

1 **Title:** Enhancing Biodiversity and Ecosystem Services in Quarry Restoration –
2 Challenges, Strategies and Practice

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4 **Running head:** Biodiversity and Ecosystem Services in Quarries

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30 **Abstract:**

31 Although covering less than 1% of the land surface, extraction activities have long-
32 lasting impacts on local ecosystems, inevitably damaging biological diversity and
33 depleting ecosystem services. Many extractive companies are now aware of their
34 impacts and, while pressured by society, demand concrete solutions from researchers to
35 reverse the effects of exploitation and restore biodiversity and ecosystems services. In
36 this paper, we compile and synthesize the contributes of the latest available research on
37 quarry restoration. We depict and discuss some of the most pressing issues regarding (1)
38 the challenges of restoring quarries, (2) the opportunities for biodiversity and ecosystem
39 services delivery, and (3) outline further research addressing current gaps. We conclude
40 that quarries pose different abiotic and biotic constraints that act interdependently,
41 hampering the attainment of effective restoration if considered solely. Such constraints
42 need to be addressed holistically to lastly encourage the self-sustainability of the system
43 by reinstating ecological processes. However, a restored site does not have to
44 specifically mimic the pristine situation, as under certain conditions alternative
45 approaches may uphold valuable natural assets contributing to the conservation of rare,
46 restricted or protected species and habitats.

47

48

49 **Key words:** Ecosystem restoration; Non-energy extractive industry; Spontaneous
50 succession; Assisted restoration; Management practices; Nature-based solutions; Nature
51 conservation

52

53 **Implications for Practice:**

- 54 1) All is connected: soil and landform directly affect productivity and
55 environmental conditions, which will determine vegetation and animal
56 communities able to colonize and thrive in these human-made systems.
- 57 2) Fully functional and self-sustainable ecosystems are usually the aim of active
58 restoration but less advanced stages may also be of restoration interest.
- 59 3) Alternative approaches, such as spontaneous succession, may provide a valuable
60 contribution to nature conservation and ecosystem services, while avoiding
61 expensive restoration or reclamation practices.
- 62 4) Quarrying activity should not prevail over conservation of threatened
63 ecosystems, especially if the post-restoration state cannot compensate for the
64 lost natural assets.

65

66

67 **Introduction**

68 Increasing demand for natural resources has been fueling the growth rates of
69 resource extraction contributing significantly to biodiversity loss (IRP, 2019). Despite
70 the several global agreements (UN, Aichi Targets), biodiversity continues to change
71 globally (Butchart et al. 2010; Dornelas et al. 2014; Tittensor et al. 2014), with ongoing
72 species loss and/or changes in communities (e.g., species turnover, homogenization).
73 Our knowledge is still too limited to understand the exact consequences of these
74 changes for human’s wellbeing and ecosystems resilience, presently and in the future
75 (Branquinho et al., 2019). In response, the United Nations advocated on March 1st,
76 2019, the 2021 – 2030 period as the “Decade on Ecosystem Restoration” (UNEA,
77 2019). Though valuable it may be, it leaves us a worrying sign: is no longer enough to
78 protect, we must also restore!

79 The non-energy mineral extraction sector, for instance, has grown 2.7% and
80 8.3% per year since 1970 for metals and non-metallic minerals, respectively (IRP,
81 2019). Although covering less than 1% of the land surface (Walker, 1999), this sector
82 has critical and long-lasting impacts on local ecosystems. Extraction activities
83 inevitably damage ecosystems resulting in biodiversity loss or change and on the
84 depletion of ecosystem services. Restoration (either spontaneous or assisted) stands as a
85 solution to reverse mining and quarrying impacts, thus contributing to improving the
86 environment and, ultimately, human health (Palmer et al., 2010). However, restoration
87 ecology has not yet reached adulthood as an academic field (Young et al., 2005; Roberts
88 et al., 2009). Although mining and quarry restoration have been targeted in many
89 studies (e.g., Prach and Tolvanen, 2016, and references therein), it still struggles with
90 naïve efforts that may hamper the attainment of effective restoration actions (Cooke et
91 al., 2019). In the brink of the Decade on Ecosystem Restoration, researchers and

92 practitioners must engage with industrial stakeholders to commit with meaningful
93 restoration targets built upon empirical and evidence-based research. This will be vital
94 to define priorities to allocate limited and precious resources into effective and
95 successful restoration actions (Cooke et al., 2019).

96 This was the motto of the Quarries Alive conference held in May 2018 in Évora,
97 Portugal. The aim was to harmonize the current demand for concrete and realistic
98 solutions to restore biodiversity and ecosystems services in the extractive industry. The
99 concept emerged from a consortium between industrial (SECIL, Companhia Geral de
100 Cal e Cimento, S.A.), and scientific stakeholders (University of Évora and the Faculty
101 of Sciences of the University of Lisbon). The conference involved about 150
102 participants (from 21 countries) representing different stakeholders including restoration
103 ecologists, industrial managers and technicians, NGO's and policymakers. Most
104 contributions to this special issue were presented at the conference, which accounted for
105 34 oral presentations and over 20 posters.

106 The collection of papers here presented depict some of the most pressing issues
107 regarding mine and quarry restoration. We assembled contributions focused on the non-
108 energy extractive sector, which include: sand or gravel pits, open-cast quarries of
109 several mineral materials (clay, granite, gypsum, limestone, and others), either active or
110 inactive, while extending it to mine by-product structures such as steep slopes,
111 stockpiles, or spoil heaps. In this paper, we describe and incorporate the main
112 considerations and implications of these studies. We synthesize this information into
113 three major topics:

- 114 • Challenges of restoring quarries: we highlight emerging techniques and
115 approaches to restore quarries in the face of demanding conditions imposed by
116 landform constraints and increased environmental degradation;

- 117 • Quarries as opportunities for biodiversity and ecosystem services delivery: open-
118 minded approaches that look for alternative solutions while taking advantage of
119 novel conditions to promote biodiversity and ecosystem services;
- 120 • Further gaps in the restoration of extractive sites: an overview of issues yet to be
121 addressed or requiring further consideration in future approaches for mining and
122 quarry restoration.

123

124 **The challenges of restoring quarries**

125 It comes as no surprise that mining and quarrying significantly alter local
126 geomorphology, soil, vegetation and fauna. The exploration is traditionally considered
127 to result in a much-degraded site that challenges restoration practitioners to overcome
128 barriers involving unstable substrates, rocky steep slopes, nutrient-poor, acidic or
129 alkaline soils, and a depleted biotic component that extends to collapsed ecosystem
130 services.

131 Whisenant (1999) in his conceptual model (see also Hobbs and Harris, 2001)
132 identifies a first barrier controlled by abiotic limitations (see Fig. 1 for an adaptation of
133 Whisenant's model to mine and quarry restoration). At this stage, restoring soil
134 properties and landforms are necessary to reinstate soil functions, thus preventing from
135 erosion, water drainage or retention, and nutrient leaching, which ultimately will
136 determine vegetation settlement.

137 As an example, Lane et al. (this issue) and collaborators found that mesotrophic
138 grasses dominated former kaolinite mining sites, instead of typical lowland heath
139 species. Such differences in vegetation composition were attributed to soil
140 characteristics, which revealed lower acidity levels and organic content. This evidence
141 was obtained for both short (2 years) and long-term (150 years) restored sites, and even

142 when stockpiled overburden was reinstated. Key nutrients and pH were, therefore,
143 affecting soil fertility and potentiating different vegetation communities at restored
144 sites, leading restoration programs to undershoot their targets. The authors concluded
145 that soil was the limiting factor and, in order to prevent its shortcomings, multiple
146 interventions were advanced to secure soil fertility, such as reducing storage time and
147 depth of overburden and admix organic material to preserve soil mesofauna and
148 microbial properties.

149 Carabassa et al. (this issue) further extended this issue by evaluating the effects
150 of manufactured soils (Technosols) on carbon sequestration and vegetation
151 development. Organic soil amendments, provided by properly treated sewage sludge,
152 revealed to outperform soils without amendments, an effect that echoed until the ten
153 years of experiment. Indeed, Technosols contributed greatly to the development of
154 vegetation and community complexity, speeding up the natural succession processes
155 without altering significantly vegetation composition. As a result, the amount of soil
156 organic carbon sequestered was three times higher due to an increase of primary
157 production supplied by the rich-nutrient content of sludge organic matter.

158 Besides vegetation, overlooking soil properties can have cascading effects on
159 animal communities. Eufrazio et al. (this issue) measured the effects of quarry
160 exploitation and restoration on population dynamics, individual movement, and habitat
161 use of a sand-dwelling beetle (*Scarites cyclops*). While thriving in surrounding areas not
162 subjected to intervention, the beetle showed lower abundance and limited dispersal
163 ability in the restored area. Further investigation places responsibility on inorganic soil
164 amendments and fertilization that altered soil texture from typical fine sand soils (grain
165 size < 4 mm) to gravel (size > 2 mm). In addition, soil amendments readily promoted an

166 abundant and displaced herbaceous cover, which hampered the ability of the species to
167 move and dig for shelter.

168 The interdependent relation between substrate, animal communities and
169 vegetation was also explored by Mexia et al. (this issue) using a multitaxa approach.
170 While looking for differences between restored and reference sites, they found that both
171 the composition and structure of the epigeal beetle community was greatly altered,
172 despite restoration practices efficiently promote a rapid recovery of vegetation cover.
173 Habitat modifications stemming from conifer plantations used to reclaim impacted sites
174 provided further changes on native vegetation while possibly affecting substrate from
175 litter accumulation and concealing natural rock outcrops. Differences on trophic guilds
176 between restored and reference sites could be thus attributable to a different availability
177 of ground food resources.

178 The relation between the abiotic and biotic components is rather complex, and
179 even more challenging when soil is lacking. Steep slopes present highly adverse edaphic
180 conditions and enhanced surface runoff. In the face of these limitations, vegetation
181 seldom establishes in such sites, and slopes are frequently left exposed and unmanaged.
182 An innovative technique that has been proposed consists of vertical greening systems,
183 or Green Walls (Medl et al., 2017). Monteiro et al. (this issue) studied the effectiveness
184 of green walls in steep slopes under different conditions, by assessing vegetation
185 establishment and spontaneous colonization. Their results validate the use of geotextiles
186 to retain the substrate. Even under different environmental constraints, the vegetation
187 evolved similarly, benefiting spontaneous colonization by native species, more adapted
188 to local conditions. Monteiro et al. (this issue) further explored the compositional and
189 functional diversity of vegetation community and found that alleviating limiting
190 conditions (e.g. using irrigation) would only increase functional redundancy. They

191 argue that low-intensity intervention was enough to ensure ecosystem services,
192 moreover being more cost-effective.

193 Another transition threshold is, thus, introduced by a biotic barrier (see Fig. 1).
194 Ecosystem functioning relies on the interplay within and between the biotic and abiotic
195 components of an ecosystem, i.e., ecological processes. Reinstating ecological
196 processes will allow the system to regulate itself without the need for active
197 intervention. With this in mind, Salgueiro V. et al. (this issue) investigated how
198 carnivore mammals could assist spontaneous restoration, while comparing emergence
199 and survival of seeds dispersed by carnivores into a quarry. The authors concluded that,
200 although endozoochorous seedlings showed greater mortality rates, still a high amount
201 of viable seeds survived, contributing to complement restoration efforts. As long as
202 restored areas remain accessible and attractive, species inhabiting the vicinity of
203 quarries can enable seed dispersal during their incursions, thus encouraging the self-
204 sustainability of the system.

205

206 **Quarries as opportunities for biodiversity and ecosystem services delivery**

207 Although mines and quarries are often regarded as biodiversity sinks, a great
208 deal of research up-rises with renewed perspectives claiming valuable natural assets
209 potentiated by post-mining or -quarrying conditions (e.g., Řehouňková et al. 2016).
210 Calvo-Robledo et al. (this issue), for instance, concluded in their study that setting
211 restoration targets that also account for nature conservation may promote ecosystem
212 services in a socio-economic perspective. While using a participatory approach they
213 were able to demonstrate potential ecosystem services provided by restoration scenarios
214 of an active quarry, weighted by cultural (nature-based recreation), regulating
215 (greenhouse gas emission and sequestration), and provisioning services (agricultural

216 production). The associated costs and annual income could be optimized if nature-based
217 solutions are taken into account while planning the post-quarry end-use.

218 Informed decisions, however, often require evidence-based approaches that can
219 advise on alternative solutions to restore and value biodiversity. Some studies figuring
220 this special issue provide approaches that detach from prevailing technical reclamation
221 techniques generally oriented to minimize visual impacts, improve safety protection of
222 the mining infrastructure, or further profit from the forest or crop harvest.

223 For example, Šebelíková et al. (this issue) found that spontaneously revegetated
224 post-mining sand and sand-gravel pits hosted more dry and mesic grassland species
225 considered rare (or threatened) and specialists of open sand habitats in early than late
226 successional stages. Although a large overlap in species composition was detected
227 between spontaneously revegetated and reclaimed sites, most of the conservation
228 interest of the later was lost after a few years following artificial afforestation. The
229 authors suggest that spontaneous revegetation should be considered in post-mining
230 areas, as long as nearby (semi-)natural habitats can source vegetation propagules into
231 mined sites. Řehouňková et al. (this issue) support this approach as well. In their study,
232 they assessed vegetation communities of 321 post-mining sites at different successional
233 stages. They identified 235 threatened plant species overall, concluding that post-
234 mining sites act as important refugia for the species. Moreover, early successional
235 habitats offer the most important conditions for threatened plants, which gradually
236 degrade with increasing woody species cover.

237 Not only plants, but also animal species may thrive in such degraded habitats.
238 Rohrer et al., Salgueiro P. et al. and Martin-Collado et al. (all in this issue) studied the
239 habitat preferences of two cliff-nesting bird species – the Sand Martin (*Riparia riparia*)
240 and the Black Redstart (*Phoenicurus ochruros*) – and a semiaquatic mammal – the

241 Eurasian otter (*Lutra lutra*). In the first two studies, both authors found that the species
242 were attracted to mined areas especially when unsuitable conditions prevailed in
243 surrounding matrix habitats. Sand Martins, for example, benefited from sand stockpiles
244 of extracted material or sandy faces of gravel pits, which were used to burrow their
245 nests. Black Redstart, on the other hand, occurred in limestone exposed slopes that
246 mimic natural steep cliffs of rocky habitats, where it naturally occurs. Regarding the
247 Eurasian otter, the authors found that gravel pit lagoons could provide suitable habitats
248 for semiaquatic species in anthropogenic landscapes, if restoration attends to certain
249 environmental features. These studies discuss the possibility of creating or enhancing
250 habitat by allowing for the preservation of novel elements in the landscape potentiated
251 by mining and quarrying activities. Such practices may also apply in the course of
252 mining operations, providing temporary habitats. This approach, despite feasible, still
253 raises some debate. One of the most relevant questions regards the promotion of
254 habitats surrounded by low-quality areas, which in the event of not providing enough
255 resources may create an ecological trap, such as discussed by Rohrer and co-workers.

256 Řehouňková et al. (this issue) refer in their study that post-mining sites,
257 especially if left to spontaneous restoration, can operate as secondary or surrogate
258 habitats for species of conservation concern. The approaches presented here can be
259 alternatives to inadequate restoration practices (Perring et al. 2013), and can contribute
260 to maximize local biodiversity (Doley & Audet 2013). However, all authors stress that
261 these alternatives are conditional and site-dependent, and may only happen ‘under
262 certain conditions’. In fact, these alternatives and opportunities may hinder or conflict
263 with some reclamation approaches.

264

265 **Further gaps in mining and quarry restoration deserving attention**

266 Mining and quarrying sites are increasingly attracting the attention of ecologists, and
267 though restoration processes are being intensively debated, much knowledge is still
268 lacking. We highlight three major issues that deserve a future commitment from
269 researchers, practitioners and other stakeholders:

270 1) Measuring the Net Impact of restoration (the gains and losses) has been a major
271 liability of most projects. Many of them have limited insight on ecosystem
272 attributes that can produce objective measures of restoration success, mostly
273 relying on vegetation structure, diversity, or indirect measures of ecological
274 processes (Ruiz-Jaen and Aide, 2005). Part of the problem relies on the
275 dimensionality of the biodiversity concept, which hampers the attainment of
276 objective measures that integrate all dimensions (Nakamura et al. 2019). Mining
277 and quarry restoration share the same problem alike. Furthermore, many
278 extractive companies already acknowledge their impacts on ecosystems and
279 biodiversity (assisted by increasing awareness and pressure from society), and
280 demand concrete solutions from researchers to disclose their efforts in restoring
281 degraded sites. Under these circumstances, developing holistic approaches
282 showing the relation between both biotic and abiotic components and how they
283 intertwine into ecological processes and socio-economic circumstances is crucial
284 to address the Net Impacts of extractive industry. In this matter, cost-benefit
285 analysis may assist weighing the Net Impact of exploitation.

286 2) Several studies in this issue showed that preserving or enhancing unique
287 characteristics of quarries can sustain threatened and other conservation interest
288 species or communities, if properly managed. However, broader scales should
289 also be considered. Throughout Europe many habitats of conservation concern
290 faced recent declines (e.g., calcareous grasslands, Wallis de Vries et al. 2002,

291 Ödman and Olsson, 2014) or are currently receding, causing important declines
292 in species, communities and ecological processes associated with them. If
293 mining or quarrying areas hold the potential to sustain such habitats, they can
294 complement existing protected habitats when oriented by suitable restoration
295 practices. The establishment of a strategic network connecting these areas as
296 stepping-stones should enable species to move across the landscape (Saura et al.
297 2014). Such an approach is considered within the Green Infrastructure concept.
298 The adoption of a Green Infrastructure Agenda can uphold the importance of
299 former mining or quarrying sites as reservoirs of specific ecosystems. Therefore,
300 spatial and temporal approaches at broader scales are necessary to design those
301 networks and set targets that guide restoration practices into wide-ranging
302 contexts.

303 3) Finally, as many studies preview in this issue, restoring mine sites or quarries
304 are among the most challenging tasks, given their starting point. In a climate
305 change scenario, the expected changes are likely to have overarching effects on
306 restored sites, mainly upon the soil, water availability and vegetation
307 establishment, and on the processes they are involved in. Many approaches are
308 now available that can simulate future climate change scenarios at micro- and
309 meso-scales (see Maestre et al., 2013, León-Sánchez et al. 2017). Such
310 approaches are of interest for companies involved in restoration activities
311 willing to adapt their practices to future climate forecasts, thus improving habitat
312 resilience.

313

314 **Concluding remarks**

315 As the demand for mineral extraction grows, restoring mines and quarries
316 emerges as a relevant issue to respond to both stakeholders and general society
317 concerns. Researchers and practitioners are being challenged to provide informed and
318 efficient solutions to respond to both ecological and economical demands.

319 This special issue on “Enhancing Biodiversity and Ecosystem Services in
320 Quarry Restoration – Challenges, Strategies and Practice” resumes several topics that
321 address the best guidance practices to restore such degraded sites. A few key practices
322 reported are as follows:

- 323 1) First interventions usually focuses on landform and soil fertility. Storage
324 of stockpiles, soil amendments, and landform will have direct impacts on
325 soil quality and edaphic conditions, which in turn will determine soil
326 biota, vegetation and animal communities able to colonize and thrive in
327 these systems. Furthermore, the soil will also provide some ecosystem
328 services.
- 329 2) However, there are also more and more evidence that low-fertility sites
330 with rough substrates can be profitable for some rare and retreating
331 specialists, both plants and animals. Thus, restoration measures that will
332 strongly change site environmental conditions should be carefully
333 considered before any decision is made.
- 334 3) Analyses of community composition and structure allow the assessment
335 of effectiveness of the restoration practices. Using local native species
336 while taking advantage of (semi-)natural conditions on the vicinity can
337 unequivocally improve restoration success.
- 338 4) Assessing how communities are performing their functions will further
339 contribute to understanding if ecological processes are being upgraded.

340 Only a fully functional system can be self-sustainable, and will further
341 allow for species to play their part in assisting restoration.

342 5) Under certain conditions, alternative approaches, namely those relying
343 on nature-based solutions (e.g., spontaneous or only slightly assisted
344 restoration), may uphold valuable natural assets in post-mining or -
345 quarrying sites, avoiding expensive (and sometimes inadequate) classical
346 restoration or reclamation practices, and further upholding ecosystem
347 services.

348 6) Nonetheless, mining or quarrying activity should not prevail over
349 threatened ecosystems or of conservation concern, especially when the
350 conditions created afterwards through restoration cannot match the lost
351 natural assets.

352

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364

365 **Literature Cited:**

- 366 Branquinho C, Serrano H, Nunes A, Pinho P, Matos P (2019) Essential biodiversity
367 change indicators for evaluating the effects of Anthropocene in ecosystems at a
368 global scale. In Casetta E, Vecchi D, da Silva JM (eds.) From Assessing to
369 Conserving Biodiversity Conceptual and Practical Challenges. Springer, 137-163.
- 370 Butchart SHM, Walpole M, Collen B, van Strien A, Scharlemann JPW, Almond REA et
371 al. (2010) Global biodiversity: indicators of recent declines. *Science* 328: 1164–
372 1168.
- 373 Cooke SJ, Bennet JR, Jones HP (2019) We have a long way to go if we want to realize
374 the promise of the “Decade on Ecosystem Restoration”. *Conservation Science and
375 Practice*. 2019;e129
- 376 Doley D, Audet P (2013) Adopting novel ecosystems as suitable rehabilitation
377 alternatives for former mine sites. *Ecological Processes* 2:22
- 378 Dornelas M, Gotelli NJ, McGill B, Shimadzu H, Moyes F, Sievers C, Magurran AE
379 (2014). Assemblage Time Series Reveal Biodiversity Change but Not Systematic
380 Loss. *Science* 344: 296-299.
- 381 Hobbs, R.J., and J.A. Harris (2001). ‘Restoration ecology: repairing the Earth’s
382 ecosystems in the new millennium’. *Restoration Ecology* 9: 239–246.
- 383 IRP (2019). *Global Resources Outlook 2019: Natural resources for the future we want*.
384 Oberle B, Bringezu S, Hatfeld-Dodds S, Hellweg S, Schandl H, Clement J et al. A
385 Report of the International Resource Panel. United Nations Environment
386 Programme. Nairobi, Kenya.
- 387 León-Sánchez L, Nicolás E, Goberna M, Prieto I, Maestre FT, Querejeta JI (2017) Poor
388 plant performance under simulated climate change is linked to mycorrhizal
389 responses in a semi- arid shrubland. *Journal of Ecology* 106:960–976.

390 Maestre FT, Escolar C, Guevara ML, Quero JL, Lázaro R, Baquerizo MD, Ochoa V,
391 Berdugo M, Gozalo B, Gallardo A (2013) Changes in biocrust cover drive carbon
392 cycle responses to climate change in drylands. *Global Change Biology*
393 19(12):3835-3847.

394 Medl A, Stangl R, Florineth F (2017a) Vertical greening systems – A review on recent
395 technologies and research advancement. *Building and Environment* 125:227-239

396 Nakamura G, Gonçalves LO, Duarte LS (2019) Revisiting the dimensionality of
397 biological diversity. *Ecography*, 00: 1–10

398 Ödman AM, Olsson PA (2014) Conservation of Sandy Calcareous Grassland: What
399 Can Be Learned from the Land Use History? *PLoS ONE* 9(3): e90998.

400 Palmer MA, Bernhardt ES, Schlesinger WH, Eshleman KN, Foufoula-Georgiou E,
401 Hendryx MS et al. (2010) Mountaintop mining consequences. *Science* 327:148 –
402 149.

403 Perring MP, Standish RJ, Hobbs RJ (2013) Incorporating novelty and novel ecosystems
404 into restoration planning and practice in the 21st century. *Ecological Processes*
405 2:18.

406 Prach K, Tolvanen A (2016) How can we restore biodiversity and ecosystem services in
407 mining and industrial sites? *Environmental Science and Pollution Research* 23:
408 13587–13590.

409 Řehouňková K, Čížek L, Řehounek J, Šebelíková L, Tropek R, Lencová K, Bogusch P,
410 Marhoul P, Máca J (2016) Additional disturbances as a beneficial tool for
411 restoration of post-mining sites: a multi-taxa approach. *Environmental Science*
412 *and Pollution Research* 23:13745–13753

413 Roberts L, Stone R, Sugden A (2009) The Rise of Restoration Ecology. *Science*
414 325(5940):555.

415 Ruiz-Jaen MC, Aide TM (2005) Restoration Success: How Is It Being Measured?
416 Restoration Ecology 13(3):569–577.

417 Saura, S., Bodin, Ö., Fortin, M.J., 2014. Stepping stones are crucial for species' long-
418 distance dispersal and range expansion through habitat networks. J. Appl. Ecol.
419 51(1), 171–182

420 Tittensor DP, Walpole M, Hill SLL, Boyce DG, Britten GL, Burgess ND et al. (2014) A
421 mid-term analysis of progress towards international biodiversity. Science 346:
422 241-244

423 UNEA (2019) United Nations Environment Agency, Press Release, March 1, 2019.
424 New UN Decade on Ecosystem Restoration offers unparalleled opportunity for
425 job creation, food security and addressing climate change.
426 [https://www.unenvironment.org/news-and-stories/press-release/new-un-decade-](https://www.unenvironment.org/news-and-stories/press-release/new-un-decade-ecosystem-restoration-offers-unparalleled-opportunity)
427 [ecosystem-restoration-offers-unparalleled-opportunity](https://www.unenvironment.org/news-and-stories/press-release/new-un-decade-ecosystem-restoration-offers-unparalleled-opportunity)

428 Walker LR (ed) (1999) Ecosystems of disturbed ground. Ecosystems of the world, vol
429 16. Elsevier, Amsterdam

430 Wallis de Vries, M.F., Poschlod, P., Willems, J.H., 2002. Challenges for the
431 conservation of calcareous grasslands in northwestern Europe: integrating the
432 requirements of flora and fauna. Biol. Conserv. 104, 265–273.

433 Whisenant, S. G. 1999. Repairing Damaged Wildlands: A process-oriented, landscape-
434 scale approach. Cambridge University Press, Cambridge.

435 Young TP, Petersen DA, Clary JJ (2005) The ecology of restoration: historical links,
436 emerging issues and unexplored realms. Ecology Letters 8: 662–673
437

438 **Figure caption**

439

440 Figure 1 – Classical conceptual model of state transition along a degraded to restored
441 gradient targeting mining and quarry restoration (adapted from Whisenant 1999, Hobbs
442 and Harris, 2001). Horizontal dashed line signals the different hypothetical equilibrium
443 states along the gradient; diagonal dashed line signals transition between states. Each
444 state transits to a more or less functional system (with increasing complexity of the
445 community dimensions) according to the interventions that promote either restoration or
446 degradation, respectively. Interventions can be grouped at three levels, according to
447 their aim and extent: (1) physical and/or chemical when the site requires landform or
448 soil preparation to assist revegetation, (2) revegetation to encourage the settlement of
449 communities attending to both composition and structure, either spontaneously or
450 actively intervening, and (3) targeted or focal, in order to allow for naturally assisted
451 restoration through reinstating ecological processes. Between each level, abiotic
452 (between level 1 and 2) or biotic (between level 2 and 3) barriers exist, marking
453 thresholds which need to be overcome to assist subsequent interventions. The intensity
454 of these thresholds may vary between site condition and overall restoration target.
455 Overcoming these barriers can be either promoted by spontaneous (supported by
456 ecological succession when barriers can be naturally overcome) or active intervention
457 (if barriers are naturally constrained). In this conceptualization, a restored site does not
458 have to specifically mimic the pristine situation, but to be fully functional, i.e., hold
459 self-sustainable communities at long-term while requiring minimal or no intervention.
460
461

462 Figure 1

