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Review article

Carbonaceous matter in the atmosphere and glaciers of the Himalayas and the Tibetan plateau: An investigative review

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ABSTRACT

Carbonaceous matter, including organic carbon (OC) and black carbon (BC), is an important climate forcing agent and contributes to glacier retreat in the Himalayas and the Tibetan Plateau (HTP). The HTP – the so-called "Third Pole" – contains the most extensive glacial area outside of the polar regions. Considerable research on carbonaceous matter in the HTP has been conducted, although this research has been challenging due to the complex terrain and strong spatiotemporal heterogeneity of carbonaceous matter in the HTP. A comprehensive investigation of published atmospheric and snow data for HTP carbonaceous matter concentration, deposition and light absorption is presented, including how these factors vary with time and other parameters. Carbonaceous matter concentrations in the atmosphere and glaciers of the HTP are found to be low. Analysis of water-insoluable organic carbon and BC from snowpits reveals that concentrations from inorganic carbon in glacier samples in arid regions of the HTP may be overestimated due to contributions from inorganic carbon in transported from outside the HTP (e.g., South Asia), although local HTP emissions may also be important at some sites. This review provides essential data and a synthesis of current thinking for studies on atmospheric transport modeling and radiative forcing of carbonaceous matter in the HTP.

1. Introduction

Carbonaceous aerosols, including organic carbon (OC) and black carbon (BC), play an important role in the climate system. Carbonaceous particles deposited on snow and ice surfaces are referred to as carbonaceous matter, which can be classified into BC, water-insoluble organic carbon (WIOC) and water-soluble organic carbon (WSOC; the equivalent of dissolved organic carbon (DOC) in aerosol studies) (Li et al., 2016c).

BC warms the atmosphere and is the second most important warming

component after CO₂ in the atmosphere (Bond et al., 2013; Ramanathan and Carmichael, 2008). Light absorption by BC is evaluated by the mass absorption cross section (MAC) of BC (MAC_{BC}) (Bond and Bergstrom, 2006). The measured MAC_{BC} value for uncoated BC particles is 7.5 \pm 1.2 m² g⁻¹ at 550 nm (Bond and Bergstrom, 2006). Generally, the MAC_{BC} increases quickly after BC is emitted from combustion activities (Gustafsson and Ramanathan, 2016).

OC has a net negative radiative forcing and generally causes cooling (IPCC, 2013). However, colored organic matter, also referred to as

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brown carbon (BrC), also absorbs sunlight and warms the atmosphere (e. g., Andreae and Gelencser, 2006; Saleh et al., 2014; Yan et al., 2015). In contrast to BC, light absorption by BrC shows a strong spectral dependence, with stronger absorption at shorter (ultraviolet) wavelengths (Helms et al., 2008). As BrC is defined by its color, not the concentration of a specific compound or class of compounds, BrC concentrations cannot be measured. Instead, the optical properties of BrC or OC are reported. Generally, the light absorption of OC is evaluated as MAC_{OC}, which can be divided into MAC_{WIOC} and MAC_{WSOC}. MAC_{WIOC} is commonly higher than MAC_{WSOC} (Chen and Bond, 2010). For instance, MAC_{WIOC} and MAC_{WSOC} of PM_{2.5} (particulate matter with an aerodynamic diameter below 2.5 μ m) in winter in Beijing were 1.66 \pm 0.48 m² g⁻¹ and 1.22 \pm 0.11 m² g⁻¹, respectively (Cheng et al., 2016).

After deposition on snow and ice, these light-absorbing components can further absorb sunlight and reduce snow and ice albedo, which accelerates the melting of snowpack and glaciers and can result in significant climatic and environmental consequences (Lau et al., 2006; Qian et al., 2011). Constraining concentrations of BC and OC in the atmosphere and snow/ice is challenging due to the complex nature and large spatial variations of carbonaceous matter. For instance, both OC and BC are composed of multiple components and often mixed together, so that no optimal method exists for accurately determining OC and BC

concentrations (Andreae and Gelencser, 2006). Meanwhile, as for some other short-lived species in the atmosphere, carbonaceous particles are often strongly concentrated around emission sources (Bond et al., 2013). Therefore, the abovementioned factors contribute to uncertainties in both estimated concentrations of carbonaceous matter and their transport, deposition, and radiative forcing. Such uncertainties are more pronounced for remote areas with harsh environments, particularly over glaciers, where it is hard to conduct in situ measurements.

The Himalayas and the Tibetan Plateau (HTP) is a remote region that has low levels of local human activity and is physically distant from regions with major emissions. The HTP is also characterized by the most extensive glacier area and snow and ice masses outside the Arctic and Antarctic. Therefore, the HTP is also referred to as the Third Pole (Yao et al., 2012) and is susceptible to environmental change due to regional and global warming. Carbonaceous particles may be an important trigger of glacial retreat in this fragile region (He et al., 2018; Xu et al., 2009), potentially affecting the seasonal supply of freshwater to billions of residents downstream (e.g., Kang et al., 2019; Kang et al., 2020; Lau et al., 2010; Menon et al., 2010; Xu et al., 2016; Zhang et al., 2017b). It is estimated that the surface radiative forcing due to BC-induced snow albedo reduction has been above 5 W m⁻² during spring and summer in many areas of the HTP (He et al., 2018; Kopacz et al., 2011; Qian et al.,



Fig. 1. (a) Schematic diagram of sources of carbonaceous particles and mineral dust in the atmosphere of the Himalayas and the Tibetan Plateau (HTP). (b) Stations and glaciers studied in this review.

2011).

Thus, an increasing number of studies have focused on carbonaceous particles, especially BC in the HTP atmosphere and glaciers. In particular, interest in this subject was boosted after the discovery of Asian Brown Clouds and the associated climate consequences (Ramanathan et al., 2001). Subsequently, numerous studies on carbonaceous particles were conducted in the Indian subcontinent (e.g., Bikkina et al., 2019; Sheesley et al., 2003), the western Himalayas (e.g., Pant et al., 2006; Ram et al., 2010) and the Central Himalayas (e.g., Decesari et al., 2010; Kaspari et al., 2014; Marinoni et al., 2010; Ming et al., 2008).

However, the knowledge of carbonaceous particles and their climatic effects over the HTP is still limited, and monitoring stations are rather sparsely distributed across the large HTP areas despite research advancements (Fig. 1, Fig. S1). Most stations and field campaign measurements are primarily distributed in the southern and eastern regions of the HTP, thus preventing accurate estimates of carbonaceous particles and their radiative/climatic effects over the entire HTP. Meanwhile, large uncertainties exist in previously reported measurement data. For instance, high atmospheric mineral dust concentrations in some regions of the HTP cause potential bias in carbonaceous particle measurements. It has been reported that BC concentrations in total suspended particles (TSP) at Nam Co and Everest stations, two typical remote stations in the middle and southern parts of the HTP, were overestimated by $52\pm35\%$ and 39 \pm 24%, respectively, due to contributions of inorganic carbon (e. g., carbonate) in mineral dust (Li et al., 2017a). Similarly, OC concentrations were overestimated by 22 \pm 10% and 22 \pm 12%, respectively, at these two stations (Li et al., 2017a). However, the contribution of inorganic carbon should be low in the southeast HTP with high forest coverage, as suggested in a previous study (Chow and Watson, 2002). In addition, mineral dust is also an important component of light-absorbing impurities in glaciers of the HTP, especially in aged snow (snow that has experienced thaw-freeze cycles) and bare ice (Li et al., 2017c; Qu et al., 2014). It should be noted that several published studies have ignored the influence of inorganic carbon in mineral dust in the HTP (e.g., Li et al., 2017c; Niu et al., 2018; Qu et al., 2014; Zhang et al., 2017b), thus overestimating OC and BC concentrations in those glaciers. Therefore, it is essential to correct this measurement bias and build an improved dataset for OC and BC concentrations in the atmosphere and glaciers of the HTP to provide better reference data for related studies. OC and BC ratios between acid-treated and untreated aerosol samples can be adapted to remove the influence of inorganic carbon on OC and BC values for snow/ice samples of glaciers.

Furthermore, although South Asia is generally considered the main source of BC in the HTP (e.g., Cong et al., 2015; Lin et al., 2017; Ming et al., 2010; Zhao et al., 2013b), recent studies have indicated that local emissions from the Tibetan Plateau are also significant at certain sites (Hu et al., 2018a; Li et al., 2016a; Li et al., 2018b). Therefore, previously reported in situ data and conclusions need to be critically investigated and constrained. This would allow a better understanding of the status of carbonaceous particles in the HTP and provide better reference values for studies focusing on radiative forcing estimates and modeling (Chung et al., 2012; He et al., 2014). After considering potential influencing factors, this review seeks to provide an improved/corrected database of the concentrations, compositions, deposition, and light-absorbing properties of carbonaceous matter in the HTP, which thus distinguishes this review from previous studies (e.g., Gertler et al., 2016; Ming et al., 2013; Yan et al., 2018; Zhang and Kang, 2017).

In this review, in situ observational data have been utilized to elucidate several fundamental characteristics of carbonaceous particles, including concentrations; relative ratios among WIOC, BC, and WSOC; light absorption characteristics; deposition; and sources of the abovementioned components in the atmosphere and glaciers of the HTP.

Characteristics of carbonaceous matter in glacier regions in the HTP can be investigated by collecting snow samples from snowpits in the accumulation zone of glaciers in spring before the onset of snowmelt (Li et al., 2016a; Ming et al., 2013). By sampling snowpits from the surface

of the snowpack down to the last autumn layer prior to spring melt, the collected samples preserve direct atmospheric deposition of carbonaceous particles in the snowpack, with minimal melt influence during winter and spring. This provides a valuable view of the original distribution of impurities within snow without considering post-depositional processes. Although carbonaceous matter in the ablation zone (e.g., in cryoconites) is also important for snow/ice albedo reduction (e.g., Gook et al., 2016; Li et al., 2017c; Takeuchi, 2002), substantial post-depositional processes make it challenging to study the direct connection between ablation-zone carbonaceous particle concentration and atmospheric deposition. Therefore, this review focuses primarily on snowpit data. This review yields an improved understanding of this topic and can be utilized by future studies, especially those focusing on modeling. Finally, a perspective on promising future research directions is presented.

2. Methodology

In this review, previously published carbonaceous matter data from the atmosphere and on glaciers across the HTP are synthesized and discussed. The sampling and measurement methods for the main particle characteristics/quantities are described in detail in Text S1 (see the supporting information). The reported data were corrected by accounting for factors such as the influence of different measurement methods and the contribution of inorganic carbon in mineral dust to reduce uncertainties (Table 1). First, data from NIOSH 5040 were adjusted to those of IMPROVE-A based on the BCIMPROVE-A TOR/ BCNIOSH $_{5040 \text{ TOT}}$ ratio (1.32 \pm 0.74) (Li et al., 2017a). Second, data from original samples were adjusted to the values for acid-treated samples according to $OC_{acid-treated}/OC_{original sample}$ (0.78 \pm 0.12) and $BC_{acid-treated}/BC_{original}$ $_{sample}$ (0.61 \pm 0.24) ratios for TSP collected at the Everest station (Li et al., 2017a). At present, BC concentrations in the atmosphere of the HTP are typically measured by either the thermal-optical method (Cao et al., 2009; Ming et al., 2010) or aethalometers (Zhang et al., 2017a; Zhao et al., 2017). BC is normally referred to as elemental carbon (EC) when measured with the thermal-optical method (Petzold et al., 2013). BC concentrations reported from aethalometers are called equivalent black carbon (eBC) (Petzold et al., 2013) and are generally comparable to or higher than those obtained from thermal-optical methods based on aerosol samples untreated by acid (Ming et al., 2010; Zhang et al., 2017a). One of the shortcomings of measuring BC with an aethalometer is that it cannot measure OC concentration simultaneously. BC concentrations in glacier snow/ice samples over the HTP are measured mainly by the thermal-optical method (e.g., Li et al., 2016a; Qu et al., 2014; Xu et al., 2009; Zhang et al., 2017b).

3. Distributions, sources and properties of carbonaceous matter in the HTP

3.1. Carbonaceous particle concentrations and compositions

3.1.1. Carbonaceous particles in the surface air

Concentrations of carbonaceous particles in the atmosphere are fundamental to understanding their sources, transport, deposition, and radiative forcing. Lhasa is the second largest city in the HTP (Fig. S2). The annual average OC and BC concentrations in PM_{2.5} samples in the ambient atmosphere of Lhasa have been reported as 3.27 and 2.24 μ g m⁻³, respectively (Li et al., 2016b), which are approximately 0.22 and 0.26 times those in Beijing, respectively (Cheng et al., 2011). Concentrations and compositions of carbonaceous particles measured at remote stations of the HTP can be used as reference data in atmospheric transport models (Ji et al., 2015; Zhang et al., 2015a) and are also crucial for investigating deposition and light absorption properties of carbonaceous matter in glaciers. Previous studies (Cong et al., 2015; Ming et al., 2010; Zhao et al., 2013a; Zhao et al., 2013b) have reported relatively high BC concentrations (e.g., 320 ng m⁻³ at Lulang station) at

Table 1

Original and adjusted concentrations of WIOC, BC, and WSOC in dif	ferent sample types for typical glaciers and monitor stations of the HTP
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Site					Original reported data				Adjusted data			
	Sample type	Method	Reference	Acid treated	WIOC + BC	WIOC	BC	WSOC	WIOC/BC	WIOC	BC	WIOC/BC
Laohugou No.12 glacier	Snowpit	А	1	Y	1.01	0.88	0.13	0.42	6.70	0.84	0.17	4.83
Xiaodongkemadi glacier	Snowpit	А	1	Y	0.52	0.47	0.05	0.25	9.61	0.46	0.06	7.04
Xiaodongkemadi glacier	Fresh snow	В	2	N	ND	0.16	0.04	ND	3.78	0.12	0.03	4.84
Zhadang glacier	Snowpit	Α	1	Y	0.64	0.56	0.08	0.37	7.33	0.54	0.10	5.31
Zhadang glacier	Fresh snow	В	3	N	ND	0.82	0.30	ND	2.71	0.64	0.18	3.47
Yulong glacier	Snowpit	Α	1	Y	0.33	0.30	0.03	0.27	9.87	0.29	0.04	7.24
Demula glacier	Snowpit	Α	1	Y	0.23	0.21	0.02	0.19	11.57	0.21	0.02	8.52
Qiangyong glacier	Snowpit	Α	1	Y	0.326	0.30	0.03	ND	10.80	0.29	0.04	7.95
East Rongbu glacier	Snowpit	Α	1	Y	0.14	0.13	0.01	0.15	12.01	0.13	0.01	9.02
Thorung glacier	Snowpit	Α	1	Y	0.18	0.16	0.02	ND	7.85	0.15	0.03	5.70
Zuoqiupu glacier	Ice core	В	4	Y	0.05	0.04	0.01	ND	4.79	0.04	0.01	4.79
Dongga glacier	Snowpit	В	5	N	0.38	0.28	0.10	ND	2.89	0.22	0.06	3.69
Renlongba glacier	Snowpit	В	5	N	1.14	0.82	0.32	ND	2.57	0.64	0.19	3.29
Demula glacier	Snowpit	В	5	N	0.30	0.18	0.13	ND	1.42	0.14	0.08	1.81
Tibet	Surface soil	В	6	Y	20.99	18.97	2.02	ND	9.39	18.97	2.02	9.39
Nam Co station	PM _{2.5}	В	7	Y	0.29	0.21	0.08	0.25	2.76	0.21	0.08	2.76
Lulang station	TSP	В	8	N	2.92	2.40	0.52	1.88	4.61	1.87	0.32	5.89
Everest station	TSP	В	9	Y	0.91	0.66	0.25	0.77	2.64	0.51	0.15	3.38
Qinghai Lake	PM _{2.5}	В	10	N	2.75	1.91	0.84	1.55	2.27	1.49	0.51	2.91
Muztagh Ata	TSP	В	11	Y	0.32	0.26	0.06	0.22	4.33	0.26	0.06	4.73
Baishui glacier	TSP	В	12	N	2.10	0.57	1.53	0.37	0.37	0.44	0.93	0.48
QSS station	PM _{2.5}	В	13	N	0.24	0.14	0.10	0.53	1.40	0.11	0.06	1.84
Nam Co station	Precipitation	В	14	Y	0.21	0.19	0.02	1.05	10.38	0.19	0.02	10.38
Lulang station	Precipitation	В	14	Y	0.36	0.33	0.03	0.83	9.53	0.33	0.03	9.53
Everest station	Precipitation	В	14	Y	0.26	0.25	0.01	0.86	18.37	0.25	0.01	18.37

Note: OC, BC and WSOC are organic carbon, black carbon, and water-soluble organic carbon, respectively. TOR and TOT are thermal-optical reflectance and thermal-optical transmittance, respectively. These estimates were derived by the IMPROVE-A TOR method. The units of water-insoluble organic carbon (WIOC) and BC in snow/precipitation samples are mg L^{-1} ; units of WIOC and BC in surface soil and aerosol samples are mg g^{-1} and μ g m⁻³, respectively. The OC/WSOC ratio for Lulang station was adapted from (Li et al., 2016d). The OC/WSOC ratio for the Muztagh Ata station was adapted from that of Qinghai Lake (Zhao et al., 2015). A and B represent NIOSH 5040 TOT and IMPROVE-A TOR, respectively. References indicated in this table are 1, (Li et al., 2016a; Li et al., 2016c); 2, (Li et al., 2017c); 3, (Li et al., 2018c); 4, (Xu et al., 2009); 5, (Zhang et al., 2017b); 6, (Gautam et al., 2019); 7, unpublished article; 8, (Zhao et al., 2013b); 9, (Cong et al., 2015); 10, (Zhao et al., 2015); 11, (Cao et al., 2009); 12, (Niu et al., 2018); 13, (Xu et al., 2015); and 14, (Yan et al., 2019). "Y" and "N" represent "yes" and "no", respectively; ND: no data.

the fringe of the HTP and low concentrations (e.g., 76 ng m⁻³ at Nam Co station) in the inner HTP because the former area is highly influenced by mineral dust and locally sourced anthropogenic emissions from nearby residential activities (Li et al., 2018b) (Fig. S1). Therefore, in contrast to Lulang and Everest, which experience high local emissions, the Nam Co station is characterized by low airborne aerosol concentrations (Yan et al., 2019). In particular, the Nam Co station is located in an open area with high vegetation coverage and few local residents; thus, local emissions are very low and easily diffuse into the atmosphere (Fig. S1C, Text. S3). Furthermore, the lowest aerosol optical depth value (0.05 at 500 nm) in the HTP has been reported at Nam Co station (Cong et al., 2009) and is comparable to that in Antarctica (Six et al., 2005), thus indicating a highly clean atmosphere. Therefore, Nam Co station is an ideal site for studying the background atmospheric environment of the HTP.

3.1.2. Carbonaceous particles in glacier snow/ice

Several factors need to be noted during studies on carbonaceous matter in glacier snow/ice over the HTP. First, data from only snow pit and fresh snow samples in glacier regions retain the information of deposited particles. Second, because most glaciers in the HTP are valley glaciers, surface soil in the glacier regions is often transported by strong winds and deposited on the glacier surfaces (Fig. S3). Because surface soil particles may also contain OC and BC (Gautam et al., 2020; Li et al., 2019a), the contribution of carbonaceous particles directly deposited on glaciers from long-range atmospheric transport can be overestimated, especially in the ablation area where deposition of soil from nearby sources can be high. However, both long-range atmospheric deposition and locally transported fine silt particles from surface soil in the glacier regions reduce the albedo of the glaciers (Li et al., 2019b) (Fig. S3). Third, it is generally assumed that high altitude glaciers are located far

from local sources, and that the dominant source of carbonaceous particles deposited on these glaciers are from long-range transport. However, this is not always the case. For instance, Mt. Nyainqêntanglha is located close to Lhasa, and one of its glaciers (Zhadang) has been shown to be influenced by Lhasa emissions (Fig. 1b) (Ming et al., 2013).

WSOC concentrations in snowpit samples of the HTP vary spatially, being higher (0.42 \pm 0.29 mg L⁻¹) toward the north and lower toward the south ($0.15 \pm 0.06 \text{ mg L}^{-1}$), a trend also seen for BC (Li et al., 2016c). As particulate WIOC and BC are hydrophobic, they tend to be retained in the surface snow and ice during melting (Xu et al., 2012), so that WIOC and BC in aged snow and ice samples become enriched compared to their levels in fresh snow and snowpit samples. For instance, BC concentrations in fresh snow and aged ice samples obtained from the same glacier (i.e., Zhadang, middle HTP) were 41.77 \pm 6.36 µg L⁻¹ and $634.27 \pm 104.42 \ \mu g \ L^{-1}$, respectively (Li et al., 2017c). It has also been found that BC is more likely to be removed by meltwater than mineral dust during the melting process, so the enrichment strength of mineral dust was higher than that of BC. Therefore, dust contributes more to glacier melt than BC after strong melt events (Hu et al., 2020). Studies of these processes for HTP glaciers are limited; thus, further studies covering more glaciers are required.

The WIOC/BC ratio differs during precipitation and snow melting processes. In general, the scavenging ratio of WIOC is higher than that of BC in the atmosphere, which has been found not only in the HTP (Yan et al., 2019), but also in other regions such as Europe (Cerqueira et al., 2010) and Japan (Huo et al., 2016). Therefore, for a given region, the WIOC/BC ratio in snow samples should be higher than the WIOC/BC ratio in atmospheric aerosols (Fig. S4a). In this study, the average WIOC/BC ratio (3.59 ± 1.47) in atmospheric aerosols obtained from 6 remote stations in the HTP (Text. S2) were utilized to compare with those of snowpit samples. The WIOC/BC ratios of snowpit samples

reported in different studies in the HTP exhibit a large range, spanning 1.81 (Demula) (Zhang et al., 2017) to 9.02 (East Rongbu) (Li et al., 2016a; Li et al., 2016b) (Table 1). WIOC/BC ratios less than 3.59 contradict the above assertion that the scavenging ratio of WIOC is higher than that of BC. However, it can be explained by melt processes in snowpits, which result in enrichment of BC compared to WIOC and thus a low WIOC/BC ratio (Xu et al., 2012) (Fig. S4b). For instance, the WIOC/BC ratio of Demula snow samples collected in June 2015 was 1.81 (Zhang et al., 2017), lower than that in another study conducted at the same glacier at May 2014 of 8.52 (Li et al., 2016a) (Fig. S4, Table 1). The reason is that higher temperatures at the collecting time in the first study (Zhang et al., 2017) caused strong melting of snow and higher enrichment of BC than that of WIOC. Meanwhile, because the former study did not remove the influence of inorganic carbon (Zhang et al., 2017), the low reported WIOC/BC ratio was also likely caused by a higher contribution of inorganic carbon to BC than WIOC (Li et al., 2017a). Therefore, not all the data reported from HTP snowpits reflect direct atmospheric deposition, and need to be re-evaluated.

3.1.3. Temporal variations in carbonaceous particles

Seasonal variations in carbonaceous particle concentrations have been observed in different areas of the HTP. Specifically, atmospheric eBC concentrations in the southern HTP indicate seasonal variations that are primarily affected by the precipitation amount. For instance, low and high atmospheric eBC concentrations at the Everest station appear during summer and spring with heavy and light precipitation, respectively (Chen et al., 2018) (Fig. 2). In contrast, high atmospheric eBC concentrations occur during summer in the northern HTP (e.g., Qilian Shan Station (QSS station)) due to the dominant westerlies (Zhao et al.,



Fig. 2. Seasonal variation in atmospheric eBC concentrations and precipitation at two remote stations (Everest and Qilian Shan Station (QSS)) under the influence of the Indian monsoon and the westerlies, respectively. Data for Everest and QSS stations have been adapted from Zhao et al. (2012) and Chen et al. (2018), respectively. The data presented here do not consider the contribution of mineral dust because BC concentrations were measured by an aethalometer.

2012).

Diurnal variations in atmospheric eBC concentrations likely reflect variations in eBC sources and transport. To date, distinct diurnal variation characteristics in eBC have been reported at different remote stations, indicating a complex influence of factors such as micro-terrains, local sources, and advection. For instance, eBC concentrations at the southern HTP presented a peak during morning hours (6–8 am) (Chen et al., 2018; Wang et al., 2016; Zhang et al., 2017a; Zhao et al., 2017), suggesting contributions from local activities such as cooking and increased traffic. Meanwhile, the diurnal eBC variation at the Nam Co station is consistent with diurnal wind speed variation (Yin et al., 2017; Zhang et al., 2017a), indicating the influence of atmospheric advection. However, diurnal variations in eBC concentrations were not obvious at the western Muztagh Ata station (Zhu et al., 2016).

3.2. Deposition of carbonaceous particles

The deposition of carbonaceous particles is closely linked to their lifetime and transport. At present, most studies concerning the deposition of carbonaceous particles in the HTP are focused on BC (e.g., Bauer et al., 2013; Han et al., 2015; Ji et al., 2015; Yasunari et al., 2013; Zhang et al., 2015a), while related studies on OC are limited (Li et al., 2017b; Yan et al., 2019; Yan et al., 2020b).

Annual BC deposition rates based on snowpit samples from eight glaciers in the HTP were 17.9 \pm 5.3 mg m⁻² a⁻¹ (Li et al., 2017a) (Fig. 3), which was far lower than that in highly polluted East China $(1160 \text{ mg m}^{-2} \text{ a}^{-1})$ (Han et al., 2016). This is because atmospheric BC concentrations in the high glaciers mainly reflect the background BC level in the middle and upper troposphere (under certain meteorological conditions). Additionally, other local factors have been shown to exert little influence. Compared to BC deposition in high-altitude glacier regions (Li et al., 2017a), corresponding data at the stations located at low altitude indicate large variability due to complex contributions from locally sourced emissions and mineral dust in addition to deposition from long - range sources (Yan et al., 2019). A detailed study on carbonaceous matter deposition directly from the atmosphere at three remote stations (Nam Co, Lulang, and Everest) of the HTP has shown that the total BC deposition at Nam Co, Lulang, and Everest stations were 15.3, 58.9, and 9.7 mg $m^{-2} a^{-1}$, respectively. These results were consistent with the spatial patterns of wet deposition of WSOC at these stations (Li et al., 2017b).

An interesting phenomenon in the HTP is the high contribution of



Fig. 3. Atmospheric BC deposition rates in the HTP derived from glaciers, models, BC concentrations in aerosol samples (Li et al., 2017a), and direct dry/ wet deposition measurements (Yan et al., 2019). Note: Deposition in the glacier region (17.9 \pm 5.3 mg m⁻² a⁻¹) is considered representative of the HTP. Lower or higher values could be attributed to local factors such as low precipitation amounts (Everest station) and local emissions (Lulang station).

dry deposition. Dry deposition accounted for approximately 62%, 41%, and 73% of the total BC deposition at Nam Co, Lulang, and Everest stations, respectively. These values are high considering that wet deposition generally accounts for a larger fraction of the total BC deposition (Jurado et al., 2008). It is likely that relatively high atmospheric mineral dust concentrations in the HTP and limited precipitation contributed to the large dry deposition of BC. For instance, the ratio of mineral dust concentration to carbonaceous particle concentration in TSP sampled from the Nam Co station is approximately 9.0 (Li et al., 2017a). Generally, the dry deposition velocity increased rapidly with increasing particle size (Zhang et al., 2001). Mean mineral dust particles of the HTP are larger than 11 µm (Dong et al., 2016), larger than those of carbonaceous particles of less than 2.5 µm, thus resulting in higher dry deposition rates. The OC and BC concentrations in fine particles from surface soil (<20 μm) in the HTP are 18.97 \pm 16.03 and 2.02 \pm 1.55 mg g^{-1} , respectively (Gautam et al., 2019), suggesting that the dry deposition of OC and BC in fine particles from surface soil significantly contributes to the dry deposition of carbonaceous matter. In addition, in situ measured ratios of BC dry and wet deposition are higher than those simulated by typical climate models (Yan et al., 2019). Therefore, current models need to incorporate the contribution of dry deposition of fine particles from surface soil containing carbonaceous matter, especially in the western and northern regions of the HTP, which are characterized by dry environments and heavy dust storms (Fig. 1).

Furthermore, the total deposition of WIOC and BC at the Everest and Lulang stations are significantly higher than deposition on nearby glaciers. This could be attributed to locally sourced coarse particles (primarily comprising OC) (Yan et al., 2019), which are difficult to transport to high-altitude glaciers. Sparser vegetation coverage around the stations is associated with higher measured atmospheric particle concentrations (Liu et al., 2017). Therefore, the total in situ WIOC and BC deposition at these stations are probably overestimated due to the contribution of the local fine particles from the surface soil of the HTP itself.

Seasonally, the wet deposition rates of WIOC and BC are influenced by their atmospheric concentrations and precipitation, with

precipitation exerting a more significant effect (Fig. 4). The wet deposition ratios of WIOC, WSOC and BC at the three sites varied with precipitation amount, with high and low values in the monsoon and nonmonsoon periods, respectively. Generally, concentrations of carbonaceous matter in precipitation indicate a significant negative correlation with precipitation amounts, especially in polluted areas with high aerosol concentrations, which depend on whether they are dominated by the scavenging or dilution effect (Rastogi and Sarin, 2005; Yan and Kim, 2012). Here, BC and WIOC concentrations in precipitation indicated a nonsignificant correlation with precipitation amounts at the three studied remote stations. It has been suggested that washout and rainout processes are both incorporated in precipitation events (Ishikawa et al., 1995). At present, although it is difficult to determine the dominant mechanism in the scavenging process at the three study stations of the HTP, the low particle concentrations in the atmosphere of Nam Co, Lulang and Everest stations should be the main factor causing the low correlation. The BC scavenging ratio is defined as the ratio of the BC concentration in the precipitation to the BC concentration in the atmosphere (Ducret and Cachier, 1992). The BC scavenging ratio at Nam Co station was 0.24×10^6 (Yan et al., 2019), consistent with those in remote regions (e.g., Schauinsland, Germany) (Cerqueira et al., 2010), thus suggesting that this station could potentially represent remote sites in the inner HTP. BC scavenging ratios at Lulang and Everest stations were lower than those at Nam Co station because of the contribution of local emissions with hydrophobic particles (e.g., those from local residential burning, vehicles, and mineral dust) (Yan et al., 2019).

3.3. Sources of carbonaceous matter

3.3.1. Biomass and fossil fuel combustion

Combustion emissions and condensed semi-volatile organic compounds are the main sources of carbonaceous particles (Pio et al., 2007; Qi and Wang, 2019; Szidat et al., 2004; Winiger et al., 2019). BC is mainly generated from biomass and fossil fuel combustion (e.g., Gustafsson et al., 2009; Szidat et al., 2004). However, sources of OC are complex and include combustion activities, fine particles from surface



Fig. 4. Seasonal wet deposition rates of WIOC, BC (Yan et al., 2019), and WSOC (Li et al., 2017b) at Nam Co, Lulang, and Everest stations.

soil, pollen, and secondary organic processes. Therefore, it is more straightforward to investigate the sources of BC than those of OC. Furthermore, the contribution of fossil and non-fossil fuel sources to WSOC in precipitation can be distinguished by the radiocarbon method (especially if the contribution of fine particles from surface soil is low) (Raymond, 2005; Stubbins et al., 2012). In general, the contribution of fossil fuel burning to BC and WSOC in precipitation decreases from the outer to the inner regions of the HTP (Li et al., 2016a; Li et al., 2018a) (Fig. 5).

Some valleys that cut across the Himalayas have been proposed as important conduits for pollution transport from highly polluted South Asia to the Tibetan Plateau (Dhungel et al., 2018; Zhang et al., 2020). It has been reported that BC atmospheric concentrations and the associated contribution from fossil fuel combustion decrease from South Asia to the inner HTP (Fig. 5) (Li et al., 2016a) for two reasons. First, pollutants from South Asia are diluted and deposited across the valleys



Fig. 5. Contributions of fossil fuel burning to BC (in aerosol, snowpit, and precipitation samples) and WSOC (in snowpit and precipitation samples) in the HTP. (a) Aerosol samples were collected from two valleys across the Himalayas (black dots), and precipitation samples (for WSOC) were obtained from Nam Co and Kathmandu (blue squares). (b) Snowpit samples were obtained from the HTP. Data for BC in aerosol and snowpit samples were adapted from (Li et al., 2016a). Data for WSOC in precipitation and snowpit samples were adapted from (Li et al., 2016a). Purple and green belts represent fossil contributions of East Asia and South Asia, respectively. "-M" represents monsoon period; TP represents the Tibetan Plateau. (For interpretation of the references to color in this figure legend, the reader is referred to the web version of this article.)

during transport (Fig. S5). Second, local non-fossil fuel sources of BC (e. g., biomass combustion) significantly contribute to atmospheric pollution in the Himalayan region. Although particles from biomass combustion can be transported over long distances (Forster et al., 2001), their strong hydrophilicity makes the long-distance transport efficiency lower than that of particles from fossil fuel combustion (Pio et al., 2007).

Similarly, the contribution of BC from fossil fuel combustion in snowpits decreases from the outer to inner regions of the HTP (Fig. 5b). Glaciers in the inner HTP are primarily influenced by local biomass combustion (Hu et al., 2018a) because local residents of the Tibetan Plateau burn yak dung and biomass for cooking and heating (Chen et al., 2015; Xiao et al., 2015). In contrast, the fraction of BC driven by fossil fuel combustion in urban cities of the HTP (i.e., Lhasa) is high (Li et al., 2016a) and similar to those in megacities of East China such as Beijing (Chen et al., 2013).

Generally, OC of a given sample (e.g., aerosol or snowpit sample) is younger and more enriched in \triangle^{14} C than that of BC, which has been shown in previous studies conducted in a number of regions in the world such as Europe (Szidat et al., 2004), South Asia (Gustafsson et al., 2009) and East China (Zhang et al., 2015b). The fossil fuel contribution to WSOC in precipitation at the Nam Co station has been reported as $15 \pm$ 6%, which is lower than that observed for typical urban cities in South Asia (Kathmandu, $33 \pm$ 9%) and East Asia (Kunming, $36 \pm$ 10%) (Li et al., 2018a) (Fig. 5).

Similar results have also been observed for snowpit samples, suggesting the presence of OC from biomass/biogenic sources in precipitation samples. Previous research has concluded that precipitation events are driven by Indian airmasses at Nam Co station that generally show higher fossil fuel contributions than by airmasses from the Westerlies (Li et al., 2018a), suggesting long-distance transport of pollutants from South Asia. However, Lhasa is approximately 100 km to the south of Nam Co station. Therefore, distinguishing fossil fuel-sourced emissions from Lhasa and long-range transported pollutants from South Asia is challenging. The δ^{13} C of WSOC is expressed as the measured ratio of stable isotopes ${}^{13}C/{}^{12}C$ relative to that of the Pee Dee Belemnite (PDB) standard by the equation $\delta^{13}C = [({}^{13}C/{}^{12}C)_{sample}/({}^{13}C/{}^{12}C)_{standard}-1] \times$ 10³ (Narukawa et al., 1999). Both aging processes and photochemical reactions (Mladenov et al., 2011) can alter the original source δ^{13} C signature of WSOC. δ^{13} C at Nam Co in precipitation samples after the potential influence of inorganic carbon in mineral dust was removed by acid has been reported as $-21 \pm 3\%$ (Li et al., 2018a). This value was similar to that of snowpit samples ($-21 \pm 3\%$) but enriched compared to those of OC in surface soil samples from the HTP (-25.4 ± 0.6 %) and precipitation WSOC in urban cities (i.e., Lhasa: $-25 \pm 2\%$) (Fig. S6), thus indicating the occurrence of nontrivial OC aging processes (Narukawa et al., 1999) and photochemical reactions (Miller and Zepp, 1995; Mladenov et al., 2011) during transport to the Nam Co station and the glaciers. Therefore, precipitation WSOC at the Nam Co station likely experienced long-distance transport.

3.3.2. Sources supplying carbonaceous particles to subregions of the HTP

It is complex and challenging to determine the regional sources of carbonaceous particles in the HTP due to: (1) the large area of the region, (2) different sub-regions of the HTP have different sources of carbonaceous particles, and (3) the network of observations and monitoring stations are sparsely distributed. Commonly used methods such as backward air mass trajectory analysis (Li et al., 2016d; Zhao et al., 2013b) and the use of fire points (Chen et al., 2018) or emission inventories (Zhang et al., 2017b) can be used to constrain subregion carbonaceous particle sources, but these methods only indirectly imply the potential sources of BC within the HTP.

Based on the δ^{18} O in the precipitation, the HTP can be divided into three subareas (i.e., northern and western region, middle region and southern region) (Yao et al., 2013), similar to the distributions of \triangle^{14} C from BC in snowpit samples. Meanwhile, an increasing body of evidence suggests significant contributions of local sources to BC concentrations at specific sites (Li et al., 2018b; Yan et al., 2019). Therefore, although long-distance transported BC significantly affects most regions of the HTP, the contribution of locally sourced BC is significant not only in cities and towns but also at some remote stations in narrow valleys or close to busy roads and emission sources. OC sources can be indirectly determined from BC sources. Because OC adheres more easily than BC on larger particles such as mineral dust, OC has less potential for longdistance transport than BC. However, fine OC still plays an important role in the aerosol compositions at remote HTP sites. For instance, the annual fine OC accounted for approximately 61% of the mass concentration of TSP at Nam Co station during 2012 (Wan et al., 2015).

3.4. Light absorption characteristics of carbonaceous matter

The light absorption ability of carbonaceous matter is assessed by MAC values, which provide fundamental input data for studies on the radiative forcing of aerosols (Petzold et al., 2013). The referenced MAC_{BC} value for freshly emitted particles is approximately $7.5 \pm 1.2 \text{ m}^2$ g⁻¹ at 550 nm (Bond and Bergstrom, 2006). Although the light absorption by OC is limited at this wavelength, it increases significantly at shorter wavelengths (Schnaiter et al., 2006). In general, MAC_{OC} is derived at 365 nm (Cheng et al., 2017; Kirillova et al., 2014). Both MAC_{BC} and MAC_{OC} values vary with particle source, particle size distribution, structures/morphology and atmospheric processes that particles experience (e.g., coating and bleaching) (Chen et al., 2017; Martinsson et al., 2015). For example, MAC_{BC} at 550 nm usually peaks around a particle diameter of 150 nm and starts to decrease as the particle diameter increases from 150 nm to larger values (Bond et al., 2006). Moreover, the particle morphology for both fresh and coated BC also significantly affects MAC_{BC}, with up to a factor of 2 variation in BC absorption due to different coating morphologies and generally larger MAC values for fully coated BC particles than for partially coated particles (Bond et al., 2013; He et al., 2015; He et al., 2016). Therefore, MAC_{BC} and MAC_{OC} indicate large spatiotemporal variations (Forrister et al., 2015; Gustafsson and Ramanathan, 2016). As a result, in situ MAC values of carbonaceous matter in the HTP are important in accurately evaluating OC and BC radiative forcing.

The MAC_{BC} can be calculated from measured data (Text S1.5) (Petzold et al., 2013). Although the MAC_{BC} based on the thermal-optical method can be influenced by many factors (e.g., decreases in incident light with increased filter loading and aerosol scattering effects) (Andreae and Gelencser, 2006; Bond and Bergstrom, 2006; Kondo et al., 2009), this method provides MAC values that are used in radiative forcing calculations (Cheng et al., 2011; Ram and Sarin, 2009).

The average MAC_{BC} values at 632 nm for Lulang and Everest stations were 6.85 ± 1.39 and 6.49 ± 2.81 m² g⁻¹, respectively. These values are comparable to values for freshly emitted BC ($6.5 \pm 1.0 \text{ m}^2 \text{ g}^{-1}$) (Table 2) (Bond and Bergstrom, 2006). These results indicate that the BC at these two stations likely originated from primary local sources with weak aging/coating influences. Previous studies on the HTP have indicated that MAC_{BC} values of fossil fuel-sourced BC at 632 nm are approximately 25% higher than those of biomass combustion-sourced BC (Hu et al., 2017). Accordingly, MAC_{BC} values at Lulang and Everest stations were lower than those reported for Lhasa and Beijing of 7.19 \pm 1.19 $m^2~g^{-1}$ and 9.41 \pm 1.92 \mbox{m}^2 g $^{-1}$, respectively (Cheng et al., 2011; Li et al., 2016b). Notably, the MAC_{BC} value for the Nam Co station has been reported as $12.2 \pm 3.4 \text{ m}^2 \text{ g}^{-1}$ (Table. 2), which is significantly higher than those of freshly emitted particles (Cheng et al., 2017). It has been suggested that some models underestimate the radiative forcing of BC primarily due to inappropriate treatments of the mixing of BC with other constituents (Bond et al., 2013). Therefore, MAC_{BC} values for Nam Co aerosols can be adapted in future models.

The MAC_{WSOC} values at 365 nm for TSP samples collected from Nam Co, Lulang, and Everest stations were 1.74 \pm 0.94 (1.22 \pm 0.44 m² g⁻¹ for PM_{2.5} samples), 0.84 \pm 0.40, and 1.18 \pm 0.64 m² g⁻¹, respectively (Li et al., 2016d). These values were comparable to those in urban cities of

Table 2

 MAC_{WSOC} , MAC_{BC} , and AAE_{WSOC} values from aerosol particles, surface soils in the HTP and previous studies.

Site/ sample	Particle	MAC _{WSOC}	AAE ₃₃₀₋ 400	MAC _{BC} Reference			
type Nam Co	PM _{2.5}	$\begin{array}{c} 1.22 \pm \\ 0.43 \end{array}$	$\begin{array}{c} \textbf{4.13} \pm \\ \textbf{1.38} \end{array}$	12.2 ± 3.36	А		
Nam Co	PM _{2.5}	$\begin{array}{c} 0.38 \pm \\ 0.16 \end{array}$	$\begin{array}{c} \textbf{6.19} \pm \\ \textbf{1.70} \end{array}$	ND	(Zhang et al., 2017c)		
Surface soil		1.67 ± 0.38	$\begin{array}{c} \textbf{4.53} \pm \\ \textbf{0.88} \end{array}$	ND	(Li et al., 2017b)		
Everest	TSP	$1.18~\pm$ 0.64	4.64 ± 1.15	$\begin{array}{c} 6.49 \\ \pm \ 2.81 \end{array}$	(Li et al., 2016b)		
Lulang	TSP	0.84 ± 0.40	5.39 ±	6.85 + 1.39	(Li et al., 2016b)		
Lulang		ND	ND	7.6	(Wang et al., 2018)		
Ngari	TSP	ND	ND	$\begin{array}{c} 14.3 \\ \pm \ 6.54 \end{array}$	(Chen et al., 2019)		
Lhasa	PM _{2.5}	$\begin{array}{c} \textbf{0.74} \pm \\ \textbf{0.22} \end{array}$	5.38	$\begin{array}{c} 7.19 \\ \pm \ 1.19 \end{array}$	(Li et al., 2016d)		
Los Angeles	PM _{2.5}	0.71	ND	ND	(Zhang et al., 2013)		
Beijing, summer	PM _{2.5}	$\begin{array}{c}\textbf{0.44} \pm \\ \textbf{0.06}\end{array}$	$\begin{array}{c} \textbf{7.18} \pm \\ \textbf{0.31} \end{array}$	9.41 ± 1.92	(Cheng et al., 2017)		
Beijing, winter	PM _{2.5}	1.54 ± 0.16	ND	ND	(Yan et al., 2015)		
Seoul	PM _{2.5}	0.28-1.18	$\begin{array}{c} \textbf{7.23} \pm \\ \textbf{1.58} \end{array}$	ND	(Kim et al., 2016)		
NCO-P	PM_{10}	$\begin{array}{c} \textbf{0.57} \pm \\ \textbf{0.18} \end{array}$	$\textbf{4.9} \pm \textbf{0.7}$	ND	(Kirillova et al., 2016)		
New Delhi (IGP)	PM_{10}	1.6 ± 0.5	5.1 ± 2.0	ND	(Kirillova et al., 2014)		
Patiala (IGP)	PM _{2.5} (day)	1.3 ± 0.7	5.1 ± 1.9	$\begin{array}{c} \textbf{4.2} \pm \\ \textbf{1.4} \end{array}$	(Srinivas et al., 2016)		
Bay of Bengal Source samples	TSP	0.4 ± 0.1	9.1 ± 3.6	ND	(Srinivas and Sarin, 2013)		
Wood	PM _{2.5}	0.13–1.1	8.6–17.8	ND	(Chen and Bond, 2010; Hu et al., 2017)		
Vehicle emission	PM _{2.5}	$\begin{array}{c}\textbf{0.71} \pm \\ \textbf{0.17} \end{array}$	$\begin{array}{c} \textbf{4.06} \pm \\ \textbf{1.11} \end{array}$	$\begin{array}{c} 6.06 \\ \pm \ 1.08 \end{array}$	(Chen and Bond, 2010; Hu et al., 2017)		

Note: A, unpublished article; AAE: absorption Ångström exponent; NCO-P: Nepal Climate Observatory-Pyramid; IGP: Indo-Gangetic Plain. ND: no data.

South Asia (Kirillova et al., 2014; Srinivas et al., 2016) but higher than those in the remote Indian Ocean (Bosch et al., 2014; Srinivas and Sarin, 2013) (Table 2), indicating a decreasing trend for aerosol MAC_{WSOC} values from continent to ocean. The higher values at the Nam Co and Everest stations than at the Lulang station indicate a significant contribution of WSOC in fine particles from surface soil (which indicates a higher MAC value there than in particles from fossil and biomass combustion) (Hu et al., 2018b) (Table 2). Furthermore, the MAC_{WSOC} values at Nam Co, Lulang, and Everest stations indicated similar seasonal variation, with high and low values occurring during non-monsoon and monsoon periods, respectively. This could be attributed to the reduction in fine particles from surface soil concentrations in the atmosphere during the monsoon period (Fig. 6). These results suggest that locally sourced fine particles from surface soil significantly contribute to spatiotemporal variations in the MAC_{WSOC} values over the HTP. Therefore, it is assumed that fine particles from surface soil exert a more significant effect on MAC_{WSOC} values in the northern and western HTP, where frequent dust storms are experienced. Meanwhile, other factors, such as increased fossil fuel combustion (Du et al., 2014) and photobleaching of light absorption components of WSOC during this period (Li et al., 2016d), can also result in reduced MAC_{WSOC}.

It has been estimated that the amount of solar energy absorbed by WSOC is $16 \pm 10\%$, $6 \pm 4\%$, and $11 \pm 7\%$ of the amount of solar energy



Fig. 6. Seasonal variation in MAC_{WSOC} at 365 nm for Nam Co, Lulang, and Everest stations (Li et al., 2016d). Note: Data for Nam Co station were obtained from $PM_{2.5}$ samples collected from December 2016 to December 2017 (unpublished data).

absorbed by BC at the Nam Co, Lulang, and Everest stations, respectively. Seasonally, absorption by WSOC and the WSOC/EC absorption ratio are highest during winter and lowest during summer (Li et al., 2016d). These results highlight the contribution of WSOC to total light absorption by carbonaceous particles of the HTP. At present, because few data on the light absorption of WIOC are available for the HTP and the existing data have large uncertainties (Yan et al., 2020a), MAC_{WIOC} is not included in the calculation of radiative forcing. Therefore, the actual relative importance of OC in the total radiative forcing of carbonaceous particles should be higher than that of the aforementioned values.

The average MAC_{WSOC} values for precipitation samples from the Nam Co, Lulang, Everest, and Yulong stations have been reported as $0.48\pm0.47~m^2~g^{-1}, 0.25\pm0.15~m^2~g^{-1}, 0.64\pm0.49~m^2~g^{-1}$, and $0.43\pm$ 0.32 m² g⁻¹, respectively (Li et al., 2017b; Niu et al., 2019). These values are significantly lower than those reported for aerosols (p < 0.01) (Fig. 7), which could be attributed to high concentrations of secondary organic aerosols and biogenic VOCs with low light absorption in precipitation (Hallquist et al., 2009; Lambe et al., 2013). Similar results have been observed in severely polluted regions (e.g., Lhasa and East China) (Cheng et al., 2011; Li et al., 2017b; Yan et al., 2015), thus indicating that this general phenomenon occurs in both remote sites and urban cities. Due to the remoteness and lack of infrastructure in the HTP, the collection of aerosol samples in the high-altitude glacier regions there is rare. However, MAC_{WSOC} for samples collected from snowpits can provide surrogate information. For instance, the MAC_{WSOC} in fresh snow at a glacier of the northern HTP was $0.38 \pm 0.06 \text{ m}^2 \text{ g}^{-1}$ (Hu et al., 2018b), which was close to the values for precipitation samples at remote stations. Thus, it is proposed that the MAC_{WSOC} values of aerosols in the atmosphere over this glacier should be comparable to those of remote stations.

The concentration and light absorption ability of WSOC can vary substantially. For instance, the concentrations and light-absorbing



Fig. 7. MAC_{WSOC} values for aerosol and precipitation samples from the Nam Co, Lulang, and Everest stations and Lhasa. Data for the Nam Co, Lulang, and Everest stations were adapted from (Li et al., 2017b); data on aerosol (PM_{10}) and precipitation samples from Lhasa were adapted from (Li et al., 2016b) and (Li et al., 2017b), respectively.

properties of WSOC in snow and ice samples vary greatly during melting. This phenomenon is also referred to as "fractionation of WSOC" (Hu et al., 2018b). Fractions of hydrophobic WSOC characterized by high light absorption attach to particles (e.g., mineral dust) during melting (Meyer et al., 2008). This mechanism also explains the lower MAC_{WSOC} values in precipitation than in aerosols. A study on Laohugou glacier No. 12 concluded that WSOC concentrations and MAC_{WSOC} in surface ice indicated decreasing and increasing trends, respectively, as the melting strength increased from the upper to the lower part of the glaciers (Hu et al., 2018b). These results indicate the occurrence of the fractionation process, which was further corroborated by a controlledstep freezing experiment with WSOC in surface ice and topsoil samples (Hu et al., 2018b). The results of this experiment showed that the WSOC concentrations of treated samples increased rapidly from the earliest frozen samples to the last frozen samples. In contrast, MAC_{WSOC} values showed the opposite trend.

4. Conclusions and perspectives

This study generates an improved dataset on carbonaceous matter in the atmosphere and glaciers of the HTP, which will provide valuable reference values for other studies, such as radiative forcing estimates and modeling. Although most of the HTP is influenced by carbonaceous particles transported long distance, locally sourced carbonaceous matter from fossil fuel and biomass combustion substantially contribute to pollutants in urban cities and some remote stations, respectively. Seasonal variations of carbonaceous matter deposition in the HTP are mainly controlled by precipitation amount, with dry deposition accounting for a larger portion than estimated by models (Yan et al., 2019). Because the HTP is surrounded by high mountains at its southern and western fringes, fine particles of surface soil significantly affect the concentration, composition, and light absorption properties of carbonaceous matter of the HTP, especially in the northern and western parts where is characterized by a dry environment.

While considerable advances have been made in characterizing HTP carbonaceous particles, further research is needed. For instance, the characteristics of locally sourced carbonaceous particles should be comprehensively studied, and it is necessary to establish high-resolution emission inventories for carbonaceous particles generated from the inner HTP. Additionally, the HTP stations and glaciers that have been sampled for carbonaceous particles are primarily distributed in the

eastern and southern regions of the HTP, and more studies need to be conducted in the western and northern regions of the HTP. Furthermore, quantities such as size distributions, morphologies, and vertical distributions of carbonaceous matter need further investigation. To produce high quality data that is representative of larger regions of the HTP, more stations with limited influence from local residents and microtopography (e.g., those located in wide areas and distant from local anthropogenic activities) should also be set up. At least three potential factors should be considered during the establishment of an "ideal" remote station: it should be in open terrain, at a suitable distance from local residents, and representative of the land type of the region. Future research that addresses these areas of research will further understanding of the sources and distribution of carbonaceous particles on the HTP, and allow a better assessment of the role of carbonaceous particles in climate change and albedo reductions in this sensitive region.

Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Appendix A. Supplementary material

Supplementary data to this article can be found online at https://doi.org/10.1016/j.envint.2020.106281.

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