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1 The revision of *HAZMAT-D-20-05548*

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3 Intra- and intersexual interactions shape microbial community dynamics in the  
4 rhizosphere of *Populus cathayana* females and males exposed to excess Zn

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23 **Abstract** Although increasing attention has been paid to rhizospheric bacteria in plants  
24 in relation to the bioavailability and tolerance of heavy metals, interactive effects  
25 between sex and microbiological processes on phytoextraction have been overlooked,  
26 especially in dioecious plants. In this study, we intended to investigate the responses of  
27 rhizospheric bacterial communities of *Populus cathayana* Rehder to excess Zn under  
28 different planting patterns. The results suggested that intersexual and intrasexual  
29 interactions strongly affect plant growth and Zn extraction in both sexes, as well as  
30 rhizosphere-associated bacterial community structures. Females had a higher capacity  
31 of Zn accumulation and translocation than males under all planting patterns. Males had  
32 lower Zn accumulation and translocation under intersexual than under intrasexual  
33 interaction; the contrary was true for females. Females harbored abundant  
34 *Streptomyces* and *Nocardioides* in their rhizosphere, similarly to males under  
35 intersexual interaction, but differed from single-sex males under excess Zn. Conversely,  
36 intersexual interaction increased the abundance of key taxa Actinomycetales and  
37 Betaproteobacteria in both sexes exposed to excess Zn. Males improved the female  
38 rhizospheric microenvironment by increasing the abundance of some key tolerance taxa  
39 of Chloroflexi, Proteobacteria and Actinobacteria in both sexes under excess Zn in  
40 intersexual interaction. This was associated with metal activation and bioavailability.  
41 Females harbored abundant *Methanothermobacter*, while males had abundant  
42 *Mycobacterium* in their rhizosphere under intrasexual interaction. These results  
43 indicated that the sex of neighboring plants affected sexual differences in the choice of  
44 specific bacterial colonizations for phytoextraction and tolerance to Zn-contaminated

45 soils, which might regulate the spatial segregation and phytoremediation potential of *P.*  
 46 *cathayana* females and males under heavy metal contaminated soils.

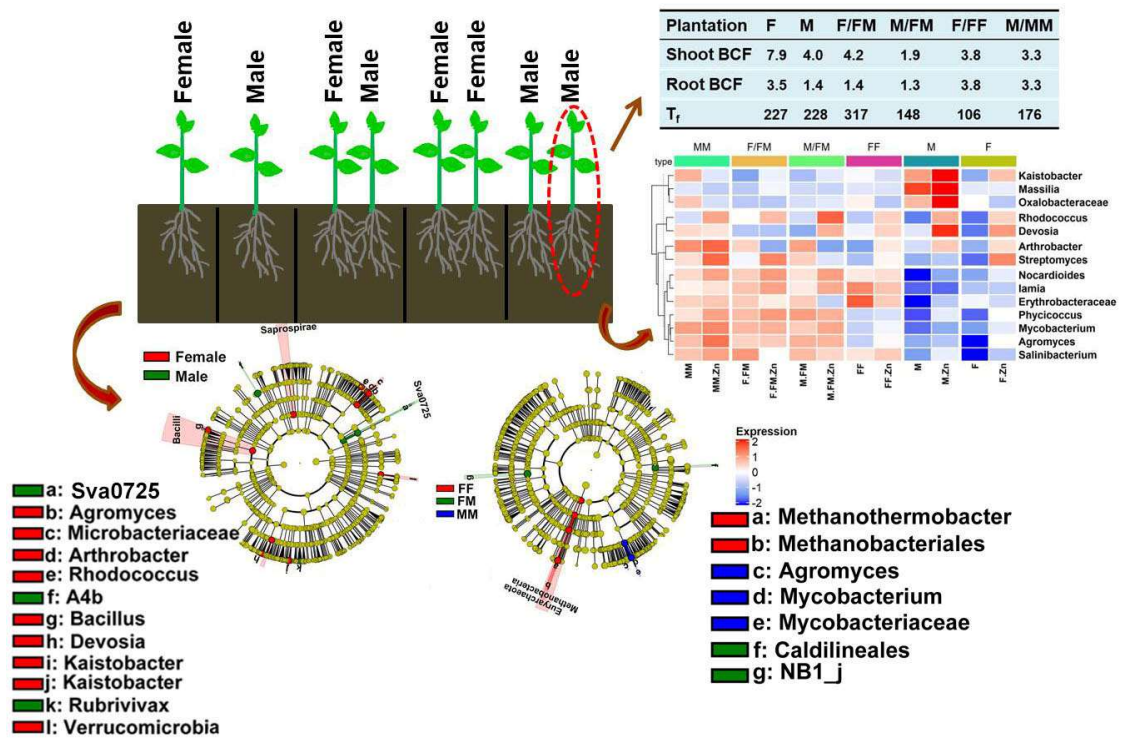
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48 **Key words:** sexual interaction; excess Zn; bacterial abundance; bacterial community;  
 49 rhizosphere.

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51 **Graphical abstract**

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55 **1. Introduction**

56

57 Males and females of dioecious plant species differ in their sexual and vegetative traits,  
58 including physiology, phenology, and allocation to reproduction and defense (Chen et  
59 al., 2014; Tonnabel et al., 2017; Xia et al., 2020). Sex-related differences in these traits  
60 are probably associated with differences in reproductive costs. Females allocate more  
61 resources to reproduction than males, since females produce not only flowers but also  
62 fruits and seeds (Juvany and Munné-Bosch, 2015). The maintenance of reproductive  
63 vigor usually limits vegetative growth and defense investment. Thus, when resources  
64 are limited, reproduction directly competes with the other two processes, leading to  
65 reduced stress tolerance (Mercer and Eppley, 2010; Tonnabel et al., 2017). It has been  
66 shown that females generally allocate more resources to reproduction and males often  
67 increase investment into defensive responses under natural conditions (Juvany and  
68 Munné-Bosch, 2015). Our previous studies have suggested that males usually exhibit a  
69 higher tolerance to stressful environments when compared with females (Zhao et al.,  
70 2012; Chen et al., 2014; Li et al., 2016; Zhang et al., 2019).

71

72 Zinc (Zn) is an essential micronutrient with a series of critical roles in living organisms;  
73 yet, Zn can be toxic at elevated concentrations (Broadley et al., 2007; Lu et al., 2013).  
74 It has been suggested that Zn toxicity symptoms usually appear when the leaf Zn  
75 content is  $> 30 \text{ mg} \cdot \text{kg}^{-1}$  of leaf dry weight, although the toxicity threshold may be highly  
76 variable, even in the same species (Ayangbenro and Babalola, 2017). Excess Zn affects

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77 plant growth and development, as well as threatens the health of animals due to  
78 excessive amounts of Zn ending up into the food chain (Duruibe et al., 2007; Carruthers,  
79 2016). The threat of excess Zn to the health of living organisms in the environment is  
80 exacerbated by its persistent nature (Ayangbenro and Babalola, 2017). Therefore, it is  
81 critical to remediate Zn-polluted soils. Several fast-growing trees, including poplars,  
82 are regarded as promising candidates for phytoremediation of heavy metal-polluted  
83 soils due to their high photoabsorption capacity and biomass, as well as reduced impact  
84 on the food chain (Chen et al., 2011; He et al., 2015; Li et al., 2017).

85

86 The migration and transformation of heavy metals in rhizospheric soils largely depend  
87 on the physical and chemical properties of soil (Rasmussen et al., 2000; Weng et al.,  
88 2014), plant types (Hou et al., 2017), rhizospheric microorganisms (Ayangbenro and  
89 Babalola, 2017) and exudates (Tao et al., 2020). Among these, rhizospheric microbial  
90 interactions with heavy metals play a critical role in heavy metal uptake and tolerance  
91 (Rajkumar et al., 2012). Microorganisms have developed variable mechanisms for  
92 maintaining heavy metal homeostasis and resistance, including biomineralization,  
93 bioaccumulation, biotransformation and biosorption (Navarro-Noya et al., 2010;  
94 Mishra et al., 2017). Heavy metals have been shown to suppress microorganisms'  
95 growth, and alter their cell morphology and biochemical characteristics, which reduces  
96 microbial biomass and diversity (Ayangbenro and Babalola, 2017). Still,  
97 microorganisms have evolved diverse defense mechanisms that help them cope with  
98 the toxic effect of heavy metals. Burd et al. (2000) have suggested that a metal-resistant

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99 bacterium *Kluyvera ascorbate* SUD 165 alleviates growth inhibition caused by a high  
100 level of nickel in *Brassica campestris*. Hou et al. (2017) have also reported that the  
101 rhizospheric bacterium *Streptomyces* probably promotes Cd accumulation in the  
102 hyperaccumulator plant *Sedum alfredii*. Therefore, the coexistence of plant and  
103 microbial systems not only facilitates survival in heavy-metal contaminated soils, but  
104 also the removal of heavy metal.

105

106 Interspecific and/or intraspecific neighbor interactions of plants play critical roles in  
107 determining the structure and function of biological communities (Chen et al., 2014).  
108 Intra- and interspecific competitive capacities have important implications for  
109 phytoremediation in metal-hyperaccumulating species (Arthur et al., 2005; Pilon-Smits  
110 and Freeman, 2006). Interactions have been shown to alter plants' responses to nickel  
111 and Zn (Koelbener et al., 2008). Zhao et al. (2017) have found that interspecific  
112 interactions can enhance antioxidant enzyme activities that increase survival and fitness  
113 in plants exposed to multiple metal stresses. Chen et al. (2016) have also shown that  
114 resource competition between consensual and heterosexual neighboring plants affect Cd  
115 allocation, biomass partitioning and carbon-nutrient balance in poplars exposed to Cd  
116 stress. Interactions not only alter plants' morphological and physiological responses,  
117 such as nutrient uptake, heavy metal availability and root exudate secretion, but also  
118 affect microbial communities in soil (Guo et al., 2019). So far, the effects of intersexual  
119 and intrasexual interactions on belowground microbial communities have been largely  
120 overlooked, especially under heavy metal stress.

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121

122 *Populus cathayana* Rehder is widely distributed in China and is regarded as a major  
123 forestry species with a high commercial and ecological value. *P. cathayana* is a  
124 common dioecious plant that displays sexual dimorphism (Chen et al., 2014; Zhang et  
125 al., 2019; Xia et al., 2020). Recently, some studies have suggested that poplar species  
126 have different tolerance mechanisms and phytoremediation potential to heavy metals,  
127 mainly based on genetic differences (Chen et al., 2017; Bi et al., 2020; Liu et al., 2020b).  
128 In addition, *Populus* females and males display sexually different physiological  
129 responses to heavy metal stress, and males usually show a stronger tolerance when  
130 compared to females (Chen et al., 2016, 2017; Liu et al., 2020a, b). Sex-specific  
131 responses to abiotic factors are affected by inter- and/or intraspecific interactions (Chen  
132 et al., 2016, 2017). However, sex-specific responses to excess Zn stress and neighbor  
133 effects have been largely overlooked in *P. cathayana*, especially in relation to  
134 rhizospheric microbial community structures. In the present study, *P. cathayana* was  
135 used to examine sexual differences in competitive ability under excess Zn conditions.  
136 We aimed to address the following questions: (i) Does excess Zn affect intra- and  
137 intersexual interactions? (ii) How do intra- and intersexual interactions affect  
138 rhizospheric bacterial communities? (iii) Do interactive effects between excess Zn and  
139 sexual interactions affect plant growth and rhizospheric bacterial community  
140 composition and diversity?

## 141 **2. Materials and methods**

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143 2.1. *Plant materials and experimental design*

144

145 Cutting of *P. cathayana* females and males were collected from 60 different trees  
146 sampled in 15 populations in the riparian and valley flat habitats of the Qinghai  
147 Province, China. Cuttings were rooted as described by Chen et al. (2016). The  
148 experimental design was completely randomized with three factors (sex, excess Zn and  
149 plantation type), i.e. two sexes (females, males), two Zn regimes (no-Zn, Zn) and five  
150 plantation types. After sprouting and growing for 4 weeks, uniform seedlings were  
151 planted into 60-L plastic pots filled with *c.* 40 kg of homogenized soil. The soil was  
152 collected from the poplar plantation at the Hangzhou Normal University, Zhejiang  
153 Province, China (30.03° N, 120.12° E). Soil samples were air-dried and sieved through  
154 2 mm sieve. The properties were as follows: 1.77 g kg<sup>-1</sup> total N, 1.80 g kg<sup>-1</sup> total P, 7.91  
155 g kg<sup>-1</sup> total K, and 0.1 g kg<sup>-1</sup> total Zn. For the excess Zn treatment, 1 L deionized water  
156 containing 100 μM ZnSO<sub>4</sub> was used to evenly irrigate the pots every two days until the  
157 final Zn level of 50 mg ZnSO<sub>4</sub> kg<sup>-1</sup> dry soil was reached, while the control plants were  
158 irrigated with equal quantities of deionized water (Chen et al., 2016).

159

160 The five plantation types were as follows: F, single-cultivated females; M, single-  
161 cultivated males; FF, female × female; MM, male × male; FM, female × male.  
162 Intrasexual neighboring plants were denoted as M/MM for males and F/FF for females.  
163 Intersexual neighboring plants were denoted as M/FM for males and F/FM for females.  
164 For single-plant cultivation, one cutting was planted per pot (a female or a male); for



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165 interactions, two plants (two females, two males or a female and a male) were cultivated  
166 10 cm apart from each other in a plastic pot (external diameter 52 cm and height 35  
167 cm). All pots were arranged randomly and each treatment was replicated four times.  
168 The experiment was performed at the Hangzhou Normal University. The plants were  
169 cultured for 120 d in a semi-controlled greenhouse at the Hangzhou Normal University  
170 (China, 30.03' N, 120.12'E) with a relative humidity of 76%-81%, a daytime  
171 temperature of 21-25 °C, a night-time temperature of 15-18 °C and a photoperiod of  
172 12-14 h throughout the growth period.

173

#### 174 *2.2. Gas exchange and fluorescence measurements*

175

176 The fully developed young leaves from each plant were used to measure gas exchange  
177 and chlorophyll fluorescence. Net CO<sub>2</sub> assimilation rate (*A*) and stomatal conductance  
178 (*g<sub>s</sub>*) were measured with a portable photosynthesis measuring system (LI-6400), as  
179 described previously by Chen et al. (2011). Chlorophyll fluorescence kinetics  
180 parameters (*ETR*, electron transport rate; quantum yield of photochemical energy  
181 conversion in *PS II*, *Y(II)*; quantum yield of regulated non-photochemical energy loss  
182 in *PS II*, *Y(NPQ)*;  $F_v/F_m$ , variable and maximum fluorescence) were measured with a  
183 PAM chlorophyll fluorometer (PAM 2100, Walz, Effeltrich, Germany). These  
184 parameters were calculated according to the method of Van Kooten and Snel (1990).

185

#### 186 *2.3. Soil sampling, plant harvesting and element measurements*

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187

188 After the measurement of photosynthesis, rhizosphere soil was collected by gently  
189 shaking a plant root; the soil adhering to the root was collected as a sample of  
190 rhizospheric soil. In total, 40 soil samples (4 replicates  $\times$  5 interaction patterns  $\times$  2  
191 treatments) were immediately sieved (4 mm) in the laboratory. The rhizospheric soils  
192 were divided into two subsamples. One subsample was oven dried at 75 °C and used  
193 for analyzing soil properties, and another subsample was stored at -80 °C until DNA  
194 extraction.

195

196 The plants were separated into leaves, stems and roots, and washed with deionized  
197 water. Dried leaves and roots were finely ground, and about 0.4 g samples ( $<$  1 mm)  
198 were dissolved in 3:1 (v/v) of HNO<sub>3</sub> and HClO<sub>4</sub>. The mixtures were carefully shaken  
199 and predigested at room temperature for 30 min. The vessels used for digestion were  
200 sealed and placed into a microwave digestion system. The microwave heating program  
201 was performed as follows: (1) 10 min at 170 °C; (2) 10 min at 190 °C; (3) 10 min at  
202 210 °C (Tokalioğlu et al., 2018). The vessels were cooled down to room temperature to  
203 avoid splashing and foaming. Clear digested solutions were transferred to 25 ml  
204 polyethylene tubes, which were filled to a volume of 25 ml with ultra-pure water for  
205 further ICP-MS analysis. Total Zn and nutrient elements were measured with ICP-MS  
206 (inductively coupled plasma mass spectrometer; Agilent 7500a; Agilent Technologies).  
207 For quality assurance/quality control (QA/QC) purposes, a blank control and standard  
208 reference material GBW10020 (GSB-11) of citrus leaves were used to validate  
209 quantification according to Xie et al. (2020). Blank control and reference materials were

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210 treated under the same conditions as the experimental samples. The recovery of the  
211 standard at a medium calibration level was checked with every 10 samples. The method  
212 detection limits ( $\text{mg kg}^{-1}$ ) were  $0.2 \text{ mg l}^{-1}$  for Zn and the recovery rates were 80-90%.  
213 All data are presented as means  $\pm$  the standard deviations, unless specified differently.

214

215 The translocation factor ( $T_f$ ) was defined as the Zn concentration in a shoot divided by  
216 the Zn concentration in roots (He et al., 2013b). The bio-concentration factor BCF was  
217 calculated as the Zn concentration in roots or shoots divided by the Zn concentration in  
218 the soil (Shi et al., 2010; He et al., 2013b).

219

#### 220 *2.4. Statistical analysis*

221

222 Statistical analyses were carried out using the SPSS software package (version 22.0).  
223 All data were checked for normality before analyses of variance (ANOVAs).  
224 Differences between means were analyzed by Duncan's tests following three-way  
225 ANOVAs, which were used to evaluate sexual interaction and excess Zn treatment  
226 effects.

227

#### 228 *2.5. DNA extraction, 16S amplification and MiSeq sequencing*

229

230 Approximately 0.5 g of each soil sample was used to extract genomic DNA utilizing  
231 the PowerSoil DNA Isolation Kit (MoBio Laboratories, Inc. Carlsbad, USA) following

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232 the manufacturer's instructions. The 16S rRNA genes were amplified using the primer  
233 pair 515F (5'-GTGCCAGCMGCCGCGGTAA-3') and 806R (5'-  
234 GGACTACHVGGGTWTCTAAT-3'). Approximately 10 ng of purified DNA was  
235 amplified with 25 µl of the PCR reaction system under the following conditions: at  
236 95 °C 2 min for denaturation, followed by 40 cycles of 10 s at 95 °C for denaturation,  
237 30 s at 56 °C for annealing, and 20 s at 72 °C for extension, with 55 to 95 °C for melting  
238 curve analyses. The PCR products were pooled and purified with a MicroElute Cycle-  
239 Pure Kit (Omega), and high-throughput sequencing was performed using the Hiseq  
240 2500 PE250 platform with 2×250 bp according to the manufacturer's manual.

241

#### 242 *2.6. Processing of high-throughput sequencing data and analysis*

243

244 Amplicon sequences were processed using the QIIME 2 version 2017.12  
245 (<https://qiime2.org/>). All sequences of 16S rRNA raw data were demultiplexed, and  
246 quality control was carried out using DADA2 with the “consensus” method, to remove  
247 chimeric and low-quality sequences (Callahan et al., 2017; Yuan et al., 2018). When the  
248 paired-ends were joined, the unreliable and low-quality sequences were detected based  
249 on the low sequence quality of the 3'-ends of the reads (Merloti et al., 2019). After that,  
250 the Amplicon Sequence Variants (ASVs) were created using the Deblur tool. The  
251 resulting final ASV table contained only high-quality reads. The taxonomic  
252 identification of ASVs (with 99% of similarity) was conducted with the VSEARCH  
253 consensus taxonomy classifier implemented in Qiime2 and the SILVA 16S rRNA

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254 database. The archaeal and chloroplast sequences were removed. Statistical analyses of  
255 the 16S rRNA microbiome sequencing data were performed using the generated  
256 taxonomic matrices. The sequencing data were submitted to NCBI (BioProject  
257 accession number: PRJNA644210).

258

### 259 *2.7. Statistical analysis of sequencing data*

260

261 The  $\alpha$ -diversity of bacteria was characterized by the Simpson's diversity index and the  
262 effective number of species, and the data were analyzed with ANOVA using Duncan's  
263 test ( $P < 0.05$ ).  $\beta$ -diversity was calculated based on weighted-UniFrac distance metrics.  
264 Principal coordinates analysis (PCoA) was used to separate the overall bacterial  
265 community structure based on the weighted-UniFrac distance with the "pcoa" function  
266 of the "ape" package in R (v3.2.2). Permutational multivariate analysis of variance  
267 (PERMANOVA) was performed to separate and evaluate the effects of sex, plantation  
268 mode and Zn treatment and their interactions on the rhizospheric bacterial communities  
269 using the "anosims" function in "vegan" package in R (v3.2.2). The linear discriminant  
270 analysis (LDA) effect size (LEfSe) algorithm was used to analyze the relative bacterial  
271 abundance in different treatments, irrespective of sexual interactions (Class: Zn  
272 treatment; Subclass: interaction patterns), as well as in different sexual interaction  
273 patterns irrespective of Zn treatment (Class: sexual interaction; Subclass: Zn treatment).  
274 The online Galaxy application (version 1.0) was used with a threshold of 1.0 and a  
275 Wilcoxon  $P$ -value of 0.05 (<http://huttenhower.sph.harvard.edu/galaxy/>). A similarity

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276 percentage (SIMPER) analysis was performed with the PRIMER 6 software to find  
277 differences between sexual interaction patterns and other treatments. The OTUs less  
278 than 1% of the relative abundance were discarded from the analysis according to  
279 Marasco et al. (2018). Correlation coefficients between the bacterial abundance at the  
280 phyla level and Zn amount were obtained with python (SciPy package). The heatmap  
281 of correlation coefficients was plotted with R (Heatmap package).

282

### 283 *2.8. PICRUSt functional prediction*

284

285 The *PICRUSt* (*phylogenetic investigation of communities by reconstruction of*  
286 *unobserved states*, v1.0.0) pipeline was used to predict the relative abundance of gene  
287 transcripts in bacteria. The *pick\_closed\_reference\_otus.py* script in QIIME (similarity  
288 threshold, 0.97) was used to cluster sequences into OTUs. The OTUs were first  
289 normalized by the copy number by removing the copy number of the 16S marker gene.  
290 The Nearest Sequenced Taxon Index (NSTI) and KEGG Ortholog (KO) were obtained  
291 by the *predict\_metagenomes.py* script. The KO level 3 within the pathway hierarchy of  
292 KEGG was collapsed using the *categorize\_by\_function.py* script (Langille et al., 2013;  
293 Hou et al., 2017). The predicted metagenomes were then annotated using the KEGG  
294 database, and the predicted relative abundances of gene transcripts for selected  
295 pathways associated with bacterial functions were analyzed and plotted using STAMP  
296 (Parks et al. 2014; Hou et al. 2017).

## 297 **3. Results**

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298

299 *3.1. Plant growth and Zn phytoextraction*

300

301 Under control conditions (no excess Zn), males showed 21% and 48% higher CO<sub>2</sub>  
302 assimilation rate (*A*) and stomatal conductivity (*g<sub>s</sub>*) respectively, than females under  
303 single-plant cultivation (Table 1). Females had lowest *A* and *g<sub>s</sub>* under intersexual  
304 interaction (F/FM) under control conditions (Table 1). Compared with controls, *A* and  
305 *g<sub>s</sub>* decreased by 29% and 18%, respectively, in females and by 19% and 22%,  
306 respectively, in males in single-sex cultivation under excess Zn (Table 1). Excess Zn  
307 decreased *A* by 18%, 11%, 12% and 16%, and *g<sub>s</sub>* by 20%, 22%, 32% and 57%, in F/FM,  
308 M/FM (males in intersexual interaction), F/FF (females in intrasexual interaction) and  
309 M/MM plants (males in intrasexual interaction), respectively (Table 1). In addition,  
310 M/FM exhibited 10% and 55% increases in *A* and *g<sub>s</sub>* compared to M/MM under  
311 excess Zn conditions (Table 1).

312

313 Excess Zn treatment decreased  $F_v/F_m$ , *ETR* and *Y (II)* by 6%, 37% and 50%,  
314 respectively, in females and by 4%, 17% and 8%, respectively, in males under single-  
315 sex cultivation (Table 1). Sexual interactions did not affect  $F_v/F_m$  in either sex under  
316 control conditions (Table 1). Excess Zn did not affect  $F_v/F_m$  and *ETR* in males (M/MM)  
317 and females (F/FF) when compared with controls. However,  $F_v/F_m$ , *ETR* and *Y (II)* were  
318 4%, 37% and 54% higher, respectively, in M/FM plants than in F/FM plants under  
319 excess Zn, and *Y (II)* of males was 13% higher when compared to females under

---

320 intrasexual interaction (Table 1). In addition, excess Zn treatment increased  $Y$  ( $NPQ$ )  
321 by 71% in females, but not in males, under single-plant cultivation (Table 1) when  
322 compared to control conditions. Sexual interactions did not affect  $Y$  ( $NPQ$ ) in either sex  
323 under control conditions. However, excess Zn increased  $Y$  ( $NPQ$ ) by 32%, 58% and 14%  
324 in F/FM, F/FF and M/MM, respectively, when compared to controls (Table 1).

325

326 Excess Zn increased the Zn content in by 22.9- and 4.3-fold, respectively, in leaves,  
327 by 3.1- and 1.4-fold, respectively, in shoots, and by 2.7- and 3.2-fold, respectively, in  
328 roots of females and males under single-sex cultivation (Fig. 1). Under excess Zn  
329 conditions, Zn levels increased by 4.6-, 29.6-, 13.6- and 13.1-fold, respectively, in  
330 leaves, by 97%, 300%, 327% and 219%, respectively, in stems, and by 160%, 240%, %  
331 and 160%, respectively, in roots of F/FM, M/FM, F/FF and M/MM when compared to  
332 controls (Fig. 1). Under excess Zn, the Zn level of females was 37% and 39% higher in  
333 leaves and roots, respectively, but 82% lower in stems when compared to males in  
334 single-sex cultivation under excess Zn conditions. The leaf and stem Zn levels of males  
335 were lower than those of females under intersexual interaction; the contrary was true  
336 for intrasexual interaction under excess Zn (Fig. 1). In addition, there was no difference  
337 in root Zn concentrations between sexes under intersexual interaction, whereas the Zn  
338 level in roots of M/MM was 26% lower than that of F/FF plants.

339

340 Females had higher  $T_f$  (the ratio of Zn in the shoots to roots) than males under single-  
341 sex cultivation.  $T_f$  was higher in both sexes under single-sex cultivation than under



---

342 intersexual interactions (except for F/FF) when exposed to excess Zn (Fig. 1). F/FM  
343 had highest  $T_f$  and F/FF showed the lowest  $T_f$  among all excess Zn treatments (Fig. 1).  
344 Under excess Zn, M/MM showed 16% higher  $T_f$  than M/FM. In single-sex cultivations  
345 under excess Zn, females had 141% and 98% increases, respectively, in root BCF (the  
346 ratio of Zn in the roots to soil) and shoot BCF (the ratio of Zn in the shoots to soil) when  
347 compared to males. F/FF had highest root BCF among all excess Zn treatments. Root  
348 and shoot BCF values were 48% and 72% higher, respectively, in M/MM than in M/FM  
349 under excess Zn, while females had 1.2-fold higher shoot BCF than males under  
350 intersexual interaction with excess Zn (Fig. 1).

351

### 352 *3.2. Dominant taxa of bacterial communities*

353

354 The bacterial communities predominantly consisted of Proteobacteria (33%),  
355 Actinobacteria (17%), Acidobacteria (16%), Chloroflexi (14%), Planctomycetes (6%),  
356 Bacteroidetes (4%), Gemmatimonadetes (2%), Crenarchaeota (2%), Verrucomicrobia  
357 (1%), and Nitrospirae (1%). These dominant species were affected by plant interactions  
358 and excess Zn (Fig. 2). Excess Zn reduced the relative abundance of the Proteobacteria  
359 phylum in the rhizosphere of M/MM and FM, but did not affect Proteobacteria in the  
360 rhizosphere of F/FF. Excess Zn reduced the abundance of Actinobacteria and  
361 Planctomycetes phyla in both sexes under intersexual interactions. The relative  
362 abundance of Proteobacteria and Actinobacteria phyla were higher under the excess Zn  
363 treatment than under control conditions in both sexes when plants were cultivated alone,

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364 while the contrary was true for the Chloroflexi phylum (Fig. 2).

365

### 366 *3.3. Bacterial community diversity and structure*

367

368 The alpha-diversities of bacterial communities (effective numbers of species) were  
369 compared in females and males under different plantation modes and excess Zn  
370 treatments. As shown in Fig. 3, intra- and intersexual interaction affected bacterial  
371 community structures, but the effects of these interaction patterns were greater under  
372 excess Zn. Permutational multivariate analysis of variance (PERMANOVA) further  
373 demonstrated that plantation modes were the largest source of variation (25.16%,  $P <$   
374 0.001; Table 2). The Zn levels were the second largest source of variation (5.00%,  $P <$   
375 0.004; Table 2).

376

### 377 *3.4. Taxonomic composition of bacterial communities*

378

379 The linear discriminant analysis (LDA) effect size analysis (LEfSe) was performed to  
380 compare the bacterial composition from phyla to genera between Zn treatments, as well  
381 as between sexes (Fig. 4). We found that bacterial compositions showed significant  
382 differences among sexual interaction patterns and Zn treatments. The orders  
383 Acidimicrobiales, Micromonosporaceae, RB40, Verrucomicrobiales, Thiotrichales,  
384 MND1, Piscirickettsiaceae and Pirellulaceae were predominant under excess Zn, while  
385 the genera *Pontibacter*, *Chryseobacterium*, *Lysobacter*, *Kaistobacter* and *Massilia*, and

---

386 the class Flavobacteria were enriched in plants without excess Zn under single  
387 cultivation (Fig. 4a). Moreover, the order Sva0725, the family A4b and the genus  
388 *Rubrivivax* were dominant in the rhizosphere of males, while other taxa were enriched  
389 in soil with females (Fig. 4b).

390

391 In sexual interaction experiments, the order Nitrososphaerales was generally more  
392 abundant in the soil of control plants (Fig. 4c). Under excess Zn, the class Saprospirae,  
393 the orders GCA004, Solirubrobacterales and AKYG1722, the families *AKIW874*,  
394 *Cellulomonadaceae*, *Mycobacteriaceae* and *Sporichthyaceae*, and the genera *Bacillus*,  
395 *Rhodoplanes* and *Mycobacterium* were predominant (Fig. 4d). Irrespective of the Zn  
396 treatment, the genera *Agromyces* and *Mycobacterium*, as well as the family  
397 *Mycobacteriaceae* were abundant in the soil of M/MM, whereas the order  
398 Methanobacteriales and the genus *Methanothermobacter* were more abundant in the  
399 soil of F/FF under intrasexual interaction (Fig. 4d). In contrast, orders Caldilineales and  
400 NB1-j were more prevalent in the soil of FM exposed to intersexual interaction (Fig.  
401 4d).

402

### 403 *3.5. Functional predictions of bacterial communities*

404

405 The PICRUSt analysis was performed to predict the metagenome gene functions. As  
406 shown in Fig. 5, the PICRUSt analysis suggested that the predicted differential gene  
407 abundances between treatments were related to carbohydrates, amino acids, lipids and

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408 secondary metabolism in the rhizosphere of females and males under single-plant  
409 cultivation. When compared to females under single-plant cultivation, males had a  
410 higher abundance of predicted genes involved in the citrate cycle and carotenoid  
411 biosynthesis in the rhizosphere in control conditions (Fig. 5a). Excess Zn stress reduced  
412 arachidonic acid metabolism, flavonoid biosynthesis and peptidoglycan, and increased  
413 the abundance of predicted genes related to glycolysis, citrate cycle and leucine  
414 synthesis in females under single-plant cultivation.

415

416 Under plant interactions, the differentially abundant genes in different treatments in the  
417 rhizospheres of both sexes were related to amino acid translation, and amino acid and  
418 lipolic acid metabolism under single-plant cultivation (Fig. 5b). Excess Zn reduced the  
419 abundance of the predicted genes related to cysteine, methionine and thiamine  
420 metabolism, especially in the rhizosphere of females exposed to intrasexual interaction.  
421 However, excess Zn elevated the abundance of predicted genes related to tyrosine  
422 metabolism in the rhizosphere of females, compared to males, both under inter- and  
423 intraspecific interaction (Fig. 5b).

424

### 425 *3.6. Relative abundance of key taxa associated with Zn levels in plants*

426

427 We examined the bacterial abundance in the rhizosphere at the genus level in relation  
428 to Zn levels in roots (Fig. 6). First, we examined the relative abundance of taxa  
429 positively associated with Zn levels. Under single-plant cultivation, females showed a

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430 relatively higher abundance of families Streptomycetaceae and Rhodobacteraceae,  
431 and the phylum Chloroflexi than males under an increasing Zn supply in the rhizosphere.  
432 In contrast, the increased Zn level raised the relative abundance of the family  
433 Sphingomonadaceae in the rhizosphere of males compared to the female rhizosphere.  
434 Under intrasexual interaction, males had a higher relative abundance of the  
435 Actinomycetales order and Phyllobacteriaceae family when compared to intersexual  
436 interaction (Fig. 6). By contrast, the family Alphaproteobacteria was more abundant in  
437 the rhizosphere of males exposed to excess Zn under intersexual interaction than under  
438 intrasexual interaction. In females, the orders Myxococcales, Roseiflexales and  
439 Actinomycetales, and the class Betaproteobacteria were predominant in the rhizosphere  
440 exposed to intersexual interaction under excess Zn relative to intrasexual interaction  
441 (Fig. 6). The increased abundance of the order Rhizobiales under excess Zn was specific  
442 to females exposed to intrasexual interaction relative to intersexual interaction. Under  
443 intersexual interaction, the abundance of the order Actinomycetales was higher in males,  
444 while the order Rhizobiales was more abundant in the rhizosphere of females under  
445 excess Zn (Fig. 6).

446

447 Then, we examined the relative abundance of bacteria negatively associated with Zn  
448 levels in roots (Fig. 6b). We found that under excess Zn, the relative abundance of most  
449 bacterial classes, such as Planctomycetia, Betaproteobacteria and Acidobacteria-6,  
450 were lower in the rhizosphere of females than in that of males under single-sex  
451 conditions. When compared to females under single-sex conditions, males had a higher

---

452 abundance of the family Hyphomonadaceae and the order WD2101 irrespective of the  
453 Zn application. Under intrasexual interaction with excess Zn, the abundance of the class  
454 Gemm-5 in the rhizosphere of males, and the abundance of the genus *Planctomyces* in  
455 females and males were higher than under intersexual interaction (Fig. 6b). On the other  
456 hand, under intersexual interaction with excess Zn, the abundance of the RB40 family  
457 was higher but the abundance of the genus *Planctomyces* was lower in the rhizosphere  
458 of females than in males.

459

#### 460 **4. Discussion**

461

##### 462 *4.1. Sexual interactions affect Zn phytoextraction in plants*

463

464 Neighbor interactions of plants affect plant growth and community composition (Hodge  
465 and Fitter, 2013; Hawkins and Crawford, 2018). In turn, these interactions could be  
466 altered by abiotic stress, e.g. heavy metals (Chen et al., 2016, 2017). It has been shown  
467 that abiotic stress can alter competitive interactions (Chen et al., 2016, 2017). Selenium  
468 has been found to strongly influence plant-plant interactions and play a vital role in  
469 elemental allelopathy (El Mehdawi et al., 2011). In this study, under excess Zn stress,  
470 F/FM accumulated more Zn in their leaves and barks compared to F/FF, while males  
471 accumulated less under intersexual interaction than under intrasexual interaction (Table  
472 1; Fig. 1). Responses of sexes to abiotic stresses depend mainly on the specific  
473 properties of stresses, the soil status and the exposure duration (Howard et al., 2000;

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474 Tonneijck et al., 2004). Zn is an essential trace element for plants and, in our study, the  
475 leaf Zn levels did not reach the toxic threshold for poplars under any interaction (Fig.  
476 1). Females and males exposed to inter- and intrasexual interaction showed strong  
477 tolerance to excess Zn (Table 1). Interestingly, we found that under intrasexual  
478 interaction both females and males accumulated more Zn in the roots, which largely  
479 inhibited Zn transportation to the shoots and leaves (Fig. 1). The reduced heavy metal  
480 accumulation in leaves facilitates heavy metal tolerance in woody plants (He et al.  
481 2013a).

482

483 The tolerance of plants to heavy metals is also reflected in leaf photosynthesis (Chandra  
484 and Kang, 2016; Salisbury et al., 2018), which indicates the ability of plants to thrive  
485 in specific environments, including heavy metal stress (Simkin et al., 2019; Hu et al.,  
486 2020). The accumulation of heavy metal in leaves damages photosynthetic electron  
487 transport and reduces photosynthesis, which are signs of stress (Aggarwal et al., 2012).  
488 We found that females had higher Zn enrichment but lower Zn toxicity tolerance than  
489 males under Zn treatments. The inhibited photosynthesis and electron transmission  
490 efficiency were associated with Zn translocation into leaves (Fig. 1; Table 1). Females  
491 had higher root-to-shoot translocation under intrasexual than intersexual interaction, as  
492 visible as elevated root BCF but lower  $T_f$  under excess Zn (Fig. 1). However, males  
493 favored to accumulate more Zn in roots, showing higher root BCF but lower shoot BCF  
494 (Fig. 1). Moreover, females had a higher Zn translocation efficiency ( $T_f$ ) under  
495 intersexual than intrasexual interaction, while the contrary was true for males (Fig. 1).

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496 Moreover, the Zn translocation efficiency was higher in females than in males when  
497 subjected to excess Zn under intersexual interaction. In previous studies, female plants  
498 have showed higher competitive ability than males (Mercer and Eppley, 2010; Sánchez-  
499 Vilas et al., 2011; Chen et al., 2016). Consistently, we found that females had a stronger  
500 competitive ability under excess Zn even when excess Zn was toxic for plants. In a  
501 previous study, we found that females had a higher expression of genes, such as those  
502 from the ZIP family, which facilitate Cd uptake and root-to-shoot translocation in *P.*  
503 *cathayana* (Liu et al., 2020b). Importantly, the potential of females and males for  
504 extracting Zn also depends on underground processes that affect Zn availability (Fig.  
505 6).

506

#### 507 *4.2. Soil microbial composition modified by plantation modes and Zn pollution*

508

509 Heavy metals affect rhizosphere-driven microbial community structures (Hou et al.,  
510 2017). Soil microorganisms are also critical drivers in modifying community structures  
511 and plant-plant interactions (Hodge and Fitter, 2013; Hawkins and Crawford, 2018).  
512 The sexually differential trade-off between plant rewards and defense has been shown  
513 to influence the colonization of host-specific microbial communities (Bever et al., 2012;  
514 Varga et al., 2017). This finding was further demonstrated by our study. We discovered  
515 that the bacterial community diversity and richness in the rhizosphere of both sexes  
516 were not affected by plantation modes or excess Zn, but probably associated with the  
517 planting time and species (Siciliano et al., 2014; Tian and Gao, 2014; Fontana et al.,



---

518 2016). However, the bacterial community structure was strongly affected by plantation  
519 modes and Zn treatments (Fig. 3), which probably affected root Zn bioavailability.

520

521 Generally, heavy metals suppress the growth and abundance of low-resistance microbes  
522 but increase the prevalence of high-resistance bacteria (Pishchik et al., 2016; Wood et  
523 al., 2016; Hou et al., 2017). Consistently, the relative abundance of Cytophagales was  
524 significantly inhibited by excess Zn in both sexes, while the abundance of key phyla,  
525 such as Betaproteobacteria, Acidobacteria, Nitrospirales and Proteobacteria, increased  
526 under excess Zn in single-cultivated males (Fig. 4; Figs S1-S2). Interestingly, the  
527 abundance of Betaproteobacteria, Acidobacteria, Nitrospirales and Proteobacteria phyla  
528 were positively associated with root Zn levels (Fig. 6), which suggested that they are  
529 probably involved in excess Zn uptake. In addition, we found that when compared to  
530 untreated soil, Zn levels in excess Zn-treated soil were 3~7 -fold higher (Fig. S5), and  
531 the level was higher than the Zn toxicity risk screening value for soil (Huang, 2014;  
532 Carruthers, 2016). Our results support the view that phytoremediation may be an  
533 effective strategy to improve the soil quality by recruiting some beneficial microbes  
534 (Ancona et al., 2020). It is evident that if attempting to use microbial remediation of  
535 metals, it would be important to investigate further the responses of specific bacterial  
536 families.

537

538 Sexual interactions alter the structure of root systems and the secretion of root exudates  
539 in *P. cathayana* females and males, thus modifying the rhizosphere ecology (Ke and

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540 Wang, 2020; Xia et al., 2020). The present study found that the Methanotheriales,  
541 Mycobacterium and Caldilineales phyla were enriched in the rhizosphere of F/FF,  
542 M/MM and FM, respectively (Fig. 4). Among these specific bacteria, the Actinobacteria  
543 and Streptomyces phyla have been identified as the most dominant taxa in soils with  
544 heavy metal contamination (Watve et al., 2001; Ellis et al., 2003; Alvarez et al., 2017),  
545 and they became more abundant under excess Zn in the rhizosphere of M/MM and  
546 M/FM (Fig. 4; Figs S3-S4). More importantly, the relative abundance of Actinobacteria  
547 showed a positive correlation with root Zn levels in M/MM and FM (Fig. 6).  
548 Actinobacteria produce siderophores for Cd acquisition and protect plants from the  
549 invasion of pathogenic bacteria (Viaene et al., 2016). The specific colonization of  
550 Actinobacteria in the rhizosphere of *Arabidopsis* confers a competitive advantage to  
551 these plants (Van der Meij et al., 2018). The increased abundance of Actinobacteria  
552 following the excess Zn treatment in the rhizosphere of FM was probably due to Zn  
553 chelation, which was likely the result of a long-term adaptation of sexes to excess Zn  
554 under intersexual interaction. *Streptomyces* is the largest genus producing antibiotics  
555 (Watve et al., 2001; Hong et al., 2009). It has been suggested that some strains  
556 belonging to *Streptomyces* promote metal solubility and auxin synthesis by stimulating  
557 siderophore synthesis (Zloch et al., 2016; Hou et al., 2017). Taking into account the  
558 higher Zn accumulation in females under single-sex and intersexual modes (Fig. 1),  
559 Streptomycetaceae probably plays a key role in the Zn accumulation in *P. cathayana*  
560 females under excess Zn.

561

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562 We also found some sex-specific bacterial colonization in the *P. cathayana* rhizosphere  
563 under different interaction patterns. The excess Zn treatment increased the abundance  
564 of the phylum Euryarchaeota in both females and males with intrasexual interaction  
565 (Figs S3-S4). In addition, the abundances of the phyla Actinobacteria and Chloroflexi  
566 were elevated under excess Zn in the rhizosphere of FM and M/MM (Fig. 4; Figs S3-  
567 S4). Overall, our results suggest that most bacteria prefer to colonize the rhizosphere of  
568 both sexes in the case of intersexual interaction, irrespective of Zn levels, as well as in  
569 the rhizosphere of males exposed to intrasexual interaction.

570

#### 571 *4.3. Microbial functional prediction highlights the role of sexual interaction patterns*

572

573 The PICRUSt analysis suggested that excess Zn and sexual interactions differentially  
574 regulate the abundance of predicted gene transcripts of bacteria in the rhizosphere. The  
575 excess Zn treatment increased the abundance of predicted gene transcripts related to  
576 amino acids and organic acids, such as valine, leucine and isoleucine biosynthesis, and  
577 the citrate cycle in both sexes under single-plant cultivation, which probably increased  
578 organic matter from dead roots, exudates and rhizodeposits (Hou et al., 2017, 2018). It  
579 is worth noting that the abundance of predicted gene transcripts related to carotenoid  
580 biosynthesis was higher in the rhizospheric bacteria of males than in those of females  
581 (Fig. 5a), which probably played an important role in heavy metal tolerance. Excess Zn  
582 increased the abundance of predicted gene transcripts related to tyrosine metabolism  
583 and secondary metabolites in males exposed to intersexual interaction and in females

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584 under both intra- and intersexual interaction, suggesting that excess Zn differentially  
585 regulates the abundance of predicted gene transcripts. Further studies are needed to  
586 explore the relationship between heavy metal stress and gene expression among the  
587 bacteria in the rhizospheres of both sexes.

588

589

590

## 591 **5. Summary**

592

593 This study suggested that excess Zn and sexual interactions strongly affect Zn  
594 accumulation, and microbial abundance and diversity. Females have higher Zn  
595 accumulation and root-to-shoot translocation under intersexual than under intrasexual  
596 interaction combined with excess Zn; the contrary was true for males. Moreover, the  
597 plantation modes and excess Zn treatment altered the bacterial structure of the  
598 rhizosphere, which largely affected the Zn availability and uptake of roots. The Zn-  
599 polluted soil with males growing under different interactions promoted some key  
600 bacterial taxa related to metal activation and chelation in the rhizosphere, e.g.,  
601 Actinobacteria and Streptomyces, while the excess Zn-treated soil with females had a  
602 lower bacterial abundance under intrasexual interaction than under intersexual  
603 interaction. To our knowledge, this is the first study that has showed a possible  
604 relationship of Zn availability with specific bacterial colonization in the rhizosphere of  
605 *P. cathayana* females and males under sexual interactions. Our study provides new

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606 insight into the interactions among plant sex, plantation types, heavy metal stress and  
607 the composition of microbial communities.

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616

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618 analysis and writing, Yuting Wang and Xiucheng Liu contributed to data collection,  
619 Helena Korpelainen contributed to the interpretation of data and manuscript preparation,  
620 and Chunyang Li (the corresponding author) had the overall responsibility for  
621 experimental design and project management.

622

623 **Conflict of interest** The authors declare that they have no conflict of interest.

624

625 **Appendix A. Supplementary data** Supplementary material related to this article can  
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918 **Figure legends**

919

920 **Fig 1.** Zinc (Zn) accumulation in leaves (a), stems (b) and roots (c), translocation factor  
921 ( $T_f$ ) (d), the root bio-concentration factor (BCF) (e) and shoot BCF (f) in *P. cathayana*  
922 females and males under different cultivation patterns and excess Zn. Values are  
923 expressed as means  $\pm$  SD ( $n = 4$ ). Different letters represent significant differences  
924 between treatments ( $P < 0.05$ ). F, female; M, male; F/FF, female under intrasexual  
925 interaction; M/FM, male under intersexual interaction; M/MM, male under intrasexual  
926 interaction; F/FM, female under intersexual interaction.

927

928 **Fig 2.** Taxonomic compositions of bacterial communities in the rhizospheres of *P.*  
929 *cathayana* females and males at the phylum level with the relative abundance over 1%  
930 under excess zinc (Zn). Rhizospheric soil samples from plants under single-sex (a) and  
931 double-sex interactions (b). Average relative abundance of key phyla in the rhizospheric  
932 soil from plants under single-sex (c) and double-sex interactions (d); FF, female with  
933 intrasexual interaction; FF+Zn, FF with excess Zn; FM, female and male under  
934 intersexual interaction; FM+Zn, FM with excess Zn; MM, male with intrasexual  
935 interaction; MM+Zn, MM with excess Zn. A, Proteobacteria; B, Actinobacteria; C,  
936 Chloroflexi; D, Acidobacteria; E, Planctomycetes; F, Bacteroidetes; G,  
937 Gemmatimonadetes.

938

939 **Fig 3.** Box plots for alpha diversity (effective number of species) (a, b) and beta-

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940 diversity (principal co-ordinates analysis) of bacteria from the rhizospheric soil of *P.*  
941 *cathayana* females and males under excess zinc (Zn) treatment. Alpha diversity of  
942 bacteria from the rhizospheric soil of plants under single-sex (a) and double-sex  
943 interaction (b). Principal coordinate analysis plots among bacterial communities in the  
944 rhizosphere of *P. cathayana* females and males under single-sex interactions (c) and  
945 double-sex interactions (d). F/FM, female under intersexual interaction; M/FM, male  
946 under intersexual interaction; FF, female under intersexual interaction; MM, male under  
947 intrasexual interaction.

948

949 **Fig 4.** LEfSe used to identify abundant taxa in the rhizospheric soil from *P. cathayana*  
950 females and males under excess zinc (Zn) treatment. A cladogram was generated by  
951 LEfSe indicating differences between bacteria at phylum, class, family and genus levels  
952 under single-plant cultivation between excess zinc (Zn) and control soil (no-Zn  
953 treatment), irrespective of sex and interaction (a), and between sexes, irrespective of Zn  
954 treatment (b). Another cladogram was generated by LEfSe indicating differences  
955 between bacteria at phylum, class, family and genus levels between excess Zn and  
956 control soil, irrespective of sex and interaction (c), and between interaction modes,  
957 irrespective of Zn treatment (d). The node colour indicates taxa enriched under different  
958 treatment and interaction patterns. Only taxa with LDA over 3 are shown. FF, female  
959 under intrasexual interaction; FM, female and male under intersexual interaction; MM,  
960 male under intrasexual interaction.

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962 **Fig 5.** Predicted functions of the bacterial communities from the rhizospheric soil of *P.*  
963 *cathayana* females and males under excess zinc (Zn) treatment. Predicted functions of  
964 the bacterial communities from the rhizospheric soil of plants from single sex (a) and  
965 double-sexual interactions (b). F, female; F+Zn, F with excess Zn; M, male; M+Zn, M  
966 with excess Zn; F/FM, females under intersexual interaction; M/FM, male under  
967 intersexual interaction; FF, female under intrasexual interaction; FF+Zn, FF with excess  
968 Zn; MM, male under intrasexual interaction; MM+Zn, male under intrasexual  
969 interaction.

970

971 **Fig 6.** Heatmaps of the average relative abundance of key bacteria at the genus level  
972 positively (a) or negatively associated (b) with Zn levels in roots of *P. cathayana*  
973 females and males exposed to different interaction patterns ( $P < 0.05$ ). F, female; M,  
974 male; FF, female exposed to intrasexual interaction; M/FM, male exposed to intersexual  
975 interaction; MM, male exposed to intrasexual interaction; F/FM, female exposed to  
976 intersexual interaction.

**Table 1** Net photosynthesis rate ( $A$ ), stomatal conductance ( $g_s$ ), and fluorescence parameters  $F_v/F_m$ ,  $ETR$ ,  $Y(II)$  and  $Y(NPQ)$  of *P. cathayana* females and males exposed to different sexual interactions under control conditions or excess zinc (Zn).  $F_v/F_m$ , maximum quantum efficiency of *PS II* photochemistry;  $ETR$ , electron transport rate;  $Y(NPQ)$ , quantum yield of regulated non-photochemical energy loss in *PS II*;  $Y(II)$ , quantum yield of photochemical energy conversion in *PS II*.

| Treatment    | $A$ ( $\mu\text{mol m}^{-2} \text{s}^{-1}$ ) | $g_s$ ( $\text{mol m}^{-2} \text{s}^{-1}$ ) | $F_v/F_m$    | $ETR$         | $Y(II)$     | $Y(NPQ)$    |
|--------------|--|---|--------------|---------------|-------------|-------------|
| Control      |  |   |              |               |             |             |
| F            | 19.21±0.35d                                  | 0.24±0.02f                                  | 0.79±0.01cd  | 51.48±0.52abc | 0.45±0.04bc | 0.28±0.01fg |
| M            | 23.31±1.41a                                  | 0.36±0.03b                                  | 0.81±0.00a   | 52.88±2.85ab  | 0.44±0.02cd | 0.33±0.01de |
| F/FM         | 19.11±0.27d                                  | 0.29±0.00de                                 | 0.80±0.01bc  | 49.10±3.12cd  | 0.44±0.02cd | 0.27±0.11fg |
| M/FM         | 22.49±0.26b                                  | 0.41±0.01a                                  | 0.80±0.01abc | 54.4±2.20a    | 0.49±0.01a  | 0.26±0.01fg |
| F/FF         | 22.15±0.93ab                                 | 0.36±0.01b                                  | 0.80±0.01bc  | 51.93±1.86abc | 0.48±0.02ab | 0.28±0.02f  |
| M/MM         | 21.64±0.48b                                  | 0.34±0.03bc                                 | 0.81±0.01ab  | 54.83±2.35a   | 0.50±0.01a  | 0.28±0.01fg |
| Zn treatment |  |   |              |               |             |             |
| F            | 13.73±0.40f                                  | 0.20±0.01g                                  | 0.75±0.02f   | 32.23±2.21g   | 0.23±0.01i  | 0.49±0.02a  |
| M            | 18.94±0.01e                                  | 0.28±0.03e                                  | 0.78±0.00d   | 43.78±1.12e   | 0.40±0.02e  | 0.35±0.01cd |
| F/FM         | 18.11±0.27e                                  | 0.23±0.02fg                                 | 0.76±0.01e   | 36.88±3.59f   | 0.27±0.04h  | 0.36±0.04c  |
| M/FM         | 20.12±0.26c                                  | 0.32±0.02cd                                 | 0.80±0.01bcd | 50.68±1.92bc  | 0.41±0.02de | 0.25±0.03g  |
| F/FF         | 19.77±0.86cd                                 | 0.25±0.04f                                  | 0.79±0.00d   | 44.33±1.36e   | 0.32±0.03g  | 0.45±0.04b  |
| M/MM         | 18.15±0.10e                                  | 0.15±0.02h                                  | 0.78±0.01d   | 47.00±1.43de  | 0.36±0.01f  | 0.32±0.02e  |

F, female; M, male; FF, female exposed to intrasexual interaction; M/FM, male exposed to intersexual interaction; MM, male exposed to intrasexual interaction; F/FM, female exposed to intersexual interaction; Different letters represent significant differences between treatments ( $P < 0.05$ ).

Values are means  $\pm$  SE ( $n = 4$ ).

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**Table 2** PERMANOVA results using Bray-Curtis as a distance metric.

| Factor                | % explained | <i>F</i> | R <sup>2</sup> | <i>P</i> |
|-----------------------|-------------|----------|----------------|----------|
| Plantation modes      | 25          | 4.38     | 0.25           | 0.001*** |
| Zn                    | 5           | 3.49     | 0.05           | 0.004**  |
| Sex                   | 5           | 3.19     | 0.05           | 0.006**  |
| Sex : Zn              | 2           | 1.28     | 0.02           | 0.227    |
| Plantation modes : Zn | 11          | 1.91     | 0.11           | 0.007**  |
| Residuals             | 52          |          | 0.52           |          |

Plantation modes: F, M, F/FF, FM, M/MM; sex: F, M; Zinc, excess Zn treatment. \*

0.01 < *P* ≤ 0.05; \*\* 0.001 < *P* ≤ 0.01; \*\*\* *P* ≤ 0.001.



Fig. 1

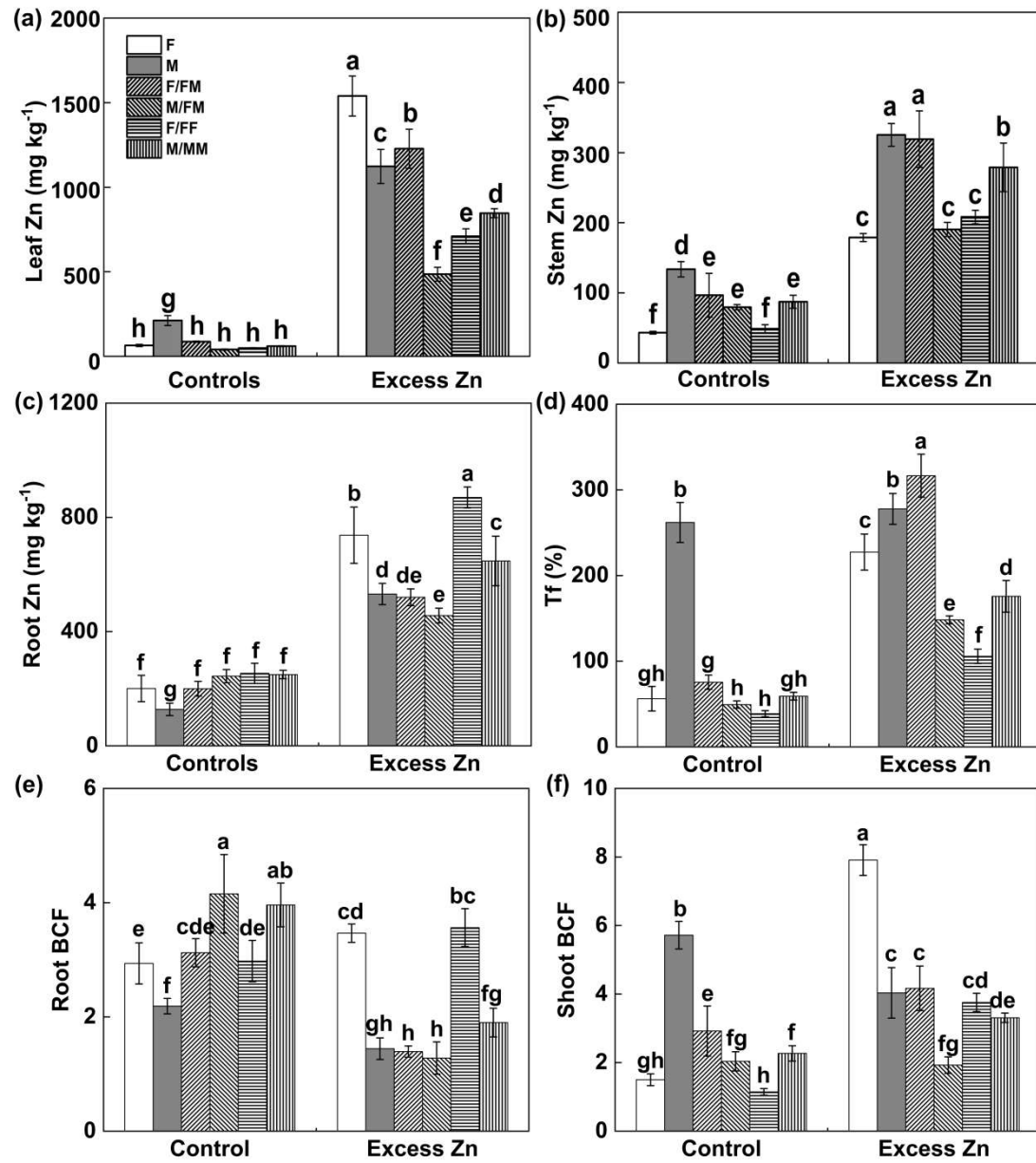


Fig. 2

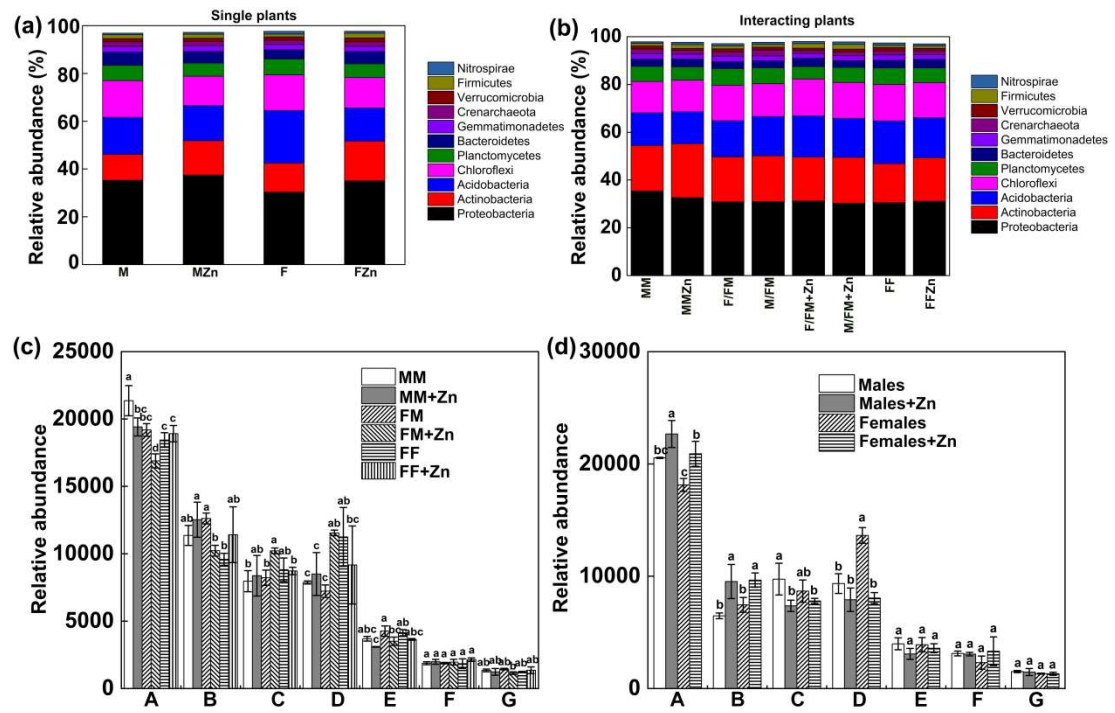


Fig. 3

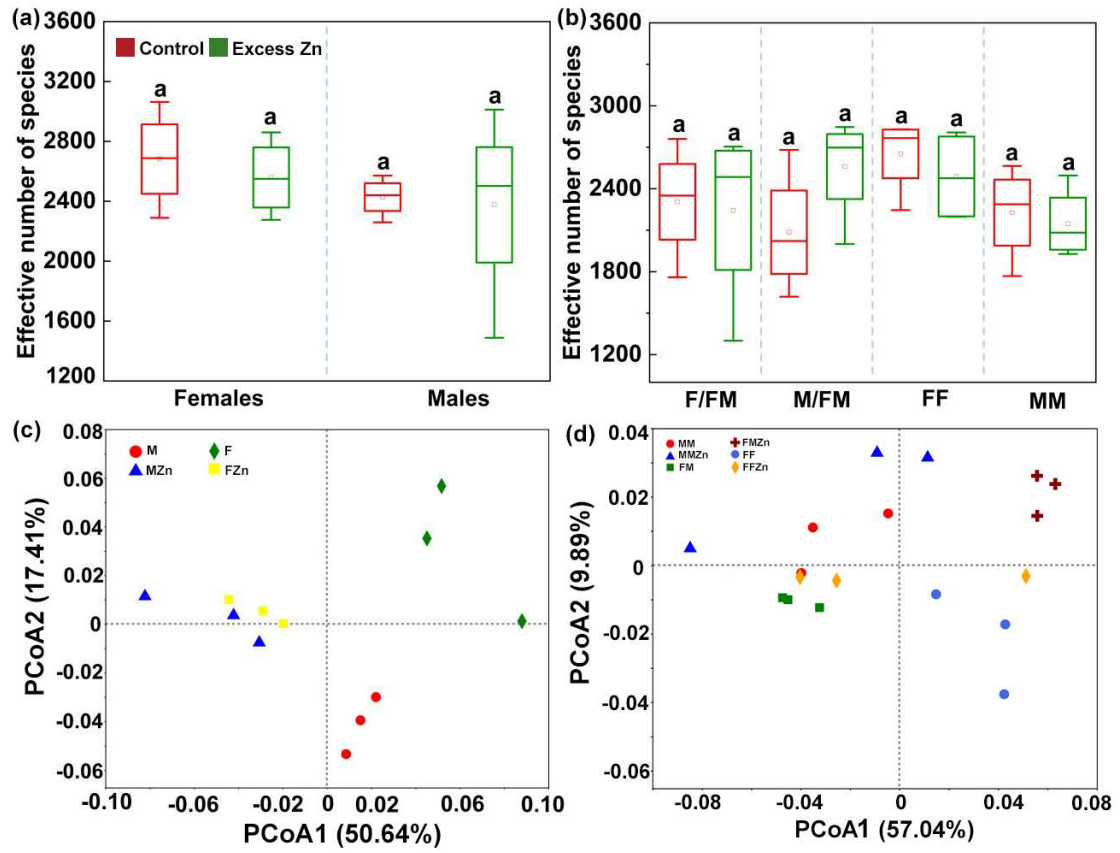


Fig. 4

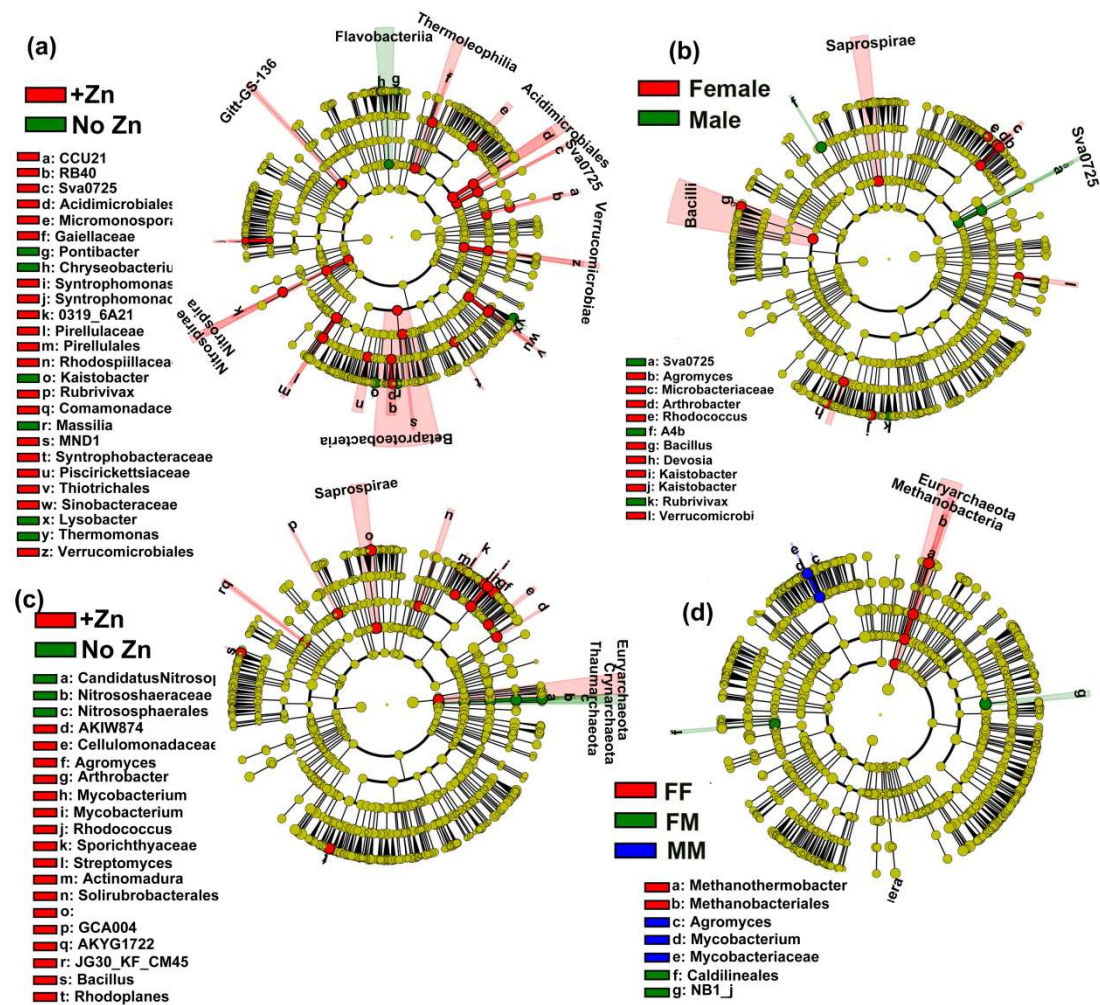


Fig. 5

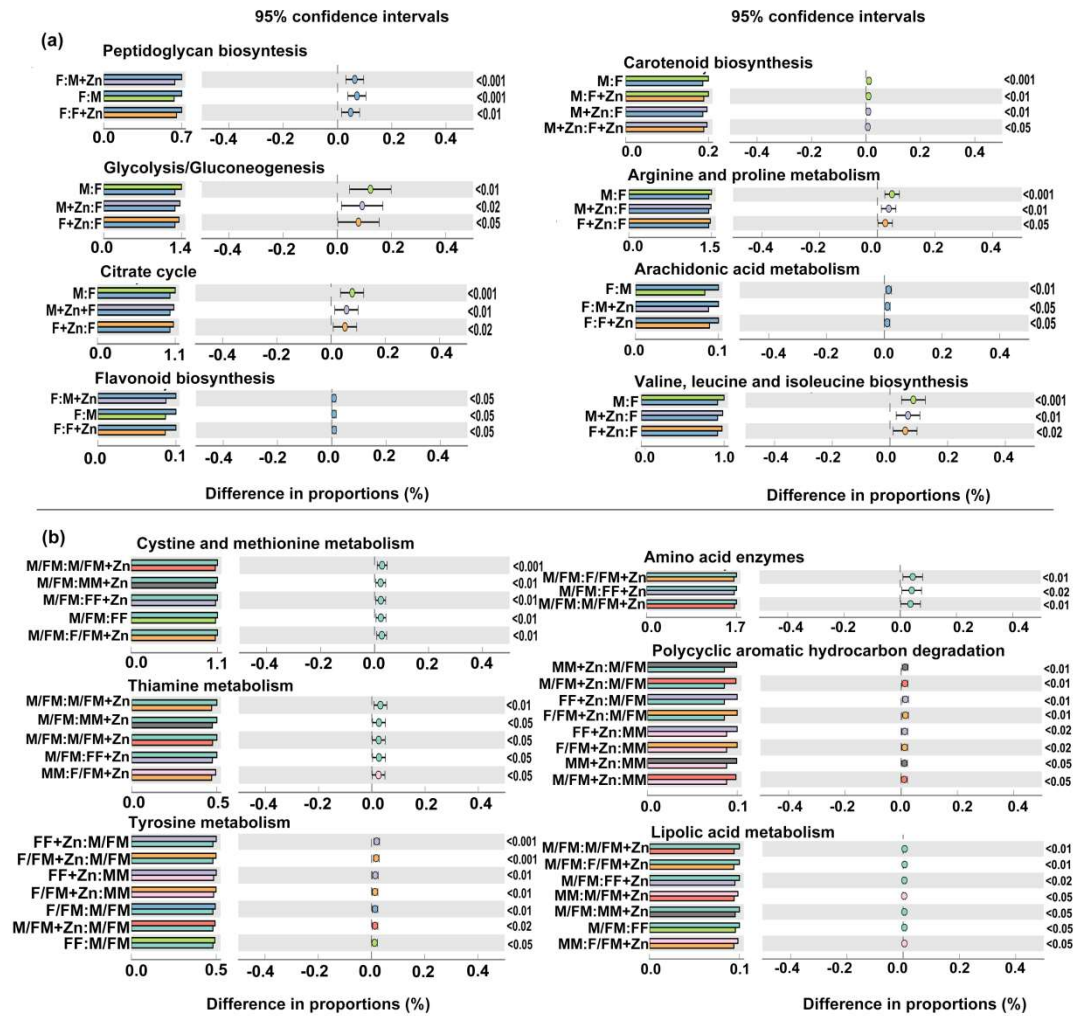




Fig. 6

