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Abstract

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Keywords

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Humans, Megafauna and environmental change in tropical Australia

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Abstract

Debate concerning the environmental impact of human arrival in Australia has continued for more than a century. Here we review the evidence for human impact and the mechanisms by which humans may have affected the environment of tropical Australia. We limit our review to tropical Australia because, over three decades ago, it was proposed that the imposition of an anthropogenic fire regime upon human occupation of the Australian continent may have resulted in profound changes in regional vegetation and climate across this region. We conclude that ecological processes and vegetation-fire-climate-human feedbacks do exist that could have driven a significant shift in boundary conditions and ecosystem state at the subcontinental scale through the sustained imposition of an anthropogenic fire regime over tens of millennia. These potential feedbacks operate through the inhibition of forest expansion both directly, by targeted burning at established forest edges and newly irrupted forest patches, and indirectly, through lengthening of the dry season as a result of changes to the timing of burning. However, the impact of any such anthropogenic forcing may have been entirely overshadowed by the effects of natural climate change and variability, as well as the generally low nutrient status of Australian soils. A robust assessment of the degree to which the environment of tropical Australia at the large scale has been modified from its 'natural' state as a result of human occupation will require new, coordinated collaborations between indigenous traditional landowners, archaeologists, anthropologists, geochronologists, geoscientists, ecologists, climatologists and modellers.

1. Introduction

'Oh no, not again' was the last thought that went through the mind of a bowl of petunias as it fell to the ground on an alien planet in 'The Hitchhiker's Guide to the Galaxy' (Adams, 1979: 103). Whatever else the petunias thought, it is clear that they had been in the same position previously. And so it is with the debate surrounding the environmental impact of the arrival of humans in Australia.

Well over a century ago, it was known that a number of large, now extinct, marsupial species had once been extant in Australia, and so began the debate as to whether climate change or 'Australoid wielders of clubs and throwing sticks' (Owen, 1877: cited in Horton, 1980: 86) were responsible for their demise (see Horton, 1980 and Johnson, 2006 for reviews).

Over four decades ago, Merrilees (1968) and Jones (1968) independently postulated that humans were implicated in megafaunal extinction, and Jones (1968: 189) suggested that it was likely that humans had arrived in Australia much earlier than previously thought: 'For the present an antiquity of 30,000 years is a reasonable claim'. Jones (1969) elaborated on his earlier work proposing that humans had manipulated the environment since arrival through 'firestick farming'.

Over three decades ago, Kershaw (1974, 1986) first suggested that Aboriginal burning had led to a sustained decline in fire-sensitive gymnosperm taxa and an increase in sclerophyll and grass taxa. This conclusion was based on a pollen record from a single site: Lynch's Crater on the Atherton Tablelands of north Queensland. Also over three decades ago, Horton (1982: 238) noted that 'There have been a few criticisms of this general thesis [that humans caused megafaunal extinction]'. On reviewing the information available at the time, he found little evidence to support human agency in either megafaunal extinction (Horton, 1980) or modification of vegetation through anthropogenic manipulation of fire regime (Horton, 1982).

Over two decades ago, Flannery (1990) proposed that after megafaunal extinction brought about by 'blitzkrieg' over-hunting, humans responded to subsequent increased fuel loads by firestick farming, thereby assisting in the maintenance of biodiversity in mediumsized mammals and some plants. Release from this anthropogenic fire regime upon arrival of Europeans resulted in another 'trophic cascade' of extinctions. Thermoluminescence (TL) and optically stimulated luminescence (OSL) dating was used by Roberts *et al.* (1990, 1994) to push back the proposed date of human occupation of mainland Australia to between 50 and 60ka, while Nanson *et al.* (1992) used TL and uranium-series dating methods to document a number of wet and dry phases in central and south-eastern Australia over the last 300 ka. Miller and Magee (1992) went further and proposed that Lake Eyre is dry in the current interglacial as a result of substantial modification of vegetation across tropical Australia following the imposition of an anthropogenic fire regime.

Over a decade ago, combinations of more robust radiocarbon pretreatment techniques, OSL and other dating techniques continued to suggest occupation of Australia occurred considerably before 40ka (Turney *et al.*, 2001a, Bird *et al.*, 2002, Bowler *et al.*, 2003), with some claims of occupation close to 60ka (Roberts *et al.*, 1998, Thorne *et al.*, 1999). Roberts *et al.* (2001) used OSL and uranium-series methods to date articulated megafaunal remains and their burial sediments at a number of sites to infer that extinction occurred in the interval of 40-51ka. Miller *et al.* (1999) and Johnson *et al.*, (1999) suggested extinction of the large flightless bird *Genyornis newtoni* and large-scale vegetation change was a direct result of human

arrival and the imposition of an anthropogenic fire regime at 45-55ka, perhaps underlain by natural climate change. We observe that, while there is abundant evidence for early human occupation across the tropical north, from the Kimberley to Cape York (Figure 1), there are comparatively few well-dated megafauna sites in tropical Australia. However, the widespread existence of notionally 'Late Pleistocene' megafaunal remains in the region (Field *et al.*, 2008: Figure 12) suggests that they were a significant component of the fauna encountered by humans on arrival.

At each stage of the debate in the twentieth century, claims for earlier arrival of humans on the Australian mainland, as well as the assertion that humans were responsible for megafaunal extinction and/or significant ecosystem modification through firestick farming, have been vigorously contested. Opposition has generally taken the form of disputing the reliability of the dating, the stratigraphic association between dated material and physical evidence, the selective use of evidence, conflicting evidence and/or insufficient evidence, coupled with the general observation that fire was a significant factor in tropical Australia prior to human arrival (Horton, 1980, 1982, O'Connell and Allen, 1998, Bowman, 1998, Mulvaney and Kamminga, 1999).

It is not our purpose to review in detail the development of the debate on the timing of human arrival and megafaunal extinction, though we do summarize debate from the last decade. Rather, we focus instead on the development of two divergent, testable, endmember hypotheses for the impact of human occupation on the landscape-scale ecology and climatology of tropical Australia. In particular:

 (i) that human dispersal through and occupation of tropical Australia had little or no impact on regional ecosystems and climate; or (ii) that human dispersal through and occupation of tropical Australia either rapidly or cumulatively over many millennia, resulted in a profound change to regional ecosystems and climate.

We focus on the literature that bears on the development of these hypotheses rather than on making a detailed assessment of the burgeoning literature surrounding the timing of human arrival and human impact on the environment following arrival. We constrain our area of interest to the Australian monsoon tropics as defined by Bowman *et al.* (2010): those areas of Australia that (currently) receive more than 85% of rainfall between November and April.

2. Timing of human colonization

Early human occupation at sites in tropical Australia such as Ngarrabullgan, Nauwalabila 1, GRE-8, Riwi and Carpenter's Gap has been established since the 1990s, although exactly how early remains the subject of debate (e.g. O'Connell and Allen, 2012). In the last decade, further sites with dates interpreted as suggesting human occupation at or before 40ka have been identified across tropical Australia: Nawarla Gabarnmang (David *et al.*, 2011), Parnkupirti (Veth *et al.*, 2009) Nonda Rock (David *et al.*, 2007), and GRE-8 (Slack *et al.*, 2004) (Figure 1) – as well as Papua New Guinea (e.g. Summerhayes *et al.*, 2010). While the absolute antiquity of initial colonization remains contested, there appears to be emerging consensus that humans had arrived in Australia and had expanded 'archaeologically instantaneously' across much of mainland Australia and Papua New Guinea by not later than 44-46ka (Hiscock, 2008, O'Connell and Allen, 2012). There is little evidence upon which to base inferences as to the number or size of colonization events or the likely population after colonization. Nevertheless, O'Connell and Allen, (2012: 12) contend that after arrival and 'despite their inherent capacity for growth, Sahul populations remained surprisingly small and spatially concentrated in persistent 'sweet spots'. Williams (2012) used 2996 radiocarbon dates from 800 archaeological sites in Australia to explore occupation trends from 40ka to the present. While subject to considerable uncertainty, the results from that study suggest approximately constant levels of occupation from the beginning of the record, with occupation levels increasing into the Holocene (see also Johnson and Brook, 2011).

For the purposes of this paper, we conclude simply that humans had arrived on the Australian mainland by at least ~45ka (Hiscock, 2008, Hiscock and Wallis, 2005), a date now consistent with the timing of migration through mainland and island SE Asia (Higham *et al.*, 2009, Demeter *et al.*, 2012) and Papua New Guinea (Summerhayes *et al.*, 2010). We further conclude that within a few millennia, humans had dispersed throughout mainland Australia (O'Connell and Allen, 2012), with the possible exception of parts of the arid interior (Smith *et al.*, 2008).

3. Megafaunal extinction: The Rasputin Syndrome 7

Recognizing that claims for earlier arrival may be correct, the apparent consensus that humans had arrived on mainland Australia by ~45ka has marginally narrowed the range of debate concerning the timing and cause of megafaunal extinction: humans and megafauna clearly co-existed for a period of time, and hence humans could potentially have been responsible for at least a component of megafaunal extinction. Gunn *et al.* (2011) report the existence of rock art in Arnhemland that appears to depict the extinct giant bird *Genyornis* *newtoni* (Figure 2) and Flannery (2012) reports rock art in the Kimberley region that has been interpreted as depicting the extinct marsupial lion (*Thylacoleo carnifex*). Both imply direct co-existence (for other rock art examples see Murray and Chaloupka, 1984). If humans and megafauna did overlap in time, it is difficult to envisage a situation where humans and megafauna could co-exist on the same continent without any form of interaction.

Correlative evidence in the last decade supporting rapid extinction coincident with human arrival has come from Tasmania (Turney *et al.*, 2008, Gillespie *et al.*, 2012) and north Queensland, the latter inferred from the abrupt decline in *Spororomiella* fungal spores – indicative of mega-herbivores – in the Lynch's Crater record at around 41ka, coincident with an abrupt increase in charcoal (Rule *et al.*, 2012). At the site of Cuddie Springs in western New South Wales, claims for survival of megafauna for at least 10 ka after human arrival (Field *et al.*, 2008) have been countered by Grün *et al.* (2010), who concluded from direct electron spin resonance (ESR) and uranium-series dating of megafauna teeth and bones that the Cuddie Springs material had been reworked from older deposits. This evidence is 7 consistent with earlier suggestions of sediment disturbance (Roberts *et al.*, 2001, Gillespie and Brook, 2006).

Arguments against a human-induced extinction 'event' have continued to appear in the last decade (Wroe and Field, 2006, Field and Wroe, 2012). Price *et al.* (2011) used OSL and uranium-series dating techniques to constrain the ages of multiple taxa in the Darling Downs in southeastern Queensland and concluded that 15 taxa disappeared from the local record over a staggered time interval between 122 and 83ka ago. The authors considered that megafaunal extinction was a process underway for some time prior to the arrival of humans, in response to a long-term trend to greater aridity. Faith and O'Connell (2011) re-examined

the data of Prideaux *et al.* (2010) from Tight Entrance Cave in southwest Western Australia and concluded that the evidence was consistent with environmentally-mediated extinction, with uncertainties in the dating making it impossible to exclude the possibility that the majority of megafauna were extinct by the penultimate glacial maximum (MIS-6), but that uncertainties in the dating made it impossible to discriminate between gradual attenuation over the last few glacial cycles and synchronous extinction between about 50 and 40ka. By contrast, Prideaux *et al.* (2010) noted that the MIS-6 had no lasting impact on the regional fauna, which argued against climate change as the sole or primary driver of the extinction. The latter interpretation is also consistent with the only other faunal sequence in Australia with a paired climate record extending back 300ka – namely, Cathedral Cave in South Australia (Prideaux *et al.*, 2007).

The absence of clear evidence of human predation on, or consumption of megafauna, also continues to be advanced as an argument against human involvement in megafaunal extinction (Davidson, 2012), although this issue had earlier been addressed by Brook and Johnson (2006), Roberts and Brook (2010) and, more recently, by Surovell and Grund (2012) who conclude that such evidence in Australia is likely to be very rare even in comparison to evidence from parts of the world occupied by humans more recently (e.g. North America and New Zealand).

Miller *et al.* (2005b), building on their earlier work, concluded that *Genyornis* became extinct as a result of anthropogenic burning between 50 and 45ka. On the other hand, Murphy *et al.* (2010), using the same stable isotope analyses of *Dromaius* eggshell reported by Miller *et al.* (2005b), concluded that ecosystem change began before the arrival of humans and could be better modelled as a gradual change in response to overall decreasing water

availability between 80 and 30ka. Hence, ecosystem change due to the imposition of an anthropogenic fire regime was not necessarily the primary cause of the extinction of *Genyornis*.

Gregori Rasputin, a monk and a controversial member of the court of the Russian Tzar through the early part of the twentieth century, was variously poisoned, shot and clubbed to death in 1916. Which of these violent actions caused his ultimate demise remains a subject of debate and it is also possible that he survived the initial attempts on his life, lingering on for a period before finally succumbing to drowning after being dumped through a hole in the ice on the Neva River. And so it is with megafaunal extinction. The potential drivers of extinction can be summarized as direct human impact in the form of direct predation or selective predation of juveniles (Brook and Johnson, 2006), indirect anthropogenic ecosystem change through modifications to fire regimes, introduced disease or natural climate change (for a review see Johnson, 2009). It has been argued that extinction followed rapidly on the heels of human occupation (Roberts et al., 2001, Miller et al., 2007, Turney et al., 2008, Gillespie et al., 2012) but the possibility cannot yet be excluded that some of the megafauna that were extant at human arrival lingered on for some time (Johnson, 2005, 2006, Field and Wroe, 2012). In that context, it is worth noting that for communities subject to non-pluvial climate regimes there is a strong correlation between precipitation, plant productivity and both total herbivore biomass (Coe et al., 1976) and large herbivore diversity (Olff *et al.*, 2002); Thus it is likely that megafaunal populations were already in decline due to increasing aridity at the time of the first human occupation. Such declining populations would be also be expected to have been more sensitive to anthropogenically-induced changes in environment (or direct hunting pressure) than would have been the case for a

more stable or expanding megafaunal population (Hubbell, 2001). Thus, the two hypotheses are not necessarily exclusive.

Field and Wroe, (2012: 69) conclude 'that the megafaunal extinction debate in Australia has a long way to go before even the most fundamental of questions are convincingly resolved'. For the purposes of this paper, the details of the mechanism and timing of extinction are not important, except insofar as determining whether human agency was involved. This is because mega-herbivores in modern tropical environments are capable of a significant degree of 'engineering' of the ecosystems they inhabit (Asner *et al.*, 2009, Staver *et al.*, 2009, Midgley *et al.*, 2010, Tanentzap and Coomes, 2012). Mega-herbivore browsers are partly responsible, along with fire and other edaphic controls, for the maintenance, and in some cases extension, of open vegetation types. The removal of megafauna from Australian tropical ecosystems is therefore likely to have resulted in some degree of ecological transformation (Johnson, 2009), and this would be considered anthropogenic if humans were responsible, in part or in whole, for megafaunal extinction.

We conclude that the results of Roberts *et al.* (2001), which were based on dating of articulated extinct megafaunal remains only, do indicate that megafauna were dramatically reduced in numbers from, to use the terminology of Johnson (2005), 'palaeontological visibility' to 'archaeological visibility' during a period of a few millennia following human arrival. As a result, and regardless of whether they became literally extinct at that time, megafauna were thereafter unlikely to be able to exert a significant influence on vegetation structure in tropical Australia. We do not speculate in detail on the likely cause of this reduction in numbers, other than to note that, of the mechanisms identified, climate change as the sole driver of extinction represents the least likely possibility; a conclusion that has

previously been reached by several authors (Owen-Smith, 1987, 1989, Miller et al., 2007, Brook and Bowman, 2004, Prideaux et al., 2007, 2009, 2010, Murphy et al., 2012).

While we acknowledge a long-term trend to increasingly arid conditions in some parts of Australia over the last 300ka (Nanson *et al.*, 1992, Magee *et al.*, 2004, Wroe and Field, 2006, Webb, 2008), there is no evidence for 'sharp, intra-decadal through millennial-scale fluctuations in temperature and precipitation' (O'Connell and Allen, 2012: 7) during the period of interest through Marine Isotope Stage 3 (MIS-3: 28-59ka). High latitude northern hemisphere ice core records and temperate northern hemisphere speleothem records do indicate significant millennial-scale variability associated with Dansgaard-Oeschger (D/O) cycles (e.g. Weninger and Jöris, 2008). However, the expression of D/O events is very muted in the southern hemisphere ice core climate records from Antarctica and the signal associated with D/O events is likely to be even more muted in low latitude locations such as tropical Australia. Hesse *et al.* (2004) concluded that the 'largest changes in circulation patterns [in Australia] over the glacial cycle probably occurred in the location and/or intensity of summer tropical convergence in northern Australia', but there is no evidence to suggest that climate variability in MIS-3 on any temporal scale was more extreme than at any time prior or subsequent to the interval during which extinction occurred.

The discussion above leads us to conclude that human occupation of the continent did play a direct and decisive, if currently unquantified, role in the rapid reduction of megafaunal populations to levels at which they could no longer exert an influence on the vegetation structure of tropical Australia. The possibility cannot be excluded, however, that other environmental factors may have played a role - as yet unquantified - in eventual megafaunal extinction (Roberts *et al.*, 2001, Koch and Barnosky, 2006, Prideaux *et al.*, 2010, Lorenzen *et* *al.*, 2011, Prescott *et al.*, 2012). In the absence of other human impacts, extinction of the
megafauna may have itself led to modifications to ecosystem function through changes to
the nature and abundance of fuel loads and changes in the density of woody vegetation
through release from browsing pressure.

273 4. Eco

4. Ecological and climatological impacts of occupation

274 Much of eastern tropical Australia has been argued to fall into the category of 'ecosystem' 275 uncertain', where the present-day climate is such that a range of vegetation states are 276 potentially possible (Bond and Keeley, 2005; Staver et al., 2012; Hirota et al., 2012). Although 277 such results have been interpreted as supporting the notion of forest and savanna 278 representing alternative steady states in such climates, with the latter induced and then 279 maintained by fire (e.g., Hoffmann et al., 2012; Murphy and Bowman, 2012), it is important 280 to recognise that factors such as soil cation status which are known to sometimes be 281 substantially higher for forest and savanna vegetation formation types (e.g. Cochrane, 1989) 282 are not accounted for in such 'climate only' simulations. Similarly, soils underlying more 283 open savanna formation types also tend to have a lower cation status than for nearby 284 woodier savanna formations (Lopes and Cox, 1977) and given the low fertility of Australian 285 tropical soils in general (McKenzie et al., 2004) edaphic factors may account for much of the 286 region currently being savanna rather than forest. Nevertheless, significant large-scale 287 anthropogenic modification to vegetation cover sustained over many millennia to the 288 present day is at least a possibility. Human occupation could potentially have modified the 289 vegetation cover of tropical Australia in three ways, with all potential mechanisms operating 290 through an influence on the dynamics of tree-grass competition.

291 First, the removal of any browsing megafauna would improve opportunities for 292 establishment of trees and shrubs (Bond and Keeley, 2005, Staver et al., 2009). Second, and 293 alternatively, if the bulk of the megafauna were browsers rather than grazers, then an 294 increased standing crop of herbaceous biomass may have resulted, with increased fire 295 frequencies/intensities ensuing (Flannery, 2012). Third, any deliberate change in intensity, 296 number or timing of fires would also be expected to modify the recruitment opportunities 297 for trees and shrubs relative to grasses (Bond, 2008, Bond and Midgley, 2012, Lehmann et al., 298 2011, Nano and Clarke, 2011, Prior et al., 2009, 2010, Russell-Smith et al., 2010, Scott et al., 299 2012). A shift towards more trees, for example, is thought to potentially result in a positive 300 feedback, leading to further vegetation thickening as a result of the suppression of grass 301 growth and, hence, the ability to carry fire, and vice versa (Sankaran et al., 2005, Bond and 302 Midgley, 2012, Higgins and Scheiter, 2012, Murphy et al., 2012).

303 It is important to note that a change in the timing of fires does not necessarily imply a 304 change in the total burnt area, and hence a change in fire regime may not be reflected in a 305 simple change in charcoal abundance in sedimentary records (Mooney *et al.*, 2011). It is also 306 important to note that any anthropogenic modification to fire regime and land cover would 307 necessarily have occurred against a background of changes due to natural forcing primarily 308 associated with changes in atmospheric CO_2 (Bond and Midgley, 2012, Higgins and Scheiter, 309 2012) and climate from the time of human occupation to the present (Mooney *et al.*, 2011).

There is no doubt that the 'indigenous' (anthropogenic) fire regime in tropical Australia is not the same as a 'natural', pre-human fire regime (e.g. Burrows and van Didden, 1991). In the monsoon tropics of north Australia, a natural fire regime would likely have been low frequency due to an absence of significant natural ignition sources (lightning) over most of

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314 the seasonal cycle, but of higher severity due to fuel accumulation between the more 315 infrequent fire returns. Most recently, Bliege Bird et al. (2012) clearly demonstrated 316 substantial differences in fire regime between areas subject to indigenous fire management 317 and areas that are not subject to management, in the spinifex grasslands of northern Western 318 Australia. Areas under indigenous fire management are subject to more but smaller fires, 319 buffering climate-driven variability in fire size and resulting in increased patch richness, 320 diversity and evenness (Yibarbuk et al., 2001, Yates et al., 2008). Price et al. (2012) note that 321 the depopulation of indigenous land custodians from northern Australia in the early to mid-322 twentieth century resulted in a shift from a larger number of small, early dry-season fires to 323 fewer but larger, more intense wildfires in the late dry season. Release from an indigenous 324 fire regime has also seen considerable additions to forest area across tropical Australia, 325 although this interpretation is complicated by the likelihood that increasing CO₂ or changes 326 in precipitation regime are also contributing to forest expansion (Brook and Bowman, 2006, 327 Donohue et al., 2009, Bowman et al., 2010, Tng et al., 2012).

328 Archibald et al. (2012) have recently suggested that humans in Africa successively acquired 329 the ability to manipulate fire frequency and season prior to 100ka and, finally, fuel 330 connectivity with the advent of agriculture/pastoralism around 10ka. They conclude that, as 331 in Australia, a lightning fire regime was characterized by seasonal fire distributions different 332 from modern anthropogenic fire regimes, with major effects on the size and intensity of fires. 333 They also conclude that substantial human impact on total burned area was probably not 334 initiated until the middle Holocene in open landscapes, whereas humans could have altered 335 fire regimes in closed/dissected landscapes by around 40ka.

336 Research into the interactions between fire, indigenous or natural, on ecosystems in 337 northern Australia can directly examine a time window encompassing the last half-century or 338 so (for review see Bowman, 1998). The ecosystems that were present at the time of 339 European arrival are generally considered to be the benchmark for what is 'natural'. The 340 extent to which indigenous fire regimes that had been in place for tens of millennia prior to 341 European arrival had led to the modification of land cover in tropical Australia is currently 342 unknown, but is of critical importance because of the potential feedbacks between land 343 cover and climate that operate on a range of spatial scales and of timescales, from annual to 344 millennial and longer (Beerling and Osborne, 2006, Bond and Midgley, 2012).

345 It has been argued that extensive regional scale tree cover can exert a direct influence on 346 meso-scale climate (Shiel and Murdiyarso, 2009), with tropical forests recycling water by 347 transpiration at higher rates compared to grasslands (Zhang et al., 2001) and extracting water 348 from greater depths within the soil (Hayden, 1998, Beerling and Osborne, 2006). 349 Vegetation-climate feedbacks have been suggested to exist for regions ranging from the 350 Amazon Basin (D'Almeida et al., 2007) to the Sahel (Los et al., 2006). Spracklen et al. (2012) 351 found that for 60% of the tropical land surface, including much of central Australia, air that 352 had previously passed over areas of extensive vegetation produced at least twice as much 353 rain as air that had passed over little vegetation in the few days prior to the rainfall. In similar 354 vein, recent studies in Australia have attributed regional reductions in rainfall to vegetation 355 clearing, where conversion of forest to croplands that could function similarly to shallow-356 rooted grasslands (Pitman et al., 2004, Deo, 2011).

357 Miller and Magee (1992) first proposed that the imposition of an anthropogenic fire358 regime in northern Australia could have reduced tree/shrub cover across tropical Australia,

359 to the extent that the degree of penetration of monsoonal rains into the continental interior 360 was reduced, resulting in the anomalous comparatively dry state of Lake Evre in the 361 Holocene (Magee et al., 2004). Miller et al. (2005a) later used a modelling approach to explore 362 the potential feedbacks between vegetation and climate in tropical Australia assuming the 363 extreme scenarios of a pre-human landscape of "broadleaf deciduous trees on sandy loam 364 soils and landscape after occupation of "desert vegetation on sandy soil". The study 365 concluded that tree cover in tropical Australia exerts a significant control on monsoon 366 penetration, with higher amounts of tree cover leading to deeper penetration of monsoon 367 rains into the continental interior. In contrast, Pitman and Hesse (2007), found little 368 response of the monsoon to changes in vegetation, a result subsequently criticized by Miller 369 et al. (2007) on the basis that Pitman and Hesse (2007) specified modern insolation for 370 Holocene simulations.

371 In another modelling study, Marshall and Lynch (2006) inferred only a muted response of 372 the monsoon to vegetation change, whereas the modelling results of Lynch et al. (2007) 373 suggested that late, high intensity fires could increase monsoon precipitation in modern northern Australia by up to 31%, (although this response was obtained from an artificially 374 375 high burnt area of 90% and high severity fire events). The most recent and comprehensive 376 simulations to date, (Notaro et al., 2011), assuming only a 20% reduction in vegetation cover, 377 found a non-significant effect of this vegetation change on peak monsoon period 378 precipitation (January to March) but a significant 40mm reduction in rainfall in the pre-379 monsoon (October to December) season, thus effectively lengthening the dry season.

380 Nevertheless, even a 20% reduction in total vegetation cover as assumed by Notaro *et al.*,
381 (2011) may be not be realistic, as the real driver of change in rainfall is change in latent and

382 sensible heat fluxes associated with the anthropogenically-induced vegetation change (e.g., 383 Grace et al., 1998). And here we note that, contrary to some claims (e.g., Flannery, 2012), 384 savanna fires tend to accelerate rates of nutrient cycling (van de Vevjer et al., 1999) with net 385 fire-induced nutrient losses usually minimal at a regional scale due to the subsequent return 386 of emitted nutrients through wet- and dry-atmospheric deposition (Delon et al., 2012; Kugbe 387 et al., 2012). Thus, with a stimulation of vegetative growth, if anything, latent heat fluxes may 388 actually be greater for invigorated savanna vegetation regrowing after low to moderate 389 severity fire events (Santos et al., 2003) and even where large effects of long-term fire regime 390 on woody vegetation cover have occurred stand-level evaporation rates are minimally 391 affected (Quesada et al., 2008).

392 The major source of natural ignition is lightning and Kilinc and Beringer (2007) found 393 that lightning strikes in the Northern Territory are concentrated in the wet season and during 394 transition periods between seasons. Very few dry season lightning strikes were observed. 395 They also found that lightning strikes were more common in grasslands relative to woodier 396 ecosystems. This was argued, at least in part, to be due to the higher sensible heat flux from 397 grasslands than surrounding vegetation types within the landscape mosaic and hence a 398 greater potential for convective activity directly above. Bowman et al. (2007) noted a link 399 between late dry season fires and rainfall with one explanation being the dynamic effects of 400 aerosols released by burning invigorating convection, with the formation of an enhanced and 401 higher cloud cover than would otherwise be the case: this leading to higher rainfall (Andreae 402 et al., 2004; Lin et al., 2006).

403 On balance, the studies cited above provide some support for the possibility that 404 anthropogenic fire regimes could have, over an extended period and through a number of

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405 fire-biosphere-atmosphere feedbacks, resulted in an appreciably different climate and land 406 cover across northern Australia than would be the case had humans not occupied the 407 continent. We explore the mechanisms in the following section, but note the following, 408 inconclusive, evidence supporting this possibility:

409 (i) the most cited evidence for a change in fire regime following human arrival leading to 410 a change in vegetation comes from the record of Lynch's Crater (Kershaw et al., 2007). In 411 this record, a sustained increase in charcoal flux at around the time of human arrival (Turney 412 et al., 2001b, Rule et al., 2012) is accompanied by a dramatic decrease in gymnosperm taxa 413 and concomitant increase in sclerophyll and grass taxa. It is also the case that fire-sensitive 414 *Callitris* species disappear from the record, probably before human arrival, but never to 415 reappear at similar levels of abundance despite having been abundant in the penultimate 416 glacial and last interglacial (Kershaw et al., 2007), This observation has also been made for 417 the Lake Frome region by (Luly, 2001).

418 The interpretation of the Lynch's Crater record as indicative of a broad change in fire 419 regime in northern Australia is complicated by an observation originally made by Kershaw 420 (1976: 492): 'This area is climatologically sensitive and it may be that the vegetation changes, 421 and the climatic shifts implied by them, were exaggerated there as compared with the rest of 422 north-eastern Australia'. The rainforest or rainforest fringe areas that are typical of the 423 Atherton Tablelands are climatically atypical of northern Australia, which is dominated by 424 lowland sclerophyll woodlands and shrublands occurring on vast sand-sheets of low relief 425 (Ash, 1983, Shulmeister, 1992). It is therefore undesirable to extrapolate northern Australian 426 palaeoenvironments solely on the basis of the Atherton Tablelands pollen record as a

427 significant impact on vegetation cover at Lynch's Crater may not equate to a significant428 impact on vegetation across the majority of tropical Australia.

429 A potentially more representative pollen record for tropical Australia is provided from the 430 marine sediment core (SHI-9014) of van der Kaars et al. (2000). This record, from the Banda 431 Sea and therefore sampling a broad area of northern Australia and the Sahul shelf (but also 432 including a southeast Asian component), is interpreted as indicating a significant expansion 433 of grassland and a concomitant decline in Eucalyptus in northern Australian and the Sahul 434 Shelf region from 37ka ago, sustained to the present-day. Further evidence for a sustained 435 change comes from the disappearance of Olea pollen, indicative of vine thickets, from 436 marine core MD98-2167 at around 45ka (Kershaw et al., 2011). In neither of these cases are 437 changes in charcoal abundance closely associated with the changes in pollen abundance, nor 438 are there abrupt changes in charcoal abundance coinciding with initial human arrival.

(ii) As discussed above, the dominantly dry modern condition of Lake Eyre is anomalous
in the context of its record of previous episodes of filling under analogous climate
conditions (Magee *et al.*, 2004). More recent work on Lake Mega-Frome, which was last
connected to Lake Eyre 50-47ka ago and is also currently dry, has suggested a greater
importance for Southern Ocean sources, but still posits a significant influence from tropical
moisture sources in the Holocene in determining its state (Cohen *et al.*, 2011).

(iii) The carbon isotope estimates of the proportion of C_4 grass in the diet of both emus and wombats, which suggested a mixed C_3 and C_4 diet prior to ~45ka and including the last interglacial and penultimate glacial periods, indicate a permanent shift to a lower proportion of C_4 biomass in the diet from 45ka until the present time (Miller *et al.*, 2005b, 2007). While these results derive from more southerly latitudes than tropical Australia, they do imply a

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450 large and permanent shift in environmental conditions across tropical Australia at some time451 after 45ka.

452 (iv) In tropical Australia, tropical dry forest types, with no perennial access to water but 453 often with a measure of fire protection, extend in patches into regions of <600mm rainfall 454 (Fensham, 1996, Miles et al., 2006) and are common but widely dispersed in regions that 455 receive >600mm (Figure 1). These patches have been shown to have both expanded and 456 contracted over adjacent areas in the recent past, suggesting there is no general soil-imposed 457 limit on their potential (local) distribution above 600mm rainfall (Bowman et al., 1994). 458 Murphy et al. (2010) conclude that both fire and soil fertility control the relative distribution 459 of mulga and spinifex in the more arid interior.

460 Sankaran et al. (2005) has argued that in African savannas areas above 650mm are 461 'disturbance' savannas, where fire and/or herbivory are required to prevent canopy closure, 462 although Lloyd et al. (2008) have pointed out that there are other explanations for their 463 observations. Nevertheless, if we accept Sankaran et al.'s (2005) thesis, then the observation 464 of Archibald et al. (2009) is pertinent - that fires become uncommon in Africa when rainfall 465 exceeds ~ 800 mm and canopy cover exceeds 40%. This is because, in the tropical Australian 466 context, Bond (2008; Figure 1) has demonstrated that, in contrast to African savannas, 467 modern Australian savannas fail to reach an African-type rainfall-determined maximum 468 cover below 650mm. Above this threshold, where closed canopy vegetation should (based 469 on the conclusions of Sankaran et al. (2005) for Africa), represent a much woodier stable 470 state in the absence of disturbance.

Thus Australian savanna continues to persist to a much higher mean annual rainfall thanAfrica with one explanation being that long-standing indigenous fire regimes have resulted in

473 large areas of savanna in northern Australia maintained by fire-mediated feedbacks 474 preventing the return of the vegetation to an alternative, forest type, vegetation formation 475 (Warman and Moles, 2009; Lawes et al., 2011; Murphy and Bowman, 2012). Nevertheless, 476 we add a note of caution that whilst the notion of forest and savanna representing two 477 alternative steady states is conceptually attractive but is contested (House et al., 2003, 478 Lehmann et al., 2011). We also note that Fensham et al. (2005) examined changes in woody 479 cover in the 500-800mm rainfall zone of central Queensland and found that neither fire nor 480 grazing exerted significant control, with variations in woody cover driven largely by 481 variations in relative rainfall since the 1940s.

482 In summary: (i) it has been suggested by some modelling studies that forest vegetation 483 can potentially exist over a larger area of tropical Australia than is currently the case, where 484 local soil conditions allow; (ii) current climate and fire regime clearly favour the maintenance 485 of open vegetation across much of tropical Australia; (iii) a number of potential feedbacks, 486 some highly non-linear, have been suggested to operate between vegetation and climate, 487 potentially resulting in alternative states of climate and vegetation; (iv) a change in fire regime 488 theoretically provides one mechanism by which one state may be advantaged over another; 489 and (v) there is evidence that fire regime has been manipulated by humans to some degree 490 since their arrival in tropical Australia and across Sahul.

491

5. Towards testable hypotheses

What is striking in the discussion to date concerning the impact of human arrival is the
relative lack of consideration of the dynamic ecological processes that would necessarily
underlie any impact associated with human arrival (but see Bowman, 1998, Johnson, 2009).
It is clear that many of the landscapes and ecosystems encountered upon European arrival

were 'naturalized' rather than 'natural' (e.g. Bliege Bird *et al.*, 2012). However, the degree to
which humans may have affected vegetation and/or climate at more than the local scale in
Australia – 'The Biggest Estate on Earth' (Gammage, 2011) – remains unresolved after
decades of debate.

Based on the discussion in the preceding sections, we elaborate below on two testable 'end-member' hypotheses concerning the broad-scale impact of an anthropogenic fire regime in the tens of millennia following human arrival, recognizing that a continuum of possible scenarios lies between them.

504

5.1 One minus one equals zero: no change

505 This scenario represents the null hypothesis and is straightforward to articulate. The 506 removal or dramatic diminution of megafauna, and particularly browsing megafaunal 507 herbivore populations, removed an impediment to tree recruitment and survival. The 508 increased potential recruitment and survival of trees in tropical Australia was offset by the 509 introduction of an anthropogenic fire regime. The net effect of an anthropogenic fire regime 510 was to more or less maintain the status quo, with changes in the 'spatial grain' of ecosystem 511 processes favouring enhanced biodiversity and resource availability at the local scale (Bliege 512 Bird et al., 2008). Whilst, over an extended period of time, this fire regime may have been 513 instituted across a large area, a population probably never numbering more than one million 514 people (Butlin, 1983), and possibly considerably fewer (Atkinson et al., 2008, Eriksson et al., 515 2012), was not capable of significant large-scale influence on ecosystems.

The relative areas of ecosystems at the broad scale were determined entirely by natural changes in regional climate, in turn forced by global changes in sea-level and insolation, modulated by the dynamics of the Asian monsoon and regional oceanic and atmospheric

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519 variability, overprinted by a long-term trend to greater aridity in Australia (Magee et al., 2004, 520 Webb, 2008, Cohen et al., 2011, Field and Wroe, 2012). A 300ka record of vegetation change 521 from the Timor Sea, northwest and downwind of tropical Australia, provided by Wang et al. 522 (1999), shows no evidence of a significant change in the balance between tree and grass 523 pollen, or in charcoal abundance, coincident with human arrival or at any time subsequent to 524 human arrival. Indeed, the records of early explorers suggest that in the arid and semi-arid 525 tropics, indigenous fire may have been infrequent and restricted to intermittent favourable 526 seasons (Silcock et al., 2013).

527 Peaks in charcoal abundance prior to human arrival in core ODP-820 from off northeast 528 Queensland (Moss and Kershaw, 2007) and core GC-17 off Western Australia (van der 529 Kaars and De Deckker, 2002) indicate that this proxy cannot be reliably interpreted as 530 indicative of anthropogenic fire. Indeed, in their review of charcoal records from the region, 531 Mooney et al. (2011) found no 'distinct change' in charcoal abundance after human arrival in 532 Australasia, though a possible trend to increased biomass burning between 50 and 40ka is 533 one possible interpretation of the composite record. They note, however, the 'considerable 534 uncertainty' associated with the chronology of records that lie beyond the limit of 535 radiocarbon dating, and their composite record for Australasia as a whole may also conceal 536 significant trends at finer spatial scales.

537 Modern tree and grass distributions in tropical Australia are largely explained by low 538 nutrient soils, high natural climate variability (Fensham *et al.*, 2003, Lehmann *et al.*, 2009, 539 Fisher *et al.*, 2012) and by the dominance of fire-adapted eucalypts (Lawes *et al.*, 2011, Crisp 540 *et al.*, 2011). Increased atmospheric CO₂ and rainfall on their own can explain the observed

- 541 forest encroachment and thickening of savannas over the twentieth century (Bond et al.,
- 542 2003, Bowman et al., 2010, Fensham et al., 2005, Lawes et al., 2011, Tng et al., 2012).
- 543 5.2 One plus one equals three: profound change

544 Significant human influence on vegetation in tropical Australia could have resulted from the 545 effects of an anthropogenic fire regime on climate-vegetation feedbacks through two 546 mechanisms. The first is indirect, operating through a change in the timing of burning, while 547 the second is direct, from targeted burning at forest boundaries and of newly irrupted forest 548 patches. Both mechanisms operate cumulatively on centennial to millennial timescales, with 549 landscape-scale change not readily observable across several human generations, although 550 local changes would be observable on decadal timescales, as is currently the case (Banfai and 551 Bowman, 2005). Neither mechanism requires 'more' fire than a natural fire regime, simply a 552 change in the frequency, timing and focus of burning. Hence, there is no requirement for a 553 discernible change in biomass burning proxies accompanying human arrival (Mooney et al., 554 2011).

555 Neither mechanism requires a large population. This is because an area under an active 556 anthropogenic fire regime by definition shares a boundary with areas not under active 557 management. Thus, human-lit fires could be expected, under favourable conditions, to carry 558 over into areas not actively managed, thereby leading to change in the timing of burning in 559 areas not actively managed and a more constant source of ignition than provided by 560 lightning alone. In glacial and early post-glacial times, arid conditions may have also served 561 to focus a comparatively small population around forest edges, which would have provided 562 access to a wider resource base and accessible surface water in riparian areas.

Bond *et al.* (2003) modelled the impact of low CO_2 and fire and found that, at the CO_2 concentrations pertaining at the LGM, seedling growth rates were too slow to ensure growth to a 'fire-escape' height. Over the 25–30ka from human arrival to the LGM, incremental introduction of fire at a higher recurrence interval, early in the dry season and at low CO_2 , may have exacerbated a natural trend to stem thinning in open environments and a reduction in forest area in response to drying and cooling of the continent (Hesse *et al.*, 2004).

570 After the LGM, climate began to warm, CO₂ began to rise and the monsoon re-571 established across tropical Australia by ~14ka (Wyrwoll and Miller, 2001). In the absence of 572 humans, this combination of factors, coupled with the absence of browsing megafauna, 573 would have improved the establishment and persistence potential for trees in northern 574 Australia. Given the vast expanse of woodlands across this region, an increase in tree cover 575 would have initiated a positive feedback, increasing the evapotranspirative movement of 576 moisture into the continental interior and enabling tree establishment further towards the 577 centre of the continent (Miller et al., 2005a). Late dry-season lightning fires had the effect of 578 reducing dry season length (Notaro et al., 2011) and/or increasing wet season precipitation 579 (Lynch et al., 2007), further favouring tree establishment. Examination of the current degree 580 of seasonality in precipitation across tropical Australia (Figure 1) indicates that even a 5% 581 change in the proportion of rain falling outside the wet season would dramatically change 582 the climate of much of northern Australia. Once tree cover passed a threshold, equivalent to 583 ~40% canopy cover in Africa (Archibald et al., 2009), the incidence of fire would be reduced, 584 further promoting the establishment of woody vegetation formations across a much broader 585 area than is currently the case.

586 The above scenario was not, however, realised because the anthropogenic fire regime 587 operating in the post-LGM times was characterized by increased frequency of ignitions, 588 decreased return interval and a shift to more common early dry-season fires. This could 589 have had the indirect result of instituting a negative feedback, constraining tree 590 establishment through the maintenance of a longer dry season as well as constraining both 591 the intensity of the monsoon and its penetration into the continental interior. As a result, 592 canopy closure was inhibited and fire, mediated by humans, maintained the system in a 593 relatively open state across a larger area than would otherwise have been the case (Murphy 594 and Bowman, 2012).

595 This indirect negative feedback, hindering forest establishment, could have been further 596 amplified by targeted anthropogenic burning along forest boundaries. Certainly, some forest 597 environments in northern Australia were difficult to exploit, with limited or no occupation 598 of forests in the wet tropics until the Holocene (Cosgrove et al., 2007). This suggests that 599 indigenous populations may have preferred to maintain an open vegetation type in the face 600 of forest encroachment. In the Northern Territory, monsoon forest patches are currently 601 valued for a number of resources, particularly yams in the wet season, and both Haynes 602 (1985) and Russell-Smith et al. (1997) found that such forest patches are actively protected by 603 early dry-season burning outwards from their edges. While such a strategy may be effective 604 in protecting existing forest areas, in post-LGM times it may also have limited the expansion 605 of forest area by inhibiting tree seedling maturation beyond forest boundaries (Woinarski et 606 al., 2004). Frequent low intensity fires in open vegetation may also have deterred forest 607 establishment through inhibition of irruption of new forest patches in the landscape, as has 608 been demonstrated in the modern environment (Russell-Smith et al., 2004a, b). Indeed, given 609 the relatively slow rates of forest advance by edge progression, generally no more than a few

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610 metres per year (e.g. Favier *et al.*, 2004b; Tng *et al.*, 2012), inhibition of the formation of new
611 forest patches across the landscape might be the dominant mechanism by which forest
612 advance could be limited by anthropogenic fire.

613 The net effect of an anthropogenic fire regime applied for tens of millennia on the 614 observed modern vegetation cover, under this scenario, would be a more highly seasonal 615 climate with lower penetration of the monsoon into the continental interior. This may in 616 turn have fed back into (i) a more limited distribution of dry/monsoon forest cover over 617 northern and northeastern Australia, (ii) savannas with comparatively low-density tree cover 618 over much of the tropical north, (iii) the expansion of spinifex grasslands at the expense of 619 shrublands in the arid interior (Miller et al., 2007) and, ultimately, (iv) to the anomalously dry 620 condition of modern Lake Eyre (Magee et al., 2004). This scenario is consistent with results 621 from the long pollen record from core SHI-9014 in the Banda Sea (van der Kaars et al., 622 2000), which shows a significant expansion of grassland and concomitant decline in 623 eucalypts in northern Australia and the Sahul Shelf region from 37ka, sustained to the 624 present-day. Indeed, it is possible that the present link between increased savanna 625 woodiness and increased rainfall in Queensland savannas in the second half of the twentieth 626 century (Fensham et al., 2005) is partly the result of release from an anthropogenic fire 627 regime in the early twentieth century – a trend that may continue for centuries to millennia.

628

629 6. Conclusions: What is natural in tropical Australia?

630 There is abundant evidence that the vegetation patterns and biodiversity across tropical
631 Australia at the time of European arrival were, to some degree, anthropogenic (Jones and
632 Bowler, 1980, Johnson, 2006, Gammage, 2011), so the question 'what is natural?' is one of

633 spatial and temporal scale. We conclude that mechanisms do exist (supported by limited 634 evidence) to support the hypothesis that human occupation resulted in profound 635 modification to broad-scale vegetation patterns and climate across tropical Australia at the 636 millennial scale. However, human activity, primarily expressed through megafaunal 637 extinction and modifications to natural fire regime, occurred against a background of 638 significant natural climate change and generally nutrient-poor soils that may ultimately have 639 mitigated against any more than local expression of this potential for significant broad-scale 640 anthropogenic change.

It is not possible to discriminate unequivocally between the two hypotheses proposed above, or any intermediate between the two, because the data required to undertake a rigorous assessment do not exist. Obtaining the data required to draw a defensible conclusion will require closer collaborations between indigenous traditional land owners, archaeologists, anthropologists, geoscientists, climatologists and ecologists than has been the case to date. The evidence required to discriminate between these hypotheses can potentially come from several avenues:

(i) geosciences – the Lynch's Crater record has existed for almost 40 years, but is not
suited to assessing broad-scale human impact across tropical Australia. There is a need for
terrestrial records spanning the last interglacial period to the present in the savannas of
northern Australia, to examine in detail the trajectory of environmental change across the
tropics. The obvious location for such records is in the regions that currently receive
>600mm of rainfall each year (Figure 1) – regions that could potentially support a higher
forest cover, but which are currently dominated by woodland and open-forest savanna of

between 20% to 60% cover with scattered patches of closed forest, including currentlyflooded continental shelf areas.

657 Two unexplored possibilities exist. The first are large sinkholes, up to 100m in depth that 658 exist across the Top End of the Northern Territory in various stages of infilling, generally in 659 limestone or dolomitic terranes. The second comprises the many swamps created by basaltic 660 volcanism in northeast Queensland over the last several million years (Stephenson, 1989, 661 Whitehead et al., 2007). There are a large number of untested potential targets that are 662 currently in savanna that could, prior to human occupation, have featured a higher woody 663 cover and potentially of non-eucalypt species (Pole and Bowman, 1996). The deposits in the 664 sinkholes and basalt-dammed swamps contain both charcoal and quartz, making them 665 amenable to numerical dating by radiocarbon and OSL. There are new geochemical 666 techniques that can provide carbon isotope fingerprints for both terrestrial carbon and 667 charcoal, to assess any long-term changes in the balance between C3 trees and C4 grasses in 668 the surrounding environment (Bird and Ascough, 2012, Wurster et al., 2012). Pollen grains 669 and Sporomiella spores can provide information on vegetation and megafaunal abundance 670 over time (Roberts and Brook, 2010, Rule et al., 2012).

The most important periods for detailed examination are not the time intervals shortly before and after human arrival, but the comparatively wet periods of MIS-5 and MIS-4, prior to human arrival, and the post-glacial period after human colonization, when climate was broadly similar. The long-term maintenance of savanna vegetation would imply little impact of human occupation, whereas the existence of forest at such sites prior to human arrival would provide further evidence of substantial human impact.

A further source of high-resolution climate proxy information is speleothems, and such records might be brought to bear on the issue of climate variability and megafaunal extinction in MIS-3, in particular. Speleothem records have proved remarkably valuable in palaeoclimate studies in southern Australia (e.g. Ayliffe *et al.*, 1998) and elsewhere (e.g. Weninger and Jöris, 2008), but the potential for long-term climate records from tropical Australian speleothems has yet to be extensively investigated.

683 (ii) archaeology and anthropology – testing the two hypotheses presented above requires 684 a more complete understanding of the evolution and distribution of population numbers in 685 prehistory to determine the plausibility of the 'profound impact' hypothesis, in particular. It 686 could also require the revisiting of available archaeological information and the material 687 excavated from sites to, for example, determine local vegetation and climate from the 688 geochemical analysis of biotic discard materials (e.g. bones, mollusc shell). There is 689 considerable space for further research into the use of forest edges across tropical Australia 690 in the past, the dynamics of and rationale behind the indigenous fire regime applied at forest 691 edges, as well as the perception of, and response to, decadal-scale change in stem density and 692 tree recruitment in savannas. This implies collaboration with contemporary traditional 693 owners to explore the meaning, uses and value of modern forest and savanna areas across 694 the tropical north (Bliege Bird et al., 2008).

(iii) environmental biology – understanding the response of tropical vegetation types to
changing climatic conditions on millennial timescales will require a better understanding of
plant, animal and ecosystem responses to fire and water stress, the historical biogeography of
Australia's tropical flora and fauna, and a better understanding of the role of soil physical

and chemical properties in influencing the structure and function of tropical vegetationformations and how soil properties interact with climate to determine vegetation type.

701 (iv) modelling - there have been significant, but largely separate, efforts across several 702 disciplines to model the interactions between vegetation, climate and fire in savannas, both 703 globally and in Australia (e.g. Archibald et al., 2012, Murphy and Bowman, 2012). There have 704 also been efforts to model the evolution of the monsoon in Australia in the Quaternary, to 705 identify the controls underlying monsoon variability in the past, including the impact of 706 vegetation (e.g. Notaro et al., 2011). The work of Archibald et al. (2012), in particular, has, for 707 the first time, explicitly included humans as an agent capable of influencing the fire-708 vegetation-climate system in prehistory. Africa has been occupied by humans for a much 709 greater length of time than Australia and saw the incremental development of human ability 710 to control fire frequency and season over the last million years with the continued existence 711 of mega-herbivores (Koch and Barnosky, 2006). The human ability to manipulate fire 712 frequency and season was imposed 'instantaneously' upon arrival in Australia. Combined 713 with the ensuing rapid loss of megafauna, the resulting trajectory of change in Australia may 714 have been substantially different from that in Africa.

Archibald *et al.* (2012) note that anthropogenic fire may have affected forest expansion and contraction in Africa in the past, but, as yet, there have been no integrated attempts to include the potential millennial-scale feedbacks between anthropogenic fire, vegetation and climate in ecosystem models. This would be a challenging undertaking for tropical Australia, but would potentially enable an assessment of the sensitivity of the vegetation–climate system to perturbations, particularly in those boundary conditions potentially capable of modification through the imposition of an anthropogenic fire regime. 722 In conclusion, we are not able, on the basis of the data currently available, to determine 723 which of the two broad hypotheses outlined above better define the actual trajectory by 724 which we arrived at the vegetation distribution we observe across tropical Australia. The 725 degree to which anthropogenic fire regimes in prehistory may have shaped the modern 726 environment is critical to understanding its trajectory under future anthropogenic climate 727 change. The possibility that human occupation resulted in profound environmental change 728 at the very least suggests that phenomena observed to be operating in the modern 729 environment cannot necessarily be extrapolated into the past, as significant anthropogenic 730 impacts on regional climate and vegetation may have considerably more time depth than is 731 currently supposed. As Midnight Oil (1987) noted, '40,000 years can make a difference to the 732 state of things ...'.

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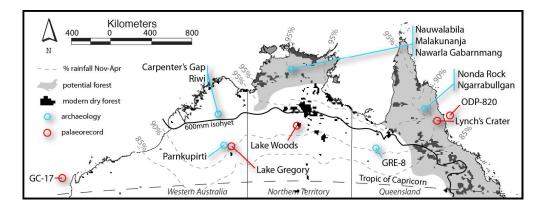
1180 Figure 1: Modern Tropical Australia. Percent of total annual precipitation falling between 1181 November and April shown as grey dashed lines (Bowman, 2002) along with the 600mm 1182 isohyet shown as a solid line. Modern dry forest area taken from Miles et al. (2006) based 1183 on 10km grid cells containing a minimum of 40% forest area. Light shading represents 1184 potential modern forest area based on the modelling of Bond and Keeley (2005). Also 1185 shown are archaeological sites and palaeoenvironmental sites mentioned in the text 1186 (locations of marine cores MD98-2167, SHI-9014 and G6-4, all from the seas off 1187 northwest Australia, are not shown).

Figure 2: Rock art thought to be a representation of *Genyornis newtoni* from Jawoyn country,
western Arnhemland (see Gunn *et al.* 2011 for discussion). Credit: Ben Gunn and Jawoyn
Association.

Figure 3: Example sinkholes in the Northern Territory. Note that scale is variable. (in white:

1192 latitude °S, longitude °E; sinkhole maximum width; land surface elevation AMSL; vertical

distance from land surface to water surface).



184x70mm (300 x 300 DPI)



199x132mm (300 x 300 DPI)



230x154mm (300 x 300 DPI)