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The Abundance and Diversity of Larval and Juvenile Fish in a Tropical Estuary

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ABSTRACT: The larval and juvenile fish of the Cayenne river estuary (French Guiana, South America) were sampled at two stations from June 1989 until October 1990. A total of 52,989 individuals from 59 species, some still incompletely identified, were collected. Three families, Engraulidae, Gobiidae and Sciaenidae, accounted for over 97% of the total number of juveniles. The analysis of data over this period showed low diversity, and a difference in diversity between the two sampling locations (H' = 1.24 and 1.68). The results conform to some theoretical models of abundance that suggest a relative equilibrium of juvenile assemblages. In contrast, the seasonal variations in diversity and abundance and the results of a correspondence analysis showed significant differences in species distribution and in their relative abundance at the two sampling locations at certain periods, mainly in the rainy season. Our study indicates that, in spite of an apparent stability, the year to year variation in salinity and freshwater inputs could affect juvenile recruitment of some species and induces modifications in the composition of larval and juvenile estuarine fish assemblages.

Introduction

A number of studies deal with fish communities within estuaries and coastal lagoons (see reviews by Yanez-Arancibia 1985; Day 1989). Such systems are known as major nursery grounds for many aquatic species. Generally rich in food, they also provide protection against predators of larvae and juveniles, thus allowing both rapid growth and a low rate of mortality. However, little is known about abundance and diversity of ichthyoplankton in estuaries, and spatiotemporal variation in these assemblages is poorly understood.

Our study was carried out in the Cayenne River estuary in French Guiana. Estuaries may be considered "unpredictable environments" in which physical and chemical factors vary widely in space and time (Bruton 1989; Whitfield 1990). The goals of our study were to characterize this variation in the Cayenne estuary and to answer the following questions about ichthyoplankton assemblages in this environment:

Is the larval and juvenile fish assemblage a stable community within an unstable environment? If only biotic factors, such as interspecific competition, play a role in the composition of a community, such a community is at an equilibrium of species richness. Yet an equilibrium may never be reached if abiotic factors are predominant.

Can this assemblage be described by any of the species abundance models currently in use (Poole 1974; Magurran 1988)? Species richness, species diversity, and equitability provide partial quantita-

tive descriptions of assemblages, but they contain no biological information. Theoretical species abundance models attempt to describe the information gathered in a community (Poole 1974; Magurran 1988) and to explain the spatial distribution of the species within a given habitat and the sharing of the available resources (Magurran 1988).

To what extent can some of the observed variations be related to simple factors like rainfall and salinity?

Materials and Methods

STUDY SITE

The study was conducted in the Cayenne River estuary (French Guiana, South America, 4°80'N 52°20'W), an area 5 km in length and 1–1.5 km in width. There is considerable tidal influence (intertidal range up to 3 m). It is an "homogenous" estuary (Yanez-Arancibia 1987) having almost no vertical salinity gradient. The rainy season occurs from December to June, with maximum rainfall generally observed in May and June.

SAMPLING METHODS

Two locations were chosen based on differences in physical characteristics (position, types of bottom, and water depth). The first (hereafter named middle), located in the middle of the estuary, is characterized by a mud bottom and an average depth of 5 m at low tide. The second location (called sandy), near the river mouth and 2.5 km

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downstream from the first location, is characterized by a sandy bottom and an average depth of 1-2 m deep at low tide. Sampling, conducted with a 1,000- μ m mesh conical zooplankton net with a 75-cm diameter mouth, was done once a month during spring tides from June 1989 to October 1990. A 3-min sample (horizontal surface haul) was taken hourly during one 12-h tidal cycle at each location and the sample was then preserved in 5% formalin. The volume of filtered water was measured using a mechanical flowmeter, and salinity readings were taken with an ATAGO refractometer.

IDENTIFICATION

Larval and juvenile fish were categorized according to Hubbs' (1943) terminology (juvenile period begins when the fins are fully differentiated). The captured fish ranged from 5 mm to about 70 mm in standard length. All juveniles were identified to species and larvae to family.

The adult keys used to make family-level identifications were from Eigenmann (1912), Puyo (1949), Le Bail et al. (1984a, b), and Rojas-Beltran (1984). Species identification (whenever possible) was done using Whitehead (1973) and Cervigon (1987) for Engraulidae, and Chao (1978) and Fischer (1978) for Sciaenidae.

CHARACTERIZATION OF JUVENILE ASSEMBLAGES: OVERALL DATA

Shannon-Wiener diversity indices (H') (Margalef 1958) were calculated (total number of juveniles of each species in each site) and compared using a Student 4-test (Hutcheson 1970).

Total number of juveniles caught at each location and the percentage of each species were calculated to plot abundance log curves, and fitted, if possible, to species abundance models (Poole 1974; Magurran 1988).

MONTHLY DATA

Monthly calculations of several diversity indices were used to highlight seasonal variations. each index being more sensitive to a different component of diversity:

Species Richness: S

This index gives only rough information and does not take into account the second component of diversity, the relative abundance of different species. Richness is calculated as the number of species.

Berger-Parker's Dominance Index: $d = N_{max} N$

This index is more sensitive to the abundance of the most common species and has the advan-

tages of being easy to calculate and of giving an intuitive image of species distribution. N_{max} is the number of individuals of the most abundant species; N is the total number of individuals.

Shannon-Wiener's Diversity Index: $H' = -\Sigma (p_i ln p_i)$

This index is more sensitive to the presence of rare species, and thus to species richness. p_i is the proportion of individuals of the ith species.

Shannon Evenness Index: E = H'/ln S

This index shows the relation between the observed diversity and the maximum diversity (when all species are equally abundant).

Statistical tables in Magurran (1988) and Rohlf and Sokal (1981) were used in determining significance.

JUVENILE AND LARVAL ASSEMBLAGES: CORRESPONDENCE ANALYSIS

The larval component being important in some samples, a correspondence analysis (CA), using the program ADDAD (Association pour le Développement De l'Analyse de Données), was done on the monthly mean densities including juveniles and larvae from both locations (monthly densities of 55 taxa over 17 mo). This essentially descriptive multivariate method has the advantage of simplifving large data sets with little loss of information and identifying interrelations among variables. The analysis produces composite factors: the first explains the maximum of variance (or inertia) in the data set along a single axis, and all subsequent factors explain the maximum amount of the remaining variance. These new factors (or axes) allow plotting of the variability of multidimensional data in a reduced space.

Results

JUVENILE ASSEMBLAGES

Overall Data-Abundance and Diversity

The list of different species collected, the total number of juveniles per species, and their relative abundance and ranking are given in Table 1. A total of 33.361 juveniles (and 19,629 larvae) from 59 species from both locations were collected and sorted. Three families, Engraulidae, Gobiidae and Sciaenidae, comprised over 97% of the total number of juveniles.

Diversity indices (H') calculated for the middle (H' = 1.24) and sandy (H' = 1.68) locations differed significantly (t = 29.14; with df = 31,379; critical value at 1%; 2.576). As a rule, Shannon diversity indices vary from 1.5 to 3.5 in the field (Margalef 1972). The Cavenne River estuary is thus a low density habitat with respect to the num-

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TABLE 1. List of the species collected at the middle and at the sandy locations. Number of individuals of a given species (N), total of juveniles (N_j), percent of a given species (N/N_j %), and rank.

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		Middle			Sandy	
Species (Juveniles)	N	N/N, %	Rank	N	N/N, %	Rank
Anchoviella lepidentostole	10,539	67.688	1	7,106	39.942	
Anchoa spinifer	1,980	12.717	2	1,347	7.571	ŝ
Gobiidae sp. 2	1,316	8.452	3	5,809	32.651	2
Gobiidae sp. 1	487	3.128	4	1,173	6.593	4
Stellifer rastrifer	351	2.254	С	110	0.618	11
Engraulidae sp. 4	194	1.246	6	270	1.518	7
Odontognathus mucronatus	123	0.790	-	18	0.101	19
Engraulidae sp. 2	120	0.771	8	412	2.316	6
Atherinidae	113	0.726	9	133	0.748	9
Gobiidae sp. 4	56	0.360	10	35	0.197	13
Macrodon ancylodon	49	0.315	11	29	0.163	14
Gobiidae sp. 3	-14	0.283	12	53	0.298	12
Colomesus psittacus	32	0.206	13	146	0.821	8
Micropogonias furnieri	23	0.148	14	782	4.395	5
Congridae	21	0.135	15	15	0.084	22
Isopisthus parmpinnis	19	0.122	16	24	0.135	16
Arius sp. 1	12	0.077	17	17	0.096	21
Soleidae	11	0.071	18	18	0.101	19
Oligophies saliens	10	0.064	19	12	0.067	24
Lycengrauus sp.	8	0.051	20	14	0.079	23
Cynoscion acoupa	6	0.039	21	126	0.708	10
Relevides	6	0.039	21	25	0.141	15
Crmoglossidae	5	0.032	23	22	0.124	17
Amphichtus muttoremtmus	5	0.032	23	20	0.112	18
Serrapidae	4	0.020	25	1	0.039	25
Stallifer microhe	4	0.020	25	5	0.034	26
Trachinotus cavennensis	3	0.019	41 97	2	0.028	29
Trichingins labrarys	3	0.019	±7 97	2	0.011	30
Borbidae	2	0.019	30	U d	0.099	22
Murenosocidae		0.013	30	4	0.022	36
Sardinella sp.	- 2	0.013	30	2	0.011	40
Arius sp. 2	2	0.013	30	1	0.000	40
Selene vomer	2	0.013	30	Ô	0.000	40
Aspredinichtys filamentosus	1	0.006	35	6 6	0.034	26
Spheroides marmoratus	1	0.006	35	5	0.028	29
Unidentified/38	1	0.006	35	3	0.017	34
Arius sp. 3	1	0.006	35	3	0.017	34
Lonchurus lanceolatus	1	0.006	35	1	0.006	40
Anchoviella guianensis	1	0.006	35	1	0.006	40
Anchovia surinamensis	1	0.006	35	1	0.006	40
Pimelodus blochii	1	0.006	35	0	_	
Unidentified/42	1	0.006	35	0		
Scombridae	1	0.006	35	0		
Unidentified/41	1	0.006	35	0		
Brachyplatystoma vaillantii	1	0.006	35	0		
Nannostomus sp.	1	0.006	35	0		
Chloroscombrus chrysurus	0	-		6	0.034	26
Aspredo aspredo	0			5	0.028	29
Pterengraulis atherinoides	0	—	بهمي.	5	0.028	29
Arius sp. 5	0			2	0.011	36
Polydactylus sp.	0	-		2	0.011	36
Unidentified/27	0			1	0.006	40
Unidendined/40	U			1	0.006	40
Arice sp. 3	U			1	0.006	40
Carany Latur	U			1	0.006	40
Linidentified /19	0			1	0.000	40
Pleuronectidae	0			1	0.000	40
Total of inveniles (Ni)	15 570	100		17 701	100	
Terel - Church	10,010	100		11,131	100	
Lotal of larvae	13,313			6,316		
Larvae and juveniles	28,883			24,107		

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Fig. 1. Rank abundance plots of middle and sandy locations.

ber of individuals collected. Fifty-nine taxa were collected, though most were represented by only 1 or 2 individuals. These can be considered as "incidental". As the Shannon-Weaver index is particularly sensitive to the presence of rare species, this measure probably overestimated diversity in our study.

The abundance of each species is plotted on a logarithmic scale against the species' rank. from the most abundant to the least abundant species (Fig. 1). It should be noted that abundance ranks for single species differed between locations (see also Table 1).

The abundance curves are obviously not straight lines (i.e., they do not fit a geometric series). Models predict that such abundance curves occur in a situation in which species arrive at an unsaturated habitat at regular intervals of time, and occupy fractions of the remaining niche hyperspace (Magurran 1988).

In Fig. 2, the observed values are compared with the theoretical log-series model. This pattern would result if the intervals between the arrival of the species were random rather than regular (Magurran 1988). In both the sandy and the middle locations, the observed values fitted the log-series model (with df = 12; middle: $c_2 = 17.83$; sandy: $c_2 = 18.97$; critical value at 5% = 21.03).

When abundance is compared with the log-nor-



Fig. 2. Middle (a) and sandy (b) species abundance and logseries distribution. The number of species observed in 13 abundance classes is plotted against the number of species predicted by the log-series model.

mal distribution, only abundance in the sandy location fit at the 5% level (with df = 10; middle: c2 = 18.86; sandy: c2 = 15.21; critical value at 5% = 18.30).

The log-series model is more adapted to less diverse communities than is the log-normal model (Magurran 1988). In our case, the goodness-of-fit of abundance in the sandy location was best with the log-normal distribution and that of the middle abundance was best with the log-series model.

MONTHLY DATA-SEASONAL VARIATIONS

The monthly mean densities of juveniles for the most common species are given in Table 2. Two species were highly dominant in the samples: Anchounella lepidentostole, a pelagic species restricted to estuaries (Rojas-Beltran 1986: Cervigon 1987) and Gobiidae sp.2. The most abundant species was al-

	06 10		1,2:18	224	0	7	16	67	435	269		951	40	16	7	75	370	138	158
	06 60		2,299	1,001	0	0	54	7	8-1-4	13		1,094	248	0	0	Ļ	53 53	215	11
	06 89		1,513	202	5 1	0	υ.	. .)	579	164		1,083	411	7	0	10	장	249	115
	90 02		411	64	0	-	-	6	951	215		631	15	51	0	51	17	9-15	75
	90 90		1,343	108	0	0	24	0	871	37		1,193	701	16	0	54	0	129	39
	90 02		642	74	0	1	0	Ψ.	1,512	353		390	55	16	0	27	21	6.17	132
	90 0-1		200	117	51	-	38	6	620	1-1-		40	25	¢1	0	55	-15 -	233	15
	90 50		3-15	6	0	_	0	-	339	58		207	ŝ	ŀ	21	ŝ	ŝ	263	25
	90 20		61	63	0		100	132	500	10		21-	Ŀ	ŝ	າ	516	697	116	14
	06 10	Middle	129	13	0	-	248	416	1,288	26	Sandy	40		31	0	414	1,075	298	30
	89 12		478	55	0	8	16	411	967	12-1		7-18	18	10	-	72	2,048	744	111
n species	68 11		892	29	0	7	39	14	17년	46		522	2:1			133	24	113	33
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	60 00		1,800	19	0	26	15	345	393	80		770	9	13	58	81	2,997	164	133
	08 80		275	37	15	245	21	155	405	244		176	26	125	51	3	216	289	316
ווינו וסיון	88) 07		1,166	110	5	52	ŝ	65	1,007	166		F97	35	661	5	19	96	215	611
(11611211) 2	68 90		661	313	0	54	ŝ	5	3,104	559		232	106	5-12	0	-17	69	346	415
BUDAU - 7 JUDI	Yeary Month:		A. lepidentostole	A. spinifer	M. Jumieri	S. rastrifer	Gobiidae sp. 2	Gobiidae sp. 1	Engraulid larvae	Scinenid larvae		A. lepidentostole	A. spinifer	AI. furnieri	S. rustrifer	Gobiidae sp. 2	Gobiidae sp. 1	Engraulid larvae	Schenid larvae

ways A. lepidentostole except when Gobiidae sp.2 predominated at the sandy location in August, September, and December 1989 and at both locations in January and February 1990. Of the other Engraulidae, only Anchoa spinifer, which is an euryhaline species (Whitehead 1973; Cervigon 1987), was well represented. Sciaenidae as a whole (of which the adult population is actively fished on a small scale by local Cayenne fishermen) dominated in June, July, and August 1989. The two main species collected were Micropogonias furnieri and Stellifer rastrifer.

Figure 3 gives the variations over time for several diversity indices. Species richness (the number of collected species) showed monthly variations between 9 and 17 at the middle location and from 10 to 22 at the sandy site. The number of species found at the sandy location was significantly higher (ANOVA on species richness: df = 1.32; Fs = 11.96; critical value at 5%: 4.15).

All the other indices showed similar trends. The Shannon diversity index varied from 0.5 to 2, and evenness varied from 0.2 (the clear dominance by a single species) to 0.7 (a more homogenous distribution of species). Indices calculated for the sandy location were generally close to or superior to those for the middle one, except in December 1989 and January and February 1990. Shannon indices (H') differed significantly between the two locations for these months (critical value at the 5% level = 1.96 with df = ∞ ; December 1989: t = 2.07; January 1990: t = 6.56; February 1990, t = 10.51). These differences reflect the species distribution and the relative dominance of Gobiidae sp.2. The total number of species collected at the sandy location was high in January and February 1990, but Gobiidae sp.2 dominated to such a degree that diversity remained low. Although Gobiidae sp.2 also dominated at the middle location, other species were also abundant and the diversity indices were higher.

A total of 19,629 larvae accounted for 37% of the total number of individuals. The distribution of larvae and juveniles of the three main families from the two sampling locations showed a consistently higher percentage of larvae at the middle location (Table 3).

The first three axes in the correspondence analysis accounted for 68% of the total inertia (respectively 34%, 19%, and 15%). They are retained hereafter to interpret the variance within the data. The most important contributions of species to total variance of each axis are given in Table 4. Go-

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Fig. 3. Temporal distribution of diversity indices of the juvenile fish assemblage. (a) Species richness, S. (b) Berger-Parker index. I_{c} d; (c) Shannon index, H': (d) Shannon evenness, E.

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TABLE 3. Larvae and juvenile distribution at the two sampling locations. N = total number; Percent = $100 \cdot N/(N \text{ larvae} + N \text{ juveniles})$.

	Mid	dle	Sandv			
	N	Percent	N	Percent		
Engraulidae larvae	11,309	47	4,776	34		
Engraulidae juveniles	12,843	53	9,156	66		
Sciaenidae larvae	1,993	82	1,538	58		
Sciaenidae juveniles	452	18	1,099	42		
Ariidae larvae	11	42	2	4		
Ariidae juveniles	15	58	54	96		
Total larvae	13,313	50	6,316	38		
Total juveniles	13,310	50	10,309	62		

biidae sp.2 contributed to 68% of the inertia of the first axis.

Contributions of stations to axes are also given in Table 4 (samples from the sandy location are indicated by S and those from the middle location by M; 1989 by A and 1990 by B). For example, the samples of sandy location in June 1989 (SA6: S = sandy location; A = 1989; 6 = June) contributed to 75% to the total variance of axis 2.

Figure 4 shows the factorial plans 1 and 2, where seasonal differences between the two locations are particularly enhanced (codes as in Table 4). On axis 1, there is a contrast between the sandy and middle samples in September (SA9 vs MA9) and in December (SA12 vs MA12) 1989, and in January (SB1 vs MB1) and in February (SB2 vs MB2) 1990. These four samples from the sandy location account for 72% of the variance of axis 1 (Table 4) and are the months with high densities of Gobiidae sp.2. On axis 2 there is a contrast between the sandy and middle locations in June (SA6), July (SA7), and August (SA8) 1989 (months with the highest numbers of M. furnieri). M. furnieri, a benthic species (Isaac 1988), was clearly dominant in June 1989, having an important juvenile recruitment. This was not observed the following year in the Cavenne River (no similar contrast on axis 2 was observed during the same period as no M. fur-

TABLE 4. Results of correspondence analysis. Contributions of the main species to the variance of each axis (in percent) (i.e., inertia (or variance) explained by species to total inertia of one axis). Contributions of the main samples to the variance of each axis (in percent) (i.e., variance explained by samples to total variance of one axis). Samples and species are independent.

Axes	Species	Percent	Samples	Percent
Axis 1	Gobiidae sp. 2	68	SA9	30
<i></i>			SA12	14
			SB1	15
			SB2	13
Axis 2	M. furnieri	59	SA6	75
Axis 3	Engraulid larvae	50	MA6	18
	A. lepidentostole	19	SB6	11



Fig. 4. Correspondence analysis performed on both stations: relative position of the monthly samples (total number of larvae and juveniles captured per 1,000 m³ in each month) on the axes 1 and 2. Codes: M (middle), S (sandy), A (1989), B (1990), 1-12 (months) (see text).

nieri were present in 1990). Precipitation during the rainy season was much greater in 1989, with much higher freshwater inputs than in 1990. Observed salinities were thus lower in 1989 than in 1990 (Table 5). In contrast to the Cayenne River, recruitment of *M. furnieri* was observed in the Sinnamary River estuary (100 km east of the Cayenne River) in 1990 (unpublished data). The Sinnamary River is larger than the Cayenne River and has a much greater freshwater flow, so freshwater input was considerable even in the less severe rainy season in 1990. High-level recruitment of juveniles of *M. furnieri* and other Sciaenidae (e.g., *S. rastrifer*) seemed to be dependent on low salinity water (see periods of low salinities in Table 5).

Two taxa contribute to the variance of axis 3, A. *lepidentostole* and unidentified Engraulidae larvae (the majority are probably A. *lepidentostole*). The peak of abundance of larvae was in June 1989 (MA6), with smaller peaks in January (MB1) and

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TABLE 5. Monthly maximum of salinity (parts per thousand) at middle location (sources our data and Lhomme personal communication).

	1986	1990
January	12	10
February	18	16
March	S	15
April	9	10
May	10	14
lune	8	25
July	24	27
August	30	30
September	31	32
October	27	34
November	32	
December	20	

May (MB5) 1990 at the middle location (top of Fig. 5). These are periods of heavy rains in the wet season.

Conclusions

Results from Shannon diversity indices and from models of abundance suggest a relative equilibrium of species at the two locations. The sandy location seemed to be more diverse and this was also confirmed by the abundance models. In a different tropical system, Pinto (1988) studied an adult fish community in a relatively stable environment (a Philippines mangrove system). He showed that species abundance conformed to the log-normal model except when an important environmental disturbance (a typhoon) disrupted the community.

Correspondence analysis and seasonal variations of diversity indices showed significant differences in species distribution and in the relative abundance of species at the two sampling locations at certain periods. Our analysis indicates that salinity and freshwater inputs played an important role.

Based on a study of a tropical brackish-water lagoon in West Africa, Albaret and Ecoutin (1990) showed that seasonal changes in the composition and the structure of fish communities are induced mainly by abundance and distribution of freshwater inputs. Similarly, different field studies in South African estuaries have shown effects of the occurrence and severity of floods on ichthyofaunal communities (Plumstead 1990).

Two types of hypotheses, deterministic and stochastic, have been considered to explain the regulation of fish assemblages (Grossman et al. 1982; Herbold 1984; Rahel et al. 1984; Yant et al. 1984). The former type emphasizes the role of biotic factors like interspecific competition that are more favorable to the settlement of a stable assemblage. These factors are the basis of the biological explanations generally applied to abundance models, but these are statistical rather than biological mod-



Fig. 5. Correspondence analysis performed on both stations: relative position of the monthly samples (total number of larvae and juveniles captured per 1,000 m² in each month) on the axes 2 and 3. Codes: M (middle), 5 (sandy), A (1989), B (1990), 1–12 (months) (see text).

els (Magurran 1988). The second type of hypothesis emphasizes the role of physical and chemical factors that are seldom stable enough to allow the settlement of a stable community.

Compared to other environments, estuaries are known as unpredictable environments where unstable physical conditions do not favor stable communities. This was shown for adult fish assemblages by Whitfield (1990), and seems to be even more important for assemblages of juvenile fish: Schlosser (1985) showed that physical (i.e., stochastic) factors are important to riverine ;uvenile fish abundance, species richness, and species composition; older age classes are more influenced by biological (i.e., deterministic) factors. Freeman et al. (1988) obtained similar results when changes in stream flow apparently affected young-of-the-year recruitment. Our study had indicated a relative stability over all the period, but year-to-year variations in the intensity of the rainy season and the level of freshwater inputs may affect juvenile recruitment of some species and lead to modifications in the composition of larval and juvenile fish assemblages.

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A.

Juvenile Fish Assemblages

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Order	Family	Species
Anguilliformes	Congridae Murenosocidae	One unidentified species One unidentified species
Clupeiformes	Clupeidae	Sardinella sp. Odontognathus mucronatus
	Engraulidae	Anchoa spinifer Anchovia surinamensis Anchoviella sevianensis
		Anchomella lepidentostole Lycengraulis sp. Pterengraulis athermoides Engraulidae sp. 2 Engraulidae sp. 4
Cypriniformes	Lebiasinidae	Nannostomus sp.
Siluriformes	Arridae	Arius sp. 1 Arius sp. 2 Arius sp. 3 Arius sp. 4 Arius sp. 5
	Aspredinidae	Aspredinichlys filamentosus Aspredo aspredo Aspredinidae sp. 3
	Pimelodidae	Brachyplatystoma vaillantu Pimelodus blochii
Batrachoidiformes	Batrachoididae	Amphichtys cryptocentrus
Atheriniformes	Belonidae Atherinidae	One unidentified species One unidentified species
Perciformes	Serranidae Carangidae	One unidentified species Caranx latus Chloroscombrus chrysurus Oligophtes saliens Selene vomer Trachinotus cayennensis
	Sciaenidae	Cynoscion acoupa Isopisthus parinpinnis Lonchurus ianceviatus Macrodon ancylodon Micropogonias furnieri Stellifer microps Stellifer rastrifer
	Mugilidae Polvnemidae Gobiidae	Mugil curema Polydactylus sp. Gobiidae sp. 1 Gobiidae sp. 2 Gobiidae sp. 3 Gobiidae sp. 4
	Trichiuridae Scombridae	One unidentified species
Pleuronectiformes	Bothidae Pleuronectidae Soleidae Cynoglossidae	One unidentified species One unidentified species Probably Achinus achinus One unidentified species
Tetraondonuformes	Tetraodonudae	Colomesus psittacus Spheroides marmoratus

APPENDIX. Taxonomic list of the collected species. (Six unidentified species are not listed. See Table 1.)