# The Abundance and Diversity of Larval and 

 Juvenile Fish in a Tropical EstuaryA. Tito de Morais<br>L. Tito de Morais<br>Hydrobiologie-UR 2C<br>Centre Orstom de Cayenne BP 165<br>97323 Cayenne Cedex<br>France


#### Abstract

The larval and juvenile fish of the Cayenne river estuary (French Guiana, South America) were sampled at two stations from June 1989 until October 1990. A total of 52,989 individuals from 59 species, some still incompletely identified, were collected. Three families, Engraulidae, Gobiidae and Sciaenidae, accounted for over $97 \%$ of the total number of juveniles. The analysis of data over this period showed low diversity, and a difference in diversity between the two sampling locations ( $\mathrm{H}^{\prime}=1.24$ and 1.68 ). The results conform to some theoretical models of abundance that suggest a relative equilibrium of juvenile assemblages. In contrast, the seasonal variations in diversity and abundance and the results of a correspondence analysis showed significant differences in species distribution and in their relative abundance at the two sampling locations at certain periods, mainly in the rainy season. Our study indicates that, in spite of an apparent stability, the year to year variation in salinity and freshwater inputs could affect juvenile recruitment of some species and induces modifications in the composition of larval and juvenile estuarine fish assemblages.


## Introduction

A number of studies deal with fish communities within estuaries and coastal lagoons (see reviews by Yanez-Arancibia 1985; Day 1989). Such systems are known as major nursery grounds for many aquatic species. Generally rich in food, they also provide protection against predators of larvae and juveniles, thus allowing both rapid growth and a low rate of mortality. However, little is known about abundance and diversity of ichthyoplankton in estuaries, and spatiotemporal variation in these assemblages is poorly understood.

Our study was carried out in the Cayenne River estuary in French Guiana. Estuaries may be considered "unpredictable environments" in which physical and chemical factors vary widely in space and time (Bruton 1989; Whitfield 1990). The goals of our study were to characterize this variation in the Cayenne estuary and to answer the following questions about ichthyoplankton assemblages in this environment:

Is the larval and juvenile fish assemblage a stable community within an unstable environment? If only biotic factors, such as interspecific competition, play a role in the composition of a community, such a community is at an equilibrium of species richness. Yet an equilibrium may never be reached if abiocic factors are predominant.

Can this assemblage be described by any of the species abundance models currently in use (Poole 1974; Magurran 1988)? Species richness, species diversity, and equitability provide partial quantita-
tive descriptions of assemblages, but they contain no biological information. Theoretical species abundance models attempt to describe the information gathered in a community (Poole 1974; Magurran 1988) and to explain the spatial distribution of the species within a given habitat and the sharing of the available resources (Magurran 1988).

To what extent can some of the observed variations be related to simple factors like rainfall and salinity?

## Materials and Methods

## Study Site

The study was conducted in the Cayenne River estuary (French Guiana, South America, $4^{\circ} 80^{\prime} \mathrm{N}$ $52^{\circ} 20^{\prime} \mathrm{W}$ ), an area 5 km in length and $1-1.5 \mathrm{~km}$ in width. There is considerable tidal influence (intertidal range up to 3 m ). It is an "homogenous" estuary (Yanez-Arancibia 1987) having almost no vertical salinity gradient. The rainy season occurs from December to June, with maximum rainfall generally observed in May and June.

## Sampling Methods

Two locations were chosen based on differences in physical characteristics (position, types of bottom, and water depth). The first (hereafter named middle), located in the middle of the estuary, is characterized by a mud bottom and an average depth of 5 m at low tide. The second location (called sandy), near the river mourh and 2.5 km
downstream from the first location, is characterized by a sandy bottom and an average depth of 1-2 m deep at low tide. Sampling, conducted with a $1,000-\mu \mathrm{m}$ mesh conical zooplankton net with a $75-\mathrm{cm}$ diameter mouth, was done once a month during spring ticles from June 1989 to October 1990. A 3-min sample thorizontal surface haul) was taken hourly during one 19 -h tidal cucle at each location and the sample was then preserved in 50 formalin. The volume of filtered water was measured using a mechanical flowmeter, and salinity readings were taken with an ATAGO retractometer.

## Identification

Larval and juvenile fish were categorized according to Hubbs ${ }^{(1943 \text { ) terminology fivenile period }}$ begins when the fins are fully differentiated). The captured fish ranged from 5 mm to about 70 mm in standard length. All juveniles were identified to species and larvae to farmitv.

The adult kevs used to make family-level identifications were from Eigenmann (1912), Puyo (1949), Le Bail et al. (1984a, bi, and Rojas-Beltran (1984. Species idencification (whenever possible) was done using Whitehead (1973) and Cervigon (1987) for Engraulidae. and Chao 11978) and Fischer (1978) for Sciaenidae.

## Characterization of Junenile Assemblages: Overall Data

Shannon-Wiener diversicy indices (H') (Margalef 1958 ) were calculated (total number of juveniles of each species in each site) and compared using a Student $t$-test (Hutcheson 1970).

Total number of juvenules caught at each location and the percentage of each species were calculated to plot abundance $\log$ curves, and fitted. if possible, to species abundance models (Poole 1974: Magurran 1988).

## Monthly Data

Monchly calculations of several diversivy indices were used to highlight seasonal variations. each index being more sensitive to a different component of diversity:

## Species Richness: $S$

This index gives only rough information and does not take into account the second component of diversitv, the relative abundance of different spectes. Richness is calculated as the number of species.

Berger-Parker's Dominance Index: $i=V_{m, n}$. $V$
This index is more sensitive to the abundance af the most common species and has the adran-
tages of being east to calculate and of giving an incuitive image of species distribution. $\mathrm{N}_{\text {min }}$ is the number of individuals of the most abundant species; $N$ is the total number of individuals.

## Shannon-Wimer's Diversity Index: $H^{\prime}=-\Sigma(p$, he $p$

This index is more sensitive to the presence of rare species, and thus to species richness. $p_{i}$ is the proportion of individuals of the ith species.

## Shannon Evenness Index: $E=H / \ln S$

This index shows the relation between the observed diversity and the maximum diversity (when all species are equallv abundanti.

Statistical tables in Magurran (1988) and Rohlf and Sokal (1981) were used in determining significance.

## Jlaenile avo Larrial Assemblages: Correspondevce Avalysis

The larval component being important in some samples, a correspondence analvsis ( $C . A$ ), using the program $A D D A D$ Association pour le Développement De l'Analyse de Donneess, sas done on the monthlv mean densities including juveniles and larvae from both locations imonthly densities of 55 taxa over 17 mol . This essentiallv descriptive multivariate method has the advantage of simplifying large data sets with hittie loss of information and identifving interrelations among variables. The analvsis produces composite factors: the first explains the maximum of variance (or inercia) in the daca set along a single axis, and all subsequent factors explain the maximum amount of the remaining variance. These new factors (or axes) allow plotting of the variabilitv of multidimensional data in a reduced space.

## Results <br> Jutentle Assemblages <br> Overall Data-Abundance and Diverszty

The list of different species collected, the total number of juveniles per species, and their relative abundance and ranking are given in Table 1. A total of 33.361 juvenules (and 19,629 larvae) from 59 species from both locanons were collected and sorted. Three families, Engraulidae, Gobiidae and Sciaenidae, comprised over $97 \%$ of the total number of juveniles.

Diversicy indices ( $\mathrm{H}^{\prime}$ ) calculated for the middle $1 \mathrm{H}^{\prime}=1.24$ ) and sandy ( $\mathrm{H}^{\prime}=1.68$ ) locations differed significantlv it $=29.14$ : with $\mathrm{df}=31,379$ : critical value at $10: 9.575$ ). As a rule, Shannon diversity indices vary from 1.5 to 3.5 in the field (Margalef 1972). The Cavenne River estuary is thus a low density habitat with respect to the num-

TABLE 1. List of the species collected at the middle and at the sandy locations. Number of individuals of a given species ( $N$ ), total of juveniles $\left(\mathrm{N}_{\mathrm{j}}\right)$, percent of a given species ( $\mathrm{N} / \mathrm{N}_{\mathrm{j}} \%$ ), and rank.

| Species (Juveniles) | Middle |  |  | Sandy |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | N | N/ ${ }^{1} \%$ | Rank | N | N/5:\% | Rank |
| Anchoviella lepidentostole | 10,539 | 67.688 | 1 | 7,106 | 39.942 | 1 |
| Anchoa spinifer | 1,980 | 12.717 | 2 | 1,347 | 7.571 | 3 |
| Gobilidae sp. 9 | 1,316 | 8.459 | 3 | 5,809 | 32.651 | 2 |
| Gobiidae sp. 1 | 487 | 3.128 | 4 | 1,173 | 6.593 | 4 |
| Stellifer rasirifer | 351 | 2.254 |  | 110 | 0.618 | 11 |
| Engraulidae sp. 4 | 194 | 1.246 | 6 | 270 | 1.518 | 7 |
| Odontognathus mucronatus | 123 | 0.790 | - | 18 | 0.101 | 19 |
| Engraulidae sp. 2 | 120 | 0.771 | 8 | 412 | 2.316 | 6 |
| Atherinidae | 113 | 0.726 | 9 | 133 | 0.748 | 9 |
| Gobiidae sp. 4 | 36 | 0.360 | 10 | 35 | 0.197 | 13 |
| Macrodon ancylodon | 49 | 0.315 | 11 | 29 | 0.163 | 14 |
| Gobiidae sp. 3 | $\pm 4$ | 0.283 | 12 | 53 | 0.298 | 12 |
| Colomesus psittacus | 32 | 0.206 | 13 | 146 | 0.821 | 8 |
| Micropogonias furnieri | 23 | 0.148 | 14 | 782 | 4.395 | 5 |
| Congridae | $\underline{9}$ | 0.135 | 15 | 15 | 0.084 | $\bigcirc 9$ |
| Isopisthus parvipinnis | 19 | 0.129 | 16 | 24 | 0.135 | 15 |
| Arius sp. 1 | 12 | 0.077 | 17 | 17 | 0.096 | 21 |
| Soleidae | 11 | 0.071 | 18 | 18 | 0.101 | 19 |
| Oligoplites saliens | 10 | 0.064 | 19 | 12 | 0.067 | 24 |
| Lycengraulis sp. | 8 | 0.051 | 20 | 14 | 0.079 | 23 |
| Cynoscion acoupa | 6 | 0.039 | 21 | 126 | 0.708 | 10 |
| Mugil curema | 6 | 0.039 | 21 | 25 | 0.141 | 15 |
| Belonidae | 5 | 0.032 | 23 | 22 | 0.124 | 17 |
| Cynoglossidae | 5 | 0.032 | 23 | 20 | 0.112 | 18 |
| Amphichtys cryptocentrus | 4 | 0.026 | 25 | 7 | 0.039 | 25 |
| Serranidae | 4 | 0.026 | 25 | 6 | 0.034 | 26 |
| Stellijer microps | 3 | 0.019 | 27 | 5 | 0.028 | 29 |
| Trachinotus cayennensis | 3 | 0.019 | 27 | 2 | 0.011 | 36 |
| Trichiurius lepturus | 3 | 0.019 | 27 | 0 | - | - |
| Bochidae | 2 | 0.013 | 30 | 4 | 0.022 | 33 |
| Murenosocidae | 2 | 0.013 | 30 | 2 | 0.011 | 36 |
| Sardinella sp. | 2 | 0.013 | 30 | 1 | 0.006 | 40 |
| Arius sp. 2 | $\stackrel{2}{2}$ | 0.013 | 30 | 1 | 0.006 | 40 |
| Selene vomer | 2 | 0.013 | 30 | 0 | - | - |
| Aspredinichtys flamentosus | 1 | 0.006 | 35 | 6 | 0.034 | 26 |
| Spheroides marmoratus | 1 | 0.006 | 35 | 5 | 0.028 | 29 |
| Unidentified/38 | 1 | 0.006 | 35 | 3 | 0.017 | 34 |
| Arius sp. 3 | 1 | 0.006 | 35 | 3 | 0.017 | 34 |
| Lonchurus lanceolatus | 1 | 0.006 | 35 | 1 | 0.006 | 40 |
| Anchoviella guianensis | 1 | 0.006 | 35 | 1 | 0.006 | 40 |
| Anchovia sumnamensis | 1 | 0.006 | 35 | 1 | 0.006 | 40 |
| Pimelodus blochii | 1 | 0.006 | 35 | 0 | - | - |
| Unidentified/42 | 1 | 0.006 | 35 | 0 | - | - |
| Scombridae | 1 | 0.006 | 35 | 0 | - | - |
| Unidentified/41 | 1 | 0.006 | 35 | 0 | - | - |
| Brachyplatystoma vaillantii | 1 | 0.006 | 35 | 0 | - | - |
| Nannostomus sp. | 1 | 0.006 | 35 | 0 | - | - |
| Chloroscombrus chrysurus | 0 | - | - |  | 0.034 | 26 |
| Aspredo aspredo | 0 | - | - | 5 | 0.028 | 29 |
| Pterengraulis athernoides | 0 | - | - | 5 | 0.028 | 29 |
| Arius sp. 5 | 0 | - | - | 2 | 0.011 | 36 |
| Polydactylus sp. | 0 | - | - | 2 | 0.011 | 36 |
| Unidentified/27 | 0 | - | - | , | 0.006 | 40 |
| Unidentified/40 | 0 | - | - | 1 | 0.006 | 40 |
| Aspredinidae sp. 3 | 0 | - | - | 1 | 0.006 | 40 |
| Arius sp. 4 | 0 | - | - | 1 | 0.006 | 40 |
| Caranx latus | 0 | - | - | 1 | 0.006 | 40 |
| Unidenuified/19 | 0 | - | - | 1 | 0.006 | 40 |
| Pleuronectidae | - | - | - | 1 | 0.006 | 40 |
| Total of juveniles ( Nj ) | 15,570 | 100 | - | 17,791 | 100 |  |
| Tocal of larvae | 13,313 | - | - | 6,316 | - | - |
| Larvae and juveniles | 28,883 | - | - | 24,107 | - | - |



Fig. i. Rank abundance plots of middle and sandv locauons.
ber of individuals collected. Fiftw-nine taxa were collected, though most were represented by only 1 or 2 individuals. These can be considered as "incidental". As the Shannon-Weaver index is particularly sensitive to the presence of rare species, this measure probably overestimated diversity in our study.

The abundance of each species is plotted on a logarithmic scale against the species rank. from the most abundant to the least abundant species (Fig. 1). It should be noted that abundance ranks for single species differed between locations isee also Table 1).

The abundance curves are obviously not straight lines (i.e. they do not fic a geomerric series). Models predict that such abundance curves occur in a situation in which species arrive at an unsaturaced habitat at regular intervais of time, and occupy fractions of the remaining niche hyperspace (Magurran 1988).

In Fig. 2 , the observed values are compared with the theoretical log-series model. This pattern would result of the intervals between the arrival of the species were random racher than regular (Magurran 1988). In both the sandy and the middle locations. the observed vaiues fitted the $\log$-sentes model iwith $\mathrm{df}=12$; middle: $\mathrm{c}^{2}=17.83$; sandy: $02=18.97$; crical vaiue at $50=21.033$.

When abundance is compared wath the log-nor-
(a) Middle

(b) Sancy

Fig. 2. Middle (a) and sandy (b) species abundance and logsenes distribution. The number of species observed in 13 abundance classes is plotted aganst the number of species predicted by the log-series model.
mal distribution, onlv abundance in the sandy location. fit at the $5 \%$ level with $\mathrm{dF}=10$; middle: c ? $=18.86$; sandy: $c=15.91$; critical value at $5 \%=$ 18.30).

The log-series model is more adapted to less diverse communities than is the log-normal model (Magurran 1988). In our case, the goodness-of-fic of abundance in the sandv location was best with the log-normal distribution and that of the middle abundance was best with the log-series model.

## Monthly Data-Seasonal Variations

The monthly mean densities of juveniles for the most common species are given in Table ?. Two species were highly dominant in the samples: $A n$ chonella lepidentostole a pelagic species restricted to estuaries (Rojas-Beltran 1986: Cervigon 1987) and Gobiidae sp.3. The most abundant species was at-
PABAE 2. Average density (per $1,000 \mathrm{~m}^{3}$ ) of the most common species.

| $\begin{aligned} & \text { Yeary } \\ & \text { Mendi: } \end{aligned}$ | $\xrightarrow{\text { 813 }}$ | $\stackrel{89}{17}$ | 89 <br> 088 | $\underset{\substack{89 \\ 09}}{ }$ | $\stackrel{19}{89}$ | $\stackrel{8!}{11}$ | 89 12 | ${ }^{90}$ | ${ }_{10}^{991}$ | ${ }_{0}^{19}$ | ${ }_{0}^{94}$ | ${ }^{(10)}$ | 9119 | ${ }_{191} 19$ | (10) | \% | !10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Midulle |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| a. Lepindentusule | 661 | 1,166 | 275 | 1,800 | 494 | 892 | 478 | 129 | 61 | 345 | 200 | 642 | 1,343 | 411 | 1,513 | 2,299 | 1,2:18 |
| d. spinijer | 313 | 110 | 37 | 19 | 347 | 29 | 55 | 13 | 63 | ${ }^{6}$ | 117 | 74 | 108 | 6.4 | 262 | 1,091 | 22.1 |
| A. furnieri | 0 | 5 | 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 2 | 0 | 0 |
| S. rastrifer | 2 | 52 | 245 | 26 | 54 | 7 | 8 | 1 | 1 | 1 | 1 | 1 | 0 | 1 | 0 | ${ }^{0}$ | 7 |
| Gobiidatesp. 2 | 5 | 3 | 21 | 15 | 17 | 39 | 91 | 248 | 100 | 0 | 38 | 0 | 2 | 1 | 5 | 2 | 16 |
| Gobiidat sp. 1 | 5 | 65 | 155 | 345 | 31 | 14 | 411 | 416 | 132 | 1 | 3 | 4 | 0 | 9 | 3 | 7 | 67 |
| Engraulid larvae | 3,10.4 | 1,007 | 405 | 393 | 562 | 47.1 | 967 | 1,288 | 500 | 339 | 620 | 1,512 | 871 | 951 | 579 | 8.44 | 435 |
| Sciaenicl larvate | 559 | 166 | 244 | 80 | 97 | 46 | 124 | 26 | 10 | 58 | 4 | 353 | 37 | 215 | 16.4 | 13 | 209 |
| Sandy |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| A. lepidentustole | 232 | 76.4 | 176 | 770 | 375 | 522 | 7.18 | 40 | 42 | 207 | 40 | 390 | 1,193 | 631 | 1,083 | 1,09.4 | 951 |
| A. spinifer | 106 | 35 | 26 | 6 | 121 | 2.4 | 18 | 1 | 1.1 | 3 | 25 | 55 | 701 | 15 | 411 | 248 | 40 |
| M. jurnieri | 542 | 194 | 125 | 13 | 1 | 1 | 10 | 31 | 3 | 4 | 2 | 16 | 16 | 12 | 7 | 0 | 16 |
| S. rastrifer | 0 | 5 | 51 | 58 | 7 | 1 | 1 | 0 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 7 |
| Gobiiidae sp. 2 | 47 | 19 | 3 | 81 | 28 | 133 | 72 | 414 | 516 | 3 | 55 | 27 | 2 | 2 | 10 | 4 | 75 |
| Cobiidae sp. 1 | 69 | 96 | 216 | 2,997 | 22 | 24 | 2,048 | 1,075 | 697 | 3 | 45 | 21 | 0 | 17 | 24 | 23 | 370 |
| Engraulid lavale | 3416 | 215 | 289 | 164 | 920 | 113 | 744 | 298 | 116 | 263 | 233 | 6.17 | 129 | 9-15 | 241 | 215 | 138 |
| Sciamid lavate | 415 | 119 | 316 | 133 | 193 | 33 | 111 | 30 | 14 | 25 | 15 | 132 | 39 | 75 | 11.5 | 11 | 158 |

ways A. lepidentostole except when Gobiidae sp. 2 predominated at the sandy location in August, September, and December 1989 and at both locations in January and February 1990. Of the other Engraulidae, only Anchoa spinifer, which is an euryhaline species (Whitehead 1973; Cervigon 1987), was well represented. Sciaenidae as a whole (of which the adult population is actively fished on a small scale by local Cayenne fishermen) dominated in June, July, and August 1989. The two main species collected were Micropogonias furnieri and Stellifer rastrifer.

Figure 3 gives the variations over time for several diversity indices. Species richness (the number of collected species) showed monthly variations between 9 and 17 at the middle location and from 10 to 22 at the sandy site. The number of species found at the sandy location was significantly higher (ANOVA on species richness: $\mathrm{df}=1.32$; $\mathrm{Fs}=$ 11.96 ; critical value at $5 \%: 4.15$ ).

All the other indices showed similar trends. The Shannon diversity index varied from 0.5 to 2 , and evenness varied from 0.2 (the clear dominance by a single species) to 0.7 (a more homogenous distribution of species). Indices calculated for the sandy location were generally close to or superior to those for the middle one, except in December 1989 and January and February 1990. Shannon indices $\left(\mathrm{H}^{\prime}\right)$ differed significantly between the two locations for these months (critical value at the $5 \%$ level $=1.96$ with $\mathrm{df}=\infty$; December 1989: $t=2.07$; January 1990: $t=6.56$; February 1990, $t=10.51$ ). These differences reflect the species distribution and the relative dominance of Gobiidae sp.2. The total number of species collected at the sandy location was high in January and February 1990, but Gobiidae sp. 2 dominated to such a degree that diversity remained low. Although Gobiidae sp. 2 also dominated at the middle location, other species were also abundant and the diversity indices were higher.

## Larval and Juvenile AssemblagesCorrespondence Analysis

A total of 19,629 larvae accounted for $37 \%$ of the total number of individuals. The distribution of larvae and juveniles of the three main families from the two sampling locations showed a consistently higher percentage of larvae at the middle location (Table 3).

The first three axes in the correspondence analysis accounted for $68 \%$ of the tocal inertia (respectively $34 \%, 19 \%$, and $15 \%$ ). They are retained hereafter to interpret the variance within the data. The most important contributions of species to total variance of each axis are given in Table 4. Go-
(a) Species richness

(c) Shannon index

months
(b) Berger-Parker index

(d) Shannon evenness

months

Fig. 3. Temporal distribuwon of diversity maices of ${ }^{\prime}$. ©) Shannon mdex. $H^{\prime}$, d) Shannon evenness. E .

TABLE 3. Larvae and juvenile distribution at the two sampling locations. $\mathrm{N}=$ total number; Percent $=100 \cdot \mathrm{~N} /(\mathrm{N}$ larvae $+\mathrm{i} \mathrm{V}$ juveniles).

|  | Middle |  |  | Sandv |  |
| :--- | ---: | :---: | :---: | :---: | :---: |
|  | N |  | Percent | N | Percent |
| Engraulidae larvae | 11,309 | 47 | 4,776 | 34 |  |
| Engraulidae juveniles | 19,843 | 53 | 9,156 | 66 |  |
| Sciaenidae larvae | 1,993 | 82 | 1,538 | 58 |  |
| Sciaenidae juveniles | 452 | 18 | 1,099 | 49 |  |
| Ariidae larvae | 11 | 42 | 9 | 4 |  |
| Ariidae juveniles | 15 | 58 | 54 | 96 |  |
| Total larvae | 13,313 | 50 | 6,316 | 38 |  |
| Total juveniles | 13,310 | 50 | 10,309 | 62 |  |

biidae sp. 2 contributed to $68 \%$ of the inertia of the first axis.

Contributions of stations to axes are also given in Table 4 (samples from the sandy location are indicated by $S$ and those from the middle location by M; 1989 by A and 1990 by B). For example, the samples of sandy location in June 1989 (SA6: $\mathrm{S}=$ sandy location; $A=1989 ; 6=$ June) contributed to $75 \%$ to the total variance of axis 2 .

Figure 4 shows the factorial plans 1 and 2, where seasonal differences between the two locations are particularly enhanced (codes as in Table 4). On axis 1 , there is a contrast between the sandy and middle samples in September (SA9 vs MLA9) and in December (SA12 vs MA12) 1989, and in January (SBl vs MB1) and in February (SB2 vs MB2) 1990. These four samples from the sandy location account for $72 \%$ of the variance of axis I (Table 4) and are the months with high densities of Gobiidae sp.2. On axis 2 there is a contrast between the sandy and middle locations in June (SA6), July (SA7), and August (SA8) 1989 (months with the highest numbers of M. furniert). M. fumieri, a benthic species (Isaac 1988), was clearly dominant in June 1989, having an important juvenile recruitment. This was not observed the following year in the Cavenne River (no similar contrast on axis 2 was observed during the same period as no M. fur-

TABLE 4. Results of correspondence analysis. Contributions of the main species to the variance of each axis (in percent) (i.e.. inertia (or variance) explained by species to total inertia of one axis). Contributions of the main samples to the variance of each axis (in percent) (i.e., variance explained by samples to total variance of one axis). Samples and species are independent.

| Axes | Species | Percent | Samples | Percent |
| :--- | :--- | :--- | :--- | :--- |
| Axis 1 | Gobiidae sp. 2 | 68 | SA9 | 30 |
|  |  |  | SA12 | 14 |
|  |  |  | SB1 | 15 |
|  |  |  | SB2 | 13 |
| Axis 2 | H. furnieri | 59 | SA6 | 75 |
| Axis 3 | Engrauhid larva | 50 | MA6 | 18 |
|  | A. lepidentostole | 19 | SB6 | 11 |



Fig. 4. Correspondence analysis performed on both stations: relative position of the monthly samples (total number of larvae and juveniles captured per $1,000 \mathrm{~m}^{3}$ in each month) on the axes 1 and 9. Codes: M (middle), S (sandy), A (1989), B (1990), 1-12 (months) (see text).
nieri were present in 1990). Precipitation during the rainy season was much greater in 1989, with much higher freshwater inputs than in 1990. Observed salinities were thus lower in 1989 than in 1990 (Table 5). In contrast to the Cayenne River, recruitment of $M$. furnieri was observed in the Sinnamary River estuary ( 100 km east of the Cayenne River) in 1990 (unpublished data). The Sinnamary River is larger than the Cayenne River and has a much greater freshwater flow, so freshwater input was considerable even in the less severe rainy season in 1990. High-level recruitment of juveniles of M. furnieri and other Sciaenidae (e.g., S. rastrifer) seemed to be dependent on low salinity water (see periods of low salinities in Table 5).

Two taxa contribute to the variance of axis $3, A$. lepidentostole and unidentified Engraulidae larvae (the majority are probably A. lepidentostole). The peak of abundance of larvae was in June 1989 (MA6), with smaller peaks in January (MB1) and

TABLE 5 . Xonthly maximum of salinity (parts per thousand) at muddle locadion isources our daca and Lhomme personal communication.

|  | 1999 | 199 |
| :--- | :---: | :---: |
| January | 12 | 10 |
| Februar | 18 | 16 |
| March | 5 | 15 |
| Aprl | 9 | 10 |
| Dav | 10 | 14 |
| June | 3 | 29 |
| Julv | 24 | 27 |
| Augusi | 30 | 30 |
| September | 31 | 32 |
| Ocrober | 27 | 34 |
| Vovember | 39 | - |
| December | 20 | - |

Mav (MB5) 1990 at the middle location (top of Fig. 5). These are periods of heavy rains in the wet season.

## Conclusions

Result.s from Shannon diversity indices and from models of abundance suggest a relative equilibrium of species at the two locations. The sandy location seemed to be more diverse and this was also confirmed bv the abundance models. In a different tropical svstem, Pinto (1988) studied an adult fish community in a relativelv stable environment (a Philippines mangrove svstemi. He showed that species abundance conformed to the log-normal model except when an important environmental disturbance (a typhoon) disrupted the community.

Correspondence analvsis and seasonal variacions of diversicy indices showed significant differences in species distribution and in the relative abundance of species at the two sampling locations at certain periods. Our analvis indicates that salinity and freshwater inputs plaved an important role.

Based on a studv of a tropical brackish-water lagoon in West Africa, Abaret and Ecoutin (1990) showed that seasonal changes in the composition and the structure of fish communities are induced mainiy by abundance and distribution of freshwater inputs. Similarly, different field studies in South African estuaries have shown effects of the occurrence and severity of floods on ichthyofaunal communities (Plumstead 1990).

Two types of hypocheses, deterministic and stochastic, have been considered to explain the regulation of fish assembiages (Grossman et al. 1982; Herbold 1984; Rahel et al. 1984: Yant et al. 1984). The former type emphasizes the role of biotic factors like interspecific competition that are more favorable to the settlement of a stable assemblage. These factors are the basis of the biological explanations generallv applied to abundance models. but these are sratistical rather than biological mod-


Fig. 5. Correspondence analvsis performed on boch stations: relanve position of the monthlv samples itoral number of larvae and juventes captured per $1,000 \mathrm{~m}^{*}$ in each month on the axes 3 and 3. Codes: $M$ (middler, S (sandy). A 19999, B 11990). 1-12 (months: see rext).
els (Magurran 1988). The second type of hypothesis emphasizes the role of physical and chemical factors that are seldom stable enough to allow the settlement of a stable community.

Compared to other environments, estuaries are known as unpredictable environments where unstable phisical conditions do not favor stable communities. This was shown for adult fish assemblages by Whitfield (1990), and seems to be even more important for assemblages of juvenile fish: Schlosser (1985) showed that physical i.e., stochastic! factors are important to riverine juvenile fish abundance. species richness, and species composition; older age classes are more influenced by biological (i.e.. deterministic) factors. Freeman et al. (1988) obtained similar results when changes in stream flow apparencly affected voung-of-the-vear recruitment. Our study had indicated a relative stability over all the period, but year-tovear variations in the intensity of the rainy season and the level of freshwater inputs may affect juvenile recruitment of some spectes and lead to modifications in the composition of larval and juvenile fish assemblages.

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$A$ PPENDIX. Taxonomic list of the collected species. Six unidentined species are not listed. See Table li;

| Order | Famaly | Speres: |
| :---: | :---: | :---: |
| Anguilliformes | Congrdae <br> Murenosocidae | One unidentified species One unidentified species |
| Cluperformes | Clupeidae <br> Engraulidae | Sardinella sp . <br> Odontognathus mucromatus Anchora spinujer <br> Anchotia surinaznenses <br> thehovrila greanensis <br> Anchovella lebrdentostole <br> Lucengraudes sp. <br> Pterngraiais athmonode: <br> Engraulidue sp. 2 <br> Engraulidae sp. $\ddagger$ |
| Cipriniformes | Lebiasinidae | Vannosiomus sp. |
| Silurformes | tridae | Arius sp. 1 <br> traus sp. ${ }^{3}$ <br> Artus sp. 3 <br> trius $s p$ ! <br> Arus sp 3 |
|  | Aspredinidae | A.predinthes filamentoms Aspredo asprecio Aspredinidae ip. 3 |
|  | Prmelodidae | Emahyphatystoma vallanoa Pamelodus blechi |
| Batrachoidiformes | Batrachoididae | Amphichtys apptocnerus |
| Atheriniformes | Belonidae Atherinudae | One undenafied species One unidentified species |
| Perciformes | Sertandae Carangldae | One unidentified species <br> Caranx latus <br> Chlorosomernes chrysurus <br> Oligophtes satiens <br> Selpne vomer <br> Trachinous cavennensts |
|  | Sciaenidae | Cinoscoon acouta <br> Isopesthus parvipnnais <br> Lonchumbs ianceviatus <br> Vacrodon anculodon <br> Macropogomas jumem <br> Stellifer micreps <br> Stellifer rastrifer |
|  | Muglidae <br> Polvnemidae <br> Gubiidae | Hugzi curema <br> Poivdactylus sp. <br> Gobuidae sp. 1 <br> Gobiidae sp. : <br> Gobiidae sp. 3 <br> Gobiidae sp. 4 |
|  | Trichiurndae Scombridae | Trichintus lopturas One undenafied species |
| Pleuronecuiformes | Borhidae Pleuronectidae Soleidae Cymoglossidae | One unidenufied species One undentified spectes Probably Achima achumes One untdenufied species |
| Terraondontformes | Terraodonudae | Colomesus patacus Spherodies marmoratus |

