generalised pattern of As uptake strategies by terrestrial plants	
	(modified from Fitz and Wenzel 2006).	33

Figure 3.1	Schematic representation of the soil sampling scheme around a cattle dip at the experimental site. The filled circles show the intensive soil sampling scheme ($n = 102$) performed to evaluate the spatial variability of As in soil. The internodal distance between soil cores reduced to 0.25 m (as shown in figure) due to the	
	presence of a buffer region.	69
Figure 3.2	The frequency distribution and summary statistics of the untransformed data (mg kg ⁻¹) for (a) total As at 0–0.2 m depth, and phosphate-extractable As at (b) 0–0.2 m (c) 0.2–0.4 m, and (d) 0.4–0.6 m depths; (e) the log-transformed data for phosphate-extractable As 0.4–0.6 m depth (log, mg kg ⁻¹).	75
Figure 3.3	The semivariogram models fitted with spatial trend by REML for mapping of total and phosphate-extractable soil As at the studied depths, using the given parameters; (a) spherical model (untransformed data, mg kg ⁻¹); (b,c) exponential models (untransformed data, mg kg ⁻¹); (d) exponential model	
	(log-transformed data, log(mg kg ⁻¹)). The parameters are; C_1 = partial sill (varian φ = distance parameter (range); C_0 = nugget. The filled circles represent the experimental semivariogram obtained using method of moments.	.ce); 76
Figure 3.4	The maps showing the spatial variability of (a) total As at 0–0.2 m depth generated using a spherical model (untransformed data); and phosphate-extracta As for (b) 0–0.2 m and (c) 0.2–0.4 m depths using an exponential model (untransformed data); (d) 0.4–0.6 m depth using an exponential model (log-transformed data were back-transformed).	
Figure 3.5	The regression and reference lines showing the distribution of observed and predicted values for the total and phosphate-extractable As at the investigated depths, where the variogram models were fitted based on the spatial (northings and eastings) trend (in Figure 3.3); (a,b,c) untransformed data (mg kg ⁻¹ As) (d) log-transformed data (log, mg kg ⁻¹ As). The leave-one-out cross validation (LOOCV) indices are also presented, ME = mean error; MSE = mean square error MSDR = mean squared deviation ratio.	

Figure 3.6	The semivariogram models fitted without spatial trend by REML using the given parameters; (a,b) spherical models (untransformed data, mg kg ⁻¹); (c) exponential model (untransformed data, mg kg ⁻¹); (d) exponential model (log-transformed data log(mg kg ⁻¹)). The parameters are; β = estimated mean; C_1 = partial sill (variance);	ta,
	φ = distance parameter (range) and C_0 = nugget.	83
Figure 3.7	The regression and reference lines showing the distribution of observed and predi- values for the total and phosphate-extractable As at the investigated depths, where the variogram models were fitted without spatial trend (see in Figure 3.6); (a,b,c) untransformed data (mg kg ⁻¹ As), (d) log-transformed data (log, mg kg ⁻¹ As). The leave-one-out cross validation (LOOCV) indices are also presented; ME = mean ere MSE = mean square error; MSDR = mean squared deviation ratio.	e
Figure 3.8	The number of samples (n) required to estimate the spatial mean of total As (mg k	
0	in soil with a known error and confidence interval (CI) at 95% confidence level (CI	0
	around mean (untransformed data).	85

Figure 4.1	A schematic representation of the experimental site at Wollongbar alon bath. The area planted with <i>P. vittata</i> (a) and <i>P. calomelanos</i> var. <i>austroan</i>	•
	June 2009, two soil cores were taken around each fern and mixed togeth	
	composite soil sample, as indicated in this figure.	96
Figure 4.2	The relationship between total soil As concentration and the frond dry	matter yield of
	(a) P. calomelanos var. austroamericana and (b) P. vittata.	106
Figure 4.3	Arsenic concentration in P. calomelanos var. austroamericana fronds in rel	ation to the (a)
	total soil As at 0–20 cm depth and phosphate-extractable As at (b) 0–20	cm (c) 20–40
	cm, and $(d) 40-60$ cm depths.	108
Figure 4.4	Arsenic concentration in the fronds of <i>P. vittata</i> in relation to the (a) to	tal soil As at 0–
	20 cm depth and phosphate-extractable As at (b) 0–20 cm (c) 20–40 cm,	, and (d) 40–60
	cm depths.	108
Figure 4.5	The frond dry biomass of <i>P. calomelanos</i> var. austroamericana and <i>P. vitta</i>	ta in relation
	to the fronds As concentration (a and b) and As uptake (c and d), respec	ctively. The
	relationship between the frond As concentration and As uptake is also	presented (e)
	P. calomelanos var. austroamericana, and (f) P. vittata.	109
Figure 4.6	Spatial distribution of (a) As concentration, and (b) As uptake in P. vitta	<i>ita</i> (Plot A) and
	<i>P. calomelanos</i> var. <i>austroamericana</i> (Plot B) after 10 months of growth.	111
Figure 4.7	The initial amount of As contained in 0–60 cm of Plot A (P. vittata) and	Plot B (P.
	calomelanos var. austroamericana) of the experimental area (a) total As, (e) phosphate-
	extractable As. The maps showing the (b) total and, (e) phosphate-extra	ctable As after
	10 months of ferns growth. The remediation estimates made for the tota	ıl As over a
	period of (c) 10 years, (d) 20 years, and for phosphate-extractable As for	a period of (g)
	4 years, are also presented.	117

Figure 5.1The spatial variability maps of As showing remediation trend in the total soil As
concentration for three depths in Plot A (*P. vittata*) and Plot B (*P. calomelanos* var.
austroamericana) of the experimental area; (a,c,e) initial As concentration in soil
determined in June 2009; (b,d,f) final As concentration in soil measured in April 2011,
after 27 month of fern growth. The two experimental plots were separated by a buffer
strip as shown in the figure above. The data presented here correspond to the mean
total soil As concentration in Table 5.2.134

CHAPTER 6

Figure 6.1	The absorbance spectra of a soil sample in the MIR spectral region (4000–400 cm ⁻¹). The blue colour line represents the raw absorbance spectrum; red colour spectrum is
	first derivative of the raw spectrum after removing the baseline. 149
Figure 6.2	Principal component analysis (PCA) performed on the (first derivative pre-treated)
	absorbance spectra of all soil samples ($n = 304$). The green filled circles are the 6
	spectral outliers in the soil samples. PC-1 = principal component-1; PC-2 = principal
	component-2. 150
Figure 6.3	Frequency distribution (histograms and box-plots) of the measured total As
	concentration in soil samples used for the MIR-PLS regression analysis for the
	development of (a) final calibration model; and for (b) validation of the calibration
	model. 151
Figure 6.4	(a) The mid infrared (MIR) absorbance spectra of the six soil samples collected from
	the surface (0–20 cm) and the subsurface (20–40, 40–60 cm) depths; and (b) the MIR
	partial least-squares (PLS) coefficients obtained in the calibration model ($n = 149$), in
	relation to the wavenumber (4000–400 cm ⁻¹) used for the MIR spectroscopy. 154
Figure 6.5	The measured values of total soil As concentration are plotted against the MIR-PLS
	model predicted As values; (a) preliminary calibration model including the 6 spectral
	outliers ($n = 155$); (b) validation of the preliminary calibration model; (c) final
	calibration model developed excluding the 6 spectral outliers spectral data ($n = 149$)
	using the calibration set data as shown in Figure 6.3a; (d) validation of the calibration
	model constructed with $n = 149$ samples. The R_c^2 and R_v^2 are coefficient of
	determination in the calibration and validation sets, respectively; SECV = standard
	error of cross validation; SEP = standard error of prediction; RPD_c and RPD_v =
	residual prediction deviation in the calibration and validation sets, respectively. 158

Figure 7.1	Locations of the As-contaminated cattle-dip sites, railway corridor and As spiked samples collected from New South Wales (NSW) and South Australia (SA) states of the state of	
		170
Figure 7.2	Sequential extraction procedure applied to the cattle-dip sites, railway corridor an As spiked soil samples ($n = 18, 3$ and 11, respectively). Means of each extracted fraction are presented as percent (%) of the total soil As concentration present in the contaminated and As spiked soils. The residual fraction of As was calculated by subtracting sum of the four fractions from the total As concentration in each of the soils. The bars represent ±SD around the mean values.	he

Figure 7.3	Arsenic K-edge XANES spectra of (a) reference compounds, and (b) selected soil samples from the cattle-dip sites, railway corridor and As spiked soils. Solid (–), dotted (…) and dashed () lines drawn at 11875.7, 11872.2 and 11868.7 eV represent the position of white lines for arsenate (As ^V), arsenite (As ^{III}), and orpiment (As ₂ ^{III} S ₃), respectively. The white line energy values (eV) are given in the parentheses for each of the reference compound. 184
Figure 7.4	Arsenic K-edge XANES spectra for Boorie (black and purple lines) and Wollongbar (red and green lines) soil samples analysed after Step-II and Step-III of the SEP. The solid (–) line is drawn at 11875.7 eV represent the position of white line for arsenate (As ^V).
Figure 7.5	Arsenic concentration in dry shoots of <i>B. juncea</i> in relation to the sequentially extracted (a) exchangeable As (Step-I); (b) specifically sorbed As (Step-II); (c) amorphous Fe oxides bound As (Step-III); (d) crystalline Fe oxides bound As (Step-IV); and (e) As determined using XANES associated with amorphous Fe oxides. Cattle-dip sites and railway corridor soils (•); As spiked soils (•). The correlations are presented excluding the As spiked soils data due the considerable differences in ageing time. 191
Figure 7.6	The relationship of amorphous (ammonium oxalate extractable) Fe concentration with the amorphous Fe oxides associated As in the studied soils determined using (a) SEP in Step-III, and (b) XANES spectroscopy. 193

LIST OF TABLES

CHAPTER 2

Table 2.1	Review of As concentrations in different materials on earth; adapted from (Smith	ı et al.
	1998; Mandal and Suzuki 2002). 11	
Table 2.2	Arsenic minerals in the soil environment; adapted from (Mandal and Suzuki	
	2002).	14
Table 2.3	Arsenic concentrations in soil adjoining the cattle-dip sites in New South Wales	
	(NSW), Australia (Smith et al. 2006).	20
Table 2.4	Review of the confirmed As-hyperaccumulating species.	46
	(NSW), Australia (Smith et al. 2006).	

CHAPTER 3

Table 3.1	Fixed effect parameters for total and phosphate-extractable As in soil at the dif	ferent
	depths.	77

CHAPTER 4

Table 4.1	Physico-chemical properties of the soil from the experimental site at Wollongbar	in
	northern NSW.	95
Table 4.2	Total and phosphate-extractable soil As concentrations (mg kg-1) in Plot A and Pl	lot B
	(n = 42 per plot). The statistical differences between the two plots, for total and	
	phosphate-extractable As, are presented based on spatial ANOVA and <i>t-test</i> .	102
Table 4.3	The frond As concentration, dry matter yield and As uptake of <i>P. vittata</i> and <i>P.</i>	
	calomelanos var. austroamericana. The differences between the two fern species are	
	evaluated using ANCOVA and <i>t-test</i> .	104
Table 4.4	The bioconcentration factors of P. vittata and P. calomelanos var. austroamericana w	vere
	calculated based on the total and phosphate-extractable soil As concentrations. The	he
	differences between both the fern species are evaluated using <i>t-test</i> .	105

Table 5.1	Dry matter yield (g plant ⁻¹), As concentration (mg kg ⁻¹ DW) and As uptake (mg plant ⁻¹) measured in fronds of <i>Pityrogramma calomelanos</i> var. <i>austroamericana</i> and <i>vittata</i> after first (10 month in December 2009), second (22 month in November 2	Pteris
	and third (27 month in April 2011) harvests.	130
Table 5.2	The concentrations of total and phosphate-extractable As in soil in Plot A and Pl of the experimental area ($n = 42$ per plot for initial sampling; $n = 15$ per plot for f sampling (as mentioned below). Initial As concentrations in soil were determine before harvesting and final after three harvests of fern fronds.	inal

	cattle- 172
Detail for the extraction steps used in the sequential extraction procedure emplo	byed
for the As speciation of As-contaminated cattle-dip sites, railway corridor and A	As
spiked soils.	175
Arsenic distribution in various phases as determined by the SEP showing the A	s
distribution in cattle-dip sites, railway corridor and As spiked soil samples	
(<i>n</i> = 32). 180	
Linear least-squares combination fits of XANES spectra of reference compound	s with
those of soil samples ($n = 30$) analysed. Fit results correspond to XANES spectra	a of
selected soil samples plotted in Figure 7.2.	186
	Arsenic distribution in various phases as determined by the SEP showing the A distribution in cattle-dip sites, railway corridor and As spiked soil samples $(n = 32)$. 180 Linear least-squares combination fits of XANES spectra of reference compound those of soil samples $(n = 30)$ analysed. Fit results correspond to XANES spectra