1	Solidified floating organic drop microextraction (SFODME) for the simultaneous
2	analysis of three non-steroidal anti-inflammatory drugs in aqueous samples by
3	HPLC
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#### 20 ABSTRACT

In this work, a liquid-liquid microextraction methodology using solidified floating 21 22 organic drop (SFODME) was combined with liquid chromatography and UV/Vis detector 23 to determine non-steroidal anti-inflammatory drugs (NSAIDs), naproxen (NPX), diclofenac (DCF) and mefenamic acid (MFN) in tap water, surface water and seawater 24 samples. Parameters that can influence the efficiency of the process were evaluated, such 25 as the type and volume of the extractor and dispersive solvents, effect of pH, agitation 26 27 type and ionic strength. The optimized method showed low detection limits (0.09 to 0.25  $\mu$ g L<sup>-1</sup>), satisfactory recoveries rates (90 % to 116 %), and enrichment factor in the range 28 29 between 149 and 199. SFODME showed simplicity, low cost, speed and high 30 concentration capacity of the analytes under study. Its use in real samples did not demonstrate a matrix effect that would compromise the effectiveness of the method, being 31 possible to apply it successfully in water samples with different characteristics. 32

33 Keywords: Floating drop; Microextraction; Non-steroidal anti-inflammatory; Water
34 sample.

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## 39 **1. INTRODUCTION**

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Pharmaceuticals present in aquatic environments are among the contaminants 40 that have most attracted attention of the scientific community. Among these substances 41 are non-steroidal anti-inflammatory (NSAIDs), widely recommended by healthcare 42 43 professionals worldwide [1]. These pharmaceuticals have great applicability in human and animal medicine due to the analgesic, antipyretic and anti-inflammatory effect [2]. 44 Because of their extensive use, the compound itself, in their unaltered form, or even their 45 metabolites can reach the environment in several ways. Inappropriate disposal of unused 46 pharmaceuticals, domestic sewage discharge, animal excretion and their use in soil 47 fertilization [2-4] can be considered the main sources of environmental contamination. 48 The presence of NSAIDs in ecosystems, even at trace levels, is potentially dangerous due 49 to their toxicity and, in some cases, bioaccumulation capacity [5,6]. This class of 50 pharmaceuticals includes diclofenac (DCF), naproxen (NPX), and mefenamic acid 51 (MFN), among others. NSAIDs are often found in the environment at concentrations in 52 the range of ng  $L^{-1}$  to  $\mu$ g  $L^{-1}$  [5,7]. Although there are some reports of high concentrations 53 54 [8-10] in aquatic environment, generally DCF, NPX and MFN are present at concentrations that reach a few  $\mu g L^{-1}$ . Moreover, various studies have observed toxic 55 effects in aquatic organisms even at these low concentrations [11-14]. Therefore, the 56 57 determination of these substances in aquatic environments is extremely important [7,15]. Since, generally the concentrations of pharmaceuticals in environmental 58 matrices are very low, it is necessary to incorporate pre-concentration and *clean-up* 59 procedures before analysis by chromatographic methods [16]. Solid-phase extraction 60 (SPE) and liquid-liquid extraction (LLE) are widely used to prepare samples from 61

different matrices [17-19], but the disadvantages are the high use of organic solvents,

large amount of waste generated and multi-stage analytical procedures with consequent 63 64 high time-consuming and labor intense. Therefore, new extraction methods have been proposed, such as solid-phase microextraction (SPME), which compared to LLE and SPE 65 requires less time and solvent consumption but has the disadvantage of the high-cost 66 [3,16,20-22]. As alternative, other microextraction techniques have experienced great 67 development lately, such as single drop microextraction (SDME) [23,24], hollow fiber 68 69 liquid-phase microextraction (HF-LPME) [25,26] and dispersive liquid-liquid 70 microextraction (DLLME) using different approaches [27].

DLLME has numerous advantages such as simplicity and low-cost, but has the 71 72 disadvantage of using toxic solvents, as well as the difficulty of the removal of the solvent 73 drop, accumulated at the bottom of the tube [28]. The DLLME technique with floating organic droplet solidification (SFODME) [29], in addition to having the advantages of 74 75 the conventional DLLME, makes use of less toxic solvents and allows easy removal of the extracting solvent, since it is completely solidified as a drop on the sample surface 76 77 [30,31]. Succinctly, a few microliters of an organic solvent (extracting solvent), with low miscibility and with lower density than water, together with a dispersive solvent, are 78 79 quickly injected into the aqueous sample, producing high turbulence cloud. This 80 turbulence causes the formation of organic droplets, which are dispersed throughout the aqueous sample. After the formation of a cloudy solution, the equilibrium state is reached, 81 the mixture is vortexed and centrifuged. After this step, the extracting solvent is 82 83 completely concentrated on the sample surface. Then, the system is placed in contact with an ice bath for a few minutes and the frozen drop of organic solvent that remains at the 84 85 top of the aqueous sample is collected into a vial for HPLC analysis. Based on this methodology, this work aims to develop an accurate, selective and sensitive SFODME 86 procedure followed by HPLC-UV/Vis analytical method for the simultaneous 87

determination of three NSAIDs in aqueous samples. The parameters that can influence
the efficiency of the extraction process were evaluated, such as the type and volume of
the extracting and dispersive solvents, the effect of pH, agitation mode and ionic strength.
The proposed method has been validated and satisfactorily tested in different aqueous
samples, such as tap water, surface water and seawater.

## 93 2. MATERIAL AND METHODS

## 94 **2.1 Reagents and samples**

Analytical standards NPX ( $\geq$  98.0 %), DCF ( $\geq$  98.0 %), MFN ( $\geq$  98.0 %) and the 95 extracting solvents 1-undecanol ( $\geq$  98.0 %) and 1-dodecanol ( $\geq$  98.0 %) were provided 96 97 by Sigma-Aldrich (United States). All of the HPLC grade organic solvents ( $\geq$  99.9 %) used, methanol (MET), acetonitrile (ACN) and ethanol (ETN) were provided by Merck 98 99 (Germany) and all other standard analytical grade reagents employed phosphoric acid (> 85%), sodium hydroxide ( $\geq$  99%) and sodium chloride ( $\geq$  99%) were from Isofar (Brazil). 100 The ultrapure water (18.2M $\Omega$ -cm) used was obtained through Milli-Q systems from 101 102 Merck Millipore (Darmstadt, Germany).

103 The individual stock solutions of NPX, DCF and MFN were prepared in 104 methanol at the concentration of 100 mg  $L^{-1}$ . Successive dilutions were prepared in 105 methanol.

#### 106 **2.2 Instrumentation**

107 A Shimadzu® liquid chromatograph (LC-20AT Prominence) equipped with a
108 UV/Vis detector SPD-20A with a slit of 8nm was used. In the first 7 minutes of analysis,
109 the 256 nm wavelength was used in the UV/Vis (for NPX) and in the last 8 minutes (for

110 DCF and MFN), the 234 nm wavelength. The separation of the compounds was 111 performed on a Luna C-18 reverse phase column (Phenomenex®) (250 mm x 4.6 mm, 112 5  $\mu$ m), maintained at 25 °C, using a mobile phase of ACN: acidified water with H<sub>3</sub>PO<sub>4</sub> 113 (pH 2.24) 60/40 ( $\nu/\nu$ ), with a flow of 1.2 mL min<sup>-1</sup>. The mobile phase was filtered through 114 0.45  $\mu$ m nylon membrane filters (Millipore). To control the equipment and obtain the 115 data, a microcomputer and LC solution® software version 1.24 SP1 of Shimadzu were 116 used.

117 An MX-S mini vortex and a USC-1400A ultrasound were used for the extraction118 process.

### 119 **2.3 SFODME procedure**

SFODME method was performed by injecting a mixture containing 30 µL of 1-120 dodecanol (extracting solvent) and 150 µL of ACN (dispersive solvent) into 5 mL of 121 122 water sample with pH adjusted to 2 with  $H_3PO_4$  85 % (v/v) and containing 2.5 % (w/v) of NaCl. Due to the presence of salt in the seawater sample, NaCl was not added. The choice 123 of using 1-dodecanol (water insoluble, melting point of 24 °C) as extracting solvent was 124 made based on preliminary experiences reported in detail in supplementary material 125 126 (SM). After addition of the extraction mixture, the tube containing the sample was 127 vortexed for 20 s and then centrifuged at 5000 rpm for 4 minutes [32]. After this step, the extractor was completely concentrated as a drop on the sample surface. To cause this drop 128 129 to solidify, the system was placed in contact with an ice bath for a few minutes. After complete solidification of the extracting solvent, it was easily removed in an interval 130 lower than 1 min with a spatula and placed in a vial at room temperature. 20µL of liquid 131 drop were then analysed by HPLC-UV/Vis. For each condition tested, five extraction 132 replicates were performed, which were analysed once by HPLC. 133

#### 134 **2.4 Optimization of extraction conditions**

## 135 *2.4.1 Effect of dispersive solvent type and volume*

The dispersive solvent must be miscible in both the organic and aqueous phases and must cause adequate dispersion of the extracting solvent throughout aqueous sample when the extraction mixture is added. It favors the contact between the extracting solvent and the sample containing the analytes, in order to improve the extraction capacity and reduce the time needed to achieve the equilibrium state. However, high volumes of dispersive solvent can also increase the solubility of the analytes in the aqueous phase [31,33].

In this work, three dispersive solvents (MET, ACN and ETN) using three different volumes (50, 150 and 200  $\mu$ L) were tested in order to evaluate the extraction capacity of the NSAIDs with each combination. The protocol consisted in the addition of 30  $\mu$ L of extracting solvent (1-dodecanol) with the different volumes of each dispersive solvent tested (MET, ACN and ETN in 50  $\mu$ L, 150  $\mu$ L and 200  $\mu$ L) to 5 mL of ultrapure water at pH 2 (adjusted with H<sub>3</sub>PO<sub>4</sub> 85 % ( $\nu/\nu$ )) fortified with 1.2  $\mu$ g L<sup>-1</sup> of each NSAIDs.

## 149 2.4.2 Effect of ionic strength

The increase in the ionic strength of the sample results in a decrease of analyte solubility and generally enhances the extraction efficiency, due to the *salting-out* effect. This effect is caused by changing the partition coefficients of the analytes between the aqueous and organic phases [29,31,34]. In order to evaluate the *salting-out* effect on the extraction efficiency, four NaCl concentrations in the range of 0 and 10 % (*w/v*) were tested. The procedure included the introduction of a mixture of 30 µL of the extracting solvent (1-dodecanol) and 150 µL of the dispersive solvent (ACN) in 5 mL of ultrapure

157	water, with pH 2 (adjusted with H <sub>3</sub> PO <sub>4</sub> 85 % ( $\nu/\nu$ )), fortified with 1.2 µg L <sup>-1</sup> of the
158	NSAIDs and containing the different amounts of NaCl ( $0\%$ , $2.5\%$ , $5\%$ , and $10\%$ ( $w/v$ )).

### 159 *2.4.3 Effect of pH*

The pH of the aqueous solution can have a large effect on the microextraction 160 capacity since it can influence the charge of the target compound and the amount 161 extracted will depend on ionization. At low pH, the NSAIDs will be in neutral form (pKa 162 NPX=4.15; pKa DCF=4.15; pKa MFN=4.2), which will facilitate their transfer to the 163 extractor solvent [3,31]. Initially, pH 2 was the condition used according to Beldean-164 165 Galea et al [3]. However, the reference used did not evaluate the extraction of MFN. Thus, 166 in this work, the extraction efficiency of the three NSAIDs under study, using three 167 different values (1.0, 2.0 and 4.0), was evaluated. The protocol included the introduction of a mixture containing 30 µL of the extracting solvent (1-dodecanol) and 150 µL of the 168 dispersive solvent (ACN) in 5 mL of ultrapure water at different pH (adjusted with H<sub>3</sub>PO<sub>4</sub> 169 85 %), with NaCl 2.5 % (w/v) and fortified with 1.2 µg L<sup>-1</sup> of the NSAIDs. 170

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The volume of extracting solvent can have a great influence on the efficiency of SFODME and it can affect the enrichment factor (ratio between the analyte concentrations in the organic ( $C_0$ ) and aqueous ( $C_{aq}$ ) phases) [31]. To verify the effect of the extracting volume, three volumes were tested: 30 µL, 40 µL and 50 µL. The procedure involved the addition of the corresponding volume of the extracting solvent (1-dodecanol) and 150 µL of dispersive solvent (ACN) in 5 mL of ultrapure water pH 2 (adjusted with 181 H<sub>3</sub>PO<sub>4</sub> 85% ( $\nu/\nu$ )), containing NaCl 2.5 % ( $w/\nu$ ) and fortified with 1.2 µg L<sup>-1</sup> of the 182 NSAIDs.

## 183 2.4.5 Stirring type effect

To evaluate the influence of the stirring type effect of vortex, ultrasound and manual agitation in the extraction efficiency, several experiments were performed. The protocol consisted in the mixture of 30  $\mu$ L of extractor (1-dodecanol) and 150  $\mu$ L of disperser (ACN) in 5 mL of ultrapure water pH 2 (adjusted with H<sub>3</sub>PO<sub>4</sub> 85% (*v*/*v*)), fortified with 1.2  $\mu$ g L<sup>-1</sup> of the NSAIDs and NaCl 2.5 % (*w*/*v*). The samples were submitted to the three mixing models separately, followed by centrifugation for 4 minutes at 5000 *rpm*.

## 191 **2.5 Determination of NSAIDs in water samples**

Finally, to evaluate the applicability of the proposed SFODME, tap water, 192 193 surface water and seawater were collected in glass containers and subjected to the optimized method described previously. Samples were collected in October 2019 in the 194 city of São Luis, Brazil. The 1 L amber flasks used in the collection were previously 195 washed and the samples were transferred, stored and refrigerated at approximately 4 °C 196 until use. Samples were filtered through 0.45µm nylon membrane filters (Millipore), 197 198 previously to the extraction procedure described in section 2.3 and submitted to HPLC-UV/Vis analysis. The evaluation of the water matrix influence on the extraction was 199 performed by spiking known amounts of NPX, DCF and MFN simultaneously on the 200 water samples and subjecting them to the previously optimized extraction procedure. 201 Spiked concentrations used were 1.2  $\mu$ g L<sup>-1</sup>, 3  $\mu$ g L<sup>-1</sup> and 5  $\mu$ g L<sup>-1</sup>. Five replicates of 202 extraction were performed for each of the three levels of fortification studied. The 203

recovery (%) was calculated as the ratio between the experimentally determined average

205 concentration and the corresponding expected concentration.

## **3. RESULTS AND DISCUSSION**

## **3.1 Optimization of extraction conditions**

## *3.1.1 Effect of dispersive solvent type and volume*

The recovery (%) (calculated as described in section 2.3) obtained for the selection of the type and volume of the dispersive solvent are shown in Fig. 1.



- Fig. 1 Effect of the type and volume of the dispersive solvent. Extraction conditions: 5 mL of ultrapure water, pH adjusted to 2, fortified with NSAIDs (1.2  $\mu$ g L<sup>-1</sup>), different volumes of dispersive solvents, and 30  $\mu$ L of 1-dodecanol. The y-axis was adjusted to 130% to allow a better visualization of DCF and MFN data. (n=5)
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Similar results, for all NSAIDs, were obtained using 50 and 150 µL of ACN. 220 221 Although the solvents ETN and MET could provide a better extraction efficiency of DCF 222 and MFN, they did not provide satisfactory recoveries for the NPX (with values in the range between 177% and 434%, resulting in an overestimation of NPX). This 223 224 overestimation might be associated with the interference of substances in 1-dodecanol that appears at the same retention time of NPX when ETN and MET were used as 225 226 dispersive solvents. Given this scenario and considering that ACN exhibited good results, 227 it was selected as dispersive solvent. Recoveries, for NPX, using 50 µL of ACN were slightly higher than 100 %, while using 150 µL the extraction recovery of the three 228 229 analytes ranged from 82 % to 103 %. For these reasons, volume of 150 µL of ACN was 230 selected for the following experiments.

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232 *3.1.2 Effect of ionic strength* 



Fig. 2 Effect of (a) NaCl (b) pH, (c) extracting solvent volume and (d) stirring type on
the extraction efficiency of NSAIDs using 5 mL of ultrapure water fortified with 1.2 μg
L<sup>-1</sup> of each compound. (n=5)

For all the NSAIDs tested, the extraction capacity improved by the addition of 2.5 % (w/v) of NaCl to the aqueous sample. This is explained by the *salting-out* effect, in which the solubility of compounds in water is reduced due to the change in their partition coefficients [31,34], thus increasing its affinity to the organic solvent. However, when

NaCl concentrations increases too much (5 % and 10 % (w/v)), a decrease in the extraction recovery was obtained. This is in agreement with the Khalili Zanjani et al [29], who attribute this reduction to restrictions on the diffusion of analytes to the organic phase due to the increase in the viscosity of the sample. Moreover, this study allows to conclude that this method should be successfully applied to seawater samples, since salinity generally is around 3.5 %.

## 249 *3.1.3 Effect of pH*

In order to evaluate the influence of aqueous pH in the extraction procedure, aqueous samples were adjusted to different pH. The chosen pH values are accordingly to  $pK_a$  of the NSAIDs under study. Results obtained are shown in Fig. 2 (b).

Recovery rates for DCF increases with increasing pH from 1 to 4. In the case of 253 NPX and MFN, there is an increase in recovery results from pH 1 to 2 and a decrease 254 when pH 4 is used. Considering the standard deviations, results obtained for DCF and 255 256 MFN at pH 2 and 4 cannot be considered different. Thus, the condition that allows a better 257 extraction of all the NSAIDs under study and chosen for the further tests was pH 2, in which recovery rates obtained, for the three NSAIDs studied, varied between 86 % and 258 259 119 %. This is in agreement with Beldean-Galea et al [3] and Shukri et al. [32] who also demonstrated in their work that pH 2 would be the best to be used for NSAIDs. In fact, 260 the NSAIDs under study present a  $pK_a$  around 4, which means that at pH lower than their 261 262  $pK_a$  values the compounds are in its neutral form, thus facilitating their transfer to the extracting solvent [31]. At higher pH values, the analytes are mostly in ionized form, 263 which negatively influences the ability to be extracted. 264

#### 265 *3.1.4 Effect of extracting solvent volume*

Fig. 2 (c) shows the results of the variation of the recovery with the increase of 266 267 the extracting volume. For all the compounds it was possible to verify that the increase in the extracting solvent volume caused a decrease in the extraction capacity. In fact, for 268 269 DCF and MFN, using 50 µL of extracting solvent, no peak was detected. The increase of 270 extracting solvent volume originates the dilution of the analyte, thus a reduction in the 271 concentration of the compound in the organic phase, which impacts the enrichment factor 272 and consequently the efficiency of the process. From the three volumes tested, the one that provided the best results was 30 µL, being the volume chosen for the next 273 experiments. 274

## 275 *3.1.5 Effect of stirring type*

276 Stirring is very important to guarantee that the organic solvent drops are totally 277 dispersed through the aqueous sample, improving the contact between the analyte to be 278 extracted and the extracting solution. Three types of stirring were tested, and results 279 obtained are shown in Fig. 2 (d).

The type of stirring that provided the best results for all the analysed compounds 280 281 was the vortex, being therefore chosen to be used in this study. Guiñez et al. [35] demonstrated in their work with nitro-PAHs that the use of vortex also generated higher 282 283 percentages of recovery compared to manual mixing and ultrasound. The vortex allows a better and more uniform dispersion of the system, favouring the contact between the 284 285 aqueous and organic phases and thus significantly improving the extraction capacity associated to the method, when compared to manual agitation and the ultrasound assisted 286 287 extraction.

### 288 **3.2 Analytical Performance SFODME**

Six standards were prepared in ultrapure water in concentrations between 0.6 to 289 290 5  $\mu$ g L<sup>-1</sup> for NPX and DCF and 1.2 to 5  $\mu$ g L<sup>-1</sup> for MFN. The standards were subjected to the optimized extraction method described in section 2.3 in order to determine the 291 corresponding calibration curves. The analytical performance of the proposed method, 292 presented on Table 1, was performed by assessing linearity (Lin (%) = 100-RSD, where 293 RSD is the slope's relative standard deviation), determination coefficients  $(r^2)$ , detection 294 295 limits (LOD) and quantification limits (LOQ), using the conditions of optimized SFODME. The LOD and LOQ were determined from the equations  $3^{*}(s/S)$  and  $10^{*}(s/S)$ , 296 respectively, where s is the estimated standard deviation of the regression equation and S 297 298 is the slope of the calibration curve [36]. Recovery (%) was calculated as the ratio between the experimental average concentration obtained from HPLC analysis and the 299 corresponding expected concentration. Enrichment factor (EF) was determined by the 300 ratio between the analyte concentration determined using the developed method and the 301 initial analyte concentration added. 302

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Table 1. Quantitative parameters for analytical curves obtained by SFODME–HPLC–
UV/Vis for NPX, DCF and MFN.

Analyte	Linear range (µg L <sup>-1</sup> )	$r^2$	Linearity (%)	LOD (µg L <sup>-1</sup> )	LOQ (µg L <sup>-1</sup> )	Extraction Recovery <sup>a</sup> (%)	Enrichment factor <sup>a</sup>
NPX	0.6 - 5	0.9964	93.04	0.09	0.29	116±11	199±18
DCF	0.6 - 5	0.9956	91.68	0.25	0.82	90±4	149±6
MFN	1.2 - 5	0.9940	94.49	0.15	0.51	109±5	182±9
306 <sup>a</sup> Mean	value $\pm$ standard de	viation (n=	5) obtained for	a concentration	n of 1.2 $\mu$ g L <sup>-1</sup>	for NPX, DCF a	nd

307 MFN. Extraction conditions: 5 mL of standard NPX, DCF and MFN with pH adjusted to 2; 30 μL of 1-

dodecanol as extracting solvent; 150  $\mu$ L of acetonitrile as dispersive solvent and NaCl 2.5 % (w/v).

The SFODME-HPLC-UV/Vis resulted in good linearity with determination 309 310 coefficients greater than 0.99, showing low dispersion between their experimental analysis. The low LODs (between 0.09 and 0.25  $\mu$ g L<sup>-1</sup>) and LOQs (between 0.29 and 311  $0.82 \mu g L^{-1}$ ) confirm the applicability to detect and quantify the NSAIDs in environmental 312 water samples. Also, the NPX showed superior EF and ER when compared to the other 313 analytes under study, but all the NSAIDs presented good recoveries, between 90 and 314 116%, and good EF values, between 149 and 199. For the optimized conditions, the 315 SFODME method presents satisfactory parameters, thus an evalution of the applicability 316 in complex natural water matrices should be performed. 317

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# 3.3 Analysis of environmental water samples

Spikes of 1.2, 3 and 5  $\mu$ g L<sup>-1</sup> of NPX, DCF and MFN were performed into three 319 different types of samples: seawater, surface water and tap water. All samples were 320 321 submitted to the optimized SFODME method described in section 2.3 and analysed by HPLC-UV/Vis. Fig. 3 shows the chromatograms obtained from the analysis of the real 322 samples with and without spiked of anti-inflammatories. Water samples chromatograms 323 324 allow us to verify that the compounds under study were not present in the samples before 325 the standard addition. On the other hand, some studies reported the occurrence of the same NSAIDs in natural waters. Sodré and Sampaio [37] quantified them in drinking 326 waters from Brazilian Federal District (Brazil) DCF in the range of  $0.0042 \ \mu g \ L^{-1}$  to 327 0.006  $\mu$ g L<sup>-1</sup>, and MFN in the range of 0.0016  $\mu$ g L<sup>-1</sup> to 0.0083  $\mu$ g L<sup>-1</sup>, but NPX was not 328 detected in that samples. Ide et al. [38] verified the presence of NPX in Iguaçu River 329 (Brazil) at maximum concentration of 0.34  $\mu$ g L<sup>-1</sup>. Pereira et al. [39] detected DCF in a 330 seawater sample from Santos Bay (Brazil) at concentration of 0.0194 µg L<sup>-1</sup> and Chaves 331

et al [40] found concentrations of DCF in a range of 0.105 to 0.463 µg L<sup>-1</sup> in surface
waters from Anil and Bacanga Rivers (Brazil).



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**Fig. 3** Chromatograms of different water samples with and without of  $5 \ \mu g \ L^{-1}$  of NSAIDs (a) seawater, (b) tap water and (c) surface water. Extraction conditions: 5 mL of sample with pH adjusted to 2, NaCl 2.5% (*w/v*) (except to seawater sample, which no salt was added), 150  $\mu$ L of ACN and 30  $\mu$ L of 1-dodecanol. (n=5).

It is possible to observe the increase in the peaks attributed to the NSAIDs in the fortified samples that were absent in the aqueous samples. The results obtained were considered satisfactory because there were no interferences at the retention time of the compounds under study.

343 The recovery results obtained for the optimized method applied to the different344 water samples are presented in Table 2.

#### 346 Table 2. Recovery (%) results obtained for the three NSAIDs spiked into the aqueous

	Recovery (%) <sup>a</sup> from Seawater			Recovery (%) <sup>a</sup> from Tap water			Recovery (%) <sup>a</sup> from Surface water			
Analyte	Spiking level (µg L <sup>-1</sup> )			Spiki	Spiking level (µg L <sup>-1</sup> )			Spiking level (µg L <sup>-1</sup> )		
	1.2	3	5	1.2	3	5	1.2	3	5	
NPX	110±6	118±1	119±1	76±2	85±3	95±5	107±3	113±1	118±3	
DCF	111±5	113±2	119±4	107±5	102±5	104±5	95±6	92±5	107±5	
MFN	114±13	$118 \pm 5$	120±3	80±6	97±2	107±2	84±9	116±3	120±1	

samples with different characteristics. 347

348 <sup>a</sup> Mean value  $\pm$  standard deviation (n=5), which corresponds to the number of extractions performed.

349 Extraction conditions: 5 mL fortified water samples, with pH adjusted to 2, NaCl 2.5 % (w/v) (except to seawater sample, which no salt was added), 150 µL of ACN and 30 µL of 1-dodecanol.

#### 3.4 Comparison with other methods 358

Table 3 shows a comparison of the methodology developed in this work with 359

others used in the determination of NSAIDs. 360

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<sup>352</sup> The results obtained for the anti-inflammatory's recovery tests were satisfactory, ranging from 110 % and 120 % for seawater, 76 % and 107 % for tap water and 84 % and 353 354 120 % for surface water, showing that the proposed method can be used to extract NSAIDs from aqueous samples with different characteristics. For the EF, values between 355 142 and 196 were reached, indicating the good ability of SFODME to concentrate the 356 evaluated compounds. 357

### 364 Table 3. Comparison of SFODME-HPLC-UV/Vis with other methods used for the

365 quantification of NSAIDs in water samples.

Method <sup>a</sup>	Analyte	Recovery (%)	LOD (µg L <sup>-1</sup> )	Sample volume (mL)	Reference
IL-DLLME-HPLC-DAD-FLD	DCF	89	95	5	[1]
MSPE-HPLC-UV	NPX, DCF	75, 83	0.06, 0.05	50	[2]
SPE-HPLC-DAD	NPX, DCF	90, 84	5.8, 23.3	1000	[41]
SPE-SUPRAS-HPLC-UV	DCF, MFN	n.a	0.4, 1.0	30	[42]
MIP-SPE-HPLC-DAD	NPX, DCF	n.a	300, 400	50	[43]
MIP-SPE-UHPLC-MS/MS	NPX, DCF	n.a	0.3, 0.7	50	[43]
US-IL-DLLME-UHPSFC-PDA	NPX, DCF	81, 100	0.31, 2.26	10	[44]
SFODME-HPLC-UV	NPX, DCF, MFN	116, 90, 109	0.09, 0.25, 0.15	5	This work

<sup>a</sup> SUPRAS: supramolecular solvent; MIP: molecular imprinted polymer extraction; IL: ionic liquids;

367 MSPE: magnetic solid-phase extraction; US-IL: ultrasound-assisted ionic liquid; UHPSFC-PDA: ultra-

368 high performance supercritical fluid chromatography coupled to photo-diode array detector; *n.a.* not

applied.

370 Comparing the hereby developed extraction method with other methods reported in literature, for the three analytes under study, and presented in Table 3 it is possible to 371 identify several advantages. Comparing with IL-DLLME-HPLC-DAD-FLD [1], our 372 study obtained similar recoveries, but a LOD 380 times lower. Moreover, the reported 373 method was only used for DCF. Alinezhad et al. [2] used a larger volume of sample, a 374 factor that can influence the LOD values, but even so the detection limit obtained in this 375 study for NPX was similar. Comparing with SPE-HPLC-DAD [41], the method presented 376 377 in this study showed similar efficiency in terms of recovery results, however lower LOD 378 were obtained. Comparing the SPE-SUPRAS-HPLC-UV [42] and the results obtained by Martinez-Sena, et al. [39] using MIP-SPE-HPLC-DAD, SFODME developed in this 379 work obtained lower detection limits, using a lower sample volume. When comparing 380 with results obtained by Martinez-Sena, et al. [43] but using a MS detector, lower 381

detection limits were obtained, even using a less expensive detection instrument.
Comparing with US-IL-DLLME-UHPSFC-PDA [44], our study presented the following
advantages: lower LOD and the use of lower sample volume.

Concluding, the method developed in this work showed a significant ability to detect low concentrations of NSAIDs in water, using low sample volume and a cheap, simple, efficient and fast technique. In general, SFODME proves to be quite advantageous concerning other techniques due to its simplicity, because it makes use of a small amount of organic solvents, has a low cost, is fast and still provides satisfactory results, being an excellent tool for the determination of anti-inflammatory drugs in aqueous samples with different characteristics.

## 392 **4. CONCLUSIONS**

A method using SFODME combined with HPLC and UV/Vis detector was 393 394 developed and validated, allowing the determination of NPX, DCF and MFN antiinflammatories, in aqueous samples. The use of 1-dodecanol as extracting solvent in the 395 optimized methodology was chosen due to low cost and low toxicity compared to other 396 organic solvents. Its capacity to melt at room temperature avoids subsequent separation 397 398 step, reducing the analysis time. The high enrichment factors obtained allowed to determine NSAIDs in different matrix of natural waters, at trace levels. The LOD and 399 LOQ reached compare with the best methods published in the scientific literature. Besides 400 401 this, the method did not demonstrate a matrix effect that would compromise its effectiveness, being possible to apply it successfully in samples of seawater, surface water 402 and tap water. 403

404

## 406 **Declarations**

## 407 **Conflicts of interests**

The authors declare that they have no known competing financial interests or personalrelationships that could have appeared to influence the work reported in this paper.

410

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#### 423 **REFERENCES**

	424	[1] Toledo-Neira	C, Álvarez-Lueje A	. Ionic liquids for	r improving the	e extraction of
--	-----	------------------	--------------------	---------------------	-----------------	-----------------

- 425 NSAIDs in water samples using dispersive liquid-liquid microextraction by high
- 426 performance liquid chromatography-diode array-fluorescence detection. Talanta
- 427 2015; https://doi.org/10.1016/j.talanta.2014.11.067
- 428 [2] Alinezhad H, Amiri A, Tarahomi M, Maleki B. Magnetic solid-phase extraction of
- 429 non-steroidal anti-inflammatory drugs from environmental water samples using
- 430 polyamidoamine dendrimer functionalized with magnetite nanoparticles as a
- 431 sorbent. Talanta 2018; https://doi.org/10.1016/j.talanta.2018.02.069

- 432 [3] Beldean-Galea MS, Coman V, Thiébaut D, Vial J. Determination of four acidic
- 433 nonsteroidal anti-inflammatory drugs in wastewater samples by dispersive liquid-
- 434 liquid microextraction based on solidification of floating organic droplet and high-
- 435 performance liquid chromatography. J. Sep. Sci. 2015;
- 436 https://doi.org/10.1002/jssc.201400933
- 437 [4] Charuaud L, Jardé E, Jaffrézic A, Liotaud M, Goyat Q, Mercier F, Le Bot B.
- 438 Veterinary pharmaceutical residues in water resources and tap water in an intensive
- 439 husbandry area in France. Sci. Total Environ. 2019;
- 440 https://doi.org/10.1016/j.scitotenv.2019.01.303
- 441 [5] Mezzelani M, Gorbi S, Da Ros Z, Fattorini D, d'Errico G, Milan M, Bargelloni L,
- 442 Regoli F. Ecotoxicological potential of non-steroidal anti-inflammatory drugs
- 443 (NSAIDs) in marine organisms: Bioavailability, biomarkers and natural occurrence

444 in *Mytilus galloprovincialis*. Mar. Environ. Res. 2016;

- 445 https://doi.org/10.1016/j.marenvres.2016.03.005
- [6] Näslund J, Fick J, Asker N, Ekman E, Larsson DGJ, Norrgren L. Diclofenac affects
- 447 kidney histology in the three-spined stickleback (*Gasterosteus aculeatus*) at low
- 448 mg/L concentrations. Aquat. Toxicol. 2017;
- 449 https://doi.org/10.1016/j.aquatox.2017.05.017
- 450 [7] Matozzo V, Rova S, Marin MG. The nonsteroidal anti-inflammatory drug,
- 451 ibuprofen, affects the immune parameters in the clam *Ruditapes philippinarum*,
- 452 Mar. Environ. Res. 2012; https://doi.org/10.1016/j.marenvres.2012.06.003
- 453 [8] Abdolmohammad-Zadeh H, Morshedzadeh F, Rahimpour E. Trace analysis of

- 454 mefenamic acid in human serum and pharmaceutical wastewater samples after pre455 concentration with Ni-Al layered double hydroxide nano-particles. J. Pharm. Anal.
- 456 2014; https://doi.org/10.1016/j.jpha.2014.04.003
- 457 [9] Bonnefille B, Gomez E, Courant F, Escande A, Fenet H. Diclofenac in the marine
- 458 environment: A review of its occurrence and effects. Mar. Pollut. Bull. 2018;
- 459 https://doi.org/10.1016/j.marpolbul.2018.04.053

- 460 [10] Wojcieszyńska D, Guzik U. Naproxen in the environment: its occurrence, toxicity
- to nontarget organisms and biodegradation. Appl. Microbiol. Biotechnol. 2020;
  https://doi.org/10.1007/s00253-019-10343-x

[11] Antonić J, Heath E. Determination of NSAIDs in river sediment samples. Anal.

- 464 Bioanal. Chem. 2007; https://doi.org/10.1007/s00216-006-0947-71
- 465 [12] Conaghan PG. A turbulent decade for NSAIDs: Update on current concepts of
- 466 classification, epidemiology, comparative efficacy, and toxicity. Rheumatol. Int.
- 467 2012; https://doi.org/10.1007/s00296-011-2263-6
- 468 [13] Parolini M, Binelli A. Sub-lethal effects induced by a mixture of three non-
- steroidal anti-inflammatory drugs (NSAIDs) on the freshwater bivalve *Dreissena*
- 470 *polymorpha*, Ecotoxicology 2012; https://doi.org/10.1007/s10646-011-0799-6
- 471 [14] Sanjuan-Reyes N, Gómez-Oliván LM, Galar-Martínez M, Vieyra-Reyes P, García-
- 472 Medina S, Islas-Flores H, Neri-Cruz N. Effluent from an NSAID-manufacturing
- 473 plant in Mexico induces oxidative stress on *Cyprinus Carpio*. Water. Air. Soil
- 474 Pollut. 2013; https://doi.org/10.1007/s11270-013-1689-8

475	[15] Xu C, Niu L, Guo H, Sun X, Chen L, Tu W, Dai Q, Ye J, Liu W, Liu J. Long-term
476	exposure to the non-steroidal anti-inflammatory drug (NSAID) naproxen causes
477	thyroid disruption in zebrafish at environmentally relevant concentrations. Sci.
478	Total Environ. 2019; https://doi.org/10.1016/j.scitotenv.2019.04.323
479	[16] Caldas SS, Rombaldi C, de Oliveira Arias JL, Marube LC, Primel EG. Multi-
480	residue method for determination of 58 pesticides, pharmaceuticals and personal
481	care products in water using solvent demulsification dispersive liquid-liquid
482	microextraction combined with liquid chromatography-tandem mass spectrometry.
483	Talanta 2016; https://doi.org/10.1016/j.talanta.2015.06.047
484	[17] Eslami A, Amini MM, Yazdanbakhsh AR, Rastkari N, Mohseni-Bandpei A,
485	Nasseri S, Piroti E, Asadi A. Occurrence of non-steroidal anti-inflammatory drugs
486	in Tehran source water, municipal and hospital wastewaters, and their
487	ecotoxicological risk assessment. Environ. Monit. Assess. 2015;
488	https://doi.org/10.1007/s10661-015-4952-1
489	[18] Paíga P, Lolić A, Hellebuyck F, Santos LHMLM, Correia M, Delerue-Matos C.
490	Development of a SPE-UHPLC-MS/MS methodology for the determination of
491	non-steroidal anti-inflammatory and analgesic pharmaceuticals in seawater. J.
492	Pharm. Biomed. Anal. 2015; https://doi.org/10.1016/j.jpba.2014.06.017
493	[19] Wolecki D, Caban M, Pazdro K, Mulkiewicz E, Stepnowski P, Kumirska J.
494	Simultaneous determination of non-steroidal anti-inflammatory drugs and natural
495	estrogens in the mussels Mytilus edulis trossulus. Talanta 2019;
496	https://doi.org/10.1016/j.talanta.2019.03.062

- 497 [20] Wang R, Li W, Chen Z. Solid phase microextraction with poly(deep eutectic
- 498 solvent) monolithic column online coupled to HPLC for determination of non-
- 499 steroidal anti-inflammatory drugs. Anal. Chim. Acta 2018;
- 500 https://doi.org/10.1016/j.aca.2018.02.024
- 501 [21] Viñas P, Campillo N, Andruch V. Recent achievements in solidified floating
- 502 organic drop microextraction. Trends Anal. Chem. 2015;
- 503 https://doi.org/10.1016/j.trac.2015.02.005
- 504 [22] Spietelun A, Marcinkowski L, de la Guardia M, Namiesnik J. Green aspects,
- 505 developments and perspectives of liquid phase microextraction techniques.
- 506 Talanta. 2014; https://doi.org/10.1016/j.talanta.2013.10.050
- 507 [23] Kokosa JM. Recent trends in using single-drop microextraction and related
- techniques in green analytical methods. Trends Anal. Chem. 2015;
- 509 https://doi.org/10.1016/j.trac.2015.04.019
- 510 [24] Tang S, Qi T, Ansah PD, Fouemina JCN, Shen W, Basheer C, Lee HK. Single-
- 511 drop microextraction. Trends Anal. Chem. 2018;
- 512 https://doi.org/10.1016/j.trac.2018.09.016
- 513 [25] Manso J, Larsson E, Jönsson JÅ. Determination of 4'-isobutylacetophenone and
- other transformation products of anti-inflammatory drugs in water and sludge from
- 515 five wastewater treatment plants in Sweden by hollow fiber liquid phase
- 516 microextraction and gas chromatography-mass spectrometry. Talanta 2014;
- 517 https://doi.org/10.1016/j.talanta.2014.02.056
- 518 [26] da Silva GS, Lima DLD, Esteves VI. Salicylic acid determination in estuarine and

519	riverine waters using hollow fiber liquid-phase microextraction and capillary zone
520	electrophoresis. Environ. Sci. Pollut. Res. 2017; https://doi.org/10.1007/s11356-
521	017-9183-2

- 522 [27] Mansour FR. Khairy MA. Pharmaceutical and biomedical applications of
- 523 dispersive liquid–liquid microextraction. J. Chromatogr. B. 2017;
- 524 https://doi.org/10.1016/j.jchromb.2017.07.055
- 525 [28] Rezaee M, Assadi Y, Milani Hosseini MR, Aghaee E, Ahmadi F, Berijani S.
- 526 Determination of organic compounds in water using dispersive liquid-liquid
- 527 microextraction. J. Chromatogr. A. 2006;
- 528 https://doi.org/10.1016/j.chroma.2006.03.007
- 529 [29] Khalili Zanjani MR, Yamini Y, Shariati S, Jönsson JÅ. A new liquid-phase
- 530 microextraction method based on solidification of floating organic drop. Anal.

531 Chim. Acta. 2007; https://doi.org/10.1016/j.aca.2006.12.049

- 532 [30] Chang CC, Huang SD. Determination of the steroid hormone levels in water
- samples by dispersive liquid-liquid microextraction with solidification of a floating
- organic drop followed by high-performance liquid chromatography. Anal. Chim.
- 535 Acta. 2010; https://doi.org/10.1016/j.aca.2010.01.003
- 536 [31] Mansour FR, Danielson ND. Solidification of floating organic droplet in dispersive
- 537 liquid-liquid microextraction as a green analytical tool. Talanta. 2017;
- 538 https://doi.org/10.1016/j.talanta.2017.03.084
- 539 [32] Shukri DSM, Sanagi MM, Ibrahim WAW, Abidin NNZ, Aboul-Enein HY. Liquid
- 540 Chromatographic Determination of NSAIDs in Urine After Dispersive Liquid–

541	Liquid Microextraction Based on Solidification of Floating Organic Droplets.
542	Chromatographia 2015; https://doi.org/10.1007/s10337-015-2920-0
543	[33] Sena LCS, Matos HR, Dórea HS, Pimentel MF, de Santana DCAS, de Santana
544	FJM. Dispersive liquid-liquid microextraction based on solidification of floating
545	organic drop and high-performance liquid chromatography to the analysis of
546	cocaine's major adulterants in human urine. Toxicology. 2017;
547	https://doi.org/10.1016/j.tox.2016.04.008
548	[34] Martín J, Santos JL, Aparicio I, Alonso E. Determination of hormones, a
549	plasticizer, preservatives, perfluoroalkylated compounds and a flame retardant in
550	water samples by ultrasound-assisted dispersive liquid-liquid microextraction
551	based on the solidification of a floating organic drop. Talanta. 2015;
552	https://doi.org/10.1016/j.talanta.2015.04.089
553	[35] Guiñez M, Martinez LD, Fernandez L, Cerutti S. Dispersive liquid-liquid
554	microextraction based on solidification of floating organic drop and fluorescence
555	detection for the determination of nitrated polycyclic aromatic hydrocarbons in
556	aqueous samples. Microchem. 2017; https://doi.org/10.1016/j.microc.2016.10.020
557	[36] Arismendi D, Becerra-Herrera M, Cerrato I, Richter P. Simultaneous determination
558	of multiresidue and multiclass emerging contaminants in waters by rotating-disk
559	sorptive extraction-derivatization gas chromatography/mass spectrometry. Talanta
560	2019; https://doi.org/10.1016/j.talanta.2019.03.120
561	[37] Sodré FF, Sampaio TR. Development and application of a SPE-LC-QTOF method
562	for the quantification of micropollutants of emerging concern in drinking waters

- from the Brazilian capital. J. Em. Con. 2020;
- 564 https://doi.org/10.1016/j.emcon.2020.01.001
- 565 [38] Ide AH, Osawa RA, Marcante LO, Pereira JC, de Azevedo JCR. Occurrence of
- 566 Pharmaceutical Products, Female Sex Hormones and Caffeine in a Subtropical
- 567 Region in Brazil. Clean Soil Air Water 2017;
- 568 https://doi.org/10.1002/clen.201700334
- 569 [39] Pereira CDS, Maranho LA, Cortez FS, Pusceddu FH, Santos AR, Ribeiro DA,
- 570 Augusto César, Guimarães LL. Occurrence of pharmaceuticals and cocaine in a
- 571 Brazilian coastal zone. Sci. Total Environ. 2016;
- 572 https://doi.org/10.1016/j.scitotenv.2016.01.051
- 573 [40] Chaves MJS, Barbosa SC, Malinowski MM, Volpato D, Castro IB, Franco TCRS,
- 574 Primel EG. Pharmaceuticals and personal care products in a Brazilian wetland of
- 575 international importance: Occurrence and environmental risk assessment. Sci.
- 576 Total Environ. 2020; https://doi.org/10.1016/j.scitotenv.2020.139374
- 577 [41] Ascar L, Ahumada I, López A, Quintanilla F, Leiva K. Nonsteroidal anti-
- inflammatory drug determination in water samples by HPLC-DAD under isocratic
- conditions. J. Braz. Chem. Soc. 2013; https://doi.org/10.5935/0103-5053.20130150
- 580 [42] Rezaei F, Yamini Y, Moradi M, Ebrahimpour B. Solid phase extraction as a
- 581 cleanup step before microextraction of diclofenac and mefenamic acid using
- nanostructured solvent. Talanta 2013; https://doi.org/10.1016/j.talanta.2012.11.035
- [43] Martinez-Sena T, Armenta S, de la Guardia M, Esteve-Turrillas FA. Determination
  of non-steroidal anti-inflammatory drugs in water and urine using selective
  - 28

585	molecular imprinted polymer extraction and liquid chromatography. J. Pharm.
586	Biomed. Anal. 2016; https://doi.org/10.1016/j.jpba.2016.08.006
587	[44] Ji Y, Du Z, Zhang H, Zhang Y. Rapid analysis of non-steroidal anti-inflammatory
588	drugs in tap water and drinks by ionic liquid dispersive liquid-liquid
589	microextraction coupled to ultra-high performance supercritical fluid
590	chromatography. Anal. Methods 2014; https://doi.org/10.1039/c4ay01305k
591	
592	