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Supplementary Information Appendix for

African biomass burning is a substantial source of phosphorus deposition to the Amazon,
Tropical Atlantic Ocean, and Southern Ocean

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21 **This PDF file includes:**

22 Supplementary information (SI) text in Section 1 and SI Appendix Fig. S1a and S1b

23 Supplementary methods in Section 2

24 SI Appendix Figs. S2 to S14 referenced in main manuscript in Section 3

25 Supplementary references for SI reference citations

26 **Supplementary Information Text**

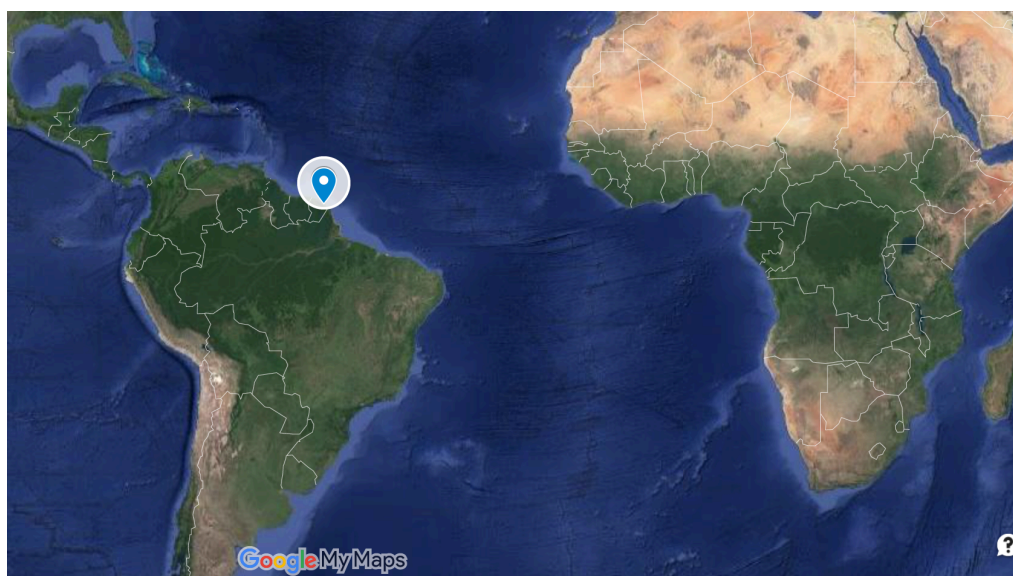
27 In section 1, the field site in Cayenne is described. Section 2 provides a detailed explanation
28 of the methods and materials used during this experiment to measure dust and total phosphorus
29 (TP) and soluble phosphorus (SP). In Section 3, the SI figures referenced in the main manuscript
30 are shown and a detailed method of the two model approaches used to estimate the total and soluble
31 P deposition to the Amazon and to the global ocean is given.

32 **1. Description of field site**

33 The blue balloon in Figure S1a shows where Cayenne, French Guiana is located in relation
34 to South America and Africa. Cayenne is the capital of French Guiana with a population of about
35 60,000 in the city proper and about 125,000 in the metro area. French Guiana (4.92°N, 52.31°W)
36 has a tropical climate that is characterized by a long, wet season from about December to June,
37 which coincides with Saharan dust transport to South America. The country of French Guiana is a
38 Department of France and maintains the same air quality standards as mainland France. As such,
39 the air quality agency ATMO-Guyane (<https://atmo-guyane.org/>) was established to monitor air
40 quality throughout the country. They maintain a suite of air quality monitoring equipment,
41 including the Thermo Scientific Tapered Oscillating Microbalance (TEOM) models 1400ab and
42 1400-FDMS used to quantify PM₁₀, which is measured every 15 minutes and used to calculate a
43 24 h daily average starting at midnight local time.

44 PM_{10} measurements collected in Cayenne have been compared to measurements made
45 outside the city in rural areas to assess the influence of local city emissions on PM_{10} concentrations.
46 Minimal differences in PM_{10} concentrations were found across the sites, which suggests that the
47 PM_{10} measurements made in Cayenne are representative of long-range transported aerosol and not
48 from local anthropogenic emissions (1). Fig. S1b shows the high-volume sampler used to collect
49 aerosol filter samples in the foreground and the Atlantic Ocean in the background.

50



51

52 **Fig. S1a.** Map showing the approximate location of the field site in Cayenne, French Guiana.



53

54 **Figure S1b.** A photo of the University of Miami high-volume sampler atop of our sampling site,
55 a 67-m coastal hill northeast of Cayenne. The filter is placed in the elevated cassette holder located
56 at the far end of the sampler box; it is protected from rain by a polycarbonate “hat”.

57

58 **2 Aerosol collection methods**

59 **2.1 Filter Sample Collection**

60 High-volume aerosol samples were collected almost daily (excluding weekends and French
61 federal holidays) in 2016. Samples were collected on 20 cm x 25 cm Whatman-41 (W-41) cellulose
62 filters with an air flow rate of approximately $0.75 \text{ m}^3 \text{ min}^{-1}$. Samples were changed at about 10
63 in the morning. At the end of each cycle, the filters were then placed in individual plastic bags and

64 periodically shipped to the University of Miami for analysis along with twice monthly collected
65 procedural blanks.

66 **2.2 Dust mass concentration**

67 One-quarter of the filter was separated from the rest of the filter for dust mass concentration
68 analysis. The filters were handled only in a laminar flow hood with gloves. Each filter section was
69 washed three times with 20 mL total of Milli-Q water (Milli-Q; Millipore, resistivity > 18.2 MΩ)
70 to remove soluble compounds, such as sea salt. The washed filters were then placed in a muffle
71 furnace for 24 hours at 500°C. The residue leftover after heating is referred to as “ash” which is
72 comprised of the filter residue and non-combustible mineral matter. Procedural blanks were also
73 run with each batch of samples and the average weight of the blank was 0.13 mg, which is roughly
74 equivalent to 0.1 mg m⁻³. The blank was subtracted from the total mass of the ash and the difference
75 is attributed to mineral dust. During the extraction and heating processes, there is some loss of
76 soluble soil minerals; therefore, an adjustment factor of 1.3 was applied to the filter ash to
77 determine the final dust mass (2). The elemental composition of western Atlantic African dust
78 determined in this way yielded an average Al concentration of 8%, close to that in the Earth’s crust
79 (3). Similarly, a large suite of elements was found to be present in ratios close to average crustal
80 abundances (3, 4).

81 **2.3 Soluble Reactive Phosphorus (SRP) quantification**

82 In this paper, SP and SRP are used interchangeably. For SRP analysis, one-eighth of the
83 W-41 filter was not washed and was instead leached in 45 mL of Milli-Q buffered with sodium
84 bicarbonate (NaHCO₃; Sigma-Aldrich, 99.5% purity) to a pH of 7, vigorously shaken, and placed
85 in a dark room at room temperature (5–7). After 16 hours of leaching, the samples were then
86 filtered with a 0.2 μm polyethersulfone syringe filter (Whatman Puradisc), 1 mL arsenate reagent
87 was added followed by 1 mL of the molybdenum reagent. The colorimetric response of the sample

88 was determined at a wavelength of 880 nm in a spectrophotometer (Shimadzu UV-1800) (8) in a
89 1 cm cuvette. The full method can be found in Strickland and Parsons (9) including the arsenate
90 correction (10). Standards were first analyzed in the sodium bicarbonate buffered aqueous solution
91 at a pH = 7 using potassium monobasic phosphate to check recovery. Procedural blanks were also
92 leached in the buffered solution and always reported as an SRP concentration below the limit of
93 detection. All reagents were ACS reagent grade. For SP the average standard deviation was 0.7 ng
94 m⁻³.

95 **2.4 Total Phosphorus (TP) quantification**

96 One-eighth of the W-41 filter was heated to 500°C for 24 hours prior to measuring TP. The
97 filter section was not washed in order to preserve any soluble forms of P on the filter. The dust
98 samples were weighed and then digested with a 3:1:1 ratio of concentrated hydrochloric acid
99 (HCl), concentrated hydrofluoric acid (HF), and concentrated nitric acid (HNO₃) in a sealed Teflon
100 Savillex vial on a hotplate heated to 195°C for ~72 hours (5). All acids were ACS trace metal
101 grade. The vials used for this analysis were precleaned with 2 mL of concentrated nitric acid,
102 capped, placed on a hot plate for 4 hours at 180°C, rinsed with Milli-Q, and dried in a laminar flow
103 hood inside a clean room. After digestion, 100 µL of the sample solution was added to 10 mL of
104 6 M nitric acid for analysis.

105 Three mineral standards (AGV-2, W2-A, and BCR-2) were also tested to ensure that the
106 method resulted in complete digestion of refractory forms of P. All mineral standards were
107 supplied by the United States Geological Survey (USGS). One sample was run in triplicate to
108 investigate the reproducibility of the digestion method resulting in a standard error of 6%. The
109 procedural blank resulted in a background level of ~1 ppm of P that was subtracted from our final
110 sample results.

111 All samples and standards were analyzed by a Thermofisher Scientific Neptune Plus multi-
112 collection inductively coupled plasma mass spectrometer (MC-ICP-MS) at the Neptune Isotope
113 Lab (NIL) at the University of Miami. Every sample was bracketed with measurements of a
114 standard solution to ensure accurate detection of instrument drift (11, 12). Detection of P was
115 measured at mass-to-charge (m/z) of 30.965 amu. For TP, the average standard deviation was 6.3
116 ng m⁻³.

117 **2.5 Scanning electron microscopy (SEM)**

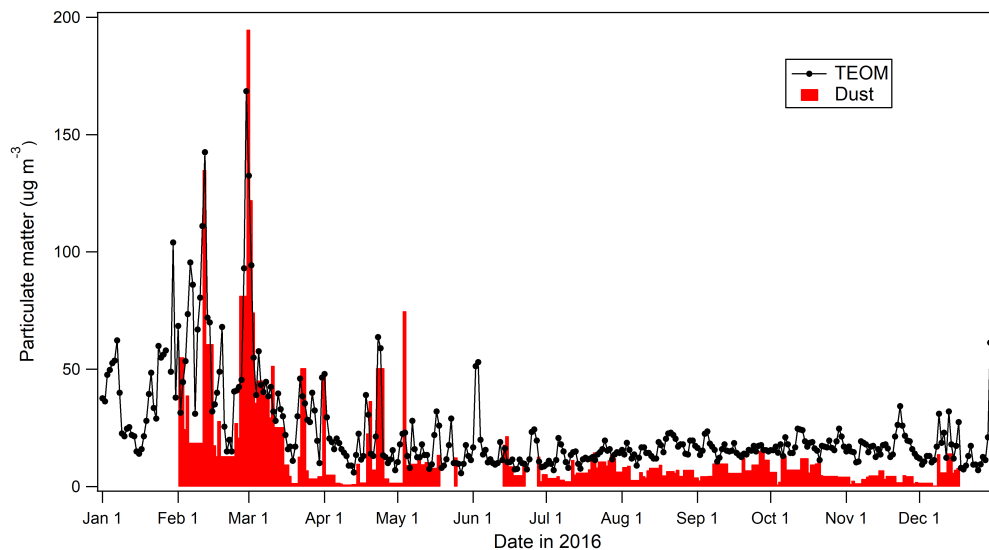
118 A ~1 cm x 1 cm piece of filter was cut and placed on an aluminum SEM stub for SEM
119 analysis. The samples were then coated with palladium in a Cressington-108 Sputter Coater and
120 imaged with a Phillips XL-30/ESEM-FEG at 20kV at the Center for Advanced Microscopy at the
121 University of Miami.

122 **3. SI figures referenced in the main manuscript**

123 **3.1. Contributions of P from African dust transported to Cayenne in Spring**

124
125 Filters collected in Cayenne, French Guiana were analyzed for dust mass concentration,
126 SP concentration, and TP concentration in Spring and Fall. Fig. S2 presents the results of all
127 measurements made at Cayenne in 2016. The filter analyses of dust are shown in red and the

128 TEOM PM₁₀ (particulate matter with an aerodynamic diameter < 10 micrometers) in black.

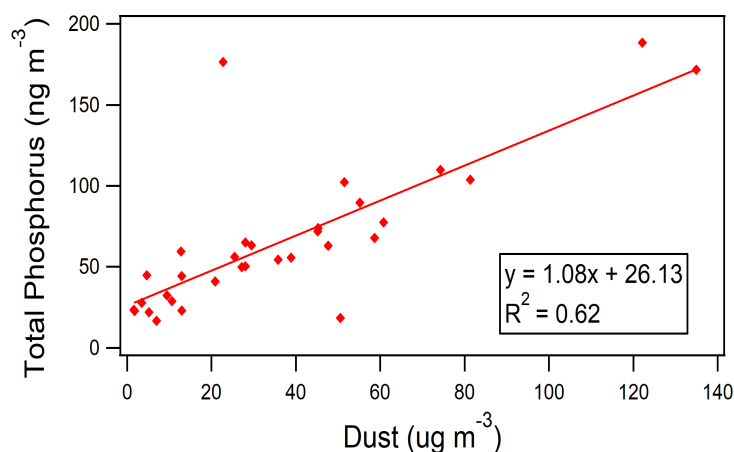


129
130 **Fig. S2.** Temporal variability of PM₁₀ concentrations, shown in black, and observed dust
131 concentrations in red.

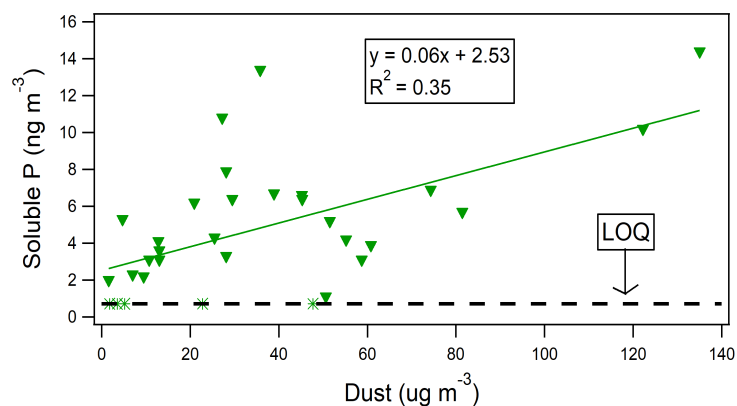
132 Fig. S3a-b below show the correlation between TP and dust as well as SP and dust for
133 Spring. P solubility of transported dust was determined for Spring by taking the average solubility
134 of major dust events (e.g., > 50 $\mu\text{g m}^{-3}$), which yielded an average P solubility of 5%.

135 The Scanning electron microscopy (SEM) images in Fig. S4 show the presence of biomass
136 burning (BB) aerosol and mineral dust from the Saharan Desert co-transported to Cayenne in the
137 Spring. This image is representative of filter images of samples collected in February and March
138 when elevated SP concentrations were measured in Cayenne. Fig. S5 shows photos of the filters.
139 The co-transport of BB emissions and dust is also indicated by the color of the filter as well. While
140 typical dust-laden filters are reddish-orange, these filters are grey-brown, which is indicative of
141 mineral dust and black carbon from BB emissions. The color of the filters shown in Fig. S5 is
142 representative of typical filters collected in February and March. Figs. S6a-c show thermal
143 anomalies detected by the Visual Infrared Imaging Radiometer Suite (VIIRS), which indicate fire
144 activity in a region. From this figure, it is evident that fire activity is prevalent in the Sahel in

145 February and March, but decreases in April in 2016. Fire data was obtained by the Fire Information
146 Resource Management System (FIRMS; <https://firms.modaps.eosdis.nasa.gov/>).

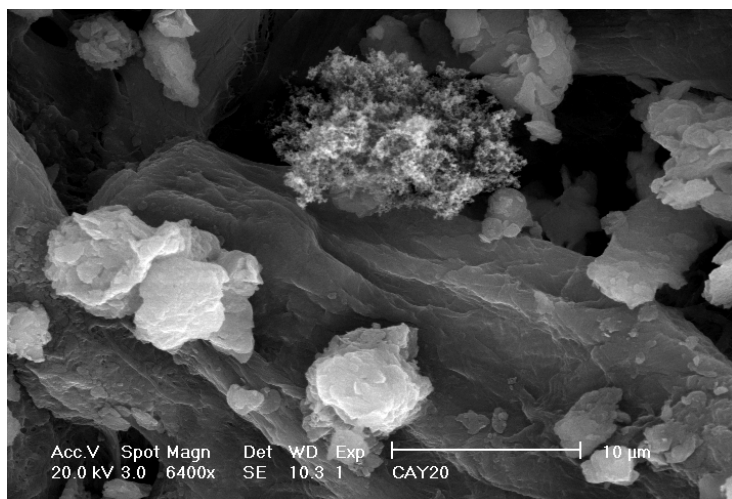


147
148 **Fig. S3a.** Correlation for dust and total phosphorus for data from Feb to Apr (Spring) 2016.



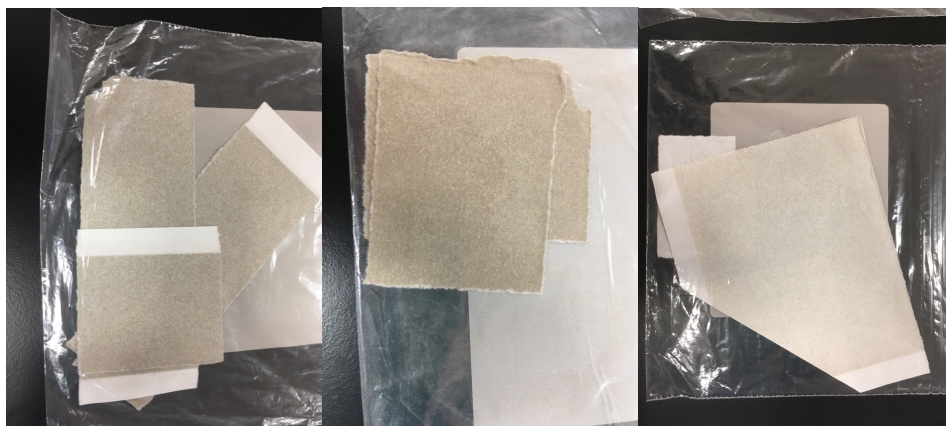
149
150 **Fig. S3b.** Correlation plot for dust and soluble phosphorus for data from Feb to Apr (Spring) 2016.
151 The LOQ is also shown as a black dashed line and has a value of 0.71 ng m^{-3} . This plot is different
152 from that in Fig. 1c in the main manuscript because it includes the points that were at or below the
153 LOQ.

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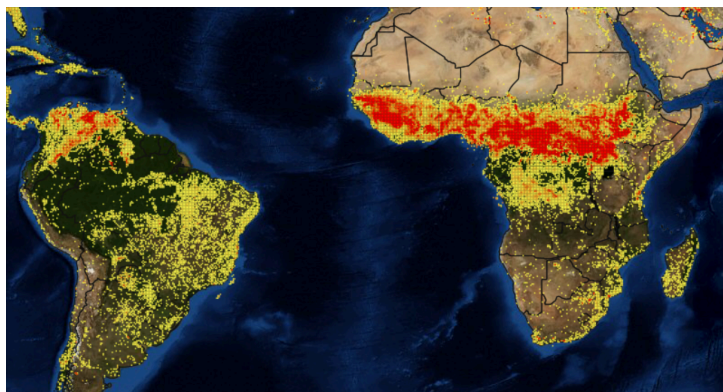
156
157 **Fig. S4.** SEM image showing transported biomass burning aerosol in the upper right and mineral
158 dust in the lower left on the same filter from February 2016. The scale bar is on the bottom and
159 represents 10 μm.

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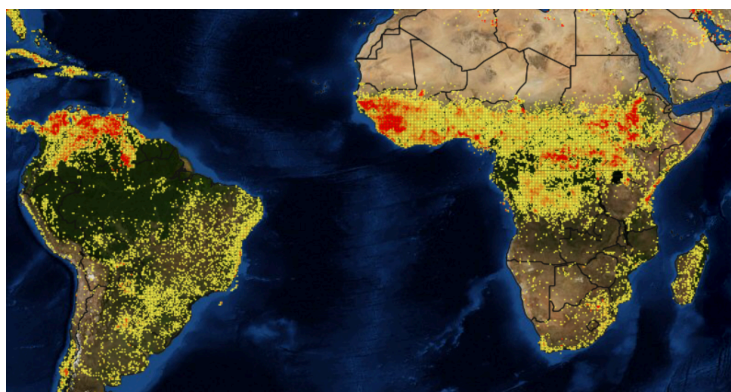


161
162 **Fig. S5.** Filters collected in Cayenne on 01, 5, and 17 February 2016 (left to right) showing grey-
163 brown hue indicative of co-transported black carbon from biomass burning emissions in the Sahel
164 and mineral dust.

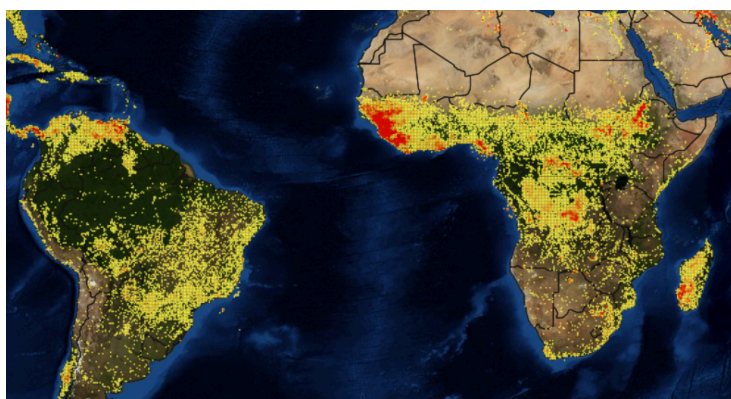
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168 **Fig. S6a-c.** The red and yellow colors in Africa shows fire activity in the Sahel in February (top),
169 March (middle), and April (bottom) 2016. The fire intensity scale is gradational with red denoting
170 greater fire activity and yellow denoting less fire activity.

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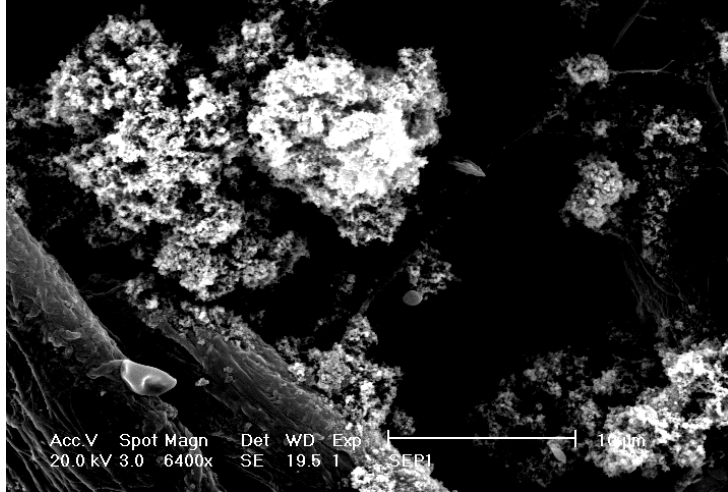
174 **3.2. An additional supply of P from southern African BB in Fall**

175 Fig. S7 shows BB aerosol imaged by SEM that was transported to Cayenne in Fall likely
176 from southern Africa. Fig. S8 shows the medium grey color of filters collected in September; the
177 grey hue is from black carbon associated with biomass burning and is representative of all filters
178 collected in Fall.

179 Fig. S9 shows the correlation in Fall between dust and SP. From Fig. S9, the relationship
180 between SP and dust is poorly correlated in Fall and many values of SP in late Oct and Nov are
181 below the LOQ. Fig. S10a shows a time-series of SP data collected in Fall and Fig. S10b shows a
182 13-day air mass back trajectory initiated on 3 October 2016 when the SP concentration is below
183 the LOQ. Figs. S10c-e show the vertical profiles of the air masses as they are transported across
184 the Atlantic Ocean. The boxed points in Fig. S10b correspond to the vertical air profiles in Fig.
185 S10c-e. This figure shows that on days when SP was measured to be below the LOQ, there is no
186 smoke transport to Cayenne.

187 SI Fig. S11 shows an air mass BT frequency plot initialized on 1 November 2016 and run
188 every 6 hours to give a spatial distribution of where air masses come from during the month of
189 November. This figure shows that the air masses no longer originate from southern Africa in
190 November. Indeed, the air masses typically come from either the nation of South Africa or from
191 near the equator.

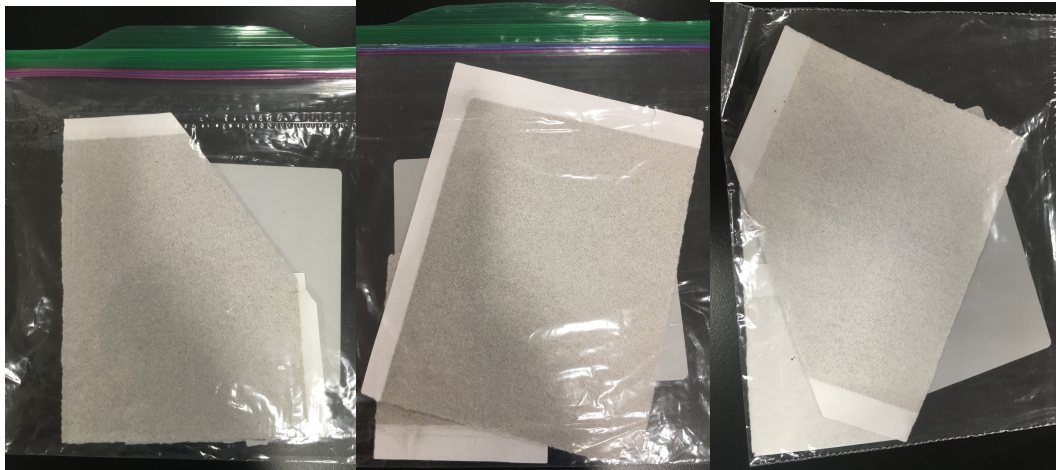
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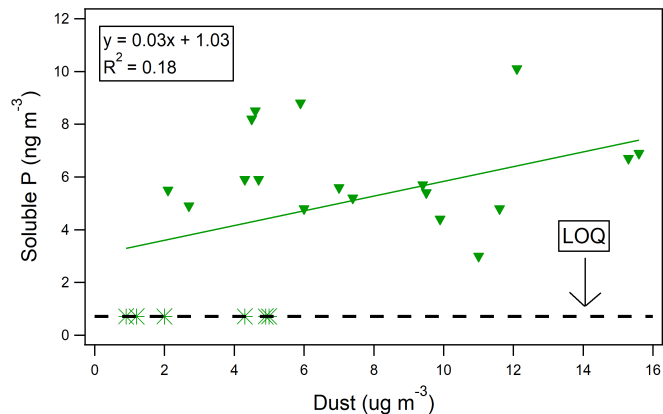
194 **Fig. S7.** Scanning electron microscopy image showing numerous biomass burning particles on
195 filters collected in Cayenne in September 2016. The scalebar is on the bottom and represents 10
196 μm .

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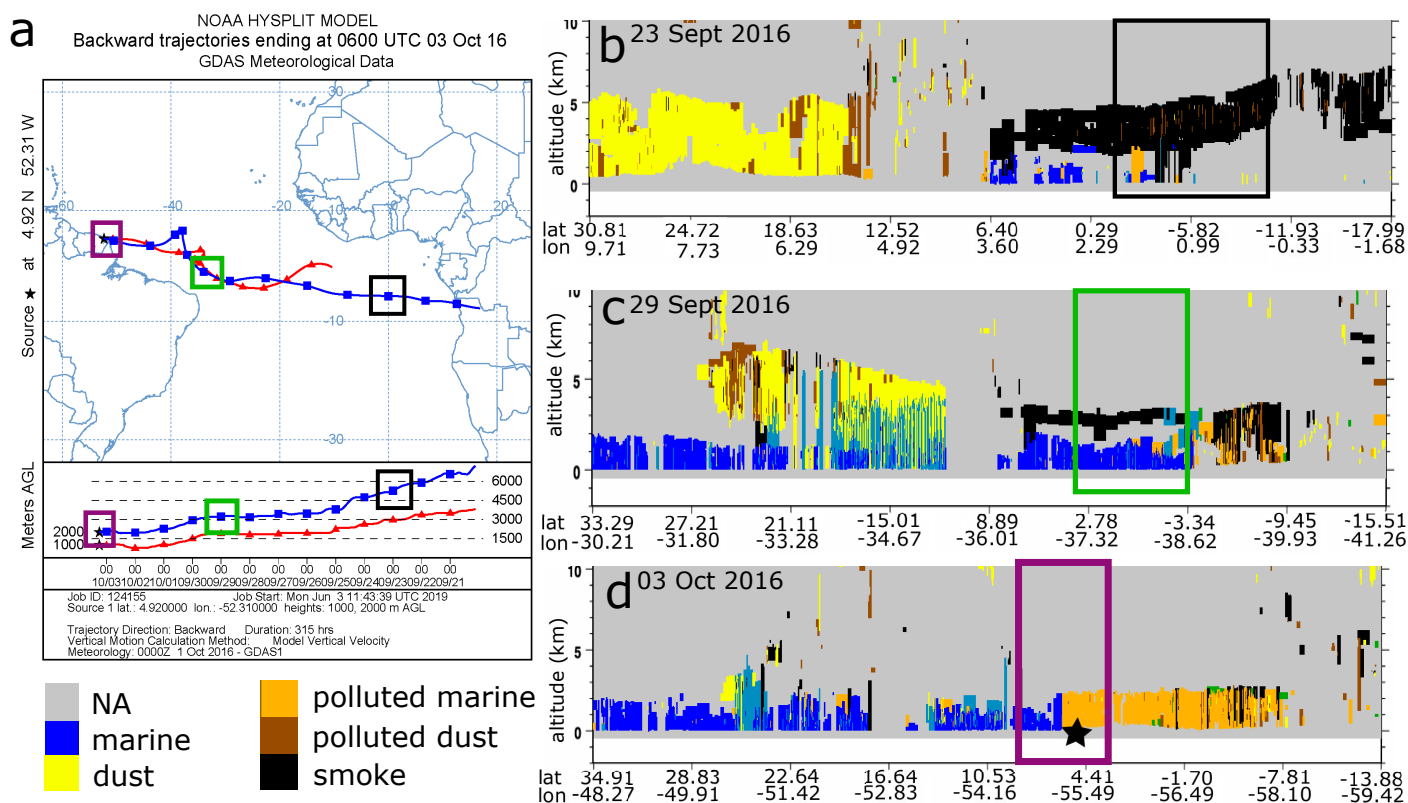


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199 **Fig. S8.** Filters collected in Cayenne on 05, 16, and 23 September 2016 (left to right) showing a
200 grey hue indicative of biomass burning emissions transported to Cayenne.

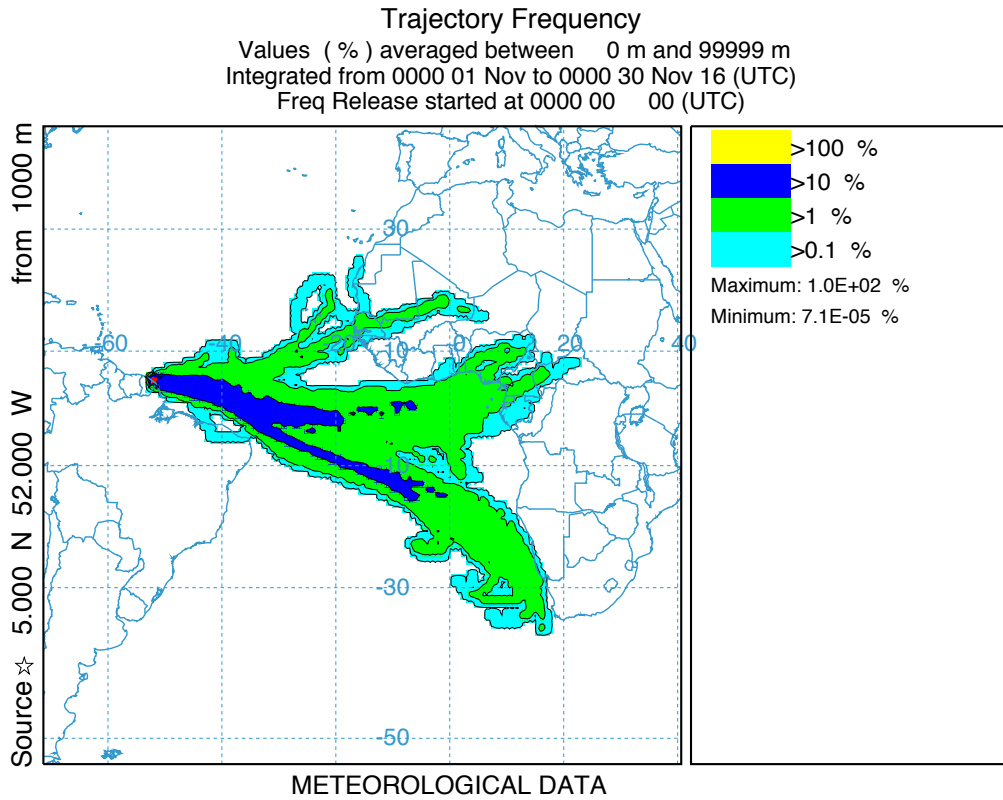


201
 202 **Fig. S9.** Correlation plot for dust and soluble phosphorus for data from Sept to Nov (Fall) 2016.
 203 The LOQ is also shown as a black dashed line and has a value of 0.71 ng m^{-3} . Most values below
 204 the LOQ are from late Oct and Nov.
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206
 207 **Fig. S10.** This figure is similar to Fig. 3 in the main manuscript except this figure shows the
 208 progression of elevated smoke on a day without measurable SP. The black arrows on the aerosol

209 vertical profiles show the approximate intersection of the BT to the CALIPSO pass. From S10d,
210 it is evident that the smoke plume is thin within a narrow altitude band. The purple box in S10e
211 emphasizes the lack of an elevated smoke layer when the air BT intersects Cayenne. The black
212 stars show the location of Cayenne.

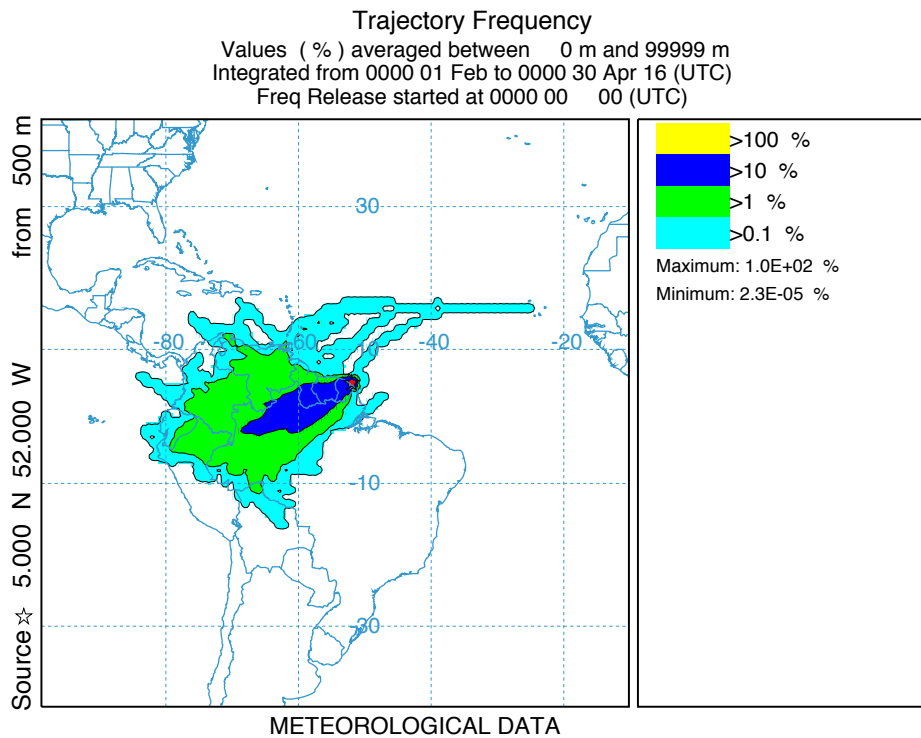


213
214 **Fig. S11.** Air mass BT frequency plot initialized on 1 November 2016 and run every 6 hours for
215 the month of November.

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222 **3.3. Estimates of P deposition to the Amazon using the MERRA-2 model and CAM**
223 **biogeochemical model**
224 **3.3.1 Using the MERRA-2 model to estimate P deposition to the Amazon**

225 Fig. S12 shows a frequency plot of 96-hour forward trajectories using the HYSPLIT model
226 initialized at 500 m from 1 February to 30 April 2016 every 6 hours. This figure shows that after
227 air masses arrive in Cayenne, they typically continue into the interior of the Amazon, which
228 demonstrates that Cayenne intercepts air masses upwind of the Amazon Basin.

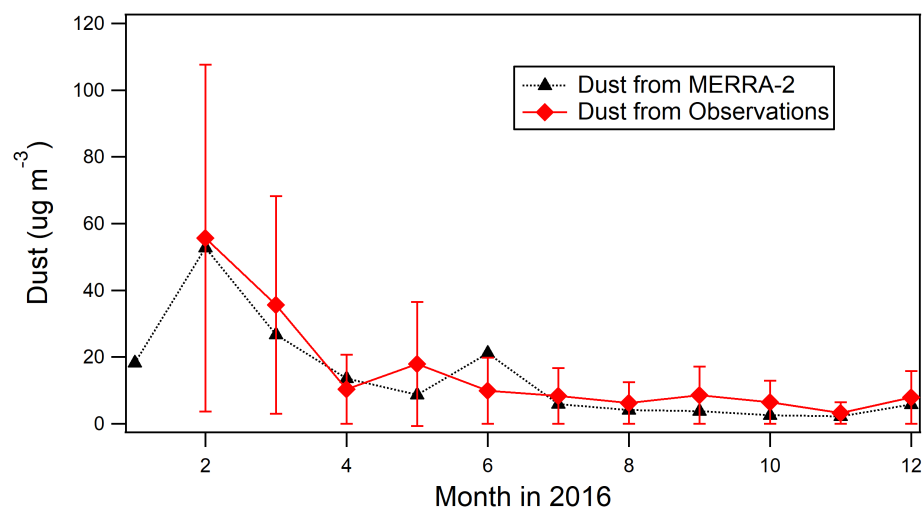


229
230 **Fig. S12.** Ninety-six-hour air mass forward trajectories initialized at 500 m on 1 February 2016 to
231 30 April 2016 every 6 hours showing that air masses originating in Cayenne typically continue
232 into the interior of the Amazon Basin.

233 We first discuss our approach using the MERRA-2 model to estimate P deposition from
234 African dust. The MERRA-2 model is based on the GEOS-5 Earth System model coupled to the
235 GOCART aerosol module (13). Version 2 performance has been substantially improved by
236 assimilating aerosol optical depth (AOD), primarily MODIS since 2002. Comparisons of

237 MERRA-2 surface aerosol model products with in situ measurements at sites in the United States
 238 show that the model provides consistently accurate estimates during African and Asian dust events
 239 (14). MERRA-2 also yielded excellent agreement with Barbados monthly means over the period
 240 1984 – 2009 (13, 15). Fig. S13 shows the MERRA-2 model output of monthly dust mass
 241 concentration for Cayenne, represented by the coordinates: -52.5°W to -51.5°W and 4.5°N to
 242 5.5°N, and measured dust mass concentration from our site in Cayenne. There was considerable
 243 agreement between our dust observations and the model output in both the magnitude and seasonal
 244 transport pattern to Cayenne. The MERRA-2 model captured the increase in dust in the Spring
 245 (February to May) when dust is transported from the Sahara to South America.

246



247

248 **Fig. S13.** Monthly average dust concentrations measured at Cayenne and monthly average dust
 249 estimates from MERRA-2 for 2016. The error bars in red are monthly standard deviations for our
 250 measured dust mass concentrations. MERRA-2 results were obtained from
 251 <https://giovanni.gsfc.nasa.gov/giovanni/>.

252 The MERRA-2 deposition product yielded a deposition rate of $4 \text{ g m}^{-2} \text{ a}^{-1}$ for the Cayenne
 253 region. The only directly measured dust deposition data in the western Atlantic is that for Florida
 254 where rates were in the range of $1.5 \text{ to } 2.0 \text{ g m}^{-2} \text{ a}^{-1}$ (16). The MERRA deposition rate for Cayenne

255 is comparable to Holocene deposition rates to the central tropical North Atlantic, $1 - 5 \text{ g m}^{-2} \text{ a}^{-1}$.
256 (17). An extensive study by Kienast et al. (18) of dust input rates to the oceans using the Thorium-
257 232 method (Thoromap) presents data for the late Holocene (0 – 4ka) in sediments off the NE
258 coast of SA directly east of Cayenne. These yield rates of $2 - 4 \text{ g m}^{-2} \text{ a}^{-1}$ and $4 - 8 \text{ g m}^{-2} \text{ a}^{-1}$. Thus,
259 the MERRA-2 rates obtained for Cayenne fall in the mid-range of those measured in regional
260 sediment cores.

261 We estimated the amount of TP deposited by African dust over the entire Amazon Basin
262 using monthly wet and dry dust deposition rates obtained from the MERRA-2 model over three
263 quadrants representative of the Amazon Basin ($-75^{\circ}\text{W}, 2^{\circ}\text{S}, -50^{\circ}\text{W}, 8^{\circ}\text{N}$), ($-75^{\circ}\text{W}, 12^{\circ}\text{S}, -50^{\circ}\text{W},$
264 2°S), and ($-50^{\circ}\text{W}, 12^{\circ}\text{S}, -40^{\circ}\text{W}, 2^{\circ}\text{S}$), as prescribed by Yu et al. (19). Using a TP to dust ratio of
265 1080 ppm (SI Fig. S3a), the TP deposition associated with wet and dry dust deposition was found.
266 This yielded a TP deposition of $0.011 \text{ Tg TP yr}^{-1}$ for the Amazon. Our TP deposition estimates for
267 African dust are within the range of Yu et al. (2015).

268 **3.3.2. Using the CAM model to estimate P deposition to the Amazon and the Oceans**

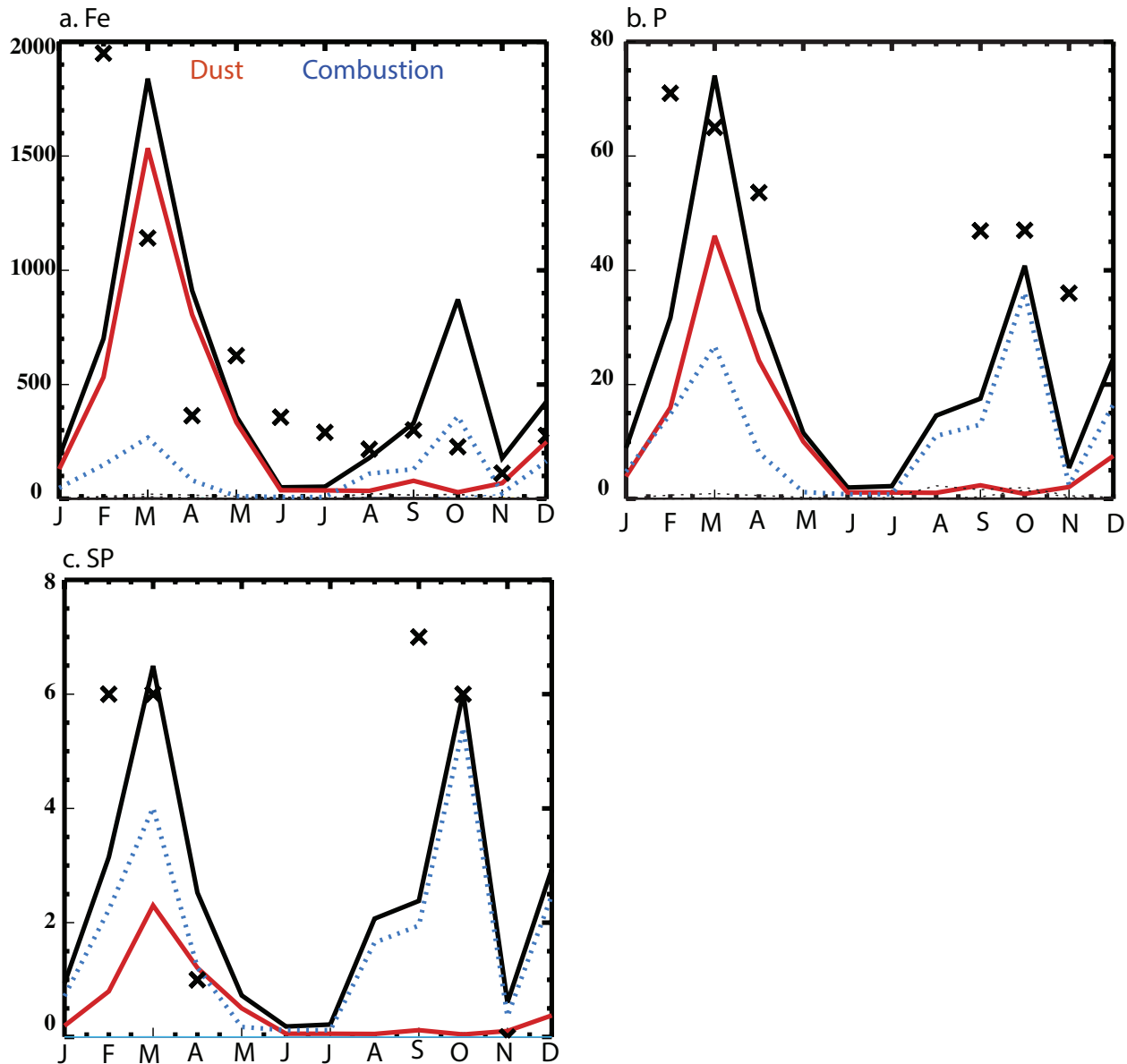
269 The second way that deposition was calculated was with version 4 of the Community
270 Atmospheric Model (CAM) using the model simulations separating each source (dust and BB
271 separately from Africa) as previously described (14). CAM has been used previously to estimate
272 P deposition to the Amazon Basin (20). For the P deposition estimates to the Amazon shown in
273 Table 1, the CAM model was tuned to match the P observations at Cayenne. For Figs. 4, 5, and 6
274 in the main manuscript, CAM used a $2.5^{\circ} \times 1.9^{\circ}$ horizontal (longitude by latitude) resolution and
275 56 vertical layers up to 2 hPa. This version of CAM had previously been used to independently
276 estimate iron deposition from dust and combustion sources (21). Meteorology (e.g., winds and
277 temperature) was nudged to the Modern-Era Retrospective analysis for Research and Applications
278 (MERRA) data for 2006-2011, and the climatological monthly mean was used in this study.

279 Sources were separated by sub-continent, and originally focused on iron, but for this study, we
280 use the source apportionment for P deposition instead.

281 To deduce the relative importance of dust and BB contributions from African sources on the
282 Amazon, we made simple assumptions about the relationship between Fe and P, based on an iron
283 content of 3.5% in dust and 1080 ppm P in dust. For biomass burning, we assume a P/BC ratio of
284 0.0029 in the submicron mode and 0.020 in the supermicron mode. The average Fe/BC in
285 combustion aerosols is assumed to be 0.2. For Fig. 4, we tuned the model to match the P
286 observations in this study (see Fig S14) by first estimating enough Fe in dust to account for the
287 observed dust mass concentrations, then converting the estimated Fe to P. To match our
288 observations, we increased the dust by 2-fold to match the P and soluble P for the boreal winter
289 months (see Figure S14a). To match our observations particularly in Fall when BB dominates,
290 current P/BC ratios alone cannot explain our P observed, and we needed to increase our P from
291 combustion by 60-fold (Figure S14b). This required enhancement can be due to a combination of
292 factors: larger aerosol emissions from biomass burning than estimated by the biomass burning
293 inventory (consistent with previous comparisons with AERONET, which required an
294 enhancement by a factor of 3 in Africa from local observations (22)), higher P emissions from
295 biomass burning than predicted by currently used P/BC ratios, or inadequate transport from Africa
296 to South America in the model. Finally, we use a SP fraction for dust and combustion of 0.05 and
297 0.15, as deduced in this study, for the deposition. The bulk (>80%) of the combustion aerosol
298 simulated in the model from the North African and South African sources is from biomass burning
299 not industrial sources. An estimate of the uncertainty in these estimates would suggest that for the
300 dust P, using the original model simulations, which were tuned to match local source strength in
301 North Africa would be a good lower bound, while tuning to the observations here (multiplying by
302 a factor of 2), would be the upper end estimate. For the biomass burning, since local observations

303 from AERONET suggest that the wildfire emissions should be multiplied by a factor of 3x (23),
304 we include this as a lower end estimate, while the higher value (multiplied by a factor of 60x) to
305 match the observations here, are used as an upper end. This gives us a range from 0.011-0.033 Tg
306 P yr⁻¹ of total P and 0.0006 to 0.002 Tg P yr⁻¹ in SP from dust deposition to the Amazon, and, in
307 comparison, 0.0026-0.052 Tg P yr⁻¹ of total P, and 0.0004-0.008 of SP from biomass burning
308 deposition in the Amazon.

309 For our global deposition estimates shown in Figs. 5 and 6, we used a slightly different set of
310 model runs, described in Brahney et al. (23) and Chien et al. (24) and did not tune CAM to our
311 observations since global sources rather than the local sources that were measured in this study
312 were accounted for. For these simulations, the winds were derived online, with fixed sea surface
313 temperatures. All sources of P were included (dust, sea salts, biomass burning, fossil fuels, primary
314 biogenic particles and volcanoes), but were not separated into different regional sources. The only
315 change from Brahney et al., (23) and Chien et al., (24) is that the soluble fraction was assumed to
316 be 0.05 for dust and 0.15 for other sources, as found in this study.



317

318 **Figure S14.** Model predictions from the Community Atmospheric Model (CAM) estimating the
 319 Fe (a), TP (b) and SP (c) concentrations (all in ng m^{-3}) from different sources at the Cayenne
 320 measurement site, versus the observations (black x). The different aerosol sources include
 321 African combustion (dotted blue line) and African dust (solid red line). The solid black line
 322 represents the total Fe (a), P (b), and SP (c) concentrations from both sources. Emissions of BB
 323 from within the Amazon are excluded in order to focus on the impact of long-range transported
 324 aerosol.

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327 African dust transport to the Caribbean Basin and South America and its impact on the
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