Responses in estuarine macrobenthic invertebrate assemblages to trace metal contaminated sediments

Anthony A. Chariton, Bach. App. Sci. (Hons)

Ecochemistry Laboratory Applied Ecology Research Group University of Canberra Canberra ACT 2601

This thesis is submitted in fulfilment of the requirements for the degree of Doctor of Philosophy in Applied Science (Environmental Chemistry). University of Canberra May, 2005 In any general discussion of structure, relating to an isolated part of the Universe, we are faced with an initial difficulty in having no a priori criteria as to the amount of structure it is reasonable to expect. We do not, therefore, always know, until we have had a great deal of empirical experience, whether a given example of structure is very extraordinary, or a mere trivial expression of something which we may learn to expect all the time.

G.A. Hutchinson, 1953

ii

Copyright

This thesis may be freely copied and distributed for private use and study. However, no part of this thesis, or the information contained herein, may be included in a publication or referred to in a publication without prior written permission from the author. Any reference must be fully acknowledged.

Anthony A. Chariton

17/5/05

Date

iv

Acknowledgments

This thesis, and the research from which it is founded, would not have come to fruition without the assistance, advice, support, skill and intellect of numerous people.

To my primary supervisor, mentor and friend Prof. Bill Maher, I thank you for your interest in my research and career, and the support you have provided throughout my academic life. You have continually permitted and supported my diverse interests, and have always endeavoured to find a way for me to continue my interests, both academically and financially. I also wish to thank you for your assistance in the field and, the 'open door' policy to queries and problems, and for your after hours commitment. I apologise for the hair loss, greying and any chronic health problems that I may have attributed too. However, your grievances about my effects on your health are correlative, and hence, unsubstantiated.

To my supervisor Dr Tony Roach (NSW Department of Environment and Conservation), your knowledge of marine ecology and biometry has been pivotal in both this project and my career as a whole. I appreciate all the time you have made available, your attention and detail in refining the chapters, and your overall commitment to the project. Finally, I wish to thank you for your continual support of my professional career, and the guidance you have provided.

I wish to express my gratitude to my collaborators and colleagues at the CSIRO Centre for Environmental Contaminants Research: Dr Graeme Batley, Dr Stuart Simpson, Dr Jenny Stauber and Dr Simon Apte, and Dr Catherine King. It has been both and honour and pleasure to work with the group, and to be able to tap into their

v

vast wealth of knowledge. I would especially like to thank Dr Graeme Batley for his guidance and unconventional approach to multi-disciplined problems. I am also most grateful to Dr Stuart Simpson for his skills and knowledge in trace metal biogeochemistry, advice on sediment spiking and his patience in teaching me various analytical skills. Thank you to Dr John Chapman, Dr Ross Hyne and Ms Sharyn Gale from the NSW EPA, it has been a pleasure to work with you all.

My deepest thanks go to Frank Krikowa (University of Canberra) who developed the analytical technique for the trace metal analysis of the porewaters, and who performed a vast majority of the analysis. I appreciate all the time, effort and skill which went into the method development. As per usual, I would be lost without your help.

I would especially like to thank my fellow environmental chemistry post-grads Simon Foster, Jason Kirby and Daniel Spooner for their commitment to my field work and assistance in the laboratory. I appreciate that you all had your own research to complete, and that field conditions were often less than desirable – however, we still always had a great time. I would need to write another chapter to thank you individually for all that you guys have done.

A special thank you to Dr Dianne Jolley (University of Wollongong), who initially directed me towards environmental chemistry and marine sciences, assisted me in the field, guided my career, and who continues to play a pivotal role in my research. Thank you to Mathew Barwick (Fisheries Research and Development Council) for his continual assistance in the field, support, fisheries permits, excellent joke telling, and friendship over the years.

Thank you to Dr Allan Jones and Ms. Anna Murray from the Australian Museum. Wayne Robinson (University of the Sunshine Coast), Dr Bob Clarke (Plymouth

vi

Marine Laboratory), Dr David Williams (University of Canberra), Dr Jim Hone (University of Canberra) for their statistical advice. Dr Phil Gibbs (NSW Fisheries) for the many discussions on Lake Macquarie. Dr Chris Hickey (NIWA) for his help with the porewater collection devices which he developed. Anne Taylor (University of Canberra) for the nutrient analysis. Huge thanks to Scott Rowan, Rod Strauss, Melissa Spooner, Ali, Rajani, Enzo, Fiona, Jules, Mirin, and Monique for their help in the field. Thanks to the technical, support and administrative staff in the School and of Health, Design and Science for their day to day help, especially Hailey, Sue, Brian and Patrick.

I would also like to thank all my friends both in Sydney and Canberra for their continual support and understanding, and I apologise for my hermit like characteristics over the last 18 months. I'm also grateful to Scott Keyworth and Jim Barrett at Murray-Darling Basin Commission for allocating me time to write up during my term of employment.

Without the support of the NSW Environmental Trust neither the research nor the collaborations would have been possible. I further wish to thank the Trust for continuing to support the research of contaminants in the marine and estuarine systems.

From the bottom of my heart I wish to thank my family for their support over the years – many more than I initially envisaged. You have continually shown great interest in what I am doing and my achievements along the way. For this, I am eternally indebt.

Finally, I wish to thank my beautiful wife Gabrielle, who would have to be one of the most tolerant people I know. I thank you all your support, increase in domestic duties,

vii

loss of holidays, boring weekends when I am writing, brilliant editing skills, commitment and love.

Abstract

Three approaches were employed to examine the effects of elevated sediment trace metal concentrations on estuarine/marine macrobenthic invertebrate assemblages. The initial study examined macroinvertebrate communities along a known polymetallic gradient, Lake Macquarie, NSW (gradient study). The second study experimentally tested if sediments sourced from different locations within Lake Macquarie differentially influenced the recolonisation of benthic invertebrates. The third study investigated the different recolonisation patterns of benthic invertebrates into sediments spiked with increasing concentrations of sediment-bound cadmium.

In the Lake Macquarie gradient study, four locations (Cockle Bay, Warner's Bay, Kooroora Bay and Nord's Wharf) were sampled in winter 2000 and summer 2003 using a hierarchical design (location > site > plot). On both sampling occasions, the sediments showed strong gradients in lead, cadmium and zinc concentrations emanating from the Cockle Bay industrialised region in the lake's north, with concentrations being significantly lower in the most southern and less urbanised location (Nord's Wharf). In general, concentrations of lead, cadmium and zinc in the sediments increased among locations in the following order: Nord's Wharf > Kooroora Bay > Warner's Bay > Cockle Bay. AVS/SEM analyses indicated that in some sites in Cockle Bay, and to a lesser extent Warner's Bay, SEM concentrations exceeded their molar equivalence of AVS, indicating the potential for trace metals to be labile within the porewaters. Granulometry also changed along the gradient, with a

viii

higher proportion of silt/clay occurring in the locations with high metal concentrations. Conversely, the percentage of total organic carbon was higher in the less contaminated locations.

In winter 2000, changes in benthic communities along the gradient supported the *a* priori hypotheses, with diversity and richness being greater in locations with lower concentrations of metals. Polychaetes were most numerous in Cockle Bay and Warner's Bay, whilst bivalves and gastropods were more abundant in Nord's Wharf and Kooroora Bay. Crustaceans were more numerous in Nord's Wharf; with all other locations having similar, lower, abundances. Ordination maps of the assemblages provided relatively clear separation of the assemblages among locations, with non-parametric multivariate analysis of variance (NPMANOVA) and subsequent pair-wise comparisons finding significant differences among the assemblages from all locations. SIMPER analyses found the highest level of dissimilarity was between the Nord's Wharf and Cockle Bay assemblages – primarily attributable to differences in the relative contributions of isopods; tellenid bivalves; and the polychaete families Spionidae, Opheliidae and Nephytidae. Weighted Spearman rank correlations (BIO-ENV) identified cadmium ($P_w = 0.74$) as the strongest environmental (single or combination) variable to correlate with biotic assemblages.

Benthic patterns along the gradient were less defined in summer 2003 due to a dramatic reduction in the abundance and diversity of fauna in Nord's Wharf. This decline was possibly attributable to a sustained reduction in salinity caused by a prolonged rainfall event. With the exception of Nord's Wharf, trends in the community indices and abundances of key taxa among the other locations were similar to those reported in winter 2000. Multivariate analyses discriminated the benthic assemblages from the four locations, with the findings from the NPMANOVA pair-wise comparisons indicating that the assemblages from all four locations were

ix

significantly different. SIMPER analyses showed the highest level of dissimilarity was between Nord's Wharf and Warner's Bay, with these differences being primarily attributable to their relative abundances of amphipods and polychaetes from the families Spionidae, Cirratulidae, Opheliidae and Capitellidae. BIOENV found that the combination of the sedimentary concentrations of cadmium and iron provided the best correlation ($P_w = 0.73$) with biotic patterns, with similar correlations occurring with the addition of lead and its covariate, zinc ($P_w = 0.72$).

The combined findings from the gradient study established a strong correlation between trace metal concentrations within the sediments and suite of univariate and multivariate measurements. The low abundance and diversity of fauna in Nord's Wharf in the summer of 2003 highlighted the dynamic changes which can occur in the distributions of macrobenthic invertebrates. Although the study indicated that there was a strong relationship between trace metal concentrations and benthic community structure, the study was correlative, and requires subsequent experimental testing to confirm the causality of the observed relationships.

The second component of the research was a translocation experiment using benthic recolonisation as an end-point. The experiment was performed to identify if the sediments, and not location, were influencing the composition of benthic assemblages in Lake Macquarie. Sediments were collected from three locations (Cockle Bay, Warner's Bay and Nord's Wharf), defaunated, and transplanted in three new locations along the south-east edge of the lake. At each location, 10 containers of each treatment were randomly placed in the sediment and allowed to recolonise for 22 weeks. Upon retrieval, the benthic communities were sampled and enumerated in conjunction with a variety of chemical and sedimentary measurements. Ten replicate invertebrate samples were also collected in the sediments adjacent to the experiment (ambient samples) at the completion of the experiment. Due to human interference,

х

the containers from only two locations were analysed.

Upon retrieval, pH and redox profiles of the sediments were similar to those expected in natural sediments. In general, concentrations of metals were low in the porewaters; however, iron precipitation on the porewater collection devices may have artificially increased the diffusion of metals, increasing concentrations near the sediment-water interface. Concentrations of SEM exceeded their AVS equivalence in some samples taken from the Cockle Bay and Warner's Bay treatments.

Two-way ANOVAs found significant interactions between location and sediment treatments in diversity, evenness and the number of polychaetes, as well as significant differences in the number of capitellids and crustaceans among locations. *Post-hoc* comparisons of means found the Nord's Wharf sediment contained a higher mean number of individuals than the other treatments, including the ambient samples. nMDS ordination plots for both locations provided poor graphical discrimination of the assemblages among treatments; however, NPMANOVA detected significant location and treatment interactions. In both locations, pair-wise comparisons indicated that the assemblages within the Nord's Wharf treatments were significantly different to the Cockle Bay, Warner's Bay and ambient assemblages. No significant differences were detected between the Cockle Bay and Warner's Bay assemblages at either location. SIMPER analyses found the highest level of dissimilarity occurred between the ambient assemblages in Location 2 and the Nord's Wharf treatment, primarily due to the relative difference in the abundances of Capitellidae, Spionidae, Oweniidae, Nereididae and isopods among the assemblages.

The findings from the translocation experiment suggest that the sediments are influencing the recolonisation of benthos. However, because differences were not detected between the Cockle Bay and Warner's Bay treatments, the approach used in

xi

the study shows potential as an *in situ* technique which could be used to assess the potential ecological risks of sediments from specific locations. Excluding cost and time considerations, the technique's primary disadvantage is the lack of a true control. As a result, the technique can only identify if the sediments are modifying benthic recolonisation, and not causality.

The final component of the research experimentally tested if elevated concentrations of sediment-bound cadmium affected benthic invertebrate recolonisation. Sediments from the south coast of New South Wales (Durras Lake) were defaunated, and spiked with cadmium under anaerobic conditions to obtain three targeted cadmium concentrations: control (<0.1 Cd μ g/g), Low-Cd (15 Cd μ g/g) and High-Cd (150 Cd μ g/g). The physio-chemical properties of the waters and porewater concentrations of cadmium were monitored over a 28-day equilibration period, with declines in pH mediated with the addition of NaOH_(aq). At the end of the equilibration period, porewater concentrations of cadmium were low in the Low-Cd and High-Cd treatments (maximum <1.5 μ g/L in High-Cd), and below the detection limit in the control. Cadmium was not detected in the control sediments, with concentrations in the Cd-Low and Cd-High sediments exceeding their targeted concentrations, with final mean concentrations of 17 μ g/g and 183 μ g/g, respectively.

The experimental design was similar to that employed in the translocation experiment, with 10 containers from each treatment transplanted into the sediments at three locations within Lake Macquarie. After 20 weeks, the containers were collected, along with benthic invertebrate samples from the ambient sediments. Data was not used from Location C due to extensive sediment deposition on the transplanted treatments. Significant declines occurred in the concentrations of cadmium in both the Low-Cd and High-Cd sediments, with the greatest loss occurring in the surficial sediments. The loss of cadmium was probably due to the differential loss of the fine

xii

fraction through physical means (hydrodynamic) rather than fluxing, as it assumed that the cadmium was primarily sediment-bound and relatively insoluble under anoxic conditions. Mean porewater concentrations of cadmium were below the detection limit in the control treatments; < 1 μ g/L in the Low-Cd treatment, and generally < 2μ g/L in the High-Cd, with the exception of some samples in Location B (maximum 5.6 μ g/L). Concentrations of ammonia were low in the porewaters from the surficial sediments, with concentrations being significantly higher, and potentially toxic, in the anoxic porewaters (7 cm depth).

In comparison to the previous recolonisation experiment, the number of individuals which recolonised the cadmium-spiked treatments was low, and significantly lower than the mean number of individuals sampled in the ambient sediments. No significant differences were detected among the treatments or locations (and their interactions) in diversity (H'), richness (d) or evenness (J). The number of polychaetes and molluses significantly differed among the treatments, with *post-hoc* analyses indicating these differences were not among the cadmium-spike treatments, but were due to a greater mean abundance of these taxa in the ambient sediments. A significant interaction between treatment and location was detected in the mean abundance of crustaceans, with the ambient sediments having significantly lower mean abundances in both Location A and B. Ordination plots of the experiments in Location A and B provided poor graphical discrimination among the spiked treatments, although the ambient assemblages appear to be separated from the cadmium-spiked assemblages. NPMANOVA detected a significant interaction between treatments and locations, as well as among treatments. In both Location A and B, pair-wise analyses found the assemblages in the ambient sediments to be significantly different to the assemblages in all three cadmium treatments, with no differences being detected among the latter. SIMPER analyses found the highest levels of dissimilarity occurred between the spike-treatments and the ambient sediments, with these differences being primarily

xiii

due to the relatively higher abundance of decapods in the spiked treatments, and capitellids in the ambient sediments.

The cadmium-spiking component of the experiment clearly illustrated that artificially increasing the trace metal concentrations of metals in estuarine sediments is a complex process which needs to be performed in a methodological manner in order to obtain homogenous treatments with low porewater concentrations, and minimal artefacts. Furthermore, the results confirmed that the equilibration time for sediments can be extensive (several weeks), even in the case of organically rich sediments. The timing of the experiment (commenced late summer, February, 2003) appears to the major factor for the relatively low recolonisation rates, with the experiment missing the main larval recolonisation period between spring and early summer. Even in the highest treatment, elevated concentrations of cadmium did not appear to affect benthic recolonisation. This finding is supported by other experimental studies which suggest that concentrations of a single isolated metal must considerably exceed current guideline values (or contain high porewater concentrations) in order to elicit a biological effect. Nevertheless, as trace metals generally co-occur with other contaminants – with the response of multiple contaminants being possibly additive or synergistic - a conservative guideline value may be suitable in the interim as a precautionary measure.

The findings of this thesis suggest that elevated concentrations of trace metal mixtures in estuarine sediments can affect the structure and composition of benthic communities; however, identifying causality is difficult. Although there has been an increase in the use of manipulative field experiments as a means of reducing the confounding influence of covariables found in field studies, this approach also has limitations, e.g. spatial and temporal scale issues, container effects, cost and biogeochemical changes to the sediments. Measuring stress at a community level is a

xiv

fundamental component of estuarine risk assessment programs; and in isolation this approach can produce subjective and confounded findings. In order to accurately assess the risks associated with trace metal contaminated sediments, an integrated approach (e.g. weight of evidence) is required, one which uses multiple lines of evidence sourced from various chemical, environmental biological measurements.

 $\mathbf{x}\mathbf{v}$

Table of Contents

Statement of originality	iii
Copyright	iv
Acknowledgments	v
Abstract	viii
List of figures	xxiv
List of tables	xxix

Chapter One Introduction

1.1 Rationale	1
1.2 Aim of the research	6

Chapter Two

Literature re	eview7
2.1 Ai	m of the literature review7
2.2 Tr	race metal biogeochemistry7
	2.2.1 Fundamental biogeochemical processes of estuarine/marine
	sediments7

xvi

2.2.2 Adsorption and partitioning of metals in sediments
2.2.3 Acid-volatile sulfide/simultaneously extractable metal model10
2.2.4 The effect of bioturbation on sediment biogeochemistry11
2.3 The ecology of estuarine macrobenthic fauna13
2.3.1 The ecological importance of macrobenthic fauna13
2.3.2 The influence of environmental variables on benthic communities
13
Salinity14
Granulometry16
Additional important environmental factors
2.3.3 Spatial and temporal patterns in benthic communities
2.3.4 Larval processes
2.4 Response of macrobenthic communities to trace metal24
2.4.1 Bioavailability and exposure routes for trace metal. accumulation
2.4.2 Soft-sediment benthic communities as an ecotoxicological end-point
2.4.3 Detecting and measuring ecological stress
Univariate measurements
Multivariate measurements
Graphical techniques32

xvii

Linking benthic	communities t	o environmental	variables33
_			

Chapter Three

Changes	in	macrobenthic	communities	along	а	polymetallic
contamin	nation	gradient, Lake M	lacquarie, Austr	alia		
3.1	l Aim o	f the study			•••••	
3.2	2 Introd	luction			· · · · · ·	
3.3	8 Mater	ials and methods.	•••••••••••••••••		• • • • • • • •	42
	3.3	8.1 Study site		· · · · · · · · · · · · · · · · · · ·		42
	3.3	3.2 Experimental de	sign			45
	3.3	3.3 Collection and i	dentification of ber	nthic mac	ofauı	na45
	3.3	3.4 Sediment collect	tion and analyses			46
		Trace metals .		•••••		46
		Acid-volatile s	ulfide/simultaneou	sly extrac	ted m	etals47
		Total organic	carbon			48
		Grain size		• • • • • • • • • • • • • • • • • • • •		48
	3.	3.5 Data analysis		•••••	•••••	48
		Univariate an	alysis	•••••		48
		Graphical and	lysis	•••••	•••••	49
		Multivariate a	nalyses		• • • • • • •	49
3.	4 Resul	lts			•••••	50
	3.4	4.1 Trace metals			•••••	50

xviii

Trace metals concentrations within the sediments (2000)50
Sediment profiles (stratification)51
Trace metals concentrations within the sediments (2003)56
AVS/SEM
Grain size59
Total organic carbon60
3.4.2 Benthic communities64
Univariate measurements of benthic assemblages64
2000 sampling occasion64
Changes in macrobenthic assemblages between 2000 and 2003
65
2003 sampling occasion65
Multivariate measurements of benthic assemblages72
SIMPER (2000)
SIMPER (2003)75
3.4.3 Relationship between biological and environmental patterns76
2000
2003
3.5 Discussion
3.5.1 Trace metal concentrations within the sediments
3.5.2 Benthic community patterns85
Benthic community patterns (2000)85
Benthic community patterns (2003)

xix

al variables and benthic communities	ı environmental	p between	Relationship	3.5.3
	• • • • • • • • • • • • • • • • • • • •			
96			on	3.6 Conclusio

Chapter Four

ł

1

Recolonisation of translocated estuarine sediments by macrobenthic
communities
4.1 Aim of the experiment100
4.2 Introduction 101
4.3 Materials and methods103
4.3.1 Sediment collection, transplant locations and design103
4.3.2 Chemical analysis of the sediments and granulometry106
4.3.3 Porewater collection and analysis106
4.3.4 Data analysis108
4.4 Results 109
4.4.1 Initial trace metal concentrations, granulometry and organic content of translocated sediments
4.4.2 Trace metal concentrations in the translocated sediments110
Location-treatment interactions in trace metal concentrations
4.4.3 Physico-chemical measurements and concentrations of lead, cadmium, zinc, copper and nickel within the porewaters
4.4.4 Acid-volatile sulfides and simultaneously extractable metals122

xx

4.4.5 Univariate measurements of biota122
Location patterns in the univariate measurements of the recolonised benthic communities
Univariate attributes of macrobenthic communities among translocated treatments
Location and treatment interactions in the univariate attributes of the recolonised communities
4.4.6 Multivariate community assemblages126
4.4.7 Composition of communities within the translocated and ambient
sediments131
4.5 Discussion
4.5.1 Characteristics of the translocated sediments
4.5.2 Recolonisation of benthic assemblages
4.6 Conclusion

Chapter Five

Recolonisation of cadmium-spiked sediments by sub-tidal macrobenth	nic
assemblages1	48
5.1 Aim of the experiment1	49
5.2 Introduction 1	50
5.3 Methods 1	54
5.3.1 Collection and spiking of sediments1	54

xxi

5.3.2 Transplant, collection and analyses of treatments156
5.4 Results 156
5.4.1 Sediment characteristics and trace metal concentrations in
control sediments (Durra Lake)156
5.4.2 Porewater measurements and cadmium concentrations of the sediments prior to deployment
5.4.3 Porewater measurements and cadmium concentrations in the sediments at the completion of the experiment
5.4.4 Univariate measurements of macroinvertebrates
5.4.5 Multivariate community assemblages173
5.4.6 Composition of macroinvertebrate communities within the
cadmium-spiked and ambient sediments173
5.5 Discussion
5.5.1 Spiking and equilibration of sediments177
5.5.2 Trace mental concentrations and porewater measurements at
the completion of the experiment179
5.5.3 Comparisons between the recolonised and the established
ambient assemblages181
5.5.4 Responses in recolonised assemblages to cadmium-spiked sediments
5.6 Conclusion

xxii

Chapter Six

Synthesis1	19	9	()	
------------	----	---	---	---	--

References1	94
-------------	----

xxiii

List of figures

Chapter One: Introduction

Figure 1.1. The decision tree employed in the Interim Sediment Quality Guidelines for the assessment of contaminated sediments
Figure 1.2. A proposed extended decision tree for the ANZECC/AMCANZ (2000) Interim Sediment Quality Guidelines for assessing the risk of trace metal contaminated sediments

Chapter Two: Literature review

1

Figure 2.1. Illustration of the stratified changes which occur in oxygen, pH and metal partitioning in estuarine sediments
Figure 2.2. Changes in species richness with salinity15
Figure 2.3. Abundance biomass curves (ABC) for unpolluted and polluted locations

xxiv

Figure 3.4. Concentrations of lead with sediment column depth from sites within locations sampled in 2000 (a) Cockle Bay; (b) Warner's Bay; (c) Kooroora Bay; (d) Nord's Wharf
Figure 3.5. Concentrations of cadmium with sediment column depth from sites within locations sampled in 2000 (a) Cockle Bay; (b) Warner's Bay; (c) Kooroora Bay; (d) Nord's Wharf
Figure 3.6. Concentrations of zinc with sediment column depth from sites within locations sampled in 2000 (a) Cockle Bay; (b) Warner's Bay; (c) Kooroora Bay; (d) Nord's Wharf
Figure 3.7. Concentrations of lead (a), cadmium (b) and zinc (c) in the sediments of the sites from four locations within Lake Macquarie sampled in 2003
Figure 3.8. Sediment characteristics from the four locations within Lake Macquarie in 2000 and 2003. (a) <63 μ m (%); (b) 63-600 μ m (%); (c) 600 μ m-2 mm (%); (d) >2mm (%); and (e) total organic carbon (%)63
Figure 3.9. The mean (a) no. of individuals; (b) diversity (H'); (c) taxa richness; (d) no. of polychaetes from the four benthic sediment locations in Lake Macquarie in 2000 (blue) and 2003 (red)
Figure 3.10. The mean (a) no. of capitellids; (b) no. of crustaceans; (c) no. of bivalves; (d) no. of gastropods from the four benthic sediment locations in Lake Macquarie in 2000 (blue) and 2003 (red)
Figure 3.11. Cumulative dominance curves (%) from benthic sediments in Cockle Bay, Warner's Bay, Kooroora Bay and Nord's Wharf, Lake Macquarie: (a) 2000 and (b) 2003
Figure 3.12. MDS ordinations of Bray-Curtis similarities from fourth-root transformed species abundance data from cores taken from four benthic sediment locations within Lake Macquarie: (a) 2000 and (b) 2003
Figure 3.13. MDS ordinations of Bray-Curtis similarities from fourth-root transformed species abundance data from sites at four benthic sediment locations within Lake Macquarie in 2000; superimposed circled areas are proportional to: (a) lead concentration of sediments ($\mu g/g$); (b) cadmium concentration of sediments ($\mu g/g$); (c) cadmium concentration of sediments ($\mu g/g$); (d) manganese concentration of sediments ($\mu g/g$)

xxv

Figure 3.15. Rainfall (mm) from Earring Bay, Lake Macquarie (1993-2003): (a) annual; (b) monthly (data source, Bureau of Meteorology 2004).....90

Chapter Four: Recolonisation of translocated estuarine sediments by

macrobenthic communities

Figure 4.1. Map of Lake Macquarie, NSW illustrating the positions of the locations (Locations 1, 2 and 3) used in the sediment translocation experiment......105

Figure 4.2. Concentrations of (a) lead, (b) cadmium and (c) zinc in sediments translocated from Nord's Wharf, Cockle Bay and Warner's Bay......115

Figure 4.3. Concentrations of (a) copper, (b) nickel (c) manganese and (d) iron in sediments translocated from Nord's Wharf, Cockle Bay and Warner's Bay.....116

xxvi

Figure 4.8. The (a) total abundance; (b) diversity (H'); (c) richness; and (d) evenness (J') measured from benthic communities sampled from the three translocated sediments and ambient sediments in Locations 1 (red) and 2 (green)129
Figure 4.9. The abundance of (a) polychaetes; (b) capitellids; (c) crustaceans; and (d) molluscs sampled from the three translocated sediments and ambient sediments in Locations 1 (red) and 2 (green)
Figure 4.10. Two dimensional nMDS ordination plots for the recolonisation experiments in Location 1 (a) and Location 2 (b)132
Chapter Five: Recolonisation of cadmium-spiked sediments by sub-tidal
macrobenthic assemblages
Figure 5.1. Map of Lake Macquarie, NSW, illustrating the positions of Locations A, B and C used in the cadmium-spiked sediment experiment157
Figure 5.2. Mean porewater pH measurements from the three treatments collected during the equilibration period
Figure 5.3. Mean concentrations of cadmium collected from the porewaters after spiking
Figure 5.4. Mean sediment cadmium concentrations (+ 1 S.E.) at different sediment depths (0-8, 0-2, 2-5, 5-8 cm) from the cadmium-spiked treatments collected at the completion of the experiment from locations A and B
Figure 5.5. Percentage reduction (%) in sediment cadmium concentrations from the cadmium-spiked treatments in locations A and B
Figure 5.6. Redox (mV) measurements obtained from the porewaters of the spiked sediments at the completion of the recolonisation experiment: (a & b) Control sediments, locations A and B; (c & d) Low-Cd spiked sediments, locations A and B; (e & f) High-Cd spiked sediments, locations A and B
Figure 5.7. pH measurements obtained from the porewaters of the spiked sediments at the completion of the recolonisation experiment: (a & b) Control sediments, Locations A and B; (c & d) Low-Cd spiked sediments, Locations A and B; (e & f) High-Cd spiked sediments, Locations A and B

xxvii

Figure 5.8. Stratified cadmium concentrations from the porewaters of the spiked	
sediments at the completion of the recolonisation experiment: (a & b) Low-Cd spike	
Locations A and B; (c & d) High-Cd spike Locations A and B166	5

Figure 5.10. The abundance of (a) polychaetes, (b) molluscs and (c) crustaceans from benthic communities sampled from the three cadmium-spiked treatments and ambient sediments in Locations A (red) and B (green)......172

xxviii

List of tables

Chapter Two: Literature review

Table 2.1. A tabulated example using multiple lines of evidence to assess theecological risks of locations under a variety of scenarios27

Chapter Three: Changes in macrobenthic communities along a polymetallic contamination gradient, Lake Macquarie, Australia

Table 3.3. Summary of results of analysis of variance for sediment AVS and SEMconcentrations from samples collected in 2003. Significant differences betweenlocations detected by the SNK multiple comparisons of means are presented in theSNK row.60
Table 3.4. Summary of results of analysis of variance for different sediment grain size classes and total organic carbon samples collected in 2000 and 200361
Table 3.5. SEM/AVS results from Lake Macquarie sediments in 200362
Table 3.6. Summary of results from a three-factor analysis of variance forcommunity indices and the abundance of selected taxa sampled in Lake Macquariebenthic sediments in 2000 and 2003
Table 3.7. Significant differences detected by SNK post-hoc tests from three-factorANOVAs on community indices and the abundance of selected taxa from LakeMacquarie in 2000 and 2003
Table 3.8. Non-parametric MANOVA on Bray-Curtis distances for assemblages of

xxix

benthic organisms from four benthic sediment locations within Lake Macquarie in 2000 and 2003......74

Table 3.9. Results from among location NPMANOVA pair-wise comparisons tests for the Lake Macquarie benthic sediments gradient study in 2000 and 2003......74

Chapter Four: Recolonisation of translocated estuarine sediments by

macrobenthic communities

Table 4.1. Trace metal concentrations (mean ± 1 S.E.) in the sediments collected from Nord's Wharf, Cockle Bay and Warner's Bay in Lake Macquarie, NSW
Table 4.2. Granulometry and total organic carbon (TOC %) (mean ± 1 S.E.) ofsediments collected from Nord's Wharf, Cockle Bay and Warner's Bay in LakeMacquarie, NSW
Table 4.3. Summary of results of analysis of variance for transformed sediment tracemetal concentrations from sediments translocated from Cockle Bay, Warner's Bayand Nord's Wharf
Table 4.4. Significant differences in sediment trace metal concentrations detected bySNK post-hoc tests from the two-factor ANOVAs
Table 4.5. SEM/AVS results from the translocated sediments (0-2 cm and 2-7 cm) atthe completion of the experiment

xxx

Table 4.7 Significant differences detected by SNK post-hoc tests in communityindices and the abundance of selected taxa sampled in the translocated and ambientsediments from Locations 1 and 2 in Lake Macquarie128

Table 4.9. Among treatments results (including ambient fauna) within locations from NPMANOVA pair-wise comparisons tests on recolonised and resident fauna.....133

Table 4.10. Among location results (including ambient fauna) within treatments from NPMANOVA pair-wise comparisons tests on recolonised and resident fauna.....133

Table 4.12. Summary of results from SIMPER analyses showing Bray-Curtis % dissimilarities among faunal assemblages in the translocated treatments and ambient sediments, and the five taxa contributing most to these dissimilarities......135

Chapter Five: Recolonisation of cadmium-spiked sediments by sub-tidal

macrobenthic assemblages

Table 5.1.Granulometry, total organic carbon and acid-volatile sulfideconcentrations of Lake Durras sediments after defaunation159

Table 5.3. Mean ammonia (μ g/L) and corrected unionised ammonia concentrations from porewaters collected at 1 and 7 cm sediment depths......166

xxxi

Table 5.4. Summary of results of from two-factor analysis of variance forcommunity indices and the abundance of selected macroinvertebrate taxa sampled incadmium-spiked and ambient sediments in Locations A and B in LakeMacquarie.169
Table 5.5. Significant differences detected by SNK <i>post-hoc</i> tests in community indices and the abundance of selected macroinvertebrate taxa sampled in the cadmium-spiked and ambient sediments from locations A and B in Lake Macquarie.
Table 5.6. NPMANOVA on Bray-Curtis distances for benthic assemblages among locations and among sediment treatments within locations
Table 5.7. Among treatments results (including ambient fauna) within locationsfrom NPMANOVA pair-wise comparisons tests of macroinvertebrate assemblages
Table 5.8. Among location results (including ambient fauna) within treatments fromNPMANOVA pair-wise comparisons tests on macrobenthic assemblages
Table 5.9. The five most dominant macroinvertebrate taxa and their relativecontribution (%) to the assemblages of the cadmium-spiked and ambient sediments
Table 5.10. Summary of results from SIMPER analyses showing Bray-Curtis (%)dissimilarities among faunal assemblages in the cadmium-spiked and ambientsediments, and the five macroinvertebrate taxa contributing most to thesedissimilarities

xxxii